

Quantifying aggregation bias in marsh carbon flux estimates caused by rhizosphere oxygen heterogeneity

Youssef Saadaoui¹, Christian Beer^{2,3}, Peter Mueller⁴, Friederike Neiske^{2,3}, Joscha N. Becker^{2,3}, Annette Eschenbach^{2,3}, Philipp Porada^{1,3}

¹Universität Hamburg; Faculty of Mathematics, Informatics and Natural Sciences; Department of Biology; Institute of Plant Science and Microbiology, Ohnhorststr. 18, 22609 Hamburg, Germany

²Universität Hamburg, Faculty of Mathematics, Informatics and Natural Sciences, Department of Earth System Sciences; Institute of Soil Science, Allende-Platz 2, 20146 Hamburg, Germany

³Universität Hamburg, Center for Earth System Research and Sustainability, Bundesstraße 53, 20146 Hamburg, Germany

⁴University of Münster, Institute of Landscape Ecology, Heisenbergstr. 2, 48149 Münster, Germany

Correspondence to: Youssef Saadaoui (youssef.saadaoui@hereon.de)

Abstract. Oxygen (O₂) in marsh rhizospheres varies strongly at millimetre scales due to radial oxygen loss (ROL), hydrology, and sediment structure. Yet many models estimate heterotrophic (microbial) respiration at a single mean O₂, which can bias CO₂ fluxes because the O₂ response of respiration is curved (nonlinear). We calibrated an enzyme kinetic respiration function to marsh-soil incubations, then compared (I) respiration evaluated at a single mean O₂ value with (II) the mean of different respiration fluxes over a prescribed rhizosphere O₂ distribution. To isolate oxygen effects, temperature and the relation between substrate and soil moisture were held fixed; incubations were used only to calibrate carbon kinetics. For a gas-phase, air-equivalent O₂ proxy, the single mean approach overestimates respiration by ~11.7% on average. This bias increases with increasing variation of O₂ and with the ratio of mean O₂ to the O₂ half-saturation constant K_{M,O₂}. In dissolved-phase, mean O₂ ≪ K_{M,O₂} and variation in O₂ is small, so the bias is negligible; when physiology is unit matched to similar relative saturation, however, the dissolved case shows a negative bias comparable to the gas-phase one. Our findings suggest reporting the mean and variation of O₂ together with a calibrated K_{m,O₂} allows modellers to anticipate or correct the bias associated with replacing heterogeneous rhizosphere conditions with a single mean.

1 Introduction

Marsh soils are major coastal carbon reservoirs where high primary production and efficient sediment trapping coincide with comparatively slow decomposition (Barbier et al., 2011; Duarte et al., 2013; Granse et al., 2024; Kirwan et al., 2016). Whether these systems store or release carbon depends not only on mean environmental conditions but also on the availability of oxygen (O₂) to microbes in the rhizosphere, which regulates aerobic heterotrophic respiration.

Rhizosphere O₂ is inherently patchy at millimetre scales. Radial oxygen loss (ROL) from roots varies across species and genotypes and with plant status (growth stage, stress), while aerenchyma formation and barriers to ROL modulate internal gas pathways. Hydrologic dynamics, such as tidal flooding and drainage, platform elevation, and episodic drying interact with sediment texture to shape air-filled porosity and diffusion pathways. In situ microelectrode and optical measurements consistently show steep radial O₂ gradients around roots and strong small scale variability within and across root

neighbourhoods (Kühl et al., 2021; Mueller et al., 2016; Van Der Nat and Middelburg, 1998; Visser et al., 2000). Thus, microbes experience a mosaic of microhabitats ranging from well oxygenated to effectively anoxic over centimetres or less. In such a setting, the curved, nonlinear shape of the O₂ response of respiration may lead to differing estimates depending on the level of aggregation of O₂ conditions.

Despite this heterogeneity, many ecosystem and land-surface models represent belowground redox drivers with a single bulk state, for example a mean O₂ level or a uniform moisture/redox scalar within a grid cell or plant functional type (Alizad et al., 2016; Kirwan et al., 2016; Morris et al., 2002; Rietl et al., 2021; Swanson et al., 2014; Zhang et al., 2002). Under a concave O₂-limitation function (saturating, Michaelis-Menten-type), evaluating respiration at the mean O₂ generally differs from averaging respiration across a heterogeneous O₂ field, a classic aggregation (upscaling) bias arising from response curvature (Jensen's inequality) (Rastetter et al., 1992). For marsh soils, strong O₂ heterogeneity is well documented and concave limitation is physiologically expected, yet the magnitude, direction, and conditions of the resulting bias remain poorly quantified (Bernal et al., 2017; Bhattacharyya and Furtak, 2022; Mueller et al., 2020; Spivak et al., 2019).

Here we quantify the aggregation bias associated with spatially heterogeneous rhizosphere O₂, focusing explicitly on aerobic heterotrophic (microbial) respiration in the root zone. While soil respiration includes both autotrophic (root) and heterotrophic (microbial) components, we focus on the heterotrophic component since the aggregation bias arises from the nonlinear O₂ limitation of microbial respiration. The drivers of autotrophic respiration, including climatic conditions and plant physiology, are outside our scope. Moreover, we do not model explicitly plant internal gas transport, or pore scale advection-diffusion. Instead, we treat ROL and hydrology as drivers that shape local O₂ distributions and ask how that heterogeneity propagates through a standard enzyme kinetic respiration formulation with an explicit O₂-limitation term (cf. Davidson et al., 2012 for the general structure). To isolate oxygen effects, we calibrate carbon kinetic parameters to marsh soil incubations and hold temperature and the substrate moisture linkage fixed.

We represent rhizosphere O₂ using two complementary proxies: (I) a gas-phase, air-equivalent proxy that captures aeration via air-filled porosity; and (II) a dissolved-phase proxy in native aqueous units. For like-for-like comparisons, we also analyse a diagnostic in which the microbial O₂ half-saturation constant K_{M,O_2} unit-matched so that the relative saturation $r = \text{mean}(O_2) / K_{M,O_2}$ equals that of the gas case. We define the uniform baseline in O₂ space as $f(\text{mean}(O_2))$ (rather than mapping mean moisture to O₂), then quantify bias as the difference between this uniform evaluation and the mean of respiration over the full O₂ distribution.

We address three questions. (Q1) How large, and in which direction is the bias when O₂ heterogeneity is replaced by a single mean? (Q2) How does the bias scale with relative saturation and with the variance of O₂ at fixed mean? (Q3) Under what conditions is the bias negligible, providing a simple rule for when single mean treatments are acceptable? Answering these provides a practical diagnostic for modellers: reporting the mean and variance of O₂ together with a prescribed K_{M,O_2} is sufficient to anticipate, and if needed correct, aggregation error in marsh carbon balance models.

2 Materials and methods

2.1 Model overview: DAMM with explicit O₂ limitation

We use the DAMM framework (Davidson et al., 2012) in which volumetric aerobic CO₂ production R_{sx} (mg C cm⁻³ h⁻¹) depends on temperature, soluble-carbon substrate, and oxygen:

$$R_{sx}(c_{org}, O_2, T) = \alpha_{sx} \times e^{\frac{-Ea_{sx}}{R_{gas}T}} \times \frac{S_x}{K_{m,s} + S_x} \times \frac{[O_2]}{K_{m,o_2} + [O_2]} \text{ Eq (1)}$$

Here C_{org} is bulk organic C (g C cm⁻³); S_x is the soluble substrate concentration (g C cm⁻³); O_2 is an air-equivalent oxygen proxy expressed as (cm³ O₂ cm⁻³); T is absolute temperature (K); α (mg C cm⁻³ h⁻¹) and Ea (J mol⁻¹) are the pre-exponential factor and activation energy; $K_{M,S}$ (g C cm⁻³) and K_{M,O_2} (cm³ O₂ cm⁻³) are half-saturation constants; and R_{gas} is the universal gas constant. Symbols, units, and baseline values are listed in Table 1.

To isolate O_2 effects in the aggregation experiment, soluble carbon is linked to bulk carbon via a fixed moisture proxy:

$$S_x = c_{org} \times p \times D_{liq} \times \theta_{ref}^3 \text{ Eq (2)}$$

where p (–) converts bulk to soluble carbon, D_{liq} (–) scales aqueous availability, and θ_{ref} volumetric water content used only inside the substrate term (to isolate oxygen effects). The oxygen-limitation term $O_2 / (K_{m,o_2} + O_2)$ is concave for $O_2 > 0$; this curvature is the source of aggregation bias examined below.

2.2 Incubation medians and calibration

We calibrated α and $K_{M,S}$ against median aerobic respiration rates derived from long-term marsh-soil incubations summarized with a two-pool ICBM framework (ICBM; (André et al., 1997; Beer et al., 2022; Knoblauch et al., 2013)). Incubation rates were converted from $\mu\text{mol CO}_2 \text{ g}^{-1} \text{ d}^{-1}$ to $\text{mg C cm}^{-3} \text{ h}^{-1}$ using soil bulk density and molar-mass factors. Parameters were estimated by bounded constrained optimization (SQP; MATLAB *fmincon*) in $\log_{10}$ space with box constraints, minimizing the RMSE between model predictions and the binned medians, at a fixed reference oxygen and $O_{2,ref}$ fixed water content θ_{ref} . The reference oxygen was computed from the gas-phase proxy defined in Sect. 2.3 evaluated at θ_{ref} . Uncertainty was quantified by non-parametric bootstrap resampling of the (C_{org} , R) pairs ($B = 500$). We report 95% pointwise bands for the calibration curve, confidence intervals for α and $K_{M,S}$, and the Spearman correlation ρ between $\log_{10} \alpha$ and $\log_{10} K_{M,S}$.

2.2.1 Incubation experiments

Marsh soils from the Elbe estuary (Germany) were gently sieved to 2 mm. For each incubation, 20 g dry-mass equivalent soil was weighed (wet masses corrected to dry-mass equivalence). The samples were then adjusted to 60 % water-holding capacity (WHC) and placed in 1,000 mL glass flasks. Incubations were conducted in the dark at 20 °C for up to 465 days. At regular intervals, 1 mL headspace samples were withdrawn and analyzed for CO₂ by gas chromatography (JAS 6890N, Germany). To maintain aerobic conditions, we kept a slight overpressure and added synthetic air whenever headspace CO₂ exceeded 2.5 %, preventing O_2 depletion. In the context of our study, these incubations provide aerobic respiration rates across a broad C_{org} range for Elbe-estuary marsh soils; they do not impose or resolve O_2 heterogeneity. We use the rates only to calibrate carbon-kinetic parameters; O_2 effects are controlled separately in the aggregation analyses (Sect. 2.3–2.6).

2.2.2 From CO₂ time series to characteristic rates (ICBM fits)

CO₂ time series were fit with the two-pool ICBM (fast/slow pools, first-order kinetics, humification fast to slow) using gradient-descent optimization to estimate pool decay and humification parameters against the GC data. For each of seven C_{org} bins, we computed the median CO₂ rate (μmol CO₂ g⁻¹ d⁻¹), then converted to mg C cm⁻³ h⁻¹ as above. These bin medians are the only empirical constraints used to fit α and K_{M, s} at fixed O_{2,ref} and θ_{ref} , thereby isolating oxygen effects in subsequent experiments.

2.3 Prescribed rhizosphere O₂ distributions

We prescribe two rhizosphere O₂ drivers and summarize each by its mean and variance.

▪ Gas-phase proxy (primary analysis).

Air-filled porosity a is approximated as

$$a = 1 - \frac{BD}{PD} - \theta \quad \text{Eq 3}$$

where θ is volumetric water content and BD and PD are bulk and particle densities (g cm⁻³), respectively. Air-equivalent oxygen is then

$$O_2^{gas} = D_{gas} \times 0.209 \times a^{\frac{4}{3}} \quad \text{Eq 4}$$

where D_{gas} (–) represents effective gas-phase diffusivity and 0.209 the atmospheric O₂ mole fraction. This proxy captures aeration in the rhizosphere and is used for all main analyses.

▪ Aqueous-phase proxy (Henry's-law dissolved O₂)

$$O_2^{aq} = \theta D_{O_2} \times C_{O_2}^* \quad \text{Eq (4b)}$$

Where D_{O_2} is the molecular diffusivity of O₂ in water (m² s⁻¹) and $C_{O_2}^* = H_{cp} P_{O_2}^{atm}$ is the henry-law equilibrium concentration (mol m⁻³) at atmospheric partial pressure. We do not rescale dissolved O₂; we retain its native magnitude relative to K_{M,O₂}:

Unit-matched physiology for fair comparison (diagnostic).

in a diagnostic used for Sect. 3.4 and SI, we keep O_2^{aq} unaltered and instead unit-match the microbial half-saturation so that relative saturation is identical across phases:

$$K_{M,O_2}^{aq,UM} = K_{M,O_2}^{gas} \times \frac{\mu(O_2^{aq})}{\mu(O_2^{gas})}$$

150 This aligns $\mu_{O_2} / K_{M,O_2}$ between gas and dissolved cases without artificially rescaling O_2 .

How distributions are obtained and summarized.

We map a uniform grid θ values to O_2 using Eqs. 3 and 4b to obtain discrete O_2 samples. In the main text we use $\theta \in [0.02, 1 - (BD / PD) - 0.02]$; as a high- θ benchmark, we also consider a narrow saturated $\theta \in [0.65, 0.68]$ (both mapped
155 through the same proxy). Each driver is summarized by its first two moments,

$$\mu_{O_2} = E [O_2], \quad \text{Var}(O_2) = E [(O_2 - \mu_{O_2})^2],$$

which are used later for use in the Jensen diagnostic.

160 2.4 Uniform-versus-heterogeneous experiment (aggregation error)

To isolate the oxygen effect, the substrate S_x is held fixed by evaluating it at a single θ_{ref} in all runs. For each C_{org} on a prescribed grid we compute two configurations of the same calibrated DAMM model:

- **Uniform configuration:**
165 $R_{\text{uni}}(C_{\text{org}}) = f(C_{\text{org}}, \mu_{O_2})$,
where $\mu_{O_2} = E [O_2]$ is the arithmetic mean of the imposed O_2 distribution
- **Heterogeneous configuration:**
 $R_{\text{het}}(C_{\text{org}}) = E [f(C_{\text{org}}, \overline{O_2})]$,
i.e., the arithmetic mean of the respiration response over the same O_2 distribution, with C_{org} and θ_{ref} held fixed.

170 Aggregation bias (percent) is calculated as:

$$\Delta_s(C_{\text{org}}) = \frac{R_{\text{het}}(C_{\text{org}}) - R_{\text{uni}}(C_{\text{org}})}{\max(R_{\text{uni}}(C_{\text{org}}), \epsilon)} \times 100\% \text{ Eq (5)}$$

175 With $\epsilon = 10^{-12} \text{ mg C cm}^{-3} \text{ h}^{-1}$ used only to avoid division by zero (in practice $R_{\text{uni}} \gg \epsilon$). In the implementation, the expectation $E[.]$ is a discrete average over the vector of O_2 values obtained by mapping θ -grid through the chosen proxy.

Crucially, the uniform baseline is defined in O_2 space as $f(C_{\text{org}}, \mu_{O_2})$, to avoid baseline reversal when the $\theta \rightarrow O_2$ mapping is nonlinear. Unless stated otherwise, main text analyses use the gas-phase driver; dissolved-phase analyses appear as
180 sensitivity/diagnostics, and where indicated also with unit matched K_{M,O_2} to equalize relative saturation $\mu_{O_2} / K_{M,O_2}$.

2.5 Dependence on O_2 affinity and transport distributions

Since K_{M,O_2} sets the relative O_2 -saturation regime, we performed a structured K_{M,O_2} variation (multipliers $\times 0.25, \times 0.5, \times 1, \times 2, \times 4$) for the gas proxy and for the aqueous proxy with unit-matched K_{M,O_2} . We further summarize bias behaviour on a response

185 surface as a function of $\text{mean}(\text{O}_2)/K_{M, \text{O}_2}$ and the coefficient of variation (CV) using a symmetric two-point distribution at fixed mean (reported in the SI).

2.6 Jensen diagnostic (mechanistic check)

190 To link bias magnitude to heterogeneity directly, we fixed the mean oxygen at $\mu = \overline{\text{O}_2}$ and constructed symmetric two-point mixtures $\{\mu - d, \mu + d\}$ (equal probability; range constrained to positive values) to vary $\text{Var}[\text{O}_2]$ while holding the mean constant. For each variance level we computed Δ from Sect. 2.5 and compared it to the quadratic approximation implied by the curvature of the O_2 -limitation term:

$$\Delta(d) \approx \frac{1}{2} \times \frac{f''(\mu) - \text{Var}[\text{O}_2(d)]}{f(\mu)} \times 100\% \text{ Eq (5)}$$

195 Where $f''(\mu)$ is the second derivative of f with respect to O_2 at fixed C_{org} and T .

2.7 Implementation details

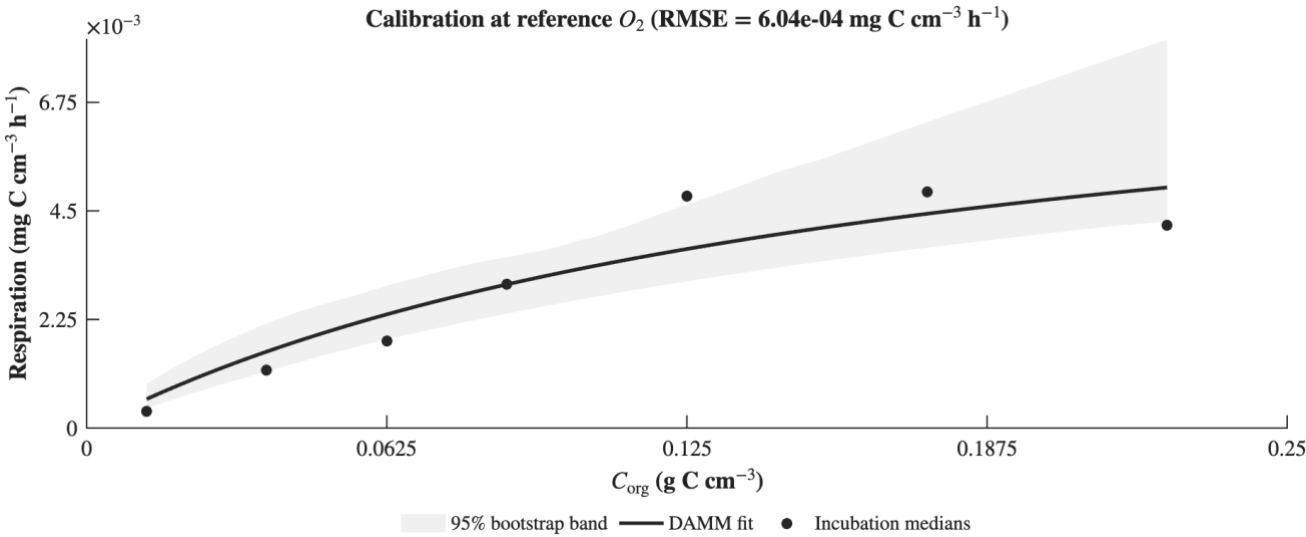
200 All analyses were performed with consistent grid resolution for C_{org} (300 points across 0.001 - 0.30 g C cm⁻³). Calibration used bounded least squares; bootstrap size was 500; a fixed random seed was used for reproducibility.

3 Results

3.1 Model calibration and parameter uncertainty

205 The respiration formulation calibrated well against incubation medians across the observed range of C_{org} (Fig. 1). Best-fit parameters are: $\alpha = 1.48 \times 10^{11}$ mg C cm⁻³ h⁻¹ (95 % CI: $1.04 \times 10^{11} - 9.93 \times 10^{13}$) and $K_{M,S} = 5.98 \times 10^{-6}$ g C cm⁻³ (RMSE = 6.04×10^{-4} mg C cm⁻³ h⁻¹), bootstrap band ($B = 500$) around the calibration curve widens at high C_{org} where constraints are sparse. A few medians fall outside the ribbon because it reflects uncertainty in the fitted curve, not confidence intervals on binned medians. Incubations were aerobic at 20 °C and span C_{org} ; these medians calibrate the carbon-kinetic terms only. In all aggregation experiments, temperature and the substrate–moisture linkage are held fixed to isolate oxygen effects; only the treatment of O_2 variability differs.

210



215

Figure 1: Calibration of the DAMM respiration model to incubation-derived respiration medians under aerobic conditions. Points show median rates for seven bins of bulk organic carbon (C_{org}) obtained from a two-pool ICBM fitted to Elbe estuary incubations (units converted to $\text{mg C cm}^{-3} \text{ h}^{-1}$). The solid line is the best DAMM fit evaluated at the air-equivalent O_2 corresponding to $\theta_{ref} = 0.30$; the shaded envelope is the 95 % bootstrap band ($B = 500$). Best-fit parameters: $\alpha = 1.475 \times 10^{11} \text{ mg C cm}^{-3} \text{ h}^{-1}$, $K_{M,S} = 5.98 \times 10^{-6} \text{ g C cm}^{-3}$; $\text{RMSE} = 6.04 \times 10^{-4} \text{ mg C cm}^{-3} \text{ h}^{-1}$. Parameter confidence intervals and provenance are summarized in Table 1.

Table 1. Parameters used in the DAMM analyses (fixed and fitted).

All constants are baseline values unless noted; fitted parameters carry 95 % bootstrap CIs.

220

Symbol	Description	Value / Estimate	Units	Status / Source
T	Temperature	293.15	K	Fixed at 20 °C
Ea	Activation energy	72.26×10^3	J mol^{-1}	Fixed; Davidson et al. (2012)
K_{M,O_2}	O_2 half-saturation	0.121	$\text{cm}^3 O_2 \text{ cm}^{-3}$ (air-equiv.)	Fixed; Davidson et al. (2012)
p	Soluble C fraction	4.14×10^{-4}	—	Fixed; Davidson et al. (2012)
D_{liq}	Solute availability scalar	3.17	—	Fixed; Davidson et al. (2012)

D_{gas}	O ₂ diffusivity scale	1.67	—	Fixed; Davidson et al. (2012)
BD	Bulk density	0.80	g cm^{-3}	Fixed; Davidson et al. (2012)
PD	Particle density	2.52	g cm^{-3}	Fixed; Davidson et al. (2012)
θ_{ref}	Reference water content (for Sx)	0.30	$\text{cm}^3 \text{H}_2\text{O cm}^{-3} \text{soil}$	Fixed by design (decouple C from O ₂)
α	Pre-exponential factor	$1.475 \times 10^{11} [1.043 \times 10^{11}, 9.925 \times 10^{13}]$	$\text{mg C cm}^{-3} \text{h}^{-1}$	Fitted
$K_{\text{M,S}}$	Carbon half-saturation	$5.98 \times 10^{-6} [2.64 \times 10^{-6}, 5.90 \times 10^{-3}]$	g C cm^{-3}	Fitted

3.2 Rhizosphere O₂ distributions

Both prescribed O₂ drivers are right-tailed on a log scale (Fig. 2). Because the dissolved proxy is $\approx 450\times$ smaller in magnitude than the gas proxy, we show them in separate panels to preserve visibility. For the gas-phase proxy [air-equivalent units], the mean and variance are $\mu_{\text{gas}} = 0.0891$ and $\text{Var}_{\text{gas}} = 3.53 \times 10^{-3} (\text{cm}^3 \text{O}_2 \text{cm}^{-3})$. For the dissolved proxy [native aqueous units], $\mu_{\text{aq}} = 1.97 \times 10^{-4}$ and $\text{Var}_{\text{aq}} = 1.16 \times 10^{-8}$. We retain each proxy in its native magnitude (no cross-rescaling) so its position

relative to K_{M,O_2} is preserved; these moments are used in the Jensen diagnostic Sect. 3.5. A “high- θ ” benchmark narrows the gas-phase distribution by mapping $\theta \in [0.65, 0.68]$ through the same proxy.

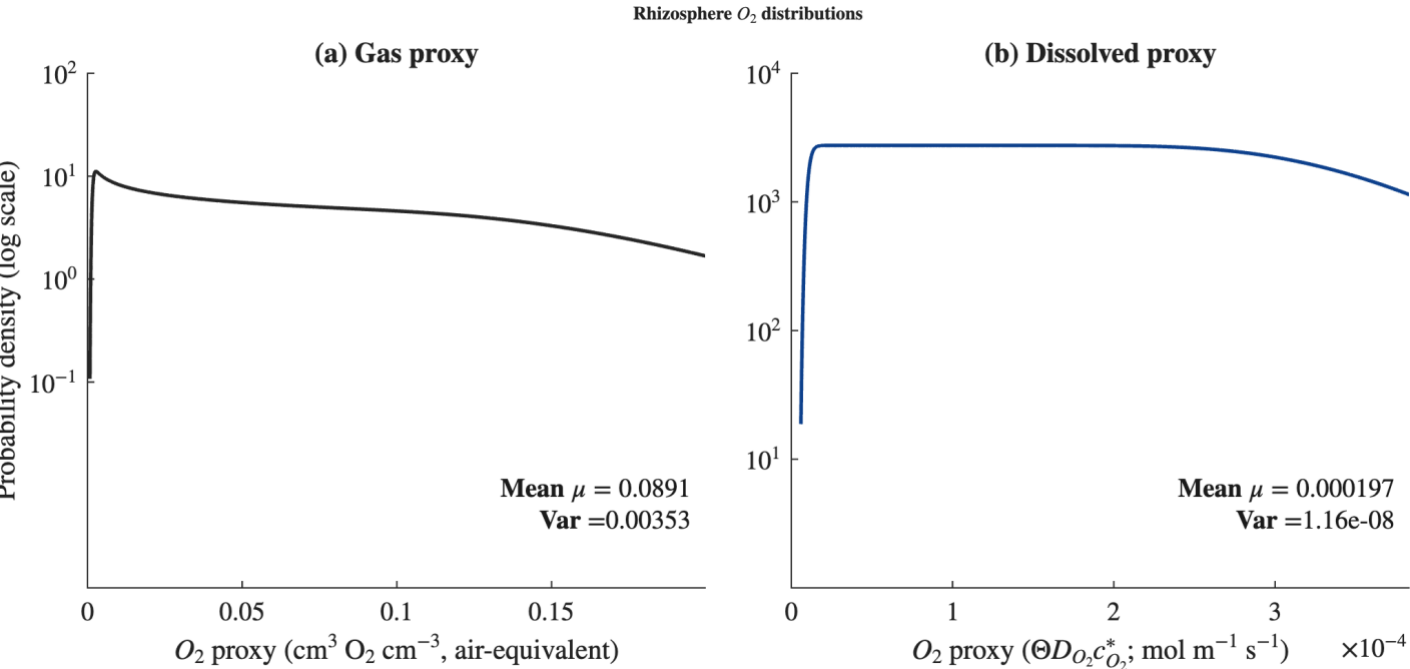


Figure 2. Prescribed rhizosphere O_2 distributions used to drive the DAMM model. (a) Gas-phase proxy in air-equivalent units ($\text{cm}^3 O_2 \text{ cm}^{-3}$), obtained from air-filled porosity and an effective gas diffusivity. (b) Dissolved-phase proxy in native aqueous units based on Henry’s law. Both panels show probability density functions on a logarithmic y-axis; in panel annotations report the mean (μ) and variance (Var) of each distribution, which are later used in the Jensen diagnostic.

3.3 Aggregation bias: uniform vs heterogeneous O_2

Across the full C_{org} range, evaluating respiration at a single mean O_2 (uniform) yields higher fluxes than averaging the model full O_2 distribution (heterogeneous) for gas-phase driver (Fig. 3). The mean aggregation bias is -11.71% , i.e., using a single mean O_2 overestimates aerobic heterotrophic respiration by about one-tenth on average under observed gas-phase heterogeneity. For the dissolved proxy in native units, $\mu_{aq} \ll K_{M, O_2}$ and Var_{aq} is small, so the response is near-linear and the mean bias is negligible (-0.05% ; Fig. S3). Tightening the gas-phase distribution in high- θ benchmark reduces variance and yields a small bias (-0.51% at $C_{org} = 0.15 \text{ g C cm}^{-3}$; Fig. S9).

Practical significance. The ~12% overestimate is the model-structural (Jensen concavity) rather than statistical: the comparison uses the same calibrated model in both configurations and differs only by how O₂ heterogeneity is treated. Hence, the bias is operationally meaningful for budgeting and benchmarking aerobic CO₂ fluxes.

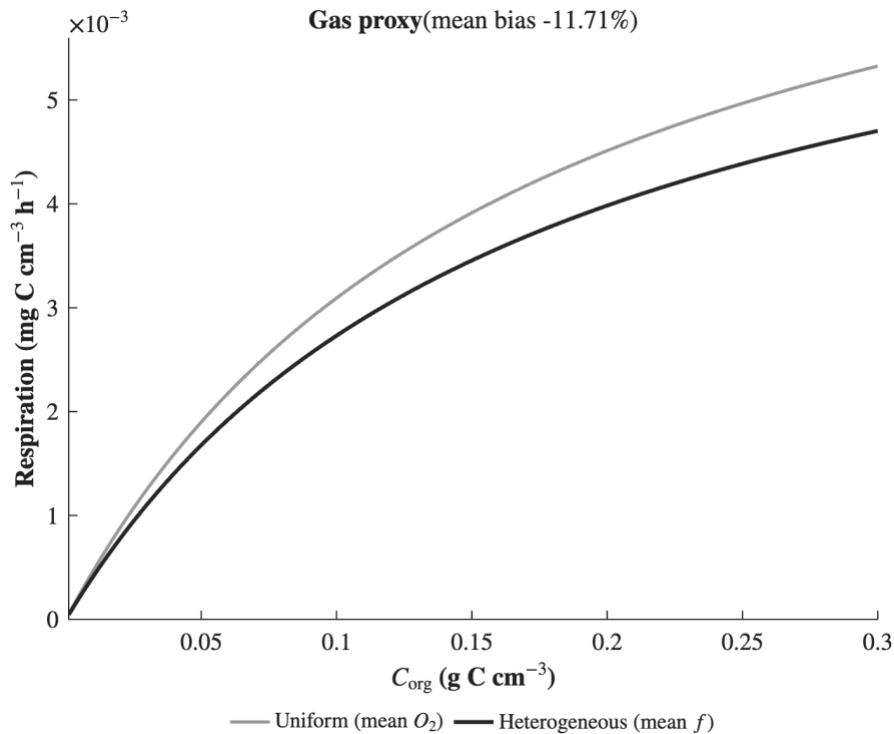


Figure 3: Aerobic respiration as a function of C_{org} when evaluated at a single mean O₂ (“Uniform”) versus averaged over the full gas-phase O₂ distribution (“Heterogeneous”). Both curves use the calibrated DAMM parameters and the gas-phase O₂ proxy. The reported mean aggregation bias (≈ 12 %) quantifies the relative difference between the heterogeneous and uniform cases over the C_{org} range considered.

3.4 Dependence on O₂ affinity and relative saturation

To separate transport from microbial physiology, we varied K_{M, O_2} while holding the O₂ distributions fixed and plot the aggregation bias against the relative saturation ratio $\text{mean}(O_2) / K_{M, O_2}$; collapses cases onto an interpretable axis (Fig. 4). For the gas proxy, varying K_{M, O_2} by factors of $\times 0.25$, $\times 0.5$, $\times 1$, $\times 2$, $\times 4$ yields a bias range from -13.15% to -5.80% . For the dissolved proxy with unit-matched K_{M, O_2} (i.e., chosen so that its mean O₂ concentration correspond to the same relative saturation as the gas case), the range is -8.97% to -3.97% . The baseline points occur at $\text{mean}(O_2) / K_{M, O_2} \approx 0.74$ for both proxies; biases differ in magnitude because the underlying O₂ PDFs differ in spread. In short, the magnitude of bias increases as the system approaches the nonlinear mid-saturation regime and as O₂ variability increases.

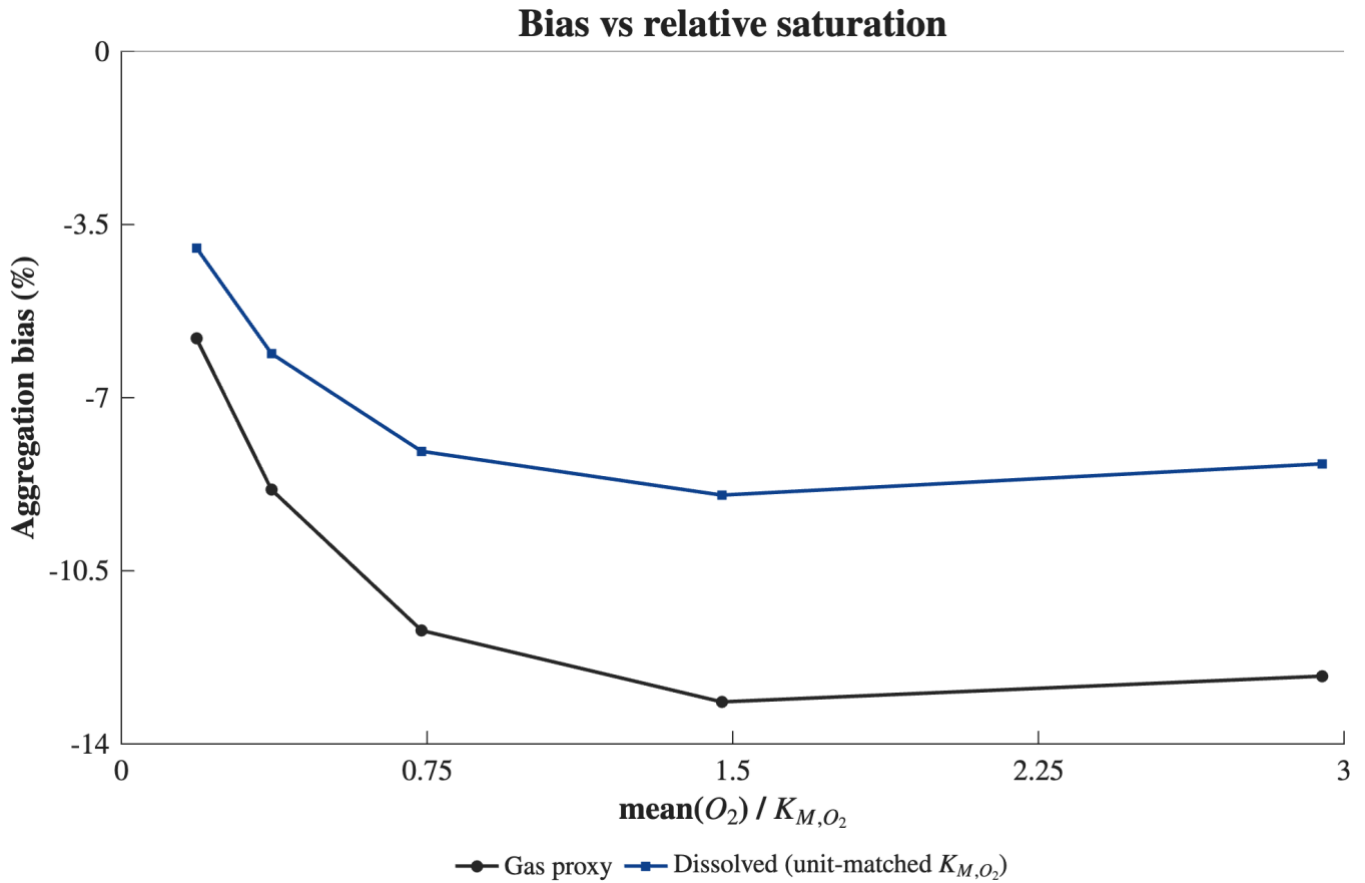


Figure 4: Aggregation bias (%) as a function of $\text{mean}(\text{O}_2) / K_{M,\text{O}_2}$ for the gas-phase O_2 proxy (circles) and the dissolved-phase proxy with unit-matched K_{M,O_2} (squares). For each proxy, K_{M,O_2} is varied via multiplicative factors around the calibrated gas-phase value. Bias is computed as the percent difference between heterogeneous (averaged over the O_2 distribution) and uniform (evaluated at the mean O_2) respiration at fixed Corg . The bias is most pronounced at intermediate relative saturation ($\text{mean}(\text{O}_2) / K_{M,\text{O}_2} \approx 1$) and small when O_2 is either much higher or much lower than K_{M,O_2} .

3.5 Jensen diagnostic

Holding the mean O_2 fixed and increasing $\text{Var}(\text{O}_2)$ with symmetric two-point mixtures produces a near-linear growth in bias (Fig. S6). A second order (quadratic) approximation based on the local curvature of the Michaelis–Menten O_2 term closely tracks the simulations, confirming that the bias is a structural consequence of concavity: averaging a heterogeneous O_2 field yields $E[f(\text{O}_2)] < f(E[\text{O}_2])$. A response-surface view in the SI (Fig. S10) shows that bias remains non-positive across the explored domain and peaks when the $\text{mean}(\text{O}_2)$ is comparable to K_{M,O_2} , and the coefficient of variation is large.

4 Discussion

4.1 Principal findings

275 Our analysis shows that spatial heterogeneity in rhizosphere oxygen can systematically bias modelled aerobic heterotrophic
respiration when the process is evaluated at a single mean oxygen concentration. Using a Dual Arrhenius-Michaelis-Menten
(DAMM) scheme calibrated to marsh-soil incubations, we compared a “uniform” configuration, in which respiration is
evaluated at the spatial mean oxygen, with a “heterogeneous” configuration, where respiration is averaged over a prescribed
distribution of oxygen in the root zone. For gas-phase (air-equivalent) oxygen distributions representative of tidally influenced
280 marsh soils, the uniform configuration overestimates aerobic CO₂ production by roughly 10-15 % across a wide range of
organic carbon concentrations. This aggregation bias arises because the oxygen-limitation term in the DAMM model is a
saturating Michaelis-Menten function of oxygen, which is concave over the range where respiration is partially oxygen-limited.

The magnitude of the bias depends on two simple descriptors of the oxygen field: (i) the ratio of the mean oxygen concentration
to the microbial half-saturation constant for oxygen, and (ii) the variance of oxygen in the rhizosphere. When mean oxygen is
285 far below the half-saturation constant, microbial respiration responds nearly linearly to oxygen and the bias is negligible. When
mean oxygen is far above the half-saturation constant, the oxygen term is close to saturation and again becomes insensitive to
spatial variation. The largest aggregation errors occur at intermediate mean oxygen, where respiration is strongly limited but
not saturated, and where the variance in oxygen is substantial. Our two point and lognormal variance experiments, which hold
the mean oxygen fixed while increasing its variance, confirm that the magnitude of the negative bias increases monotonically
290 with oxygen variance and that the sign of the bias does not change within the Michaelis-Menten framework (Figs. S6, S7).

The comparison between gas-phase and dissolved-phase oxygen drivers illustrates the same mechanism from a complementary
angle. When the dissolved-phase proxy is used in its native units, its mean and variance are both much smaller than the half-
saturation constant, so respiration is effectively in the linear regime of the oxygen response, and the aggregation bias is close
to zero (Fig. S3). When we adjust the microbial half-saturation constant so that the dissolved-phase proxy has a similar degree
295 of oxygen limitation to the gas-phase proxy (i.e. a similar ratio of mean oxygen to half-saturation), the dissolved-phase
simulations exhibit biases comparable in sign and magnitude to the gas-phase case (Fig. 4). The difference between the gas
and aqueous results therefore reflects how each proxy is positioned relative to microbial oxygen affinity, rather than a change
in model structure.

Overall, the marsh-soil case study provides a concrete example of a general mechanism: when a saturating oxygen-limitation
300 term is applied to a spatially heterogeneous rhizosphere, evaluating respiration at a single bulk oxygen level tends to

overestimate aerobic heterotrophic CO₂ production. The size of that overestimate can be understood and, in principle, approximated from simple statistics of the oxygen distribution and plausible ranges of the half-saturation constant.

4.2 Relation to process understanding and prior work

305 Tidal marsh soils are characterized by strong small-scale redox heterogeneity. Hydrologic forcing through tidal inundation, drainage, and groundwater flow drives rapid switches between oxic and anoxic conditions, while plant traits such as radial oxygen loss from aerenchymatous roots create localized oxic zones around roots embedded in largely reduced sediment. Microelectrode and planar optode studies in flooded and marsh soils consistently show millimetre-centimetre oxygen mosaics, with steep gradients around roots and patches where oxygen fluctuates over diel and tidal cycles (Koop-Jakobsen et al., 2018). Our imposed gas-phase oxygen distributions are designed to be consistent with this process understanding: they arise from
310 variation in water content and air-filled porosity, both of which are strongly affected by tidal inundation and drainage, and they are intended to bracket realistic degrees of rhizosphere oxygen heterogeneity in tidally influenced marshes.

Previous work has emphasized that tidal marshes can be strong carbon sinks because high primary production and efficient sediment trapping are combined with slow anaerobic decomposition. At the same time, a growing literature highlights how root-mediated oxygen supply can locally accelerate aerobic decomposition and alter greenhouse gas production, especially in
315 systems with high root biomass and pronounced aerenchyma development. However, most marsh carbon models represent belowground redox conditions using a single bulk variable (such as average water content, a redox class, or an empirical soil-moisture scalar), and they typically apply this scalar uniformly within a grid cell or plant functional type. In doing so, they effectively assume that respiration responds to a uniform oxygen environment, even though observations show that roots and microbes experience a patchwork of oxic and anoxic microhabitats (Pezeshki and DeLaune, 2012).

320 The broader ecological modelling literature has long recognized that using mean environmental conditions in nonlinear response functions can lead to systematic upscaling biases. (Rastetter et al., 1992), for example, showed that aggregating nonlinear responses to water and nutrients can misrepresent ecosystem responses to climate change. Yet, to our knowledge, the specific consequences of small-scale rhizosphere oxygen heterogeneity for marsh carbon flux estimates have not been quantified using a physiologically based respiration scheme calibrated to marsh soils. Our results therefore connect a well-
325 established mathematical principle (Jensen's inequality) to a specific, process rich context in which oxygen heterogeneity is both expected and extensively documented. They complement previous work on micro-scale redox and organic-matter dynamics in soils and sediments by translating that heterogeneity into explicit consequences for ecosystem-scale flux calculations (Wilson and Gerber, 2021), and recent three dimensional reactive transport simulations that resolve plant mediated oxygen inputs and fine scale porewater biogeochemistry in coastal wetlands (Zhou et al., 2024) by translating those insights
330 into an explicit correction for marsh carbon flux calculations.

Finally, by framing the bias in terms of the mean and variance of oxygen and the microbial half-saturation constant, our results fit naturally alongside DAMM-type applications in terrestrial and wetland systems. DAMM has been used to represent soil heterotrophic respiration in land surface and ecosystem models, where it replaces simple Q₁₀ moisture scalars with a mechanistic treatment of temperature, substrate, and diffusion constraints. Our analysis extends this line of work by showing

335 that, once oxygen limitation is represented explicitly, the spatial distribution of oxygen within a model grid cell becomes a
first-order control on predicted respiration, rather than a secondary details (Moyano et al., 2018).

4.3 Implications for modelling and observation

340 The most direct implication for marsh process models is that replacing heterogeneous rhizosphere oxygen with a single mean
value will, under realistic conditions, tend to overestimate aerobic heterotrophic CO₂ production. In our marsh case study, gas-
phase oxygen heterogeneity produces a mean aggregation bias of -11.7 %, that is, a uniform-oxygen configuration
overestimates respiration by roughly one tenth. Because net carbon sequestration reflects the balance between inputs and
heterotrophic losses, this implies a systematic underestimation of the marsh carbon sink. If a bias of this magnitude were
broadly representative of tidal marshes, then, given published global salt-marsh carbon burial estimates of about 10 Tg C yr⁻¹
345 to 21 Tg C yr⁻¹ (Alongi, 2020; Ouyang and Lee, 2014), correcting for aggregation bias would increase inferred global marsh
sequestration by on the order of 1-2.5 Tg C yr⁻¹, corresponding to roughly 0.1-0.25 Gt C over a century. Given the wide
uncertainty in global salt-marsh carbon burial, with CAR-based estimates spanning roughly 0.9-31.4 Tg C yr⁻¹ (Chastain et
al., 2022), these numbers should be interpreted as order of magnitude extrapolations rather than precise forecasts. Nevertheless,
they indicate that aggregation bias is non negligible for coastal carbon budgets and for greenhouse gas inventories that include
blue carbon ecosystems, even though the corrected flux remains small compared with anthropogenic fossil fuel emissions
350 (Intergovernmental Panel on Climate Change (IPCC), 2023).

At the land surface and Earth system model scale, most soil biogeochemical schemes still employ relatively simple
representations of heterotrophic respiration, often combining a temperature response with an empirical soil moisture scalar.
More mechanistic schemes, including DAMM and related enzyme kinetic formulations, are now being incorporated into
regional and global models to capture interactions among temperature, moisture, and substrate supply more explicitly. Recent
355 work using a reduced complexity carbon-climate model shows that replacing a first order respiration formulation with a
Michaelis-Menten formulation can substantially increase the strength of the terrestrial carbon-climate feedback and raise the
remaining carbon budget to meet a given temperature target (Beer, 2025). This underlines that large scale projections are
structurally sensitive to the mathematical form chosen for respiration. Our results add that, when such schemes include an
explicit oxygen limitation term but still operate at coarse spatial resolution, they also inherit an additional structural error that
360 depends on how oxygen is aggregated within each grid cell. Regions with shallow water tables, strong hydrologic variability,
and high root biomass, such as tidal marshes, riparian wetlands, peatlands, and some rice systems, are likely to be particularly
sensitive.

In practical terms, the modelling framework developed here can be used to diagnose where aggregation bias is likely to matter
and to design simple mitigation strategies. First, modellers can characterize or bracket the “relative saturation” of oxygen, that
365 is, the ratio between the mean oxygen in the rhizosphere and a plausible microbial half-saturation constant, together with a
measure of subgrid variability. A convenient measure of variability is the coefficient of variation, defined as the standard
deviation of oxygen divided by its mean. When relative saturation is very low or very high, or when variability is small, our

results suggest that aggregation error will be negligible. When relative saturation is intermediate and variability in oxygen is substantial, a negative aggregation bias should be expected if a single bulk oxygen value is used.

370 Second, if subgrid oxygen distributions cannot be resolved explicitly, our Jensen-type diagnostics offer a way to correct for at
least part of the bias or to represent it as an uncertainty range. For small to moderate variability, a Taylor series approximation
using the local curvature of the oxygen-limitation term can be used to derive a multiplicative correction factor for the uniform-
oxygen flux, which we demonstrate for the gas-phase proxy (Fig. S11). When variability is large or poorly constrained, it may
375 be more appropriate to propagate a plausible range of oxygen variance into a range of predicted fluxes, thereby making the
aggregation error explicit in model uncertainty budgets.

Finally, we note that a conceptually similar issue arises when oxygen limitation is parameterized as a function of another state
variable, such as water content or water-table depth. If oxygen is first expressed as a nonlinear function of water content, and
then the model uses mean water content within a grid cell, the resulting “mean water” baseline can differ from the baseline
defined by mean oxygen. In extreme cases, this can even alter the apparent sign of the aggregation effect. Our analysis suggests
380 that it is preferable to define the uniform reference case directly in oxygen space (that is, evaluate respiration at the mean
oxygen concentration) rather than in terms of the mean of a proxy variable such as water content.

4.4 Uncertainties and limits

Several limitations of our approach should be kept in mind when interpreting the results. First, the gas-phase and dissolved-
phase oxygen distributions we impose are proxies that reflect hydrologic and structural controls (through water content,
385 porosity, and effective diffusivity) but do not resolve root architecture, advective transport, or biogeochemical feedback
between reactions and oxygen supply. They are designed to constrain plausible degrees of rhizosphere heterogeneity rather
than to reproduce site specific oxygen probability density functions. Fully coupled transport–reaction models that explicitly
resolve root geometry, gas-phase diffusion, and solute flows would be needed to predict detailed oxygen distributions for
particular marshes or plant communities.

390 Second, we treat microbial oxygen affinity through a single half-saturation constant and explore its influence via sensitivity
sweeps, rather than by refitting oxygen kinetics for each hydrologic or redox regime. In reality, microbial communities and
enzyme systems can acclimate or adapt to chronic hypoxia or frequent redox oscillations, shifting effective half-saturation
constants and potentially altering the degree of concavity in the oxygen response. Our sensitivity analysis indicates that the
sign of the aggregation bias is robust as long as the oxygen-limitation term retains its saturating form, but the exact magnitude
395 of the bias in any specific marsh will depend on the local microbial community and redox history.

Third, our calibration focuses on aerobic heterotrophic respiration derived from long term soil incubations under controlled
temperature and water content. Autotrophic root respiration and anaerobic pathways (e.g. fermentation, sulfate reduction,
methanogenesis) are outside the scope of the present study. In the field, these processes coexist and interact; root respiration
can covary with oxygen and hydrology, while anaerobic pathways contribute substantially to CO₂ and CH₄ production under

400 persistent anoxia. Our results should therefore be interpreted as applying to the aerobic heterotrophic component of
belowground CO₂ production, not to total soil respiration or full greenhouse gas budgets.

Lastly, we use a single marsh system and a fixed temperature as our calibration context. While the DAMM formulation is
designed to be portable across sites and temperature regimes, additional calibration and validation against independent datasets
from other marsh types, climates, and plant communities would be needed to confirm the quantitative magnitude of aggregation
405 bias across the global diversity of coastal wetlands.

4.5 Observational priorities and data need

Our diagnostics highlight a specific set of observational needs that would allow the aggregation mechanism to be better
constrained in marshes and other wetland ecosystems. First, direct measurements of oxygen distributions in marsh rhizospheres
are still rare relative to the spatial and temporal variability expected from hydrology and plant traits. Fine scale microelectrode
410 and planar optode measurements (Koop-Jakobsen and Wenzhöfer, 2015) have proven capable of resolving steep oxygen
gradients and dynamic oxic-anoxic interfaces in sediments and wetland soils. Extending these techniques across tidal
elevations, plant functional types, and seasons would provide the empirical oxygen means, variances, and coefficients of
variation needed to apply our bias diagnostics more directly.

Second, joint observations of oxygen, water content, and CO₂ fluxes in the root zone would permit explicit tests of the DAMM
415 based aggregation framework. For example, combining high resolution oxygen imaging with controlled manipulations of water
level and plant species could reveal how hydrologic state and radial oxygen loss map into the statistical properties of
rhizosphere oxygen distributions, and how those properties in turn control aerobic respiration.

Third, improved constraints on microbial oxygen kinetics in marsh soils through laboratory respiration assays across oxygen
gradients, in combination with molecular or enzymatic measurements would narrow the plausible range of half saturation
420 constants and inform when and where the concavity driven aggregation bias is likely to be large. Together, such observational
efforts would reduce structural uncertainty in marsh carbon models, enable more targeted use of correction factors in DAMM
type schemes, and provide a foundation for representing rhizosphere oxygen heterogeneity in land surface and Earth system
models.

5 Conclusions

425 We examined how small-scale heterogeneity in rhizosphere oxygen affects modelled aerobic heterotrophic respiration in tidal
marsh soils. Using a Dual Arrhenius-Michaelis-Menten formulation calibrated to long term marsh soil incubations, we
compared respiration evaluated at a single mean oxygen concentration with respiration averaged over heterogeneous oxygen
distributions generated by hydrologic and structural variation in the root zone. For gas-phase oxygen fields representative of
tidally influenced marsh soils, evaluating respiration at the mean oxygen led to an overestimate of aerobic CO₂ production on
430 the order of 10-15 %.

435 The magnitude of the bias is controlled by the ratio between mean oxygen and the microbial half-saturation constant for oxygen, together with the variance of oxygen in the rhizosphere. Biases are negligible when oxygen is either extremely scarce or effectively saturating, or when spatial variability is small, and they are largest at intermediate oxygen levels and substantial variability.

440 These findings have direct implications for marsh carbon budgets and for the representation of wetland soils in biogeochemical and Earth system models. In models that employ mechanistic, oxygen-limited respiration schemes but represent oxygen as a single bulk variable per grid cell, the concavity driven aggregation bias identified here can introduce a structural overestimate of aerobic heterotrophic CO₂ production in systems with strongly heterogeneous rhizospheres. Our diagnostic framework provides a simple way to anticipate when this bias is likely to be important namely, where mean oxygen in the root zone is comparable to microbial half-saturation constants and where variability in oxygen is nontrivial and to design correction factors or uncertainty ranges accordingly.

445 Finally, the work points to clear observational and modelling priorities. High resolution measurements of rhizosphere oxygen distributions, coupled with constraints on microbial oxygen kinetics, would allow aggregation effects to be quantified and corrected for specific marsh types. Extending the present framework to include autotrophic root respiration, anaerobic pathways, and fully coupled transport reaction models would further clarify how plant traits, hydrology, and microbial communities jointly shape carbon fluxes in coastal wetlands. By explicitly linking rhizosphere oxygen heterogeneity to aggregation bias in a widely used respiration scheme, this study provides a template for incorporating small scale redox dynamics into ecosystem and Earth system scale assessments of marsh carbon cycling.

Appendix:

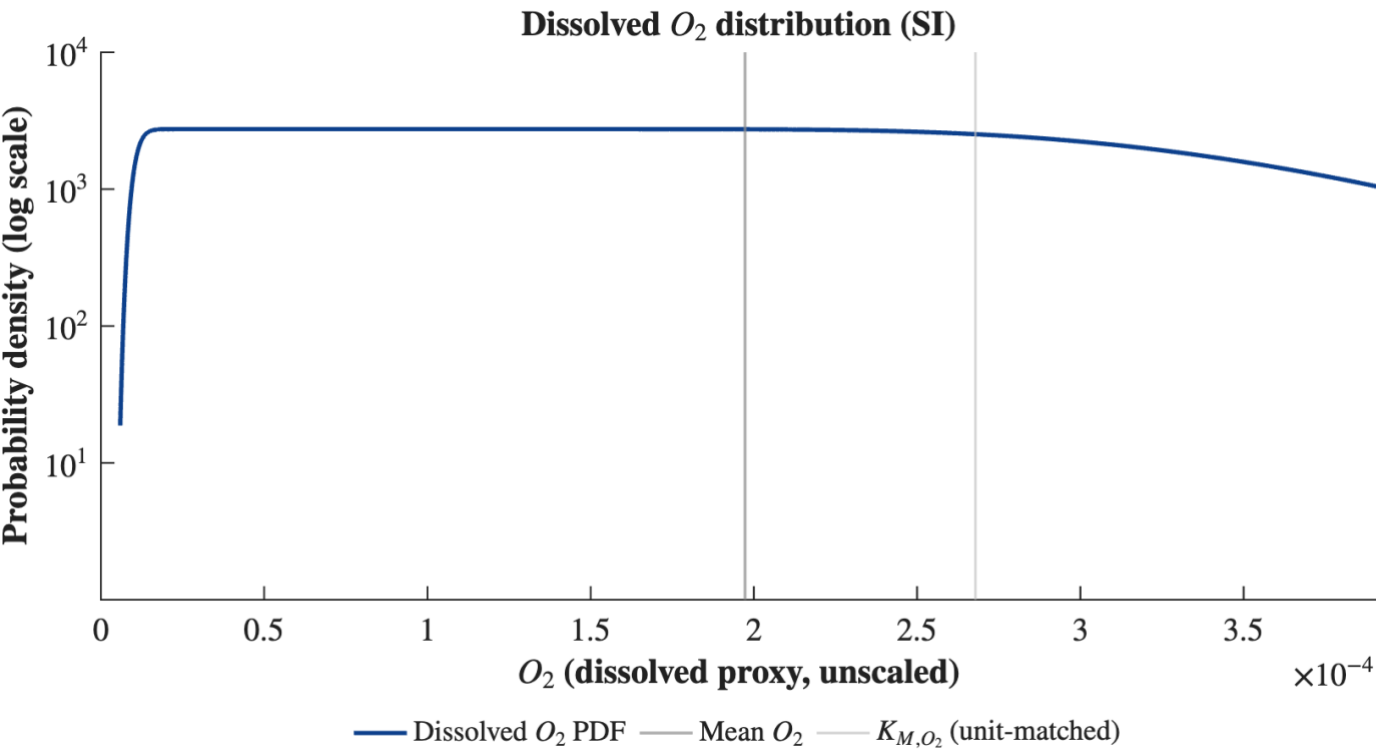


Figure S2: Dissolved-phase O₂ proxy distribution with the same units and proxy definition as in Fig. 2b, shown on a logarithmic y-axis. The vertical lines indicate the distribution mean (μ) and the unit-matched K_{M,O_2} value used in the dissolved sensitivity sweep in Fig. 4, placing the gas and dissolved cases on a comparable relative O₂ saturation scale.

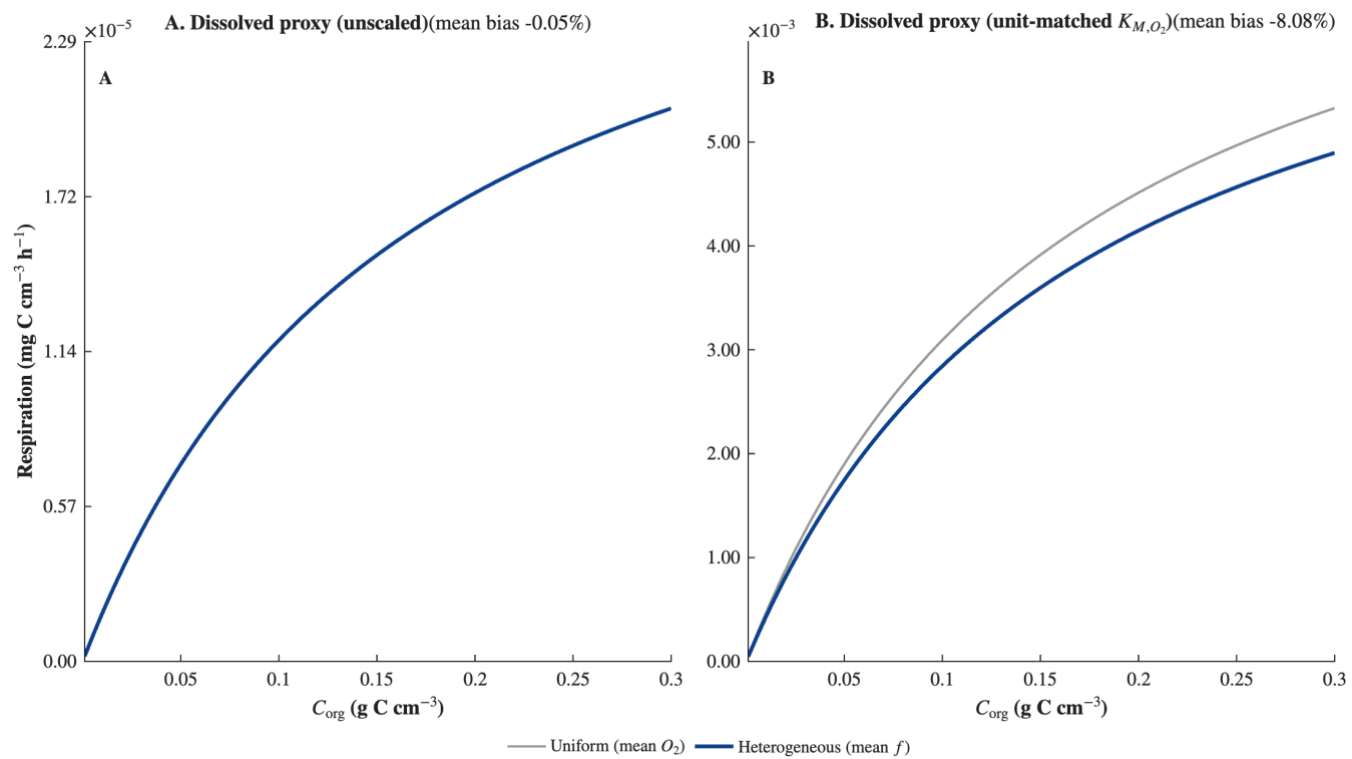


Figure S3: Sensitivity of aggregation bias to the choice of dissolved-phase O₂ driver. (a) Uniform (mean O₂) versus heterogeneous (full distribution) DAMM respiration curves driven by the unscaled dissolved O₂ proxy. Because dissolved O₂ is orders of magnitude below the calibrated K_{M,O_2} , the O₂ limitation is effectively linear and the bias is negligible. (b) Same analysis but with K_{M,O_2} rescaled to match the mean relative saturation of the gas-phase case. Under this unit-matched K_{M,O_2} , the dissolved proxy exhibits a negative aggregation bias similar in magnitude to the gas-phase proxy.

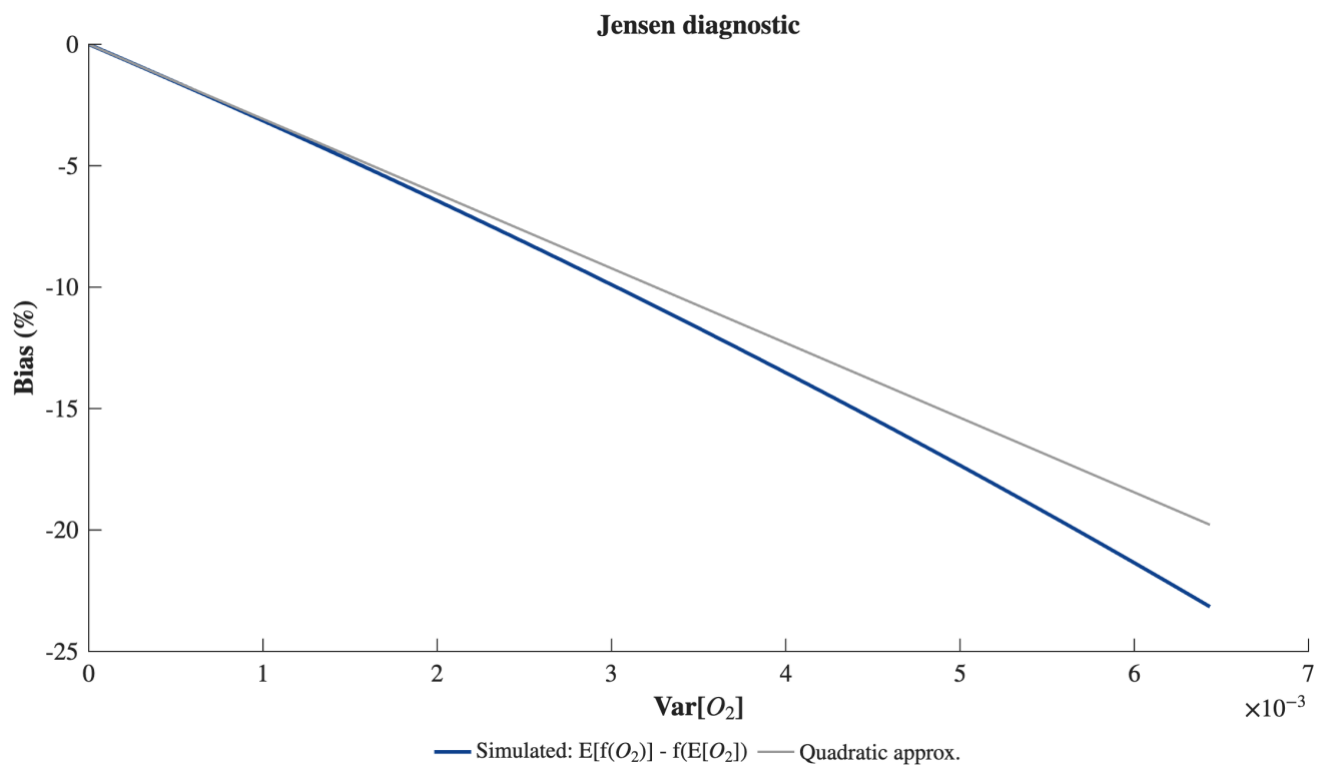


Figure S6: Jensen diagnostic for a symmetric two-point O₂ family. Bias in respiration $E[f(O_2)] - f(E[O_2])$ versus $\text{Var}[O_2]$ at fixed mean O₂ and C_{org}, computed from explicit DAMM evaluations (primary curve) and from the second-order Taylor approximation based on the curvature of O₂ limitation (secondary curve). The close agreement over a broad variance range supports the use of a curvature-based Jensen correction.

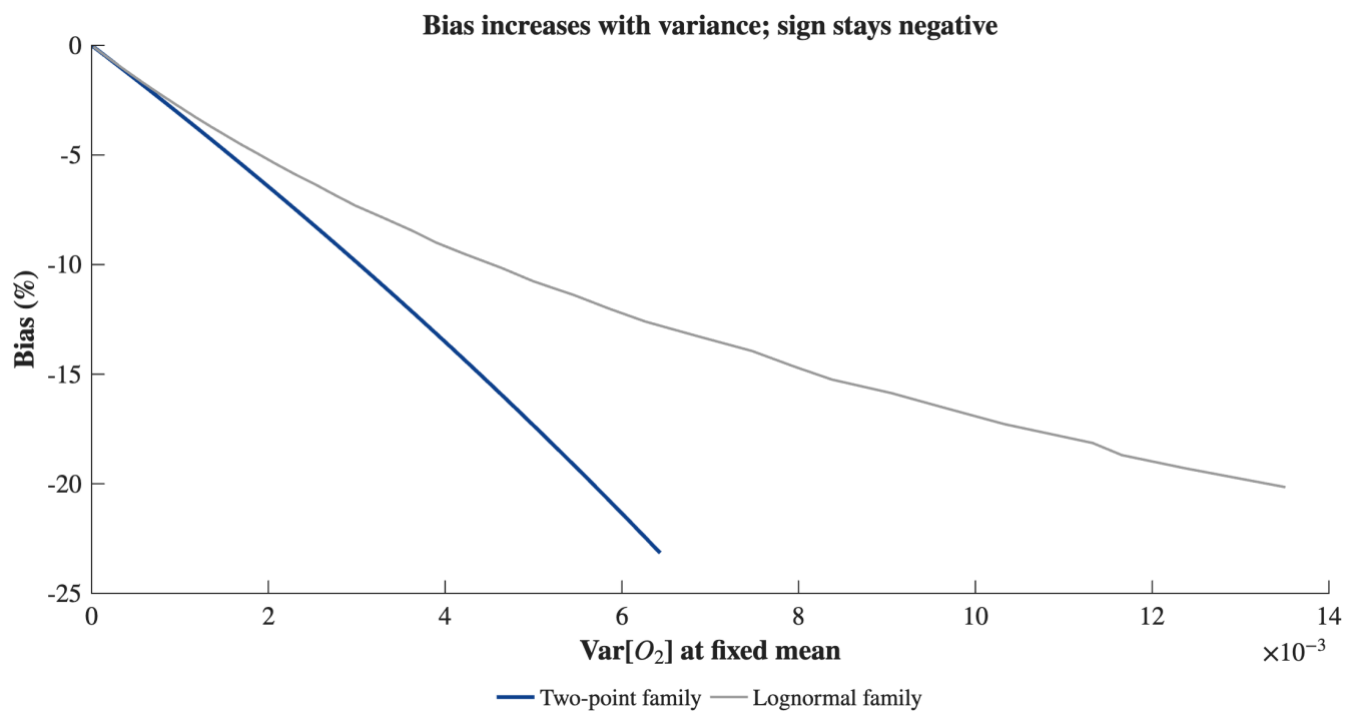


Figure S7: Aggregation bias versus $\text{Var}[O_2]$ at fixed mean O_2 for two families of O_2 distributions: symmetric two-point mixtures and lognormal distributions. For each family, the mean O_2 is held constant and the variance is varied. In both cases, the bias remains negative and its magnitude increases monotonically with $\text{Var}[O_2]$, indicating that the concavity mechanism is robust to the choice of O_2 distribution shape.

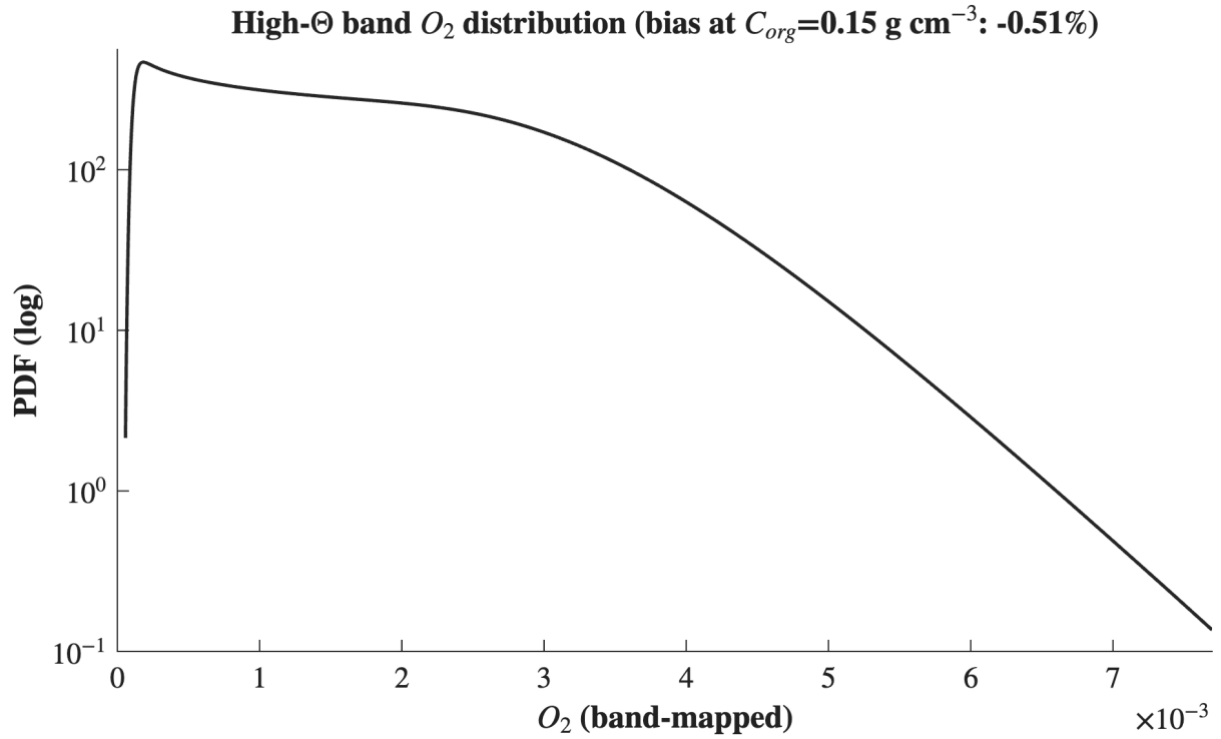


Figure S9: Probability density of O_2 obtained by mapping a narrow water-content band, $\theta \in [0.65, 0.68]$, to gas-phase O_2 using the same proxy as in the main analysis; the density is shown on a log scale. This “high- θ ” band has much lower variance than the full rhizosphere distribution. Evaluated at $C_{org} = 0.15 \text{ g C cm}^{-3}$, the respiration predicted at the mean O_2 (uniform configuration) is slightly larger than the heterogeneous average $E[f(O_2)]$, giving an aggregation bias of -0.51% (uniform overpredicts). Numerically: $R_{uniform} = 1.2657 \times 10^{-4}$ and $R_{heterogeneous} = 1.2592 \times 10^{-4} \text{ mg C cm}^{-3} \text{ h}^{-1}$ (calibrated DAMM parameters; substrate term evaluated at $\theta_{ref} = 0.30$).

500

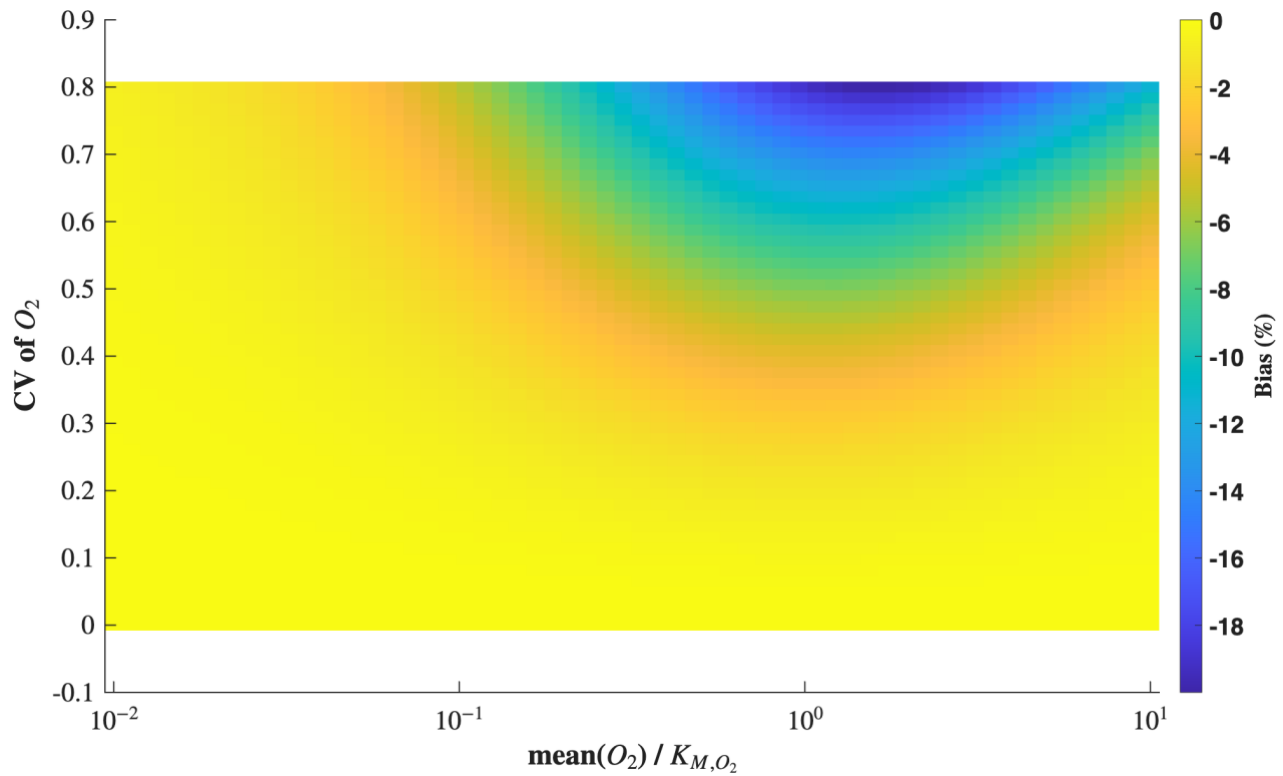


Figure S10: Response surface of aggregation bias as a function of $\text{mean}(O_2) / K_{M,O_2}$ and the coefficient of variation (CV) of O_2 , for a symmetric two-point O_2 family at fixed mean O_2 . Colour shading indicates bias (%) with negative values only, consistent with concave O_2 limitation. Bias magnitude increases with CV and peaks at intermediate relative saturation, summarising how physiology K_{M,O_2} and O_2 heterogeneity jointly control aggregation error.

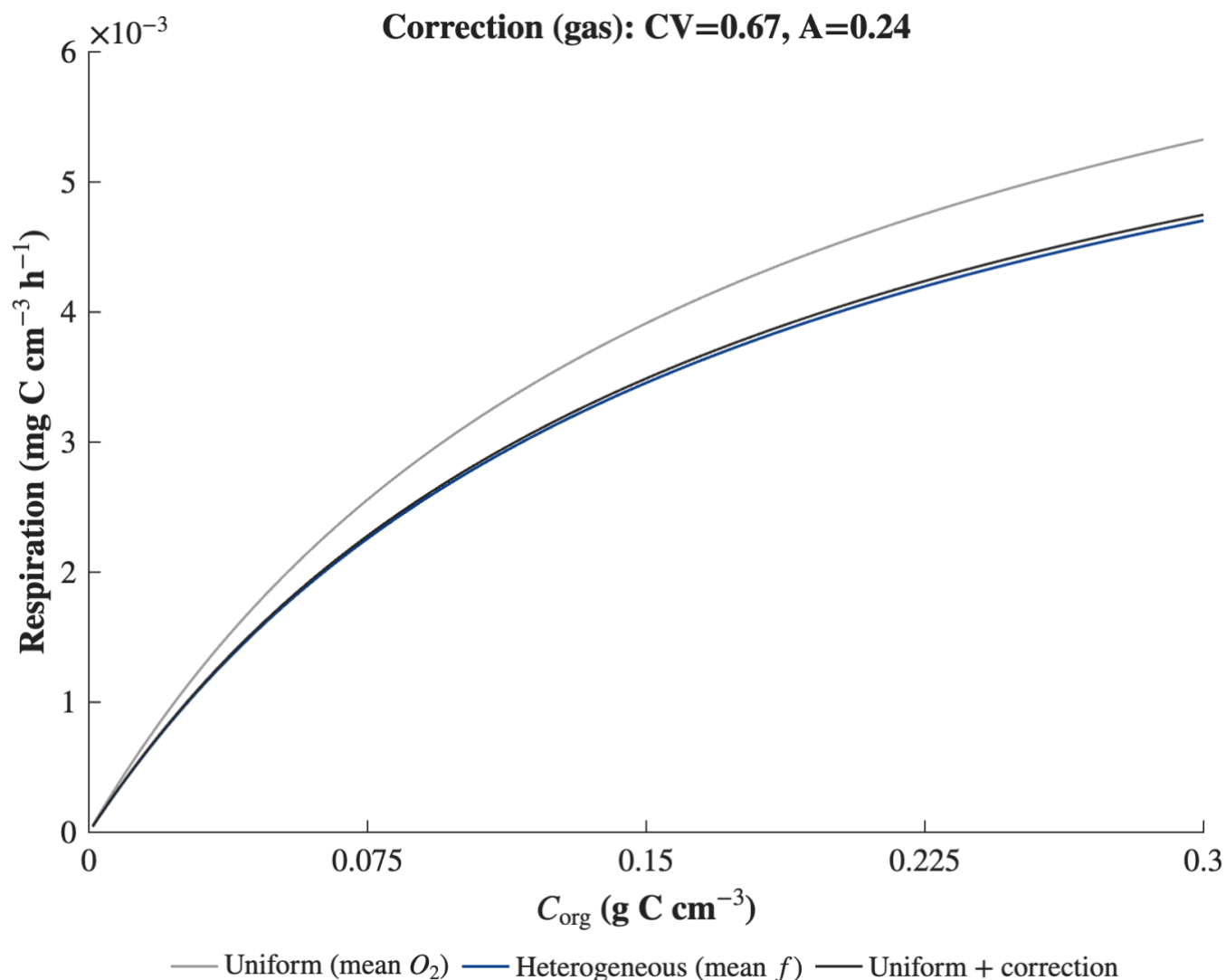


Figure S11: Example of a small variance Jensen correction applied to the gas-phase O_2 proxy. The uniform curve shows respiration evaluated at the mean O_2 , the heterogeneous curve shows the explicit average over the prescribed O_2 distribution, and the corrected curve applies a second-order curvature-based factor using the local O_2 variance and curvature. For the modest O_2 variability considered, the correction removes most of the negative aggregation bias, illustrating a lightweight approach for models that cannot explicitly resolve subgrid O_2 .

Code availability

520 The MATLAB code corresponding to this publication, including all scripts used for the analyses and figures, is archived in the institutional research data repository of Universität Hamburg and is available at <http://doi.org/10.25592/uhhfdm.17912>

Data availability

Summary datasets used in this study, including incubation medians and ICBM outputs for model calibration, are archived in the institutional research data repository of Universität Hamburg (<http://doi.org/10.25592/uhhfdm.14367>). Additional raw measurements are available from the corresponding author upon reasonable request.

525 Author contribution

YS, PP, CB and PM designed the study and YS, PP, and CB carried out the modelling. FN, JNB and AE designed the incubation experiment, FN investigated and provided the incubation data. YS wrote the manuscript with input from all co-authors.

Competing interests

530 The authors declare that they have no conflict of interest.

Acknowledgments

This project was funded by the Deutsche Forschungsgemeinschaft (DFG, German Research Foundation) as a part of the research training group (RTG2530) “Biota-mediated effects on Carbon Cycling in Estuaries (grant no. 407270017).

535 P.M was supported by Deutsche Forschungsgemeinschaft (DFG) in the framework of the Emmy Noether program (502681570).

C.B. acknowledges financial support by the Heisenberg professorship (DFG-BE 6485/4-1).

We thank Daniel Schwarze for supporting laboratory work.

540 References

Alizad, K., Hagen, S. C., Morris, J. T., Bacopoulos, P., Bilskie, M. V., Weishampel, J. F., and Medeiros, S. C.: A coupled, two-dimensional hydrodynamic-marsh model with biological feedback, *Ecological Modelling*, 327, 29–43, <https://doi.org/10.1016/j.ecolmodel.2016.01.013>, 2016.

545 Alongi, D. M.: Carbon balance in salt marsh and mangrove ecosystems: A global synthesis, *Journal of Marine Science and Engineering*, 8, 1–21, <https://doi.org/10.3390/jmse8100767>, 2020.

André, O., And, N., Ka“tterer, T., Ka“tterer, K., Andren, O., and Katterer, T.: ICBM: The Introductory Carbon Balance Model for Exploration of Soil Carbon Balances, *Ecological Applications*, 7, 1226, <https://doi.org/10.2307/2641210>, 1997.

- Barbier, E. B., Hacker, S. D., Kennedy, C., Koch, E. W., Stier, A. C., and Silliman, B. R.: The value of estuarine and coastal ecosystem services, *Ecological Monographs*, 81, 169–193, <https://doi.org/10.1890/10-1510.1>, 2011.
- 550 Beer, C.: Carbon–climate feedback higher when assuming Michaelis–Menten kinetics of respiration, *Earth System Dynamics*, 16, 1527–1537, <https://doi.org/10.5194/esd-16-1527-2025>, 2025.
- Beer, C., Knoblauch, C., Hoyt, A. M., Hugelius, G., Palmtag, J., Mueller, C. W., and Trumbore, S.: Vertical pattern of organic matter decomposability in cryoturbated permafrost-affected soils, *Environmental Research Letters*, 17, 104023, <https://doi.org/10.1088/1748-9326/ac9198>, 2022.
- 555 Bernal, B., Megonigal, J. P., and Mozdzer, T. J.: An invasive wetland grass primes deep soil carbon pools, *Global Change Biology*, 23, 2104–2116, <https://doi.org/10.1111/GCB.13539>, 2017.
- Bhattacharyya, S. S. and Furtak, K.: Soil–Plant–Microbe Interactions Determine Soil Biological Fertility by Altering Rhizospheric Nutrient Cycling and Biocrust Formation, *Sustainability* 2023, Vol. 15, Page 625, 15, 625, <https://doi.org/10.3390/SU15010625>, 2022.
- 560 Chastain, S. G., Kohfeld, K. E., Pellatt, M. G., Olid, C., and Gailis, M.: Quantification of blue carbon in salt marshes of the Pacific coast of Canada, *Biogeosciences*, 19, 5751–5777, <https://doi.org/10.5194/bg-19-5751-2022>, 2022.
- Davidson, E. A., Samanta, S., Caramori, S. S., and Savage, K.: The Dual Arrhenius and Michaelis-Menten kinetics model for decomposition of soil organic matter at hourly to seasonal time scales, *Global Change Biology*, 18, 371–384, <https://doi.org/10.1111/j.1365-2486.2011.02546.x>, 2012.
- 565 Duarte, C. M., Losada, I. J., Hendriks, I. E., Mazarrasa, I., and Marbà, N.: The role of coastal plant communities for climate change mitigation and adaptation, *Nature Climate Change*, 3, 961–968, <https://doi.org/10.1038/NCLIMATE1970>, 2013.
- Granse, D., Wanner, A., Stock, M., Jensen, K., and Mueller, P.: Plant-sediment interactions decouple inorganic from organic carbon stock development in salt marsh soils, *Limnology And Oceanography Letters*, <https://doi.org/10.1002/lol2.10382>, 2024.
- 570 Intergovernmental Panel on Climate Change (IPCC): Global Carbon and Other Biogeochemical Cycles and Feedbacks, in: *Climate Change 2021 – The Physical Science Basis*, Piers Forster, 673–816, <https://doi.org/10.1017/9781009157896.007>, 2023.
- Kirwan, M. L., Walters, D. C., Reay, W. G., and Carr, J. A.: Sea level driven marsh expansion in a coupled model of marsh erosion and migration, *Geophysical Research Letters*, 43, 4366–4373, <https://doi.org/10.1002/2016GL068507>, 2016.
- 575 Knoblauch, C., Beer, C., Sosnin, A., Wagner, D., and Pfeiffer, E. M.: Predicting long-term carbon mineralization and trace gas production from thawing permafrost of Northeast Siberia, *Global Change Biology*, 19, 1160–1172, <https://doi.org/10.1111/gcb.12116>, 2013.
- Koop-Jakobsen, K. and Wenzhöfer, F.: The Dynamics of Plant-Mediated Sediment Oxygenation in *Spartina anglica* Rhizospheres—a Planar Optode Study, *Estuaries and Coasts*, 38, 951–963, <https://doi.org/10.1007/s12237-014-9861-y>, 2015.
- 580 Koop-Jakobsen, K., Mueller, P., Meier, R. J., Liebsch, G., and Jensen, K.: Plant-Sediment Interactions in Salt Marshes - An Optode Imaging Study of O₂, pH, and CO₂ Gradients in the Rhizosphere, *Front Plant Sci*, 9, 541, <https://doi.org/10.3389/fpls.2018.00541>, 2018.

- 585 Kühl, M., Scholz, V., Sorrell, B. K., Koop-Jakobsen, K., Meier, R. J., and Mueller, P.: Plant-Mediated Rhizosphere Oxygenation in the Native Invasive Salt Marsh Grass *Elymus athericus*, *Frontiers in Plant Science*, 12, 669751, <https://doi.org/10.3389/FPLS.2021.669751>, 2021.
- Morris, J. T., Sundareshwar, P. V., Nietch, C. T., Kjerfve, B., and Cahoon, D. R.: Responses of coastal wetlands to rising sea level, *Ecology*, 83, 2869–2877, [https://doi.org/10.1890/0012-9658\(2002\)083%255B2869:ROCWTR%255D2.0.CO;2](https://doi.org/10.1890/0012-9658(2002)083%255B2869:ROCWTR%255D2.0.CO;2), 2002.
- 590 Moyano, F. E., Vasilyeva, N., and Menichetti, L.: Diffusion limitations and Michaelis–Menten kinetics as drivers of combined temperature and moisture effects on carbon fluxes of mineral soils, *Biogeosciences*, 15, 5031–5045, <https://doi.org/10.5194/bg-15-5031-2018>, 2018.
- Mueller, P., Jensen, K., and Megonigal, J. P.: Plants mediate soil organic matter decomposition in response to sea level rise, *Global Change Biology*, 22, 404–414, <https://doi.org/10.1111/gcb.13082>, 2016.
- 595 Mueller, P., Mozdzer, T. J., Langley, J. A., Aoki, L. R., Noyce, G. L., and Megonigal, J. P.: Plant species determine tidal wetland methane response to sea level rise, *Nature Communications* 2020 11:1, 11, 1–9, <https://doi.org/10.1038/s41467-020-18763-4>, 2020.
- Ouyang, X. and Lee, S. Y.: Updated estimates of carbon accumulation rates in coastal marsh sediments, *Biogeosciences*, 11, 5057–5071, <https://doi.org/10.5194/bg-11-5057-2014>, 2014.
- Pezeshki, S. R. and DeLaune, R. D.: Soil Oxidation-Reduction in Wetlands and Its Impact on Plant Functioning, *Biology*, 1, 196–221, <https://doi.org/10.3390/biology1020196>, 2012.
- 600 Rastetter, E. B., King, A. W., Cosby, B. J., Hornberger, G. M., O'Neill, R. V., and Hobbie, J. E.: Aggregating fine-scale ecological knowledge to model coarser-scale attributes of ecosystems, *Ecological Applications*, 2, 55–70, <https://doi.org/10.2307/1941889>, 1992.
- 605 Rietl, A. J., Megonigal, J. P., Herbert, E. R., and Kirwan, M. L.: Vegetation Type and Decomposition Priming Mediate Brackish Marsh Carbon Accumulation Under Interacting Facets of Global Change, *Geophysical Research Letters*, 48, 1–8, <https://doi.org/10.1029/2020GL092051>, 2021.
- Spivak, A. C., Sanderman, J., Bowen, J. L., Canuel, E. A., and Hopkins, C. S.: Global-change controls on soil-carbon accumulation and loss in coastal vegetated ecosystems, *Nature Geoscience*, 12, 685–692, <https://doi.org/10.1038/s41561-019-0435-2>, 2019.
- 610 Swanson, K. M., Drexler, J. Z., Schoellhamer, D. H., Thorne, K. M., Casazza, M. L., Overton, C. T., Callaway, J. C., and Takekawa, J. Y.: Wetland Accretion Rate Model of Ecosystem Resilience (WARMER) and Its Application to Habitat Sustainability for Endangered Species in the San Francisco Estuary, *Estuaries and Coasts*, 37, 476–492, <https://doi.org/10.1007/s12237-013-9694-0>, 2014.
- Van Der Nat, F. J. W. A. and Middelburg, J. J.: Effects of two common macrophytes on methane dynamics in freshwater sediments, *Biogeochemistry*, 43, 79–104, <https://doi.org/10.1023/A:1006076527187/METRICS>, 1998.
- 615 Visser, E. J. W., Colmer, T. D., Blom, C. W. P. M., and Voesenek, L. A. C. J.: Changes in growth, porosity, and radial oxygen loss from adventitious roots of selected mono- and dicotyledonous wetland species with contrasting types of aerenchyma, *Plant, Cell and Environment*, 23, 1237–1245, <https://doi.org/10.1046/j.1365-3040.2000.00628.x>, 2000.
- Wilson, C. H. and Gerber, S.: Theoretical insights from upscaling Michaelis–Menten microbial dynamics in biogeochemical models: a dimensionless approach, *Biogeosciences*, 18, 5669–5679, <https://doi.org/10.5194/bg-18-5669-2021>, 2021.

- 620 Zhang, Y., Li, C., Trettin, C. C., Li, H., Sun, G., Zhang, C. :, Li, C., Trettin, C. C., Li, H., and Sun, G.: An integrated model of
soil, hydrology, and vegetation for carbon dynamics in wetland ecosystems, *Global Biogeochemical Cycles*, 16, 9-1-9-17,
<https://doi.org/10.1029/2001gb001838>, 2002.
- Zhou, Y., O'Meara, T., Cardon, Z. G., Wang, J., Sulman, B. N., Giblin, A. E., and Forbrich, I.: Simulated plant-mediated
oxygen input has strong impacts on fine-scale porewater biogeochemistry and weak impacts on integrated methane fluxes in
625 coastal wetlands, *Biogeochemistry*, 167, 945-963, <https://doi.org/10.1007/s10533-024-01145-z>, 2024.