

# Formation drivers and photochemical effects of ClNO<sub>2</sub> in a coastal city of Southeast China

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**Abstract.** Nitryl chloride ( $\text{ClNO}_2$ ) is an important precursor of chlorine ( $\text{Cl}$ ) radical, significantly affecting ozone ( $\text{O}_3$ ) formation and photochemical oxidation. However, the key drivers of  $\text{ClNO}_2$  production are not fully understood. In this study, the field observations of  $\text{ClNO}_2$  and related parameters were conducted in a coastal city of Southeast China during the autumn of 2022, combining with machine learning and model simulations to elucidate its key influencing factors and atmospheric impacts. Elevated concentrations of  $\text{ClNO}_2$  ( $> 500$  ppt) were notably observed during nighttime in late autumn, accompanied by increased levels of dinitrogen pentoxide ( $\text{N}_2\text{O}_5$ ) and nitrate ( $\text{NO}_3^-$ ). Nighttime concentrations of  $\text{ClNO}_2$  peaked at 3.4 ppb, while its daytime levels remained significant, reaching up to 100 ppt and sustaining at approximately 40 ppt at noon. Machine learning and field observations identified nighttime  $\text{N}_2\text{O}_5$  heterogeneous uptake as the predominant pathway for  $\text{ClNO}_2$  production, whereas  $\text{NO}_3^-$  photolysis may contribute to its daytime generation. Additionally, ambient temperature ( $T$ ) and relative humidity ( $\text{RH}$ ) emerged as primary meteorological factors affecting  $\text{ClNO}_2$  formation, mainly through their effects on thermal equilibrium and  $\text{N}_2\text{O}_5$  hydrolysis processes, respectively. Ultraviolet ( $\text{UV}$ ) radiation was found to play a dual role in  $\text{ClNO}_2$  concentrations around noon. Box model simulations showed that under high  $\text{ClNO}_2$  conditions, the rates of alkane oxidation by  $\text{Cl}$  radical in the early morning exceeded those by  $\text{OH}$  radical. Consequently,  $\text{VOC}$  oxidation by  $\text{Cl}$  radical contributed  $\sim 19\%$  to  $\text{RO}_x$  production rates, thereby significantly impacting  $\text{O}_3$  formation and atmospheric oxidation capacity. This research enriched the understanding of  $\text{ClNO}_2$  generation and loss pathways, providing valuable insights for the regulation of photochemical pollution in coastal regions.

## 1 Introduction

Chlorine ( $\text{Cl}$ ) radical, as an important atmospheric oxidant, can react with volatile organic compounds ( $\text{VOCs}$ ) to affect  $\text{RO}_x$  (including  $\text{OH}$ ,  $\text{HO}_2$ , and  $\text{RO}_2$ ) radicals and ozone ( $\text{O}_3$ ) formation (Yi et al., 2023), thereby perturbing atmospheric chemical components and air quality (Peng et al., 2021; Li et al., 2020). The reaction rates between  $\text{Cl}$  radical and some alkanes are several orders of magnitude faster than those involving  $\text{OH}$  radical (Atkinson et al., 2006). Furthermore, the related studies indicated that the production rates of  $\text{Cl}$  radical in the early morning could significantly exceed the production rates of  $\text{OH}$  radical formed via  $\text{O}_3$  photolysis (Phillips et al., 2012; Tham et al., 2016), thereby enhancing the atmospheric oxidation capacity.

Nitryl chloride ( $\text{ClNO}_2$ ) is one of the major  $\text{Cl}$  radical precursors in the tropospheric atmosphere (Thornton et al., 2010; Xue et al., 2015; Liu et al., 2017). It is mainly generated by the heterogeneous uptake of dinitrogen pentoxide ( $\text{N}_2\text{O}_5$ ) on chloride-containing aerosols (Finlayson-Pitts et al., 1989; Thornton et al., 2010), among which  $\text{N}_2\text{O}_5$  is produced through the equilibrium reaction with nitrogen dioxide ( $\text{NO}_2$ ) and nitrate ( $\text{NO}_3$ ) radical. Since Osthoff et al. (2008) firstly detected over 1 ppb of  $\text{ClNO}_2$  in the urban outflows

of America (Osthoff et al., 2008), significant production of ClNO<sub>2</sub> has been widely observed in the polluted coastal and inland areas with abundant anthropogenic emissions and chloride sources, with concentrations ranging from tens of ppt to several ppb (Riedel et al., 2012; Mielke et al., 2013; Mielke et al., 2011; Phillips et al., 2012; Bannan et al., 2015; Wang et al., 2016; Xia et al., 2020; Xia et al., 2021; Yun et al., 2018; Wang et al., 2022; Li et al., 2023). For the diurnal profile of ClNO<sub>2</sub>, its concentrations generally peaked and accumulated at midnight, then rapidly decreased to low levels due to strong photolysis after sunrise (Ma et al., 2023; Mielke et al., 2011; Xia et al., 2020). However, elevated daytime concentrations of ClNO<sub>2</sub> have been observed in field studies, mainly attributed to reduced photolysis rates under heavy cloud or fog cover, as well as contributions from horizontal and vertical transport (Tham et al., 2016; Xia et al., 2021; Jeong et al., 2019; Mielke et al., 2013; Bannan et al., 2015). Notably, the recent laboratory research demonstrated that nitrate (NO<sub>3</sub><sup>-</sup>) photolysis can generate ClNO<sub>2</sub> alongside Cl<sub>2</sub> (Dalton et al., 2023), yet this mechanism has not been confirmed under real atmospheric conditions.

At present, the observation studies of ClNO<sub>2</sub> focused on investigating its influencing factors, such as the N<sub>2</sub>O<sub>5</sub> uptake coefficient and the production yield of ClNO<sub>2</sub> (Thornton et al., 2003; Tham et al., 2018). The field and laboratory studies have indicated that ClNO<sub>2</sub> production was mainly affected by ambient temperature (T), relative humidity (RH), and particle components (e.g., chloride (Cl<sup>-</sup>), NO<sub>3</sub><sup>-</sup>, and liquid water content) (Bertram and Thornton, 2009; Wang et al., 2023; Wang et al., 2020). In addition to influencing factors, the photochemical effects of ClNO<sub>2</sub> photolysis have been extensively evaluated (Xue et al., 2015; Xia et al., 2021; Tham et al., 2016). Cl radical released by ClNO<sub>2</sub> photolysis will oxidize VOCs to promote the formation of RO<sub>2</sub> radical and O<sub>3</sub>, greatly compensating for the underestimation of RO<sub>2</sub> radical and O<sub>3</sub> generation in model simulations (Peng et al., 2021; Ma et al., 2023). The field measurements of ClNO<sub>2</sub> have been conducted in different atmospheric environments, while the key drivers of ClNO<sub>2</sub> chemistry were still not well recognized. Moreover, it is pertinent to explore whether there are additional and unrecognized sources of ClNO<sub>2</sub> beyond its heterogeneous generation from N<sub>2</sub>O<sub>5</sub>.

In this study, the comprehensive measurements of ClNO<sub>2</sub> and related parameters were conducted in a coastal city of Southeast China during the autumn of 2022. Field observations, combined with a machine learning model, were used to reveal the key driving factors of ClNO<sub>2</sub> formation. Furthermore, we further investigated the potential mechanisms driving daytime ClNO<sub>2</sub> generation. Additionally, we also assessed the photochemical impacts of ClNO<sub>2</sub> based on a box model. Overall, this study underscored the important role of NO<sub>3</sub><sup>-</sup> in the ClNO<sub>2</sub> chemistry.

## 2 Materials and methods

### 2.1 Field Measurements

The intensive field measurements of ClNO<sub>2</sub>, related precursors, and meteorological parameters from October 9th to December 5th, 2022 were performed at an urban site (Institute of Urban Environment,

Chinese Academy of Sciences) in a coastal city (Xiamen) of Southeast China (Fig. S1). Here, ClNO<sub>2</sub>, N<sub>2</sub>O<sub>5</sub>, gaseous pollutants (volatile organic compounds (VOCs), NO<sub>x</sub>, SO<sub>2</sub>, CO, and O<sub>3</sub>), aerosol mass concentrations, ionic components, size distribution, and meteorological factors were simultaneously detected. Meanwhile, an iodide-adduct time-of-flight chemical ionization mass spectrometer (I<sup>-</sup>-ToF-CIMS) was used to measure ClNO<sub>2</sub> and N<sub>2</sub>O<sub>5</sub>. The principles and settings of I<sup>-</sup>-ToF-CIMS were similar with previous studies (Ma et al., 2023; Yan et al., 2023). Detailed descriptions of this observation site and instruments have been provided in previous work (Chen et al., 2024; Hu et al., 2022), Text S1, and Table S1. For the calibrations of ClNO<sub>2</sub> and N<sub>2</sub>O<sub>5</sub>, ClNO<sub>2</sub> was produced by passing Cl<sub>2</sub> (6 ppm in N<sub>2</sub>) through a moist mixture of sodium nitrite (NaNO<sub>2</sub>) and sodium chloride (NaCl) (Thaler et al., 2011; Wang et al., 2022), and N<sub>2</sub>O<sub>5</sub> was synthesized by the reactions of O<sub>3</sub> and excessive NO<sub>2</sub> (Tham et al., 2016; Wang et al., 2016). The dependences of ClNO<sub>2</sub> and N<sub>2</sub>O<sub>5</sub> sensitivities on relative humidity are presented in Fig. S2. The uncertainties of the ClNO<sub>2</sub> and N<sub>2</sub>O<sub>5</sub> measurements were estimated to be ~15 and 12 %, respectively. The details of ClNO<sub>2</sub> and N<sub>2</sub>O<sub>5</sub> calibrations and uncertainty analysis are displayed in Text S2.

## 2.2 Machine Learning model

Here, the extreme gradient boosting (XGBoost) model coupling with the Shapely additive explanations (SHAP) model (the XGBoost-SHAP model) was used to identify the key influencing factors of ClNO<sub>2</sub> formation. Meanwhile, the XGBoost model was applied to establish the predictive model of ClNO<sub>2</sub> based on the observed data of gaseous precursors and meteorological factors; the SHAP model was employed to evaluate the importance of each feature affecting the simulated concentrations of ClNO<sub>2</sub>. The SHAP model is an interpretability tool designed to analyze the contributions of individual features to model predictions. It employs an additive explanatory framework that considers all features as contributors, drawing inspiration from cooperative game theory. For each predicted instance, SHAP assigns a Shapley value, representing the cumulative contribution of each feature. Positive SHAP values indicate that a feature increases the model's predicted outcome, signifying a positive contribution. Conversely, negative SHAP values suggest that the feature reduces the predicted value, reflecting a negative contribution. The absolute value of the SHAP score reflects the magnitude of the contribution, regardless of direction, offering insight into the overall importance of the feature. The true value, on the other hand, reveals the direction of the contribution (positive or negative), facilitating a clearer understanding of the relationship between the feature and the prediction. Besides, the partial dependence plot (PDP) analysis offers a visual representation of the marginal effect that the factors have on the model's predicted outcome. It is based on the principle of stabilizing the values of non-target features, and systematically altered the target feature's values according to the model's algorithmic framework to derive the predicted values.

ClNO<sub>2</sub> concentrations served as the dependent variable, with trace gases (SO<sub>2</sub>, CO, NO<sub>2</sub>, NO, O<sub>3</sub>, and N<sub>2</sub>O<sub>5</sub>), PM<sub>2.5</sub> and its inorganic compositions (NO<sub>3</sub><sup>-</sup>, SO<sub>4</sub><sup>2-</sup>, NH<sub>4</sub><sup>+</sup>, and Cl<sup>-</sup>), and meteorological parameters (T, RH, UV, WS, WD, and BLH) acting as independent variables. The simulated ClNO<sub>2</sub> concentrations by the

XGBoost model were highly similar with the observed values ( $R^2=0.91$ ), indicating the good performance of the XGBoost model (Fig. S3). Detailed introductions and settings of the XGBoost-SHAP model are provided in Text S3.

## 2.3 The box model

The observation-based model (OBM) was utilized to assess the impacts of  $\text{ClNO}_2$  on photochemically atmospheric oxidation. As delineated in earlier studies (Xue et al., 2015; Tham et al., 2016; Xia et al., 2021; Peng et al., 2021; Peng et al., 2022), the Master Chemical Mechanism (MCM, version 3.3.1) was adopted, and established chlorine chemistry mechanisms have been integrated. The Tropospheric Ultraviolet and Visible Radiation (TUV) model was used to calculate  $\text{ClNO}_2$  photolysis rates ( $J\text{ClNO}_2$ ) under clear-sky conditions. The simulated  $J\text{ClNO}_2$  values were then scaled based on field-measured  $J\text{NO}_2$  values. A thorough exposition of the box model configuration can be found in our previous publications (Liu et al., 2022b; Liu et al., 2022a) and Text S4. Observation data, including  $\text{ClNO}_2$ , VOCs, HCHO, HONO, CO,  $\text{O}_3$ , NO,  $\text{NO}_2$ ,  $\text{SO}_2$ , along with meteorological factors as constraint were input into the box model at an hourly resolution (Table S2). Two scenarios were examined: one representing observation-average conditions from October 9th to December 5th, the other reflecting a high  $\text{ClNO}_2$  case observed on November 28th.

This study focused on elucidating the influence of  $\text{ClNO}_2$  on the formation of  $\text{RO}_x$  radical and  $\text{O}_3$ . The  $\text{O}_3$  production rate minus the  $\text{O}_3$  loss rate was used to calculate the net  $\text{O}_3$  production rate (Eq. S1-3). The AOC is calculated by the sum of the rates of  $\text{CH}_4$ , CO, and VOCs oxidized by atmospheric oxidants ( $\text{O}_3$ , OH, Cl, and  $\text{NO}_3$  radical) (Eq. S4) (Xue et al., 2015; Yi et al., 2023). Both scenarios were evaluated with and without including  $\text{ClNO}_2$  inputs to assess its impacts on these processes.

## 3 Results and discussion

### 3.1 Overview of observations

Fig. 1 displays the time series of  $\text{ClNO}_2$ ,  $\text{N}_2\text{O}_5$ , and related parameters including  $\text{O}_3$ ,  $\text{NO}_x$ ,  $\text{PM}_{2.5}$ ,  $\text{Cl}^-$ ,  $\text{NO}_3^-$ , and meteorological parameters during the autumn observation period. Our observation shows a decline in T and UV values from October to November, with average RH values increasing from ~ 60% in October to ~ 70% in November (excluding rainy days). During the entire measurement period,  $\text{ClNO}_2$  concentrations exhibited significant variability, with elevated levels (> 500 ppt) frequently observed in late autumn, particularly after November 10th. The elevation of  $\text{ClNO}_2$  concentrations coincided with increased levels of  $\text{N}_2\text{O}_5$  and  $\text{NO}_3^-$  during late autumn. The concentrations of  $\text{ClNO}_2$  at our study site reached several ppb, compared with previous field measurements conducted at urban, suburban, rural, background, and mountain sites (Table S3), indicating its widespread presence in diverse atmospheric environments. The highest concentrations of  $\text{ClNO}_2$  were detected during the night of November 27th, with a maximum hourly average of 3.4 ppb. Peak concentrations of  $\text{N}_2\text{O}_5$  and  $\text{NO}_3^-$  were also observed on that night (Fig. 1). On the evening of November 27th,  $\text{N}_2\text{O}_5$  concentrations rapidly decreased after 7 p.m., while  $\text{ClNO}_2$  and  $\text{NO}_3^-$

concentrations significantly increased, reflecting fast  $\text{N}_2\text{O}_5$  heterogeneous hydrolysis and effective formation of  $\text{ClNO}_2$ . Notably, on the following day (November 28th) (Fig. 2a),  $\text{ClNO}_2$  concentrations sustained above 100 ppt around noon, partially related with weakened UV values ( $\sim 14 \text{ W}\cdot\text{m}^{-2}$ ) under heavy fog and cloud cover, with the RH values of  $\sim 70\%$  at that time. Similar research in California has shown  $\text{ClNO}_2$  concentrations exceeding 100 ppt after sunrise 4 hours due to reduced photolysis (Mielke et al., 2013).

The average diurnal changes of  $\text{ClNO}_2$  and related parameters during the entire measurement campaign are depicted in Fig. 2b. As expected,  $\text{ClNO}_2$  exhibited a distinct diurnal variation, peaking and accumulating after sunset and decreasing in the early morning. However,  $\text{ClNO}_2$  concentrations remained  $\sim 40$  ppt around noon, different with some studies that  $\text{ClNO}_2$  concentrations decreased to near the detection limit around midday (Wang et al., 2022; Niu et al., 2022). Similar observation in North China declared  $\text{ClNO}_2$  concentrations above 60 ppt in the afternoon (Liu et al., 2017). Previous studies have indicated that abundant  $\text{ClNO}_2$  may be transported from upper atmosphere or air mass, contributing to the elevated  $\text{ClNO}_2$  concentrations in the early morning (Tham et al., 2016; Xia et al., 2021; Jeong et al., 2019). However, the explanations for the concentrations of  $\text{ClNO}_2$  around noon remained elusive.

To evaluate the contribution of the heterogeneous  $\text{N}_2\text{O}_5$  uptake to daytime  $\text{ClNO}_2$  levels, we calculated  $\text{ClNO}_2$  production using a box model, considering (1) the contribution of heterogeneous  $\text{N}_2\text{O}_5$  uptake to  $\text{ClNO}_2$  production, and (2)  $\text{ClNO}_2$  loss via photolysis, aerosol uptake, and reaction with  $\text{OH}\cdot$  (Text S5 and S6). We used a  $\gamma(\text{N}_2\text{O}_5)$  value of 0.06, a  $\phi(\text{ClNO}_2)$  value of 1.0, and a  $\gamma(\text{ClNO}_2)$  value of 0.006 in our calculations, which represent upper-end estimates based on previous field studies (McDuffie et al., 2018a; McDuffie et al., 2018b; Tham et al., 2016). However, as shown in Fig. 3, the simulated daytime  $\text{ClNO}_2$  concentrations were lower than the observed values. Therefore, we believe that the observed daytime  $\text{ClNO}_2$  levels, particularly around noon, cannot be adequately explained by heterogeneous  $\text{N}_2\text{O}_5$  uptake alone, suggesting the presence of additional sources contributing to the formation of daytime  $\text{ClNO}_2$ .

### 3.2 Key drivers of $\text{ClNO}_2$ formation

The XGBoost-SHAP model was employed to investigate the major drivers of  $\text{ClNO}_2$  production during the whole observation period. The average absolute SHAP value of each feature was ranked to determine the key drivers of  $\text{ClNO}_2$  formation, with larger SHAP values suggesting greater contributions (Fig. 4a). Additionally, features with positive SHAP values (depicted as red points) indicate that higher values of those features positively affect  $\text{ClNO}_2$  concentrations, and vice versa (Fig. 4b). Overall,  $\text{N}_2\text{O}_5$ ,  $\text{NO}_3^-$ , T, RH, and UV were the most important features affecting  $\text{ClNO}_2$  concentrations. Notably, these factors exhibited varied behaviors between daytime and nighttime periods.

In our study,  $\text{N}_2\text{O}_5$  was identified as the most important influencing factor, consistent with its role in  $\text{ClNO}_2$  formation through heterogeneous uptake processes (Thornton et al., 2010; Finlayson-Pitts et al., 1989). After sunset,  $\text{ClNO}_2$  concentrations markedly increased due to active nighttime  $\text{N}_2\text{O}_5$  chemistry, while this heterogeneous uptake process was hindered after sunrise as  $\text{N}_2\text{O}_5$  concentrations decreased

significantly (Fig. 1) (Niu et al., 2022; Wang et al., 2020; Tan et al., 2022). Indeed, the concentrations of ClNO<sub>2</sub> were evidently increased when N<sub>2</sub>O<sub>5</sub> concentrations exceeded ~13 ppt, predominantly during the nighttime (Fig. 5a). Conversely, in Northern Europe, the ClNO<sub>2</sub> concentrations were mainly controlled by O<sub>3</sub> and NO<sub>2</sub>, rather than by the heterogeneous uptake of N<sub>2</sub>O<sub>5</sub> (Sommariva et al., 2018). In Heshan of South China, chloride and PM<sub>2.5</sub> were the major factors affecting ClNO<sub>2</sub> formation (Wang et al., 2022). Differently, the relative importance of NO<sub>3</sub><sup>-</sup> derived from the XGBoost-SHAP result indicated that elevated ClNO<sub>2</sub> concentrations were associated with high concentrations of NO<sub>3</sub><sup>-</sup> besides N<sub>2</sub>O<sub>5</sub>. According to Fig. 5b, high NO<sub>3</sub><sup>-</sup> concentrations (> 3.7 μg·m<sup>-3</sup>) are accompanied by the elevation of ClNO<sub>2</sub>, especially its concentrations reaching 6.2 μg·m<sup>-3</sup>. Previous studies suggested that increased concentrations of NO<sub>3</sub><sup>-</sup> decreased γ(N<sub>2</sub>O<sub>5</sub>), which would limit the production of ClNO<sub>2</sub> (Wahner et al., 1998; Mentel et al., 1999; Bertram and Thornton, 2009). As depicted in Fig. S4, the dependence of γ(N<sub>2</sub>O<sub>5</sub>) on NO<sub>3</sub><sup>-</sup> concentrations follows the nitrate suppression effect. Therefore, the importance of nighttime NO<sub>3</sub><sup>-</sup> for ClNO<sub>2</sub> levels is that they are co-products from the processes of N<sub>2</sub>O<sub>5</sub> heterogeneous uptake. As shown in Fig. 1, compared to low NO<sub>3</sub><sup>-</sup> conditions, ClNO<sub>2</sub> production was enhanced in high NO<sub>3</sub><sup>-</sup> conditions. Especially in late autumn, increased aerosol abundances and N<sub>2</sub>O<sub>5</sub> levels enhanced N<sub>2</sub>O<sub>5</sub> uptake further promoting ClNO<sub>2</sub> and NO<sub>3</sub><sup>-</sup> production. Considering the limited contribution of N<sub>2</sub>O<sub>5</sub> hydrolysis to daytime NO<sub>3</sub><sup>-</sup> levels (Yan et al., 2023; Zang et al., 2022; Chen et al., 2020), the impact of high NO<sub>3</sub><sup>-</sup> concentrations on daytime ClNO<sub>2</sub> concentrations warrants further analysis.

The simulated concentrations of ClNO<sub>2</sub>, based on the XGBoost-SHAP model, were significantly elevated when NO<sub>3</sub><sup>-</sup> concentrations were higher than 3.7 μg·m<sup>-3</sup> (Fig. 5b). Consequently, the average daily concentrations of NO<sub>3</sub><sup>-</sup> were classified as high (> 3.7 μg·m<sup>-3</sup>) and low (< 3.7 μg·m<sup>-3</sup>) to further elucidate the impacts of NO<sub>3</sub><sup>-</sup> on the formation of ClNO<sub>2</sub>. Fig. 6 presents the diurnal variations in the relative importance of factors based on the SHAP values under high and low NO<sub>3</sub><sup>-</sup> concentrations. Unexpectedly, daytime NO<sub>3</sub><sup>-</sup> was the dominant influencing factors for daytime ClNO<sub>2</sub> (Fig. 6a). High concentrations of daytime NO<sub>3</sub><sup>-</sup> positively affected the daytime concentrations of ClNO<sub>2</sub>, independent of N<sub>2</sub>O<sub>5</sub> uptake processes. As depicted in Fig. 6a, daytime N<sub>2</sub>O<sub>5</sub> did not promote the elevation of daytime ClNO<sub>2</sub>. Negative SHAP values for N<sub>2</sub>O<sub>5</sub> during the daytime indicate that the contribution of N<sub>2</sub>O<sub>5</sub> chemistry to daytime ClNO<sub>2</sub> levels was limited. Therefore, it is very likely that high concentrations of daytime NO<sub>3</sub><sup>-</sup> participated in daytime ClNO<sub>2</sub> production. A recent study suggested that nitrate photolysis produced ClNO<sub>2</sub> in addition to Cl<sub>2</sub> (Dalton et al., 2023), while it has been not verified by field observations. Fig. 7 shows that daytime ClNO<sub>2</sub> concentrations correlated well (R=0.62) with the product of a proxy of NO<sub>3</sub><sup>-</sup> photolysis (NO<sub>3</sub><sup>-</sup>×JNO<sub>2</sub>×S<sub>a</sub>) on aerosol surfaces, implying that the photolysis of NO<sub>3</sub><sup>-</sup> likely resulted in the daytime formation of ClNO<sub>2</sub> at our study site. Furthermore, high concentrations of NO<sub>3</sub><sup>-</sup> and Cl<sup>-</sup>, along with large values of S<sub>a</sub> (Fig. 7a, b, c and Fig. S5) in the daytime accelerated NO<sub>3</sub><sup>-</sup> photolysis, promoting the formation of ClNO<sub>2</sub>, while ClNO<sub>2</sub> concentrations exhibited a weak correlation with JNO<sub>2</sub>. It should be emphasized that the weak correlation

between  $\text{JNO}_2$  and  $\text{ClNO}_2$  concentrations does not deny the potential contribution of nitrate photolysis, which could be explained by the fact that  $\text{ClNO}_2$  concentrations are affected by both its production and loss processes. Specifically, photolysis rates exert dual effects on daytime  $\text{ClNO}_2$  concentrations: positive effects through photochemical production pathways and negative effects through direct  $\text{ClNO}_2$  photolytic loss. Given the short daytime lifetime of  $\text{ClNO}_2$ , we calculated the missing  $\text{ClNO}_2$  production rate (production rate minus loss rate) to assess the contributions from unknown sources (Text S6). The production rates of unknown sources showed a good correlation with  $\text{JNO}_2$  ( $R=0.41$ ) (Fig. S6), indicating that photochemical processes may enhance  $\text{ClNO}_2$  production. Notably, the strong correlation between the observed concentrations of  $\text{ClNO}_2$  and the  $\text{NO}_3^-$  photolysis proxy ( $\text{NO}_3^- \times \text{JNO}_2 \times S_a$ ) has revealed the possibility of the contribution of  $\text{NO}_3^-$  photolysis to the unknown daytime  $\text{ClNO}_2$  source.

In term of meteorological factors, UV, T, and RH were the major influencing factors. The photolysis was the most important sink of  $\text{ClNO}_2$  in the daytime, leading to a rapid reduction in  $\text{ClNO}_2$  concentrations, particularly in the early morning (Fig. 5e and Fig. 6). However, it is crucial to understand the dual role of photolysis intensity in determining daytime  $\text{ClNO}_2$  levels. As mentioned before, photolysis can contribute to the generation of  $\text{ClNO}_2$  by promoting  $\text{NO}_3^-$  photolysis, while also causing the rapid decomposition of  $\text{ClNO}_2$ . As reported in California (Mielke et al., 2013), reduced photolysis rates even increased daytime  $\text{ClNO}_2$  levels by decreasing  $\text{ClNO}_2$  loss through photolysis. The impact of ambient temperature on  $\text{ClNO}_2$  was probably reflected in its thermal equilibrium with  $\text{N}_2\text{O}_5$ . Elevated daytime ambient temperature suppressed the formation of  $\text{N}_2\text{O}_5$ , resulting in low  $\text{N}_2\text{O}_5$  concentrations, which further limited the contribution of heterogeneous  $\text{N}_2\text{O}_5$  uptake to daytime  $\text{ClNO}_2$  generation (Fig. 5c and Fig. 6). During the whole observation period from October to November, the drop in ambient temperature facilitated  $\text{ClNO}_2$  production by decreasing the thermal decomposition process. Increased RH values provided favorable conditions for the nighttime  $\text{N}_2\text{O}_5$  hydrolysis reactions, thereby affecting  $\text{ClNO}_2$  production (Fig. 5d and Fig. 6), while high RH ( $> 80\%$ ) also weakened the generation of  $\text{ClNO}_2$ . Notably,  $\text{Cl}^-$  was not the most important factors of  $\text{ClNO}_2$  formation at our study site (Fig. 4), likely attributed to the abundant chlorine source in coastal regions (Peng et al., 2022).

### 3.3 Impact of $\text{ClNO}_2$ photolysis on $\text{RO}_x$ budget

The photochemical effects of  $\text{ClNO}_2$  were evaluated under the observation-average condition and the high  $\text{ClNO}_2$  case based on the box model. The largest Cl production rates ( $P(\text{Cl})$ ) contributed from  $\text{ClNO}_2$  photolysis were  $0.05 \text{ ppb}\cdot\text{h}^{-1}$  for the observation-average condition, which was lower than  $0.19 \text{ ppb}\cdot\text{h}^{-1}$  for the high  $\text{ClNO}_2$  case. The difference led to variable levels of atmospheric oxidation capacity induced by Cl radical. Cl radical released via the photolysis of  $\text{ClNO}_2$  initiated the oxidation of VOCs. Among VOC groups (including alkanes, alkenes, alkynes, aromatics, and OVOCs), Cl radical primarily oxidized alkanes ( $\sim 65.0\%$ ), followed by OVOCs ( $\sim 12.7\%$ ) for both the observation-average condition and the high  $\text{ClNO}_2$  case (Fig. 8a, b). The contributions of Cl radical and other atmospheric oxidants (including OH radical and



O<sub>3</sub>) to daytime VOC oxidation were also compared (Fig. 8c, d and Table 1). In our study, the oxidation of alkanes by Cl radical for the observation-average condition were about 11.7%, which increased by 44.8% for the high ClNO<sub>2</sub> case, were higher than those in London (Bannan et al., 2015), Weybourne (Bannan et al., 2017), Boston (Rutherford et al., 1995), and LA (Fraser et al., 1997), lower than that in Hong Kong (Xue et al., 2015). It should be noticed that the rates of Cl radical reacting with alkanes even exceeded those of OH radical in the early morning for the high ClNO<sub>2</sub> case. The largest rates of alkanes oxidized by Cl radical were approximately twice as high as those of OH radical at 10 a.m. (Fig. 8e, f), highlighting that the photochemical effects of Cl radical released via ClNO<sub>2</sub> photolysis were particularly important for VOC oxidation during the morning hours at our study site.

The oxidation of VOCs by Cl radical further affects the generation of RO<sub>x</sub> (OH + HO<sub>2</sub> + RO<sub>2</sub>) radicals. The RO<sub>x</sub> radical production rates for the high ClNO<sub>2</sub> case were evidently lower than that under the observation-average condition, primarily due to reduced photolysis rates on that day. However, the total RO<sub>x</sub> radical production rates averagely increased by 23.8% with ClNO<sub>2</sub> photolysis for the high ClNO<sub>2</sub> case, higher than a 4.9% increase for the observation-average condition (Fig. S7). For the observation-average condition, O<sub>3</sub> (32.7%), HONO (31.7%), and OVOCs (21.6%) photolysis were the most significant contributors to RO<sub>x</sub> radical production in the early morning (7-10 a.m.), with VOC oxidation by Cl radical contributing only 3.7% (Fig. 9a). However, for the high ClNO<sub>2</sub> case, VOC oxidation induced by Cl radical in the early morning accounted for 19.1% of RO<sub>x</sub> radical production, which was higher than O<sub>3</sub> (7.4%) and HCHO (4.1%) photolysis, close to OVOCs (19.0%) photolysis (Fig. 9b). The contributions of ClNO<sub>2</sub> photolysis to the RO<sub>x</sub> radical production rates in our study were larger than previous results observed in autumn of Heshan (Wang et al., 2022) and North China (Xia et al., 2021), similar with that in summer of Wangdu (Tham et al., 2016). Thus, the concentrations of OH, HO<sub>2</sub>, and RO<sub>2</sub> radicals in the box model with ClNO<sub>2</sub> inputs averagely increased by 17.9%, 34.6%, and 54.3% for the high ClNO<sub>2</sub> case, higher than the increases of 3.7%, 7.1%, and 10.3% contributed from the observation-average conditions, respectively (Fig. S8). The uplift in the concentrations of RO<sub>x</sub> radicals also accelerated the generation of O<sub>3</sub>. The increase in the net O<sub>3</sub> production rates (P(O<sub>3</sub>)) for the observation-average condition averagely reached 0.13 ppb·h<sup>-1</sup> (15.8 %) in the daytime (Fig. 10a), while larger elevations in the net P(O<sub>3</sub>) were observed for the high ClNO<sub>2</sub> case (Fig. 10b), with a maximum of 0.64 ppb·h<sup>-1</sup> (120 %) at 10 a.m. As a result, increased RO<sub>x</sub> radical and O<sub>3</sub> greatly enhanced the atmospheric oxidation capacity (Fig. 10c, d), especially for the high ClNO<sub>2</sub> case (up to 65%).

Table 2 summarizes the impacts of ClNO<sub>2</sub> photolysis on RO<sub>x</sub> radical and O<sub>3</sub> production in our study and previous observations around the world (Xia et al., 2021; Wang et al., 2022; Tham et al., 2016; Wang et al., 2016; Xue et al., 2015; Bannan et al., 2017; Jeong et al., 2019), indicating that the photochemical impacts of ClNO<sub>2</sub> were variable in different atmospheric environments. At our study site, the effects of ClNO<sub>2</sub> photolysis on RO<sub>x</sub> radical production were important, especially in the early morning. The enhanced

RO<sub>x</sub> radical production induced by ClNO<sub>2</sub> photolysis accelerated the chemical generation of O<sub>3</sub>. Primary RO<sub>x</sub> radical production rates (including O<sub>3</sub>, HONO, HCHO, OVOCs, and ClNO<sub>2</sub>) were considered as one of the most important parameters to O<sub>3</sub> formation (Lu et al., 2023). Therefore, the considerable contribution of ClNO<sub>2</sub> photolysis to primary RO<sub>x</sub> radical production in the early morning may bring new challenges for O<sub>3</sub> alleviation.

## Conclusions

In conclusion, we present two months of field measurements in the coastal area of Southern China during the autumn, coupled with machine learning and model simulations, providing new insights into ClNO<sub>2</sub> chemistry. Our observation shows the increase in the concentrations of ClNO<sub>2</sub> were accompanied by elevated concentrations of N<sub>2</sub>O<sub>5</sub> and NO<sub>3</sub><sup>-</sup>, low values of T and UV, and high values of RH. The nighttime heterogeneous uptake of N<sub>2</sub>O<sub>5</sub> was identified as the major source of ClNO<sub>2</sub>, while NO<sub>3</sub><sup>-</sup> photolysis served as a potential daytime ClNO<sub>2</sub> source. Cl radical released by ClNO<sub>2</sub> photolysis after sunrise had important photochemical effects in the early morning. The photolysis of high ClNO<sub>2</sub> concentrations resulted in net O<sub>3</sub> production rates and atmospheric oxidation capacity levels increasing by 120% and 65%, respectively. Our results enhanced the understanding of ClNO<sub>2</sub> chemistry in coastal regions, calling for more observations and laboratory research to fully reveal its exact role in different atmospheric environments.

**Data availability.** Data are available upon request to Jinsheng Chen ([jschen@iue.ac.cn](mailto:jschen@iue.ac.cn)).

**Author contributions.** JC provided funding support for field measurements, designed this study, and revised this manuscript. GC designed this study, analyzed the data, and wrote this manuscript. HW helped perform the calibrations and revised this manuscript. XF revised this manuscript. XF, HW, YT, ZL, XJ, LX, BH contributed to discussions of this manuscript.

**Competing interests.** The authors declare that they have no conflict of interest.

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## References

Atkinson, R., Baulch, D. L., Cox, R. A., Crowley, J. N., Hampson, R. F., Hynes, R. G., Jenkin, M. E., Rossi, M. J., Troe, J., and Subcommittee, I.: Evaluated kinetic and photochemical data for atmospheric chemistry: Volume II – gas phase reactions of organic species, *Atmos. Chem. Phys.*, 6, 3625-4055, <https://doi.org/10.5194/acp-6-3625-2006>, 2006.

Bannan, T. J., Bacak, A., Le Breton, M., Flynn, M., Ouyang, B., McLeod, M., Jones, R., Malkin, T. L., Whalley, L. K., Heard, D. E., Bandy, B., Khan, M. A. H., Shallcross, D. E., and Percival, C. J.: Ground and Airborne U.K. Measurements of Nitryl Chloride: An Investigation of the Role of Cl Atom Oxidation at Weybourne Atmospheric Observatory, *J. Geophys. Res. Atmos.*, 122, 11,154-11,165, <https://doi.org/10.1002/2017jd026624>, 2017.

Bannan, T. J., Booth, A. M., Bacak, A., Muller, J. B. A., Leather, K. E., Le Breton, M., Jones, B., Young, D., Coe, H., Allan, J., Visser, S., Slowik, J. G., Furger, M., Prévôt, A. S. H., Lee, J., Dunmore, R. E., Hopkins, J. R., Hamilton, J. F., Lewis, A. C., Whalley, L. K., Sharp, T., Stone, D., Heard, D. E., Fleming, Z. L., Leigh, R., Shallcross, D. E., and Percival, C. J.: The first UK measurements of nitryl chloride using a chemical ionization mass spectrometer in central London in the summer of 2012, and an investigation of the role of Cl atom oxidation, *J. Geophys. Res. Atmos.*, 120, 5638-5657, <https://doi.org/10.1002/2014jd022629>, 2015.

Bertram, T. and Thornton, J.: Toward a general parameterization of N<sub>2</sub>O<sub>5</sub> reactivity on aqueous particles: the competing effects of particle liquid water, nitrate and chloride, *Atmos. Chem. Phys.*, 9, 8351-8363, <https://doi.org/10.5194/acp-9-8351-2009>, 2009.

Chen, G., Ji, X., Chen, J., Xu, L., Hu, B., Lin, Z., Fan, X., Li, M., Hong, Y., and Chen, J.: Photochemical pollution during summertime in a coastal city of Southeast China: Ozone formation and influencing factors, *Atmos. Res.*, 301, 107270, <https://doi.org/10.1016/j.atmosres.2024.107270>, 2024.

Chen, X., Wang, H., Lu, K., Li, C., Zhai, T., Tan, Z., Ma, X., Yang, X., Liu, Y., Chen, S., Dong, H., Li, X., Wu, Z., Hu, M., Zeng, L., and Zhang, Y.: Field Determination of Nitrate Formation Pathway in Winter Beijing, *Environ. Sci. Technol.*, 54, 9243-9253, <https://doi.org/10.1021/acs.est.0c00972>, 2020.

383 Dalton, E. Z., Hoffmann, E. H., Schaefer, T., Tilgner, A., Herrmann, H., and Raff, J. D.: Daytime  
 384 Atmospheric Halogen Cycling through Aqueous-Phase Oxygen Atom Chemistry, *J. Am. Chem. Soc.*, 145,  
 385 15652-15657, <https://doi.org/10.1021/jacs.3c03112>, 2023.

386 Finlayson-Pitts, B. J., Ezell, M. J., and Pitts, J. N.: Formation of chemically active chlorine compounds by  
 387 reactions of atmospheric NaCl particles with gaseous  $\text{N}_2\text{O}_5$  and  $\text{ClONO}_2$ , *Nature.*, 337, 241-244,  
 388 <https://doi.org/10.1038/337241a0>, 1989.

389 Fraser, M. P., Cass, G. R., Simoneit, B. R., and Rasmussen, R.: Air quality model evaluation data for  
 390 organics. 4. C2–C36 non-aromatic hydrocarbons, *Environ. Sci. Technol.*, 31, 2356-2367,  
 391 <https://doi.org/10.1021/es960980g>, 1997.

392 Hu, B., Duan, J., Hong, Y., Xu, L., Li, M., Bian, Y., Qin, M., Fang, W., Xie, P., and Chen, J.: Exploration of  
 393 the atmospheric chemistry of nitrous acid in a coastal city of southeastern China: results from measurements  
 394 across four seasons, *Atmos. Chem. Phys.*, 22, 371-393, <https://doi.org/10.5194/acp-22-371-2022>, 2022.

395 Jeong, D., Seco, R., Gu, D., Lee, Y., Nault, B. A., Knote, C. J., McGee, T., Sullivan, J. T., Jimenez, J. L.,  
 396 Campuzano-Jost, P., Blake, D. R., Sanchez, D., Guenther, A. B., Tanner, D., Huey, L. G., Long, R.,  
 397 Anderson, B. E., Hall, S. R., Ullmann, K., Shin, H., Herndon, S. C., Lee, Y., Kim, D., Ahn, J., and Kim, S.:  
 398 Integration of airborne and ground observations of nitryl chloride in the Seoul metropolitan area and the  
 399 implications on regional oxidation capacity during KORUS-AQ 2016, *Atmos. Chem. Phys.*, 19, 12779-  
 400 12795, <https://doi.org/10.5194/acp-19-12779-2019>, 2019.

401 Li, F., Huang, D. D., Nie, W., Tham, Y. J., Lou, S., Li, Y., Tian, L., Liu, Y., Zhou, M., and Wang, H.:  
 402 Observation of nitrogen oxide-influenced chlorine chemistry and source analysis of  $\text{Cl}_2$  in the Yangtze River  
 403 Delta, China, *Atmos. Environ.*, 306, 119829, <https://doi.org/10.1016/j.atmosenv.2023.119829>, 2023.

404 Li, Q., Badia, A., Wang, T., Sarwar, G., Fu, X., Zhang, L., Zhang, Q., Fung, J., Cuevas, C. A., Wang, S.,  
 405 Zhou, B., and Saiz-Lopez, A.: Potential Effect of Halogens on Atmospheric Oxidation and Air Quality in  
 406 China, *J. Geophys. Res. Atmos.*, 125, e2019JD032058, <https://doi.org/10.1029/2019JD032058>, 2020.

407 Liu, T., Chen, G., Chen, J., Xu, L., Li, M., Hong, Y., Chen, Y., Ji, X., Yang, C., Chen, Y., Huang, W., Huang,  
 408 Q., and Wang, H.: Seasonal characteristics of atmospheric peroxyacetyl nitrate (PAN) in a coastal city of  
 409 Southeast China: Explanatory factors and photochemical effects, *Atmos. Chem. Phys.*, 22, 4339-4353,  
 410 <https://doi.org/10.5194/acp-22-4339-2022>, 2022a.

411 Liu, T., Hong, Y., Li, M., Xu, L., Chen, J., Bian, Y., Yang, C., Dan, Y., Zhang, Y., Xue, L., Zhao, M., Huang,  
 412 Z., and Wang, H.: Atmospheric oxidation capacity and ozone pollution mechanism in a coastal city of  
 413 southeastern China: analysis of a typical photochemical episode by an observation-based model, *Atmos.*

Chem. Phys., 22, 2173-2190, <https://doi.org/10.5194/acp-22-2173-2022>, 2022b.

Liu, X., Qu, H., Huey, L. G., Wang, Y., Sjostedt, S., Zeng, L., Lu, K., Wu, Y., Hu, M., Shao, M., Zhu, T., and Zhang, Y.: High Levels of Daytime Molecular Chlorine and Nitryl Chloride at a Rural Site on the North China Plain, Environ. Sci. Technol., 51, 9588-9595, <https://doi.org/10.1021/acs.est.7b03039>, 2017.

Lu, K., Zhou, H., Lee, J., Nelson, B., and Zhang, Y.: Ozone mitigations beyond the control of nitrogen oxides and volatile organic compounds, Sci Bull, 68, 1989-1992, <https://doi.org/10.1016/j.scib.2023.07.051>, 2023.

Ma, W., Chen, X., Xia, M., Liu, Y., Wang, Y., Zhang, Y., Zheng, F., Zhan, J., Hua, C., and Wang, Z.: Reactive Chlorine Species Advancing the Atmospheric Oxidation Capacities of Inland Urban Environments, Environ. Sci. Technol., 57, 14638-14647, <https://doi.org/10.1021/acs.est.3c05169>, 2023.

McDuffie, E. E., Fibiger, D. L., Dubé, W. P., Lopez Hilfiker, F., Lee, B. H., Jaeglé, L., Guo, H., Weber, R. J., Reeves, J. M., Weinheimer, A. J., Schroder, J. C., Campuzano-Jost, P., Jimenez, J. L., Dibb, J. E., Veres, P., Ebben, C., Sparks, T. L., Wooldridge, P. J., Cohen, R. C., Campos, T., Hall, S. R., Ullmann, K., Roberts, J. M., Thornton, J. A., and Brown, S. S.: ClNO<sub>2</sub> Yields From Aircraft Measurements During the 2015 WINTER Campaign and Critical Evaluation of the Current Parameterization, J. Geophys. Res. Atmos., 123, 12,994-913,015, <https://doi.org/10.1029/2018JD029358>, 2018a.

McDuffie, E. E., Fibiger, D. L., Dubé, W. P., Lopez-Hilfiker, F., Lee, B. H., Thornton, J. A., Shah, V., Jaeglé, L., Guo, H., Weber, R. J., Michael Reeves, J., Weinheimer, A. J., Schroder, J. C., Campuzano-Jost, P., Jimenez, J. L., Dibb, J. E., Veres, P., Ebben, C., Sparks, T. L., Wooldridge, P. J., Cohen, R. C., Hornbrook, R. S., Apel, E. C., Campos, T., Hall, S. R., Ullmann, K., and Brown, S. S.: Heterogeneous N<sub>2</sub>O<sub>5</sub> Uptake During Winter: Aircraft Measurements During the 2015 WINTER Campaign and Critical Evaluation of Current Parameterizations, J. Geophys. Res. Atmos., 123, 4345-4372, <https://doi.org/10.1002/2018JD028336>, 2018b.

Mentel, T. F., Sohn, M., and Wahner, A. J. P. C. C. P.: Nitrate effect in the heterogeneous hydrolysis of dinitrogen pentoxide on aqueous aerosols, Phys. Chem. Chem. Phys., 1, 5451-5457, <https://doi.org/10.1039/A905338g>, 1999.

Mielke, L. H., Furgeson, A., and Osthoff, H. D.: Observation of ClNO<sub>2</sub> in a Mid-Continental Urban Environment, Environ. Sci. Technol., 45, 8889-8896, <https://doi.org/10.1021/es201955u>, 2011.

Mielke, L. H., Stutz, J., Tsai, C., Hurlock, S. C., Roberts, J. M., Veres, P. R., Froyd, K. D., Hayes, P. L., Cubison, M. J., Jimenez, J. L., Washenfelder, R. A., Young, C. J., Gilman, J. B., Gouw, J. A., Flynn, J. H., Grossberg, N., Lefer, B. L., Liu, J., Weber, R. J., and Osthoff, H. D.: Heterogeneous formation of nitryl

chloride and its role as a nocturnal NO<sub>x</sub> reservoir species during CalNex-LA 2010, *J. Geophys. Res. Atmos.*, 118, 10,638-610,652, <https://doi.org/10.1002/jgrd.50783>, 2013.

Niu, Y.-B., Zhu, B., He, L.-Y., Wang, Z., Lin, X.-Y., Tang, M.-X., and Huang, X.-F.: Fast Nocturnal Heterogeneous Chemistry in a Coastal Background Atmosphere and Its Implications for Daytime Photochemistry, *J. Geophys. Res. Atmos.*, 127, e2022JD036716, <https://doi.org/10.1029/2022JD036716>, 2022.

Osthoff, H. D., Roberts, J. M., Ravishankara, A. R., Williams, E. J., Lerner, B. M., Sommariva, R., Bates, T. S., Coffman, D., Quinn, P. K., Dibb, J. E., Stark, H., Burkholder, J. B., Talukdar, R. K., Meagher, J., Fehsenfeld, F. C., and Brown, S. S.: High levels of nitryl chloride in the polluted subtropical marine boundary layer, *Nat. Geosci.*, 1, 324-328, <https://doi.org/10.1038/ngeo177>, 2008.

Peng, X., Wang, T., Wang, W., Ravishankara, A., George, C., Xia, M., Cai, M., Li, Q., Salvador, C. M., and Lau, C.: Photodissociation of particulate nitrate as a source of daytime tropospheric Cl<sub>2</sub>, *Nat. Commun.*, 13, 1-10, <https://doi.org/10.1038/s41467-022-28383-9>, 2022.

Peng, X., Wang, W., Xia, M., Chen, H., Ravishankara, A. R., Li, Q., Saiz-Lopez, A., Liu, P., Zhang, F., Zhang, C., Xue, L., Wang, X., George, C., Wang, J., Mu, Y., Chen, J., and Wang, T.: An unexpected large continental source of reactive bromine and chlorine with significant impact on wintertime air quality, *Natl. Sci. Rev.*, 8, nwaa304, <https://doi.org/10.1093/nsr/nwaa304>, 2021.

Phillips, G. J., Tang, M. J., Thieser, J., Brickwedde, B., Schuster, G., Bohn, B., Lelieveld, J., and Crowley, J. N.: Significant concentrations of nitryl chloride observed in rural continental Europe associated with the influence of sea salt chloride and anthropogenic emissions, *Geophys. Res. Lett.*, 39, L10811, <https://doi.org/10.1029/2012gl051912>, 2012.

Riedel, T. P., Bertram, T. H., Crisp, T. A., Williams, E. J., Lerner, B. M., Vlasenko, A., Li, S. M., Gilman, J., de Gouw, J., Bon, D. M., Wagner, N. L., Brown, S. S., and Thornton, J. A.: Nitryl chloride and molecular chlorine in the coastal marine boundary layer, *Environ. Sci. Technol.*, 46, 10463-10470, <https://doi.org/10.1021/es204632r>, 2012.

Rutherford, J. A., Koehl, W. J., Benson, J. D., Burns, V. R., Hochhauser, A. M., Knepper, J. C., Leppard, W. R., Painter, L. J., Rapp, L. A., and Rippon, B.: Effects of Gasoline Properties on Emissions of Current and Future Vehicles-T50, T90, and Sulfur Effects-Auto/Oil Air Quality Improvement Research Program, SAE Technical Paper, 0148-7191, <https://doi.org/10.4271/952510>, 1995.

Sommariva, R., Hollis, L. D. J., Sherwen, T., Baker, A. R., Ball, S. M., Bandy, B. J., Bell, T. G., Chowdhury,

474 M. N., Cordell, R. L., Evans, M. J., Lee, J. D., Reed, C., Reeves, C. E., Roberts, J. M., Yang, M., and Monks,  
 475 P. S.: Seasonal and geographical variability of nitryl chloride and its precursors in Northern Europe, *Atmos.*  
 476 *Sci. Lett.*, 19, e844, <https://doi.org/10.1002/asl.844>, 2018.

477 Tan, Z., Fuchs, H., Hofzumahaus, A., Bloss, W. J., Bohn, B., Cho, C., Hohaus, T., Holland, F., Lakshmisha,  
 478 C., Liu, L., Monks, P. S., Novelli, A., Niether, D., Rohrer, F., Tillmann, R., Valkenburg, T. S. E., Vardhan, V.,  
 479 Kiendler-Scharr, A., Wahner, A., and Sommariva, R.: Seasonal variation in nitryl chloride and its relation to  
 480 gas-phase precursors during the JULIAC campaign in Germany, *Atmos. Chem. Phys.*, 22, 13137-13152,  
 481 <https://doi.org/10.5194/acp-22-13137-2022>, 2022.

482 Thaler, R. D., Mielke, L. H., and Osthoff, H. D.: Quantification of nitryl chloride at part per trillion mixing  
 483 ratios by thermal dissociation cavity ring-down spectroscopy, *Anal. Chem.*, 83, 2761-2766,  
 484 <https://doi.org/10.1021/ac200055z>, 2011.

485 Tham, Y. J., Wang, Z., Li, Q., Wang, W., Wang, X., Lu, K., Ma, N., Yan, C., Kecorius, S., Wiedensohler, A.,  
 486 Zhang, Y., and Wang, T.: Heterogeneous N<sub>2</sub>O<sub>5</sub> uptake coefficient and production yield of ClNO<sub>2</sub> in polluted  
 487 northern China: roles of aerosol water content and chemical composition, *Atmos. Chem. Phys.*, 18, 13155-  
 488 13171, <https://doi.org/10.5194/acp-18-13155-2018>, 2018.

489 Tham, Y. J., Wang, Z., Li, Q., Yun, H., Wang, W., Wang, X., Xue, L., Lu, K., Ma, N., Bohn, B., Li, X.,  
 490 Kecorius, S., Größ, J., Shao, M., Wiedensohler, A., Zhang, Y., and Wang, T.: Significant concentrations of  
 491 nitryl chloride sustained in the morning: investigations of the causes and impacts on ozone production in a  
 492 polluted region of northern China, *Atmos. Chem. Phys.*, 16, 14959-14977, [https://doi.org/10.5194/acp-16-](https://doi.org/10.5194/acp-16-14959-2016)  
 493 [14959-2016](https://doi.org/10.5194/acp-16-14959-2016), 2016.

494 Thornton, J. A., Braban, C. F., and Abbatt, J. P.: N<sub>2</sub>O<sub>5</sub> hydrolysis on sub-micron organic aerosols: the effect  
 495 of relative humidity, particle phase, and particle size, *Phys. Chem. Chem. Phys.*, 5, 4593-4603,  
 496 <https://doi.org/10.1039/B307498F>, 2003.

497 Thornton, J. A., Kercher, J. P., Riedel, T. P., Wagner, N. L., Cozic, J., Holloway, J. S., Dubé, W. P., Wolfe, G.  
 498 M., Quinn, P. K., Middlebrook, A. M., Alexander, B., and Brown, S. S.: A large atomic chlorine source  
 499 inferred from mid-continental reactive nitrogen chemistry, *Nature.*, 464, 271-274,  
 500 <https://doi.org/10.1038/nature08905>, 2010.

501 Wahner, A., Mentel, T. F., Sohn, M., and Stier, J.: Heterogeneous reaction of N<sub>2</sub>O<sub>5</sub> on sodium nitrate aerosol,  
 502 *J. Geophys. Res. Atmos.*, 103, 31103-31112, <https://doi.org/10.1029/1998JD100022>, 1998.

503 Wang, H., Yuan, B., Zheng, E., Zhang, X., Wang, J., Lu, K., Ye, C., Yang, L., Huang, S., and Hu, W.:



504 Formation and impacts of nitryl chloride in Pearl River Delta, *Atmos. Chem. Phys.*, 22, 14837-14858,  
505 <https://doi.org/10.5194/acp-22-14837-2022>, 2022.

506 Wang, H., Wang, H., Lu, X., Lu, K., Zhang, L., Tham, Y. J., Shi, Z., Aikin, K., Fan, S., Brown, S. S., and  
507 Zhang, Y.: Increased night-time oxidation over China despite widespread decrease across the globe, *Nat.*  
508 *Geosci.*, 16, 217-223, <https://doi.org/10.1038/s41561-022-01122-x>, 2023.

509 Wang, H., Chen, X., Lu, K., Tan, Z., Ma, X., Wu, Z., Li, X., Liu, Y., Shang, D., Wu, Y., Zeng, L., Hu, M.,  
510 Schmitt, S., Kiendler-Scharr, A., Wahner, A., and Zhang, Y.: Wintertime N<sub>2</sub>O<sub>5</sub> uptake coefficients over the  
511 North China Plain, *Sci. Bull.*, 65, 765-774, <https://doi.org/10.1016/j.scib.2020.02.006>, 2020.

512 Wang, T., Tham, Y. J., Xue, L., Li, Q., Zha, Q., Wang, Z., Poon, S. C. N., Dubé, W. P., Blake, D. R., Louie, P.  
513 K. K., Luk, C. W. Y., Tsui, W., and Brown, S. S.: Observations of nitryl chloride and modeling its source and  
514 effect on ozone in the planetary boundary layer of southern China, *J. Geophys. Res. Atmos.*, 121, 2476-2489,  
515 <https://doi.org/10.1002/2015JD024556>, 2016.

516 Xia, M., Peng, X., Wang, W., Yu, C., Wang, Z., Tham, Y. J., Chen, J., Chen, H., Mu, Y., and Zhang, C.:  
517 Winter ClNO<sub>2</sub> formation in the region of fresh anthropogenic emissions: seasonal variability and insights  
518 into daytime peaks in northern China, *Atmos. Chem. Phys.*, 21, 15985-16000, [https://doi.org/10.5194/acp-](https://doi.org/10.5194/acp-21-15985-2021)  
519 [21-15985-2021](https://doi.org/10.5194/acp-21-15985-2021), 2021.

520 Xia, M., Peng, X., Wang, W., Yu, C., Sun, P., Li, Y., Liu, Y. A. H. C. t. A.-P. O. A. C., Xu, Z., Wang, Z., Xu,  
521 Z., Nie, W., Ding, A., and Wang, T.: Significant production of ClNO<sub>2</sub> and possible source of Cl<sub>2</sub> from N<sub>2</sub>O<sub>5</sub>  
522 uptake at a suburban site in eastern China, *Atmos. Chem. Phys.*, 20, 6147-6158, [https://doi.org/10.5194/acp-](https://doi.org/10.5194/acp-20-6147-2020)  
523 [20-6147-2020](https://doi.org/10.5194/acp-20-6147-2020), 2020.

524 Xue, L. K., Saunders, S. M., Wang, T., Gao, R., Wang, X. F., Zhang, Q. Z., and Wang, W. X.: Development  
525 of a chlorine chemistry module for the Master Chemical Mechanism, *Geosci. Model Dev.*, 8, 3151-3162,  
526 <https://doi.org/10.5194/gmd-8-3151-2015>, 2015.

527 Yan, C., Tham, Y. J., Nie, W., Xia, M., Wang, H., Guo, Y., Ma, W., Zhan, J., Hua, C., and Li, Y.: Increasing  
528 contribution of nighttime nitrogen chemistry to wintertime haze formation in Beijing observed during  
529 COVID-19 lockdowns, *Nat. Geosci.*, 16, 975-981, <https://doi.org/10.1038/s41561-023-01285-1>, 2023.

530 Yi, X., Sarwar, G., Bian, J., Huang, L., Li, Q., Jiang, S., Liu, H., Wang, Y., Chen, H., and Wang, T.:  
531 Significant Impact of Reactive Chlorine on Complex Air Pollution Over the Yangtze River Delta Region,  
532 China, *J. Geophys. Res. Atmos.*, 128, e2023JD038898, <https://doi.org/10.1029/2023JD038898>, 2023.

533 Yun, H., Wang, T., Wang, W., Tham, Y. J., Li, Q., Wang, Z., and Poon, S. C. N.: Nighttime NO<sub>x</sub> loss and



ClNO<sub>2</sub> formation in the residual layer of a polluted region: Insights from field measurements and an iterative box model, *Sci. Total Environ.*, 622-623, 727-734, <https://doi.org/10.1016/j.scitotenv.2017.11.352>, 2018.

Zang, H., Zhao, Y., Huo, J., Zhao, Q., Fu, Q., Duan, Y., Shao, J., Huang, C., An, J., Xue, L., Li, Z., Li, C., and Xiao, H.: High atmospheric oxidation capacity drives wintertime nitrate pollution in the eastern Yangtze River Delta of China, *Atmos. Chem. Phys.*, 22, 4355-4374, <https://doi.org/10.5194/acp-22-4355-2022>, 2022.

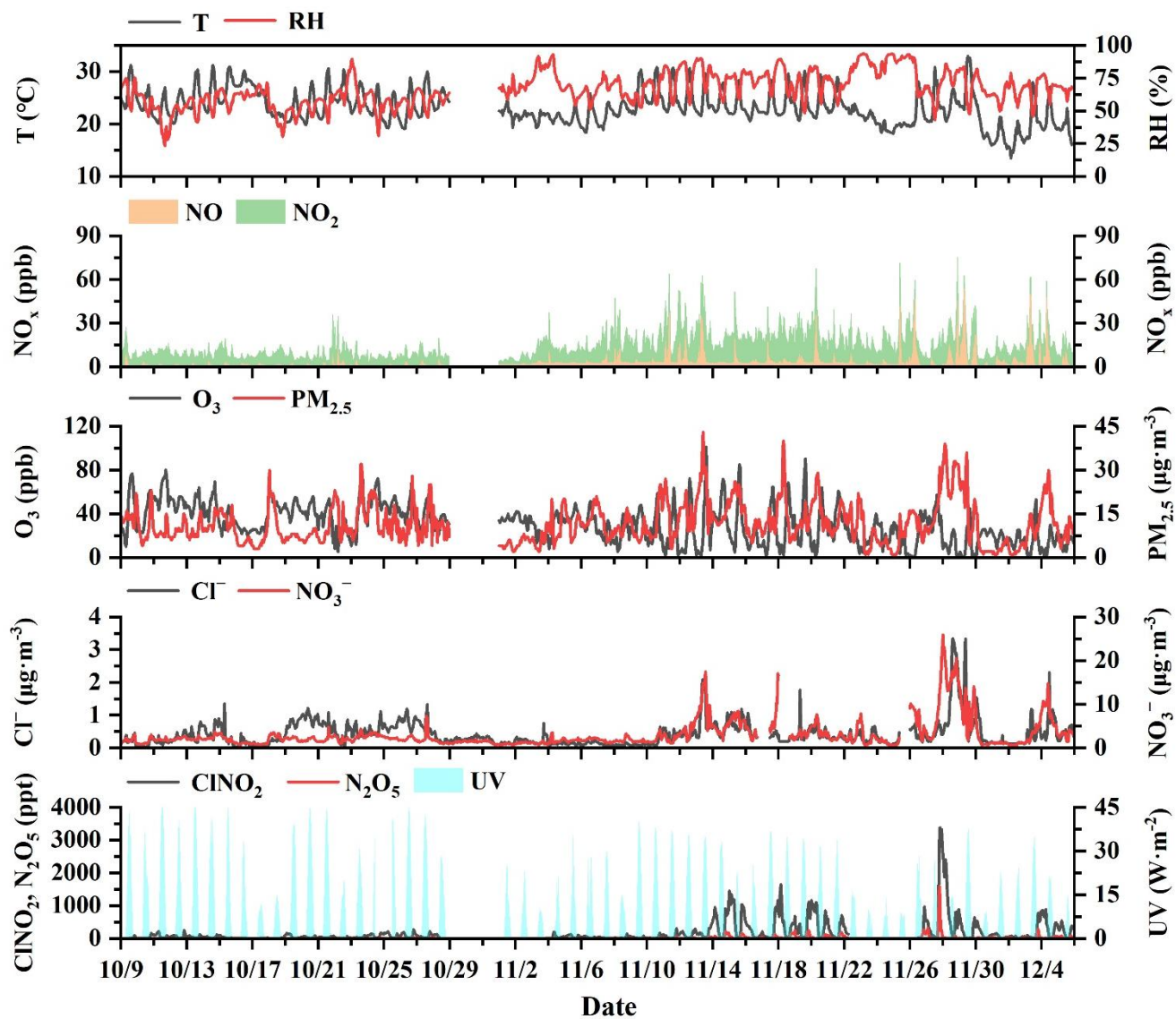


Figure 1. The time series of ClNO<sub>2</sub>, related precursors, and meteorological parameters during the autumn observation period.

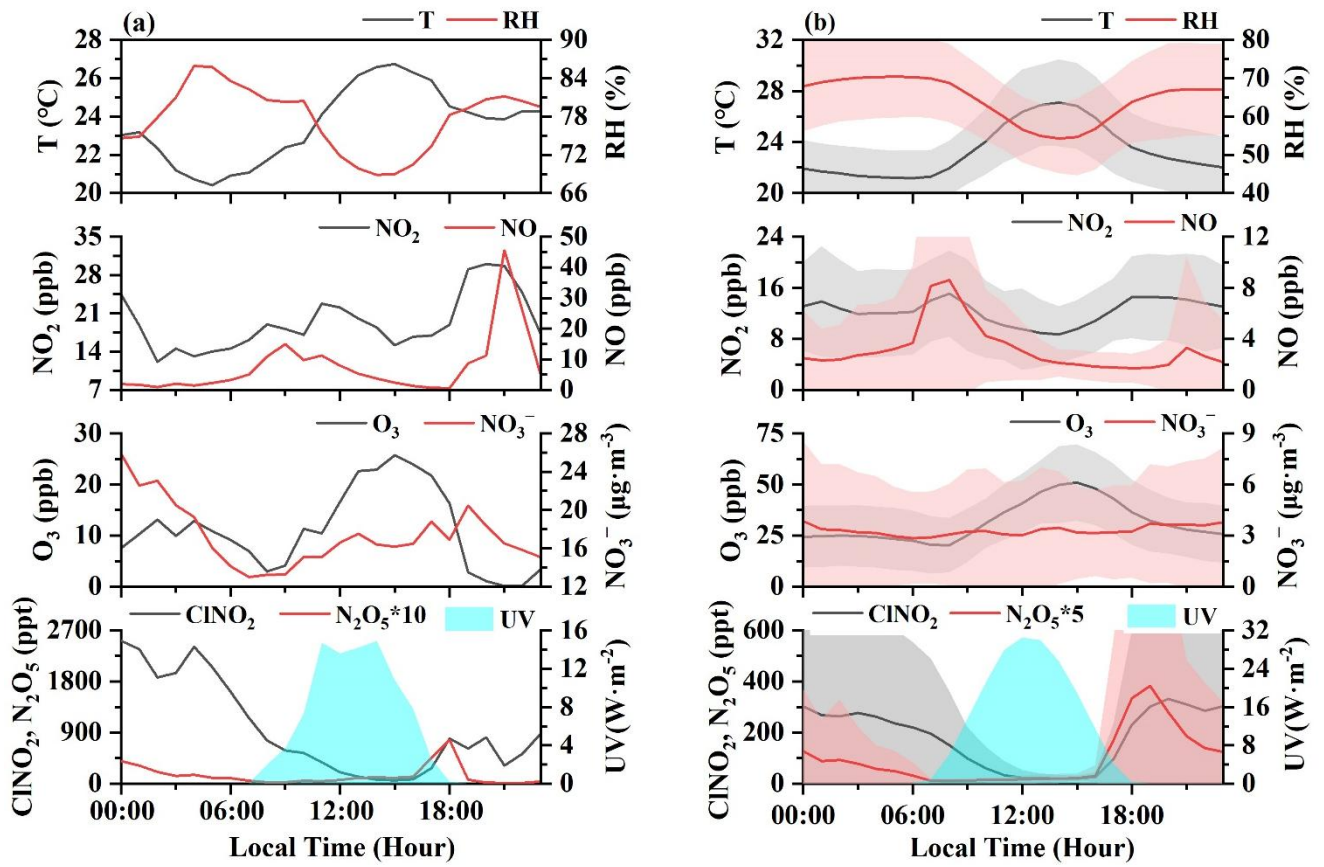


Figure 2. Diurnal variations of ClNO<sub>2</sub> and other related parameters for the highest concentrations of ClNO<sub>2</sub> (case) on November 28th (a) and the observation-average condition (from 9 October to 5 December) (b).

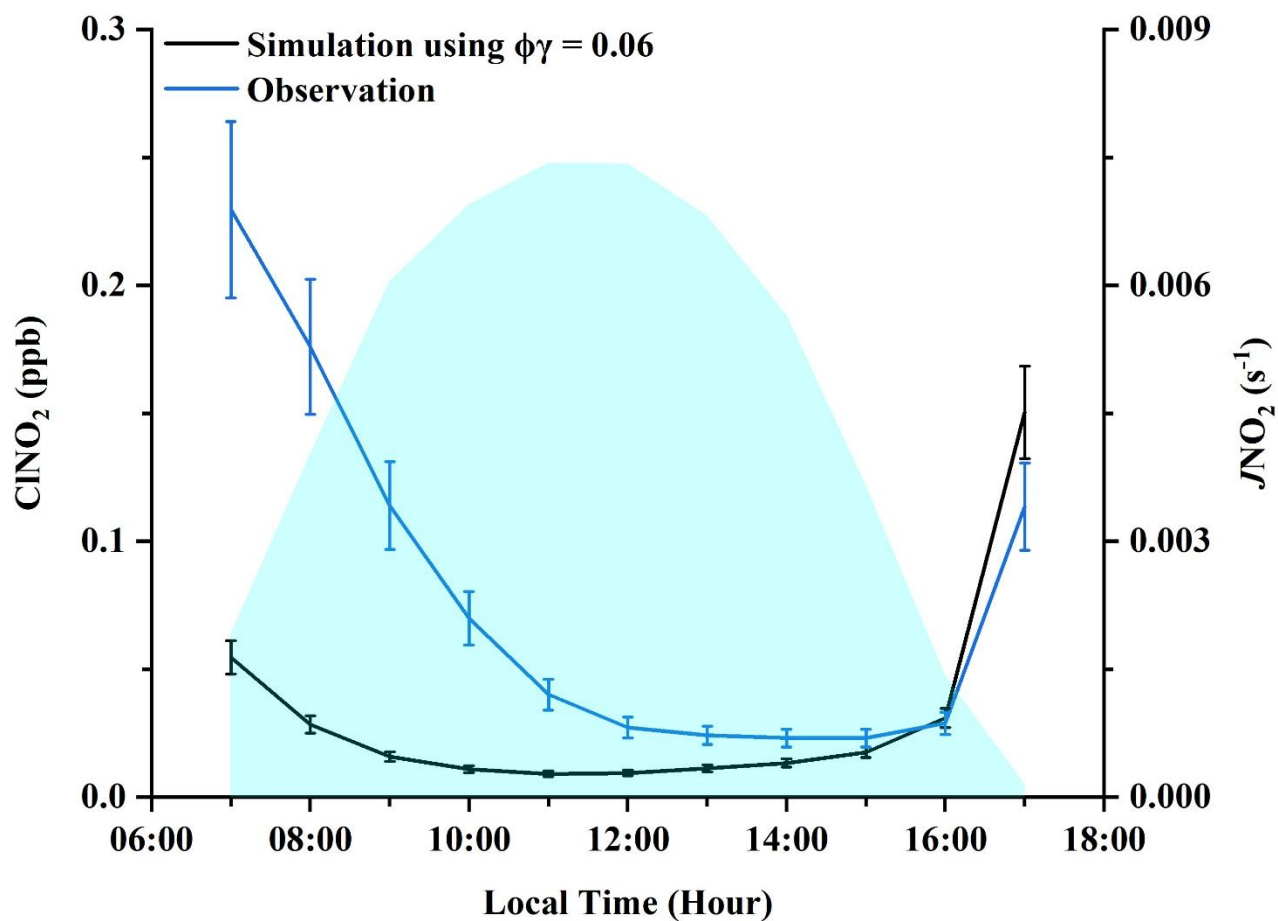
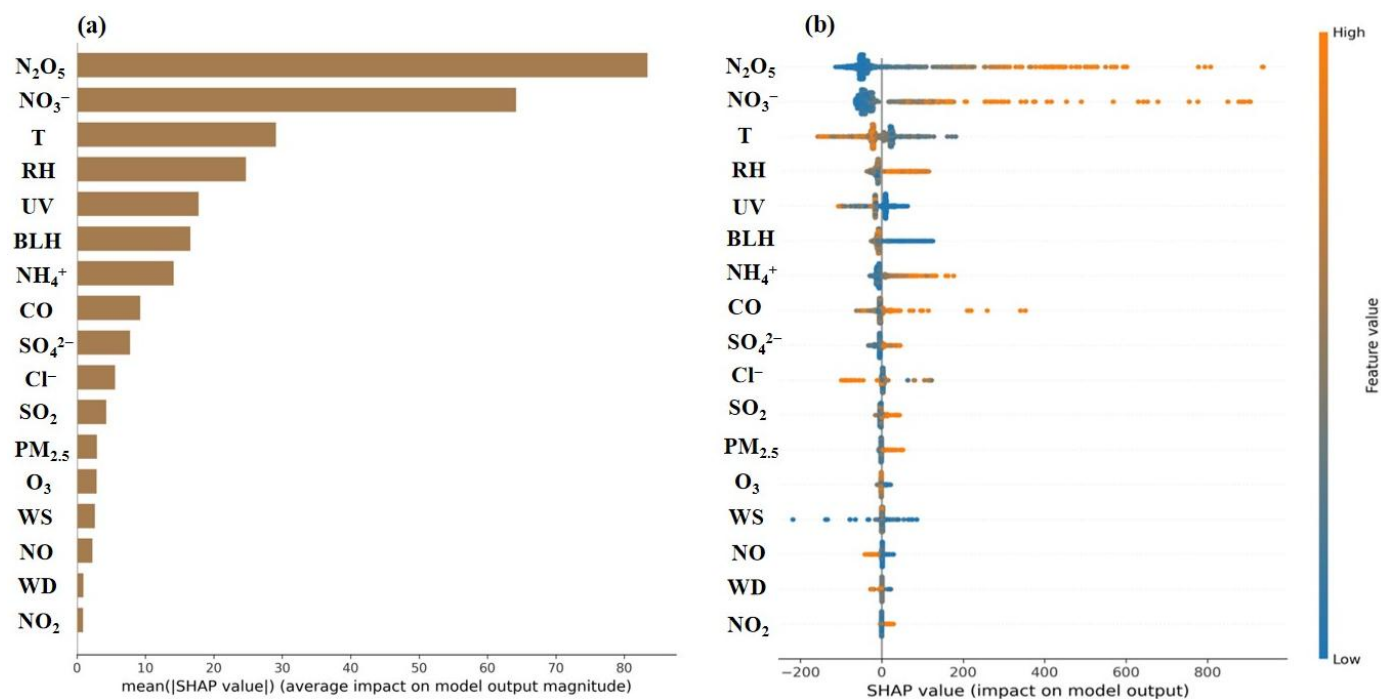


Figure 3. Comparisons of daytime  $\text{ClNO}_2$  levels between observation and simulation with a  $\phi(\text{ClNO}_2)$  of 1.0 and a  $\gamma(\text{N}_2\text{O}_5)$  of 0.06 ( $\phi\gamma = 0.06$ ).



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590 Figure 4. Relative importance of each feature to  $\text{ClNO}_2$  using XGBoost-SHAP during the autumn  
 591 observation period. The mean absolute SHAP value (a), summary plot of SHAP values of each feature (b).

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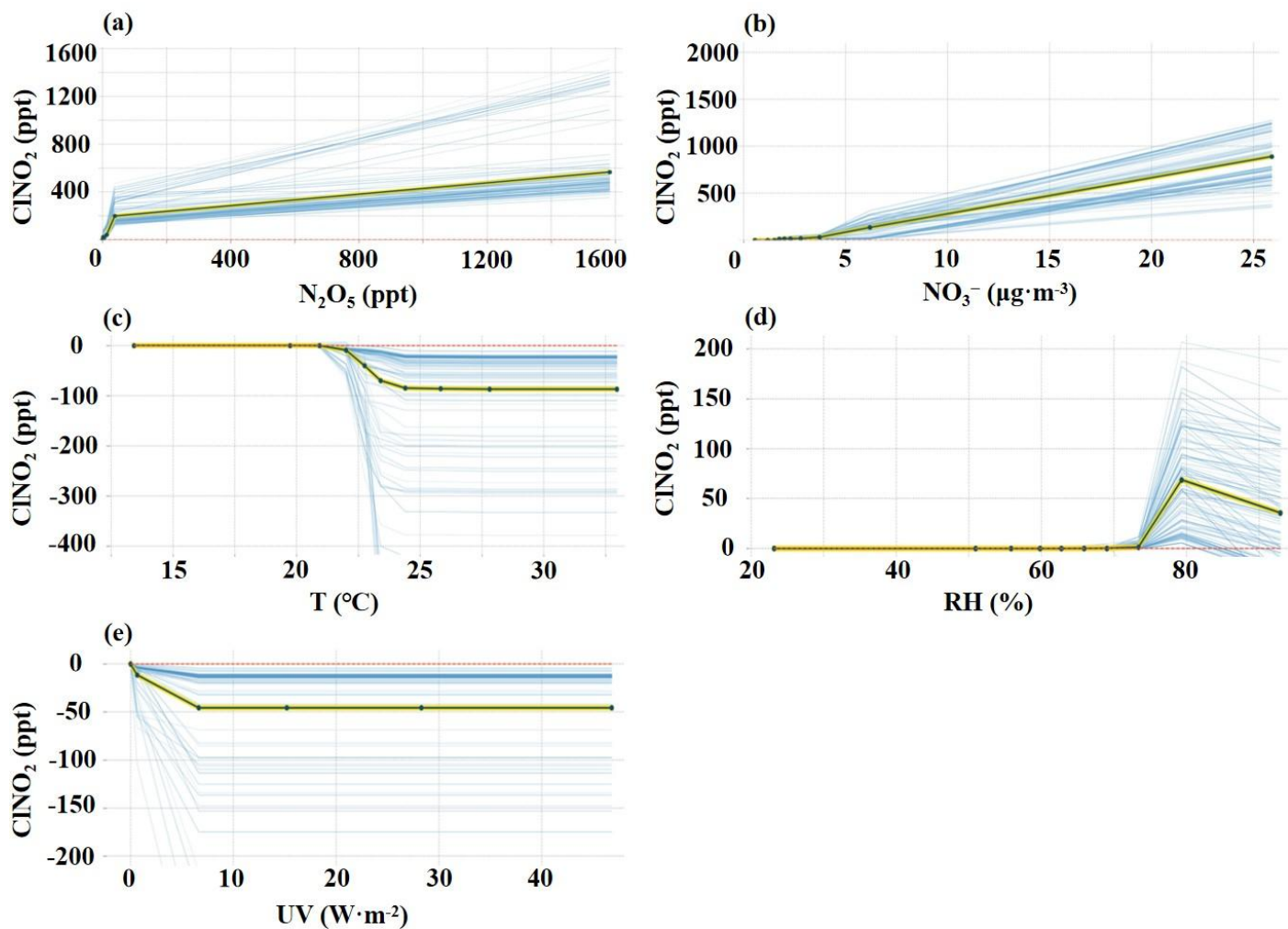


Figure 5. Isolation plots of PDP for  $\text{N}_2\text{O}_5$  (a),  $\text{NO}_3^-$  (b),  $T$  (c),  $\text{RH}$  (d), and  $\text{UV}$  (e). The average variations of simulated  $\text{ClNO}_2$  with factors' changes spline are indicated by the yellow and black curve, and blue curves presents all situations during the whole observation period.

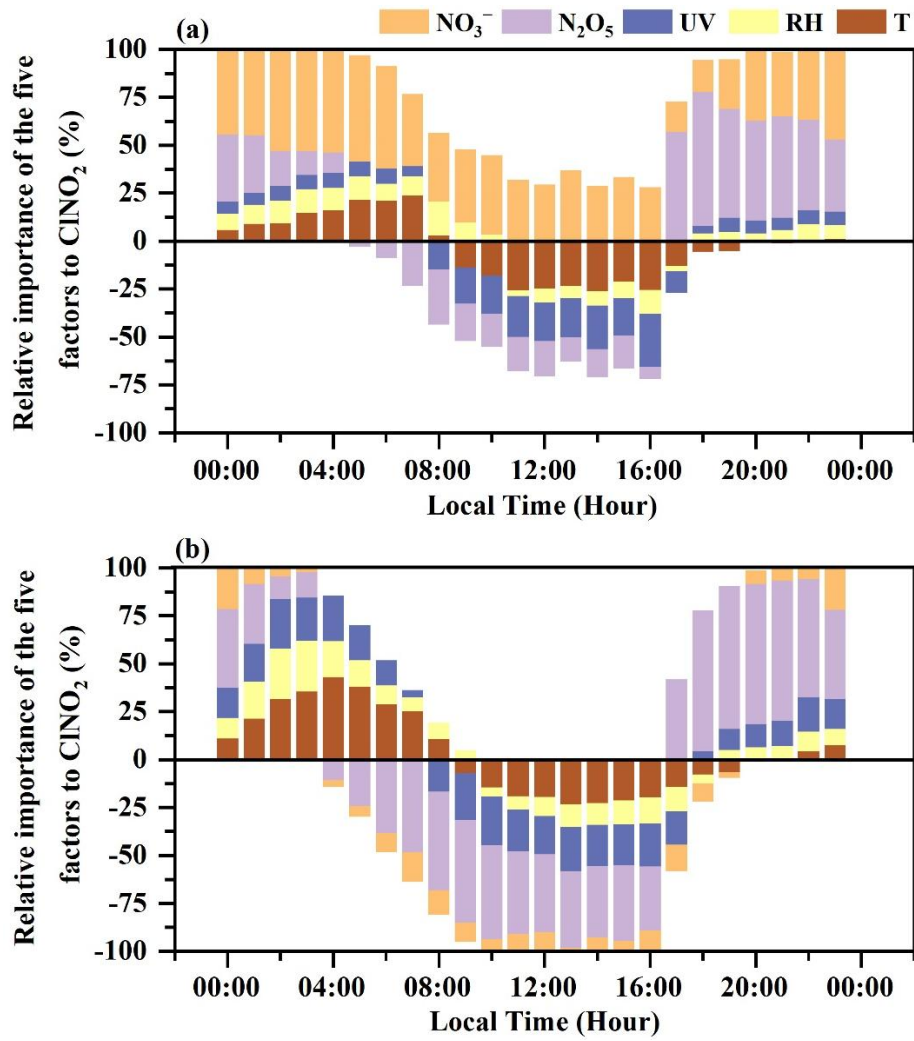


Figure 6. The diurnal variations of the relative importance of factors to  $\text{ClNO}_2$  based on the SHAP values under the high ( $> 3.7 \mu\text{g}\cdot\text{m}^{-3}$ ) (a) and low ( $< 3.7 \mu\text{g}\cdot\text{m}^{-3}$ ) (b)  $\text{ClNO}_2$  concentrations.



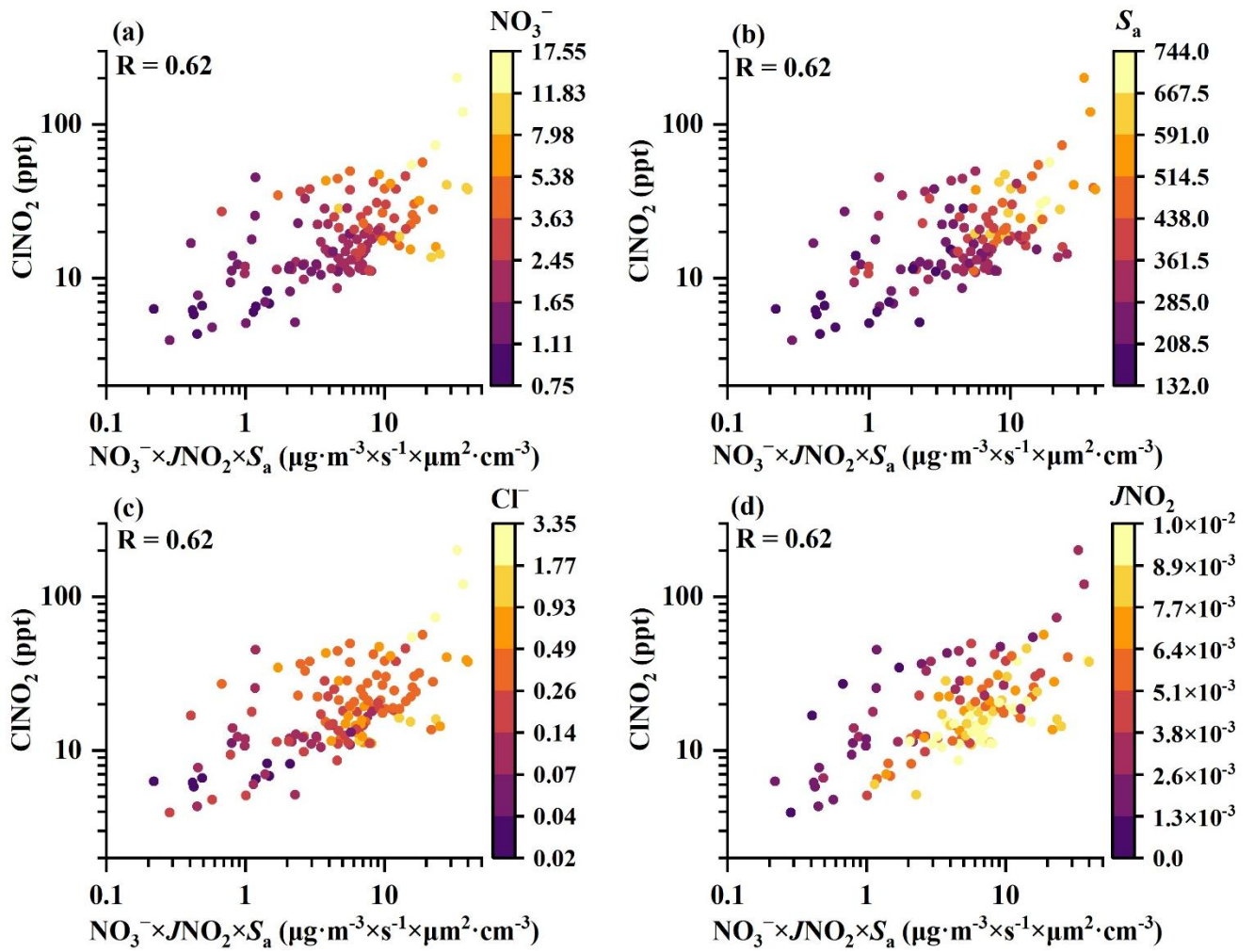


Figure 7. The relationship of daytime  $\text{ClNO}_2$  concentrations (12:00-15:00 Local Time) and a proxy of nitrate ( $\text{NO}_3^-$ ) photolysis ( $\text{NO}_3^- \times J\text{NO}_2 \times S_a$ ). The color of the dots respects the  $\text{NO}_3^-$  (a),  $S_a$  (b),  $\text{Cl}^-$  (c),  $J\text{NO}_2$  (d), respectively.



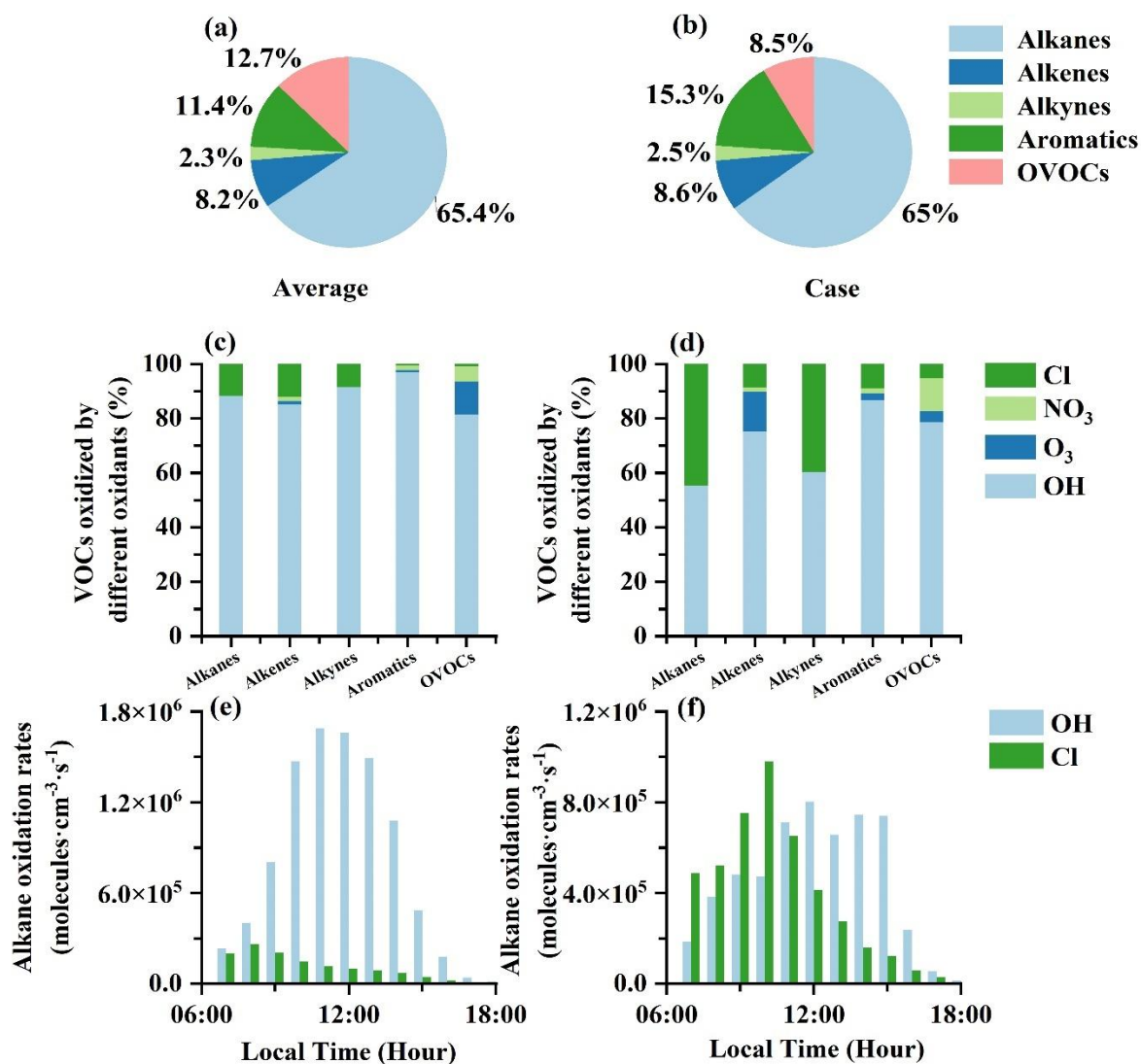


Figure 8. The impacts of Cl radicals released by ClNO<sub>2</sub> photolysis and other atmospheric oxidants (including OH, NO<sub>3</sub>, and O<sub>3</sub>) on VOC oxidation under the observation-average condition and high ClNO<sub>2</sub> case, respectively. The contributions of different VOC groups oxidized by Cl radical during the observation-average (a). The contributions of different VOC groups oxidized by Cl radical during the case (b). The contributions of different atmospheric oxidants (including OH, Cl, NO<sub>3</sub>, and O<sub>3</sub>) to VOC groups during the observation-average (c). The contributions of different atmospheric oxidants (including OH, Cl, NO<sub>3</sub>, and O<sub>3</sub>) to VOC groups during the case (d). Comparisons of alkane oxidation rates (molecules·cm<sup>-3</sup>·s<sup>-1</sup>) by OH and Cl radical during the observation-average (e). Comparisons of alkane oxidation rates by OH and Cl radical (molecules·cm<sup>-3</sup>·s<sup>-1</sup>) during the case (f).

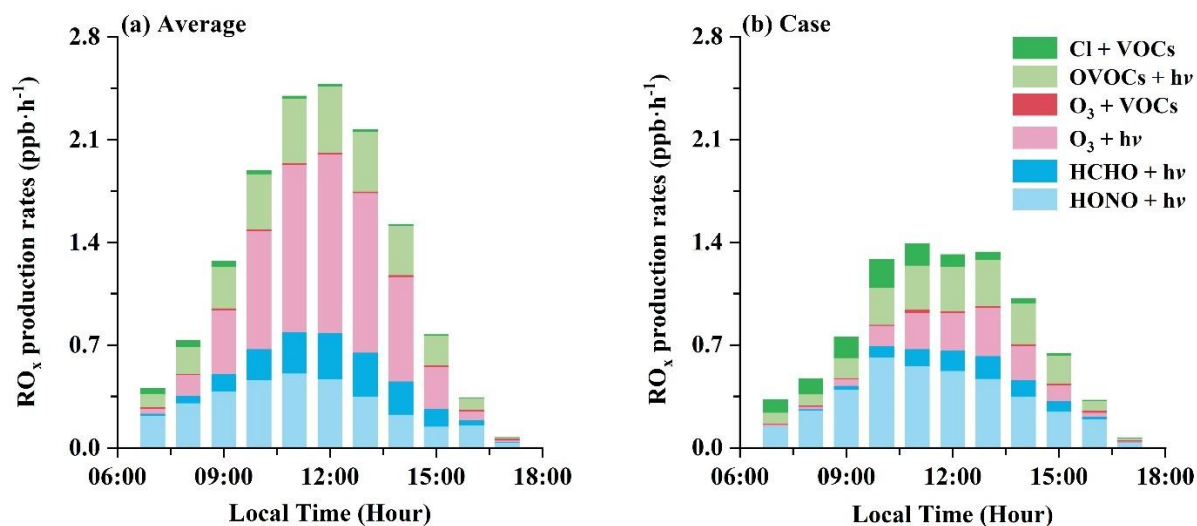


Figure 9. The contributions of different production pathways to RO<sub>x</sub> production rates under the observation-average condition (a) and high ClNO<sub>2</sub> case (b), respectively.

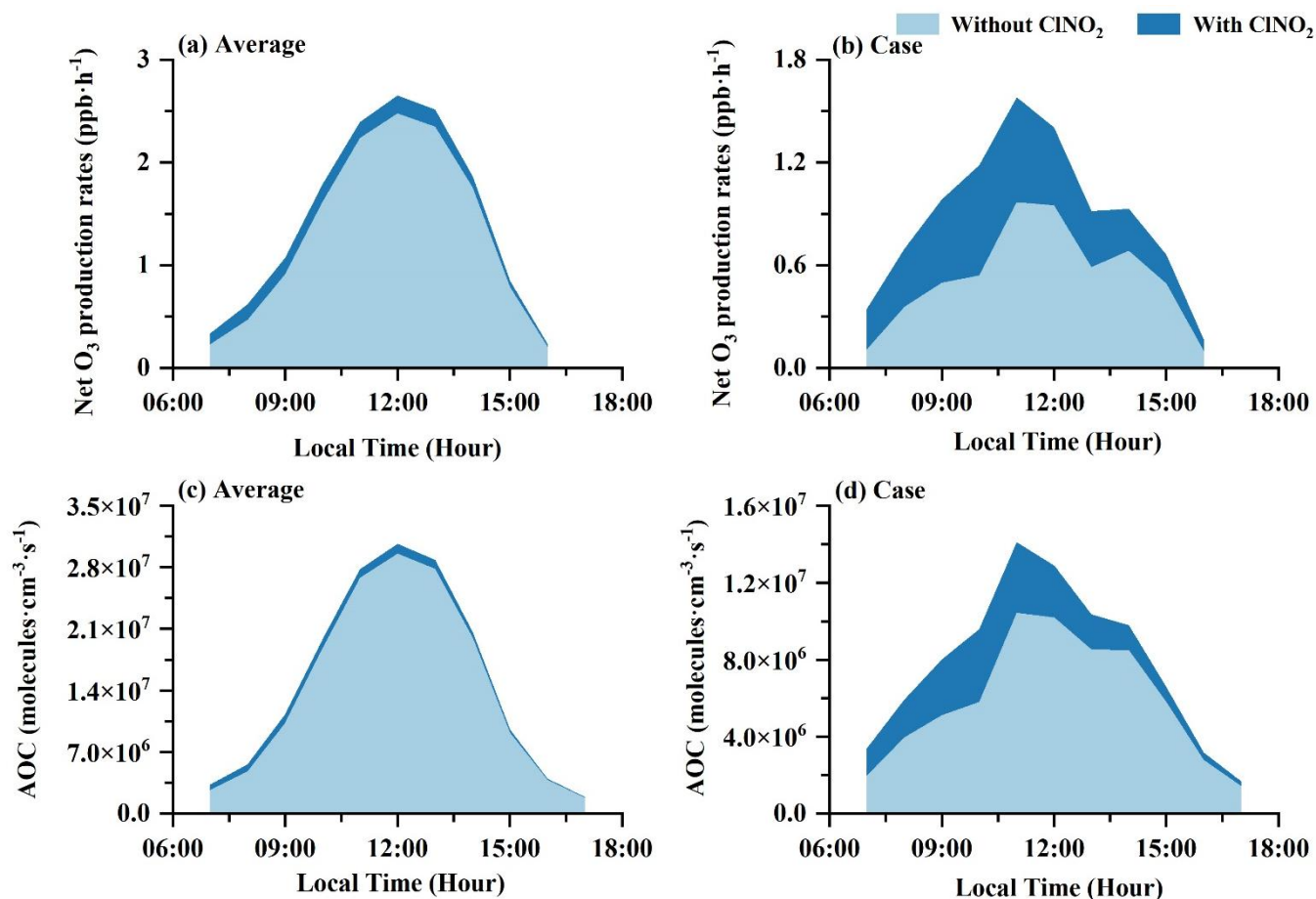


Figure 10. The impacts of Cl radicals released by ClNO<sub>2</sub> photolysis on net O<sub>3</sub> production rates and the AOC levels under the observation-average condition (a, c) and high ClNO<sub>2</sub> case (b, d), respectively.

Table 1. Relative importance of Cl, OH, and O<sub>3</sub> to the daytime oxidation of VOC groups (including alkanes, alkenes, alkynes, aromatics, and OVOCs) around the world (Xue et al., 2015; Bannan et al., 2015; Bannan et al., 2017; Rutherford et al., 1995; Fraser et al., 1997).

	Xiamen (average)	Xiamen (case)	Hong Kong (max)	London (average)	Weybourne (average)	Boston	LA
Alkane Cl%	11.7	44.8	53.0	3.5	1.0	8.5	9.9
Alkane OH%	88.3	55.2	47.0	96.5	99.0	91.5	90.1
Alkane O <sub>3</sub> %	-	-	-	-	-	-	-
Alkene Cl%	12.2	8.7	14.0	0.6	0.4	0.3	0.3
Alkene OH%	85.0	75.2	81.0	77.9	78.3	33	31.3
Alkene O <sub>3</sub> %	1.2	14.7	5.0	21.5	21.4	66.7	68.4
Alkyne Cl%	8.5	40.0	-	7.0	2.6	8.7	8.7
Alkyne OH%	91.5	60.0	-	91.8	96.7	89.7	89.7
Alkyne O <sub>3</sub> %	-	-	-	1.2	0.7	1.6	1.6
Aromatics Cl%	0.7	9.1	11.0	-	-	-	-
Aromatics OH%	97.0	86.6	89.0	-	-	-	-
Aromatics O <sub>3</sub> %	0.7	2.6	-	-	-	-	-
OVOCs Cl%	0.9	5.2	6.0	-	-	-	-
OVOCs OH%	81.4	78.7	85.0	-	-	-	-
OVOCs O <sub>3</sub> %	12.0	3.9	-	-	-	-	-

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694 Table 2. The impacts of ClNO<sub>2</sub> photolysis on RO<sub>x</sub> (OH, HO<sub>2</sub>, and RO<sub>2</sub>) levels, P(RO<sub>x</sub>), and P(O<sub>3</sub>) around the  
 695 world (Xia et al., 2021; Wang et al., 2022; Tham et al., 2016; Wang et al., 2016; Xue et al., 2015; Bannan et  
 696 al., 2017; Jeong et al., 2019).

Study Area	Season	OH	HO <sub>2</sub>	RO <sub>2</sub>	P(RO <sub>x</sub> )	P(O <sub>3</sub> )
Xiamen (average)	Autumn	3.7%	7.1%	10.3%	4.9%	6.7%
Xiamen (case)	Autumn	17.9%	34.6%	54.3%	23.8%	41.7%
Wangdu/Beijing/Mt. Tai	Winter	15.0%–22.0%	24.0%–31.0%	36.0%–52.0%	1.3%–3.8%	1.3%–6.2%
Heshan	Autumn	1.5%–2.6%	1.9%–4.6%	3.0%–6.8%	< 2.2%	1.0%–4.9%
Wangdu	Summer	-	-	-	10%–30%	3.0%–13.0%
Mt. Tai Mo Shan, Hong Kong	Winter	40.0%–77.0%	53.0%–106.0%	-	-	11.0%–41.0%
Hok Tsui, Hong Kong	Summer	6.6%	12.2%	45.1%	-	10.3%
Weybourne	Spring	5.0%	7.0%	9.0%	-	-
Seoul	Spring	-	-	-	-	1.0%–2.0%

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