Non-negligible secondary contribution to brown carbon in autumn and winter: inspiration from particulate nitrated and oxygenated aromatic compounds in urban Beijing

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Abstract

Nitrated aromatic compounds (NACs) and oxygenated derivatives of polycyclic aromatic hydrocarbons (OPAHs) play vital roles within brown carbon (BrC), influencing both climate dynamics and human health to a certain degree. The concentrations of these drug classes were analyzed in $PM_{2.5}$ from an urban area in Beijing during the autumn and winter of 2017 and 2018. There were four heavy haze episodes during the campaign; two of which happened prior to heating whereas the other two during heating. During the entire course of sampling, the mean total concentrations of the nine NACs and the eight OPAHs were 1.2-263 and 2.1-234 ng m^{-3}, respectively. The concentrations of both NACs and OPAHs were approximately 2-3 times higher in the heating period than before heating. For NACs, the relative molecular composition did not change significantly before and during heating, with 4-nitrocatechol and 4-nitrophenol demonstrating the highest abundance. For OPAHs, 1-Naphthaldehyde was the most abundant species before and during heating, while the relative proportion of Anthraquinone increased by more than twice, from 13% before heating to 31% during the heating. In Beijing's urban area during autumn and winter, significant sources of NACs and OPAHs have been traced back to automobile emissions and biomass burning activities. Interestingly, it was observed that the contribution from coal combustion increased notably during heating. It is worth noticing that the secondary generation of BrC was important throughout the whole sampling period, which was manifested by the photochemical reaction before heating and the aqueous

reaction during heating. It was further found that the haze in autumn and winter was nitrate-driven before heating and SOC-driven during heating, and the secondary formation of BrC increased significantly in pollution events, particularly during heating.

1 Introduction

As an important light-absorbing material, brown carbon (BrC) has garnered increasing attention in recent years (Jiang et al., 2023; Song et al., 2022; Zhang et al., 2021b; Liu et al., 2023; Ren et al., 2023; Ren et al., 2022; Chen et al., 2022). BrC could not only directly absorb solar energy, but also indirectly contribute to climate change by promoting the evaporation of water and the dispersal of clouds (Laskin et al., 2015; Huang et al., 2018). BrC also has potential adverse effects on human health on account of its strong mutagenic, cytotoxic, and carcinogenic properties (Teich et al., 2016).

Primary as well as secondary sources contribute to the atmospheric accumulation of BrC (Zhu et al., 2021). Direct emissions of primary BrC come from burning biomass and combustion of fossil fuels (Ni et al., 2021; Wang et al., 2020a; Lu et al., 2019a; Lu et al., 2019b). Secondary BrC in the atmosphere is produced from oxidation and aging processes (Wang et al., 2019; Wang et al., 2020c; Cheng et al., 2021; Jiang et al., 2023; Cai et al., 2022). Toluene, phenol, benzene, and other aromatic hydrocarbons can be oxidized to produce nitrophenol or nitrocatechol by NO_3 or OH radical vapor phase in the presence of NO_x (Olariu et al., 2002; Sato et al., 2007; Iinuma et al., 2010; Ji et al., 2017). VOCs can be oxidized to produce nitro-aromatic hydrocarbons when emitted

3

59　during biomass combustion and pyrolysis (such as cresol, catechol, methyl catechol,

60　etc.) (Iinuma et al., 2010; Claeys et al., 2012; Finewax et al., 2018). Research on the

61　source analysis of brown carbon (BrC) frequently focuses on examining two key

62　constituents: the carbon component within humic-like substances (HULIS-C) and

63　water-soluble organic carbon (WSOC). These components are often studied to

64　understand the origins and properties of BrC in various environmental contexts.

65　Secondary generation and burning of biomass are the two main sources of HULIS in

66　Guangzhou and Shanghai (Fan et al., 2016; Zhao et al., 2016). In comparison with the

67　water-insoluble BrC in the winter, the contribution of non-fossil sources (for instance

68　burning biomass) to water-soluble BrC was sometimes as high as 70% or more (Liu et

69　al., 2018; Song et al., 2018). Coal combustion is presumably a significant source of

70　HULIS in the winter, in addition to burning biomass and secondary generation (Tan et

71　al., 2016). According to multiple studies conducted in Beijing, the primary contributor

72　to WSOC is secondary generation, accounting for 54% of its composition. Following

73　this, biomass burning contributes approximately 40%, while other primary emission

74　sources contribute a smaller proportion, making up only 6% (Du et al., 2014). In Beijing,

75　the percentages of biomass burning, coal combustion, and secondary generation that

76　contribute to atmospheric HULIS are 47%, 15%, and 39%, respectively. The primary

77　origins of HULIS show minimal association with motor vehicles and industrial

78　emissions (Li et al., 2019). According to Ma et al. (2018), secondary generation is

79　responsible for over 50% of HULIS in the non-heating season. Biomass burning

80　represents 21% of the HULIS content during this period. However, in the heating

81   season, approximately 40% of HULIS originates from biomass burning, while the

82   remaining 60% is contributed by diverse combustion sources like coal burning, waste

83   incineration, and vehicular emissions. Within this season, secondary generation

84   accounts for about 19% of the HULIS content (Ma et al., 2018).

85   The research suggests that various sources contribute to BrC, but their relative

86   impact varies depending on time and location. As a result, the chemical makeup, light

87   absorption characteristics, and concentrations of BrC show considerable variability.

88   This variability poses challenges in accurately assessing and forecasting the influence

89   of these sources on radiation and climate changes (Wang et al., 2020b; Yan et al., 2018;

90   Laskin et al., 2015). However, until recently, there was only a limited volume of

91   research pertaining to the sources and pathways of BrC leading to their generation in

92   the densely populated city environment. NACs and OPAHs are the primary focus of

93   this study because several studies have noted that nitrogen-containing aromatics,

94   polycyclic aromatic hydrocarbons (PAHs), and their derivatives are significant BrC

95   chromophores (Huang et al., 2018; Wu et al., 2020; Liu et al., 2023; Wang et al., 2020b;

96   Xie et al., 2017). The average contribution of OPAHs (five species) to the solar-

97   spectrum weighed absorption coefficient of water-insoluble BrC in summer is

98   0.51±0.28% during daytime and 0.34±0.19% during nighttime. The contribution of

99   NACs to light absorption of water-soluble BrC is on average 2.5 times higher during

100   nighttime (3.47±2.03%) than during the day (1.41±0.29%) in winter, and the fractions

101   are much higher in winter than in summer (0.12±0.03%) (Li et al., 2020a). It is well-

102   established that residential heating plays a significant role in the substantial increase of

anthropogenic pollutant emissions during the winter season. There is a substantial rise

in the emission of aromatics-derived secondary organic aerosol from autumn to winter

(Ding et al., 2017), and particle BrC is often detected especially in haze periods (Liu et

al., 2023). This study was conducted in the autumn and winter of 2017 and 2018 in

Beijing. Nine NACs and eight OPAHs were measured in $PM_{2.5}$ samples, with a focus

on examining their sources, compositions, and concentration variations under various

air conditions. Specifically, emphasis was placed on investigating the contribution of

secondary generation to these two typical BrC species, particularly their involvement

in particle pollution processes during autumn and winter.

2 Materials and Methods

2.1 Field observations

PM2.5 was sampled at a height of 10m on the rooftop of a building at the Chinese

Research Academy of Environmental Sciences (CRAES), Beijing, China (40º02´N,

116º24´E). Using a high-volume sampler (1.13 m^3 min^{-1}, Thermofisher Co., USA),

$PM_{2.5}$ specimens were collected in the autumn and winter of 2017 and 2018. The

sampling process was executed from 8:00 to 19:30 during the day and from 20:00 to

7:30 in the subsequent morning. The specimens and blanks were gathered using a pre-

combusted quartz fiber filter (at 450 °C for 6 h). A total of 4 field blanks and 122 PM2.5

samples were acquired. Individual filters were sealed in a bag made from an aluminum

foil bag before sampling and analysis and placed in a freezer set at a temperature of -

123 20 °C.

124 Using automatic equipment (CRAES Supersite for Comprehensive Urban Air

125 Observation and Research), meteorological parameters such as air temperature (T, °C)

126 and relative humidity (RH, %) along with gaseous pollutants (SO_2, NO_2, O_3, and CO)

127 were observed and measured at the same time.

128 **2.2 Chemical analysis**

129 The present study employed a pre-treatment comprising ultrasonic extraction and

130 derivatization in an attempt to analyze the organic species in the specimens. The details

131 of specimen extraction and derivatization have already been published (Wang et al.,

132 2009; Ren et al., 2021; Ren et al., 2023). In brief, filter aliquots were sectioned and

133 extracted with a methanol and dichloromethane (1:2 v/v,) mixture. Following the

134 concentration of the extracts to dryness, derivatization was carried out using a mixture

135 of N, O-bis-(trimethylsilyl) trifluoroacetamide [BSTFA+TMCS, (99:1), v/v] and

136 pyridine (5:1, v/v). Lastly, the derivatized samples were examined using gas

137 chromatography coupled with a mass spectroscopy detector (GC/MS: HP 7890A, HP

138 5975C, Agilent Co., USA). The extraction and derivatization methods described above

139 allowed for the simultaneous measurement of the samples' polar and non-polar

140 constituents.

141 Given that OPAHs and NACs were the main points of focus, this study

142 investigated a total of eight OPAHs and nine NACs. The nine NACs included 2,4-

143 dinitrophenol (2, 4-DNP), 4-nitrophenol (4NP), 3-methyl-4-nitrophenol (3M4NP), 4-

144 nitrocatechol (4NC), 4-methyl-5-nitrocatechol (4M5NC), 4-nitroguaiacol (4NGA), 5-

145 nitroguaiacol (5NGA), 3-nitro-salicylic acid (3NSA), and 5-nitro-salicylic acid (5NSA),

146 while the eight OPAHs encompassed 9-fluorenone (9-FO), benzanthrone (BZA), 1-

147 Naphthaldehyde (1-NapA), anthraquinone (ATQ), 1,4-chrysenequione (1,4-CQ),

148 benzo(a)anthracene-7,12-dione (7,12-BaAQ), 5,12-naphthacenequione (5,12-NAQ)

149 and 6H-benzo(cd)pyrene-6-one (BPYRone).

150   The elemental carbon (EC) and organic carbon (OC) content of individual $PM_{2.5}$

151 filter samples were analyzed using an Atmoslytic Inc. DRI model 2001 Carbon

152 Analyzer. This analysis followed the Interagency Monitoring of Protected Visual

153 Environments (IMPROVE) thermal/optical reflectance (TOR) protocol, involving the

154 examination of a 0.526 cm^2 punch from each specimen. The specifics of the above-

155 described techniques have been documented in literature (Li et al., 2016; Ren et al.,

156 2021).

157 **2.3 Evaluation of secondary BrC**

158   In this study, the contributions of secondary oxidation to the detected NACs and

159 OPAHs were evaluated by using a CO-tracer method, which is comparable to the EC-

160 tracer used for secondary OC quantification. Various methodologies have been

161 similarly adopted successfully in other studies (Liu et al., 2023; Cai et al., 2022).

162 Equation 1 and Equation 2 were respectively used to evaluate the secondary formation

163 of NACs and OPAHs.

164     $$[NACs]_s = [NACs]_t − ([NACs]_t /[CO])_{pri} \times [CO] \tag{1}$$

165     $$[OPAHs]_s = [OPAHs]_t − ([OPAHs]_t /[CO])_{pri} \times [CO] \tag{2}$$

166 $[NACs]_s$ and $[NACs]_t$ in Equation 1 refer to the NACs concentration produced by

167   secondary oxidation and the total amount of NACs, respectively. ([NACs] /[CO]) $_{pri}$

168   represents the primary emission ratio of NACs in relation to combustion. This

169   calculation assumes that the primary source was predominant during the period, with

170   minimal secondary production. The ([NACs] /[CO]) $_{pri}$ was calculated in this work by

171   fitting the 15% lowest [NACs] $_{t}$ /[CO] ratios observed during the entire sampling

172   duration. In equation 2 the concentration of OPAHs produced by secondary oxidation

173   and the total observed OPAHs are denoted by [OPAHs]$_s$ and [OPAHs]$_t$ respectively.

174   The concentration of CO is denoted by [CO], while the primary emission ratio of

175   OPAHs in relation to combustion is represented by ([OPAHs] /[CO]) $_{pri}$, which was

176   calculated by fitting the lowest 15% [OPAHs]$_t$ /[CO] ratios observed in the entire

177   sample interval.

178   **3 Results and discussion**

179   **3.1 Variations in major components of PM$_{2.5}$ with respect to meteorological**

180   **conditions and gaseous pollution**

181   Based on the Beijing heating time, the entire period of the study was divided into

182   two phases: before heating (18 October to 14 November 2017) and during heating (15

183   to 23 November 2017; 23 December 2017 to 17 January 2018). Table 1 and Fig. 1

184   present the temporal fluctuations in meteorological factors, gaseous pollutant

185   concentrations, and the main PM$_{2.5}$ components in the two phases. The temperature (T)

186   and relative humidity (RH) were higher before heating (11 ± 3.8 °C and $49 \pm 26\%$) than

187   during heating (1.9 ± 4.4 °C and 23 ± 15%), with average values amounting to 5.9 ±

188   5.9 °C and 35 ± 25%, respectively. SO_2 concentrations during heating (4.3 ± 1.5 ppb)

189   were more than twice that before heating (2.1 ± 0.8 ppb), presumably because of the

190   increase in household coal burning for heating. The levels of NO_2 and NO remained

191   consistent before and during heating, suggesting that these pollutants were minimally

192   impacted by heating and were primarily influenced by mobile sources in Beijing. This

193   pattern seems to remain stable in the short term.

194   Fig.2 shows the variation in the chemical makeup of $PM_{2.5}$ in the whole sampling

195   period, before and during hearting, respectively. Secondary inorganic aerosols (SIA, i.e.

196   SO_4^{2-}, NH_4^+, and NO_3^-) were identified as the leading constituents of $PM_{2.5}$, followed

197   by OM (1.6 times OC), with an average of 31.5% and 20.4% in the whole sampling,

198   respectively (Fig. 2a). Even though the $PM_{2.5}$ concentrations remained relatively stable

199   during this period (as indicated in Table 1 and Fig. 2), there were significant changes

200   observed in the concentrations of SIA and OM, as well as their relative contributions to

201    $PM_{2.5}$. SIA accounted for 41.9% of $PM_{2.5}$ before heating, which notably decreased to

202   23.1% during heating. This decline was primarily evident in the reduction of NO_3^-.

203   SO_4^{2-}, NO_3^-, and NH_4^+ were measured at 5.5, 16, and 5.4 µg m^{-3}, respectively, with

204   constituted 8.7%, 24.7%, and 8.5% of $PM_{2.5}$ before heating (Fig. 2b). And their

205   concentrations decreased to 4.3, 6.8, and 4.2 µg m^{-3} (Table 1). The relative contributions

206   of NO_3^- to $PM_{2.5}$ dropped dramatically to 10.3% during heating, amounting to a drop

207   of nearly 60%. Both SO_4^{2-} and NH_4^+ experienced a roughly 25% decrease in their

208   relative contributions to $PM_{2.5}$, as illustrated in Fig. 2c. The relative abundance of OM

to PM$_{2.5}$ increased from 18.6% before heating to 21.9% during heating, with the average

mass concentration of OC showing an increase from 7.4 to 9.1 μg m^{-3} in the

corresponding duration. The OC/EC ratio also increased by 63% from 2.7 ± 3.3 before

heating to 4.4 ± 3.7 during heating. These significant changes in SIA and OM, including

concentrations and the relative contributions to PM$_{2.5}$, showed that primary organic

aerosols and/or VOCs emissions were the leading contributors during the heating

seasons due to household heating (Tan et al., 2018). It is important to note that the

decrease proportion of SIA in PM$_{2.5}$ may also be affected by meteorological conditions

during heating, which ofter influencing their production (Wang et al., 2016; An et al.,

2019). For example, increased sulfate is often accompanied by a high RH in the urban

atmosphere and the reaction rate of nitrate formation is accelerated by increasing RH

(Zhang et al., 2015; Sun et al., 2013). However, the RH were lower during than before

heating. Aside from coal combustion, the rise in mass concentrations of K$^+$ and Cl$^-$

indicated additional burning activities occurring during heating, such as biomass

burning (Bai et al., 2023; Li et al., 2022).

Table 1. Gaseous pollution concentrations and meteorological parameters and chemical constituents of PM$_{2.5}$ during the sampling periods in Beijing.

	The whole sampling N=122	Before heating period 18/10–14/11, 2017 N=56	During heating period 15/11–23/11, 2017 23/12, 2017–17/1, 2018 N=66
Meteorological parameters			
Temperature, ℃	5.9 ± 5.9 ((-7.5) – 16)	11 ± 3.8 (1.2 – 16)	1.9 ± 4.4 ((-7.5) – 11)
Relative humidity, %	35 ± 25 (7.1 – 99)	49 ± 26 (11 – 99)	23 ± 15 (7.1 – 67)
Gaseous pollutants, ppb			
SO$_2$	3.2 ± 1.6 (1.1 – 7.9)	2.1 ± 0.8 (1.1 – 4.8)	4.3 ± 1.5 (2.2 – 7.9)

NO$_2$	26 ± 13 (4.6 − 56)	25 ± 11 (4.6 − 43)	26 ± 14 (5.5 − 56)
NO	26 ± 28 (2.4 − 136)	28 ± 30 (2.4 − 136)	25 ± 26 (2.7 − 116)
CO	0.64 ± 0.55 (0.03 − 2.7)	0.81 ± 0.42 (0.12 − 1.6)	0.50 ± 0.61 (0.03 − 2.7)
Major components of PM$_{2.5}$, μg m^{-3}			
PM$_{2.5}$	65 ± 40 (6.1 − 195)	64 ± 39 (6.1 − 175)	66 ± 41 (8.6 − 195)
OC	8.3 ± 5.0 (0.99 − 26)	7.4 ± 3.9 (1.0 − 18)	9.1 ± 5.8 (1.8 − 26)
EC	4.7 ± 4.7 (0.11 − 25)	4.9 ± 3.8 (0.11 − 17)	4.5 ± 5.3 (0.18 − 25)
OC/EC	3.7 ± 3.6 (0.96 − 21)	2.7 ± 3.3 (0.96 − 21)	4.4 ± 3.7 (1.0 − 17)
SO$_4^{2-}$	4.8 ± 4.2 (0.85 − 25)	5.5 ± 3.5 (0.86 − 13)	4.3 ± 4.7 (0.85 − 25)
NO$_3^-$	11 ± 14 (0.09 − 58)	16 ± 16 (0.09 − 58)	6.8 ± 8.8 (0.29 − 37)
NH$_4^+$	4.7 ± 4.9 (0.02 − 20)	5.4 ± 5.4 (0.02 − 20)	4.2 ± 4.5 (0.19 − 20)
K$^+$	0.43 ± 0.39 (0.02 − 2.2)	0.38 ± 0.27 (0.03 − 1.1)	0.48 ± 0.46 (0.02 − 2.2)
Cl$^-$	1.5 ± 1.6 (0.06 − 9.2)	1.0 ± 0.98 (0.06 − 4.5)	1.9 ± 2.0 (0.13 − 9.2)

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Time (YYYYMMDD; D: daytime, N: nighttime)
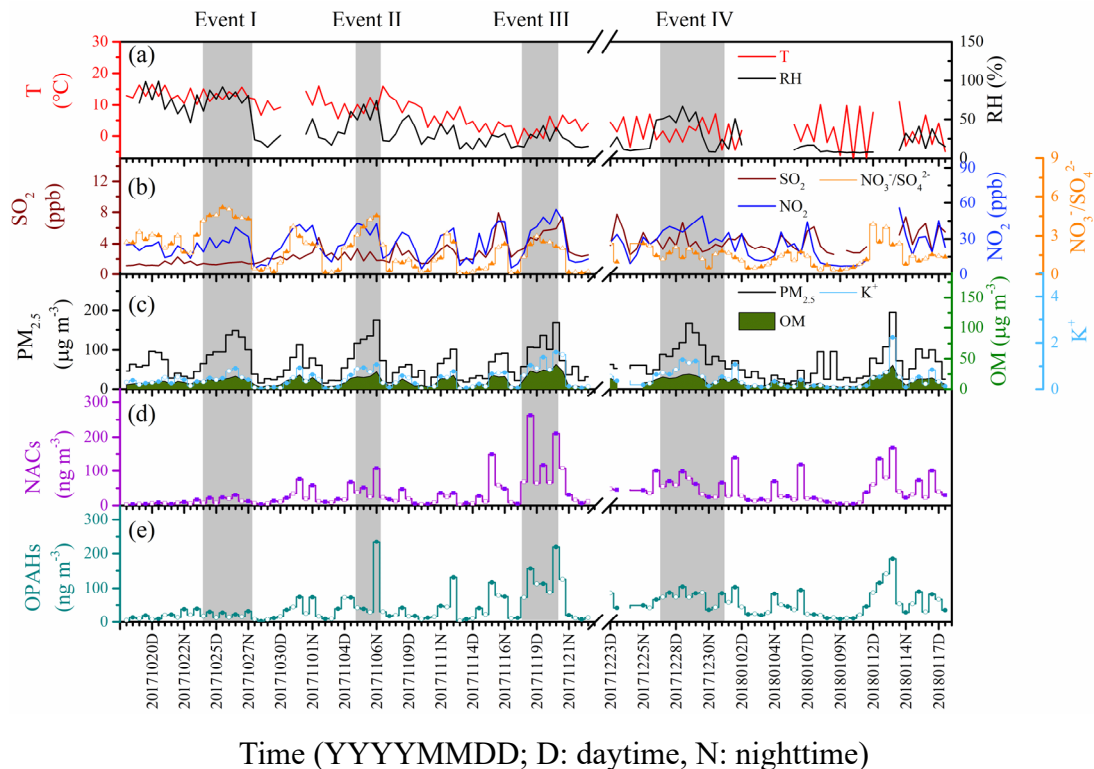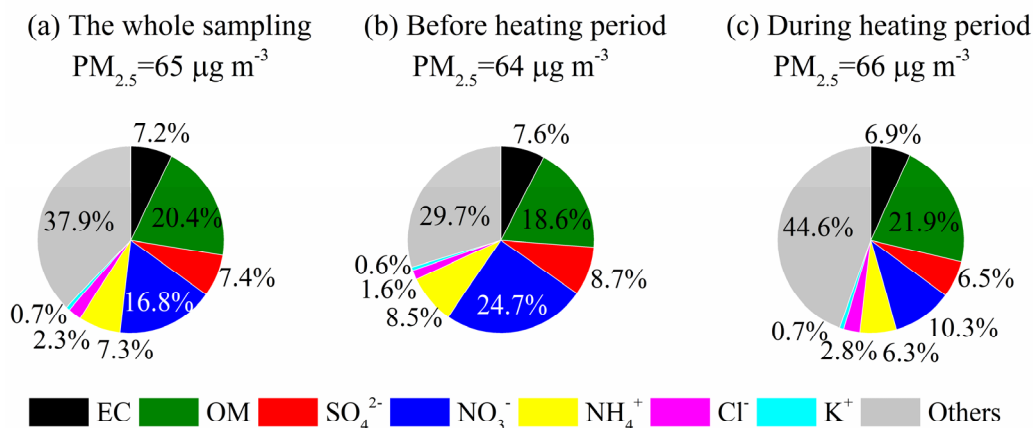
Fig.1 Time series of (a) RH and T, (b) SO$_2$ and NO$_2$, (c) PM$_{2.5}$, OM, and K$^+$, (d) NACs and (e) OPAHs in the autumn and winter of urban Beijing. (Daytime is denoted by empty marks and the nighttime is represented by solid marks in the panel b–e. The pollution episodes, with elevated concentrations of daily PM$_{2.5}$ more than 75μg m^{-3} in two successive days, have been marked in light gray).

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(a) The whole sampling
PM$_{2.5}$=65 µg m^{-3}

(b) Before heating period
PM$_{2.5}$=64 µg m^{-3}

(c) During heating period
PM$_{2.5}$=66 µg m^{-3}

EC ■ OM ■ SO$_4^{2-}$ ■ NO$_3^-$ ■ NH$_4^+$ ■ Cl$^-$ ■ K$^+$ ■ Others

Fig.2 Chemical constitution of PM$_{2.5}$ in the entire sampling period (a), before (b), and during (c) heating periods, respectively.

3.2 Concentration and composition variations of BrC compounds

This work quantified nine NACs and eight OPAHs. The corresponding concentrations and compositions have been presented in Fig. 3 and Table S1 (supporting information).

As seen in Table S1, during the entire sampling, the total concentrations of NACs ($\sum$9NACs) and their corresponding contribution to OM ($\sum$9NACs/OM) respectively averaged to 38 (1.2–263) ng m^{-3} and 0.25 (0.03–0.86) %. $\sum$9NACs and $\sum$9NACs/OM respectively averaged 53 (4.5–263) ng m^{-3} and 0.33 (0.09–0.86) % during heating, both values are two times higher in magnitude in comparison to those measured before heating (averaged 20 (1.2–108) ng m^{-3} and 0.15 (0.03–0.4) %, respectively). $\sum$9NACs exhibited a nighttime increase, reaching approximately twice the levels observed during daytime throughout the entire campaign (Fig. 3a). The observed difference between day and night is consistent with our previous research (Ren et al., 2022). However, the relative molecular composition of the total nine NACs in PM$_{2.5}$ did not manifest any

significant change (Fig.4a, c), 4-Nitrophenol (4NP) was found to have the highest concentration among all species, accounting for 44% and 42% of the total NACs before and during heating, followed by 4-nitrocatechol (4NC) which accounted for 21% before heating and 23% during heating. These findings align with the dominant species observed in previous studies (Ren et al., 2022; Ren et al., 2023; Li et al., 2020b). However, the values were much higher in comparison to those found in our earlier work (Ren et al., 2022) at the same sample site during the spring (8.6 (0.48–27) ng m^{-3}) and summer (8.5 (1.0–16) ng m^{-3}). It's plausible that seasonal variations in NACs are linked to emission sources, formation pathways, and weather conditions. In this study, the overall abundance of the $\sum$9NACs appeared to align closely with measurements from earlier studies conducted during winter in Beijing (74 $\pm$ 51 ng m^{-3} in winter, 20 $\pm$ 12 ng m^{-3} in autumn) (Li et al., 2020b) and Jinan (48 $\pm$ 26 ng m^{-3} in winter, 9.8 $\pm$ 4.2 ng m^{-3} in autumn,) (Wang et al., 2018), but are significantly higher than those measured for Xi'an (17 $\pm$ 12 ng m^{-3}) and Hong Kong (12 $\pm$ 14 ng m^{-3}) in winter (Wu et al., 2020; Chow et al., 2015). In contrast to studies conducted abroad, the levels of $\sum$9NACs in this particular study tended to be higher Germany showed 16 ng m^{-3}, while in the UK, levels were around 19 ng m^{-3}. Belgium recorded levels of 32 ng m^{-3} in winter and 13 ng m^{-3} in autumn. (Teich et al., 2017; Mohr et al., 2013; Kahnt et al., 2013). This indicates that it is urgent to further reduce the concentration of contaminant precursors in China.
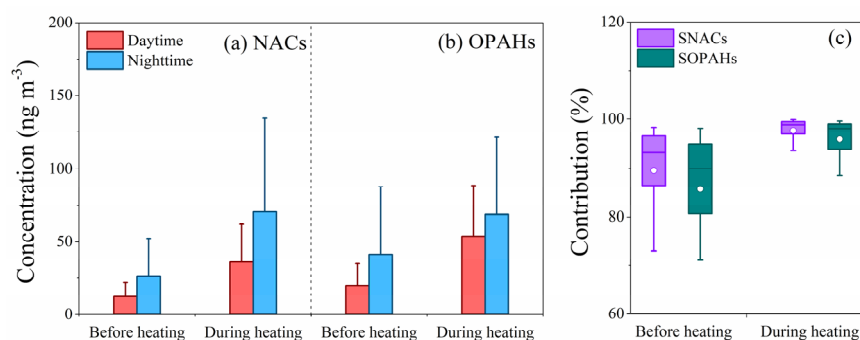
Fig.3 NACs and OPAHs concentrations (a,b) and contributions of secondary formation (SNACs, SOPAHs) to the total (c) before and during heating periods. The mean values are represented by the markers and the 25th and 75th percentiles are represented by whiskers.

Throughout the sampling, the total concentrations of OPAHs ($\sum$8OPAHs) averaged 47 (1.2–234) ng m^{-3} whereas the mean value for their total contribution to OM ($\sum$8OPAHs/OM) was 0.33 (0.06–0.81) % (Table S1). These values were both slightly higher than those of NACs in this work. $\sum$8OPAHs and $\sum$8OPAHs/OM respectively averaged 61 (6.9–218) ng m^{-3} and 0.40 (0.18–0.58) % during heating. These values are almost twice as much as those measured before heating (averaging 31 (2.1–234) ng m^{-3} and 0.24 (0.06–0.81) %, respectively). Like the $\sum$9NACs, the combined levels of $\sum$8OPAHs were higher during nighttime compared to daytime, averaging about twice as high before heating and 1.3 times during heating, as indicated in Fig. 3b. Among the eight OPAHs studied, 1-NapA constituted the highest proportion before (60%) and during (36%) heating. However, the relative proportion of ATQ more than doubled, increasing from 13% before heating to 31% during heating, as depicted in Fig. 4b and d. The average concentrations of $\sum$8OPAHs were higher than those recorded for other Chinese urban sites, including Guangzhou (23 ng m^{-3}) and Xi'an (54 ng m^{-3}) (Ren et al., 2017) as well as higher than those documented for the south (41.8 ng m^{-3}, traffic

site) (Alves et al., 2017) and central (~10 ng m^{-3}) European cities (Lammel et al., 2020).

The average concentrations of ∑8OPAHs were also higher than those recorded for

Mainz, Germany (0.047-1.6 ng m^{-3}) and Thessaloniki, Greece (0.86-4.3 ng m^{-3})

(Kitanovski et al., 2020).
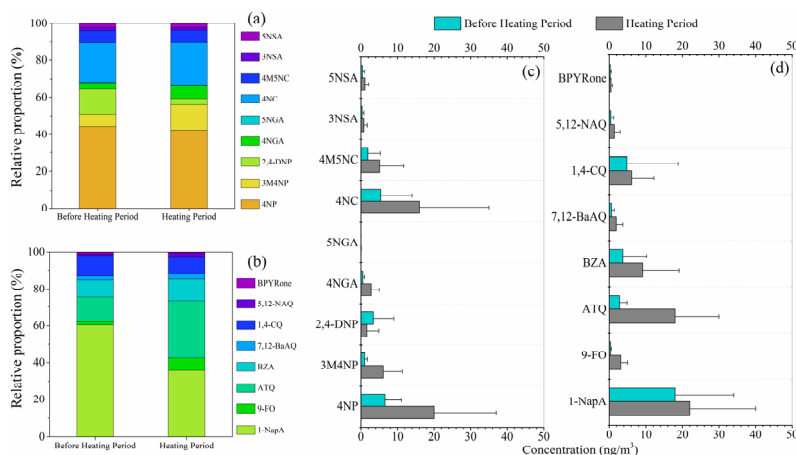
Fig.4 Comparison of measurements before and during the heating period at the urban site of Beijing, including (a) Relative proportion of NAC species, (b) Relative proportion of OPAH species, (c) NAC concentrations, and (d) OPAH concentrations. (4NP: 4-nitrophenol, 3M4NP: 3-methyl-4-nitrophenol, 2, 4-DNP: 2,4-dinitrophenol, 4NGA: 4-nitroguaiacol, 5NGA: 5-nitroguaiacol, 4NC: 4-nitrocatechol, 4M5NC: 4-methyl-5-nitrocatechol, 3NSA: 3-nitro-salicylic acid, 5NSA: 5-nitro-salicylic acid; 1-NapA: 1-Naphthaldehyde, 9-FO: 9-fluorenone, ATQ: anthraquinone, BZA: benzanthrone, 7,12-BaAQ: benzo(a)anthracene-7,12-dione, 1,4-CQ: 1,4-chrysenequione, 5,12-NAQ: 5,12-naphthacenequione, and BPYRone: 6H-benzo(cd)pyrene-6-one)

3.3 Sources and formation of BrC compounds

The relation between individual and total species and the associated pollutants—

levoglucosan, K$^+$, SO$_2$, NO$_2$, O$_3$, RH, and SIA—was examined according to the data

findings for the Pearson correlations shown in Table 2 (for NACs) and Table 3 (for

OPAHs) to provide additional clarity regarding the source and formation of NACs and

OPAHs. There were strong correlations between levoglucosan (an organic tracer

312   associated with biomass burning), K^+ (an inorganic tracer linked to biomass burning),

313   and NO_2 with total NACs and all identified NAC species, which indicated that both

314   automobile emissions and biomass burning played significant roles in the accumulation

315   of NACs in urban Beijing throughout the entire campaign. The correlation between

316   NACs and SO_2 was significantly higher during heating (r=0.275, $p<0.05$) compared to

317   pre-heating (r=0.210, $p>0.05$), suggesting that coal combustions play a more significant

318   role in NAC formation after heating commences.

319   In addition to these primary pollutants, NACs were also significantly correlated

320   with some secondary pollutants. Before heating, there existed a strong positive

321   association (r=0.692, $p<0.01$) between NACs and O_3. However, this association

322   changed considerably after heating, becoming notably negative (r=-0.303, $p<0.05$). The

323   negative correlation between them may be related to the meteorological conditions

324   during the sampling period, e.g. lower temperature and weak solar irradiation, which

325   suppress the photo-degradation process of NACs (Li et al., 2020a). NACs and RH

326   concurrently displayed a strong positive correlation (r=0.548, $p<0.01$) during heating.

327   Along with SO_4^{2-}, NO_3^-, and NH_4^+, total NACs also exhibited high positive correlation,

328   particularly while heating (r=0.373, $p<0.01$; r=0.504, $p<0.01$; r=0.513, $p<0.01$,

329   respectively). The overall concentrations of OPAHs and NACs throughout the

330   campaign exhibited substantial correlations (r=0.830, $p<0.01$ before heating; r=0.895,

331    $p<0.01$ during heating) (Table 3, Fig. S1). This suggests that their sources and/or

332   influencing variables were comparable. Specifically, throughout the entire campaign,

333   both total OPAHs and all identified OPAH species exhibited a strong correlation with

17

334 levoglucosan, K^+, and NO_2. This implies that automobile emissions and biomass

335 burning played significant roles as sources of OPAHs. OPAHs and SO_2 (r=0.365,

336 *p*<0.01) were determined to be more strongly correlated during heating than before

337 heating, suggesting the contribution of coal combustions to OPAHs becomes significant

338 during heating. Moreover, the correlation between OPAHs and O_3 was significantly

339 positive before heating (r=0.563, *p*<0.01), whereas it was significantly negative during

340 heating (r=-0.385, *p*<0.01). Furthermore, it was discovered that throughout the heating

341 phase, OPAHs and RH had a substantial positive correlation (r=0.578, *p*<0.01). Total

342 OPAHs also showed good correlations with SO_4^{2-}, NO_3^-, and NH_4^+, especially during

343 heating period (r=0.477, *p*<0.01; r=0.658, *p*<0.01; r=0.658, *p*<0.01; respectively).

344   The observed phenomena, involving photooxidation before heating and aqueous

345 reactions during heating, strongly suggest a significant role in the secondary creation

346 of BrC throughout the entire sampling period. Earlier studies have highlighted that in

347 certain regions, the primary mechanism driving the formation of nitro-aromatic

348 hydrocarbons involves the gaseous phase oxidation of VOC precursors from

349 anthropogenic sources, such as toluene and benzene (Olariu et al., 2002; Sato et al.,

350 2007; Yuan et al., 2016; Ji et al., 2017; Liu et al., 2023). According to a recent study,

351 for instance, NACs are mostly generated at a rural location on China's Chongming

352 Island through gaseous-phase photooxidation (Liu et al., 2023). Aqueous reaction is

353 also a key pathway for the formation of BrC (Zhang et al., 2020; Cheng et al., 2021;

354 Jiang et al., 2023). Previous studies suggested that aqueous-phase reaction is an

355 important partant pathway for secondary BrC formation during the winter season

356 (Zhang et al., 2020; Li et al., 2020a). Wang et al.'s field observations in urban Beijing

357 revealed that the aqueous reaction is a significant mechanism for the secondary

358 synthesis of nitro-aromatic hydrocarbons during summer temperatures with high

359 relative humidity (Wang et al., 2019). Furthermore, spherical primary OM particles (i.e.,

360 tarballs), which are mainly from residential coal burning especially during heating

361 period in the North China Plain, usually contain BrC species and their aqueous

362 formation could occur during the long-range transport (Zhang et al., 2021a; Zhang et

363 al., 2023).

Table 2 Correlations between NACs and meteorological parameters, gas pollutants, and aerosol components before (n=56) and during the heating period (n=66).

Before heating period		levoglucosan	K^+	SO_2	NO_2	O_3	RH	SO_4^{2-}	NO_3^-	NH_4^+
NACs	∑9NACs	0.897**	0.738**	0.210	0.714**	0.692**	0.170	0.359**	0.369**	0.190
	4NP	0.784**	0.699**	0.249	0.715**	0.649**	0.118	0.345**	0.372**	0.207
	3M4NP	0.752**	0.526**	0.290*	0.575**	0.511**	-0.011	0.184	0.191	0.029
	2,4-DNP	0.436**	0.353**	0.310*	0.463**	0.492**	-0.151	0.034	-0.048	-0.166
	4NGA	0.545**	0.361**	0.438**	0.560**	0.524**	-0.137	-0.016	-0.025	-0.131
	5NGA	0.582**	0.355**	0.114	0.343**	0.433**	0.005	0.120	0.139	0.044
	4NC	0.897**	0.748**	0.064	0.641**	0.617**	0.308*	0.448**	0.486**	0.325*
	4M5NC	0.885**	0.668**	0.076	0.579**	0.577**	0.252	0.364**	0.413**	0.261
	3NSA	0.791**	0.678**	0.129	0.553**	0.495**	0.214	0.457**	0.515**	0.331*
	5NSA	0.737**	0.596**	0.219	0.594**	0.553**	0.085	0.279*	0.316*	0.125
During heating period		levoglucosan	K^+	SO_2	NO_2	O_3	RH	SO_4^{2-}	NO_3^-	NH_4^+
NACs	∑9NACs	0.888**	0.786**	0.275*	0.481**	-0.303*	0.548**	0.373**	0.504**	0.513**
	4NP	0.812**	0.725**	0.262*	0.471**	-0.296*	0.586**	0.390**	0.489**	0.511**
	3M4NP	0.756**	0.655**	0.248	0.374**	-0.225	0.613**	0.318**	0.397**	0.462**
	2,4-DNP	0.537**	0.495**	0.280*	0.417**	-0.304*	0.199	0.136	0.247*	0.136
	4NGA	0.672**	0.406**	0.229	0.274*	-0.206	0.201	-0.047	0.074	0.081
	5NGA	0.275*	0.208	-0.028	0.190	0.026	-0.006	0.114	0.100	0.125
	4NC	0.894**	0.804**	0.248	0.454**	-0.290*	0.520**	0.378**	0.523**	0.530**
	4M5NC	0.882**	0.736**	0.246	0.434**	-0.244	0.430**	0.283*	0.421**	0.422**
	3NSA	0.788**	0.910**	0.348**	0.681**	-0.410**	0.577**	0.707**	0.888**	0.828**
	5NSA	0.820**	0.866**	0.268*	0.629**	-0.377**	0.599**	0.680**	0.846**	0.828**

**significant correlation at the 0.01 level;

*significant correlation at the 0.05 level;

364   From the above analysis, it is evident that there is a good correlation between these

365   two aromatic compounds and levoglucosan, as long-lived and inert chemicals in the

366   atmosphere (Cai et al., 2022). Therefore, it was not possible to determine with certainty

367   whether NACs and OPAHs originated predominantly from direct emission from the

368   biomass combustion or by secondary oxidation of the precursors produced as a result

369   of the process. Equations 1 and 2's outcomes indicated that in Beijing's urban areas

370   during fall and winter, NACs and OPAHs were predominantly of secondary origin.

371   Throughout the entire sampling period, secondary formation accounted for 17% to 99%

372   (average of 80%) of NACs and 8.9% to 99% (average of 73%) of OPAHs, as depicted

373   in Fig. 3c. Notably, the secondary fraction for OPAHs increased by 10.4% from 86% to

374   96%, while the secondary fraction for NACs rose by 8.9% from 90% before heating to

375   98% during heating. Earlier studies have highlighted the presence of significant levels

376   of secondary particle BrC during autumn and winter, particularly during haze periods

377   (Ding et al., 2017; Liu et al., 2023), and the results of this work corroborate well with

378   the earlier studies. Moreover, the good correlations between OPAHs and NACs with O_3

379   before heating and with RH during heating, confirm the importance of photochemical

380   and aqueous oxidation in these two different periods.

381

Table 3 Correlations between OPAHs and meteorological parameters, gas pollutants, and aerosol components before (n=56) and during the heating period (n=66).

Before heating period		$\sum$9NACs	levoglucosan	K^+	SO_2	NO_2	O_3	RH	SO_4^{2-}	NO_3^-	NH_4^+
	$\sum$8OPAHs	0.830**	0.865**	.605**	0.188	0.563**	0.563**	0.143	0.244	0.283*	0.139
OPAHs	1-NapA	0.844**	0.870**	.621**	0.211	0.640**	0.622**	0.174	0.213	0.238	0.096
	9-FO	0.775**	0.785**	.646**	0.283*	0.558**	0.573**	0.059	0.183	0.235	0.119
	ATQ	0.633**	0.694**	.497**	0.392**	0.477**	0.483**	-0.018	0.042	0.061	-0.024

BZA	0.686**	0.759**	.573**	0.232	0.537**	0.594**	0.110	0.177	0.206	0.117
7,12-BaAQ	0.821**	0.865**	.685**	0.189	0.591**	0.622**	0.187	0.325*	0.356**	0.224
1,4-CQ	0.636**	0.646**	.406**	0.041	0.303*	0.290*	0.102	0.259	0.309*	0.174
5,12-NAQ	0.694**	0.752**	.563**	0.227	0.511**	0.559**	0.091	0.184	0.214	0.110
BPYRone	0.827**	0.870**	.662**	0.131	0.590**	0.616**	0.226	0.345**	0.398**	0.255
During heating period	∑9NACs	levoglucosan	K^+	SO_2	NO_2	O_3	RH	SO_4^{2-}	NO_3^-	NH_4^+
∑8OPAHs	0.895**	0.931**	0.877**	0.365**	0.678**	-0.385**	0.578**	0.477**	0.658**	0.658**
1-NapA	0.752**	0.774**	0.659**	0.248	0.547**	-0.378**	0.332*	0.332**	0.498**	0.446**
9-FO	0.478**	0.457**	0.342**	0.305*	0.302*	-0.020	0.131	-0.013	0.167	0.250*
ATQ	0.780**	0.797**	0.815**	0.426**	0.668**	-0.372**	0.656**	0.466**	0.642**	0.684**
BZA	0.840**	0.881**	0.829**	0.332**	0.570**	-0.273*	0.577**	0.431**	0.568**	0.577**
7,12-BaAQ	0.801**	0.856**	0.896**	0.391**	0.633**	-0.299*	0.655**	0.531**	0.689**	0.710**
1,4-CQ	0.703**	0.780**	0.791**	0.244	0.597**	-0.282*	0.624**	0.560**	0.647**	0.675**
5,12-NAQ	0.777**	0.818**	0.869**	0.365**	0.601**	-0.293*	0.535**	0.444**	0.619**	0.614**
BPYRone	0.858**	0.857**	0.845**	0.339**	0.588**	-0.320*	0.478**	0.396**	0.612**	0.568**

Note: OPAHs label spans the rows from ∑8OPAHs to BPYRone.
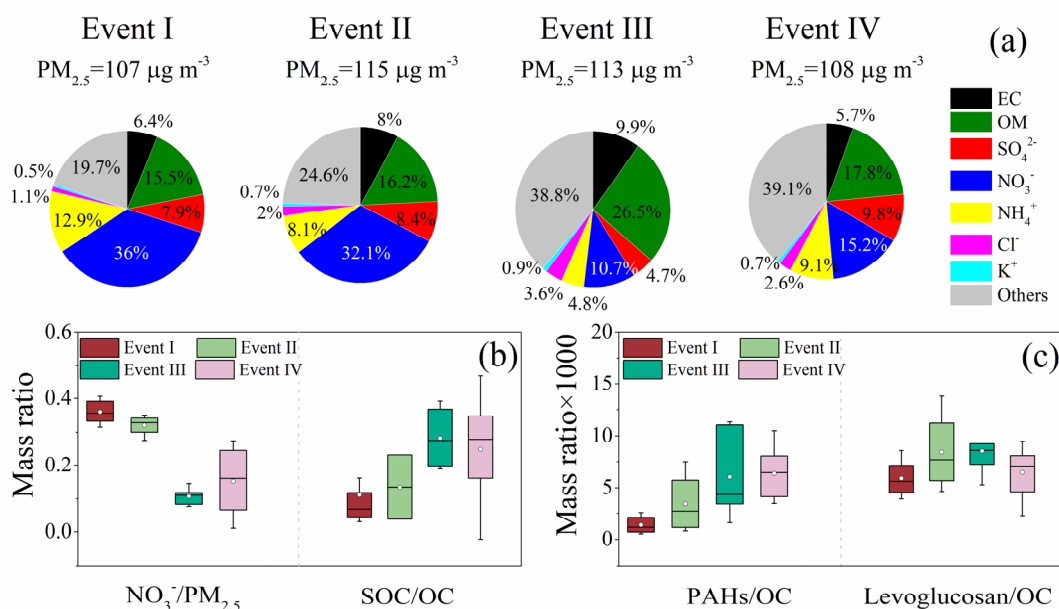
**significant correlation at the 0.01 level;

*significant correlation at the 0.05 level;

382

3.4 Different pollution characteristics in haze events

383

384 From Fig.1, it can be found that PM$_{2.5}$ shows four equivalent maxima lasting for

385 two to five days. Among the four pollution events, two occurred before heating (24-27,

386 October and 5-6, November) and the other two occurred during heating (18-20,

387 November and 27-31 December). PM$_{2.5}$ was significantly different in terms of its

388 chemical constituents before and during heating although the mass concentration of

389 PM$_{2.5}$ was rather similar (respectively averaging 107, 115, 113, and 108 µg m^{-3} for

390 Event I, II, III, and IV) (Fig. 5a). In the two events before heating, OM existed as the

391 second most dominant species in PM$_{2.5}$, with the respective relative abundance of 15.5%

392 and 16.2% in Events I and II. In contrast, OM surfaced as the most dominant species of

393 PM$_{2.5}$ during heating. The relative abundance of OM (26.5%) during Event III was

higher (17.8%) than that during Event IV (Fig.5a). The ratios of $NO_3^-/PM_{2.5}$ were higher

in Events I and II as compared to Events III and IV, with the ratios of SOC/OC showing

the opposite trend (Fig.5b), suggesting a significant increase in the concentration of

secondary organic compounds after heating. In the context of fossil fuel combustion,

PAHs serve as markers for coal burning, while levoglucosan acts as a significant tracer

for biomass smoke. Figure 5c shows that the ratios of PAHs to organic carbon mass in

$PM_{2.5}$ (PAHs/OC) were higher during Events III and IV compared to Events I and II.

This underscores the heightened emissions from household burning of coal for heating

purposes. Levoglucosan/OC, the mass ratio of levoglucosan to OC in $PM_{2.5}$, did not,

however, rise considerably over the same period (Fig. 5c), indicating a similar degree

of emissions from burning biomass before and during heating. This result was

consistent with our earlier research from the 2014 APEC meeting (Wang et al., 2017).

Fig.5 Comparative analysis of the chemical composition of $PM_{2.5}$ during four distinct events of air pollution. (a) Relative percentages of major species in $PM_{2.5}$; (b, c) mass ratios of the key species and organic tracers in $PM_{2.5}$. The mean values are represented by the markers and the 25th and 75th percentiles are represented by whiskers.

According to the majority of research, Beijing's haze is distinguished by intense secondary formation (Zhang et al., 2018; Xu et al., 2017; Sun et al., 2016; Guo et al., 2014). According to several research studies, organic materials (OM) predominates in the autumn and winter, while secondary SIA is the most prevalent species in the summer (Renhe et al., 2014). Additionally, according to a few investigators, SIA has a major role in wintertime pollution episodes (Guo et al., 2014; Wang et al., 2016). Furthermore, a recent investigation identified the species responsible for Beijing haze, and listed distinct haze-driving species operative over the year: The haze is primarily OM-driven during winter and late fall, nitrate-driven in early fall, sulfate-driven in summer, whereas it is driven primarily by nitrates during spring (Tan et al., 2018). Table 4 and Fig.5a depict that $PM_{2.5}$ was enriched with SIA especially NO_3^- during Events I and II, but enriched OM with higher levels of SOC was observed during Events III and IV. The findings strongly indicated that haze during fall and winter in urban Beijing was primarily influenced by nitrate before heating and shifted to being driven by SOC during heating. Table 4 illustrates that T and RH were notably higher during Events I and II compared to Events III and IV. These warmer and moister conditions favored photochemical oxidation, leading to an increased abundance of SIA during the same period. Home heating activities, such as burning residential coal, were increased during the heating period. This resulted in massive emissions of SO_2, NO_x, VOCs, and primary particles, all of which were conducive to the generation of SOC. As a result, during Events III and IV, SOC concentrations and relative abundances were higher than during Events I and II. Furthermore, Fig. 6a shows that the NACs/OC and OPAHs/OC ratios

were significantly higher in Events III and IV compared to Events I and II. Figure 6b

displays a parallel trend in the relative contributions of secondary formation for both

events, highlighting a notable increase in the secondary formation of BrC during

pollution events, particularly evident during heating periods.
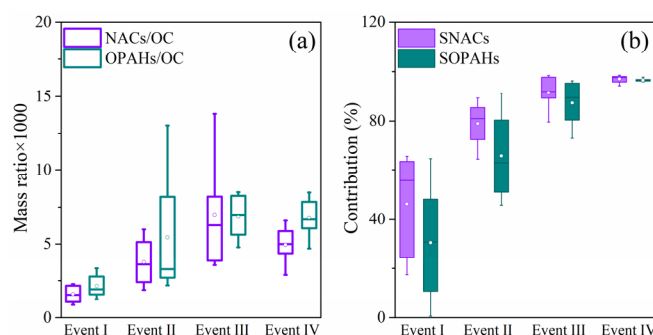
Fig. 6 Comparative analysis of the chemical composition of BrC during four distinct air pollution events. (a) Mass ratios of NACs and OPAHs to OC in $PM_{2.5}$. (b) Relative contributions of secondary formation (SNACs/OPAHs) to the total NACs/OPAHs in the fine particulate. The mean values are represented by the markers and the 25th and 75th percentiles are represented by whiskers.

Table 4. Meteorological parameters, chemical components ($\mu g\ m^{-3}$) of $PM_{2.5}$, and concentrations of gaseous pollutants (ppb) among four pollution episodes in Beijing.

	Before Heating Period		During Heating Period	
	Event I	Event II	Event III	Event IV
	24/10–27/10	5/11–6/11	18/11–20/11	27/12–31/12
	N=8	N=4	N=6	N=10
$PM_{2.5}$	107 ± 29	115 ± 48	113 ± 37	108 ± 35
Temperature, ℃	13 ± 1.6	8.8 ± 4.2	1.8 ± 2.6	1.8 ± 3.8
Relative humidity, %	79 ± 10	55 ± 26	29 ± 10	36 ± 22
SO_2	1.4 ± 0.1	1.9 ± 1.1	5.0 ± 1.0	4.0 ± 1.2
NO	27 ± 19	27 ± 15	51 ± 42	47 ± 31
NO_2	30 ± 6.7	33 ± 14	40 ± 12	37 ± 8.1
SIA [a]	62 ± 20	58 ± 22	23 ± 9.9	39 ± 30
NO_3^-	39 ± 13	38 ± 15	12 ± 5.2	18 ± 14
SOC [b]	1.9 ± 1.3	2.3 ± 1.8	7.0 ± 2.8	3.9 ± 2.5
NACs ($ng\ m^{-3}$)	17 ± 8.3	41 ± 36	131 ± 85	55 ± 27
OPAHs ($ng\ m^{-3}$)	21 ± 7.6	62 ± 85	127 ± 53	74 ± 23

[a] SIA: secondary inorganic aerosols (the sum of sulfate, nitrate, and ammonium).

[b] SOC: secondary organic carbon ([SOC]=[OC]-[EC]×([OC]/[EC])pri). [OC]/[EC]pri was estimated from the fitting of the minimum [OC]/[EC] ratio, assuming that the primary source dominated the period with minimal secondary formation. In this work, ([OC]/[EC])pri was estimated from the fitting of the lowest 15% [OC]/[EC] ratios during the whole sampling period.

4 Conclusions

The current study determined the concentrations of $PM_{2.5}$-bound nine NACs and eight OPAHs in autumn and winter in Beijing urban areas. The OPAHs and NACs concentrations were much higher during heating than before heating. These species have a distinct diurnal variation, with higher concentrations at night compared to day. 4-Nitrophenol, 4-nitrocatechol, and 1-Naphthaldehyde were the most abundantly existing species in the whole campaign.

The primary sources of NACs and OPAHs were biomass combustion and automobile emissions, with the secondary generation of BrC being the predominant contributor across the entire sampling period. Our results underscore the significant role of secondary generation in producing BrC, particularly its heightened contribution in pollution events during heating. A comparative analysis of the chemical constitution of $PM_{2.5}$ and BrC in four different haze events also revealed that the haze was caused by SOC during heating and by nitrate before heating in autumn and winter. Increased attention should be directed towards reducing the emissions of aromatic hydrocarbons and other anthropogenic volatile organic compounds (VOCs) when heating commences. This focus is crucial for effectively mitigating pollution and ensuring the preservation of human health. There is still only a limited volume of research on the molecular makeup of BrC, and further research is needed to identify more impactful chromophores at a molecular level. Additionally, a comprehensive exploration of the secondary generation pathways and key influencing factors of BrC through field

observations and laboratory simulations is essential. This investigation is crucial for accurately assessing the environmental and human health impacts of BrC.

Data availability

The field observational and the lab experimental data used in this study are available from the corresponding author upon request (Hong Li via lihong@craes.org.cn).

Author contributions

Yanqin Ren, Gehui Wang and Hong Li designed the research; Yanqin Ren, Yuanyuan Ji and Zhenhai Wu collected the samples; Yanqin Ren, Fang Bi and Hao Zhang conducted the experiments; Yanqin Ren and Gehui Wang analyzed the data, Yanqin Ren wrote the paper; Gehui Wang, Junling Li, Haijie Zhang and Hong Li contributed to the paper with useful scientific discussions and comments.

Competing interests

The authors declare that they have no conflict of interest.

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