Rethinking the deployment of static chambers for CO2 flux measurement in dry desert soils

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Abstract. The mechanisms underlying the soil CO_2 flux (Fs) in dry desert soils are not fully understood. To better understand these processes, we must accurately estimate these small fluxes. The most commonly used method, static chambers, inherently alter the conditions that affect the flux and may introduce errors of the same order of magnitude as the flux itself. Regional and global assessments of annual soil respiration rates are based on extrapolating point measurements conducted with flux chambers. Yet, studies conducted in desert ecosystems rarely discuss potential errors associated with using static chambers in dry and bare soils. We hypothesized that a main source of error is the collar protrusion above the soil surface. During the 2021 dry season, we deployed four automated chambers on collars with different configurations in the Negev Desert, Israel. Fs exhibited a repetitive diel cycle of nocturnal uptake and daytime efflux. CO_2 uptake measured over the conventionally protruding collars was significantly lower than over the collars flushed with the soil surface. Using thermal imaging, we proved that the protruding collar walls distorted the ambient heating and cooling regime of the topsoil layer, increasing the mean surface temperatures. Higher soil temperatures during the night suppressed the flux driving forces, i.e., soil-atmosphere CO_2 and temperature gradients, ultimately leading to an underestimation of up to 50% of the actual Fs. Accordingly, the total daily CO_2 uptake by the soil in the conventionally deployed collars was underestimated by 35%. This suggests that desert soils are a larger carbon sink than previously reported and that drylands, which cover approximately 40% of Earth's terrestrial surface, may play a significant role in the global carbon balance.

1 Introduction

Soil respiration, i.e., the carbon dioxide (CO_2) efflux from the soil to the atmosphere, is among the largest components of the carbon balance in terrestrial ecosystems, contributing approximately 60 PgC to the atmosphere every year (Houghton, 2007). In arid and semi-arid environments, soil respiration is mostly considered to be restricted to short pulses of increased moisture availability from rainfall events, during which microbial metabolic activity increase rapidly, followed by long periods of desiccation and low to negligible soil respiration rates (Cable et al., 2008; Austin et al., 2004). In the last two decades, studies carried out in several deserts have challenged this paradigm, reporting a diel course of CO_2 exchange, consisting of nocturnal CO2 uptake and daytime efflux (Ball et al., 2009; Sagi et al., 2021; Lopez-Canfin et al., 2022). Researchers usually attribute this diel cycle to changes in soil temperatures and soil air pressure that leads to cycles of expansion/contraction of soil air, following the ideal gas law (Yang et al., 2020). These cycles change the surface CO_2 concentration and may generate a soil-atmosphere pressure gradient (Ganot et al., 2014), both driving forces for soil CO_2 flux (F_s). Another explanation is based on Henry's Law. It states that diurnal fluctuations of soil temperatures change the solubility of soil CO_2 in water films, which changes the concentration of gaseous CO_2 in soil pores, leading to the exchange of CO_2 between the soil and the atmosphere by diffusion (Fa et al., 2016). In saline/alkaline soils, this process is thought to cause a diel cycle of calcium carbonate ($CaCO_3$) precipitation/dissolution, which enhances F_s (Hamerlynck et al., 2013; Fa et al., 2016). Yet, the factors controlling F_s in desert soils and the partitioning between them are still under debate.

Furthermore, the ability to accurately estimate the soil CO_2 flux in desert soils at the very dry-end is controversial due to the potential for measurement-induced modifications to soil and atmospheric conditions that can introduce errors of the same order of magnitude as the flux being measured. This problem is exacerbated when using static chambers to measure flux, as the chambers inherently alter the conditions that affect the flux (Pumpanen et al., 2010 ; Parkin et al., 2012). During efflux, CO_2 concentration in the chamber builds up, decreasing the diffusion gradient between CO_2 in the soil pores and the chamber headspace, thereby altering CO_2 concentration within the top soil layer and reducing the flux (Pumpanen et al., 2004). Artificial changes in air pressure within the chamber headspace compared to the ambient atmosphere are another source of error (Bain et al., 2005; Lund et al., 1999).

There are additional sources of errors associated with the chamber-soil contact method (Ngao et al., 2006; Baram et al., 2022). Flux chambers are typically deployed on a collar (i.e., PVC pipe) that is inserted into the soil, with the upper 3-5 cm of the collar protruding above the soil surface to allow for chamber deployment. This practice modifies the soil surface temperature by shading a portion of the measured surface area. The non-representative soil surface temperature results in modified heat exchange between the soil and the atmosphere, as well as a modified soil temperature profile (Ninari and Berliner, 2002). Soil microbial and physical processes that drive F_s are susceptible to changes in soil temperature (Cable et al., 2011), and thus shading the soil surface can lead to errors in F_s measurements. These errors may intensify in high-latitude cold deserts, in which the low angle of insolation will dictate a larger shaded surface area for longer periods during the day. F_s was shown to be particularly affected by fluctuations in soil temperatures in cold deserts (Parsons et al., 2004; Ball et al., 2009).While these effects are likely minimal in temperate, vegetated areas, they could be significant in bare soil, partly because fluctuations in surface temperatures are not regulated by vegetation cover as in humid environments. Desert soils also have lower specific heat capacity than soils in humid environments due to lower

water content (Hillel, 1998). The lower water content also means that a larger portion of the available energy is invested in soil heating rather than stored as latent heat during evaporation (Brutsaert, 1982). However, studies using static chambers in desert ecosystems rarely discuss potential errors associated with the unique characteristics of desert soils. Moreover, to our knowledge, the effect of collar height above the surface on soil surface temperature and, consequently, on F_s was never studied.

Under dry soil conditions, the depth to which the collar is inserted can also significantly influence the flux measurements. The ideal insertion depth is debatable, as both shallow and deep collar insertion depths can lead to errors, depending on climate and soil conditions. Inserting the collar to a shallower depth than the depth to which feedback from the chamber still affects gas concentrations may result in lateral diffusion, leading to underestimation of the vertical flux (Healy et al., 1996). However, insertion depth of only 2.5 cm and a measurement period of 10 minutes will reduce this underestimation to 1% for a soil with air-filled porosity of 0.3 m^3 m^{-3} (Hutchinson and Livingston, 2001). Hence, for short measurement periods (common today) and soils with low effective diffusivity, errors resulting from lateral diffusion may be insignificant. With current static chamber systems, even small F_s measured in dry desert soils can be accurately quantified with much shorter measurement periods of only 1-2 minutes (Yang et al., 2022), thus overcoming a significant drawback of the shallow collars. Deep collar insertion, on the other hand, can lead to either overestimation or underestimation of the flux by generating vertical mass flow of air along the collar walls or by facilitating root cutting, respectively (Heinemeyer et al., 2011). Still, in most studies, collars are inserted to a depth of ~5-10 cm into the soil and, in some cases, to a depth of 30-60 cm, while more than a third of all authors fail to report the collar insertion depth (Rochette and Eriksen-Hamel, 2008; Cable et al., 2011; Fa et al., 2018; Jian et al., 2020; Sagi et al., 2021; Yang et al., 2022).

In this paper, we aimed to investigate the effect of collar height above the soil surface and collar depth of insertion on F_s in a dry bare desert soil. Given the small fluxes in these conditions, and the fact that regional and global assessments of annual soil respiration are based on extrapolating point measurements conducted with flux chambers (Jian et al., 2020), minimizing measurement errors associated with the collar deployment technic is critical. Arid and semi-arid regions, which comprise approximately 40% of Earth's terrestrial surface, constitute the largest uncertainty on mean annual soil respiration estimations (Stell et al., 2021). Improving the accuracy of F_s measurements in desert environments is essential for enhancing our understanding of the terrestrial carbon balance and our ability to predict climate change.

2   Materials and Methods

2.1   Research site

The study was carried out at the Wadi Mashash Experimental farm in the Northern Negev, Israel (31°04'14''N, 34°51'62''E; 360 m.a.s.l; 65 km SE of the Mediterranean Sea). The climate in the research site is arid, with an average annual rainfall of 116 mm (IMS, 2021), occurring between October and April. The daily mean maximum and minimum temperatures for January (winter) are 15.9 C° and 8.0 C°, respectively, while those for August (summer) are 33.3 C° and 20.7 C°. During the summer season, the prevailing wind direction is NW due to the sea breeze carrying water vapor from the Mediterranean Sea inland. The sea breeze reaches its peak at a wind speed of 7 m s^{-1} (at 10 m height) in the afternoon. The research is located on a largely bare plain of sandy-loam loess

99   soil (72.5% sand, 15% silt and 12.5% clay), partly covered by a biological soil crust over a thin physical crust,
100 with dry annual grasses and Shrubs.

2.2   Meteorological measurements

103 Air temperature and relative humidity (100K6A1A, BetaTherm, USA) were monitored along with wind speed
104 and direction as part of an eddy-covariance system (IRGASON, Campbell Scientific Inc.). Air temperature was
105 measured at 5-second intervals and averaged over 15-minute periods. Wind speed and direction were determined
106 from high-frequency measurements of 3D wind speed taken at 20 Hz intervals, then averaged over 30-minute
107 periods and stored in a data logger (CR6, Campbell Scientific Inc.). Net radiation was measured at a height of 2.4
108 m using a 4-component net radiometer (SN-500-SS, Apogee instrument Inc, USA) at 10-second intervals,
109 averaged over 15-minute periods, and stored in a data logger (CR5000, Campbell Scientific Inc.).

2.3   Soil CO_2 flux measurements

111 We measured F_s using a non-dispersive Infrared Gas Analyzer with a range of 0-20,000 ppm and an accuracy of
112 1.5% of reading. The analyzer was connected to four automated non-steady-state chambers (LI 8100A- 104C, LI-
113 COR, Lincoln, USA). The chambers were closed on a pre-inserted collar every 30 minutes for a measurement
114 period of 60 seconds, with a 10-second dead band period to allow homogeneous air mixing within the system.
115 Each measurement started with a 90-second pre-purge and ended with a 45-second post-purge period.

116 We deployed the chambers on three types of collars (i.e., treatments): (1) The conventional type (CONV) - an 11
117 cm long collar, inserted 7.5 cm into the soil, leaving 3.5 cm of collar above the soil surface (Fig. 1); (2) The deep
118 type (DEEP) - an 11 cm long collar completely inserted into the soil, leaving the top of the collar flush with the
119 soil surface; and (3) The shallow type (SHAL) - a 2.5 cm long collar completely inserted into the soil, with the
120 top of the collar flush with the soil surface. Three collars from each type (1-3) were inserted into the soil two
121 months before measurements started. All collars had a inner diameter of 20 cm.
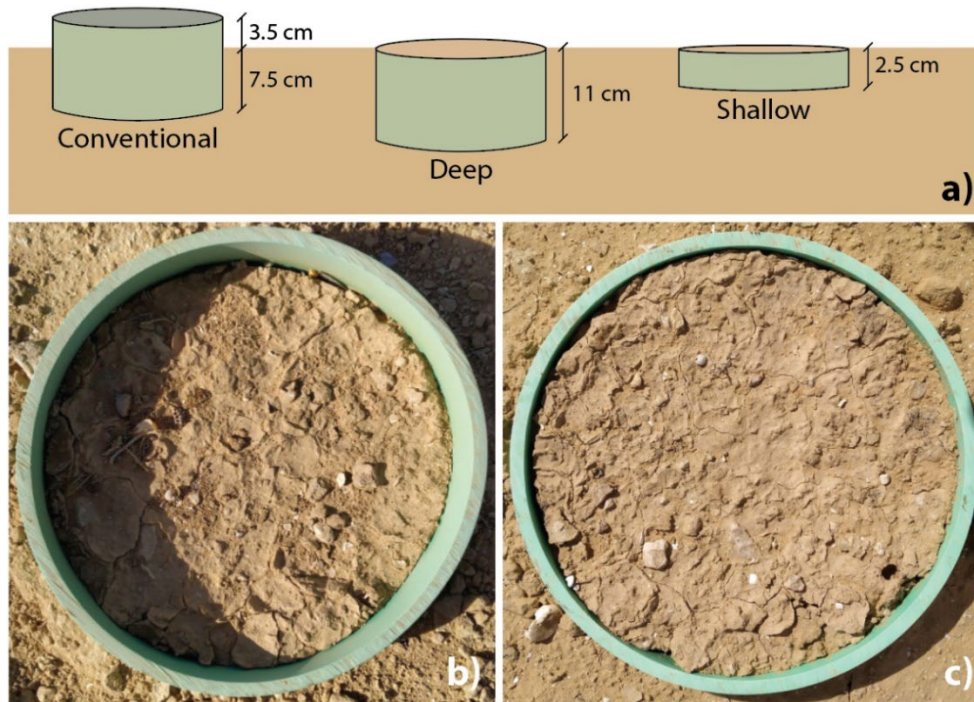
122
123 **Figure 1: a) The three types of collars used in this experiment. b) Photo of a conventional (CONV) collar. C) Photo of**
124 **a collar flashed with the soil surface, representing the DEEP and SHAL treatments.**

125 We collected data between May and June of the 2021 dry season. Three chambers were rotated between the collars

126 on a near-weekly basis (periods 1-6; Table 1), ensuring that each period consisted of at least five full and

127 representative days. The fourth chamber was placed on an additional DEEP collar for the whole experiment

128 duration (the permanent type - PERM). The chambers were rotated in two configurations (Table 1): during periods

129 1, 3 and 5, each chamber was set over a different treatment, e.g., in period 1, chambers were placed over collars

130 CONV1, DEEP1, SHAL1; and during periods 2, 4 and 6, the three chambers were placed on the same treatment

131 (SAME), e.g., in period 2, chambers were placed over collars CONV1, CONV2, CONV3.

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Table 1. Chamber placement during the 6 measurement periods - 12/05-29/06/2021

Period	1	2	3	4	5	6
Dates	12-18/05	18-22/05 27-30/05	30/05-03/06 06-09/06	09-16/06	16-22/06	24-29/06
Analyzed days	12-16/05	19-21/05 28-29/05	31/05-02/06 07-08/06	09-14/06	17-21/06	25-29/06
Treatment and replicate	CONV1 DEEP1 SHAL1	CONV 1-3	CONV2 DEEP2 SHAL2	DEEP 1-3	CONV3 DEEP3 SHAL3	SHAL 1-3

One chamber (PERM) continuously measured soil CO_2 flux on the same collar throughout the experiment.

133 **2.4   Ancillary soil measurements**

134 The temperature profile in the soil was measured by self-made T-type thermocouples buried at depths of 0.5, 1,

135 2, 3, 4, 5, 10, 15, 20, 30 and 50 cm. The thermocouple buried at 0.5 cm provided a proxy for the soil surface

136 temperature. The soil heat flux was derived using the combination method with three repetitions, using a soil heat

5

flux plate (HFT3, Campbell Scientific Inc.) buried at a depth of 5 cm. Heat storage above the plates was derived from two self-made T-type thermocouples buried at depths of 1.25 and 3.75 cm, and soil water content was measured with a time-domain reflectometer (TDR-315, Acclima, Inc., USA) installed at a depth of 3 cm. The volumetric water content of the soil was lower than 3% throughout the experiment. Temperature profile and water content data were collected at 10-second intervals, and 15-minute averages were stored in a data logger (CR1000X, Campbell Scientific Inc.) and multiplexer (AM 16/32B, Campbell Scientific Inc.). Soil heat flux data were also collected at 10-second intervals, and 15-minute averages were stored in a data logger (CR5000, Campbell Scientific Inc.).

2.5   Radiometric surface temperature

A 24-hour field campaign was conducted on August 17-18, 2021. During the campaign, the surface radiometric temperature of the collars was acquired hourly using a thermal infrared camera (A655sc, FLIR, Wilsonville, USA), immediately before taking F_s measurements.

2.6   Data analysis

To calculate F_s, a linear function was fitted to the change in CO_2 mole fraction over time for each measurement, using the software LI-COR SoilFluxPro 5.2.0 (LI-COR, Lincoln, USA). The fitting period, which usually lasted 20 seconds, started after air mixing within the chamber was achieved.

To decipher the differences between collars, and given the limited number of chambers, we derived an "average-day" for each collar type (CONV, DEEP, and SHAL). First, five full representative days from each experiment period (Table 1) were analyzed. Then, for each of the four chambers, an average diel course was calculated from the 5 analyzed days, resulting in 4 average days per period. All average days from all periods (4 treatments × 6 periods= 24 average days) were then divided into 3 groups based on collar type (6 average days per treatment), and a single average day per treatment was calculated as the mean of the 6 average days. Each time point in the three treatment average days consists of 30 values (6 average days × 5 days per average).

The differences between the treatments were tested for significance using linear mixed models (LMMs), following the approach developed by Spyroglou et al. (2021). We built a statistical model Using LMMs that predicted the response variable (i.e., the mean daily cycle of F_s) as a function of treatment and time as fixed factors (fixed for all data points), and each collar as a subject-specific factor (random effect). This allowed us to assess the effect of treatment, but also the effect of time and individual collars on F_s, while incorporating all 24-hour time series into a single model. Still, this model fails to defuse the autocorrelation between data points in each time series. To address this, the LMM residuals were passed through an Autoregressive Integrated Moving Average (ARIMA) model and then incorporated within the LMM as errors. The predicted F_s values produced by the corrected model were compared between treatments for each time interval separately using a two-tale t-test with a 95% confidence interval. To avoid type I errors, the p-value was divided by the number of tests performed on each time point according to the Bonferroni correction. Therefore, the corrected p-value used here is 0.05/6=0.008. The differences between the treatments were also tested by comparing peak daily and daily accumulated efflux and uptake value. This was executed using one-way ANOVA and a post hoc Tukey test with a 95% confidence interval. The modeling process and statistical analysis were performed using "stats", "lme4" and "forecast" packages in RStudio 4.1.1.

175 To analyze the collars surface temperature, the region of interest (ROI) for each thermal image was defined for

176 the collar's inner surface area using FLIR ResearchIR Max 4.40.35. The surface temperature of all pixels within

177 the ROI were then exported to RStudio to calculate statistical parameters used to compare treatments. The soil

178 surface emissivity was set to 0.95 for all images (Li et al., 2013).

179 **3   Results**

180 **3.1   Meteorological and soil conditions**

181 The experiment period was characterized by clear sky days, with similar diel patterns and magnitudes of incoming

182 short-wave and net-radiation (Fig. 2). Solar noon occurred at 11:30 every day of the experiment (UTC+02:00).

183 Sunrise and sunset occurred at 04:30-05:00 and 19:00, respectively. The daily minimum and maximum air and

184 soil surface temperatures were 19.45±2.3 and 34.5±2.7 C° (air) and 17.7±2 to 49.6±2.2 C° (soil surface),

185 respectively. The mean daily range was 13.7±1.0 and 31.8±1.2 C°, for the air and the soil surface respectively,

186 with a slight variation between the experiment weeks. The soil surface temperature regularly dropped below air

187 temperature at night (Fig. 2B). The prevailing wind direction was NW, peaking in the afternoon at a mean speed

188 of 6.2±0.2 m s^{-1} (2 m height).
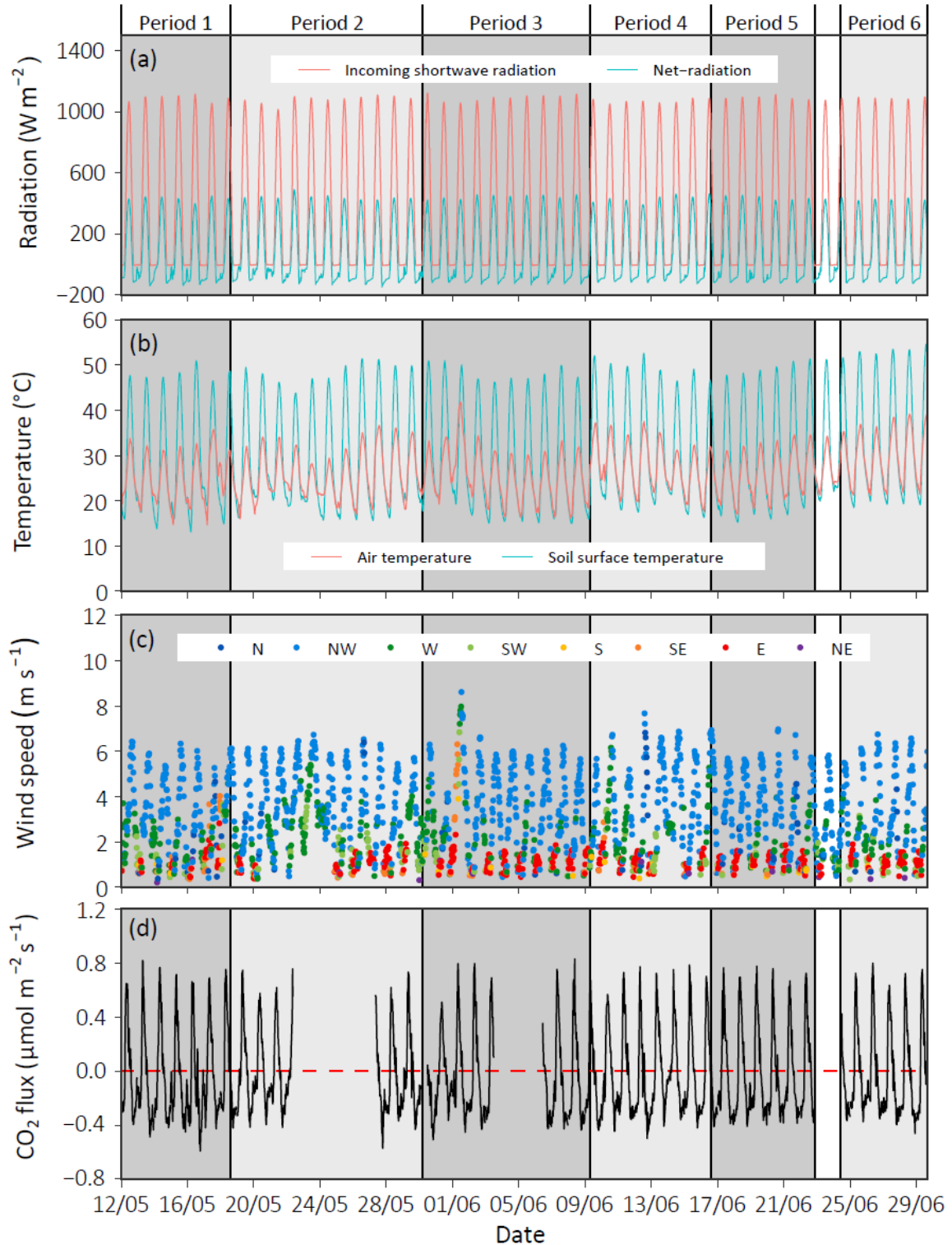
Figure 2: Time series with half hourly data of environmental variables measured at the Wadi Mashash Experimental farm during the 2021 summer season. A) Incoming shortwave radiation and net radiation. B) Air and soil surface temperatures measured at 0.5 cm depth. C) Wind speed is color-coded according to wind direction: north (N), north-west (NW), west (W), south-west (SW), south (S), south-east (SE), east (E), and north-east (NE). D) The soil CO_2 flux measured by the permanent chamber (PHARM).

Soil CO_2 flux measured on the permanent collar followed a consistent diurnal pattern throughout the experiment (Fig. 2d), confirming that the weekly periods can be used to test differences between treatments. Starting from the afternoon (mean time 13:30), negative CO_2 flux (i.e., uptake; from the atmosphere to the soil) occurred, peaking,

on average, at a flux of -0.4±0.04 μmol m^{-2} s^{-1} (at 18:30). Then in the early morning (06:00), the flux reversed, and positive CO_2 flux (i.e., efflux; from the soil to the atmosphere) increased sharply until 08:30, when a daily maximum of 0.71±0.08 μmol m^{-2} s^{-1} was observed. After that, efflux gradually decreased until the afternoon.

3.2   The effect of collar type on soil CO_2 flux

The daily temporal dynamic of F_s shows little variation among the different treatments. However, the rate of increasing CO_2 efflux in the early morning, measured by the CONV collars, was lower than in the other treatments, as evidenced by the curve's concave nature (Fig. 3). Consequently, the daily maximum CO_2 efflux of CONV occurred at 08:30, an hour later than in the other treatments. The SHAL collars were also different from the other treatments in the timing of CO_2 uptake onset, occurring each day between 12:00-12:30, two hours before uptake started in the other treatments (Fig. 3).
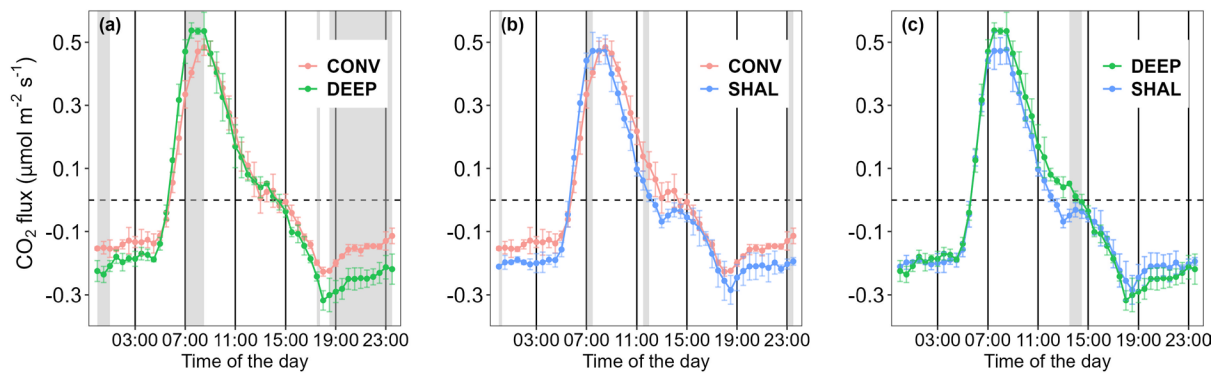
Fig. 3. Mean daily cycles of the soil CO_2 flux measured in the following collar types- A) The conventional (CONV) and deep (DEEP) insertion types. B) The conventional (CONV) and shallow (SHAL) types. C) The shallow (SHAL) and deep (DEEP) types. Error bars denote two standard deviations (n=30). Gray areas represent periods in which differences between the treatments were statistically significant (p-value<0.008).

The LMM model, combined with time series analysis, yielded statistically significant results (P<0.008) for the differences in F_s between CONV and DEEP during the morning (07:00-08:30) and the evening/night (17:30-01:00). In fact, F_s of CONV were consistently lower than in DEEP. The relative differences peaked at 06:00 and 23:30, when mean daytime CO_2 efflux and nocturnal CO_2 uptake were 56 and 53% lower in the CONV than in the DEEP. F_s measured in the CONV collars were also significantly lower than SHALL, by a maximum of 41%, but for shorter periods around noon and midnight. F_s measured in the DEEP collars were only significantly different from SHAL (P<0.008) from 13:30 to 14:30.

The mean peak daily efflux measured in the DEEP treatment differed significantly from the other two treatments (p<0.05), while no statistically significant difference in peak efflux was found for SHAL and CONV (one-way ANOVA and Tukey post hoc test). The differences between the total daily amount of CO_2 emitted during the day measured in SHAL and CONV were also insignificant (p>0.05; Table 2). In contrast, the total daily amounts of CO_2 uptaken by the soil in the CONV collars were significantly lower than in the SHAL and the DEEP collars (Table 2), which may lead to erroneous estimations of daily net CO_2 exchange.

**Table 2. Summary of main features- the mean daily cycles of Fs**

Period	Treatment	Max CO_2 efflux	Max CO_2 uptake	Total uptake	Total efflux
		$\mu mol\ m^{-2}\ s^{-1}$	$\mu mol\ m^{-2}\ s^{-1}$	$g\ m^{-2}$	$g\ m^{-2}$
1	CONV1	0.51±0.08	-0.28±0.04	0.43±0.077	0.29±0.04
	DEEP1	0.61±0.06	-0.38±0.05	0.54±0.05	0.39±0.06
	SHAL1	0.59±0.06	-0.38±0.06	0.57±0.10	0.32±0.05
2	CONV1	0.47±0.04	-0.26±0.03	0.30±0.05	0.33±0.04
	CONV2	0.51±0.06	-0.26±0.03	0.28±0.04	0.37±0.03
	CONV3	0.52±0.07	-0.27±0.02	0.31±0.06	0.32±0.04
3	CONV2	0.57±0.09	-0.25±0.04	0.36±0.11	0.39±0.04
	DEEP2	0.61±0.07	-0.36±0.03	0.53±0.13	0.34±0.09
	SHAL2	0.58±0.10	-0.35±0.03	0.49±0.12	0.33±0.08
4	DEEP1	0.64±0.08	-0.38±0.04	0.47±0.11	0.41±0.05
	DEEP2	0.67±0.11	-0.40±0.03	0.52±0.14	0.40±0.05
	DEEP3	0.57±0.10	-0.34±0.07	0.43±0.14	0.33±0.05
5	CONV3	0.55±0.04	-0.27±0.03	0.41±0.04	0.33±0.03
	DEEP3	0.60±0.01	-0.30±0.02	0.47±0.03	0.36±0.04
	SHAL3	0.48±0.04	-0.28±0.03	0.44±0.02	0.28±0.04
6	SHAL1	0.56±0.03	-0.32±0.01	0.46±0.11	0.34±0.01
	SHAL2	0.52±0.04	-0.28±0.03	0.37±0.08	0.28±0.03
	SHAL3	0.48±0.02	-0.27±0.02	0.35±0.09	0.29±0.03

228   Each value in the table is an average of 5 days ± one standard deviation.

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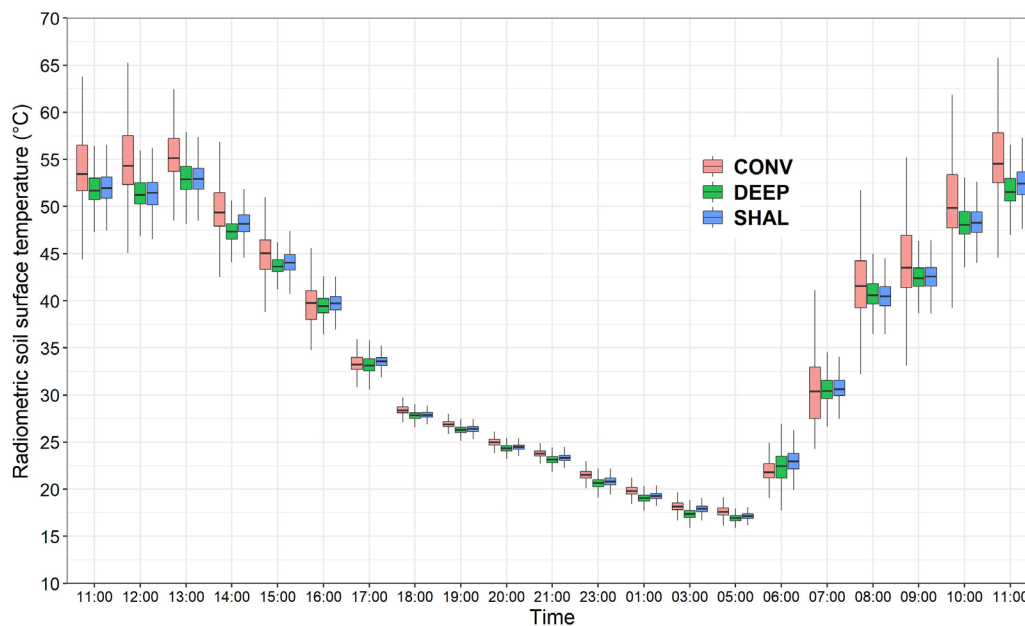230   **3.3   The effect of collar type on the radiometric soil surface temperature**

231   The mean and range of soil radiometric surface temperatures in the CONV collars were higher than in the DEEP

232   and SHAL collars, even at midday (Fig. 4). At 16:00, the three treatments all exhibited a mean surface temperature

233   of 40 °C, but the range of surface temperatures in the CONV collars doubled those of the other treatments. During

234   the night, the mean surface temperature of the CONV collars was 0.5-1 °C higher than in the DEEP collars and

235   0.5-0.9 °C higher than in the SHAL collars. After sunrise, the surface temperatures of the CONV and SHAL

236   increased faster than in the CONV collars up to 07:00. Later, the mean surface temperature of DEEP and SHAL

237   maintained a similar distribution over time, while the range and mean surface temperature in the CONV increased

238   sharply (Fig. 5).
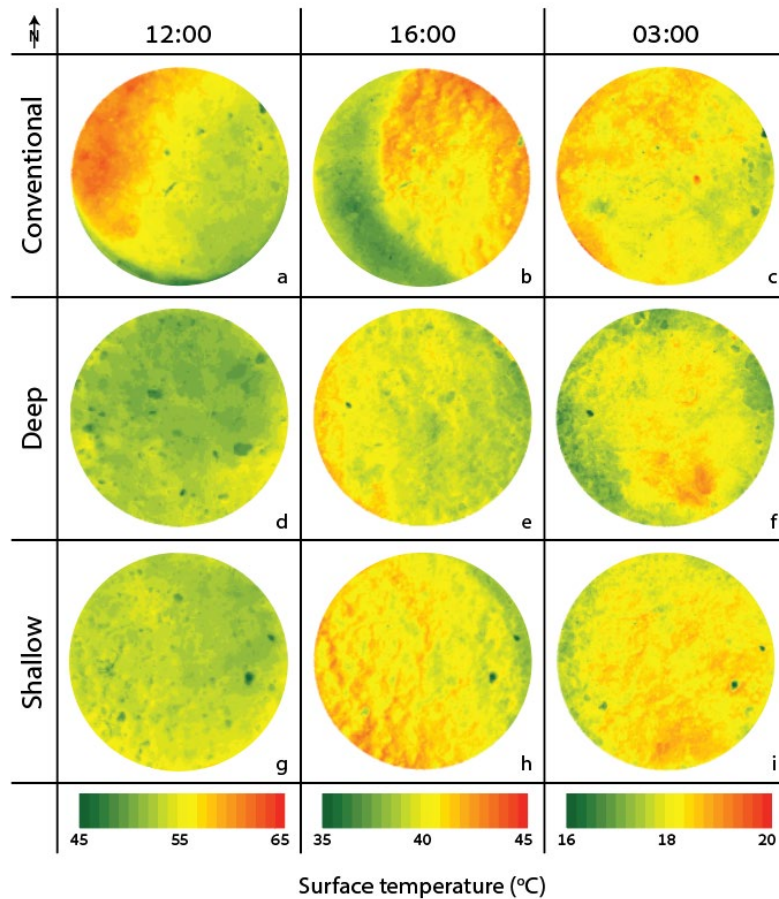
Figure 4: Thermal images of the soil surface radiometric temperature of one collar for each treatment in example hours of the day. A-C) The conventional treatment. D-F) The deep treatment. G-I) The shallow collar treatment. Note that each hour has a different temperature range.

Figure 5: Box plot and whiskers of the radiometric soil surface temperatures measured within the 3 types of collars on the 17-18/08/2021.

3.4   The effect of the soil heat flux on soil CO_2 flux

Changes in soil surface temperature induced by the collar treatment significantly affected F_s. Nonetheless, F_s and soil surface temperatures were uncoupled throughout the day and therefore may not be the sole variable that explains F_s dynamics (Figs. 3 and 5). For example, while the soil surface temperature decreased throughout the night, F_s decreased until the evening (18:00) and slowly increased during the night. However, the soil surface temperature has a prime effect on the temperature profile within the soil, as well as the direction and magnitude of soil heat flux. In fact, fig. 6 shows that F_s was linearly correlated with the soil heat flux, during the night and morning efflux. Later, F_s decreased earlier than the soil heat flux, resulting in a daytime hysteresis relationship (Fig. 6b).
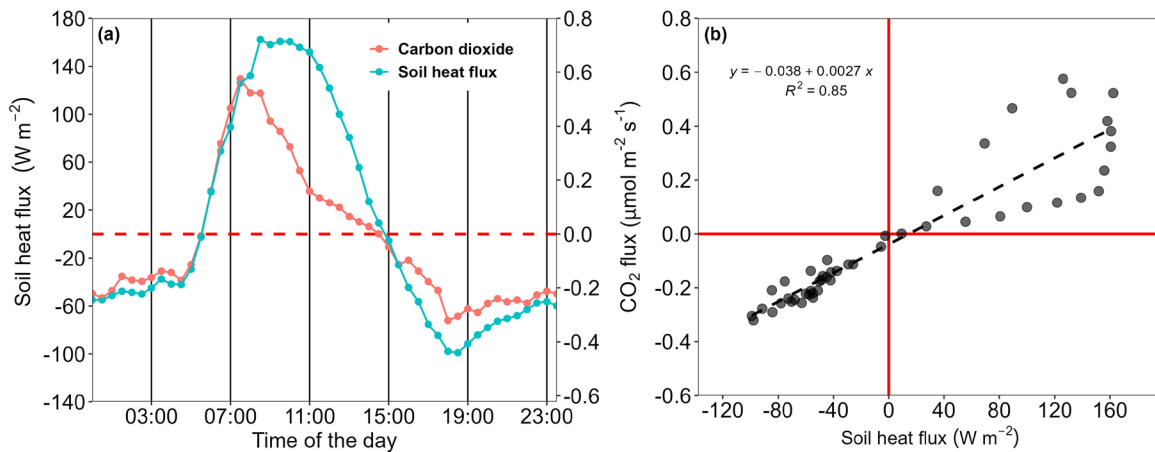
Figure 6: Relationship between the mean days of F_s and the soil heat flux for period 4 (9-16/06/2021). Note that positive F_s values indicate that the direction of the flux is from the soil to the atmosphere and vice versa for negative F_s values. Positive and negative soil heat flux values indicate the opposite directions than F_s values.
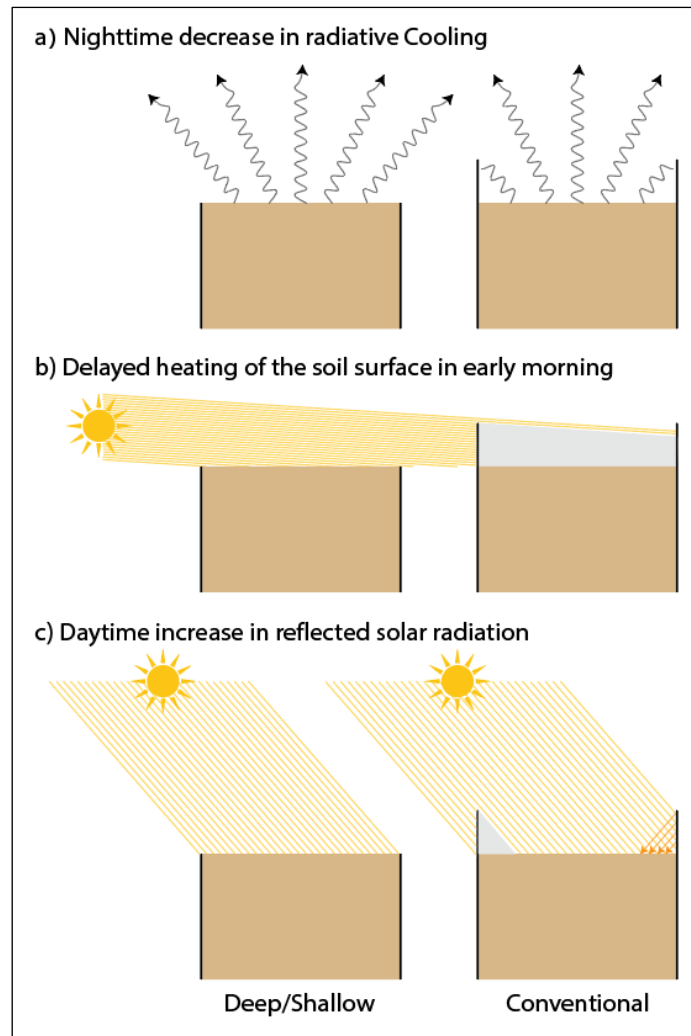
4   Discussion

Our study's results indicate that in dry and bare desert soils, using collars that protrude over the soil surface (CONV) can decrease F_s. This finding is consistent with a prior global assessment that identified a negative correlation between collar height above the soil surface and mean annual soil respiration rates (Jian et al., 2020). However, while we found that protruding collars resulted in significant errors of nearly 50% in F_s (Fig. 3 and table 2), Jian et al. (2020) demonstrated that collar height leads to a much smaller bias of only ~10% in annual soil respiration rates. They explained this bias by nonuniform air mixing within the chamber system resulting from the larger system volume but did not consider the potential effect of elevated collars on soil surface temperatures. Moreover, 85% of the annual soil respiration rate values Jian et al. (2020) used were estimated based on a limited number of instantaneous CO_2 efflux measurements, which were usually performed during the daytime, and, therefore, overlook diurnal dynamics in F_s. Since F_s is not constant throughout the day in desert soils but varies between daytime efflux and nocturnal uptake (Fig. 3), a small discontinuous number of daytime measurements will fail to capture errors in flux measurements. Finally, while most studies discussing potential sources of errors in F_s measurements were conducted in conditions where the dominant flux is a result of microbial respiration, in dry desert soils F_s is primarily driven by an abiotic process governed by changes in soil temperatures (Soper et al., 2017). Therefore, errors associated with using static chambers in dry desert soils are likely related

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276 to alteration of geochemical processes in the soil rather than affecting the factors that influence soil microbial
277 activity.

278 The abiotic process driving nocturnal CO_2 uptake in desert soils is often explained by the combined effect of
279 contraction and dissolution of gaseous CO_2 in soil water. These processes decrease gaseous CO_2 concentration in
280 the soil surface layer, forming an atmosphere-to-soil concentration gradient and CO_2 diffusion into the soil (Yang
281 et al., 2020; Sagi et al., 2021). Contraction of soil air may decrease CO_2 concentration in the soil surface layer and
282 lead to atmosphere-to-soil pressure gradient and thermal convection, which further contributes to CO_2 uptake
283 (Ganot et al., 2014). Soil temperature negatively affects both contraction and dissolution. Higher temperature
284 result in less contraction and dissolution, thus a higher CO_2 concentration in the surface air-filled soil-pores,
285 ultimately leading to a smaller soil-atmosphere CO_2 gradient, and lower F_s. It is therefore expected that a
286 modification of the surface temperature by the collar will affect the magnitude of the flux.

287 The elevated walls in the CONV collars limit nocturnal radiative cooling of the topsoil layer, resulting in higher
288 soil temperatures that suppress the CO_2 concentration gradient and the actual CO_2 uptake from the atmosphere
289 (Fig. 4 and fig. 7). Following sunrise, soil temperature increases in the DEEP and SHAL collars, promoting CO_2
290 expansion and outgassing from water films, rapidly increasing CO_2 efflux (Fa et al., 2016). This process is delayed
291 in the CONV collars because the surface is entirely shaded by the collar walls (Fig. 7b), resulting in a lower mean
292 temperature and a narrower overall range of surface temperatures (Fig. 5; 06:00 and Fig. 7b). As a result, CO_2
293 efflux increases at a slower rate (Fig. 3). When the sun elevation increases, solar radiation is reflected off the
294 collar walls into the measured area, increasing the radiation flux in the unshaded soil surface and, consequently,
295 increasing the mean and range of soil surface temperatures compared to the DEEP and SHAL collars (Figs. 4A-
296 B, 5 and 7). Thus, lower surface temperatures cannot explain the significantly lower CO_2 efflux measured in the
297 CONV collars between 07:00 and 08:30. Instead, it is probably related to the significantly lower total nighttime
298 CO_2 uptake, which leads to a faster depletion of soil CO_2 in the following morning (Table 2).

a) Nighttime decrease in radiative Cooling

b) Delayed heating of the soil surface in early morning

c) Daytime increase in reflected solar radiation

Deep/Shallow   Conventional

Figure 7: Conceptual model showing the effects of collar deployment on soil surface radiative heating and cooling during the night (a), early morning (b), and daytime (c).

The results of our study indicate that lateral diffusion is not a significant concern in dry, bare desert soils when the measurement period (i.e., the length of time during which the chamber is closed over the collar) is short, as demonstrated by the insignificant differences between F_s measured over the SHAL and DEEP collars. This confirms the findings of Hutchinson and Livingston, (2001). Although statistically insignificant, the mean CO_2 efflux in the SHAL collars was consistently lower than in the DEEP collars between 7:00 and 14:30 (Fig. 3 and table 2). Additionally, the flux direction measured over the SHAL collars, consistently changed from efflux (positive) to uptake (negative) earlier than in the other treatments, and earlier than the soil heat flux changed from positive to negative (Fig. 6). A change in the soil heat flux sign indicates that temperatures in the uppermost soil layer are decreasing, promoting the removal of gaseous CO_2 from the soil air phase, followed by CO_2 uptake from the atmosphere. Hence, when soil temperatures are undisturbed (e.g., by the presence of a collar), we expect the onset of CO_2 uptake to coincide with the change in soil heat flux direction (Fig. 6). The only difference between the SHAL and DEEP collars was their insertion depth (in both the collar's top end was flashed with the soil surface). Root cutting, which is often suggested as an explanation for lower F_s measured over deeper collars (Heinemeyer et al., 2011), is inapplicable when the soil is sparsely vegetated. Furthermore, our results show higher F_s values when measured over deeply inserted collars (DEEP) then when measured over shallow collars (SHAL).

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Potential overestimation of F_s resulting from enhanced air flow along the collar walls in the DEEP collars was minimized by inserting the collars more than two months prior the measurements, a sufficiently long time to allow the soil to settle around them (Hutchinson and Livingston, 2001). Lateral diffusion below the shallow collars therefore remains the most probable explanation. As suggested by Healy et al. (1996), lateral movement likely decreased the CO_2 concentration in the soil top layer during CO_2 efflux, decreasing the concentration gradient between the soil and the chamber headspace, resulting in an underestimation of F_s. The lower soil CO_2 concentration beneath the SHAL collars caused the concentration gradient that drives the vertical flux to reverse direction toward the soil, starting CO_2 uptake earlier than in the other treatments (fig. 3).

The conventionally deployed collars (CONV) underestimated the instantaneous CO_2 uptake and thus the total CO_2 uptake during the night (table 2). This suggests that the actual carbon sequestration by desert soils is higher than previously reported. In some cases, the net daily exchange measured in the CONV collars is even positive, indicating a net efflux of CO_2 to the atmosphere (Table 2). Note, however, that the net daily values measured by the CONV collars are very small, thus more susceptible to errors, to the point of flipping the direction, and concluding from the absolute daily net values must be done with caution. Theoretically, if F_s in dry desert soils is derived by abiotic geochemical processes, a balanced net daily cycle would be expected, where nocturnal CO_2 uptake is compensated by daytime efflux. Even in alkaline soils, such as the ones in our study site, where the nocturnal dissolution of $CaCO_3$ may sustain CO_2 uptake from the atmosphere, the reverse reaction should occur when water evaporates and $CaCO_3$ precipitates, promoting CO_2 efflux and system equilibrium (Roland et al., 2013). This hypothesis was supported by Hamerlynck et al. (2013) who found that a soil in the Chihuahuan Desert, USA, only serves as a minor carbon sink (0.88 g C m^{-2} accumulated over three months) and concluded that this contribution is insignificant to the global carbon balance. Contrarily, in the Taklamakan (Yang et al., 2020) and the Gubantonggut (Xie et al., 2009) Deserts in China, nocturnal CO_2 uptake led to a mean annual uptake of 7.11 and 62-622 g C m^{-2}, respectively. This gave rise to the hypothesis that nocturnal CO_2 uptake by desert soils might explain a substantial portion of the global missing sink. However, they did not provide a mechanism to explain where the carbon is stored, especially given that the leaching of dissolved carbonates to groundwater is limited in space and time (Ma et al., 2014; Schlesinger, 2017; Yang et al., 2022). Furthermore, the abiotic component of F_s contributed 21% of mean CO_2 efflux in a semi-arid pine forest located ~35 km north-east of our study site and therefor functioned as a source for atmospheric carbon rather than as a sink in that ecosystem (Qubaja et al., 2020). Either way, no conclusions can be drawn about the role desert soils play in the global carbon balance until a methodology to measure these small fluxes is proved to be accurate. Our study shows that instantaneous F_s and F_s daily balance could be significantly affected by even as small as a few centimeters difference in collar height and depth. This implies that previous estimates of the carbon balance of desert ecosystems using static chambers need to be carefully considered.

In fact, studies show that the abiotic mechanisms involved in F_s are not restricted to dry desert conditions but rather play a significant role in Fs in deserts under wet soil conditions (Fa et al., 2016). This was found for both a semi-arid pine forest(Qubaja et al., 2020), and a temperate grassland (Plestenjak et al., 2012). Hence, the collar disruption to abiotic processes likely affects the carbon balance in various ecosystems beyond the scope of deserts during the dry season. Alteration of F_s due to collar insertion is not restricted to abiotic processes. The soil biological processes, and specially activity of biological soil crust, may be significantly affected by altered soil

surface conditions. Since they cover a vast area of Earth's drylands, and play a significant role in desert ecosystem's carbon balance (Wilske et al., 2008), it is important to consider these effects.

5   Summary and Conclusions

The drivers of abiotic soil CO_2 flux observed in dry desert soils are yet far from being understood. Further research is needed to reconcile the discrepancy between the theoretical basis, which suggests a balanced daily cycle, and field measurements, which often show net uptake by the soil in both diel and annual scales. Particularly, studies should focus on improving our understanding of CO_2 in the soil profile in desert soils, and on allocating the sources of water that are assumed to act as a solvent for CO_2 even when the soil is dry. None of these questions, however, can be addressed without an accurate methodology to measure the small F_s characterizing bare desert soils.

During a two months measurement period in the summer of 2021, the soil in the Wadi Mashash Experimental farm exhibited a repetitive diel cycle of CO_2 flux that consisted of nocturnal CO_2 uptake and daytime efflux, driven by a combination of physical and geochemical processes in the soil. We show here for the first time that collar deployment practices significantly affect this abiotic diel cycle by altering the factors that drive F_s. Notably, morning CO_2 efflux and nocturnal CO_2 uptake were underestimated when measured on conventionally inserted collars because the elevated collar walls distorted the ambient surface temperature regime. We conclude that in bare desert soils collars should be deployed flashed with the soil surface to prevent distortion of heat exchange between the soil and the atmosphere and between soil layers, two important drivers of the abiotic F_s. Lateral diffusion under shallow collars may occur and affect F_s' temporal dynamics. However, we found this to be of a lesser concern in compact soils and short measurement periods. Still, in dry desert soils, the collar insertion depth should exceed the depth at which the fluctuations in soil CO_2 concentration that drive F_s occur, roughly 2 cm (Hamerlynck et al., 2013).

Deployment protocols of flux chambers should be adapted to the unique characteristics of desert soils rather than follow standard procedures suitable for mesic environments. We conclude that using collars with at least 3 cm length inserted flush with the soil surface will minimize measurement errors of CO_2 flux and will pave the way to accurate estimates of the carbon balance of desert ecosystems.

6   Code/data availability

Code and data will be provided upon request.

7   Author contributions

Nadav Bekin: Conceptualization, Data curation, Formal analysis, Investigation, Methodology, Writing - original draft. **Nurit Agam**: Conceptualization, Funding acquisition, Methodology, Project administration, Resources, Supervision, Writing - review & editing.

8   Competing interests

The authors have no competing interests.

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10   References

Austin, A. T., Yahdjian, L., Stark, J. M., Belnap, J., Porporato, A., Norton, U., Ravetta, D. A., and Schaeffer, S. M.: Water pulses and biogeochemical cycles in arid and semiarid ecosystems, Oecologia, 141, 221–235, https://doi.org/10.1007/s00442-004-1519-1, 2004.

Bain, W. G., Hutyra, L., Patterson, D. C., Bright, A. V., Daube, B. C., Munger, J. W., and Wofsy, S. C.: Wind-induced error in the measurement of soil respiration using closed dynamic chambers, Agric. For. Meteorol., 131, 225–232, https://doi.org/10.1016/j.agrformet.2005.06.004, 2005.

Ball, B. A., Virginia, R. A., Barrett, J. E., Parsons, A. N., and Wall, D. H.: Interactions between physical and biotic factors influence CO_2 flux in Antarctic dry valley soils, Soil Biol. Biochem., 41, 1510–1517, https://doi.org/10.1016/j.soilbio.2009.04.011, 2009.

Baram, S., Bar-Tal, A., Gal, A., Friedman, S. P., and Russo, D.: The effect of static chamber base on N_2O flux in drip irrigation, Biogeosciences, 19, 3699–3711, https://doi.org/10.5194/bg-19-3699-2022, 2022.

Brutsaert, W.: Evaporation into the Atmosphere, https://doi.org/10.1007/978-94-017-1497-6, 1982.

Cable, J. M., Ogle, K., Williams, D. G., Weltzin, J. F., and Huxman, T. E.: Soil texture drives responses of soil respiration to precipitation pulses in the sonoran desert: Implications for climate change, Ecosystems, 11, 961–979, https://doi.org/10.1007/s10021-008-9172-x, 2008.

Cable, J. M., Ogle, K., Lucas, R. W., Huxman, T. E., Loik, M. E., Smith, S. D., Tissue, D. T., Ewers, B. E., Pendall, E., Welker, J. M., Charlet, T. N., Cleary, M., Griffith, A., Nowak, R. S., Rogers, M., Steltzer, H., Sullivan, P. F., and van Gestel, N. C.: The temperature responses of soil respiration in deserts: A seven desert synthesis, Biogeochemistry, 103, 71–90, https://doi.org/10.1007/s10533-010-9448-z, 2011.

Fa, K., Zhang, Y., Lei, G., Wu, B., Qin, S., Liu, J., Feng, W., and Lai, Z.: Underestimation of soil respiration in a desert ecosystem, Catena, 162, 23–28, https://doi.org/10.1016/j.catena.2017.11.019, 2018.

Fa, K. Y., Zhang, Y. Q., Wu, B., Qin, S. G., Liu, Z., and She, W. W.: Patterns and possible mechanisms of soil CO_2 uptake in sandy soil, Sci. Total Environ., 544, 587–594, https://doi.org/10.1016/j.scitotenv.2015.11.163, 2016.

Ganot, Y., Dragila, M. I., and Weisbrod, N.: Impact of thermal convection on CO_2 flux across the earth-atmosphere boundary in high-permeability soils, Agric. For. Meteorol., 184, 12–24, https://doi.org/10.1016/j.agrformet.2013.09.001, 2014.

17

421   Hamerlynck, E. P., Scott, R. L., Sánchez-Cañete, E. P., and Barron-Gafford, G. A.: Nocturnal soil CO_2 uptake

422   and its relationship to subsurface soil and ecosystem carbon fluxes in a Chihuahuan Desert shrubland, J. Geophys.

423   Res. Biogeosciences, 118, 1593–1603, https://doi.org/10.1002/2013JG002495, 2013.

424   Healy, R. W., Striegl, R. G., Russell, T. F., Hutchinson, G. L., and Livingston, G. P.: Numerical Evaluation of

425   Static-Chamber Measurements of Soil-Atmosphere Gas Exchange: Identification of Physical Processes, Soil Sci.

426   Soc. Am. J., 60, 740–747, https://doi.org/10.2136/sssaj1996.03615995006000030009x, 1996.

427   Heinemeyer, A., Di Bene, C., Lloyd, A. R., Tortorella, D., Baxter, R., Huntley, B., Gelsomino, A., and Ineson,

428   P.: Soil respiration: Implications of the plant-soil continuum and respiration chamber collar-insertion depth on

429   measurement and modelling of soil CO_2 efflux rates in three ecosystems, Eur. J. Soil Sci., 62, 82–94,

430   https://doi.org/10.1111/j.1365-2389.2010.01331.x, 2011.

431   Houghton, R. A.: Balancing the global carbon budget, Annu. Rev. Earth Planet. Sci., 35, 313–347,

432   https://doi.org/10.1146/annurev.earth.35.031306.140057, 2007.

433   Jian, J., Gough, C., Sihi, D., Hopple, A. M., and Bond-Lamberty, B.: Collar Properties and Measurement Time

434   Confer Minimal Bias Overall on Annual Soil Respiration Estimates in a Global Database, J. Geophys. Res.

435   Biogeosciences, 125, 1–13, https://doi.org/10.1029/2020JG006066, 2020.

436   Li, Z. L., Wu, H., Wang, N., Qiu, S., Sobrino, J. A., Wan, Z., Tang, B. H., and Yan, G.: Land surface emissivity

437   retrieval from satellite data, Int. J. Remote Sens., 34, 3084–3127, https://doi.org/10.1080/01431161.2012.716540,

438   2013.

439   Lopez-Canfin, C., Lázaro, R., and Sánchez-Cañete, E. P.: Water vapor adsorption by dry soils: A potential link

440   between the water and carbon cycles, Sci. Total Environ., 824, https://doi.org/10.1016/j.scitotenv.2022.153746,

441   2022.

442   Lund, C. P., Riley, W. J., Pierce, L. L., and Field, C. B.: The effects of chamber pressurization on soil-surface

443   CO_2 flux and the implications for NEE measurements under elevated CO_2, Glob. Chang. Biol., 5, 269–281,

444   https://doi.org/10.1046/j.1365-2486.1999.00218.x, 1999.

445   Ma, J., Liu, R., Tang, L. S., Lan, Z. D., and Li, Y.: A downward CO_2 flux seems to have nowhere to go,

446   Biogeosciences, 11, 6251–6262, https://doi.org/10.5194/bg-11-6251-2014, 2014.

447   Ngao, J., Longdoz, B., Perrin, D., Vincent, G., Epron, D., Le Dantec, V., Soudani, K., Aubinet, M., Willim, F.,

448   and Granier, A.: Cross-calibration functions for soil CO_2 efflux measurement systems, Ann. For. Sci., 63, 477–

449   484, https://doi.org/10.1051/forest:2006028, 2006.

450   Ninari, N. and Berliner, P. R.: The role of dew in the water and heat balance of bare loess soil in the Negev Desert:

451   Quantifying the actual dew deposition on the soil surface, Atmos. Res., 64, 323–334,

452   https://doi.org/10.1016/S0169-8095(02)00102-3, 2002.

453   Parkin, T. B., Venterea, R. T., and Hargreaves, S. K.: Calculating the Detection Limits of Chamber-based Soil

454   Greenhouse Gas Flux Measurements, J. Environ. Qual., 41, 705–715, https://doi.org/10.2134/jeq2011.0394, 2012.

455   Parsons, A. N., Barrett, J. E., Wall, D. H., and Virginia, R. A.: Soil Carbon Dioxide Flux in Antarctic Dry Valley

456   Ecosystems, Ecosystems, 286–295, https://doi.org/10.1007/s10021-003-0132-1, 2004.

457   Plestenjak, G., Eler, K., and Vodnik, D.: Sources of soil CO_2 in calcareous grassland with woody plant
458   encroachment, J. soils sediments, 1327–1338, https://doi.org/10.1007/s11368-012-0564-3, 2012.

459   Pumpanen, J., Kolari, P., Ilvesniemi, H., Minkkinen, K., Vesala, T., Niinistö, S., Lohila, A., Larmola, T., Morero,
460   M., Pihlatie, M., Janssens, I., Yuste, J. C., Grünzweig, J. M., Reth, S., Subke, J. A., Savage, K., Kutsch, W.,
461   Østreng, G., Ziegler, W., Anthoni, P., Lindroth, A., and Hari, P.: Comparison of different chamber techniques for
462   measuring soil CO_2 efflux, Agric. For. Meteorol., 123, 159–176,
463   https://doi.org/10.1016/j.agrformet.2003.12.001, 2004.

464   Pumpanen, J., Longdoz, B., and Kutsch, W. L.: Field measurements of soil respiration: Principles and constraints,
465   potentials and limitations of different methods, Soil Carbon Dyn. An Integr. Methodol., 16–33,
466   https://doi.org/10.1017/CBO9780511711794.003, 2010.

467   Qubaja, R., Tatarinov, F., Rotenberg, E., and Yakir, D.: Partitioning of canopy and soil CO_2 fluxes in a pine
468   forest at the dry timberline across a 13-year observation period, Biogeosciences, 4, 699–714, 2020.

469   Rochette, P. and Eriksen-Hamel, N. S.: Chamber Measurements of Soil Nitrous Oxide Flux: Are Absolute Values
470   Reliable?, Soil Sci. Soc. Am. J., 72, 331–342, https://doi.org/10.2136/sssaj2007.0215, 2008.

471   Roland, M., Serrano-Ortiz, P., Kowalski, A. S., Goddéris, Y., Sánchez-Cañete, E. P., Ciais, P., Domingo, F.,
472   Cuezva, S., Sanchez-Moral, S., Longdoz, B., Yakir, D., Van Grieken, R., Schott, J., Cardell, C., and Janssens, I.
473   A.: Atmospheric turbulence triggers pronounced diel pattern in karst carbonate geochemistry, Biogeosciences, 10,
474   5009–5017, https://doi.org/10.5194/bg-10-5009-2013, 2013.

475   Sagi, N., Zaguri, M., and Hawlena, D.: Soil CO_2 influx in drylands: A conceptual framework and empirical
476   examination, Soil Biol. Biochem., 156, 108209, https://doi.org/10.1016/j.soilbio.2021.108209, 2021.

477   Soper, F. M., McCalley, C. K., Sparks, K., and Sparks, J. P.: Soil carbon dioxide emissions from the Mojave
478   desert: Isotopic evidence for a carbonate source, Geophys. Res. Lett., 44, 245–251,
479   https://doi.org/10.1002/2016GL071198, 2017.

480   Spyroglou, I., Skalák, J., Balakhonova, V., Benedikty, Z., Rigas, A. G., and Hejátko, J.: Mixed models as a tool
481   for comparing groups of time series in plant sciences, Plants, 10, 1–16, https://doi.org/10.3390/plants10020362,
482   2021.

483   Stell, E., Warner, D., Jian, J., Bond-Lamberty, B., and Vargas, R.: Spatial biases of information influence global
484   estimates of soil respiration: How can we improve global predictions?, Glob. Chang. Biol., 27, 3923–3938,
485   https://doi.org/10.1111/gcb.15666, 2021.

486   Wilske, B., Burgheimer, J., Karnieli, A., Zaady, E., Andreae, M. O., Yakir, D., and Kesselmeier, J.: The CO_2
487   exchange of biological soil crusts in a semiarid grass-shrubland at the northern transition zone of the Negev desert
488   , Israel, Biogeosciences, 1411–1423, 2008.

489   Xie, J., Li, Y., Zhai, C., Li, C., and Lan, Z.: CO_2 absorption by alkaline soils and its implication to the global
490   carbon cycle, Environ. Geol., 56, 953–961, https://doi.org/10.1007/s00254-008-1197-0, 2009.

491   Yang, F., Huang, J., He, Q., Zheng, X., Zhou, C., Pan, H., Huo, W., Yu, H., Liu, X., Meng, L., Han, D., Ali, M.,

492   and Yang, X.: Impact of differences in soil temperature on the desertcarbon sink, Geoderma, 379, 114636,

493   https://doi.org/10.1016/j.geoderma.2020.114636, 2020.

494   Yang, F., Huang, J., Zheng, X., Huo, W., Zhou, C., Wang, Y., Han, D., Gao, J., Mamtimin, A., Yang, X., and

495   Sun, Y.: Evaluation of carbon sink in the Taklimakan Desert based on correction of abnormal negative CO_2 flux

496   of IRGASON, Sci. Total Environ., 838, 155988, https://doi.org/10.1016/j.scitotenv.2022.155988, 2022.

497

498

499