



1 **Changes in Aerosol/Gas-Phase Distribution Ratio of Semi-Volatile Products**

2 **Affect Secondary Organic Aerosol Formation with NO_x from α -Pinene**

3 **Photooxidation**

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5 **Shijie Liu^{1,2}, Xinbei Xu¹, Si Zhang¹, Rongjie Li¹, Zheng Li¹, Can Wu¹, Rui Li¹,**
6 **Guiqin Zhang², and Gehui Wang^{1,3*}**

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8 ¹Key Laboratory of Geographic Information Science of the Ministry of Education,
9 School of Geographic Sciences, East China Normal University, Shanghai 200241,
10 China

11 ²Resources and Environment Innovation Institute, Shandong Jianzhu University,
12 Jinan 250101, China

13 ³Institute of Eco-Chongming, Cuiniao Road, Chenjia Zhen, Chongming, Shanghai
14 202150, China

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23 *Correspondence to: Prof. Gehui Wang (ghwang@geo.ecnu.edu.cn)

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25 **Abstract**

26 Atmospheric α -pinene is one of the most important precursors of secondary
27 organic aerosols (SOA). The formation of α -pinene derived SOA is strongly affected
28 by NO_x. However, we still do not comprehensively understand the effects of NO_x on
29 α -pinene derived SOA formation. Therefore, we conducted α -pinene photooxidation
30 experiments in an atmospheric chamber at different NO_x concentrations. The yields of
31 α -pinene SOA increased with NO_x concentration under low-NO_x conditions, but were
32 suppressed under high-NO_x conditions. The maximum SOA yields were 8.0% and 26.2%
33 in the low- and high-volatility organic compound (VOC) experiments, respectively. We
34 found the increased SOA yields under low-NO_x conditions were related to increased
35 consumption of α -pinene. The products of α -pinene photooxidation were mainly semi-
36 volatile, and the change in the aerosol/gas-phase distribution ratio as the formation of
37 α -pinene photooxidation products increased was identified as the main reason for the
38 enhanced SOA yields with increasing NO_x. The sensitivity of the SOA yield to changes
39 in NO_x and VOCs under different experimental conditions was also analyzed. This
40 study also quantified the nitrogen-containing organic compound (NOC) concentrations.
41 The mass fraction of NOCs in SOA increased monotonically with NO_x in the α -pinene
42 photooxidation process, and the maximum NOC mass fraction made up as much as
43 two-fifths of the α -pinene SOA.

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46 **1. Introduction**

47 Fine particulate matter (PM_{2.5}) is an important atmospheric pollutant. Recently,
48 PM_{2.5} has attracted attention because of its negative impacts on human health, air
49 quality, atmospheric chemical reactions, and even climate radiation balances (Bellouin
50 et al., 2020; Wang et al., 2016). Secondary organic aerosols (SOA), which form via the
51 atmospheric photooxidation of volatile organic compounds (VOCs), are some of the
52 most significant sources of PM_{2.5} (Lv et al., 2022; Mcfiggans et al., 2019; Zhang et al.,
53 2018). However, due to the diversity of SOA production processes and influencing
54 factors, much researches are still needed for the fully understand the formation
55 mechanisms of SOA.

56 NO_x, which is mainly emitted from human activities, is a key substance that
57 facilitates photooxidation. Differences in concentrations of NO_x can lead to significant
58 changes in photooxidation processes and further affect SOA formation (Sarrafzadeh et
59 al., 2016; Nussbaumer et al., 2022). Organic peroxy radicals (RO₂) are the main
60 intermediate oxidation product when VOCs are oxidized by OH. The interaction
61 between NO_x and RO₂ plays a key role in the SOA formation process, and nonlinear
62 relationships between NO_x and SOA yields have been observed in a large number of
63 experimental investigations (Xu et al., 2021; Zhao et al., 2018; Sarrafzadeh et al., 2016;
64 Xu et al., 2014). Hydroperoxides (ROOH), which are formed from the reactions of RO₂
65 with HO₂ or RO₂, have low volatilities and are responsible for SOA formation. However,
66 as the concentration of NO_x increases, RO₂ reacts with NO and is rapidly converted to
67 alkoxy radicals (RO). RO act as an intermediate and goes on to produce more high
68 volatility compounds, which suppresses SOA formation (Sarrafzadeh et al., 2016;
69 Atkinson, 2000). Highly oxygenated organic molecules (HOMs) formed by the
70 autoxidation of RO₂ are also key compounds involved in SOA formation (Rissanen,
71 2021). HOM formation is suppressed when the NO₂ concentration increases and RO₂
72 + NO becomes the dominant sink of RO₂ in the photooxidation process, which also
73 contributes to the inhibition of SOA formation as NO_x concentrations increase.



74 Clearly, the effects of NO_x on photooxidation and SOA formation are quite
75 complicated, but they have been widely studied in controlled chamber experiments.
76 Based on the semi-volatile partitioning theory in SOA formation, it has been established
77 that SOA yield is a function of SOA mass concentration when other experimental
78 conditions are held constant (Odum et al., 1996; Takeuchi et al., 2022). Furthermore,
79 the mass concentration of formed SOA is directly related to the mass concentration of
80 available VOC precursors. However, the SOA yield is often discussed as a constant,
81 and the nonlinear relationships between SOA yield and initial NO_x concentration
82 reported in chamber studies do not account for the consumption of VOCs (Chen et al.,
83 2022b; Aruffo et al., 2022; Qi et al., 2020). To control for the impact of the reacted
84 VOC concentration on SOA yields under different NO_x conditions, Sarrafzadeh et al.
85 (2016) conducted experiments that maintained similar levels of VOC consumption.
86 However, due to differences in OH concentrations during photooxidation, there were
87 clear differences in the reaction times in each experiment. Under normal circumstances,
88 when exploring the impact of NO_x on SOA yield, SOA yields are normalized based on
89 the reacted VOCs concentration, but this does not account for differences in the
90 consumption of VOCs. This approach inevitably overlooks the impact of semi-volatile
91 partitioning theory in SOA formation under different NO_x conditions. The roles of
92 chemical processes are often considered due to the impacts of NO_x on SOA yields, but
93 physical processes in SOA formation are equally significant and should be given more
94 attention.

95 α -Pinene is one of the most abundant monoterpene VOCs in the atmosphere
96 (Sindelarova et al., 2014; Guenther et al., 2012). Due to the high concentrations and
97 high SOA formation potential, α -pinene is one of the most important sources of SOA
98 in the atmosphere (Xu et al., 2015a; Zhang et al., 2018). In fact, α -pinene derived SOA
99 is commonly utilized in atmospheric models to represent biogenic SOA (Henry et al.,
100 2012). However, further in-depth research on the mechanisms of α -pinene SOA
101 formation is required to improve the fundamental basis and accuracy of atmospheric
102 models. In order to fully examine the effects of NO_x on SOA formation, we used an



103 indoor chamber to investigate the yields of α -pinene derived SOA with different initial
104 NO_x concentrations. The relationships among NO_x concentration, VOCs consumption,
105 distribution of photooxidation products, and SOA mass yield were considered. Based
106 on the semi-volatile partitioning theory, this study aimed to better characterize the
107 mechanisms by which NO_x affect SOA yield.

108 **2. Experimental methods**

109 **2.1. Chamber studies**

110 A series of α -pinene photooxidation experiments initiated by NO_x were performed
111 in a temperature controlled photooxidation chamber. The chamber and its
112 characterization capabilities have been described in detail in our previous studies (Liu
113 et al., 2021a; Liu et al., 2021b). In brief, the photooxidation chamber was constructed
114 of Teflon-FEP film (0.06 mm) and surrounded with black light lamps (GE F40BLB) as
115 the light source for the photooxidation reaction. The black light lamps provided an
116 effective light intensity (J_{NO_2}) of 0.165 min^{-1} at full illumination. Before each
117 experiment, the chamber was cleaned by first evacuating all air and then filling it with
118 purified air. This filling-purging cycle was repeated 5 times between experiments to
119 ensure the residual particulate and α -pinene concentrations were less than 5 cm^{-3} and
120 0.5 ppb , respectively. 5 m^3 zero air, which was supplied by a zero-air generator (111-
121 D3N, Thermo Scientific™, USA), was introduced into the chamber for the
122 photooxidation experiment. Analytically-pure, liquid α -pinene (Sigma-Aldrich) was
123 injected into a Teflon tube using a micro syringe, and then the α -pinene was evaporated
124 into gas and flushed into the chamber with zero air. NO_x (Air Liquid Shanghai, 510 ppm
125 NO_2 in N_2) were introduced directly into the chamber. After all the reactants were well
126 mixed, photooxidation of α -pinene was initiated by turning the black light lamps on.
127 All the experiments were conducted in dry conditions with a relative humidity of $15 \pm$
128 5% . Both OH and O_3 were formed under black light irradiation in the presence of NO_x .
129 The detailed conditions for the α -pinene photooxidation experiments are listed in Table
130 1. The concentrations of α -pinene were kept as constant as possible across different



131 experiments to ensure the effects of NO_x were not obscured. No seed particles were
 132 used in the chamber experiments.

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134 **Table 1.** Details of the α -pinene/NO_x systems used in the photooxidation experiments.

No.	α -pinene (ppb)	$\Delta\alpha$ -pinene (ppb)	NO _x (ppb)	[VOCs] ₀ /[NO _x] ₀	SOA mass conc. ^a ($\mu\text{g m}^{-3}$)	SOA yield
Exp. 1	116.2	67.3	12	9.7	26.0	6.5%
Exp. 2	117.2	96.2	25	4.7	43.8	7.7%
Exp. 3	117.8	115.3	68	1.7	54.3	8.0%
Exp. 4	119.3	117.6	150	0.8	34.6	4.9%
Exp. 5	116.9	114.9	337	0.3	15.3	2.2%
Exp. 6	115.6	112.8	600	0.2	11.4	1.7%
Exp. 7	258.3	70.7	8	32.3	20.9	5.0%
Exp. 8	251.0	168.7	26	9.7	144.0	14.5%
Exp. 9	263.4	234.0	52	4.1	370.3	26.2%
Exp. 10	250.8	245.3	113	2.2	347.6	23.7%
Exp. 11	247.8	244.2	237	1.5	253.5	17.1%
Exp. 12	260.0	256.0	369	0.7	226.4	14.6%
Exp. 14	254.2	249.9	669	0.4	175.0	11.8%

135 ^a All mass concentrations were wall-loss corrected.

136 2.2. On-line instruments

137 The concentration of NO_x was constantly monitored with a NO/NO₂/NO_x analyzer
 138 (Model 42C, Thermo Corporation, USA). The α -pinene concentration was measured
 139 throughout the photooxidation process with a proton-transfer-reaction time-of-flight
 140 mass spectrometer (PTR-tof-MS, Ionicon Analytik, Austria) using hydronium (H₃O⁺)
 141 ions. The drift tube of the PTR-tof-MS was operated at 60.0 °C (T_{drift}), 2.30 mbar (P_{drift}),
 142 and 600V (U_{drift}). Three ions which are commonly observed in ambient air, i.e., H₃O⁺
 143 (21.0226 *m/z*), NO⁺ (29.9980 *m/z*), and C₃H₆O⁺ (59.0497 *m/z*), were used for PTR-tof-
 144 MS mass calibration.

145 The formation and chemical characteristics of bulk SOA were measured with a
 146 high-resolution time-of-flight aerosol mass spectrometer (HR-tof-AMS, Aerodyne
 147 Research Inc., USA). During which, the AMS was only operated in the V-mode in order
 148 to avoid W-mode data too noisy to analyze. The ionization efficiency (IE) of the AMS
 149 was calibrated via monodisperse with dried ammonium nitrate (AN) aerosols (300 nm).

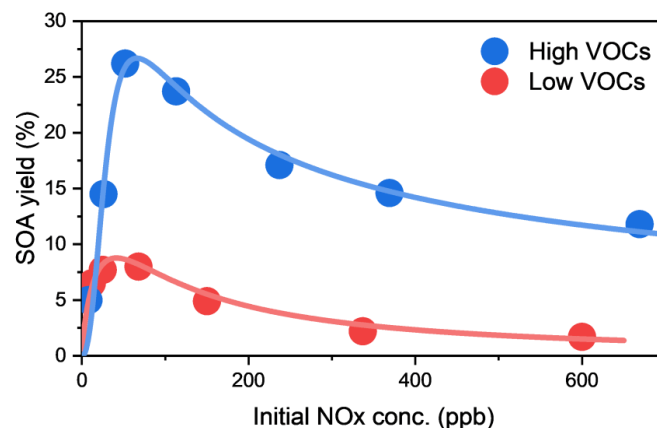


150 The SOA mass concentrations obtained from AMS measurements were compared and
151 corrected through a scanning mobility particle sizer. For SOA mass concentration
152 calculations, the relative ionization efficiencies (RIE) of 1.4 and 1.1 were applied for
153 organic compounds and nitrate, respectively (Middlebrook et al., 2012). The AMS data
154 were analyzed based on the standard applications of SQUIRREL 1.51H and PIKA
155 1.10H in Igor Pro (WaveMetrics, Inc., USA). $\text{NO}_2^+(m/z = 46)/\text{NO}^+(m/z = 30)$ was used
156 to differentiate inorganic and organic nitrate in the AMS measurements (Xu et al.,
157 2015b; Kiendler-Scharr et al., 2016; Ng et al., 2017).

158 **3. Result and discussion**

159 **3.1 Effect of NO_x on α -pinene SOA formation**

160 The relationship between SOA yields and NO_x is shown in Fig. 1. Here, SOA yield
161 was calculated as the SOA mass concentration divided by the reacted VOCs. Clearly,
162 the SOA yield increased rapidly at first, reached a maximum and then decreased
163 gradually with increasing initial NO_x concentration under both high- and low-VOC
164 conditions. In this study, we defend the positive correlation between SOA yield and
165 NO_x as low-NO_x conditions, and a negative correlation of them as high-NO_x
166 conditions. In the low- and high-VOC experiments, the maximum SOA yields were
167 8.0% and 26.2% respectively. Since the change trend in SOA yield with NO_x was not
168 affected by initial VOC concentration, the effect of NO_x on SOA formation is discussed
169 in this section using the data from the low-VOCs experiments. The concentration of
170 VOCs in the chamber is generally higher than that in the real atmospheric environment.
171 Lower VOC concentrations would be closer to their levels in the actual atmosphere,
172 which would make experimental results more representative of the real atmospheric
173 environment.



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Figure 1. SOA yield from α -pinene photooxidation with different initial NO_x concentrations under two levels of VOCs.

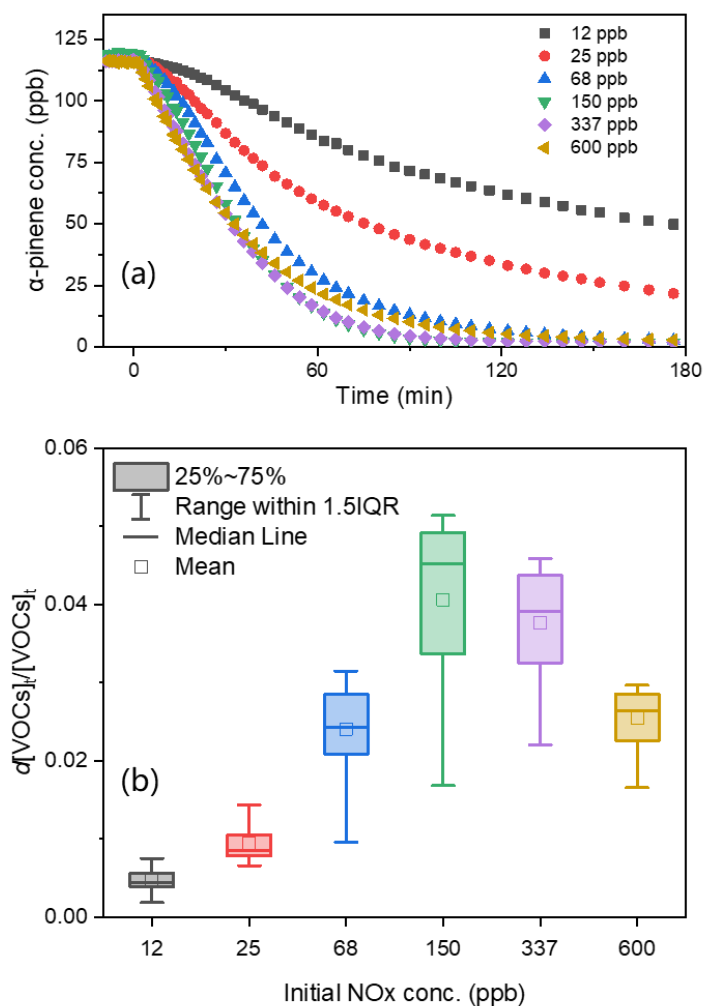
178 Like our study, similar relationships between SOA yields and initial NO_x
179 concentrations have been widely observed in previous studies (Aruffo et al., 2022; Liu
180 et al., 2019b; Zhao et al., 2018; Lane et al., 2008; Ng et al., 2007; Kroll et al., 2006).
181 RO₂ was the main intermediate in the photooxidation of VOCs and the concentration
182 and fate of RO₂ depended on the NO_x concentrations. RO₂ mainly reacted with NO and
183 was rapidly converted to alkoxy radicals (RO) under high-NO_x conditions. RO
184 intermediates produce more high-volatility compounds than hydroperoxides (ROOH),
185 which form through the reaction of RO₂ with RO₂/HO₂, and decrease SOA yields
186 (Atkinson, 2000; Sarrafzadeh et al., 2016). Furthermore, the autooxidation of RO₂ can
187 be inhibited through the RO₂ + NO / NO₂ reaction, and the reduction in HOMs further
188 contributes to the decreased SOA yields under high-NO_x conditions (Yu et al., 2022;
189 Laskin et al., 2018).

190 However, the mechanisms by which SOA yields increase with increasing NO_x
191 concentrations under low-NO_x conditions remain poorly understood (Camredon et al.,
192 2007; Kroll et al., 2006). The atmospheric oxidizing capacity (AOC), which indicates
193 the oxidizing ability of the atmosphere, is significantly influenced by NO_x (Wang et al.,
194 2023). In this study, we assessed AOC based on the decay ratio of VOCs ($AOC_t = -d$



195 [VOCs]_t / [VOCs]_i). The time-dependent consumption curves of VOC and the decay
196 ratios of VOCs under different NO_x concentrations are shown in Fig. 2. All experiments
197 were allowed the same photooxidation time. For the low-NO_x experiments, the
198 consumption of α -pinene increased with NO_x concentration. Only when the initial NO_x
199 concentration was higher than 68 ppb was α -pinene completely consumed by the end
200 of the photooxidation period. The average decay ratio of VOCs increased from 4.75×10^{-3}
201 to 4.53×10^{-2} as the initial NO_x increased from 12 ppb to 150 ppb, and gradually
202 decreased to 2.64×10^{-2} with an initial NO_x of 600 ppb. Hence, the AOC in the chamber
203 also showed a trend of first increasing and then decreasing with increasing NO_x
204 concentration. Sarrafzadeh et al. (2016) noted that the OH recycling reaction $\text{NO} + \text{HO}_2$
205 $\rightarrow \text{NO}_2 + \text{OH}$ was responsible for the NO_x-induced increase in OH concentration. Chen
206 et al. (2022b) revealed that the increasing NO_x level obviously enhanced the
207 atmospheric oxidation ability in a NO_x-sensitive (low-NO_x) regime. The higher
208 consumption of VOCs and the faster VOC oxidation rate both suggested there was a
209 higher atmospheric oxidation ability in the reaction system (Ng et al., 2007). In addition,
210 NO_x is an important sink for OH, so the competition between NO_x and VOCs for OH
211 might have been responsible for the decreases in VOC consumption and AOC under
212 high-NO_x conditions.

213 Although there were similar trends in the yield of SOA and AOC as the NO_x
214 concentration increased, no correlation between SOA yield and AOC was observed in
215 our study (Fig. S1). It has been shown that increases in AOC are essential drivers of
216 increases in SOA mass concentration in the troposphere (Feng et al., 2019; Li et al.,
217 2023). The higher SOA mass concentration under high AOC conditions was due to
218 more VOCs being oxidized, rather than an increase in SOA yield. Therefore, the
219 increased AOC was not the direct mechanism by which increasing NO_x concentration
220 influenced SOA yield.



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Figure 2. Changes in α -pinene concentrations over time (a) and the decay ratios of the VOCs (b) with different initial NO_x concentrations

225 The function of SOA yield with SOA mass concentration (M_0) was recalculated
226 based on Odum's SOA yield model (Fig. 3). In each experiment, an increase in M_0
227 directly resulted in an increase in SOA yield. Notably, if the photooxidation products
228 can be classified as low-volatility oxidation products, the SOA yield should remain
229 constant with increasing M_0 (Krechmer et al., 2015; Ehn et al., 2014; Odum et al., 1996).
230 However, the gas-particle distribution coefficients of semi-volatile substances are



231 directly related to their concentrations, and the distribution coefficients of semi-volatile
 232 substances into the aerosol phase are larger at higher concentrations (Akherati et al.,
 233 2019; Odum et al., 1996). Hence, the increasing distribution ratios of semi-volatile
 234 organic products between aerosol and gas phases at high M_0 were responsible for the
 235 increasing SOA yields with increasing photooxidation time (Kolesar et al., 2015;
 236 Valorso et al., 2011; Takeuchi et al., 2022).

237 A two-product model is an effective method for fitting the relationship between
 238 SOA yield and M_0 (Liu et al., 2019a; Odum et al., 1996). The model was calculated via
 239 equation (1), shown below:

$$Y = M_0 \times \left(\frac{\alpha_1 K_{om,1}}{1 + K_{om,1} M_0} + \frac{\alpha_2 K_{om,2}}{1 + K_{om,2} M_0} \right) \quad (1)$$

240 where, M_0 is SOA mass concentration (mg m^{-3}); α_1 and α_2 are the mass-based
 241 stoichiometric coefficients of species with low-volatility and semi-volatility products,
 242 respectively; and $K_{om,1}$ and $K_{om,2}$ ($\text{m}^3 \mu\text{g}^{-1}$) are the gas-particle partitioning equilibrium
 243 constants for low-volatility and semi-volatility products, respectively. The α_1 , α_2 , $K_{om,1}$,
 244 and $K_{om,2}$ of α -pinene SOA formed under different initial NO_x concentrations are shown
 245 in Table 2, and the simulated SOA yields with changing M_0 are shown in Fig. 3.

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Table 2 Parameters of the two product model for α -derived SOA under
 different initial NO_x concentration.

Initial NO_x conc. (ppb)	α_1	$K_{om,1}$ ($\text{m}^3 \mu\text{g}^{-1}$)	α_2	$K_{om,2}$ ($\text{m}^3 \mu\text{g}^{-1}$)
12	0.048	0.19	0.28	0.0040
25	0.038	0.19	0.30	0.0039
68	0.028	0.19	0.32	0.0037
150	0.019	0.19	0.33	0.0031
337	0.017	0.19	0.35	0.0019
600	0.014	0.19	0.38	0.0016

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250 It was clear that the curves of SOA yield to M_0 moved lower on the SOA yield axis
 251 in Fig. 3 with increasing NO_x concentration. According to Table 2, α_1 decreased while
 252 α_2 increased as the initial NO_x concentration increased. For the experiments conducted
 253 with 12 ppb NO_x , the α_1/α_2 ratio of α -pinene derived SOA was 0.17. This α_1/α_2 ratio
 254 decreased to 0.13, 0.086, 0.058, 0.049, and 0.037 as the initial NO_x concentration



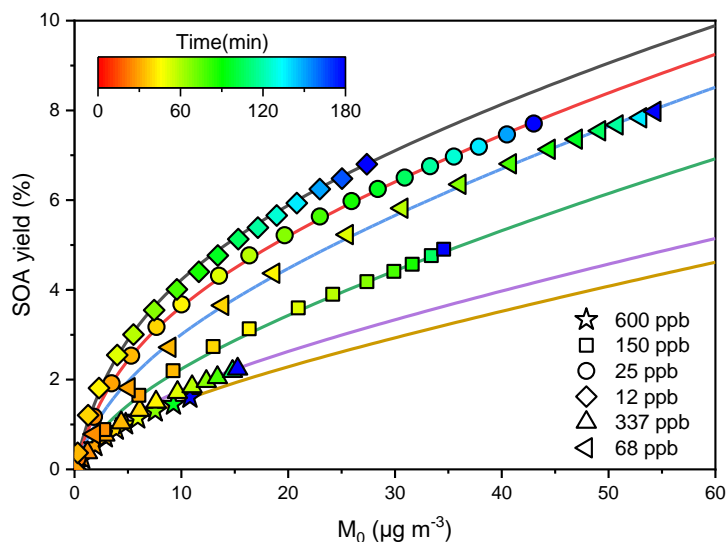
255 increased to 25, 68, 150, 337, and 600 ppb, respectively. Lower α_1 values indicate lower
256 proportions of low-volatility products in the SOA, while higher α_2 values indicate
257 higher proportions of semi-volatility products in SOA. In addition, the $K_{om,2}$ value also
258 decreased, dropping continuously from 0.0040 to 0.0016 $\text{m}^3 \mu\text{g}^{-1}$ as the initial NO_x
259 concentration increased from 12 to 600 ppb. This meant that the volatility of semi-
260 volatility products produced through α -pinene photooxidation increased with initial
261 NO_x concentration. Overall, with increasing initial NO_x concentration, the low-
262 volatility products made up decreasing proportions (lower α_1) of the total products
263 while semi-volatility products made up increasing proportions (higher α_2) and higher
264 volatilities (lower $K_{om,2}$).

265 To determine why the volatility of SOA increase with increasing NO_x, we
266 analyzed the consumption of VOCs under different NO_x conditions (Fig. 2). Due to the
267 lower consumption rate of VOCs and low AOC, α -pinene was not completely
268 consumed at the end of the photooxidation period under low-NO_x conditions. Increased
269 consumption of VOCs will lead to higher concentrations of photooxidation products
270 generated in the chamber, so when the initial NO_x concentration increased from 12 ppb
271 to 25 ppb to 68 ppb, the SOA mass concentration increased from 26.0 $\mu\text{g m}^{-3}$ to 43.8
272 $\mu\text{g m}^{-3}$ to 54.3 $\mu\text{g m}^{-3}$. As mentioned above, SOA yield increased with the mass
273 concentration of SOA. Even if the fitting curve of the two-product model gradually
274 moved lower on the graph with increases in NO_x, the SOA yield still increased from
275 6.5% to 8.0% when the initial NO_x concentration increased from 12 ppb to 68 ppb.
276 Hence, under low-NO_x conditions, more SOA was generated due to the increased AOC
277 and VOC consumption with increasing NO_x. The enhancement of the SOA yield with
278 increasing NO_x concentrations can be attributed to the increased ratio of the aerosol/gas
279 phase distribution resulting from higher concentrations of semi-volatile photooxidation
280 products. Chen et al. (2022c) compared the relative content of intermediate products
281 with different volatilities and found that the proportion of semi-volatile oxidized
282 products in gas-phase intermediate products was lower when experiments had higher
283 VOC consumption and SOA yields. Assuming that the proportions of different volatile



284 oxidation products remain constant, the smaller proportion of semi-volatile organic
285 products in the gas phase suggested that a larger proportion of semi-volatile organic
286 products was condensed into the particulate phase when more VOCs were consumed.
287 This result also supported our conclusion that the increased aerosol/gas phase
288 distribution ratio of semi-volatile products was the dominant driver underlying the
289 enhanced SOA yields with increased VOC consumption under low-NO_x conditions.

290 It should be noted that if the SOA yield is simulated based only on the two-product
291 model, the SOA yield should increase by 34% and 51% when NO_x increases from 12
292 ppb to 25 ppb and 68 ppb, respectively. However, in reality, the SOA yield only
293 increased 18% and 23%, respectively (Fig. S2). The real SOA yields were not only
294 lower than the simulated yields, but the discrepancy between the real and simulated
295 SOA yields increased with increasing NO_x concentration. As mentioned earlier,
296 reactions between RO₂ and NO_x form highly volatile oxidized compounds. Therefore,
297 the increasing discrepancy between real and simulated SOA yields with increasing NO_x
298 concentration indicated that inhibition of SOA formation through the RO₂ + NO_x
299 reaction pathway was always occurring. This indicated that, under low-NO_x conditions,
300 the increase in SOA yield resulted from a combination of effects, the positive effect of
301 the increased aerosol/gas phase distribution ratios of semi-volatile products and the
302 negative effect of the RO₂ + NO_x reaction.



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Figure 3. SOA yields as a function of organic aerosol mass concentration M_0 of α -pinene at different initial NO_x concentrations. The simulated SOA yields based on the two-product model are shown by the solid lines.

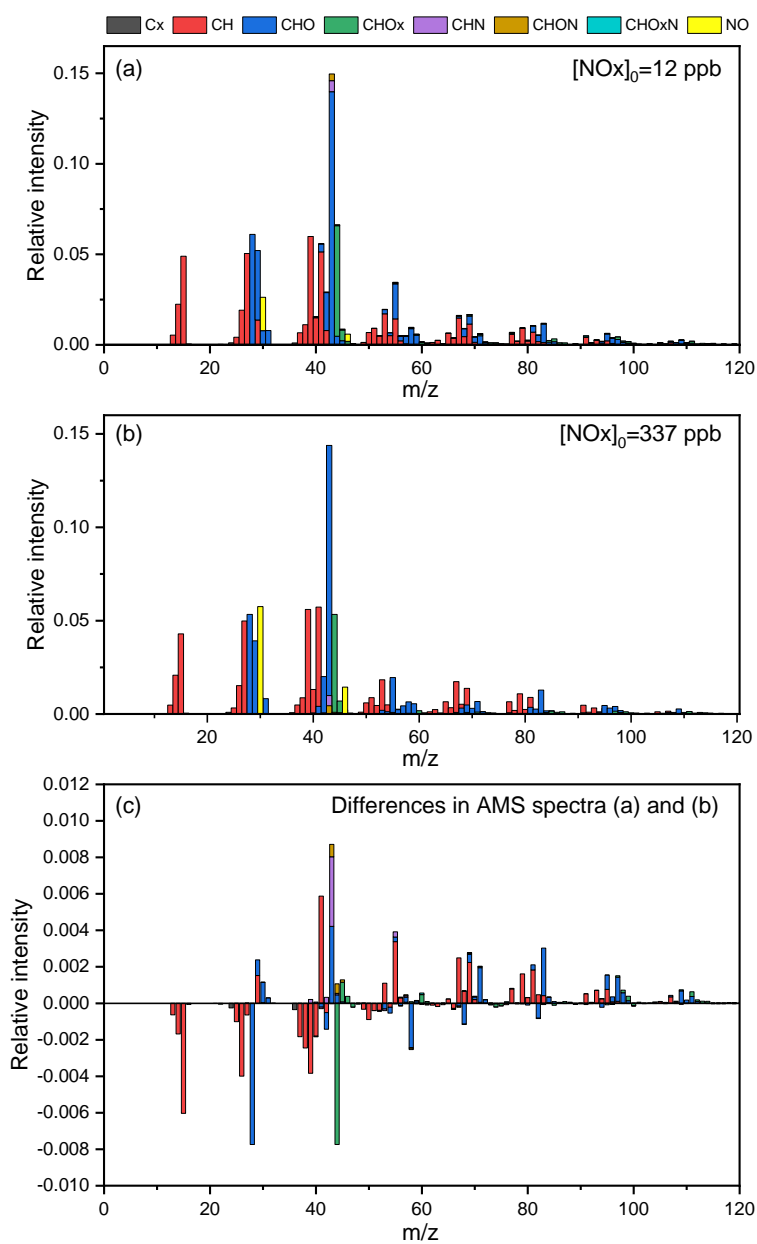
308 The effects of NO_x on the yields and chemical properties of SOA have been well
309 documented in previous studies (Chen et al., 2020; Eddingsaas et al., 2012; Sarrafzadeh
310 et al., 2016; Zhao et al., 2018). However, few studies have focused on how NO_x
311 influences VOC consumption during photooxidation, so the relationship between the
312 concentrations of NO_x and oxidation products has not been fully elucidated. For SOA,
313 which is mainly composed of semi-volatile oxidation products, the gas-particle
314 distribution coefficients of semi-volatile substances are directly related to their
315 concentrations. Therefore, SOA yield can be affected by the concentrations of semi-
316 volatile oxidation products, and this should be taken into consideration when studying
317 SOA formation process. Hence, future studies of photooxidation and SOA formation
318 should pay more attention to the formation processes of semi-volatile substances.

319 3.2 SOA chemical composition at different NO_x concentrations

320 SOA chemistry must be studied at the molecular level to better understand the
321 characteristics of SOA formation. The bulk chemical properties of SOA generated



322 under low- and high-NO_x concentrations are shown in Fig. 4.



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Figure 4. Typical AMS spectra of α -pinene SOA formed under low-NO_x conditions (a), high-NO_x conditions (b), and the differential spectra comparing the low- and high-NO_x conditions (c).



328 Organic fragments of $C_2H_3O^+$ ($m/z = 43$), which originated from the fragmentation
329 of aldehydes and ketones, were the dominant peaks in the AMS. The strong organic
330 AMS signal at $m/z = 43$ (f_{43}) indicated that the SOA oxidation level was relatively low.
331 The intensity of the CO_2^+ ($m/z = 44$), representing the thermal decarboxylation of
332 organic acid groups, was rather low, which indicated there was low carboxylic acid
333 content in the α -pinene derived SOA. Carboxylic acid has a relatively low volatility and
334 enters the particulate phase relatively easily, which means that the carboxylic acid
335 produced through the photooxidation of α -pinene was likely less than that detected in
336 the AMS. In our study, the f_{43} and f_{44} of α -pinene SOA ranged from 0.160 to 0.175 and
337 from 0.069 to 0.074, respectively. According to the “triangle plot” of the AMS, the SOA
338 derived from α -pinene photooxidation mainly fell in the lower area designated semi-
339 volatile oxygenated organic aerosols (SV-OOA) (Ng et al., 2010). The AMS results
340 provided direct evidence that semi-volatile products were the main components of α -
341 pinene SOA formed through NO_x photooxidation. This result further supported our
342 observations in Section 3.1, wherein the aerosol/gas-phase distribution ratio of
343 photooxidation products increased with increasing SOA mass concentration and VOC
344 consumption, resulting in the enhancement of SOA yield with increasing NO_x at low-
345 NO_x conditions.

346 The differences in the AMS spectra between low- and high- NO_x conditions are
347 compared in Fig. 4 (c). Although the inhibitory effect of the reaction between NO_x and
348 RO_2 on SOA formation was confirmed, the chemical compositions of the SOA formed
349 under different NO_x conditions showed only minor differences. Photooxidation
350 products formed through the reaction of NO with RO_2 radicals usually have relatively
351 high volatilities. However, the α -pinene derived SOA was mainly composed of semi-
352 volatile photooxidation products, and these photooxidation products formed under
353 high- NO_x conditions with higher volatilities may not easily be condensed into the
354 particulate phase, so they were mainly present in the gas phase. This means that the
355 high volatility products formed via the $RO_2 + NO$ reaction path under high- NO_x
356 conditions did not affect the chemical composition of SOA. Therefore, the AMS spectra



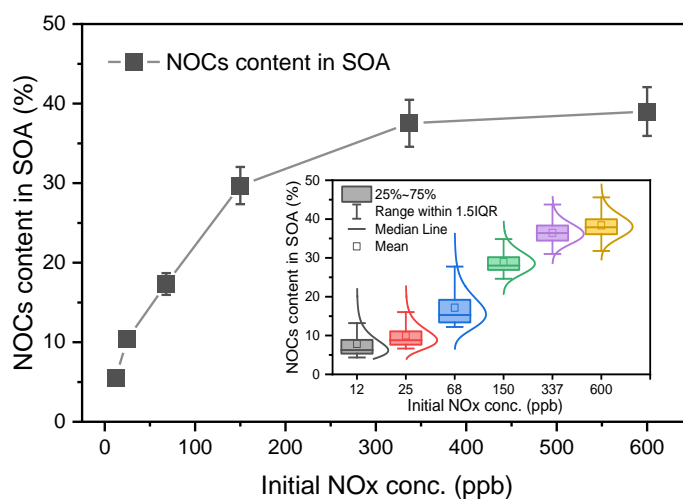
357 of the α -pinene SOA were not significantly different between low- and high-NO_x
358 conditions.

359 In addition to the generation of RO, the reaction of RO₂ with NO can also form N-
360 containing compounds (NOCs). The NOC contents in SOA under different NO_x
361 conditions were calculated based on the AMS fragments of NO⁺ and NO₂⁺. NO⁺ and
362 NO₂⁺ are characteristic fragments of NOCs, but the same NO⁺ and NO₂⁺ fragments can
363 also be detected by AMS in the forms of inorganic nitrates. In order to distinguish
364 organic nitrates from total nitrates, we used the same method as described in our
365 previous study based on the differences in the ratios of NO⁺/NO₂⁺ for organic and
366 inorganic nitrates in the AMS mass spectra (Xu et al., 2015b; Ng et al., 2017; Day et
367 al., 2022). The concentrations of NO⁺ and NO₂⁺ from organic nitrates were calculated
368 using the following equations:

$$\text{NO}_{2,\text{org}} = \frac{\text{NO}_{2,\text{meas}} \times (R_{\text{meas}} - R_{\text{AN}})}{R_{\text{ON}} - R_{\text{AN}}} \quad (4)$$

$$\text{NO}_{\text{org}} = R_{\text{ON}} \times \text{NO}_{2,\text{org}} \quad (5)$$

369 where R_{meas} is the NO⁺/NO₂⁺ ratio from the AMS results; R_{ON} and R_{AN} are the
370 NO⁺/NO₂⁺ ratios for organic nitrate and ammonium nitrate, respectively; and $\text{NO}_{2,\text{meas}}$
371 represents the total NO₂⁺ fragments obtained from the AMS data. In this study, R_{AN} was
372 about 1.1 based on the measurements and R_{ON} was assumed to be about 10 referring to
373 the study by Takeuchi and Ng (2019). Previous studies have shown monoterpene
374 hydroxyl nitrate ($m/z = 215$), pinene keto nitrate ($m/z = 229$), and monoterpene
375 dicarbonyl nitrate ($m/z = 247$) to be the main NOCs from α -pinene photooxidation
376 (Chen et al., 2022a; Li et al., 2018). Here, the formation of NOCs was calculated based
377 on organic nitrate contents and the molecular weights of NOCs.



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Figure 5 Content and mass concentration of NOCs under different initial NOx conditions.

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The contents and mass concentrations of NOCs under different NOx levels are shown in Fig. 5. The increase in NOC contents with NOx could be divided into two stages. The first stage occurred when the initial NOx concentration increased from 12 ppb to 150 ppb. During this stage the content of NOCs in the SOA increased linearly, indicating that the NOx was the limiting factor for NOC formation in low-NOx conditions. The second stage occurred as NOx concentrations continued to rise to 600 ppb. During this stage, the growth in NOC content gradually slowed while approaching the maximum value. Based on the nonlinear fit between NOx concentration and NOC content, the maximum value of NOCs content in SOA was predicted to be about $39 \pm 3.8\%$. This indicated that, under extremely high-NOx conditions, NOCs could account for up to two-fifths of the α -pinene SOA. These results not only showed that NOCs play a more important role in biogenic SOA formation under high-NOx conditions, but can also serve as a basis for estimating the generation of NOCs in SOA.

394

3.3 Effect of initial VOC concentration on SOA formation

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The SOA yields with different α -pinene concentrations were also compared in this study. As shown in Fig. 1, the change trends in SOA yields with increasing NOx



397 concentrations were similar across all experiments, including low- and high-VOC
398 conditions. However, higher initial VOC concentrations led to increases in both
399 maximum SOA yield and the initial NO_x concentration required for the SOA yield to
400 reach its maximum value. Under low-VOCs conditions, based on the no-linear fitting
401 of SOA yield with initial NO_x, the peak SOA yield was 8.8% when the initial NO_x
402 concentration was 41 ppb. The maximum SOA yield increased to 26.7% under high-
403 VOC conditions, and the concentration of NO_x corresponding to the maximum SOA
404 yield increased to 66 ppb. Under high-VOC conditions, the maximum SOA mass
405 concentration was 370.3 μg m⁻³. However, the maximum mass concentration of SOA
406 formed under high-VOC conditions was only 54.3 μg m⁻³, which was 6.8 times lower
407 than that under high-VOCs.

408 Previous studies have suggested that the formation of SOA from α-pinene
409 photooxidation is primarily limited by NO_x concentration under low-NO_x conditions,
410 and variation in VOC concentrations have little effect. However, as shown in Fig. 1, the
411 SOA yield under high-VOC conditions was consistently higher than that under low-
412 VOC conditions. Moreover, the ratio and gap in SOA yields between high- and low-
413 VOC conditions significantly increased with increasing NO_x concentrations when NO_x
414 concentrations were lower than 60 ppb (Fig. S3). Therefore, it can be concluded that
415 under low-NO_x conditions, the formation of SOA from α-pinene photooxidation is
416 limited by both VOCs and NO_x concentration. Based on differences in VOC
417 consumption between high- and low-VOCs conditions (Fig. S4), it appeared that, at the
418 same NO_x concentration, the consumption of VOCs was higher under high-VOC
419 conditions and the difference in VOC consumption between high- and low-VOC
420 conditions gradually increased with increasing NO_x. According to the aerosol/gas-
421 phase distribution ratio of semi-volatile products described in Section 3.1, higher VOC
422 consumption can generate more semi-volatile oxidation products, which can enhance
423 the SOA yield. The continuously increasing concentration gradient in VOC
424 consumption between high- and low-VOC conditions indicated that the SOA yield
425 under low-NO_x conditions was also affected by the concentration of VOCs.

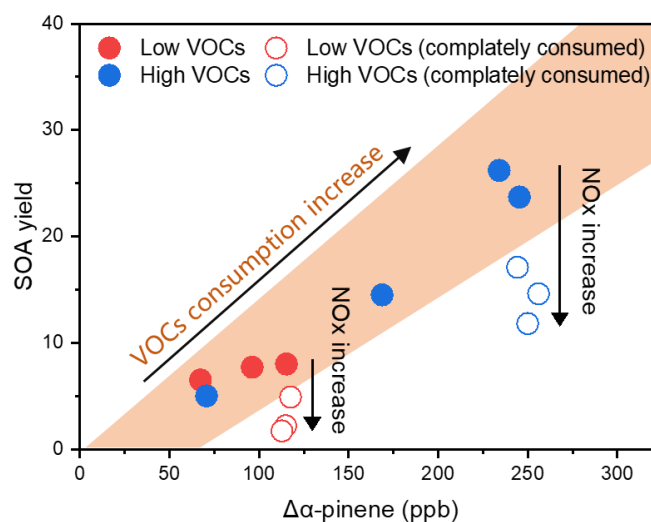


426 When the initial NO_x concentration was higher than 100 ppb, the growth rate of
427 the SOA yield ratio between high- and low-VOC conditions was 67.7% lower than that
428 under low-NO_x conditions. The ratio of SOA yield from high-VOC experiments was
429 about 3–8 times higher than that from the low-VOC experiments, which surpassed the
430 VOC ratio between different VOC conditions. This indicated that, under high-NO_x
431 conditions, SOA yield has a high sensitivity to changes in VOC concentrations, and the
432 inhibition effect of reducing VOCs on SOA yields would be more pronounced under
433 high-NO_x conditions. It should be noted that, while the difference in VOC consumption
434 did not change further under high-NO_x conditions, the ratio of the SOA yields between
435 high- and low-VOCs conditions continued to increase with increasing NO_x
436 concentrations. At lower VOC conditions, the RO₂ + NO pathway was more
437 competitive, which enhanced the production of high-volatile products compared to in
438 the higher VOC experiments. This sustained increase in the ratio of SOA yields may
439 have been mainly due to the differences in volatilities of the oxidation products
440 produced under high- and low-VOC conditions. Therefore, as illustrated in Fig. S3, it
441 can be inferred that changes in VOC concentrations have a more pronounced effect on
442 SOA yield under higher NO_x conditions.

443 Although the trend in SOA yields with increasing NO_x was similar under different
444 VOC conditions, the change trend in AOC with increasing NO_x was different (Fig. S5).
445 First, under low-NO_x conditions, the AOC under high-VOC conditions was lower than
446 when under low-VOCs. The consumption of α -pinene was higher in the high-VOC
447 experiments than in the low-VOC experiments and more oxidants were consumed by
448 α -pinene under high-VOC conditions, which may have led to the inhibition of AOC.
449 Secondly, AOC only increased with increasing NO_x concentrations under high-VOC
450 conditions. According to the fitting results in Fig. S5, under low-VOC conditions, AOC
451 initially increased with NO_x and then decreased, with an inflection point at VOCs/NO_x
452 ≈ 0.47 . However, under high-VOC conditions, AOC increased with initial NO_x
453 concentration across the entire NO_x gradient. Although the rate of change in AOC
454 gradually slowed as NO_x increased, no inflection point occurred in the AOC change



455 trend even when NO_x reached 669 ppb, at which point VOCs/NO_x decreased to 0.38.
 456 These results suggested that AOC is jointly affected by NO_x and VOCs, and that
 457 VOCs/NO_x is not sufficient as a direct basis for the evaluation of AOC. This
 458 demonstrated the complicated nature of accurately simulating AOC, which is an
 459 important research topic that needs to be studied further.



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Figure 6 Relationship between consumed VOCs and SOA yields under different NO_x and VOC conditions

464 The SOA yield trends at different AOCs are shown in Fig. S1. Based on the
 465 photooxidation experiments with different VOC concentrations, there was no clear
 466 correlation between SOA yield and AOC. This result further explains why increases in
 467 AOC are not the direct cause of increased SOA yields from α -pinene photooxidation.
 468 We further compared the consumption of VOCs with SOA yield (Fig. 6) and found a
 469 significant positive correlation between the two. This result was consistent with the
 470 study by Jiang et al. (2022) which reported that SOA yield increased with increasing
 471 consumption of VOCs, even when the OH concentration was held constant across
 472 experimental photooxidation groups. The continuous growth trend in SOA yield with
 473 increasing consumption of VOCs, which overlapped among different VOC conditions,
 474 further supported the processes discussed in Section 3.1; that is to say, changes in the
 475 aerosol/gas-phase distribution ratio for semi-volatile products with changing VOC



476 concentrations lead to increased SOA yields with increasing NO_x concentration under
477 low-NO_x conditions.

478 However, when the NO_x concentrations exceeded 68 ppb and 113 ppb under low-
479 and high-VOC conditions, respectively, α -pinene was almost entirely consumed during
480 the photooxidation process. This meant that, although AOC continued to increase with
481 increasing NO_x, further increases in AOC only accelerated VOC consumption time
482 without impacting total VOC consumption or photooxidation products formation. As a
483 result, the SOA yield did not continue to increase. In contrast, more NO_x would
484 promote the RO₂ + NO reaction, which would result in a decreasing trend in the SOA
485 yield.

486 **4. Atmospheric implications**

487 NO_x is a key substrate for photooxidation in the atmosphere. Changes in NO_x
488 concentrations can lead to significant changes in the photooxidation of VOCs and affect
489 subsequent SOA formation. In this study, SOA yields were observed to increase sharply
490 at first, and then decrease gradually with increasing NO_x concentration. The maximum
491 SOA yields in the low- and high-VOC conditions were 8.0% and 26.2%, respectively.
492 Based on the relationship between SOA yield and M₀, as well as the SOA chemical
493 composition, semi-volatile oxidation products were the main components of α -pinene
494 derived SOA. Under low-NO_x conditions, increasing NO_x concentration increased the
495 consumption of VOCs during photooxidation and lead to an increase in the SOA yield.
496 We concluded that the combination of increased distribution ratios of the semi-volatile
497 oxidation products with increasing amounts formed was key in the promotion of SOA
498 yield with increasing NO_x. Conversely, the effect of NO_x on AOC did not directly cause
499 the increase in SOA yield. NO_x can significantly increase the content of NOCs in SOA.
500 At their highest, under extremely high-NO_x concentrations, NOC contents made up as
501 much as two-fifths of α -pinene derived SOA.

502 With the progression of urbanization and the changes to the natural environments
503 surrounding urban areas, atmospheric environments of have become quite different (Xu



504 et al., 2015b; Domínguez-López et al., 2014; Shon et al., 2007; Agbo et al., 2022). Most
505 rural areas exhibit NO_x limitation, while urban areas more often exhibit VOC limitation
506 (Tan et al., 2018; Santos et al., 2021; Hui et al., 2018). In areas with low-NO_x
507 concentrations, the consumption of biogenic VOCs might not be significant despite
508 there being high biogenic VOC emissions, which would limit the formation of SOA.
509 However, if unreacted VOCs are transported to areas with high-NO_x concentrations (i.e.,
510 from rural to urban areas), or vice versa, there could be explosive increases in SOA
511 concentrations and noticeable changes of SOA chemical compositions. Therefore, the
512 regional transport of pollutants between urban and forested areas will significantly
513 impact the formation of atmospheric aerosols and environmental pollution. This was
514 supported by our study, which demonstrated how human activities can enhance SOA
515 formation from biogenic emissions.

516 Finally, as a typical goal, research often works toward creating or improving
517 models to simulate atmospheric phenomena. However, past simulations have generally
518 underestimated SOA concentrations when compared to field observations, especially
519 during pollution periods (Ling et al., 2022; Kelly et al., 2018). Our study noted that,
520 under high-NO_x and high-VOC conditions, SOA yield exhibited significant changes
521 with changes in reaction conditions (Fig. S3). Under polluted conditions, the faster
522 consumption of VOCs might lead to underestimates of VOC concentrations in the
523 atmosphere and, subsequently, the generation of oxidation products, ultimately
524 resulting in underestimates of SOA yields. Hence, our research provides crucial
525 information for improving the accuracy of air quality models simulating SOA formation.

526

527 **Data availability**

528 The data used to support the conclusions in this study are available at a public data
529 repository of Figshare via [https://figshare.com/articles/dataset/NOx_photooxidation_](https://figshare.com/articles/dataset/NOx_photooxidation_pinene/25200929)
530 [pinene/25200929](https://figshare.com/articles/dataset/NOx_photooxidation_pinene/25200929) (S. Liu, 2024)

531

532 **Author contributions**



533 SL and GW designed the whole work and wrote the paper. SL did the experiment,
534 collected the samples, conducted the sample analysis and performed the data
535 interpretation. All authors contributed to the paper with useful scientific discussions
536 and comments.

537

538 **Competing interests**

539 The authors declared that they have no conflict of interest.

540

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