

1 **Exploring the Crucial Role of Atmospheric Carbonyl**
2 **Compounds in Regional Ozone heavy Pollution: Insights**
3 **from Intensive Field Observations and Observation-**
4 **based modelling in the Chengdu Plain Urban**
5 **Agglomeration, China**

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20 **Abstract.** Gaseous carbonyl compounds serve as crucial precursors and intermediates
21 in atmospheric photochemical reactions, significantly contributing to ambient ozone
22 formation. To investigate the impact of gaseous carbonyls on regional ozone pollution,
23 simultaneous field observations and observation-based modelling of ambient carbonyls
24 were conducted at nine sites within the Chengdu Plain Urban Agglomeration (CPUA),
25 China during August 4-18, 2019, when three episodes of regional heavy ozone pollution
26 occurred across eight cities within CPUA. Throughout the study, the total mixing ratios
27 of 15 carbonyls ranged from 10.70 \pm 4.16 to 35.18 \pm 13.37 ppbv, in which formaldehyde
28 (48.1%), acetone (19.9%), and acetaldehyde (17.5%) were most abundant within the
29 CPUA. Ambient levels of carbonyls and ozone showed some positive correlations in
30 space (especially pronounced around Chengdu in both northern and southern directions)

31 and in diurnal variations with higher concentrations of carbonyls during ozone pollution
32 episodes. Photochemical reactivity analysis emphasized the significant contributions of
33 carbonyls, especially formaldehyde and acetaldehyde, to ozone formation. Sites with
34 higher average ozone concentrations during observations were mainly in the VOCs-
35 limited regime, while others were in the transitional regime. Local primary emissions,
36 mutual air transportation among cities within the CPUA and photochemical secondary
37 processes were recognized to contribute significantly to the production or the
38 contamination of carbonyls in ambient air, with alkenes and alkanes being important
39 precursors of secondary carbonyls. This study highlights the pivotal role of carbonyls
40 in heavy ozone pollution within the CPUA, China, providing valuable scientific insights
41 to guide the development of effective countermeasures for regional ozone pollution
42 control in the future.

43 **Keywords:** Gaseous Carbonyls; Ozone Heavy Pollution; Pollution Characteristics;
44 Atmospheric Photochemical Reactivity; Source Analysis; The Chengdu Plain Urban
45 Agglomeration, China

46 1. Introduction

47 Atmospheric carbonyl compounds play a pivotal role in tropospheric chemistry,
48 acting as crucial precursors to both ozone (O₃) and secondary organic aerosols (SOA),
49 a fact recognized for decades (Altshuller, 1993; Grosjean and Seinfeld, 1989). Their
50 importance has been confirmed by numerous studies over the years(Guo et al., 2004;
51 Hallquist et al., 2009; Wang et al., 2020; Ye et al., 2021; Coggon et al., 2019),
52 highlighting their significant contribution to atmospheric photochemistry and air
53 pollution. Over the past two decades, severe air pollution in China has driven substantial
54 research efforts to understand the contributions of carbonyl compounds to these
55 environmental challenges. Studies have shown that photolysis of carbonyl compounds
56 is a major source of RO_x radicals (Grosjean and Seinfeld, 1989; Zhang et al., 2016).
57 These compounds can be photolyzed and react with OH radicals to form a large number
58 of HO₂ and RO₂ radicals, which increase the atmospheric oxidation capacity and

删除了: The ozone formation sensitivity for sites experiencing severe ozone pollution were classified as VOCs-limited regime, while others were categorized as transitional regime.

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删除了: Atmospheric carbonyl compounds are an important group of volatile organic compounds (VOCs) and are common precursors of ozone and secondary organic aerosols. this category of VOCs has become a research hotspot in the field of atmospheric environmental science due to their key role in tropospheric chemistry(Guo et al., 2004) . form a large number of HO₂ and RO₂ radicals.

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74 participate in the NO_x photochemical cycle, leading to ozone formation (Zhang et al.,
 75 2016; Meng et al., 2017). Additionally, dialdehydes such as glyoxal and methylglyoxal
 76 undergo heterogeneous reactions with aqueous particulate matter, rapidly forming SOA
 77 (Lou et al., 2010; Xue et al., 2016; Yuan et al., 2012). Ambient carbonyl compounds
 78 not only affect the environment but also pose direct health risks to humans. They can
 79 harm ecosystems through deposition and adsorption processes (Yang et al., 2018). They
 80 also pose direct health risks to humans, including sensitization, carcinogenesis, and
 81 mutagenicity (Fuchs et al., 2017).

82 Recent research has increasingly focused on understanding the spatial and
 83 temporal variability of carbonyl compounds in highly polluted regions, particularly in
 84 China, where rapid industrialization has led to severe air quality challenges. Xue et al.
 85 (2013) and Duan et al. (2012) reported typical ambient concentrations of carbonyl
 86 compounds ranging from a few $\mu\text{g}\cdot\text{m}^{-3}$ to tens of $\mu\text{g}\cdot\text{m}^{-3}$ in urban areas, depending on
 87 the specific compounds and regions studied. For example, formaldehyde concentrations
 88 in highly polluted areas can exceed $10\ \mu\text{g}\cdot\text{m}^{-3}$. Shen et al. (2013) and Fu et al. (2008)
 89 observed significant diurnal variation, with higher concentrations of carbonyl
 90 compounds during the daytime, particularly in the afternoon, driven by photochemical
 91 production. Concentrations can increase by as much as 50-100% during peak sunlight
 92 hours compared to nighttime levels. Pang and Mu (2006) and Rao et al. (2016)
 93 identified key sources of carbonyl compounds, including vehicular emissions, industrial
 94 activities, and secondary formation from VOC oxidation in the atmosphere. In urban
 95 environments, vehicular emissions are often a dominant primary source, while
 96 secondary formation contributes significantly during daytime due to photochemical
 97 processes. The results highlight severe and spatiotemporal variations of carbonyl
 98 pollution in China. High levels are found mainly in the North China Plain(NCP), the
 99 Yangtze River Delta(YRD), and the Pearl River Delta(PRD)(Duan et al., 2008; Shao et
 100 al., 2009; Tan et al., 2018; Wang et al., 2018; Xue et al., 2014, 2013; Yang et al., 2017).
 101 Urban areas generally exhibit higher carbonyl levels than suburban and rural areas due

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删除了: In addition, ambient carbonyl compounds could also be removed through wet and dry deposition, and soil and ocean adsorption, thereby adversely affecting the ecological environment

下移了 [1]: The sources of Atmospheric carbonyl compounds can be divided into primary source and secondary source. Primary sources mainly include incomplete combustion of fossil fuels and biomass, industrial emissions, catering business emissions, and plant releases, etc; And secondary source is mainly from the secondary formation of atmospheric photochemical oxidation of VOCs (Xue et al., 2013).

上移了 [2]: Atmospheric carbonyl compounds could be photolyzed and react with OH radicals to form a large number of HO₂ and RO₂ radicals. Studies have shown that photolysis of carbonyl compounds is a major source of RO_x radicals (Guenther et al., 2012; Y. Zhang et al., 2016). The free radicals formed from carbonyl compounds can not only increase atmospheric oxidation, but also participate in the

删除了: In addition, toxic and harmful atmospheric carbonyl compounds could have potential hazards such as ...

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207 to human activities(Xue et al., 2013). Despite the progress made, significant gaps
208 remain in understanding the spatiotemporal distribution and source apportionment of
209 carbonyl compounds, particularly in urban agglomerations. Existing research has
210 primarily focused on urban areas in rapidly developing regions like the NCP, YRD, and
211 PRD. Moreover, studies have often emphasized the overall role of VOCs in ozone
212 pollution, with less attention given to specific carbonyl compounds and their individual
213 contributions to atmospheric oxidation capacity and ozone formation (Meng et al.,
214 2017).

215 Monitoring carbonyl compounds in the atmosphere is challenging due to their
216 typically low concentrations (ppt-ppb levels), necessitating highly sensitive analytical
217 methods. The diversity of carbonyl compounds, including multiple isomers, requires
218 highly selective analytical techniques for differentiation. Current measurement
219 technologies limit our understanding of the spatiotemporal distribution of carbonyl
220 compounds, affecting the accurate assessment of their environmental behavior, sources,
221 and transport (Xue et al., 2013; Sahu and Saxena, 2015). While numerous studies have
222 explored the role of carbonyl compounds in ozone production, many focus on general
223 mechanisms rather than specific compounds or regional variations (Atkinson and Arey,
224 2003; Monks et al., 2015).

225 Atmospheric carbonyl compounds originate from both primary and secondary
226 sources (Pang and Mu, 2006; Rao et al., 2016). Primary sources include the incomplete
227 combustion of fossil fuels and biomass, industrial emissions, emissions from the
228 catering industry, and releases from plants. Secondary sources arise from the
229 atmospheric photochemical oxidation of VOCs (Xue et al., 2013), particularly
230 alkenes, aromatics, and isoprene, which typically dominate the secondary formation of
231 carbonyls. However, distinguishing between primary and secondary contributions
232 remains challenging. Existing source apportionment methods, such as characteristic
233 species ratios and multiple linear regression, often lack the resolution to differentiate
234 these sources accurately, especially for non-vehicular emissions and secondary

上移了 [3]: (Xue et al., 2013; Duan et al., 2012).

上移了 [5]: (Duan et al., 2008; Shao et al., 2009; Tan et al., 2018; Wang et al., 2018; Xue et al., 2014, 2013; Yang et al., 2017)

上移了 [4]: (Shen et al., 2013; Fu et al., 2008)

删除了: At present, there are many studies on the concentration, diurnal variation and sources of carbonyl compounds worldwide. Studies show that compared with other cities in the world, the concentrations of carbonyl compounds in Chinese cities are currently at relatively higher level (Xue et al., 2013; Duan et al., 2012). Areas with high ambient carbonyl compounds are mainly in the North China Plain, the Yangtze River Delta and the Pearl River Delta in China, among which, the pollution of carbonyl compounds in the Pearl River Delta is relatively light

删除了: . In the meantime, carbonyl compounds levels in urban areas are generally higher than those in suburban and rural areas in China due to anthropogenic sources (Xue et al., 2013). The diurnal variation of carbonyl compounds is usually consistent with the diurnal variation of ozone, but there are some differences in the concentrations of carbonyl compounds among different regions due to the effects of anthropogenic emissions and meteorological conditions on the diurnal variation (Shen et al., 2013; Fu et al., 2008).

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260 formation. The limitations of these methods underscore the need for more advanced
261 approaches to better quantify the secondary formation mechanisms of carbonyl
262 compounds and their regional impact on ozone formation. Despite significant
263 advancements in studying atmospheric carbonyls, key gaps remain in understanding
264 their precise spatiotemporal distribution and source apportionment. Specifically, there
265 is a need for studies that examine how carbonyls vary across different environments—
266 urban, suburban, and rural—and during varying pollution events. Without such targeted
267 analysis, our understanding of the behavior of carbonyl compounds and their
268 contribution to ozone pollution remains incomplete, particularly in regions
269 experiencing severe pollution.

270 In this context, this study focuses on atmospheric carbonyl compounds and their
271 roles in photochemical pollution within the Chengdu Plain Urban Agglomeration
272 (CPUA) of China. The CPUA includes eight cities: Chengdu, Mianyang, Deyang,
273 Leshan, Meishan, Yaan, Suining, and Ziyang. This region has a developed economy
274 and a high degree of internationalization. The CPUA is located on the western edge of
275 the Sichuan Basin, surrounded by mountain ranges, which easily block airflow. The
276 unique climatic environment of the CPUA features low wind speeds year-round, high
277 frequency of static winds, short hours of sunshine, frequent winter inversions, and a
278 pronounced heat island effect in summer. These climatic characteristics significantly
279 impact the variations in air pollutant concentrations, making the region prone to ozone
280 pollution in summer and haze pollution in winter. (Li et al., 2013; Hu et al., 2017; Zhang
281 et al., 2010). Although previous studies have shown that ozone formation in urban
282 Chengdu is primarily VOCs-limited, (Tan et al., 2018), with aromatic hydrocarbons and
283 alkenes contributing significantly to ozone generation in summer (Xu et al., 2020),
284 these studies mainly focus on single cities and overall VOCs. There is limited
285 understanding of the distribution, sources, and specific roles of carbonyl compounds
286 across the entire CPUA and their contributions to regional ozone pollution and mutual
287 air transport mechanisms.

删除了: At present, the analysis methods for the sources of carbonyl compounds mainly include characteristic species ratio, source tracer proportion, multiple linear regression, parameterization method based on photochemical age, and acceptor model. There are usually differences in the source classifications obtained in different studies, due to the different methods of source analysis employed. But generally, for formaldehyde and acetaldehyde with high reactivities, their sources vary significantly across studies; and for acetone with low reactivity, the sources are relatively similar. However, current knowledge of the sources of carbonyl compounds is far from providing scientific support for air quality improvements. Meanwhile, most of the research on their health effects has focused on indoor carbonyl compounds (Yuan et al., 2013). Research on the health effects of atmospheric carbonyl compounds should be strengthened.

删除了: The Chengdu Plain Urban Agglomeration (CPUA) of China is centered

删除了: with Chengdu and includes eight cities at and above the prefecture level, including CDHKY, MY, DY, LS, MS, YA, SN and ZY, with a developed economy and a high degree of internationalization. The CPUA is located on the western edge of the Sichuan Basin, surrounded by mountain ranges, thus the airflow is easily blocked by the high

删除了: In the autumn of 2016, Peking University and Chengdu Academy of Environmental Sciences carried out a comprehensive observation experiment of photochemical pollution at one urban site and three suburban sites, and

删除了: based on the observation-based model (OBM)

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删除了: analyzed the online observation data of VOCs in Chengdu from June to September 2019 and found that the aromatic hydrocarbons and alkenes in VOCs contributed greatly to the formation of ozone in Chengdu in summer.

删除了: However, the previous studies on the causes and sources of ozone pollution in the CPUA mainly focused on the contribution and sources of ozone formation by VOCs other than carbonyl compounds in Chengdu urban area. And

355 To address these research gaps, this study involves an intensive field observation
 356 experiment conducted by the Sichuan Academy of Environmental Sciences, Peking
 357 University, Sichuan University and Chinese Academy of Environmental Sciences.
 358 Atmospheric carbonyl compounds were observed at nine sites in eight cities within the
 359 CPUA for 15 days during a period of heavy ozone pollution in August 2019. Samples
 360 were analyzed using 2,4-dinitrophenylhydrazine solid phase adsorption/high
 361 performance liquid chromatography (HPLC). The study aims to characterize the
 362 atmospheric carbonyl compounds in the CPUA, assess their influence on
 363 photochemical pollution, identify key carbonyl compounds that may play crucial roles
 364 in heavy ozone pollution in the CPUA, and evaluate the contribution of primary
 365 emissions, air pollution transport, and secondary generation to key carbonyl compounds
 366 through a combination of multivariate linear regression modeling and Observation-
 367 Based Modeling (OBM). This research aims to provide technical support for controlling
 368 carbonyl compounds pollution in the CPUA and to reduce their contributions to ozone
 369 pollution.

370 **2. Materials and methods**

371 **2.1 Observation Sites Profile**

372 In this study, a total of 9 off-line sampling sites for atmospheric carbonyl
 373 compounds were set up in 8 cities in the CPUA from August 4th to 18th, 2019(table S1).
 374 Considering that this study focused on the pollution characterization of carbonyl
 375 compounds in urban areas, one urban site was selected in each city. In addition, in order
 376 to compare and study the pollution characteristics of carbonyl compounds in the
 377 suburbs, a suburban site was set up in Xinjia County, Chengdu. For the selection of
 378 urban sites in each city, priority is given to those choices of set-up in the vicinity of the
 379 state-controlled site, and the perimeter of the sites should be open, unobstructed and no
 380 obvious pollution sources, with convenient transportation and power supply. The
 381 distribution of specific sites is shown in Fig. 1.

删除了: In order to study the multi-dimensional causes of atmospheric ozone pollution and its management and prevention system in the CPUA, Sichuan Academy of Environmental Sciences

删除了: jointly carried out an intensive field observation experiment on the formation mechanisms of photochemical pollution in summer in the CPUA.

删除了: In this study, atmospheric carbonyl compounds were observed at 9 sites in 8 cities in the CPUA for 15 days during a period of heavy ozone pollution in August 2019, and the samples were analyzed by

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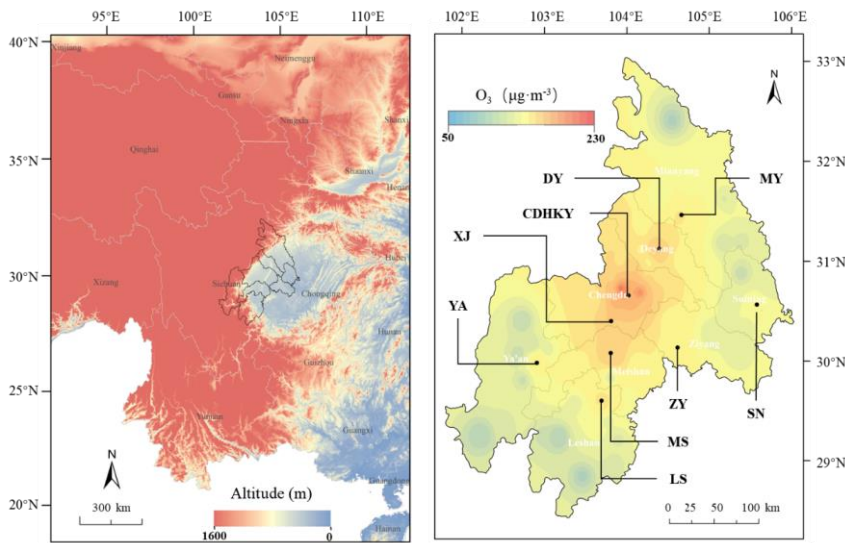
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414 Ozone concentrations were measured using the UV absorption method with a
 415 Thermo O₃ analyzer (Model 49i), with data sourced from national control stations near
 416 each sampling site. Nitrogen dioxide (NO₂) was measured by chemiluminescence
 417 following chemical conversion to nitric oxide (NO) using a molybdenum catalyst;
 418 however, this method is known to have interferences from other NO_x species. Carbon
 419 monoxide (CO) was measured via infrared absorption with a Thermo instrument
 420 (Model 20). All Thermo instruments were carefully maintained and calibrated daily at
 421 01:00 to ensure measurement accuracy. Measurements for ozone, NO₂, and CO were
 422 collected with a time resolution of one hour. Simultaneously, meteorological
 423 parameters—temperature, relative humidity (RH), wind speed, and direction—were
 424 recorded at each observation site using an automatic weather station (PC-4, JZYG,
 425 China), also at a one-hour resolution.

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426 **Figure 1.** Distribution of sampling sites. The left panel shows the elevation map of the Sichuan
 427 Basin, highlighting the geographical features of the region, with elevation data sourced from the
 428 Geospatial Data Cloud (<https://www.gscloud.cn/#page1/2>). The right panel presents the spatial
 429 distribution of ozone concentrations in the CUA during the observation period (August 4–18,
 430 2019), with ozone data obtained from national control stations near each sampling site. Black dots
 431 represent the locations of the sampling sites, labeled as follows: MY (Mianyang), DY (Deyang),
 432 CDHKY (Chengdu Environmental Science Research Institute), XJ (Xinjin), SN (Suining), ZY
 433 LS

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436 (Ziyang), MS (Meishan), YA (Ya'an), and LS (Leshan). The color bar in the top left corner
437 corresponds to interpolated ozone concentrations, with each color representing a concentration
438 gradient.

439 2.2 Samples Collection

440 The sampling of atmospheric carbonyl compounds mainly referred to the TO-11A
441 standard of the United States Environmental Protection Agency (US EPA) and the
442 Chinese environmental protection standard HJ 683-2014 High Performance Liquid
443 Chromatography Method for the Determination of Atmospheric Carbonyl Compounds,
444 and the sampling was carried out by using silica gel sampling tubes (IC-DN3501 from
445 Tianjin Bonna-Agela) coated with DNPH (2,4-dinitrophenylhydrazine). In this study,
446 an automatic sampler for carbonyl compounds (Zhang et al., 2019) was used to
447 continuously collect atmospheric carbonyl compounds. From August 4th to 18th, 2019,
448 air samples were collected every 2 hours with a sampling flow rate of 0.8 L·min⁻¹. In
449 addition, in order to prevent the impact of ozone and rainwater in the atmospheric air
450 on sample collection, a potassium iodide ozone removal column (KI 140 from Tianjin
451 Bonna-Agela) was installed and a water removal agent made by ourselves (Bao et al.,
452 2022; Wang et al., 2020) was added at the front end of the sample tube. Two blank
453 samples were collected before and after the sampling, and blank samples were also
454 collected for different batches of sampling tubes. The samples were frozen at -18°C and
455 analyzed within one month.

456 Atmospheric VOCs were sampled using SUMMA tanks, stainless steel tanks with
457 electropolished and silanized inner walls, manufactured by Entech in the United States,
458 with a sampling volume of 3.2 liters. The sampling was controlled by a constant current
459 integral sampler to sampling for an average of 1 hour. From August 4th to 18th, 2019,
460 two VOCs samples were collected each day at each site, at 8:00-9:00 and 14:00-15:00
461 (no samples were taken under special weather conditions, such as rain). On August 11th,
462 12th and 16th, six samples were collected per day to capture diurnal variations under
463 ozone pollution events, at the following times: 8:00-9:00, 10:00-11:00, 12:00-13:00,
464 14:00-15:00, 16:00-17:00, and 18:00-19:00.

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475 2.3 Samples Analysis

476 The carbonyl compounds samples were qualitatively and quantitatively analyzed
477 by using High Performance Liquid Chromatography (HPLC) (LC-20AD, Shimadzu,
478 Japan) and an ultraviolet detector (SPD-20A, Shimadzu, Japan), mainly based on the
479 US EPA TO-11A standard and the Chinese HJ 683-2014 standard. The DNPH sampling
480 column after sampling was slowly eluted into a volumetric flask using acetonitrile
481 (chromatographically pure, Thermo Fisher Scientific China) to 5.0 mL. Then 1.5 mL
482 sample was taken into an HPLC sample bottle, and sealed and stored in a refrigerator
483 at <4 °C to complete the pre-treatment. Prior to sample analysis, a standard solution of
484 the concentration gradient was prepared using TO-11A standard solution (Supelco,
485 USA) and used as the external standard. The correlation coefficient (R^2) of the standard
486 curve was greater than 0.995. The limit of detection of the device was 0.56~5.57
487 $\text{ng}\cdot\text{mL}^{-1}$, and the limit of quantification was 1.87~18.56 $\text{ng}\cdot\text{mL}^{-1}$ (Table S2). Then 20
488 μL of the pretreated sample was extracted through the autosampler and injected into the
489 HPLC/UV system, detected by a UV detector with a wavelength of 360 nm, qualified
490 by retention time value, quantified by peak area value, and the qualitative and
491 quantitative analysis data of carbonyl compounds were obtained after conversion. The
492 HPLC conditions referred to Chinese environmental protection standard HJ 683-2014:
493 binary gradient washing was performed using acetonitrile and water, 60% acetonitrile
494 was held for 20 mins, acetonitrile was increased linearly from 60% to 100% within 20-
495 30 mins, and acetonitrile was reduced to 60% again within 30-32 mins and held for 8
496 mins; the column oven was kept at 40 °C.

497 The atmospheric VOCs were analyzed using the TO-14 and TO-15 methods,
498 which are recommended by the US EPA. These methods involve frozen
499 preconcentration coupled with gas chromatography and mass spectrometry (GC-MS).
500 TO-15 is a method for detecting and quantifying a wide range of VOCs from air samples.
501 The VOCs were pre-concentrated by the Entech7100 system at a low temperature, then
502 quantified by an Agilent GC-MS. During the sample analysis, four internal standard

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510 gases (bromochloromethane, 1,4-difluorobenzene, chlorobenzene-d5, and 4-
511 bromofluorobenzene) were used. A multi-point calibration curve was created using a
512 standard gas containing 118 VOCs, including PAMS compounds, TO-15 target analytes,
513 and carbonyl compounds. PAMS (Photochemical Assessment Monitoring Stations)
514 compounds are a subset of hydrocarbons known to contribute to ozone formation, such
515 as ethane, ethylene, propane, and others."

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删除了: Ambient levels comparison

516 2.4 Data Analysis

517 2.4.1 Ozone pollution assessment criteria

518 According to the Technical Regulation on Ambient Air Quality Index (on trial),
519 National Environmental Protection Standard of the People's Republic of China HJ
520 633—2012, days with an ozone pollution index (IAQI) of 100 or higher during the
521 observation period were designated as pollution days, while days with an IAQI below
522 100 were considered clean days. This study compared the pollution characteristics of
523 carbonyl compounds between pollution days and clean days. Additionally, the
524 concentrations of formaldehyde, acetaldehyde, and acetone observed during the
525 summer of 2009-2013 in economically developed and industrialized areas such as
526 Beijing, Shanghai, and Guangzhou in China, as well as locations in South America
527 (Brazil), Asia (Thailand), Europe (France), and North America (United States), were
528 selected and compared.

529 2.4.2 Ozone formation sensitivity

530 Previous studies have shown that the formaldehyde to NO₂ ratio (FNR) can be
531 used to determine the sensitivity of O₃-NO_x-VOCs (Schroeder et al., 2017; Tonnesen
532 and Dennis, 2000; Vermeuel et al., 2019). Most studies used satellite remote sensing-
533 based FNR, but the FNR column concentration ratios inverted by satellite remote
534 sensing mainly represented the average photochemical of the troposphere, and the
535 concentration distributions of HCHO and NO₂ in the vertical direction were

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552 inconsistent (Hong et al., 2022; Schroeder et al., 2017). So, there is a large uncertainty
553 to develop ground-level ozone pollution prevention and control measures. In this study,
554 sensitivity analysis of ground-level ozone formation was carried out based on the ratio
555 of ground-level HCHO to NO₂ during the observation period at the 9 sites of 8 cities in
556 the CPUA. FNR < 0.55±0.16 and FNR > 1.0±0.3 were defined to VOCs-limited and
557 NO_x-limited, respectively, and FNR ratio ranged from 0.55±0.16 to 1.0±0.3 defined to
558 NO_x and VOCs co-limited (Liu et al., 2021; Zhang et al., 2022).

559 2.4.3 Exploration of Secondary Formation Mechanisms

删除了: Secondary formation mechanism investigation

560 (1) Atmospheric chemical reactivity

561 In this study, the contribution of atmospheric chemical reactivity of carbonyl
562 compounds to ozone formation was evaluated using the OH free radical consumption
563 rate (L_{OH}) and ozone formation potential (OFP):

$$564 L_{OH} = [OVOC]_i \times k_i(OH) \quad (1)$$

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565 Where, [OVOC]_i was the observed concentration of the ith (i=1 to n) carbonyl
566 compound, in molecule·cm⁻³; k_i(OH) was the rate constants of the ith carbonyl
567 compound reacting with OH radicals, in cm³·(molecule·s)⁻¹. The unit of L_{OH} is s⁻¹,
568 representing the rate of OH radical consumption. The selected k_i(OH) values were from
569 literature (Atkinson and Arey, 2003).

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$$570 OFP = MIR_i \times [OVOC]_i \quad (2)$$

571 Where, MIR was the maximum incremental reactivity of the ith carbonyl
572 compound, in g O₃·(g VOC)⁻¹ (grams of ozone formed per gram of volatile organic
573 compound), and the MIR values of each species were from California Code of
574 Regulations (<https://govt.westlaw.com>); [OVOC]_i was the mass concentration of the ith
575 carbonyl compound, in μg·m⁻³. The unit of OFP is μg·m⁻³, representing the potential
576 ozone formation.

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577 (2) Observation-based model (OBM)

586 The Observation-Based Model (OBM) is a box model that uses actual
587 observational data to evaluate the sensitivity of secondary pollutant formation
588 mechanisms to their precursor emissions. By constraining the model with atmospheric
589 observation data, typical secondary pollutants and parameters such as NO_x, SO₂, CO,
590 VOCs, temperature, humidity, pressure, and JNO₂ are input into the model as hourly
591 observational data to calculate the chemical formation and consumption of secondary
592 pollutants and free radicals. In this study, the OBM model used the Master Chemical
593 Mechanism (MCM) (v3.3.1, mcm.leeds.ac.uk), which is a nearly detailed chemical
594 mechanism that describes the chemical processes of 143 VOC species from emission
595 to degradation in the atmosphere, including approximately 6,700 species and 17,000
596 inorganic and organic reactions. The MCM chemical mechanism can simulate
597 atmospheric photochemical reaction processes under near-real conditions and calculate
598 the concentrations of highly reactive species, quantifying the reaction rates of all
599 species involved.

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600 Relative Incremental Reactivity (RIR) was first used by Cardelino and Chameides
601 (1995) to simulate the response of ozone to precursor changes through scenario tests
602 using box model calculations. RIR was calculated by assuming that the concentration
603 of a given carbonyl compound precursor decreased by a certain proportion could cause
604 the change of the concentration of the carbonyl compound, so as to further judge the
605 effect of VOCs on the formation of carbonyl compounds. Combining the concentrations
606 and activity levels of 15 carbonyl compounds during the observation period, this study
607 focused on formaldehyde, acetaldehyde, and acetone as the primary research targets.
608 The impacts of various AVOCs (anthropogenic VOCs), including alkanes, alkenes,
609 alkynes, and aromatic hydrocarbons, as well as BVOCs (biogenic VOCs) like isoprene,
610 on the formation of formaldehyde, acetaldehyde, and acetone were assessed using
611 observation-based OBM classification. Specific species of anthropogenic source VOCs
612 (alkanes, alkenes, alkynes, and aromatic hydrocarbons) and biogenic VOCs (isoprene)
613 are detailed in Table S3.

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616 VOCs observations, conventional gases (NO₂, CO and SO₂) and meteorological
617 parameters (temperature, relative humidity and pressure) were imputed into the model.
618 It was assumed that the pollutants are well mixed. Under the constraints of the measured
619 hourly concentration data of pollutants, the atmospheric chemical process was
620 simulated to obtain the source-effect relationship of the measured pollutants. By
621 assuming the reduction of the source effect, the RIRs of different carbonyl compounds
622 precursors were calculated, and the sensitivities of carbonyl compounds to different
623 pollutants were obtained, and then the secondary formation mechanism of carbonyl
624 compounds was determined. The formula to calculate the RIR is as follows:

$$625 \quad RIR(X) = \left[\frac{\Delta P_Y(X)/P_Y(X)}{\Delta S(X)/S(X)} \right] \quad (3)$$

$$626 \quad P_Y = Y_{\text{net formation}} - Y_{\text{net consumption}} \quad (4)$$

627 Where X was a specific species; P_Y(X) was the net formation rate of species y;
628 S(X) was the total amount of emissions of species X in a certain period, i.e., the source
629 effect of species X. ΔS(X) was the change in total emissions of X caused by the
630 hypothetical change in source effect, ΔP_Y(X) was the change in P_Y(X) after the change
631 in source effect S(X), and RIR(X) was the relative incremental reactivity of species X.
632 The species Y in this study were formaldehyde, acetaldehyde and acetone, respectively,
633 and pollutant X was reduced by 20%.

634 The absolute RIR of the precursor reflects the sensitivity of carbonyl compounds
635 formation to the precursor. The higher the absolute RIR, the more sensitive the carbonyl
636 compounds formation to the precursor. A positive RIR value indicates that reducing the
637 species can reduce the formation rate of species Y, and a negative RIR value indicates
638 that reducing the species can increase the formation rate of species Y.

639 2.4.4 Sources Analysis

640 (1) Multi-linear regression model

641 There is a good correlation between concentrations of compounds of the same or

642 similar source in the atmosphere. Based on this property, it was assumed that the
 643 primary and secondary sources of carbonyl compounds were linearly correlated with
 644 the selected tracers, and then a quantitative source model was established by multiple
 645 linear statistical regression analysis (Kanjanasiranont et al., 2016a; Li et al., 2010; Ling
 646 et al., 2017; Luecken et al., 2012; Lui et al., 2017; Wang et al., 2017). In general, CO is
 647 the marker product of typical anthropogenic combustion source emissions, mainly from
 648 vehicle exhaust emissions and coal combustion. Ozone, as an indicator of
 649 photochemical smog, is a typical secondary formation pollutant. In this study, CO and
 650 ozone were selected as the tracers of primary source and secondary source of carbonyl
 651 compounds, respectively. The formula is as follows:

$$652 \quad [carbonyl] = \beta_0 + \beta_1[CO] + \beta_2[O_3] \quad (6)$$

653 Where [carbonyl], [CO] and [O₃] represented the observed mixing ratios of
 654 carbonyl compounds, CO and ozone, respectively, in ppbv. β_0 , β_1 and β_2 were
 655 coefficients obtained by multiple linear regression fitting model, in ppbv/ppbv. β_0
 656 represented the background concentration of a given carbonyl compound, β_1
 657 represented the emission ratio of the carbonyl compound relative to CO. $\beta_1[CO]$ and
 658 $\beta_2[O_3]$ represented the concentrations of carbonyl compound in primary emission and
 659 secondary formation, respectively, in ppbv.

660 In addition, the relative contribution of primary emissions, secondary formation
 661 and background concentrations of carbonyl compounds can be calculated using the
 662 following formula:

$$663 \quad P_{primary} = \frac{\beta_1[CO]_i}{(\beta_0 + \beta_1[CO]_i + \beta_2[O_3]_i)} \times 100\% \quad (7)$$

$$664 \quad P_{secondary} = \frac{\beta_2[O_3]_i}{(\beta_0 + \beta_1[CO]_i + \beta_2[O_3]_i)} \times 100\% \quad (8)$$

$$665 \quad P_{background} = \frac{\beta_0}{(\beta_0 + \beta_1[CO]_i + \beta_2[O_3]_i)} \times 100\% \quad (9)$$

666 Where, $P_{primary}$ represented the contribution of the primary emission of a given
 667 carbonyl compound, %; $P_{secondary}$ represented the contribution of the secondary
 668 formation of the carbonyl compound species, %; $P_{background}$ represented the contribution

669 of the carbonyl compounds species from sources other than primary emissions and
670 secondary formation, %.

671 (2) Backward trajectory model

672 The effects of long-distance air mass transport on the pollution of carbonyl
673 compounds in the CPUA were studied using MeteoInfo software and TrajStat plug-in
674 (<http://www.meteothink.org/downloads/index.html>). In this model, meteorological
675 data were relevant meteorological data from the global data assimilation system (GDAS)
676 database (<ftp://arlftp.arl.hq.noaa.gov/pub/archives/gdasl>). A trajectory simulation height
677 of 500 m was selected. The duration of backward trajectory was 48 h. The daily start
678 time was 00:00 UTC. The analog frequency was 2 h. The backward trajectory diagram
679 was calculated. Meanwhile, the clustering method in TrajStat software and the
680 Euclidean distance algorithm were used to cluster the airflow trajectory to the CPUA.
681 And then the statistical analysis was carried out in combination with the corresponding
682 pollutant mass concentration characteristics.

683 3. Results and Discussion

684 3.1 Overview of air quality during observation period

685 Due to the influence of cooling and precipitation caused by cold air intrusion, the
686 early observation period (from August 4th to 6th, 2019) in the Chengdu Plain Urban
687 Agglomeration (CPUA) experienced slightly lower temperatures (25.1°C) and higher
688 humidity (87.6%). ~~These conditions were unfavorable for ozone formation. Although
689 ozone itself is not easily removed by rain, precipitation reduces ozone pollution by
690 washing away its precursors, such as nitrogen oxides (NOx) and volatile organic
691 compounds (VOCs), decreasing sunlight exposure, and enhancing atmospheric
692 dispersion. However, as temperatures increased and humidity dropped in the
693 subsequent days, more favorable conditions for ozone formation emerged,~~ leading to
694 heavy and persistent regional ozone pollution in the CPUA. By August 12th, the mean

删除了: These conditions were not conducive to ozone formation. However, as

删除了: temperatures rose and humidity decreased thereafter, favorable conditions for ozone generation emerged

699 temperature had gradually increased to 29.1°C, while it averaged 27.7°C from August
700 13th to 14th. During this time, cumulative precipitation reached 975 mm, resulting in
701 temporary alleviation of ozone pollution. Subsequently, temperatures rose again from
702 August 15th to 18th, with the mean temperature persisting above 28.4°C for several days,
703 accompanied by a decrease in humidity to a minimum of 64.8% on August 17th. Overall,
704 during the observation period (from August 4th, 2019, 0:00 to August 18th, 2019, 24:00),
705 three episodes of severe ozone pollution occurred, namely EP1 (August 7th to 9th), EP2
706 (August 10th to 13th), and EP3 (August 15th to 18th), as depicted in Fig. 2.

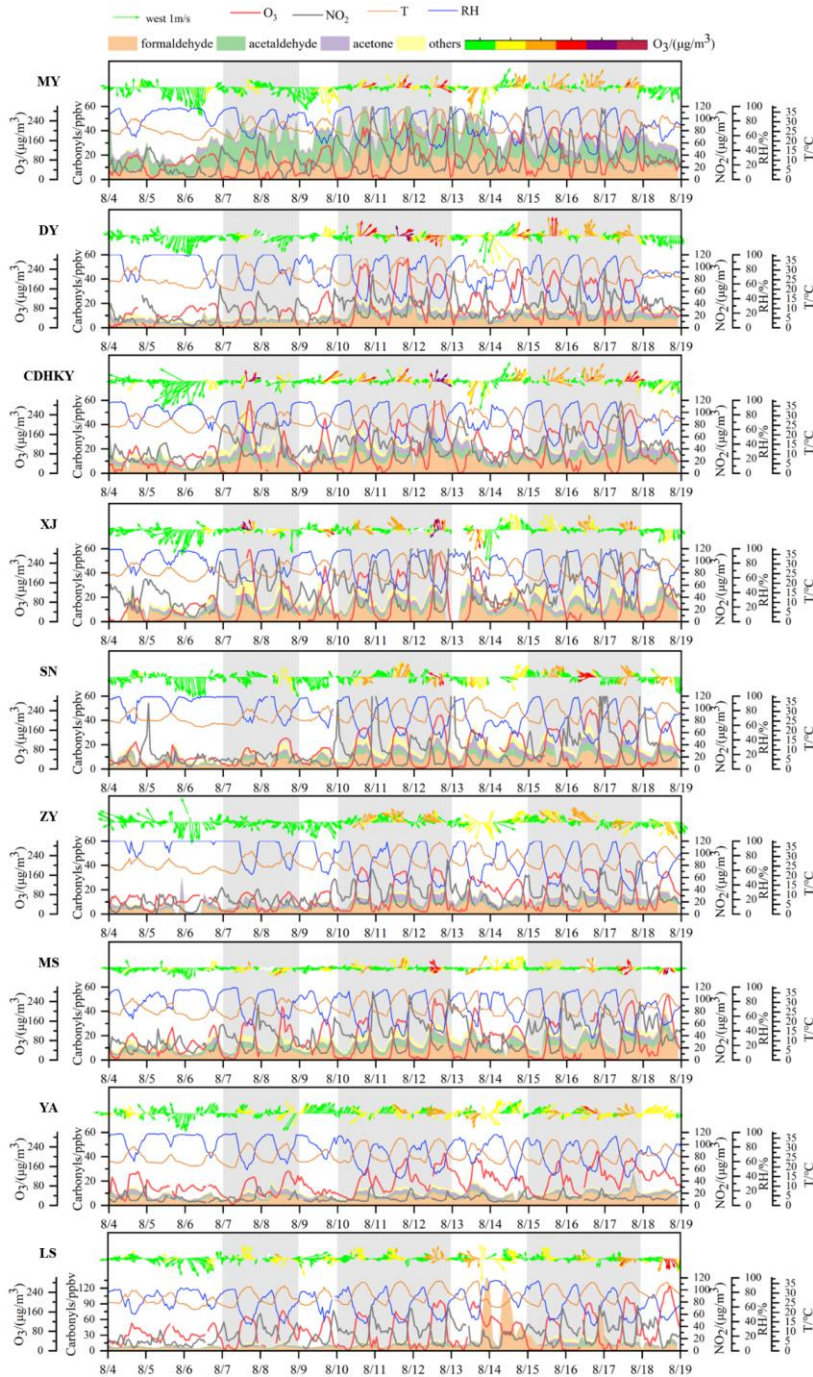
707 Fig.3 illustrates the temporal and spatial variations of ozone and NO₂
708 concentrations, as well as temperature and humidity at each site during the observation
709 period. After observing the spatial distribution of ozone concentration during EP1, it's
710 evident that the severity of pollution reached heavily polluted levels, with Chengdu
711 recording an **MDA8** concentration of 297 $\mu\text{g}\cdot\text{m}^{-3}$ on August 7th. This distribution
712 demonstrated a radial decrease from Chengdu to the surrounding areas. However, the
713 subsequent episodes, EP2 and EP3, exhibited even broader ranges of ozone pollution
714 and more pronounced spatial movements. During the early stages of EP2 and EP3 (from
715 August 10th to 11th and from August 14th to 15th, respectively), high ozone
716 concentrations were observed in the Chengdu-Deyang-Mianyang region. In the middle
717 stages (August 12th and from August 16th to 17th, respectively), influenced by northerly
718 airflow, regions with high ozone concentrations expanded to the central (Meishan,
719 Ziyang, and Suining) and southwestern (Leshan and Ya'an) parts of the CPUA. In the
720 later stages (August 13th and August 18th), under the influence of northwesterly airflow,
721 regions with high ozone concentrations (Meishan and Leshan) moved southward again,
722 while ozone pollution in other areas of the CPUA gradually weakened. On August 11th
723 to 12th and August 16th to 17th, ozone concentrations in the eight cities of the CPUA
724 reached light pollution levels or higher, with the heaviest pollution recorded on August
725 12th. Specifically, Deyang, Mianyang, Suining, and Meishan reached moderate
726 pollution levels, while Chengdu reached heavy pollution with a concentration of 324

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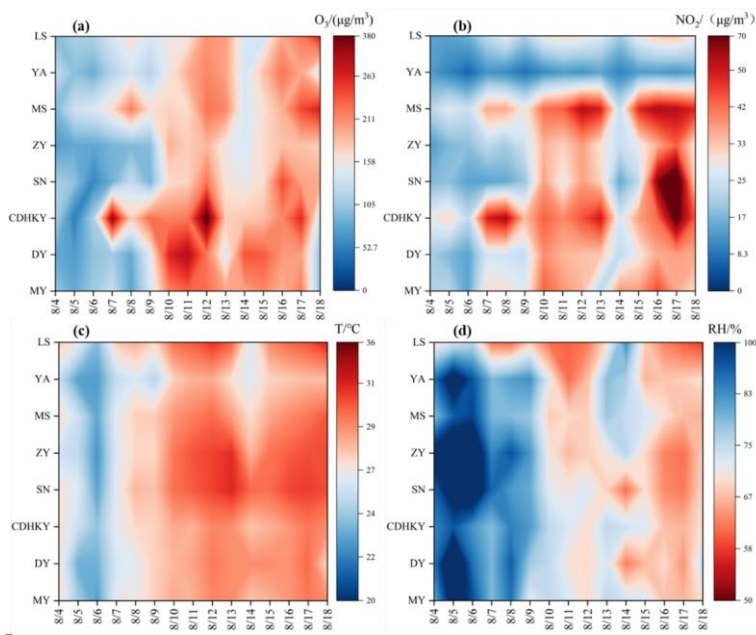
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732 **Figure 2.** Overview of air quality at each site during the observation period. The gray shaded parts
 733 respectively represent the three heavy ozone pollution episodes (EP1, EP2, EP3).

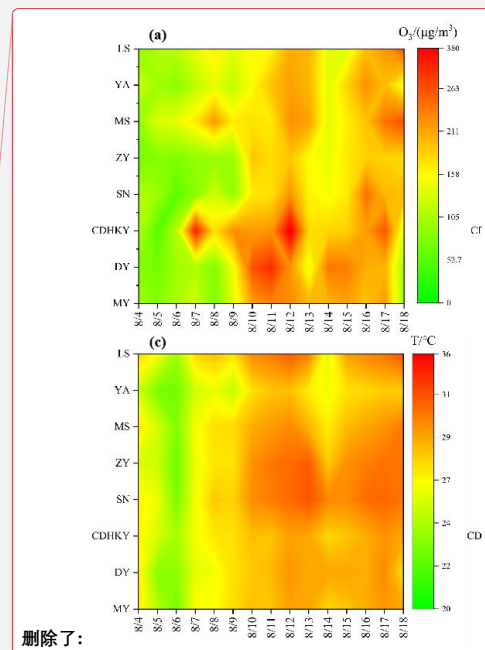


734
 735 **Figure 3.** Temporal and spatial variations of (a) ozone concentration, (b) NO₂ concentration, (c)
 736 temperature and (d) humidity in the CUPA during the observation period.

737 **3.2 Comparative characterization of carbonyl compounds**

738 **3.2.1 Ambient levels**

739 During the observation period, we utilized 2,4-dinitrophenylhydrazine (DNPH)
 740 cartridge and high-performance liquid chromatography (HPLC) analysis technique to
 741 quantify 15 carbonyl compounds. The concentrations and relative proportions of these
 742 compounds are summarized in Table 1. The average total concentration of the 15
 743 carbonyls in the CUPA was 17.35 ± 5.31 ppb. Overall, areas with elevated
 744 concentrations of carbonyl compounds were primarily concentrated in and around
 745 Chengdu in both northern and southern directions. MY site, located to the north of
 746 Chengdu, exhibited the highest concentration of carbonyl compounds (35.18 ± 13.37



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749 ppb), while YA site, situated southwest of Chengdu, showed the lowest concentration
 750 (10.70 ± 4.16 ppb).

751

752 **Table 1.** Daily mean \pm standard error of carbonyl compound mixing ratios (ppbv) at each site in
 753 the CUPA during the observation period. Sum: the total sum of carbonyl compound mixing ratios
 754 across all compounds at each site.

删除了: Daily mean mixing ratio of carbonyl compounds at each site in the CUPA during the observation period (ppbv)

Carbonyls	MY	DY	CDHKY	XJ	SN	ZY	MS	YA	LS
formaldehyde	12.82 \pm 6.52	6.06 \pm 2.82	10.09 \pm 4.21	8.87 \pm 4.39	6.98 \pm 3.56	5.84 \pm 2.69	8.47 \pm 4.15	6.36 \pm 2.40	6.55 \pm 3.35
acetaldehyde	16.65 \pm 7.38	1.54 \pm 0.77	3.65 \pm 2.15	2.33 \pm 1.07	2.62 \pm 1.74	1.40 \pm 0.61	3.24 \pm 1.60	0.88 \pm 0.68	1.63 \pm 1.32
acetone	4.36 \pm 1.70	2.80 \pm 1.19	4.51 \pm 2.25	3.70 \pm 1.21	3.14 \pm 1.70	3.23 \pm 1.73	2.15 \pm 1.14	2.18 \pm 1.08	2.91 \pm 1.63
propionaldehyde	0.41 \pm 0.22	0.24 \pm 0.14	0.39 \pm 0.27	0.39 \pm 0.17	0.34 \pm 0.22	0.28 \pm 0.14	0.41 \pm 0.18	0.20 \pm 0.15	0.31 \pm 0.16
crotonaldehyde	0.20 \pm 0.21	0.10 \pm 0.11	0.23 \pm 0.34	0.05 \pm 0.07	0.23 \pm 0.08	0.19 \pm 0.27	0.15 \pm 0.21	0.36 \pm 0.24	0.12 \pm 0.24
butyraldehyde	0.22 \pm 0.48	0.22 \pm 0.28	0.40 \pm 0.57	0.94 \pm 1.67	0.26 \pm 0.18	0.06 \pm 0.18	0.44 \pm 0.46	0.25 \pm 0.16	0.02 \pm 0.06
benzaldehyde	0.00 \pm 0.04	0.02 \pm 0.06	0.04 \pm 0.11	0.21 \pm 0.20	0.08 \pm 0.10	0.00 \pm 0.01	0.00 \pm 0.01	0.00 \pm 0.00	0.01 \pm 0.04
isovaleraldehyde	0.01 \pm 0.14	0.03 \pm 0.09	0.08 \pm 0.14	0.08 \pm 0.13	0.05 \pm 0.10	0.01 \pm 0.05	0.68 \pm 0.42	0.04 \pm 0.07	0.06 \pm 0.12
valeraldehyde	0.00 \pm 0.00	0.25 \pm 0.09	0.30 \pm 0.59	0.63 \pm 0.36	0.85 \pm 0.65	0.00 \pm 0.00	0.00 \pm 0.00	0.00 \pm 0.02	0.77 \pm 0.47
o-Tolualdehyde	0.46 \pm 0.52	0.36 \pm 0.29	0.45 \pm 0.19	0.00 \pm 0.00	0.00 \pm 0.00	0.23 \pm 0.17	0.43 \pm 0.33	0.18 \pm 0.22	0.16 \pm 0.17
m-Tolualdehyde	0.00 \pm 0.02	0.04 \pm 0.10	0.04 \pm 0.09	0.17 \pm 0.17	0.30 \pm 0.13	0.00 \pm 0.03	0.00 \pm 0.02	0.00 \pm 0.02	0.01 \pm 0.05
p-Tolualdehyde	0.00 \pm 0.00	0.01 \pm 0.05	0.01 \pm 0.04	0.00 \pm 0.00	0.00 \pm 0.00	0.00 \pm 0.00	0.01 \pm 0.04	0.00 \pm 0.02	0.00 \pm 0.02
hexaldehyde	0.00 \pm 0.01	0.34 \pm 0.25	0.41 \pm 0.69	0.57 \pm 0.47	0.95 \pm 0.65	0.02 \pm 0.18	0.78 \pm 0.58	0.00 \pm 0.01	0.10 \pm 0.32
2,5-dimethylbenzaldehyde	0.01 \pm 0.03	0.00 \pm 0.01	0.00 \pm 0.01	0.05 \pm 0.12	0.00 \pm 0.00	0.00 \pm 0.01	0.01 \pm 0.02	0.00 \pm 0.01	0.00 \pm 0.01
MACR	0.03 \pm 0.20	0.14 \pm 0.17	0.26 \pm 0.34	1.05 \pm 1.10	0.26 \pm 0.21	0.19 \pm 0.16	0.42 \pm 0.36	0.24 \pm 0.22	0.81 \pm 0.88
Sum	35.18\pm13.37	12.16\pm4.84	20.84\pm8.85	19.04\pm8.1	16.05\pm7.73	11.47\pm4.89	17.19\pm7.61	10.70\pm4.16	13.46\pm6.12

755

756

759

760 Fig.S1 illustrates the relationship between ozone concentration and carbonyl
761 compounds concentration at each site during the observation period. It is evident that
762 the spatial distribution of carbonyl compound concentrations is similar to that of ozone
763 concentration. Regions with severe ozone pollution tend to exhibit higher
764 concentrations of carbonyl compounds. The variation in carbonyl compound
765 concentrations is primarily attributed to anthropogenic emissions and prevailing
766 summer wind directions in the CPUA. Chengdu is the most economically developed
767 city in the CPUA, with notably higher GDP and industrial production values than other
768 regions. Chengdu's major industries include coal-fired power plants, chemical plants,
769 metallurgy and building materials plants, and high concentrations of carbonyls were
770 observed in here. The unique basin climate of the CPUA, characterized by intense
771 sunlight and stable atmospheric conditions, facilitates the accumulation of pollutants.
772 Large amount of industrial emissions and strong photochemical reaction contributes to
773 ozone pollution. Additionally, during the summer, prevailing northerly winds in the
774 CPUA facilitate the downwind transport of pollutants from upwind sources, leading to
775 regional pollution. It is noteworthy that the concentration of carbonyl compounds at the
776 MY site significantly exceeds that at the CDHKY site. Mianyang, with its industrial
777 roots, consistently maintains its position as the second-highest GDP contributor in
778 Sichuan Province. The electronics information industry stands as Mianyang's primary
779 economic driver, constituting approximately half of the city's total output value. Studies
780 investigating the volatile organic compound (VOC) source profile in Chengdu(Zhou et
781 al., 2021) reveal that ethanol and carbonyls predominantly characterize electronics
782 manufacturing emissions.

783 3.2.2 Compositional characteristics

784 According to the composition characteristics of 15 carbonyl compounds in the
785 ambient air of each city during the observation period (Table S4) . Formaldehyde was

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787 the most abundant carbonyl, found in these sites followed by acetone and acetaldehyde,
788 which is widely observed in previous studies. The measured ratios of formaldehyde,
789 acetone, and acetaldehyde across different sites ranged from 36.4% to 59.4% (average
790 48.1%), 12.4% to 28.1% (average 19.9%), and 8.2% to 47.3% (average 17.5%),
791 respectively. In this study, the total measured of formaldehyde, acetaldehyde, and
792 acetone (FAT) account for over 78% of the total carbonyls concentrations. At the MY
793 and ZY sites, this proportion even exceeded 90%. It is noteworthy that isobutyraldehyde
794 (MACR) ranks fourth in the volume concentration of 15 carbonyls measured in the
795 ambient air surrounding XJ, accounting for 5.3%. MACR, a characteristic product of
796 isoprene photooxidation from biogenic sources, possibly originates from the abundant
797 vegetation surrounding XJ. It reflects the period's relatively active photochemical
798 reactions, with substantial contributions from secondary formation to the measured
799 carbonyls composition.

800 The observed levels of FAT in different areas were influenced by various factors
801 including sampling period, geographic location, meteorological conditions, chemical
802 removal, and source emissions(Z. Zhang et al., 2016). Despite these influences,
803 comparisons remain valuable in providing an overview of ambient carbonyl levels in
804 the CPUA. During the summer of 2010, a national wide survey of ambient
805 monocarbonyl compounds were conducted simultaneously in nine sites (Ho et al.,
806 2015)found that the total FAT concentration was highest in Chengdu (14.96 ppb),
807 followed by Beijing (11.83 ppb), and Wuhan (11.70 ppb). Beijing, as the capital of
808 China, and Wuhan, being one of the top ten most populous cities in China, played
809 significant roles in this comparison. In our study, the CDHKY site within CPUA
810 exhibited the highest FAT concentration, with values of 18.25 ppb, surpassing those
811 recorded in 2010. Furthermore, the total FAT concentrations observed at the CPUA and
812 XJ sites, with values of 14.99 ppb and 14.90 ppb respectively in our study, closely
813 resemble those reported in August 2010 in Chengdu. The consistently high levels of
814 carbonyl compounds observed in Chengdu, both in 2010 and our current study, indicate

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818 that the city likely experiences higher concentrations of these pollutants compared to
819 other regions across the country. However, more extensive temporal data would be
820 beneficial to fully validate this pattern at a national scale. Comparing our findings to
821 international studies, the FAT concentrations at the CDHKY site were lower than those
822 reported in Rio De Janeiro, Brazil(da Silva et al., 2016), during July to October 2013
823 (35.43 ppb), but higher than those in Bangkok, Thailand(Kanjanasiranont et al., 2016b),
824 Orleans, France(Jiang et al., 2016), and the United States(Murillo et al., 2012), with
825 values of 9.05 ppb, 6.12 ppb, and 5.76 ppb, respectively.

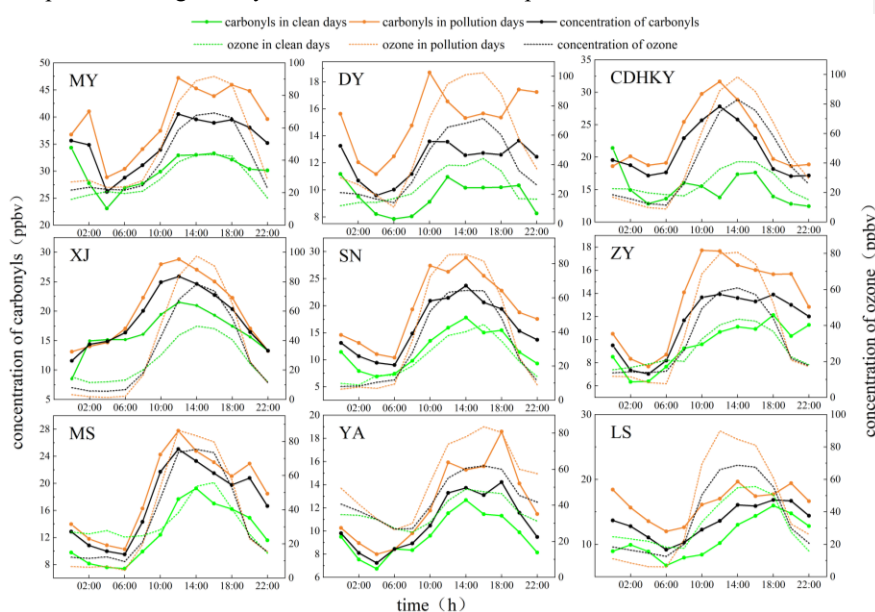
826 3.3 Temporal variations of carbonyl compounds

827 The diurnal variation of the total mixing ratio of ambient carbonyl compounds and
828 ozone concentration around each site in the CPUA during the observation period is
829 shown in Fig. 4. According to the observation results, the diurnal trend of ozone
830 concentration at each site showed a "unimodal" variation characteristic, that was, it
831 gradually increased from the morning to the peak of one day at noon, and then decreased.
832 The diurnal variation of the total mixing ratio of carbonyl compounds at each site
833 generally showed a characteristic of high during the daytime and low at night. The
834 concentration of carbonyl compounds during the day (6:00-16:00) was 48.8% higher
835 than that at night (18:00-4:00) at the XJ site. This indicated that the concentration of
836 carbonyl compounds increased by photochemical production during the daytime.
837 Additionally, deposition processes, particularly dry deposition at night, likely
838 contribute to the observed diurnal variation in carbonyl levels. The diurnal variation
839 characteristics of each site were different. For example, the diurnal variation
840 characteristics of carbonyl compounds concentration at CDHKY, XJ and SN sites were
841 consistent with those of ozone. The diurnal variation of carbonyl compounds
842 concentrations at other sites showed "double peaks", peaking at 10:00-12:00 and 18:00-
843 20:00, respectively. The concentrations of carbonyl compounds at night were also
844 higher at MY, DY and LS sites. The diurnal minimum values of the total concentration
845 of carbonyl compounds and ozone concentration appeared at similar time, usually at

删除了: This suggests that elevated concentrations of carbonyl compounds in Chengdu have been a longstanding issue on a national scale.

849 4:00 a.m. or 6:00 a.m. The first peak of the total mixing ratio of carbonyl compounds
 850 occurred earlier than the maximum ozone concentration of the day. The first peak of
 851 the total mixing ratio of carbonyl compounds mostly occurred between 10:00 and 12:00.
 852 And the maximum ozone concentration mostly occurs between 14:00 and 16:00. This
 853 was related to the fact that carbonyl compounds were important precursors of ozone.

854 In general, the diurnal variation of the total concentration of carbonyl
 855 compounds on pollution days and clean days was high during the daytime and low at
 856 night. The total mixing ratio of carbonyl compounds on pollution days was 22.8%-
 857 66.2% higher than that on clean days. At the same time, the increase of concentration
 858 of carbonyl compounds during the daytime on pollution days was higher than that on
 859 clean days. This suggested that the increase in the concentration of carbonyl
 860 compounds during the daytime contributed to ozone pollution.



861
 862 **Figure 4.** Diurnal variations of carbonyl compounds and ozone concentrations at each site in the
 863 CUA during the observation period

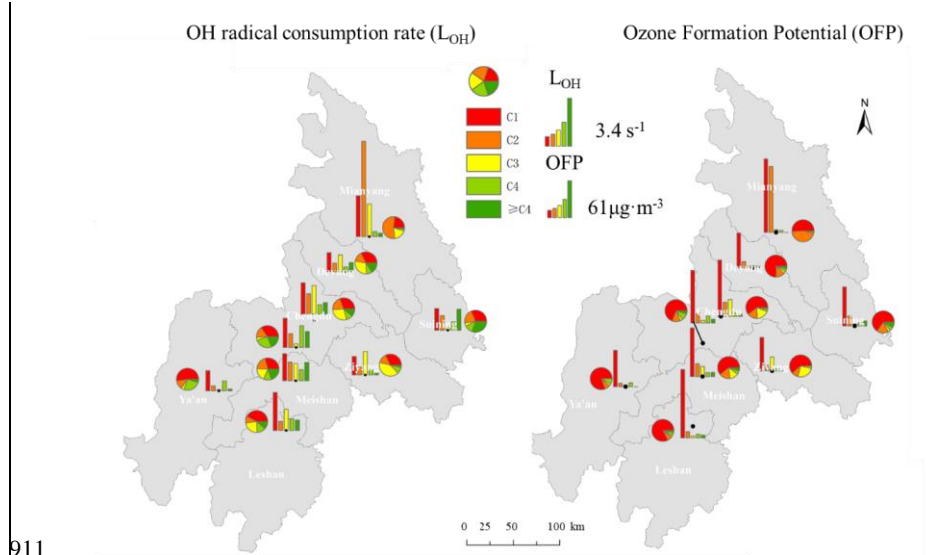
864 **3.4 Atmospheric photochemical reactivity of carbonyl compounds**

865 During the observation period, the total OH radical consumption rate (LOH) and

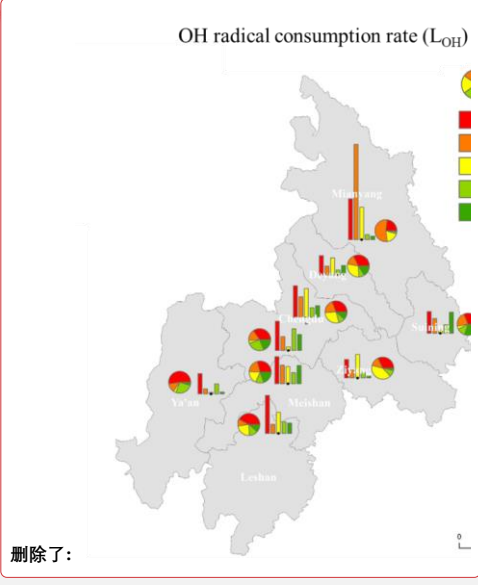
删除了: The diurnal variation of the mixing ratio of ambient carbonyl compounds on weekdays and weekends in the eight cities of the CUA is shown in Fig. S2. The total concentration of carbonyl compounds at each site on weekends was higher than that on weekdays, and the increase in carbonyl compounds at 0:00 (36.3%), 10:00 (16.3%) and 18:00-22:00 (17.6%) on weekends was higher than that on weekdays. Except for the XJ site, the increase in the concentration of carbonyl compounds at 0:00 on weekends was significantly higher than that on weekdays, which was mainly related to the increase of acetaldehyde, propionaldehyde and acetone on weekends. At 10:00, the higher increase at DY, CDHKY and SN sites was mainly related to the increase of propionaldehyde, acetaldehyde and formaldehyde concentrations. From 18:00 to 22:00, the higher increase at DY and YA sites was mainly related to the increase in the concentrations of propionaldehyde, acetone and acetaldehyde. Acetaldehyde, acetone and propionaldehyde were mainly from vehicle exhaust. In particular, when ethanol gasoline and biodiesel were used as alternative fuels, the content of acetaldehyde and acetone in the exhaust gas would be significantly increased. Therefore, the increase in the concentration of carbonyl compounds on weekends might be related to the increase in traffic at 10:00 and at night. In addition, the peak concentration of carbonyl compounds on weekends (10:00) was earlier than that on weekdays (12:00-14:00) at CDHKY, XJ and SN sites, and the diurnal trend of carbonyl compounds concentrations on weekdays and weekends had little difference at other sites.

895 total ozone formation potential (OFP) of the 15 carbonyl compounds at each site are
 896 depicted in Fig.5. The ranking of total L_{OH} and total OFP at each site is consistent,
 897 except for the YA and ZY sites with lower concentrations of carbonyl compounds,
 898 where the atmospheric photochemical reactivity ranking also aligns with the
 899 concentration. Among all sites, the MY and CD sites display the highest reactivity,
 900 while the YA and ZY sites exhibit the lowest reactivity. During the observation period,
 901 carbonyl compounds significantly contributed to ozone formation. The contributions to
 902 total VOCs (alkanes, alkenes, alkynes, aromatics, and carbonyl compounds) OFP at the
 903 MY, SN, ZY, YA, and LS sites ranged from 19.5% to 48.6%. Formaldehyde and
 904 acetaldehyde were identified as the most reactive species in the atmosphere, surpassing
 905 other carbonyl compounds in reactivity due to their higher concentrations and inherent
 906 reactivity, especially formaldehyde. However, acetone exhibited high inertness and a
 907 prolonged atmospheric lifetime, leading to its accumulation in ambient air with
 908 concentrations higher than other carbonyl compounds except for formaldehyde and
 909 acetaldehyde. Thus, despite its elevated concentration, acetone's reactivity remained
 910 relatively low.

删除了: Contrasting the L_{OH} and OFP during clean and polluted periods reveals higher values during ozone pollution periods than clean days. L_{OH} and OFP during different pollution periods show a strong positive correlation with the severity of ozone pollution; the heavier the ozone pollution, the higher the L_{OH} and OFP at the sites. Regardless of clean or polluted periods, the L_{OH} and OFP at the MY site are higher than other sites. However, despite this, the average ozone concentration at the MY site ranks lower among the nine sites observed. This might be associated with higher concentrations of aldehyde compounds at the MY site.



911
 912 **Figure 5.** L_{OH} and OFP of carbonyl compounds at each site in the CPUA during the observation



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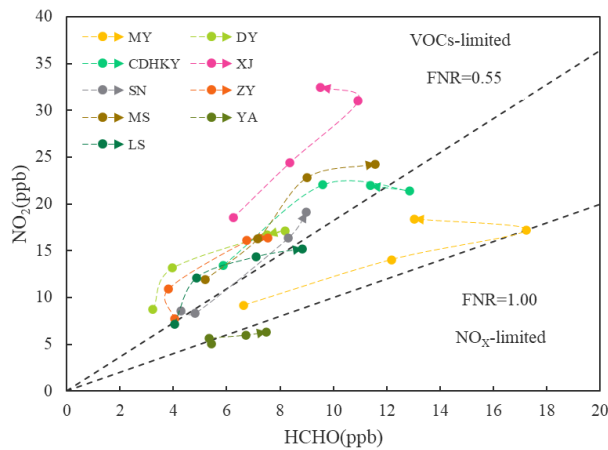
926

period

927 **3.5 Sensitivity analysis of ozone formation based on formaldehyde to NO₂ ratio (FNR)**

928 The change of O₃ formation sensitivity of each site in the CPOA during the
929 observation period is shown in Fig. 6. As can be seen from the Fig. 6, most sites remain
930 in the VOCs-limited regime during the cleaning period and EP1 to EP3. Economically
931 developed city such as Chengdu, Meishan, with high levels of formaldehyde and NO₂,
932 remain in the VOCs-limited regime. Ya'an as a city with the lowest GDP ranking in the
933 CPOA, with low levels of formaldehyde and NO₂, remain in the transitional regime.

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934

935 **Figure 6.** The change of O₃ formation sensitivity of each site in the CPOA during the observation
936 period. The arrows represent time step from clean period to EP1 to EP2 to EP3.

937 The daily variation of O₃ formation sensitivity and ozone concentration at each
938 site in the CPOA during the observation period is shown in Fig. S4. The mean FNR of
939 each site ranged from 0.48 to 1.29 during the observation period. The FNRs were lower
940 than 0.55±0.16 at XJ, DY, ZY, CDHKY, and MS, and higher than 1.0 at LS, SN, YA
941 and MY. At the same time, the mean ozone concentration at each site was between 138
942 and 192 $\mu\text{g}\cdot\text{m}^{-3}$. The mean ozone concentration in XJ, DY, CDHKY and MS was 166-
943 192 $\mu\text{g}\cdot\text{m}^{-3}$, it was 150-164 $\mu\text{g}\cdot\text{m}^{-3}$, in LS, SN, YA and MY. Therefore, it could be seen
944 that most of the sites with high mean ozone concentrations during the observation

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950 period, like CDHKY, XJ, MS and Deyan sites, were in the VOCs-limited regime, and
951 most of the stations with low mean ozone concentrations during the observation period
952 such as YA, SN, MY and LS were in the transitional regime. It was worth noting that
953 the mean ozone concentration at ZY site (only 138 $\mu\text{g}\cdot\text{m}^{-3}$) during the observation
954 period was much lower than that of other sites, but most of the ZY site was in VOCs-
955 limited regime, which was mainly related to the low concentration of formaldehyde. In
956 addition, the FNR value of the MY site was also relatively high, which was mainly
957 caused by the high concentration of formaldehyde.

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958 Based on the ratio of formaldehyde to NO_2 mixing ratio, most sites remain in the
959 VOCs-limited regime during the observation period. And the sites with heavy ozone
960 pollution were in the VOCs-limited regime, and the sites with light ozone pollution
961 were in the transitional regime. Photochemical reactivity (L_{OH} and OFP) analysis
962 showed that formaldehyde and acetaldehyde contributed significantly to the
963 enhancement of atmospheric oxidation and ozone formation potential. Therefore, when
964 heavy ozone pollution occurs in the CPUA, special attention should be paid to the
965 control of VOCs, especially formaldehyde and acetaldehyde in carbonyl compounds,
966 under the coordinated control of NO_x and VOCs. Overall, this study reveals the
967 important contribution of carbonyl compounds to ozone pollution in the CPUA, and
968 provides scientific support for the establishment of ozone pollution prevention and
969 control measures.

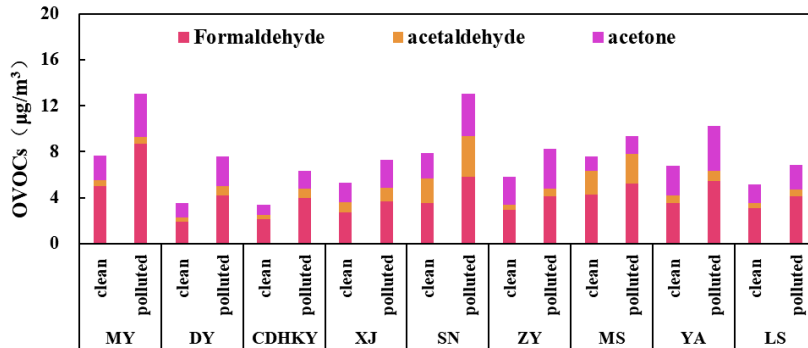
970 3.6 Source Analysis of carbonyl compounds

971 3.6.1 Quantitative source analysis of key carbonyl compounds

972 The table S7 provides a summary of the background and primary emissions
973 concentrations of formaldehyde, acetaldehyde, and acetone at nine sites across the eight
974 cities of the CPUA, along with the proportion of secondary formation contributing to
975 their concentrations. Background concentrations and primary emissions of
976 formaldehyde, acetaldehyde, and acetone ranged from 50% to 80%, 46% to 83%, and

978 45% to 78%, respectively. Secondary formation accounted for 20% to 50%, 17% to
 979 54%, and 22% to 55% of their concentrations, respectively. Notably, in SN and YA, the
 980 secondary formation of formaldehyde contributed half of the observed concentration,
 981 indicating it as the predominant source, while acetaldehyde's secondary formation also
 982 prevailed in these sites. Conversely, acetone, with lower reactivity, primarily originated
 983 from background concentrations and primary emissions at other sites except YA.
 984 Moreover, background concentrations and primary emissions were identified as the
 985 main contributors to carbonyl compounds in XJ and LS.

986 Fig.7 illustrates the secondary formation concentrations of formaldehyde,
 987 acetaldehyde, and acetone at each site in the CPUTA under both clean and polluted
 988 conditions. Under polluted conditions, the secondary concentrations of formaldehyde,
 989 acetaldehyde, and acetone exceeded those in clean conditions by 52.4%, 80.3%, and
 990 58.5%, respectively. The most significant increases in secondary concentrations were
 991 observed at the SN site, while relatively smaller increases were observed at LS and XJ.



992
 993 **Figure 7.** Concentrations of formaldehyde, acetaldehyde and acetone in secondary formation
 994 under different pollution conditions at each site in the CPUTA during the observation period

995 **3.6.2 Exploration of secondary formation mechanism of key carbonyl compounds**

996 In this study, we utilized VOC data collected on August 11, 12, and 16, when all
 997 eight cities in the CPUTA were experiencing mild to severe ozone pollution. We
 998 calculated the Relative Incremental Reactivity (RIR) of formaldehyde, acetaldehyde,

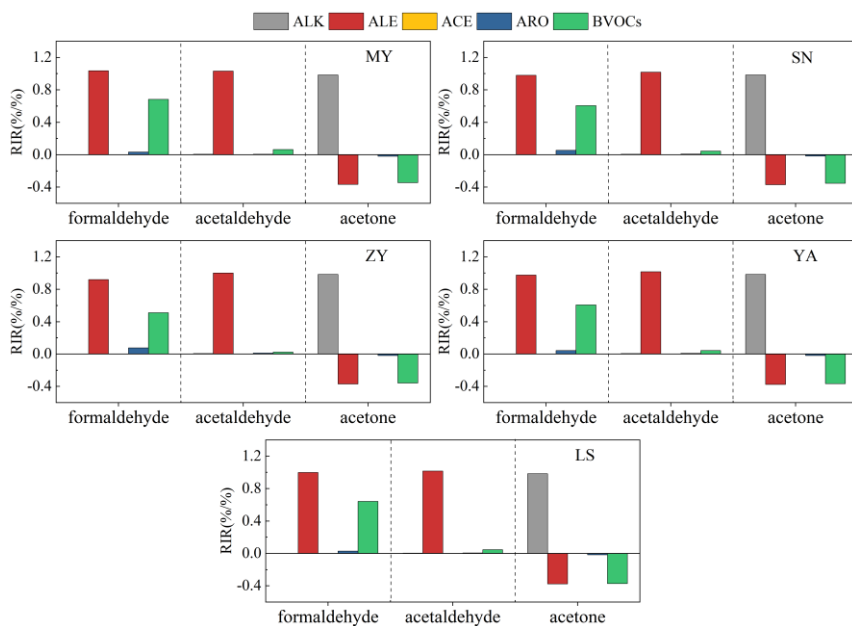
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删除了: The effects of anthropogenic source VOCs and plant source VOCs on the formation of formaldehyde, acetaldehyde and acetone at MY, SN, ZY, YA and LS sites were researched during a regional ozone pollution period when all 8 cities of the CPUTA had mild or above ozone pollution (August 11th, 12th and 16th)

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1006 and acetone at the MY, SN, ZY, YA, and LS sites on these days. The OBM analysis
 1007 allowed us to assess the impact of anthropogenic VOCs (alkanes, alkenes, alkynes,
 1008 aromatics) and biogenic VOCs (e.g., isoprene) on carbonyl compound formation in the
 1009 context of regional ozone pollution events (Fig.8). Overall, the sensitivities of different
 1010 anthropogenic source and plant source VOCs to formaldehyde, acetaldehyde and
 1011 acetone was consistent among sites. For formaldehyde, reducing alkenes in
 1012 anthropogenic source VOCs and plant VOCs was the most effective way to control
 1013 formaldehyde concentration, while reducing alkenes in anthropogenic source VOCs
 1014 was also beneficial to reduce the formation of acetaldehyde. For acetone with low
 1015 reactivity, the alkanes in anthropogenic source VOCs were the most sensitive to the
 1016 formation of acetone, followed by alkenes and BVOCs. Only the RIR value of alkanes
 1017 were greater than zero, and the RIR values of both alkenes and BVOCs were less than
 1018 zero, indicating that reducing alkanes could reduce the formation of acetone, while
 1019 reducing alkenes and BVOCs was not conducive to acetone concentration control.

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1020
 1021 **Figure 8.** Mean RIRs of formaldehyde, acetaldehyde and acetone to different anthropogenic
 1022 source VOCs and biogenic source VOCs at MY, SN, ZY, YA and LS sites on August 11th, 12th and

1023

16th

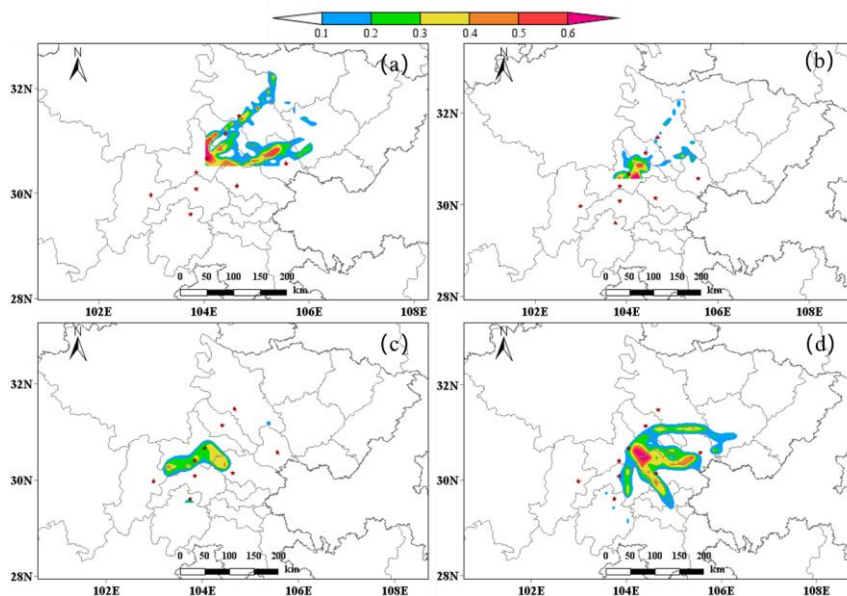
1024 **3.6.3 Influence of regional transportation contribution**

1025 The TrajStat trajectory model was used to calculate and cluster the 24-hour
1026 backward trajectories of air quality at the sampling sites. The backward trajectory
1027 during sampling is shown in Fig.S5. During the observation period, the pollution of
1028 carbonyl compounds in the cities of the CPUTA was affected by the mutual transport
1029 among cities in Sichuan Province, especially along the MY-DY-CDHKY route. In
1030 addition, the surrounding provinces and cities of Sichuan Province (Gansu and
1031 Chongqing) also contributed to the carbonyl compounds of the CPUTA.

1032 The potential sources of carbonyl compounds at different pollution stages at the
1033 CDHKY during the observation period are shown in Fig. 9. It can be seen from the
1034 figure that there are differences in the potential sources of carbonyl compounds among
1035 different pollution stages at the CDHKY site. The concentration of local carbonyl
1036 compounds in CDHKY was high during the early observation period and EP1, which
1037 existed local sources, and was also affected by the northern airflow, and carbonyl
1038 compounds was also affected by the transport from MY, DY and other northern regions.
1039 Under the effect of the continuous northern airflow, the local source emissions
1040 decreased during EP1, and the potential source of carbonyl compounds changed to from
1041 the junction between CDHKY and ZY. During EP3, under the combined influence of
1042 the western airflow, the contribution of transport from SN and ZY to carbonyl
1043 compounds increased, while emissions from local sources also increased.

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删除了: The potential sources of carbonyl compounds at different pollution stages at the Chengdu Institute of Environmental Sciences site during the observation period are shown in Fig. 9. It can be seen from the figure that there are differences in the potential sources of carbonyl compounds among different pollution stages at the CDHKY site. The concentration of local carbonyl compounds in CDHKY was high during the early observation period and EP1, which existed local sources, and was also affected by the northern airflow, and carbonyl compounds was also affected by the transport from MY, DY and other northern regions. Under the effect of the continuous northern airflow, the local source emissions decreased during EP1, and the potential source of carbonyl compounds changed to from the junction between CDHKY and ZY. During EP3, under the combined influence of the western airflow, the contribution of transport from SN and ZY to carbonyl compounds increased, while emissions from local sources also increased.



1063
1064 **Figure 9.** Analysis of potential sources of carbonyl compounds at different periods at the CDHKY
1065 site during the observation period (a) August 4th-6th (b) August 7th-9th (c) August 10th-13th (d)
1066 August 15th-18th

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1067 **4. Conclusions**

1068 During a concurrent atmospheric observation campaign conducted at nine sites in
1069 the CPUA from August 4th to 18th, 2019, three regional heavy ozone pollution episodes,
1070 labeled EP1 to EP3, were observed. This study extensively examines the concentration
1071 variations, atmospheric chemical reactivity, and sources of carbonyls during this period.
1072 The average total concentrations of 15 carbonyl compounds across the nine sites within
1073 eight cities of the CPUA were measured at 17.35 ± 5.31 ppb. Spatial analysis revealed
1074 a positive correlation between carbonyl levels and ozone concentrations, particularly
1075 concentrated around Chengdu in both northern and southern directions. Formaldehyde
1076 (36.4%-64.3%), acetone (12.4%-28.1%), and acetaldehyde (8.2%-47.3%) constituted
1077 the predominant species by volume concentration. Intriguingly, Chengdu exhibited FAT
1078 concentrations surpassing national and international levels, indicating heightened levels

1082 compared to other regions. Diurnal variations showed peaks during the day and lows at
1083 night, with notable spikes on ozone pollution days. A distinctive "weekend effect" was
1084 observed, particularly evident in carbonyl compounds associated with motor vehicle
1085 emissions, such as acetaldehyde and acetone, peaking during morning rush hours and
1086 nighttime on weekends. This suggests significant contributions from both daytime
1087 photochemical processes and nighttime vehicular emissions to carbonyl compounds. At
1088 the MY site, 48.6% of the total volatile organic compounds (VOCs) ozone formation
1089 potential (OFP) was attributed to the 15 carbonyl compounds, emphasizing their
1090 substantial impact on ozone formation, especially formaldehyde and acetaldehyde.

1091 Ground-level observations of FNR were utilized to assess the sensitivity of
1092 ground-level ozone formation. FNR from ground-level observations were used to
1093 determine the sensitivity of ground-level ozone formation. Analysis of FNR revealed
1094 that sites experiencing heavy ozone pollution exhibited lower FNRs, indicating a
1095 VOCs-limited regime, while sites with lighter ozone pollution were categorized into a
1096 transitional regime. Carbonyl compound sources include primary emissions and
1097 secondary formation processes. Multivariate linear regression quantitatively analyzed
1098 formaldehyde, acetaldehyde, and acetone sources. Secondary formation contributed
1099 over 30% on average to formaldehyde, acetaldehyde, and acetone, despite primary
1100 emissions being primary sources. OBM modeling revealed that formaldehyde and
1101 acetaldehyde primarily originated from the secondary formation of alkenes and BVOCs,
1102 while acetone mainly stemmed from the secondary formation of alkanes. Furthermore,
1103 it is recommended to establish a scientific control mechanism for both NO_x and VOCs,
1104 with special attention to formaldehyde, acetaldehyde, and acetone, and their alkenes
1105 precursors. Additionally, considering the regional nature of pollution, this study
1106 suggests that carbonyl compound pollution is influenced by mutual transport among
1107 cities within the CPUA, notably along the MY-DY-CDHKY route. Establishing a
1108 collaborative prevention and control mechanism among cities within the CPUA and
1109 neighboring provinces and cities is crucial to effectively address carbonyl compounds

1110 and ozone pollution in the region in the future.

1111

1112 **Data availability.** Observational data including meteorological parameters and air
1113 pollutants used in this study are available from the corresponding authors upon request
1114 (lihong@craes.org.cn).

1115

1116 **Author contributions.** Hong Li and Jiemeng Bao designed this study. Xin Zhang,
1117 Zhenhai Wu, Jiemeng Bao, Li Zhou, Qinwen Tan, and Fumo Yang coordinated the
1118 selection of field observation sites, including locations for both VOCs and carbonyls
1119 grid sampling. Qinwen Tan and Hefan Liu supported the collection of carbonyls at one
1120 site. Zhenhai Wu and Xin Zhang assisted in carbonyls sampling; Xin Zhang and
1121 Yunfeng Li assisted in carbonyls sample analysis and data collection. Li Zhou and
1122 Hefan Liu organized the analysis of VOCs measurements. Jun Qian, Junhui Chen, and
1123 Liqun Deng provided support in project funding application. Jiemeng Bao performed
1124 the data analysis and wrote the paper with contributions from all co-authors; Hong Li
1125 reviewed the paper, provided comments and finalized it.

1126

1127 **Competing interests.** The contact author has declared that none of the authors has any
1128 competing interests.

1129

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1144

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