Supplement of

# The Role of Naphthalene and Its Derivatives in the Formation of Secondary Organic Aerosols in the Yangtze River Delta Region, China 

Fei Ye et al.

Correspondence to: Jingyi Li (jingyili@nuist.edu.cn), Jianlin Hu (jianlinhu@nuist.edu.cn)

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## Text S1. Fitting the SOA yield by one-product and two-product methods

In the CMAQ model, counter species were used to calculate the production of SOA through gas-particle partitioning based on yields $\left(\alpha_{i}\right)$ and partitioning coefficients ( $K_{\mathrm{om}, i}, \mathrm{~m}^{3} \mu \mathrm{~g}^{-1}$ ) of condensable organic products derived from chamber experiment data. The definition of $K_{\mathrm{om}, i}$ followed Pankow (1994) as shown below:

$$
\begin{equation*}
K_{\mathrm{om}, i}=\frac{1}{C_{\mathrm{sat}, i}^{*}}=\frac{R T}{10^{6} \mathrm{MW}_{\mathrm{om}} \xi_{i} P_{\mathrm{L}, i}^{0}} \tag{1}
\end{equation*}
$$

where $\overline{\mathrm{MW}_{\mathrm{om}}}\left(\mathrm{g} \mathrm{mol}^{-1}\right)$ is the average molecular weight of the organic phase, $\xi_{i}$ is the activity coefficient of species $I$ in the absorbing organic phase, $P_{\mathrm{L}, i}^{0} \quad(\operatorname{atm})$ is the vapor pressure of species $I$ at temperature $T(\mathrm{~K})$, and $R\left(8.314 \mathrm{~J} \mathrm{~mol}^{-1} \mathrm{~K}^{-1}\right)$ is the gas constant. Factor $10^{6}$ is needed for unit conversion. The inverse of the saturation concentration of species $i, C_{\mathrm{sat}, i}^{*}\left(\mu \mathrm{~g} \mathrm{~m}^{-3}\right)$, is equivalent to $K_{\mathrm{om}, i}$. The temperature dependence of $C_{\text {sat }, i}^{*}$ was calculated by the Clausius-Clapeyron equation (Hayes et al., 2015):

$$
\begin{equation*}
C_{\mathrm{sat}, i}^{*}=C_{0, i}^{*} \frac{T_{0}}{T} \exp \left[\frac{\Delta H_{\mathrm{vap}, i}}{R}\left(\frac{1}{T_{0}}-\frac{1}{T}\right)\right] \tag{2}
\end{equation*}
$$

where $C_{0, i}^{*}$ is the effective saturation concentration of condensable species $i$ at the reference temperature $T_{0}(\mathrm{~K})$, usually taken as $298 \mathrm{~K} . \Delta H_{\text {vap, } i}\left(\mathrm{~J} \mathrm{~mol}^{-1}\right)$ is the enthalpy of vaporization.

Assuming that the condensable organic products are $\mathrm{P}_{1}, \mathrm{P}_{2}, \ldots \mathrm{P}_{\mathrm{n}}$, the total SOA yield (Y) of a certain parent VOC was calculated as follows (Odum et al., 1996):

$$
\begin{equation*}
Y=\sum_{i} Y_{i}=M_{0} \sum_{i} \frac{\alpha_{i} K_{\mathrm{om}, i}}{1+M_{0} K_{\mathrm{om}, i}} \tag{3}
\end{equation*}
$$

where $\alpha_{i}$ is the mass-based stoichiometric coefficient of semi-volatile product i. $M_{0}$ is the total mass concentration of the absorbing aerosol medium. The SOA formation scheme is named the n-product method based on the total number (n) of $\mathrm{P}_{i}$. Accordingly, the total SOA yield of the two-product and one-product methods can be derived as follows:

$$
\begin{gather*}
Y=\frac{\alpha_{1}}{1+\frac{1}{M_{0} K_{\mathrm{om}, 1}}}+\frac{\alpha_{2}}{1+\frac{1}{M_{0} K_{\mathrm{om}, 2}}}  \tag{4}\\
Y=\frac{\alpha_{1}}{1+\frac{1}{M_{0} K_{\mathrm{om}, 1}}} \tag{5}
\end{gather*}
$$

## Text S2. Estimation of MN emissions

To obtain the detailed emissions of $1-\mathrm{MN}$ and $2-\mathrm{MN}$, the ratios of $1-\mathrm{MN} / 2 \mathrm{MN}$ to non-methane volatile organic compounds (NMVOCs) of different sources were calculated as follows:

$$
\begin{equation*}
p=\sum_{i=1}^{n}\left(a_{i} * W_{i}\right) \tag{6}
\end{equation*}
$$

where p is the mass proportion of $1-\mathrm{MN} / 2-\mathrm{MN}$ to total NMVOC emissions in a major source, $a_{i}$ is the mass ratio of 1-MN/2-MN to NMVOCs in a subcategorized source $i$, and $W_{i}$ is the weight percent of the total emitted NMVOCs from a subcategorized source $i$ to that of the major source. For a specific source $i, a_{i}$ was obtained from the US EPA (Environmental Protection Agency) repository of organic gas and PM speciation profiles of air pollution sources (SPECIATEv5.2), and $W_{i}$ was based on the information reported by An et al. (2021) for the YRD and Li et al. (2014) for other regions, respectively.

Table S1. The yield ( $\alpha$ ), equilibrium partitioning coefficient ( $K_{\mathrm{om}}$ ), and enthalpy of SOA precursors from Nap, 1-MN, and 2-MN.

|  | Species | $\alpha_{1}$ | $\begin{gathered} K_{\mathrm{om}, 1} \\ \left(\mathrm{~m}^{3} \mu \mathrm{~g}^{-1}\right) \end{gathered}$ | $\alpha_{2}$ | $K_{\mathrm{om}, 2}\left(\mathrm{~m}^{3}\right.$ <br> $\left.\mu \mathrm{g}^{-1}\right)$ | Enthalpy (kJ <br> $\mathrm{mol}^{-1}$ ) | References for $\alpha$ and $\mathrm{C}_{\text {star }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| High $\mathrm{NO}_{\mathrm{x}}$ | Nap | 0.210 | 0.602 | 1.07 | $3.77 \mathrm{E}-3$ | 18 | CMAQ |
| (one-product) | $1-\mathrm{MN}$ | 0.500 | 0.110 | - | - | 16.6 | Chan et al. (2009) |
|  | 2-MN | 0.550 | 0.130 | - | - | 17.5 | Chan et al. (2009) |
| High $\mathrm{NO}_{\mathrm{x}}$ | Nap | 0.210 | 0.602 | 1.07 | $3.77 \mathrm{E}-3$ | 18 | CMAQ |
| (two-product) | 1-MN | 0.206 | 0.193 | 0.107 | 0.001 | 16.6 | Shakya and Griffin (2010) |
|  | 2-MN | 0.247 | 0.193 | 0.0920 | 0.001 | 17.5 | Shakya and Griffin (2010) |
| Low $\mathrm{NO}_{\mathrm{x}}$ | Nap | 0.730 | a |  |  |  | CMAQ |
|  | 1-MN | 0.680 | a |  |  |  | Chan et al. (2009) |
|  | 2-MN | 0.580 | a |  |  |  | Chan et al. (2009) |

${ }^{a}$ Constant yield

Table S2. Sector-based mass ratios of Nap, 1-MN, and 2-MN to NMVOC emissions in the emis-orig case.

| Emission | Species | residential | road transport | industry | industrial | residential |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | solvent | solvent | wildfire ${ }^{\text {c }}$ |
| region |  |  |  |  | use | use |  |
| YRD region ${ }^{\text {a }}$ | Nap |  | The original Nap emission in YRD |  |  |  | 0.391 |
|  | 1-MN | 0.0739 | 0.0143 | $2.62 \mathrm{E}-3$ | $5.47 \mathrm{E}-4$ | 0.0115 | 0.134 |
|  | 2-MN | 0.0898 | 0.0232 | 0.0116 | $2.38 \mathrm{E}-3$ | 0.0500 | 0.170 |
| Other regions ${ }^{\text {b }}$ | Species | residential | transportation | industry | power | agriculture | wildfire $^{\text {c }}$ |
|  | Nap | 0.0252 | 0.0317 | 0.0983 | - | - | 0.391 |
|  | 1-MN | $3.95 \mathrm{E}-3$ | $4.38 \mathrm{E}-3$ | $1.67 \mathrm{E}-4$ | - | - | 0.134 |
|  | 2-MN | 0.0108 | $7.24 \mathrm{E}-3$ | $1.11 \mathrm{E}-3$ | - | - | 0.170 |

[^0]Table S3. Settings of all the scenarios.

| Case | Emission setting | SOA |
| :---: | :---: | :---: |
| base1 | Set emissions of Nap and MN to zero based on the emis-adjust | one-product method |
| case-1 product-orig | Use the original Nap emissions from the 2017 YRD inventory | one-product method |
| case-2products-orig | and new-added MN emissions (emis-orig) | two-product method |
| case-1product | The anthropogenic emissions of Nap and MN in the YRD | one-product method |
| case-2products | region from emis-orig were multiplied by 5 and 7, | two-product method |

Table S4. Model performance of MDA8 $\mathrm{O}_{3}$ and daily average $\mathrm{PM}_{2.5}, \mathrm{SO}_{2}, \mathrm{NO}_{2}$, and CO in several cities listed in Fig. S2 in case-1product. OBS and PRE represent observation and prediction concentrations, respectively. The benchmarks follow Emery et al. (2017). Performances exceeding the benchmark are represented in bold font.

| Species | Metrics | Suzhou | Nanjing | Hangzhou | Hefei | Shanghai | Benchmark |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| MDA8 | OBS | 68.37 | 75.22 | 70.74 | 71.29 | 65.85 |  |
| $\mathrm{O}_{3}(\mathrm{ppb})$ | PRE | 71.54 | 71.93 | 73.49 | 68.48 | 47.50 |  |
|  | NMB | 0.05 | -0.04 | 0.04 | -0.04 | -0.28 | $< \pm 0.15$ |
|  | NME | 0.25 | 0.17 | 0.22 | 0.23 | 0.33 | $<0.25$ |
|  | R | 0.40 | 0.73 | 0.62 | 0.56 | 0.25 |  |
| PM 2.5 | OBS | 42.16 | 38.41 | 37.65 | 46.54 | 42.04 |  |
| $\left(\mu \mathrm{g} \mathrm{m}{ }^{-3}\right)$ | PRE | 31.83 | 36.79 | 31.91 | 48.53 | 23.99 |  |
|  | NMB | -0.24 | -0.04 | -0.15 | 0.04 | -0.43 | $< \pm 0.30$ |
|  | NME | 0.33 | 0.26 | 0.39 | 0.32 | 0.47 | $<0.50$ |
|  | R | 0.36 | 0.39 | 0.21 | 0.38 | 0.47 |  |
| $\mathrm{SO}_{2}$ | OBS | 2.33 | 3.42 | 3.03 | 2.50 | 3.80 |  |
| (ppb) | PRE | 2.51 | 3.55 | 2.36 | 2.70 | 1.76 |  |
|  | NMB | 0.08 | 0.04 | -0.22 | 0.08 | -0.54 |  |
|  | NME | 0.35 | 0.20 | 0.24 | 0.36 | 0.54 |  |
|  | R | 0.49 | 0.70 | 0.64 | 0.50 | 0.44 |  |


| $\mathrm{NO}_{2}$ | OBS | 18.77 | 17.84 | 18.54 | 19.80 | 17.34 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (ppb) | PRE | 12.56 | 13.84 | 10.92 | 15.03 | 17.57 |
|  | NMB | -0.33 | -0.22 | -0.41 | -0.24 | 0.01 |
|  | NME | 0.34 | 0.24 | 0.41 | 0.25 | 0.18 |
|  | R | 0.49 | 0.67 | 0.33 | 0.83 | 0.83 |
| CO | OBS | 549.52 | 590.07 | 673.79 | 644.89 | 495.34 |
| (ppb) | PRE | 471.97 | 611.71 | 483.32 | 507.03 | 307.43 |
|  | NMB | -0.14 | 0.04 | -0.28 | -0.21 | -0.38 |
|  | NME | 0.26 | 0.17 | 0.30 | 0.25 | 0.39 |
|  | R | 0.73 | 0.68 | 0.36 | 0.50 | 0.78 |

Table S5. Model performance of daily ARO1, ARO2MN', BENZ, $\mathrm{NO}_{2}$, and NO at the Taizhou site in case-1product. The units of OBS (observation) and PRE (prediction) are ppb.

|  | ARO1 | ARO2MN' | BENZ | $\mathrm{NO}_{2}$ | NO |
| :---: | :---: | :---: | :---: | :---: | :---: |
| OBS | 0.05 | 0.37 | 0.52 | 9.13 | 1.42 |
| PRE | 0.71 | 0.45 | 0.61 | 7.25 | 0.96 |
| NMB | 13.20 | 0.23 | 0.19 | -0.21 | -0.32 |
| r | 13.20 | 0.36 | 0.37 | 0.27 | 0.62 |

## Aerosol Module

(a)


(c)


Figure S1. Scheme of SOA module in CMAQv5.2. (a) pre-existing SOA formation pathways for Nap fitted by two products under high $\mathrm{NO}_{\mathrm{x}}$; (b) newly added SOA formation pathways for 1-MN fitted by two products under high $\mathrm{NO}_{\mathrm{x}}$; (c) newly added SOA formation pathways for 1-MN fitted by one product under high $\mathrm{NO}_{\mathrm{x}}$. The values of $\alpha$ refer to Table S1.


Figure S2. The modeling domain and locations of Taizhou city (red dot) and other sites (Hefei, Nanjing, Suzhou, Shanghai, and Hangzhou; blue dots) for model performance evaluation. The color bar on the right represents the topography height (in meters).


Figure S3. Regional distributions of Nap, 1-MN, and 2-MN emissions in emis-orig and emis-adjust.


Figure S4. Observed and simulated hourly concentrations of MN, Nap, OC, $\mathrm{PM}_{2.5}$, and $\mathrm{O}_{3}$ based on emis-adjust (red) and emis-orig (blue) at the Taizhou site. Model performances of daily MN, Nap, OC, $\mathrm{PM}_{2.5}$, and MDA8 $\mathrm{O}_{3}$ are shown in blue for case-2products-orig and red for case-2products. OBS and PRE represent averaged concentrations of observations and predictions, respectively.
$\qquad$






Figure S5. Observed and predicted temporal variations of OH and $\mathrm{HO}_{2}$ radicals based on emis-adjust (left column) and emis-orig (right column) at the Taizhou site. Model performances of OH and $\mathrm{HO}_{2}$ radicals are shown in blue for case-1product or case-1product-YRD and red for case-2products or case-2products-YRD when compared with the observations from the EXPLORE-YRD campaign. OBS and PRE represent the average of observation and prediction, respectively.


Figure S6. Episode-averaged composition of OC and SOC-aromatics in case-1product and case-2products. SOC-aromatics represent the SOC generated by all the aromatic precursors. SOC-other represents SOC minus SOC-aromatics.


Figure S7. Contributions of Nap, 1-MN, and 2-MN to aromatic-derived SOC in the case-2products ( $\mathrm{a}-\mathrm{c}$ ) and the case-1product ( $\mathrm{d}-\mathrm{f}$ ) and to the total aromatic emissions ( $\mathrm{g}-\mathrm{i}$ ).


Figure S8. Episode-averaged concentrations of Nap, 1-MN, and 2-MN in case-1 product (first column), concentrations of SOC-Nap, SOC-1MN, and SOC-2MN in case-1product (second column), and relative differences of SOC-1MN and SOC-2MN between case-2products and case-1product calculated as (case-2products -case-1product)/ case-1product $\times 100 \%$ (third column).


Figure S9. Episode-averaged concentrations of SOC and $\mathrm{O}_{3}$ in case-1product (left column) and the changes of $\mathrm{O}_{3}$ and SOC in case-2products relative to case-1product (right column).

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[^0]:    ${ }^{\text {a }}$ The YRD inventory was used for the YRD region and was divided into 13 sources (ship, nonroad transport, road transport, industry, industrial solvent use, residential solvent use, agriculture, residential, industry boiler and kiln, power, dust, oil storage and transportation, gasoline vehicle evaporative emissions), among which only sources with none zero mass ratios of Nap and MN were listed in the Table. ${ }^{\mathrm{b}}$ The MEIC inventory was used for other regions and was divided into 5 sources. ${ }^{\mathrm{c}}$ The ratios of Nap and MN for wildfire emissions in the YRD region were the same as in other regions.

