

1 Diel and seasonal methane dynamics in the shallow and 2 turbulent Wadden Sea

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14
15 **Abstract.** The Wadden Sea is a coastal system fringing the land-sea borders of Denmark, Germany, and the
16 Netherlands. The Wadden Sea is extremely productive and influenced by strong variations in physical and
17 biological forcing factors that act on time scales of hours to seasons. Productive coastal seas are known to dominate
18 the ocean's methane emission to the atmosphere, but knowledge on controls and temporal variations of methane
19 dynamics in these vastly dynamic systems are scarce. Here we address this knowledge gap by measuring methane
20 inventories and methanotrophic activity at a temporal resolution of two hours over a time period of two days,
21 repeatedly during four successive seasons in the central Dutch Wadden Sea. We found that methane dynamics
22 varied between colder and warmer seasons, with generally higher water column methane concentrations and
23 methanotrophic activity in the warmer seasons. Efflux of methane to the atmosphere was, on the other hand, lower
24 in the warmer seasons because of lower wind speeds. On a diel scale, tides controlled methanotrophic activity,
25 which increased ~ 40 % at low tide compared to high tide. We estimate that methane oxidizing bacteria reduce the
26 methane budget of the Dutch Wadden Sea by only 2 %, while ~ 1/3 escapes to the atmosphere and ~ 2/3 are flushed
27 out into the open North Sea at ebb tide. Our findings indicate that tides play a key role in controlling methane
28 dynamics and methanotrophic activity and highlight the importance of high resolution and repeated sampling
29 strategies to resolve methane dynamics in fast-changing coastal systems.

30 1 Introduction

31 32 1.1 Methane and methane oxidation

33 Atmospheric methane (CH₄) concentrations have been increasing since industrial times, surpassing 1900 ppb in
34 2021 (Lan et al., 2022) and contributing more than 20 % of total radiative forcing in the atmosphere (Etminan et
35 al., 2016). Due to its relative short atmospheric lifetime of ~10 years (Canadell et al., 2021), reducing methane
36 emissions to the atmosphere could play a key role in global warming mitigation strategies. However,
37 implementation of such strategies requires a thorough understanding of methane sources and sinks. Anthropogenic
38 methane emissions (336 – 376 Tg y⁻¹) are rather well constrained and constitute ~60 % of the total atmospheric
39 budget (Saunois et al., 2020). Individual natural sources, on the other hand are associated with comparably large
40 uncertainties. This is particularly true for methane emissions originating from marine environments (5 to 28 Tg
41 CH₄ y⁻¹; (Rosentretter et al., 2021)).

42 The inner shelf (0 – 50 m water depth) only account for ~ 3 % of the global ocean surface but are a main source
43 of marine methane emissions to the atmosphere (Weber et al., 2019). In these shallow ecosystems, light availability
44 as well as terrestrial inputs of nutrients support a high diversity of producers and consumers that generate huge
45 quantities of organic matter (Philippart et al., 2009; Beck and Brumsack, 2012). Consequently, rates of organic
46 matter degradation, including methanogenesis in anoxic sediments are high, often leading to elevated levels of free
47 and dissolved methane in sediments and pore waters (Bange et al., 1994; Røy et al., 2008; Wu et al., 2015).
48 Transport of methane-rich porewaters and ebullition of methane bubbles, in return, lead to elevated methane
49 concentrations in the water column (Reeburgh, 2007; Grunwald et al., 2009; James et al., 2016). It is estimated
50 that ~ 5 % of shelf seas surface waters have methane concentrations above 100 nM (Weber et al., 2019).
51 Nevertheless, a substantial amount of dissolved methane is oxidized by aerobic methanotrophic bacteria (MOB),
52 which mediate the aerobic oxidation of methane (MOx) (Reeburgh, 2007):



54 Similar to other metabolic processes involving small molecules, MOx discriminates against isotopically heavy
55 methane (i.e. containing ¹³C and ²H (D) instead of ¹²C and ¹H) so that the residual methane pool successively
56 becomes ¹³C and D enriched as a result of ongoing MOx (Barker and Fritz, 1981; Whiticar, 1999).

57 MOB typically belong to the Gamma- (type I and type X), Alphaproteobacteria (type II), Verrucomicrobia and
58 members of candidate division NC10 (Hanson and Hanson, 1996; Knief, 2015). MOBs build a microbial methane
59 filter in the water column that functions as the ultimate sink for oceanic methane before reaching the atmosphere.
60 Yet, little is known about the controls and capacity of this microbial filter in the inner shelf ecosystems where the
61 vertical distance between the sedimentary source and the atmosphere is short. Factors such as oxygen (Boetius and
62 Wenzhöfer, 2013; Steinle et al., 2017) and methane availability (Mau et al., 2013; James et al., 2016) affect MOx,
63 but also increasing water temperatures play a role by impacting metabolic rates of MOB (He et al., 2012). The
64 capacity of the microbial methane filter in the water column is typically higher during extended periods of
65 continuity, i.e., when the water column is more stagnant (Steinle et al., 2015; James et al., 2016). This increases
66 the contact time of MOB with methane-rich waters so that the size of the MOB standing stock increases. However,
67 water mass movement induced by destratification, or seasonal winds, leads to shifting mixing regimes that disrupt
68 continuity on a seasonal scale (Gründger et al., 2021). On a daily scale, tides induce currents, which also disrupt
69 continuity and hence can affect MOx, too (Steinle et al., 2015). This disruption of continuity is particularly strong
70 in the extremely dynamic inner-shelf seas where rapid changes in environmental conditions can lead to rapid
71 changes in water column dynamics.

72 The Wadden Sea, a UNESCO heritage site that consists of the largest continuous tidal flat area worldwide (14.900
73 km²), is an extremely dynamic system, with major hydrological changes occurring at seasonal to diel time scales.
74 The Wadden Sea stretches for about 500 km along the coast of the Netherlands, Germany and Denmark. Here, we
75 investigated methane dynamics in the Dutch part of the Wadden Sea, that is separated from the North Sea by five
76 barrier islands (Fig. 1). Our aim was to temporally resolve methane dynamics from an hourly to a seasonal scale
77 to determine key controls on methane dynamics and to establish a methane budget for the Dutch Waddensea

Deleted: , but large uncertainties exist around the strength of individual natural sources (Saunois et al., 2020). Especially, methane emissions from marine environments

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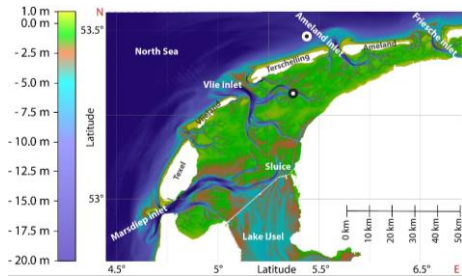
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89 **Figure 1.** Bathymetry of the western sector of the Dutch Wadden Sea between the Marsdiep and Friesche inlet (modified from
 90 Materić et al., 2022). Tidal inlets between barrier islands facilitate water exchange with the open North Sea. The time-series
 91 station is located south of the island Terschelling (black mark; 53°19.015 N, 5°22.071 E). The offshore reference station is
 92 located 8 km north of Terschelling (white mark; 53°29.190 N, 5°21.449 E).
 93

94 **2 Materials and methods**

95 **2.1 Experimental design**

96 A chain of 5 barrier islands (located 5 to 30 km offshore) shelters the Dutch Wadden Sea from waves and strong
 97 westerly winds. Between these barrier islands **and with the rhythm of the tides**, large volumes of water are
 98 transported in and out the Dutch Wadden Sea through deep tidal inlets, such as the Marsdiep (most western point
 99 of the Dutch Wadden Sea) and the Vlie inlet (Duran-Matute et al., 2014). Our fixed mooring station (53°19.015
 100 N, 5°22.071 E) is in a branch of the Vlie inlet between the island of Terschelling and the mainland, roughly in the
 101 middle of the Dutch Wadden Sea (Fig. 1). This location was chosen as it remains submerged at low tide and lays
 102 in-between the Wadden Sea's landward and offshore termination. The water flowing by this station thus equally
 103 integrates the tidal flat area, mostly during ebb tide, as well as the inflowing North Sea water during high tide.
 104 Also, the station was relatively far away from the port of Harlingen (~ 20 km) so that a potential influence of
 105 methane rich port waters is minimized. The reference station was located 8 km north of the island Terschelling in
 106 the North Sea (53°29.190 N, 5°21.449 E).

107 Samples were recovered with the R/V Navicula during 4 sampling campaigns, respectively in winter (19 February
 108 2019 – 21 February 2019), spring (23 April 2019 – 25 April 2019), summer (22 July 2019 – 24 July 2019) and
 109 autumn (11 November 2019 – 13 November 2019). During each campaign, we conducted hourly CTD casts with
 110 discrete water sampling over a two-day period. During CTD casts, water mass properties (temperature, salinity,
 111 **depth**) and oxygen concentrations were measured continuously using a Sea-Bird (SBE911) + conductivity-
 112 temperature-depth (CTD) system. Discrete water samples were recovered with Niskin bottles from 1 and 3 m
 113 water depth and, upon recovery, immediately sampled for subsequent analyses of water column constituents
 114 (methane concentrations, methane isotopic composition and methane oxidation rates).

115 Sediment samples were retrieved using a boxcorer, and upon recovery, subsampled with small pushcores (diameter
 116 7 cm, ~ 18 cm sediment recovery). Pushcores were subsampled for methane concentrations by taking every 2 cm
 117 5 mL of sediment that was quickly added to 60 mL glass bottles containing 30 mL of a saturated NaCl brine
 118 solution and the bottles were immediately sealed with butyl rubber stoppers. Atmospheric flask samples (250 ml)
 119 were taken hourly at ~ 10 m above the sea surface, in winter and spring. In summer and autumn, atmospheric
 120 methane concentrations were continuously measured using a cavity ringdown spectrometer (CRDS, Picarro model
 121 G2301).

122 **2.2 Dissolved methane concentrations and stable isotope ratios**

123 Dissolved methane concentrations were determined using a headspace (HS) technique (Green, 2005). In brief,
 124 immediately upon CTD recovery, 260 mL glass serum bottles were filled HS-free, closed with black-butyl rubber
 125 stoppers (Rubber B.V. the Netherlands) and crimp-top sealed. Next, we added a 5 mL N₂ headspace and fixed the
 126 sample with 5 mL NaOH solution (25 % w/v). HS methane concentrations of sediments and dissolved methane
 127 were measured in our home laboratories with a gas chromatograph (GC; Thermo Scientific FOCUS GC equipped
 128 with a Restek stainless steel column HS-Q 80/100 SS GEN config (length 2 m, 2 mm ID, 1/8 OD) with flame
 129 ionization detection). The instrument was calibrated with a certified 100 ppm methane standard (Scott Specialty
 130 Gases Netherlands B.V.).

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134 Similarly, seawater aliquots were taken for methane stable carbon and hydrogen isotope measurements, but these
 135 samples were fixed with 60 μl HgCl_2 (2 mM). A continuous flow isotope ratio mass spectrometry (CF-IRMS)
 136 system was used to quantify D- CH_4 in the gas phase (Thermo Delta Plus XL, Thermo Fisher Scientific Inc.,
 137 Germany) as described previously (Röckmann et al., 2016; Jacques et al., 2021). Isotopic values are represented
 138 in the delta notation against the international reference standard VSMOW (δD). To monitor precision and
 139 accuracy, sample measurements were alternated with measurements of an inhouse air standard (cross calibrated
 140 against certified reference standards) containing 1975.5 ppb methane with a $\delta\text{D} -90.81 \pm 1.1$ ‰. We constructed a
 141 two-endmember mixing model (Mariotti et al., 1981; Jacques et al., 2021) and a Rayleigh fractionation model.
 142 This was done to investigate whether enrichment of D in the residual methane was caused by MOx, which is
 143 known to discriminate against heavy isotopes (Barker and Fritz, 1981; Whiticar, 1999), or by mixing with
 144 comparably heavy atmospheric methane (see supplementary methods).
 145

146 2.3 Methane oxidation rate measurements

147 MOx was determined by ex-situ incubations with trace amounts of ^3H -labelled methane as described previously
 148 (Niemann et al., 2015). Briefly, aliquots from each Niskin bottle were filled HS-free in 20 mL glass vials in
 149 triplicate, sealed with grey-bromobutyl stoppers that are known to not hamper methanotrophic activity and
 150 amended with 5 μL of $^3\text{H}\text{-CH}_4/\text{N}_2$ (4.5 kBq, American Radiolabeled Chemicals, USA). ~~The samples were~~
 151 ~~incubated in a temperature-controlled incubator for 72 hours in the dark, maintaining in situ temperature~~
 152 ~~conditions.~~ Activities of residual C^3H_4 and the MOx product $^3\text{H}_2\text{O}$ were measured by liquid scintillation counting.

153 MOx first-order rate constant (k) was determined from the fractional tracer turnover (Reeburgh, 2007):

$$154 k = \frac{^3\text{H}_2\text{O}}{^3\text{H}_2\text{O} + \text{C}^3\text{H}_4} \times \frac{1}{t} \quad (2)$$

155
 156 where t is incubation time in days. k was corrected for (negligible) tracer turnover in killed controls (KC, fixed
 157 with 100 μl HgCl_2 directly after sampling) and multiplied with dissolved methane concentrations $[\text{CH}_4]$, yielding
 158 MOx:

$$159 \text{MOx} = (k - k_{\text{KC}}) \times [\text{CH}_4] \quad (3)$$

160
 161 **2.4 Diffusive fluxes of methane**
 162 The diffusive sea-air methane flux was calculated based on a boundary layer model that consider the relation
 163 between wind, temperature and methane concentrations in the atmosphere and a well-mixed surface water layer
 164 (Wanninkhof, 2014):

$$165 F = (p\text{CH}_{4w} - p\text{CH}_{4a}) K_0 k_{\text{CH}_4} \quad (4)$$

166 F denotes the diffusive methane flux, $p\text{CH}_{4a}$ and $p\text{CH}_{4w}$ (in atm) are the partial pressures of methane in the air and
 167 in the well-mixed surface water layer, respectively. $p\text{CH}_{4a}$ was measured with a Picarro G2301 gas concentration
 168 analyser on board. $p\text{CH}_{4w}$ was determined from surface water methane concentrations (see above). K_0 is the
 169 methane solubility in $\text{mol m}^{-3} \text{atm}^{-1}$ (Wiesenberg and Guinasso, 1979) and was calculated from temperature and
 170 salinity obtained from corresponding CTD casts. k_{ch_4} is the methane gas transfer velocity in m d^{-1} which was
 171 calculated using wind speed (U), the Schmidt number (Sc_{CH_4}) and the normalised gas transfer velocity (k_{660})
 172 according to (Wanninkhof, 2014):

$$173 k_{\text{CH}_4} = 0.251 U^2 \left(\frac{\text{Sc}_{\text{CH}_4}}{660} \right)^{-0.5} \quad (5)$$

174 Wind speed was measured on board at 10 m above sea level. The Schmidt number describes the ratio between
 175 kinematic viscosity of water and the gas diffusion coefficient, which relates the different k -values for different
 176 gases (Jähne et al., 1987; Wanninkhof, 2014).

177 2.5 Statistical analysis

178 A principal component analysis (PCA) was carried out to study the relationship between environmental variables
 179 and methanotrophic activity. The input variables for the PCA were temperature, salinity, density, k , MOx, and
 180 dissolved methane concentrations. Prior to running the PCA, the variables were centered and scaled. We utilized
 181 the R software and the ‘FactoMineR’ package (Lê et al., 2008) for the PCA analyses.

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185 **3 Results**

186 **3.1 Dynamics of sea water properties**

187 Water column temperature varied between seasons and ranged from 6.3 °C to 24 °C (Fig. 2, Table 1). A clear
188 distinction could be made between colder seasons (autumn and winter) in which temperature ranged from 6.3 °C
189 to 9.1 °C and warmer (spring and summer) seasons where temperatures ranged from 14.2 °C to 24 °C. Water
190 temperatures at the reference station were similar in winter (6.9 °C) but colder in spring (10.5 °C) and summer
191 (20.3 °C) and warmer in autumn (11.8 °C) when compared to the Wadden Sea.

192 **Table 1.** Average seawater temperature, salinity, and density at the time-series station (central Dutch Wadden Sea) and
193 reference station (offshore Terschelling, North Sea). For the time-series station, values are presented as the mean ± standard
194 deviation for the ~2 d measurement period during a given season. At the reference station, we only measured one CTD cast
195 per season.

	Autumn	Winter	Spring	Summer
Temperature (°C)	8.4 ± 0.4	6.7 ± 0.2	15.0 ± 0.6	22.1 ± 0.8
Salinity (psu)	22.4 ± 1.9	23.4 ± 1.8	31.3 ± 0.2	30.6 ± 0.6
Density (σ _t)	17.4 ± 1.4	18.3 ± 1.4	23.1 ± 0.2	20.8 ± 0.6
	Autumn Ref st.	Winter Ref st.	Spring Ref st.	Summer Ref st.
Temperature (°C)	11.8	6.8	10.4	20.0
Salinity (psu)	31.3	32.2	31.8	32.3
Density (σ _t)	23.7	25.2	24.4	22.7

196
197 On a diel scale, variation in water temperature were related to the tidal phase. In winter, spring, and summer,
198 maximum water temperatures were observed around low tide (LT, here defined as the time when we encountered
199 the lowest water depth during CTD casts, Fig. 2). This was 7.2 °C in winter, 17.3 °C in spring and 24.1 °C in
200 summer. Minimum water temperatures were around high tide (HT, high tide, here defined as the time when we
201 encountered maximum water depth during CTD casts, Fig. 2). This was 6.3 °C winter, 14.2 °C in spring, and 20.9
202 °C in summer. In autumn, this pattern was inverse with minimum water temperatures at LT (7.6 °C) and maximum
203 at HT (9.1 °C).

204 Like temperature, salinity differed strongly between colder (18 - 27 psu) and warmer seasons (29 - 32 psu; Fig. 2,
205 Table 1). Furthermore, salinity was higher during HT irrespective of season. Changes in density were caused by
206 salinity rather than temperature during all four seasons, with one exception in spring: after 28 hours of the time
207 series, salinity remained stable, but water temperatures decreased which lowered water density. Salinity levels at
208 the reference station in the North Sea were stable (31.3 – 32.3 psu) without obvious seasonal fluctuations.

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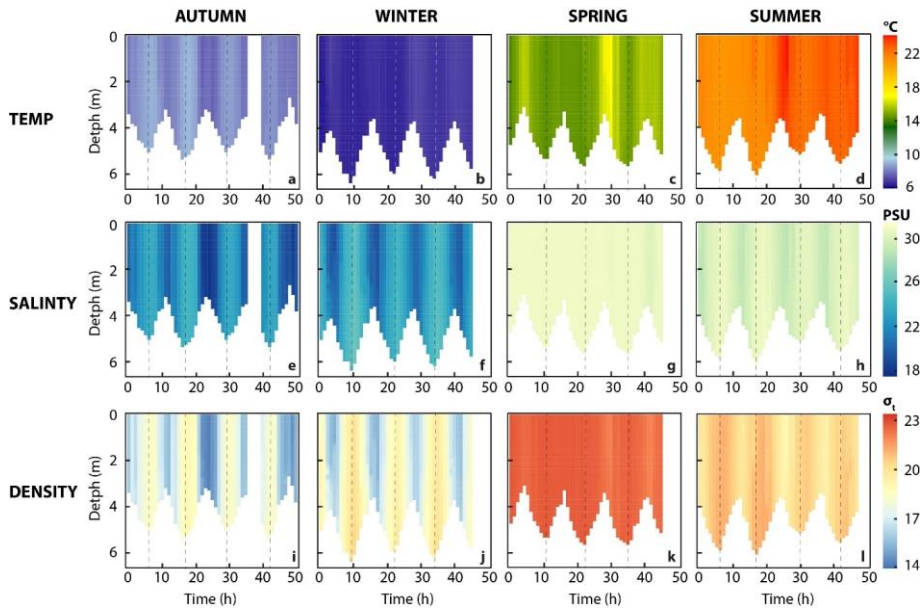
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225

226 **Figure 2.** Properties of sea water. (a-d) Spatiotemporal distribution of temperature, (e-h) salinity and (i-l) density. Dashed
 227 line indicates high tide.

228 3.2 Methane dynamics

229 3.2.1 Methane concentrations in the water column and in sediments

230 Water column methane concentrations showed a high degree of variability and were clearly distinguishable
 231 between the colder and warmer seasons (Fig. 3A-D, Table 2). We found a significant difference in average methane
 232 concentrations between 1 m (16.0 nM) and 3 m (17.6 nM) water depth in winter ($p \leq 0.007$, Welch's t-test). In
 233 autumn, methane concentrations were also lower at 1 m (15.5 nM) than at 3 m (16.2 nM) water depth, but the
 234 difference was not significant. However, it is noteworthy that the methane concentrations at the beginning of the
 235 time-series were around 35 nM and rapidly decreased to values below 15 nM within one day. During warmer
 236 seasons, average methane concentrations were similar at the surface and in deeper waters, i.e. 40.9 nM (1 m) and
 237 41.3 nM (3 m) in spring and 69.2 nM (1 m) and 69.4 (3 m) in summer. Methane concentrations at our reference
 238 station were ~ 3 nM in winter, spring, and autumn and ~ 6 nM in summer and thus far lower when compared to
 239 the Wadden Sea.

240 On a diel scale, methane concentrations varied during all seasons, roughly matching the tidal regime. In spring at
 241 LT, depth-averaged methane concentrations were 42.6 nM, but decreased by $\sim 25\%$ to 34.2 nM at HT. This pattern
 242 also occurred in autumn where methane concentrations decreased by 21% from 17.4 nM at LT to 14.4 nM at HT.
 243 In winter (14.7 nM at LT and 14.3 nM at HT) and summer (72.5 nM at LT and 71.3 nM at HT), the difference
 244 between LT and HT was smaller (Table 2).

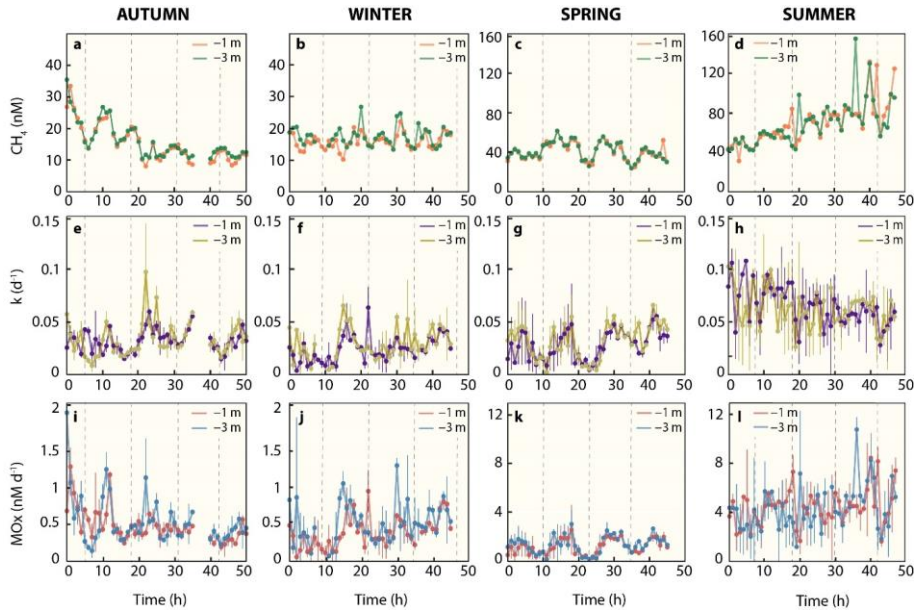
245 Sediment methane concentrations increased with depth during all seasons (Fig. S1 in the Supplement).
 246 Concentrations were similar in autumn (0.5 – 2.2 μM), winter (0.4 - 0.6 μM) and spring (0.5 - 0.9 μM) but in
 247 summer, we found highly elevated sediment methane concentrations ranging from 3.6 to 18.7 μM . The high
 248 concentrations in sediments during the summer season are in line with an increase in dissolved methane
 249 concentrations in the water column.

250

251 **Table 2. Methane dynamics in the Dutch Wadden Sea.** Average and standard deviation of methane concentrations, k , MOx
 252 and $\delta\text{D-CH}_4$ during four seasons in 2019. Values represent averages for 1 and 3 m water depth (averaged over the two-day time
 253 series recorded for each season) as well as for low and high tide only (averaged over depth). LT = minimal water depth during
 254 CTD casts, HT = maximum water depth during CTD cast. Average wind speed and methane efflux to the atmosphere are
 255 averaged over the two-day time series recorded for each season. ns = not sampled. [The reference station represents a single](#)
 256 [time point.](#)
 257

	Autumn	Winter	Spring	Summer
Methane concentration (nM)				
1 m water depth	15.5 ± 5.8	16.0 ± 2.4	40.9 ± 9.2	69.2 ± 21.4
3 m water depth	16.2 ± 5.7	17.6 ± 3.0	41.3 ± 8.9	69.4 ± 22.4
Low tide	17.4 ± 9.7	14.7 ± 2.1	42.6 ± 6.9	72.5 ± 36.1
High tide	14.4 ± 1.6	14.3 ± 0.6	34.2 ± 10.7	71.3 ± 27.4
Reference station	3.3	3.1	3.7	6.6
k (d⁻¹)				
1 m water depth	0.03 ± 0.01	0.02 ± 0.01	0.03 ± 0.01	0.07 ± 0.02
3 m water depth	0.03 ± 0.02	0.03 ± 0.01	0.03 ± 0.02	0.06 ± 0.02
Low tide	0.05 ± 0.01	0.03 ± 0.01	0.05 ± 0.01	0.08 ± 0.02
High tide	0.03 ± 0.01	0.03 ± 0.02	0.02 ± 0.01	0.06 ± 0.02
Reference station	0.01	0.0004	0.02	0.04
MOx (nM d⁻¹)				
1 m water depth	0.48 ± 0.22	0.39 ± 0.21	1.16 ± 0.61	4.41 ± 1.49
3 m water depth	0.54 ± 0.34	0.52 ± 0.27	1.33 ± 0.71	4.33 ± 1.84
Low tide	1.05 ± 0.48	0.47 ± 0.24	2.02 ± 0.42	5.24 ± 2.33
High tide	0.50 ± 0.16	0.43 ± 0.31	0.59 ± 0.19	4.23 ± 2.13
Reference station	0.03	0.001	0.07	0.23
$\delta\text{D-CH}_4$ (‰)				
1 m water depth	-219 ± 31	ns	ns	-250 ± 17
3 m water depth	-224 ± 27	ns	ns	-250 ± 14
Low tide	-208 ± 41	ns	ns	-227 ± 1
High tide	-227 ± 13	ns	ns	-265 ± 3
Methane sea-air flux ($\mu\text{mol m}^{-2} \text{d}^{-1}$)				
Wind speed (m s ⁻¹)	8.0 ± 2.1	8.3 ± 1.4	7.9 ± 2.7	3.8 ± 1.6
Methane flux	40.2 ± 28.1	38.7 ± 14	144.8 ± 98	72.9 ± 52
Atmosphere conc. (ppm)	2.0 ± 0.03	2.12 ± 0.19	2.02 ± 0.15	2.14 ± 0.15

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259

260 **Figure 3.** Methane dynamics. (a-d) Dissolved methane concentration, (e-d) first-order rate constant, (i-l) methane oxidation
 261 rates. Note that for dissolved methane concentrations in colder seasons, autumn and winter, the y-axis differs from warmer
 262 seasons, spring, and summer. Dashed line indicates high tide.

263 3.2.2 Methane oxidation rates

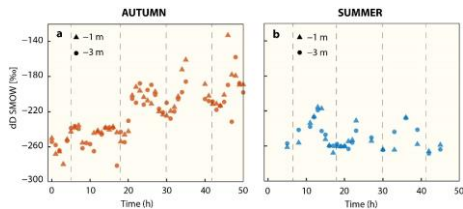
264 Similar to methane concentrations, we observed strong seasonal differences in MOx (Fig. 3I-L, Table 2). Depth-
 265 averaged MOx in spring (1.2 nM d⁻¹) and summer (4.4 nM d⁻¹) was ~ 3 and ~ 9 – fold higher than in winter (0.5
 266 nM d⁻¹) and autumn (0.5 nM d⁻¹). MOx at 1 m and 3 m water depth statistically differed from each other in winter
 267 ($p \leq 0.01$, Welch's t-test), but not in spring, summer, and autumn. MOx at the reference station was < 5 % of MOx
 268 in the Wadden Sea, with maxima found in summer (0.2 nM d⁻¹).

269 On a diel scale, MOx showed fluctuations during all seasons. In general, depth-averaged MOx was higher during
 270 LT compared to HT. In autumn average MOx at LT (0.79 nM d⁻¹) was about 2 – fold higher and significant different
 271 from MOx at HT (0.38 nM d⁻¹, $p \leq 0.03$, Welch's t-test).

272 In winter, the difference between MOx at LT (0.47 nM d⁻¹) and HT (0.43 nM d⁻¹) was small. In spring, depth-
 273 averaged MOx at LT (2.02 nM d⁻¹) was about 4 – fold and significantly ($p \leq 6.4 \times 10^{-6}$, Welch's t-test) higher than
 274 during HT (0.58 nM d⁻¹). In summer, MOx was high at both, LT (5.2 nM d⁻¹) and HT (5.4 nM d⁻¹). Similarly, *k*
 275 was substantial higher (16 – 63 %) at LT than HT in all seasons (Fig. 3E-H, Table 2). In fact, the difference in
 276 depth-averaged *k* between LT and HT was significant in autumn ($p \leq 0.003$, Welch's t-test) and spring ($p \leq 6 \times$
 277 10^{-5} , Welch's t-test, Table S1 in the Supplement).

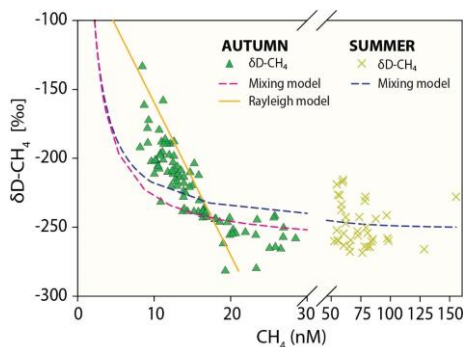
278 3.2.3 Stable hydrogen isotope signatures

279 The stable hydrogen isotope composition of dissolved methane was only measured in autumn and summer (Fig.
 280 4, Table 2). In autumn, average $\delta D-CH_4$ over the entire time-series was -219 ‰ at 1 m water depth and -224 ‰
 281 at 3 m water depth, but there was a generally strong trend towards higher $\delta D-CH_4$ values over the two-day period
 282 from about -260 ‰ to about -180 ‰. In summer the mean $\delta D-CH_4$ values were homogenous throughout the water
 283 column (-250 ‰) and generally lower than in autumn. Except for the first full tidal cycle in autumn, the results
 284 showed a tidal imprint on $\delta D-CH_4$ values with higher $\delta D-CH_4$ values at LT and lower values at HT independent
 285 of depth and season (Fig. 4).



286
287 **Figure 4.** Progression of $\delta\text{D-CH}_4$ signatures in (a) autumn and (b) summer at 1 m and 3 m water depth. Vertical dashed line
288 indicates high tide.

289 In addition to tidal patterns, the $\delta\text{D-CH}_4$ values in autumn were substantially higher at lower methane
290 concentrations (< 21 nM, Fig. 5). Linear mixing alone of (i) well-mixed surface waters in equilibrium with
291 atmospheric methane and (ii) the maximum methane concentration in the water column, both concentrations with
292 their associated isotopic signatures, would result in concentration/isotope data as depicted by the mixing lines in
293 Fig. 5. Results in autumn clearly deviated from this mixing line at low methane concentrations. On the other hand,
294 the open system Rayleigh fractionation model that we ran for low methane concentration in autumn yielded an ϵ
295 value of -97 ‰ and matched the steep rise in $\delta\text{D-CH}_4$ with decreasing methane concentration much better ($R^2 =$
296 0.79). This directly indicates that MOx is the dominant mechanism driving $\delta\text{D-CH}_4$ to higher values at low
297 concentrations.



298
299 **Figure 5.** Methane concentration versus $\delta\text{D-CH}_4$ - mixing and oxidative removal in autumn and summer. Dashed lines show
300 methane concentration/isotope dynamics determined with a two-endmember mixing model considering (i) well-mixed Wadden
301 Sea surface waters and (ii) methane charged waters as endmembers. Methane concentration and stable hydrogen isotope
302 composition following oxidative removal according to a Rayleigh model for low methane concentrations are depicted as a solid
303 line. Samples with methane concentrations < 21 nM ($\delta\text{D-CH}_4 = \sim -217$ ‰) in autumn and < 61 nM ($\delta\text{D-CH}_4 = \sim -244$ ‰) in
304 summer were defined as the methane source signal and thus starting point of the Rayleigh fractionation model. The apparent
305 isotope enrichment (ϵ see also Fig. S2 in the Supplement) was -97 ‰ in autumn with an R^2 of 0.79 . Neither the mixing nor the
306 Rayleigh model are well constrained for $\delta\text{D-CH}_4$ in summer; the mixing line is thus only shown for comparison and ϵ could
307 not be calculated.

308 3.2.4 Diffusive efflux to the atmosphere

309 The water column in the Wadden Sea was consistently methane supersaturated (> 8 nM) with respect to
310 atmospheric equilibrium (~ 2.6 nM) during all sampling campaigns (Fig. 3, Table 2), which indicates a continuous
311 release of methane from the water to the atmosphere throughout the measurement series. Atmospheric
312 concentrations were similar ranging from 1.8 to 2.6 ppm, with relatively constant concentrations in autumn and
313 more erratic concentrations in winter, spring, and summer (Fig. 6e-h, Table 2). Noteworthy is the sharp increase
314 of atmospheric methane from 2 to 2.6 ppm between 29 and 38 hours in summer before decreasing again to 2 ppm.

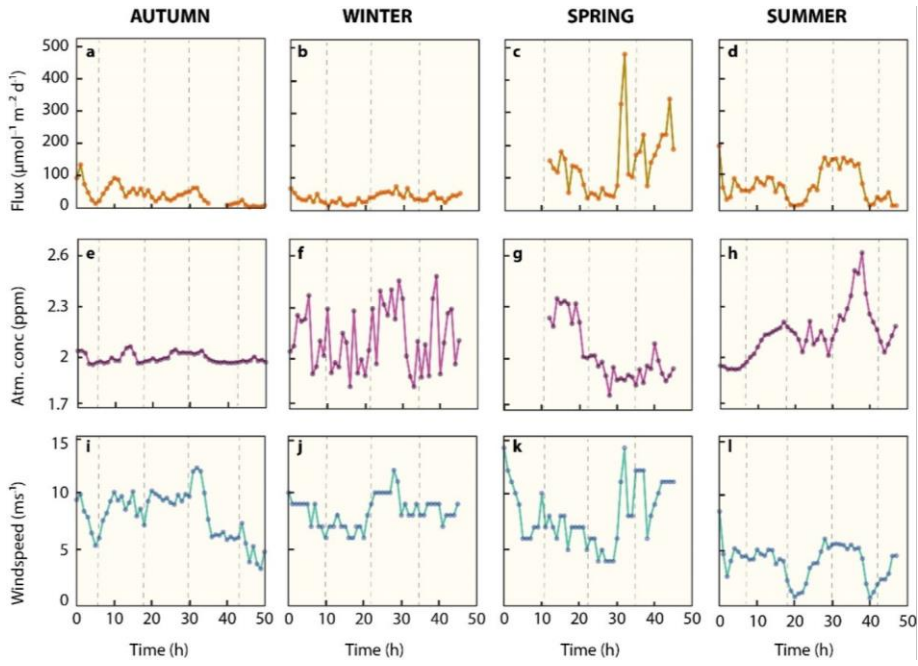
315 Windspeeds in autumn, winter, and spring were relatively high (typically > 5 ms^{-1}) when compared to calmer
316 conditions in summertime (typically < 5 ms^{-1} , Fig. 6i-l, Table 2). As a result of the strong but variable wind forcing,
317 diffusive methane fluxes fluctuated in magnitude within, and between season (Fig. 6a-d, Table 2). Average
318 diffusive fluxes in autumn and winter were with < 40 $\mu\text{mol m}^{-2} \text{d}^{-1}$ about 4-fold lower than in spring and 2-fold
319 lower than in summer. Maximum efflux (479 $\mu\text{mol m}^{-2} \text{d}^{-1}$) in spring occurred after the wind velocity increased

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322 rapidly from 6 m s^{-1} to 14 m s^{-1} within two hours and methane concentrations slightly increased from 38 nM to 45
323 nM.

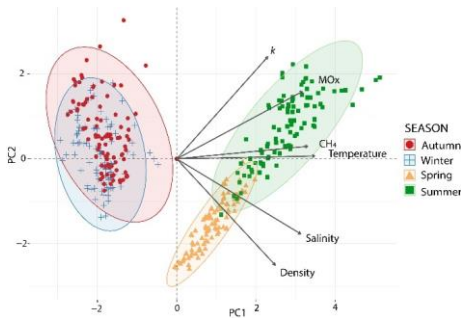


324
325 **Figure 6.** Diffusive methane flux. (a-d) Sea surface atmosphere methane fluxes, (e-h) Seasonal atmospheric methane
326 concentrations, (i-l) local wind speed. Vertical dashed lines indicate high tide.

327

328 **3.3 Statistical Analysis**

329 To study the relationship between environmental variables and methanotrophic activity, we conducted a Principal
330 Component Analysis (PCA). The outcome explained 92 % of the data variability on the first two components (Fig.
331 7, Table S2 in the Supplements). The main gradient (PC1: 69 %) showed a contrast between autumn/winter and
332 summer and spring. Temperature, salinity, methane concentrations and MOx peaked in summer and spring, while
333 lower values were measured in winter/autumn. The relatively small ellipse in spring indicates that samples show
334 more similarity than in other seasons. The second gradient distinguished the spring samples from the summer
335 samples, with higher k values observed in summer and greater density in spring (PC2: 23 %).



336 **Figure 7.** Principal Component Analysis (PCA) of environmental conditions across seasons in the Dutch Wadden Sea. Biplot
337 of a PCA of the explanatory variables as vectors (in black) and observations (marks) of each season on the first (x-axis, PC1)
338 and second principal component (y-axis, PC2). Coloured concentration ellipses (size determined by a 0.95-probability level)
339 show the observations grouped by season. The magnitude of the vectors (line length) shows the strength of their contribution
340 to the PCs. Vectors pointing in similar directions indicate positively correlated variables and vectors at angles $> 90^\circ$ indicate
341 no correlation.
342

343 **4 Discussion**

344 The Wadden Sea is a highly productive ecosystem (Philippart et al., 2009) where the decay of organic matter
345 supports high rates of methanogenesis in sediments (Røy et al., 2008; Wu et al., 2015), which in return leads to
346 high methane concentrations in the Wadden Sea's water column (Grunwald et al., 2007; Grunwald et al., 2009).
347 Little knowledge, however, exists on the variability of methane dynamics on short time scales of hours to days or
348 between seasons and the underlying controls on this variability. Here, we measured water column methane
349 concentrations, methane oxidation and the oceanographic regime as well as atmospheric methane mixing ratios
350 and wind velocity in the Dutch sector of the Wadden Sea for two days during four consecutive seasons in 2019.

351 **4.1 Water column properties**

352 In general, we found a clear distinction between colder (autumn and winter) and warmer (spring and summer)
353 seasons (Figs. 2, 7). North Sea waters with incoming tide, flow through tidal inlets that in turn branch into
354 successively smaller tidal creeks in which water-flow direction alternates with the tidal phase. This then led to
355 increasing water temperatures in autumn but decreasing water temperatures in winter. The high temperature of
356 Wadden Sea waters during incoming tide in autumn can be explained by the fact that the shallow Wadden Sea
357 cools rapidly once the summer is over, while the North Sea's large water volume takes longer to cool down.

358 Salinity levels were on average lower in colder seasons compared to warmer seasons, likely because land runoff
359 and ground water discharge are typically higher in autumn and winter because of the overall higher precipitation
360 levels during the cold seasons (Van Aken, 2008). A higher level of freshwater inflow from land was also evident
361 from the rapidly dropping salinity levels during falling and LT in autumn and winter (Fig. 2). This freshening
362 effect is amplified at times when the Dutch Ministry of Infrastructure and Water management (Rijkswaterstaat)
363 opens water gates to discharge excess water from lake IJssel (Fig. 1), which occurs more often in colder seasons
364 due to increased input of precipitation, groundwater discharge as well as surface and riverine discharge to the lake.
365 During the warmer and dryer seasons, water gates are mostly kept closed to ensure that the lake's water level stays
366 high. However, freshwater inflow into the Wadden Sea was evident during all seasons because incoming North
367 Sea water generally increased salinity levels at HT independent of sampling time. [The North Sea water mass](#)
368 [entering the Wadden Sea during incoming tide hence becomes overprinted in the Wadden Sea area as a result of](#)
369 [mixing with waters from terrestrial sources.](#)

370 **4.2 Differences in methane concentrations and isotopic signatures on time scales of seasons**

371 Sediment and water column methane concentrations were highly elevated in summer (Figs. 3, S1 in the
372 Supplement and Table 2). In fact, average sediment methane concentrations increased 17 – fold in summer
373 compared to spring two months earlier. This increase is probably related to the remineralization of the spring
374 phytoplankton bloom that takes place in the months of April and May (Philippart et al., 2009) leading to elevated
375 rates of methanogenesis in anaerobic sediments (Beck and Brumsack, 2012). A time lag of one to two months
376 between the peak of the spring bloom and methane release from sediments was also observed in the Baltic Sea
377 (Bange et al., 2010). In the Wadden Sea, where sediments are generally silty and organic-rich, it is likely that
378 temperature plays a crucial role in controlling methanogenesis, in addition to the elevated inputs of organic matter.
379 As water temperatures increase towards summer, microbial methanogenesis in the sediments is further enhanced
380 (Yvon-Durocher et al., 2014; Borges et al., 2018). We indeed observed lower methane concentrations in autumn
381 and winter compared to spring and summer, which is most likely related to both reduced organic matter input and
382 colder temperatures. It has to be noted that the sediment methane concentrations presented here are comparably
383 low as sediment methane concentrations close to saturation levels were previously found at other locations in
384 Wadden Sea sediments (Røy et al., 2008; Wu et al., 2015). We did not measure sulphate concentrations, but the
385 methane profiles indicate that we only reached the upper part of the methane-sulphate transition zone below of
386 which methanogenesis proceeds. Also, sediment methane concentrations can be variable on spatial scales of
387 metres. Depending on the hydrographic regime, the methane-sulphate transition zone can be metres below the tidal
388 flat sediments (Wu et al., 2015), but pore water flow can also transport reduced compounds such as sulphide and
389 methane to the sediment surface (Røy et al., 2008).

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390 Methane release from sediments and the relatively low wind speed (and thus relatively low forcing to drive
391 diffusive efflux) in summer led to charging of the water column with methane. MOx discriminates against
392 isotopically heavy methane and thus causes an isotopic enrichment of residual methane. The isotopic
393 discrimination effect manifests more pronouncedly at low methane concentrations. Indeed, we found more
394 pronounced MOx induced isotopic discrimination effects in autumn at low methane concentrations (< 21 nM). At
395 higher methane concentrations (> 21 nM) values were more depleted and were comparable to summer $\delta\text{D-CH}_4$
396 values. We relate the $\delta\text{D-CH}_4$ values (~ -217 ‰ in autumn and ~ -244 ‰ in summer) at higher methane
397 concentrations (> 21 nM in autumn and > 61 nM in summer) to the $\delta\text{D-CH}_4$ source signal (Fig. 4A, 5). At these
398 concentrations, the isotope effect imposed by MOx is masked by the high background methane and/or is
399 overprinted by methane entering the water column from sediments.

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400 **4.3 Differences in MOx on seasonal time scales**

401 The activity of MOB in the water column is determined by the availability of methane, oxygen, nutrients, and the
402 size of the standing stock of the MOB community (Reeburgh, 2007; Crespo-Medina et al., 2014; Steinle et al.,
403 2015). The Wadden Sea water column is a nutrient rich and typically oxygenated environment, we hence argue
404 that nutrient and O_2 availability are not a limiting factor for MOB activity. However, MOB in the Wadden Sea
405 need to cope with high fluctuations in temperature, salinity, and methane availability (see above).

406 We did not measure the size of the MOB community; nevertheless, it seems likely that the highly variable water
407 column properties with admixture of different water masses and resuspension of particles effects the standing stock
408 of the MOB community and/or its activity. Notably, North Sea waters with potentially low MOB standing stock
409 (indicated by the low k value at the reference station) enter the Wadden Sea during incoming tides. As these waters
410 traverse through the Wadden Sea, they acquire methane and likely carry microbes irrigated from sediments
411 and/or originating from mixing with terrestrial waters; incoming North Sea waters hence undergo oceanographic (see
412 above) and biogeochemical overprinting. On short time scales, microbes carried with the tidal current through the
413 Wadden Sea will consequently be exposed to variable conditions regarding salinity and temperature levels, and
414 methane concentration.

415 Previous studies showed that elevated salinity often led to an immediate decrease in MOx in terrestrial/lacustrine
416 systems (Ho et al., 2018; Zhang et al., 2023). Likewise, marine methanotrophs seem to function best at salinity
417 levels of > 20 psu (Osudar et al., 2017), while a sudden decrease in salinity can strongly inhibit MOx (Hirayama
418 et al., 2013; Tavormina et al., 2015). This begs the question if waters, with rapidly changing salinity levels such
419 as the Wadden Sea, are environments that are rather not conducive for MOx, in particular in colder months where
420 salinity levels may drop to ~ 20 psu because of elevated freshwater influx (see above). While MOx was indeed
421 lower in autumn and winter, the relative decrease in MOx was moderate in comparison to the previous literature
422 findings (Osudar et al., 2017; Zhang et al., 2023). Also, autumn and winter are colder and defined by lower methane

427 levels, which likely reduces MOx further. Across seasons, the PCA (Fig. 7), and Pearson correlation coefficients
428 of pairs of variables (Fig. S3 in the Supplement) indicated that MOx (or k) and salinity (or density) are not or only
429 weakly correlated. The Wadden Sea thus seems to host a euryhaline MOB community that contrast with MOB
430 communities from terrestrial/lacustrine (Zhang et al., 2023) and oceanic origin (Osudar et al., 2015), which seem
431 less able to cope with varying salinity levels.

432 Sediments and the water column in the Wadden Sea are increasingly fuelled with methane when ambient
433 temperatures rise. The higher availability of methane could then enhance methanotrophic activity (Reeburgh,
434 2007). Indeed, we found a seasonal imprint with highest MOx levels in summer that were 3-fold higher than those
435 observed in spring, 9-fold higher than in autumn and 10-fold higher than in winter (Table 2). A correlation between
436 methane, temperature and MOx was also apparent from the PCA (Figs. 7, S3 in the Supplement). We note that not
437 only MOx, but also the first order rate constant k was stimulated by higher methane concentrations (MOx is a
438 function of k and $[\text{CH}_4]$, see eq. 3) and temperature. A positive effect of methane on MOx and k is often associated
439 with changes in methane concentrations over several orders of magnitude (Crespo-Medina et al., 2014; James et
440 al., 2016). Here we found that k doubled in summer compared to spring, while methane concentrations were only
441 30 nM higher, i.e., 1.7-fold. This suggests that a combination of methane availability and temperature determined
442 k in our study. I.e., the MOB's may have been stimulated on the enzymatic level. However, the fact that k remained
443 stable in colder seasons with low water temperatures, suggest that additional factors, likely MOB community size
444 (Steinle et al., 2015), might play a more important role in maintaining k . For example, MOB's from sediments can
445 be resuspended into the water column due to tidal currents, or transported from sediments to the water-column
446 with bubbles as has been found at other cold seeps (Steinle et al., 2016; Jordan et al., 2020; Jordan et al., 2021).
447 Resuspension could thus be a key driver of the Wadden Sea water column MOB communities, with major
448 consequences for maintaining a microbial filter under less favourable conditions.

449 **4.4 Methane dynamics on times scales of hours to days.**

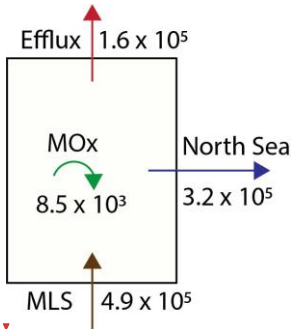
450 Strong hydraulic dynamics are an important characteristic of the Dutch Wadden Sea, with tidal currents
451 interchanging a large water volume with the North Sea twice per day (Gräwe et al., 2016). With the change in tidal
452 phase, the hydrostatic pressure changes rapidly with water depth, which triggers porewater flow (tidal pumping,
453 (Røy et al., 2008; Santos et al., 2015)) but may also trigger expansion and ebullition of gas bubbles (Schmale et
454 al., 2015; Jordan et al., 2020). Similar effects are caused by tidal currents flowing over bathymetric features, which
455 triggers pore water flow, too, and additionally resuspends sediments and MOB's into the water column (Bussmann,
456 2005; Abril et al., 2007; Røy et al., 2008). On the other hand, incoming water from the open North Sea contains
457 relatively low amounts of methane (<6 nM as measured at our reference station), hence, this will dilute the Dutch
458 Wadden Sea's methane content, and outflowing water will export methane from the Dutch Wadden Sea main
459 water body.

460 Temporal patterns of methane concentration and MOx indeed correlated well with tidal oscillation (Figs. 3, S1 in
461 the Supplements, Table 2). Independent of the seasons, methane concentrations and MOx were elevated at LT.
462 The tidal effect seemed most pronounced in spring where at LT, methane concentrations (1.3 – fold), k (2.5 – fold)
463 and MOx (4 – fold) were higher than at HT, independent of depth. We found it surprising that, just like methane
464 concentrations, k also was substantially higher during LT compared to HT independent of seasons and despite an
465 overall lower salinity at low tide (Figs. 3, S1 in the Supplement, Table 2). To the best of our knowledge, this has
466 not been described before. It seems unlikely that the MOB community substantially grew or that the velocity of
467 the MOB's metabolism increased/decreased in a time frame of a few hours. We rather argue that the observed
468 oscillation is caused by a likewise oscillation of shear force and hydrostatic pressure, leading to resuspension of
469 MOB's from sediments as well as elevated release of methane from the sea floor.

470 Grunwald et al. (2007 and 2009) conducted time-series measurements in the German sector of the Wadden Sea
471 near the island of Spiekeroog. There, absolute methane concentrations were ~ 3 – fold higher in spring and summer
472 and ~ 15 – fold higher in winter when compared to our study. This might be related to local factors, for example
473 the vicinity of the estuaries of the rivers Eems and more importantly Weser close by Spiekeroog, which increase
474 the background methane concentrations in this sector of the Wadden Sea. Like in our study, Grunwald et al. (2007,
475 2009) also reported on a strong influence of tides, with highest methane concentrations at low tide, probably related
476 to tidal pumping, while inflowing waters showed concentrations typical for the open North Sea in the German
477 Bight. The temporal aspects and processes determining methane dynamics discussed in our work are thus not a
478 local feature but applicable to the entire Wadden Sea and likely to other mud flat areas influenced by tides, too.

479 **4.5 Methane emissions from the Dutch Wadden Sea**

480 Surface waters were supersaturated in methane with respect to the atmospheric equilibrium during all seasons; the
481 Wadden Sea is consequently a constant source of methane to the atmosphere. Just as for dissolved methane
482 concentrations and MOx, methane efflux to the atmosphere was higher during warmer seasons compared to colder
483 seasons. This was primarily driven by methane concentrations rather than wind velocity: wind speeds were similar
484 in autumn, winter, and spring, but 2-fold higher methane concentrations in spring translate to a 4-fold higher sea-
485 air flux when compared to autumn and winter. In summer, meteorological conditions were dominated by a heat
486 wave with extremely low wind speeds. This resulted in a comparably low methane efflux to the atmosphere
487 (though still higher than during the colder seasons) leading to an accumulation of methane in the water column.
488 Previously described diffusive methane fluxes at coastal systems vary over several orders of magnitude and appear
489 site specific. For instance, at the Baltic sea coast, fluxes of up to $15 \mu\text{mol m}^{-2} \text{d}^{-1}$ have been reported (Bange et al.,
490 2010; Steinle et al., 2017), while in arctic shelf seas, diffusive fluxes of up to $240 \mu\text{mol m}^{-2} \text{d}^{-1}$ were found
491 (Thornton et al., 2016). In the Southern bight of the North Sea, reported fluxes at the coast were up to $345 \mu\text{mol}$
492 $\text{m}^{-2} \text{d}^{-1}$ (Borges et al., 2018). In comparison, estuarine research along the European Atlantic coast found a median
493 flux of $130 \mu\text{mol m}^{-2} \text{d}^{-1}$ (Middelburg et al., 2002), which is similar to the fluxes that we found in the Dutch
494 Wadden Sea (~ 39 to $145 \mu\text{mol m}^{-2} \text{d}^{-1}$). Globally, tidal flats were estimated to emit CH_4 at a median rate of ~ 3.6
495 $\text{mg m}^{-2} \text{d}^{-1}$ ($226 \mu\text{mol m}^{-2} \text{d}^{-1}$; (Rosentreter et al., 2021)), which is similar (1.5 to 6-fold higher) than our flux
496 estimates from the Wadden Sea.



497
498 **Figure 8.** Methane budget for the Dutch Wadden Sea is calculated based on values for the Wadden Sea's geometry, tidal
499 displacement volume, and biogeochemical parameters as discussed in the text. MLS stands for methane liberation from
500 sediments. All values are presented as $\text{mol CH}_4 \text{d}^{-1}$.

501 Towards a roughly estimated methane budget for Dutch Wadden Sea, we combined our diffusive flux, MOx and
502 methane concentration data (Fig 8.) as well as estimates of the Wadden Sea water volume and tidal prism. Our
503 flux estimates (Table 2) translate to an annual average sea surface-atmosphere flux of $74 \mu\text{mol m}^{-2} \text{d}^{-1}$.
504 Extrapolating this to the area of Dutch sector of the Wadden Sea ($\sim 2200 \text{ km}^2$; (Materić et al., 2022) implies that
505 $1.6 \times 10^5 \text{ mol CH}_4 \text{d}^{-1}$ escapes from the Dutch Wadden Sea to the atmosphere (Table 2). The average water volume
506 of the Dutch Wadden Sea is about 5.15 km^3 (Materić et al., 2022); hence the annual average of 1.7 nM d^{-1} of MOx
507 translates to $0.09 \times 10^5 \text{ mol CH}_4 \text{d}^{-1}$ that is oxidized in the water column by MOBs. In addition to atmospheric
508 efflux and microbial consumption, methane rich waters are also flushed into the North Sea. To estimate this, we
509 simplified that the total tidal prism of 4.5 km^3 (Gräwe et al., 2016) is an approximation of the net amount of water
510 that leaves the Wadden Sea during LT. With respect to our measured mean methane concentration (36.8 nM),
511 about $1.6 \times 10^5 \text{ mol}$ of methane are thus flushed towards the North Sea twice daily, i.e., $3.2 \times 10^5 \text{ mol}$ per day. A
512 large uncertainty in this calculation is caused by the delay of ~ 3 hours in tidal phases between the Western and
513 Eastern part of the Dutch Wadden Sea. In other words, methane-rich waters are flowing out of the tidal inlet in the
514 West can be entrained in the current that starts flowing back into the Wadden Sea at eastern tidal inlets. Therefore,
515 the net loss of methane to the Wadden Sea is probably lower than described above. Still, data from our reference
516 station show only slightly oversaturated methane concentrations ($< 6 \text{ nM}$) suggesting that the amount of methane
517 flowing back into the Wadden Sea is rather low. A similar observation was found during a tidal inlet study in the
518 German Wadden Sea (Grunwald et al., 2009). Though overall methane concentrations were higher, methane
519 concentrations in North Sea waters flowing into the Wadden Sea were 60 % lower compared to waters flowing
520 out of the Wadden Sea at low tide. Excluding allochthonous methane sources (for example methane influx with

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525 freshwater from Lake IJssel), the Dutch Wadden Sea's methane budget must be supported by a total rate of
526 methanogenesis that at least equals the sum of methane efflux to the atmosphere, water column methanotrophy
527 and methane outflow to the North Sea; together these amount to $4.9 \times 10^5 \text{ mol d}^{-1}$. This is comparable to
528 methanogenesis rates in the Eckernförde Bay in the Baltic Sea in the Baltic Sea (Maltby et al., 2018). Note that
529 this accounts for the amount of methane liberated from sediments, while it neglects methane oxidation in sediments
530 (dominantly anaerobic oxidation of methane), which can retain a substantial fraction of methane in sediments
531 (Reeburgh, 2007). Hence, the total rate of methanogenesis in the Wadden Sea is consequently much higher.

532 Taken all methane export terms/sinks considered together (MOx, efflux and tidal displacement amounting to 4.9
533 $\times 10^5 \text{ mol d}^{-1}$), MOx reduces roughly 2 % of the Wadden Sea's methane budget, while about 1/3rd of methane
534 escapes to the atmosphere and the remaining ~ 2/3rd is flushed into the North Sea (where it may be further oxidised
535 and/or released to the atmosphere). The effect of MOx on the Wadden Sea's methane budget is low when compared
536 to the global ocean, where an estimated >90 % of water column methane is consumed by MOx (Reeburgh, 2007).
537 As the Wadden Sea is very shallow, liberation of methane from sediments to the atmosphere is fast; in other words,
538 MOBs have a very limited time to consume methane released from the sediments before it is liberated to the
539 atmosphere or flushed with tides to the North Sea. ~~In a meta study, Rosentreter et al. (2021) estimated a global
540 median methane efflux from tidal flats (covering ~128000 km² globally; (Murray et al., 2019)), to the atmosphere
541 of 0.17 Tg y⁻¹. We found a total annual diffusive sea-air flux from the Dutch sector of the Wadden Sea (2200 km²)
542 of ~ 0.001 Tg y⁻¹, which alone already accounts for 0.6 % of the global methane emissions from tidal flats to the
543 atmosphere.~~

544 5 Summary and conclusion.

545 Our work revealed substantial variations in methane dynamics when comparing colder and warmer seasons; in
546 warmer seasons, methane concentrations, efflux and MOx were higher compared to colder seasons. Still during
547 colder seasons waters were continuously supersaturated with methane and higher wind speeds in these seasons led
548 to substantial amounts of methane released to the atmosphere. We show that tidal dynamics are a key control for
549 methanotrophic activity and methane distribution. Although changing water column properties and methane
550 concentrations do not provide continuity, the capacity of the microbial methane filter is seemingly stable, with an
551 active MOB community even under unfavourable conditions. Nevertheless, MOx only consumes a minor fraction
552 of the methane inventory of the highly dynamic Wadden Sea, while most is or liberated to the atmosphere and
553 flushed out with tidal currents into the neighbouring North Sea. It appears likely that the contribution of the
554 Wadden Sea to the global atmospheric methane budget will alter in the future due to global warming, and changes
555 in nutrient availability and more frequently occurring storm events. Our results finally highlight the importance of
556 repeated high frequency sampling strategies in dynamic coastal waters to resolve temporal patterns on diel and
557 seasonal scales.

558 *Data availability.* All data will be archived and made publicly available in the data base DAS (Data Archive
559 System, www.nioz.nl/en/research/dataportal/das).

560 The supplement related to this article is available online.

561
562
563 *Author contributions.* The study was designed by Tim de Groot, Thomas Röckmann, and Helge Niemann. On-
564 board sampling was performed by Tim de Groot, Anne Mol, Katherine Mesdag, Julia Engelmann, Pierre Ramond,
565 and Helge Niemann. Further geochemical analysis was conducted by Tim de Groot, Anne Mol, Katherine Mesdag,
566 and Rachel Ndhlovu. Microbial rates were measured by Tim de Groot and Anne Mol. Statistical analysis was
567 carried out by Tim de Groot and Pierre Ramond. Helge Niemann supervised the research project. The manuscript
568 was prepared by Tim de Groot with input from all authors.

569 *Competing interests.* The authors disclose that at least one of the (co-)authors holds a position on the editorial
570 board of Biogeosciences.

571
572 *Acknowledgements.* Our gratitude goes to the captain and crew of R/V Navicula, as well as the staff of the
573 geochemical, radioisotope, and atmospheric laboratories at NIOZ and IMAU, for their exceptional support. We
574 would also like to extend our appreciation to Eric Wagemaak for regularly calibrating the CTD.

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Deleted: et al. (2019) estimated that coastal systems globally contribute between 0.8 and 3.8 Tg y⁻¹ of methane to the atmospheric budget. We found a total annual sea-air flux from the Dutch sector of the Wadden Sea (2200 km²) of ~ 0.001 Tg y⁻¹. The Dutch sector of the Wadden Sea alone may thus already account for 0.03 % to 0.1 % of the global methane emission from the global coastal ocean.

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