Title page

Effect of colloidal particle size on physicochemical properties and aggregation behaviors of two alkaline soils

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Abstract

 Colloidal particles are the most active soil components, and they vary in elemental composition and environmental behaviors with the particle size due to the heterogenous nature of natural soils. The purposes of the present study are to clarify how particle size affects the physiochemical properties and aggregation kinetics of soil colloids, and to further reveal the underlying mechanisms. Soil colloidal fractions, from two alkaline 26 soils—Anthrosol and Calcisol were subdivided into three ranges: $d < 2 \mu m$, $d < 1 \mu m$ and *d* < 100 nm. The organic and inorganic carbon contents, clay mineralogy, surface electrochemical properties, including surface functional groups and zeta potentials, were characterized. Through time-resolved light scattering technique, the aggregation kinetics of soil colloidal fractions were investigated, and their critical coagulation concentrations (CCCs) were determined. With decreasing colloidal particle diameter, the total carbon content, organic carbon, organic functional groups content and illite content all increased. The zeta potential became less negative and the charge variability decreased with decreasing particle diameter. The CCC values of Anthrosol and Calcisol 35 colloids followed the descending order of $d < 100$ nm, $d < 1$ µm, $d < 2$ µm. Compared 36 with the course factions $(d < 1 \mu m)$ and $d < 2 \mu m$, soil nanoparticles were more abundant in organic carbon and more stable clay minerals (*d* < 100 nm), thus they exhibited strongest colloidal suspension stability. The differences in organic matter contents and clay mineralogy are the fundamental reasons for the differences in colloidal suspension stability behind the size effects of Anthrosol and Calcisol colloids. The present study revealed the size effects of two alkaline soil colloids on carbon content, clay minerals,

 surface properties and suspension stability, emphasizing that soil nanoparticles are prone to be more stably dispersed instead of being aggregated. These findings can provide references for in-depth understanding of the environmental behaviors of the heterogeneous soil organic-mineral complexes.

 Keywords: Nanoparticles; Clay mineral composition; zeta potential; Critical coagulation concentration

1. Introduction

 Soils contain a series of solid particles in continuous sizes, ranging over six orders of magnitude from nanometers to millimeters (Lead and Wilkinson, 2006; Li et al., 2011), among which soil colloids are the most reactive fractions. Soil colloids are characterized by high surface area and abundant surface charges, exhibiting high potential for carbon sequestration and strong adsorption capacity, which can largely determine the fate and transport of pathogens, nutrients, heavy metals and organic pollutants, and might cause environmental problems to adjacent water bodies or groundwater (Baalousha et al., 2009; Calabi-Floody et al., 2011). Due to their high reactivity and fluidity in aqueous environment, soil colloids play an important role in physical, chemical and biogeochemical processes in natural environment (Schäfer et al., 2012; Mayordomo et al., 2016). The capacity of soil colloids in mobilizing bound nutrients and pollutants is closely related to their dispersion stability under various environmental conditions (Won and Burns, 2018). Therefore, studies on the dispersion stability of soil colloids have attracted extensive attention.

 Currently, the definition of soil and environmental colloidal fractions is ambiguous. 64 Soil colloidal fractions are defined as soil particles in diameter of $\leq 1 \mu m$ (Lead and Wilkinson, 2006; Weil and Brady, 2016), and also being of < 2 μm (Zhang et al., 2021); 66 while in some extreme cases, they can refer to the particles in diameter of $5-10 \mu m$ (Yin et al., 2010). Such discrepancies are seen among publications due to the fact that colloids are defined based on the particle diameter range within which they can display colloidal properties. Since for different materials, e.g., metal (Fe/Al/Ti) oxides, silica

gel, phyllosilicates, the specific colloidal range differs greatly.

 Compared with engineered nanoparticles with known mineralogical organization, natural soils are much more heterogeneous (Cárdenas et al., 2010); their elemental composition and clay mineralogy of soil colloids change with particle size. Tsao et al., (2013) found that quartz and feldspar were mainly dominant in colloidal particles of < 2 μm and 450–2000 nm in red soil (Ferralsols, WRB, IUSS Working Group WRB. 2022), while illite and montmorillonite were the main clay minerals in nanoparticles (1–100 nm). In addition, the mineral structure at nanometer scale also changes. Compared with colloidal particles of < 2 μm, the Si/Al ratio in nanoparticles increased, and the surface area, morphology, crystallinity, surface atomic structure and frame structure were significantly different (Tsao et al., 2011). Furthermore, particle size also affects the surface potential of soil colloids. Tang et al., (2015) investigated the surface 82 potential variations with particle size $(1-10 \mu m, 0.5-1 \mu m, 0.2-0.5 \mu m, < 0.2 \mu m)$ for variably-charged yellow soil (Lixisols) and permanently-charged purple soil (Leptosols); among the colloidal fractions, the absolute surface potential of the finest particles of purple soil (Leptosols) was lowest while that of the yellow soil (Lixisols) was the largest, caused by the differences in surface charge density. Thus, the influences of particle size on elemental composition and surface properties of soils should be further studied.

 In recent years, great progress has been made in the study of dispersion stability of soil clay minerals, such as montmorillonite, kaolinite, illite or hematite, and soil nanoparticles (Xu et al., 2018; Sun et al., 2020; Wei et al., 2021; Zhu et al., 2014). He

 et al., (2008) demonstrated that hematite nanoparticles with various particle diameters showed different surface properties and aggregation behaviors under the same pH conditions; moreover, the critical coagulation concentrations (CCCs) of hematite decreased with the decrease of particle diameter. Zhou et al., (2013) compared the 96 CCCs of ten different TiO₂ nanoparticles with varying sizes and indicated that crystal 97 structure and particle diameter both affected the aggregation behaviors of $TiO₂$. Zhang et al. (2016) confirmed that the types of clay minerals for two Alfisols changed from smectite and vermiculite to kaolinite and illite when the particle size varied from colloids to nanoparticles. Therefore, the dependence of physiochemical properties, surface properties and environmental behaviors on particle size for heterogeneous soil colloidal particles needs systematic investigation.

 In the present study, soil colloidal particles of two alkaline soils—Anthrosol and 104 Calcisol were subdivided into three ranges: $d < 2 \mu m$, $d < 1 \mu m$ and $d < 100$ nm. Their organic fraction and clay mineralogy, surface electrochemical properties and colloidal stability were studied. This study selected two representative calcareous soils to verify the following scientific hypothesis: soil colloids are organic-inorganic composites. As particle diameter decreases (from colloid particles to nanoparticles), the number of organic functional groups on the surface of soil colloids increases, and the type of clay minerals shifts towards finer clay particles, e.g. illite, resulting in increased specific surface area and decreased charge density, and thus enhanced suspension stability, meaning particle diameter influences the composition of soil colloidal fractions, thereby changing surface properties and suspension stability. The findings can have important implications for predicting the environmental performances of colloids and colloid-facilitated nutrients, pollutes and pathogens in natural soil and water environment.

- **2. Materials and methods**
- **2.1 Soil sampling**

 The study collected two surface soil samples (0–20 cm), Lou soil and Cinnamon soil, being calcareous soils, by mixing soils form 5–10 sampling points using a stainless-steel auger at Yangling District (N38°18′14″ and E108°2′30″), and Zhouzhi Country (N34°8′8″ and E108°3′10″), in Shaanxi province, northwest China, respectively. According to World Reference Base for Soil Resources (WRB, IUSS Working Group WRB. 2022), the Lou soil and Cinnamon soil are classified as Calcic Protocalcic Calcisols (Loamic, Lixic, Humic) and Hortic Endoanthric Anthrosols (Loamic, Luvic, Eutric, Calcic), respectively. Both types of soil developed from loess parent material. The typical soil profile configuration for the tested Anthrosol is Ap1- Ap2-Bt-Bk-C, while for the Calcisol, it is Ah-Bt-Bk-C.

 Soils samples were taken back to laboratory for air-drying and sieving. The basic soil properties were determined based on standard methods. Soil pH was measured with a pH electrode, employing a solution-to-soil ratio of 2.5:1. Soil organic carbon (SOC) 132 was determined using the $K_2Cr_2O_7$ oxidation method. The cation exchange capacity 133 (CEC) of soil was measured with exchange method. The $CaCO₃$ content was determined by gasometric method. The free Fe/Al oxides were extracted by dithionite-citrate-bicarbonate (DCB) solution. The particle size distribution was measured using

2.2 Extraction of soil colloidal fractions in different size ranges

 The soil colloidal particles were extracted based on the Stokes' law, and the detailed procedures can be found in our previous publication (Hu et al., 2022). Briefly, 50 g of dry soil was weighed into a beaker containing 500 mL of distilled water, and put the suspension under sonication for an hour using the ultrasonic cell disrupter (XO- 900D, Nanjing Xianou Instruments Corporation, China) while maintaining the temperature below 30°C. Afterwards, the suspension was transferred to a larger beaker and distilled water was added to make up the total volume of 5 L. The suspension was further dispersed using an electronic blade stirrer (JB-200, Shanghai Nanhui Huiming Apparatus, China) for one hour, before being sieved through sieve with a pore size of 53 μm, and the upper suspensions containing soil colloidal particles in different diameters were collected by centrifugation. Based on the equation (1), centrifugation speed and time for colloidal particles of *d* < 2 μm, < 1 μm and < 100 nm were calculated and shown in Table S1.

$$
t = \frac{\eta \lg \frac{R_2}{R_1}}{3.81N^2 r^2 \Delta d}
$$
(1)

159 in which, *t* is time for centrifugation (s); R_1 is the distance from the surface of the liquid 160 to the center of the axis of the centrifuge, here is 5.7 cm; R_2 is distance from the particles 161 to the center of the axis of the centrifuge, here is 10.5 cm; N (rev·s⁻¹) is the centrifuge 162 speed; *r* (cm) is the desired colloidal particle radius; Δd is the difference in density 163 between the soil particles (2.65 g·cm⁻³) and water (1 g·cm⁻³), while Δ*d* is 1.65 g·cm⁻³; 164 *η* is the water viscosity coefficient, here is 0.00839 g·cm⁻¹·s⁻¹ at 25 °C.

165 **2.3 Characterization of soil colloidal fractions in different size ranges**

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 The initial particle diameters of soil colloids were determined by a time-resolved dynamic light scattering (DLS) apparatus (Nanobrook Omni, Brookhaven, USA). The organic carbon contents in soil colloids were determined by potassium dichromate external heating method and total carbon content was determined by elemental analyzer (Elementar Vario EL III, Germany). The inorganic carbon content was calculated by subtraction method (Wang et al., 2011). The clay mineralogy of soil colloids was determined by the XRD (Ultima-IV, Rigaku, Japan), and by comparing the intensity of the dominant X-ray diffraction peak of the soil mineral colloid to a standard mineral reference (Database ICDD 2004), the relative percentage content of the minerals was 175 determined. The specific surface areas of the soil colloids were measured by $BET-N₂$ method (ASAP 2460, Micromeritics instrument, USA). High-resolution spectra of C1s and O1s of soil colloids were acquired by X-ray photoelectron spectroscopy (XPS) (Thermo Scientific K-Alpha, USA) (Luo et al., 2019), and the Gaussian−Lorentzian curve-fitting program (XPSPEAK 4.1) was used to analyze the XPS spectra. The zeta potentials of soil colloids were measured by Zeta PALS equipped with a BI-ZTU 181 Autotitrator (ZetaPALS, Brookhaven, USA) with 1 mmol·L⁻¹ NaCl solution as the background electrolyte; and the pH range of colloidal suspension was set to 3–10 183 adjusted with 0.1 mol·L⁻¹ HCl and NaOH. The concentrations of K⁺, Na⁺, Ca²⁺, and Mg^{2+} in soil colloidal suspensions were measured by flame atomic absorption spectrophotometry (PinAAciie 900F, USA).

2.4 Aggregation kinetics of soil colloidal fractions

 The aggregation kinetic curves of soil [colloidal particles](javascript:;) in different electrolytes were determined by time-resolved DLS measurements. The incident wavelength was 635 nm and the scattering angle was 90°. The stock colloidal suspensions with particle 190 concentration of 200 mg⋅L⁻¹ were mixed with electrolyte solutions with equal volume. The suspension pH was adjusted to 8.0, which was close to the pH value of natural soil 192 with addition of 0.1 mol⋅L⁻¹ HCl or NaOH before measurement. The chosen electrolyte 193 concentrations for NaCl and CaCl₂ were 200–2000 and 2–20 mmol⋅L⁻¹. The effective 194 diameter (D_h) of the mixed sample was automatically recorded every 2 min, and an aggregation kinetic curve was obtained in 30 min monitoring.

2.5 Calculation of critical coagulation concentration

 According to the particle interaction theory, the aggregation kinetic curves under electrolyte conditions can be divided into reaction-limited aggregation (RLA) stage under low concentration which was affected by electrolyte conditions and diffusion-limited aggregation (DLA) stage under high concentration which was not affected by electrolyte concentration. The CCC is the critical electrolyte concentration when the 202 aggregation process changes from the RLA state $(a < 1)$ to the DLA state $(a = 1)$. Attachment efficiency (*α*) represents the bonding probability of particle collisions and can be calculated for each electrolyte concentration by using equation 2, which allowed 205 the curve of α as a function of electrolyte concentration to be plotted (Xu et al., 2020a; Hu et al., 2022).

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$$
\alpha_{\exp} = \frac{1}{W} = \frac{k_{11}}{(k_{11})_{\text{fast}}} = \frac{\frac{1}{N_0} \left(\frac{d a_h(t)}{dt} \right)_{t \to 0}}{\frac{1}{(N_0)_{\text{fast}}} \left(\frac{d a_h(t)}{dt} \right)_{t \to 0, \text{fast}}} \tag{2}
$$

208 where D_h is the effective diameter of particles, *t* is the time (min); N_0 is the density of 209 particles; K_{11} is the aggregation rate of RLA; $(K_{11})_{\text{fast}}$ is the aggregation rate of DLA. 210 The intersection of RLA regime and DLA regime is the CCC.

 The aggregation rates were calculated by the average of the last 5 effective diameters divided by the aggregation time at specific electrolyte concentration. The fractal dimension in the DLA regime was obtained based on the method proposed by Wang et al. (2013).

215
$$
D(t) = b * t^{n} + D_{0}
$$
 (3)

216 in which, $D(t)$ is the colloidal effective diameter at time t (min), D_0 is the initial effective 217 diameter of colloids, and *b* and *n* are constants determined by the aggregation curves.

218 The fractal dimension is $d_f = 1/n$ in the DLA regime.

²¹⁹ **3. Results and discussion**

3.1 Particle size and distribution characteristics of Anthrosol and Calcisol colloidal fractions

 The average diameters of Anthrosol and Calcisol colloids were measured by time- resolved DLS, and the results were shown in Table 1. The number-weighted diameters 224 for Anthrosol colloids of $d < 2$ μm were about 1.04 times of $d < 1$ μm, and about 1.84 times of d < 100 nm, respectively. The intensity-weighted diameters for Anthrosol colloids were 294.10~396.81 nm. For Calcisol colloidal fractions, the number-weighted 227 diameters for colloids of $d < 2$ μ m were about 1.07 times of $d < 1$ μ m, and about 1.65 times of d < 100 nm, and the intensity-weighted diameters were 312.25~439.20 nm. The intensity-weighted diameters were generally higher than the number-weighted diameters, especially in polydisperse system (Xu et al., 2020b). The colloidal particles in the soil solution were in constant Brownian motion, upon illumination by light, these colloidal particles scatter light, causing variations in light intensity. This phenomenon allowed for the calculation of the effective diameter of the particles, which was the intensity-weighted diameter (Filella et al., 1997). Given that particle diameter is proportional to the sixth power of light intensity. Consequently, in polydisperse systems where larger particles were present, the number-weighted diameter provided typically a more accurate representation of the true diameter of colloidal particles (Xu et al., 2015).

(*Insert Table 1 near here*)

240 From table 1, it can be seen that the average colloidal diameters of $d < 2 \mu m$ were 241 close to that of $d < 1$ µm, and they were both significantly higher than that of the nano sized fraction. From the particle size distribution characteristics, it is clear that the size 243 range indicated by the differences of D_{90} and D_{10} increased with intended particle diameter. For Anthrosol and Calcisol, 74.69% and 63.55% of all particles contained in 245 the colloidal suspensions of $d \le 100$ nm were actually less than 100 nm, respectively, indicating the complexity of soil colloidal particle irregularity.

3.2 Physiochemical properties and clay mineralogy of Anthrosol and Calcisol colloids

 Table 2 shows the physiochemical properties of soil colloidal fractions. The yields of each colloidal fraction of Anthrosol were slightly larger than that of Calcisol, 251 respectively. The yields of colloidal particles of $d < 2$ µm were about 1.3~1.4 times of $d \le 1$ µm, and about 4.0~4.9 times of $d \le 100$ nm, respectively. With the decreasing colloidal particle diameter, the total carbon content, organic carbon and inorganic carbon content all increased, suggesting the finer particles were richer in carbon. This tendency is in agreement with other publications (Zhang et al., 2021; Said-Pullicino et 256 al., 2021; Hu et al., 2022). The specific surface areas for colloidal fractions of $d < 1 \mu m$ were largest of all, which may be related to the structures of formed clusters while drying the samples for observation under microscopy (Yu et al., 2017; Weissenberger et al., 2021). Furthermore, Anthrosol and Calcisol nanoparticles exhibited the lowest specific surface area. This phenomenon raised from Organic substances adsorb relatively little inorganic nitrogen (Li et al., 2013; Wilson et al., 2008). Therefore, However, to our knowledge, no other better method has been reported for measuring the specific surface area of natural nanoparticles.

 nm showed a different trend, compared with that in Anthrosol colloids. The relative contents of organic oxygen-containing functional groups gradually increased with the decrease of diameter. This trend was particularly pronounced in fraction of *d* < 100 nm, and contents of C–O and COO- were highly increased (Table S2).

(*Insert Figure 1 near here*)

 Oxygen-containing functional groups of C–O, C=O and COO- are electronegative functional groups, hydroxyl and carboxyl groups can lose protons and make the surface of soil colloidal particles carry negative charges (Audette et al., 2021). Functional groups of C–O, C=O and COO- can affect the negative charges carried on the colloidal surface by forming hydrogen bonds, and their polarity can also affect the negative charges on the surface when O atom combines with C and H. The electrons will lean towards the O atom with stronger electronegativity, which also makes the colloidal surface carry negative charges (Tan et al., 2019). The contents and types of oxygen- containing functional groups are one of the main factors affecting colloid charge and aggregation.

 The zeta potential values of different colloidal fractions at the pH range of 3–10 are shown in Fig. 2. Zeta potentials of the colloidal particles were negative, indicating that they were negatively-charged. The zeta potentials of Anthrosol and Calcisol colloidal particles were more negative with increasing solution pH, due to the deprotonation of the surface (Moayedi and Kazemian, 2013; Dong et al., 2019). Compared with the Calcisol colloids, the zeta potentials of Anthrosol colloidal particles were more negative, due to Anthrosol possessed a higher surface charge density. For Calcisol colloids, the differences among colloidal fractions were larger.

(*Insert Figure 2 near here*)

 In general, zeta potential became more negative with increasing particle diameter. When the pH changed from 3 to 10, for every pH unit increased, the zeta potential 313 values of Anthrosol colloids of $d \le 2$ um, ≤ 1 um, and ≤ 100 nm would be increased by 2.14 mV, 2.09 mV and 1.89 mV; and for Calcisol colloids, those variation rates were 2.15 mV, 1.45 mV and 1.37 mV, respectively. Those data demonstrate that the charge variability decreasing with the decreasing particle diameter. Song et al., (2019) compared the zeta potential of wheat straw biochar nanoparticles (< 100 nm) and colloidal particles (< 1000 nm), and found that the absolute values of colloidal particles were larger at same pH, which was explained by the differences in the number of surface carboxyl and hydroxyl groups. The zeta potential of colloidal particles is proportional to charge density, which means that it is related to both charge quantity and specific surface area (Hou et al., 2009). Therefore, the size effect of zeta potential of Anthrosol and Calcisol colloidal particles is mainly related to the reduction of charge density caused by larger specific surface area of nanoparticles (Xu et al. 2020b).

3.4 Aggregation kinetics curves of Anthrosol and Calcisol colloids in NaCl and CaCl² solutions

 The aggregation kinetics of Anthrosol and Calcisol colloids in NaCl and CaCl² solutions are shown in Figs. S1 and S2. The aggregation process of soil colloids was divided into RLA and DLA stages. The RLA stages for Anthrosol colloids of *d* < 2 μm, *d* < 1 μm and *d* < 100 nm in NaCl solution were 0–80, 0–80 and 0–100 mmol⋅L⁻¹, respectively, during which repulsive forces existed between the particles and

 Aggregation rates of soil colloids varied with particle diameters at the same [electrolyte](javascript:;) concentration, which was particularly evident in RLA stage (Table 4). With decreasing particle diameter, the aggregation rates of Anthrosol and Calcisol colloids 349 in 50 mmol⋅L⁻¹ NaCl and 1 mmol⋅L⁻¹ CaCl₂ solutions exhibited a corresponding 350 decline. In addition, in 50 mmol⋅L⁻¹ NaCl solution, the aggregation rates for Anthrosol 351 colloids of $d < 2$ µm, $d < 1$ µm and $d < 100$ nm were about 2.17, 2.09 and 1.95 times those of Calcisol colloids, while the aggregation rates were about 1.46, 1.57 and 1.91 353 times those of Calcisol colloids in 1 mmol⋅L⁻¹ CaCl₂ solution, respectively. Therefore,

neglected.

(*Insert Figure 4 near here*)

 The CCCs of Anthrosol and Calcisol colloids increased with decreasing diameter; that is, the CCCs of Anthrosol and Calcisol colloids both showed the size effects. Hsu and Kuo (1995) demonstrated that the CCCs would generally decrease with the increasing particle diameter because smaller particles possess thicker double electric layers and higher electrolyte concentration is needed to neutralize charges on the surface, which were consistent with the results of Anthrosol and Calcisol colloids. The above explanation by Hsu and Kuo (1995) was derived from homogenous particles whose composition does not change with particle diameter. The results of this paper show that, for those two alkaline soils being such heterogeneous materials, when the organic matter contents and mineral types changed with colloidal particle diameter, the CCCs in monovalent and divalent solutions also decreased with increasing particle diameter.

 In this paper, the organic matter contents of soil nanoparticles were the highest, so the CCCs were the largest, which were 1.7 and 2.4 times of the corresponding colloidal 409 particles of $d < 2$ µm. The suspension stability of different clay minerals has been 410 reported to vary with the mineralogical structure. The CCC of illite (\approx 100 mM) in 411 NaCl solution was significantly higher than that of kaolinite (\approx 20 mM) (Jiang et al., 2012; Xu et al., 2017), indicating that the stability of illite suspensions is significantly higher than that of kaolinite. So another possible reason for the higher stability of soil nanoparticles is the increase of illite content and the decrease of kaolinite content. Therefore, the differences in organic matter contents and clay mineralogy are the fundamental reasons for the differences in colloidal suspension stability behind the size effects of Anthrosol and Calcisol colloids.

4. Conclusion

This study obtained soil colloidal fractions with three different particle sizes from

 Anthrosol and Calcisol using high-speed centrifugation and revealed the particle size effects on the soil constitutes, surface properties, and aggregation behavior of heterogeneous soil colloids. The results showed that, compared to coarse colloids (*d* < 2 μm and *d* < 1 μm), the organic carbon contents of Anthrosol and Calcisol 424 nanoparticles were higher, at 27.38 $g \cdot kg^{-1}$ and 28.31 $g \cdot kg^{-1}$, respectively, approximately twice and three times that of the coarse colloids and the bulk soils, indicating that nanoparticles exhibit a strong potential for carbon sequestration. The absolute zeta potential values of soil nanoparticles decreased with decreasing average particle diameter, indicating a reduction in charge density. Anthrosol and Calcisol 429 nanoparticles exhibited greater suspension stability in NaCl and CaCl₂ solutions. On one hand, this was due to the increased thickness of the double electric layer on the surface of soil nanoparticles, resulting in stronger repulsive forces between particles; on the other hand, it was due to the presence of more illite, which has higher CCC compared to other clay minerals. In conclusion, for such a highly heterogeneous system as soil, the size effects on soil colloidal suspension stability are strongly influenced by variations driven by essentially mineral composition. Future studies should further explore the mechanisms underlying the size effects of the typical soils on particle interactions, coagulation, and transport behaviors under environmentally relevant conditions.

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Author Contributions

Declaration of Interest Statement

 The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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Table captions

- **Table 1** The average diameters and distribution patterns of soil colloids.
- **Table 2** The physiochemical properties of soil colloids.
- **Table 3** The dominant clay minerals of soil colloidal fractions (shown in mass
- fraction, %).
- **Table 4** The aggregation rates of soil colloids.

Soil type	Colloidal fractions	Number-weighted diameter (nm)	Intensity-weighted diameter (nm)	D_{10} (nm)	D_{90} (nm)
	$d < 2 \mu m$	133.16 ± 24.28	396.81 ± 12.34	71.53	232.49
Anthrosol	$d < 1$ µm	127.84 ± 20.29	371.45 ± 11.88	67.64	219.87
	$d < 100$ nm	72.47 ± 7.04	294.10 ± 15.80	38.74	136.72
	$d < 2 \mu m$	141.23 ± 24.33	439.20 ± 12.72	78.29	244.97
Calcisol	$d < 1$ µm	131.67 ± 15.77	372.07 ± 8.23	75.84	231.64
	$d < 100$ nm	85.48 ± 16.93	312.25 ± 33.58	47.84	158.99

603 Note: D_{10} , D_{50} and D_{90} represent diameter of particles with a cumulative distribution of 10%, 50% and

604 90%, respectively.

606 **Table 2** The physiochemical properties of soil colloids

Soil type	Colloidal fractions	Yield (%)	Total carbon content $(g \cdot kg^{-1})$	Organic carbon content $(g \cdot kg^{-1})$	CaCO ₃ content $(g \cdot kg^{-1})$	Specific surface area $(m^2 \cdot g^{-1})$
Anthrosol	$d < 2 \mu m$	25.12	20.90 ± 0.30	10.90 ± 1.29	10.00	65.37
	$d \leq 1 \text{ }\mu\text{m}$	18.76	20.65 ± 0.15	10.91 ± 0.43	9.74	72.99
	$d < 100$ nm	6.32	58.25 ± 0.35	27.38 ± 0.21	30.87	45.28
Calcisol	$d < 2 \mu m$	23.17	24.00 ± 0.30	11.66 ± 0.60	12.34	49.99
	$d \leq 1 \text{ }\mu\text{m}$	16.20	22.30 ± 0.20	12.76 ± 0.11	9.54	61.88
	$d < 100$ nm	4.70	76.30 ± 0.40	28.31 ± 0.15	47.99	34.53

Soil type	Colloidal fractions		Illite Kaolinite	Chlorite	Vermiculite
	$d < 2 \mu m$	34	23		9
Anthrosol	$d \leq 1$ µm	30	22		11
	$d < 100$ nm	37	14	16	3
	$d < 2 \mu m$	24	22	29	16
Calcisol	$d \leq 1 \text{ }\mu\text{m}$	31	19	25	12
	$d < 100$ nm	37	16	17	

608 **Table 3** The dominant clay minerals of soil colloidal fractions (shown in mass fraction, %)

Figure captions

- **Fig. 1.** The photoelectron spectrum C1s and O1s peak diagram of Anthrosol and
- 614 Calcisol colloids. C1s of Anthrosol colloids, a. $d < 2 \mu m$, b. $d < 1 \mu m$, c. $d < 100 \text{ nm}$;
- 615 C1s of Calcisol colloids, d. $d < 2$ µm, e. $d < 1$ µm, f. $d < 100$ nm; O1s of Anthrosol
- 616 colloids, g. $d < 2 \mu m$, h. $d < 1 \mu m$, i. $d < 100 \text{ nm}$; O1s of Calcisol colloids, j. $d < 2 \mu m$,
- k. *d* < 1 μm, l. *d* < 100 nm.
- **Fig. 2.** The zeta potential of Anthrosol (a) and Calcisol (b) colloids of *d* < 2 μm, < 1 μm,
- and < 100 nm at different pH
- 620 **Fig. 3.** The CCCs of Anthrosol (a) and Calcisol (b) colloids of $d < 2 \mu m, < 1 \mu m$, and $<$
- 100 nm in NaCl solution
- **Fig. 4.** The CCC of Anthrosol (a) and Calcisol (b) colloids of $d < 2 \mu m, \leq 1 \mu m$, and \leq
- 100 nm in CaCl2 solution

 Fig. 1 The photoelectron spectrum C1s and O1s peak diagram of Anthrosol and Calcisol 626 colloids. C1s of Anthrosol colloids, (a). $d < 2 \mu m$, (b). $d < 1 \mu m$, (c). $d < 100 \text{ nm}$; C1s of Calcisol colloids, (d). *d* < 2 μm, (e). *d* < 1 μm, (f). *d* < 100 nm; O1s of Anthrosol colloids, (g). *d* < 2 μm, (h). *d* < 1 μm, (i). *d* < 100 nm; O1s of Calcisol colloids, (j). *d* < 2 μm, (k). *d* < 1 μm, 629 (1). $d < 100$ nm.

 Fig. 2 The zeta potential of Anthrosol (a) and Calcisol (b) colloids of *d* < 2 μm, < 1 μm, and < 100 nm at different pH

in CaCl₂ solution.