

1 **Source differences in the components and cytotoxicity of PM_{2.5} from**
2 **automobile exhaust, coal combustion, and biomass burning**
3 **contributing to urban aerosol toxicity**

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15 **Abstract.** Although air quality guidelines generally use the atmospheric concentration of fine particulate matter (PM_{2.5}) as the
16 metric for air pollution evaluation and management, the fact can't be ignored that different particle toxicity is unequal and
17 significantly related to their sources and chemical compositions. Therefore, judging the most harmful source and identifying
18 the toxic component would be helpful to optimize air quality standards and prioritize targeted PM_{2.5} control strategies to protect
19 public health more effectively. Since the combustions of fuels, including oil, coal, and biomass, are main anthropogenic sources
20 of environmental PM_{2.5}, their discrepant contributions to health risks of mixed ambient aerosol pollution dominated by
21 respective emission intensity and unequal toxicity of chemical components need to be identified. In order to quantify the
22 differences among these combustion primary emissions, ten types of PM_{2.5} from each typical source group, i.e., vehicle exhaust,
23 coal combustion, and plant biomass (domestic biofuel) burning, were collected for comparative study with toxicological
24 mechanisms. Totally thirty type individual combustion samples were inter-compared with representative urban ambient air
25 PM_{2.5} samples, which chemical characteristics and biological effects were investigated by component analysis (carbon, metals,
26 soluble ions) and *in vitro* toxicity assays (cell viability, oxidative stress, inflammatory responses) of human lung
27 adenocarcinoma epithelial cells (A549). Carbonaceous fractions were plenteous in automobile exhaust and biomass burning,
28 while heavy metals were more plentiful in PM_{2.5} from coal combustion and automobile exhaust. The overall ranking of mass-
29 normalized cytotoxicity for source-specific PM_{2.5} was automobile exhaust > coal combustion > domestic plant biomass
30 burning > ambient urban air, possibly with differential toxicity triggers, that the carbonaceous fractions (organic carbon, OC;
31 elemental carbon, EC) and redox-active transition metals (V, Ni, Cr) assisted by water-soluble ions (Ca²⁺, Mg²⁺, F⁻, Cl⁻) might

32 play important roles in inducing cellular reactive organic species (ROS) production, causing oxidative stress and inflammation,
33 resulting in cell injury and apoptosis, thus damage human health. Coupled with the source apportionment results of typical
34 urban ambient air PM_{2.5} in eastern China, reducing toxic PM_{2.5} from these anthropogenic combustions will be greatly beneficial
35 to public health. Besides the air pollution control measures that have been implemented, like strengthening the vehicle emission
36 standards, energy switching from coal to gas and electricity, and controlling the open incineration of agricultural straws, further
37 methods could be considered especially through preferentially reducing the diesel exhaust, then lessening the coal combustion
38 by replacement with low-ash clean coals, and depressing the rural crop straw biomass burning emissions.

39

40 **1 Introduction**

41 As a mixture of multiple sources, ambient particulate matter (PM) arise from anthropogenic activities are continuously
42 deteriorating the urban air quality, particularly in developing countries. Among these, fine PM with an aerodynamic diameter
43 of less than 2.5 μm (PM_{2.5}) is recognized as a serious public health concern due to its long persistence in air, carcinogenicity
44 and acute toxicity to humans (Al-Kindi et al., 2020). There were extensive epidemiological evidences that airborne PM can
45 cause serious negative effects on human health, such as respiratory and cardiovascular diseases, genetic mutations, and
46 developmental disorders (Chowdhury et al., 2022;Lelieveld et al., 2021;Smith, 2021;Clemens et al., 2017). Currently, either
47 the world air quality guidelines or the national air quality standards use the mass concentration of PM_{2.5} as the metric for PM_{2.5}
48 pollution evaluation and management, however, the particle toxicity are unequal and significantly related to their sources and
49 chemical compositions varying with space and time (Shiraiwa et al., 2017). Therefore, to identify which component(s) and
50 source(s) of ambient PM are most harmful to health, will be helpful to evaluate air quality and prioritize targeted PM control
51 strategies for protecting public health more effectively.

52 Besides natural sources, most aerosols come from anthropogenic activities especially energy consumption, including the
53 combustion of fossil fuels causing industrial emissions and automobile exhaust, and biomass burning (McDuffie et al.,
54 2021;Wu et al., 2022). These diverse sources make the ambient air PM_{2.5} become a complex mixture with multiple chemical
55 components, such as salts, organic carbon (OC), elemental carbon (EC), mineral and trace metals (Bari and Kindzierski, 2016).
56 The physiological mechanisms of PM-induced cell toxicity in respiratory system have been continuously investigated with
57 some progresses (Kelly and Fussell, 2012, 2020; Shiraiwa et al., 2017; Mack et al., 2020; Li et al., 2022b), such as the metabolic
58 activation, oxidative stress, inflammatory response, and apoptosis, focused on by current study. In brief, after inhalation and
59 deposition onto the epithelium, redox-active materials in PM_{2.5} can induce the release of reactive organic species (ROS), which
60 cause oxidative stress (an imbalance between ROS and antioxidants, i.e., disequilibrium of the redox state of a cell) followed
61 by inflammation and cell death. The ROS can mediate subsequent signaling pathways leading to biomolecule damage (e.g.,
62 DNA, lipid, and protein) and cellular injury, through mediating inflammatory responses including the release of pro-
63 inflammatory cytokines like IL-6 and TNF- α by epithelial cells (Sabbir Ahmed et al., 2020; Landwehr et al., 2021). For

64 instance, oxidative stress could trigger the induction of pro-inflammatory transcription factors, such as nuclear factor (NF)- κ B,
65 via the mitogen-activated protein kinase (MAPK) signaling pathway. Components adsorbed on particle surface, such as redox-
66 active metals (transition metals, Fe, Ni, V, Cr, Cu), organic compounds (polycyclic aromatic hydrocarbons, PAHs; quinones),
67 or even carbonaceous core of particles, are responsible for oxidative stress (Cachon et al., 2014; Sabbir Ahmed et al., 2020).
68 The non-redox active metals (Zn, Pb, Al) can also influence the toxic effects of transition metals by exacerbating or lessening
69 the production of free radicals. The EC may not be a directly toxic component of PM_{2.5} but rather operate as a universal carrier
70 of combustion-derived chemicals (semi-volatile organic fractions, transition metals) of varying toxicity (Kelly and Fussell,
71 2020). Inorganic soluble sulphates and nitrates are acidic and can interact with and influence the solubility other compositions
72 like metal bioavailability (Fang et al., 2017; Weber et al., 2016). However, besides the well-known toxic pollutants in
73 environment like heavy metals and PAHs, which specific components and which particular sources are the most critical factors
74 dominating the ambient aerosols' health risks, still need be explored.

75 Past studies performed in various countries have focused on physicochemical characterization or biological effects of
76 ambient air PM_{2.5} respectively (Weagle et al., 2018; Jia et al., 2017; Wang et al., 2020). For example, the source analysis of
77 PM_{2.5} by photochemical modelling (Bao et al., 2018), chemical composition of regional PM_{2.5} (Chi et al., 2022), and the
78 mechanism of PM_{2.5} toxicity was independently reported recently (Jia et al., 2020). Because differences in particle composition,
79 sources, and toxicity appear in different urban environments (Zhao et al., 2019; Borlaza et al., 2018), the source profiles of
80 different emission inventories were needed to elucidate the local aerosol pollution characteristics for control strategies. For
81 instance, it was reported that increased hospital admission risks were significantly associated with sources of vehicle exhaust,
82 coal combustion, and secondary inorganic aerosols; in particular, coal combustion was positively correlated with increases in
83 mortality risks (Du et al., 2022). Coal combustion and vehicle exhaust contributed more significantly to cancer risks of
84 respiratory exposure to atmospheric heavy metals in Tianjin city of the northern China during the cold seasons (31% and 11%)
85 than the warm seasons (11% and 4%) (Tian et al., 2021); while in Nanjing city of the eastern China, traffic emissions and non-
86 traffic combustion (coal/waste/biomass) contributed 35% and 31% to carcinogenic risks of urban PM_{2.5}-associated metals
87 respectively (Xie et al., 2020). Traffic was suggested playing the most crucial role in enhancing the toxicity of fine particles
88 (Park et al., 2018). The particle composition of motor vehicle exhaust was related to automobile types with various fuels,
89 engines, and loads (Lin et al., 2020). A strong catalytic reactivity of metals in PM emitted from diesel vehicles was observed
90 by dithiothreitol (DTT) assay (Jesus et al., 2018). It was found that straw burning during the harvest season is a major trigger
91 of severe air pollution in many regions (Sahu et al., 2021). Aerosols from open biomass burning in the Amazon had a stronger
92 ability to induce ROS than laboratory-generated secondary organic aerosols (Tuet et al., 2019). Although there were emerging
93 studies on particle emission from single source, quantitatively comparative studies on multi-source pollutants as well as the
94 differential composition and unequal toxicity of various sources are still limited.

95 The main objective of current study was to compare the chemical components and corresponding mass-normalized
96 toxicological effects of individual PM_{2.5} from various combustion sources and their unequal contributions to ambient aerosol

97 health risks. The aim is to provide experimental evidences supporting the targeting control of specific anthropogenic sources
98 with prominent risks based on their pivotal toxic components. Therefore, we collected both representative ambient PM_{2.5}
99 samples (n = 16) from urban air and typical source PM_{2.5} samples (n = 30) from automobile exhaust, coal combustion, and
100 plant biomass burning. Their independent profiles of chemical compositions and *in vitro* cytotoxicity (cell viability, oxidative
101 stress, and inflammatory responses) were investigated and intercompared, to assess the differences in source-to-receptor
102 toxicity and to infer the core toxic components and respective harmful contribution. The pivotal toxic components were
103 identified based on the source-sink bi-directional composition-effect results, which were further used to assess the health
104 toxicity contribution of various emission sources to ambient air PM_{2.5}, supported by its source apportionment through positive
105 matrix factorization (PMF) and chemical mass balance (CMB) models.

106 **2 Materials and methods**

107 **2.1 Collection of PM_{2.5} samples from primary emissions of 30 typical combustion sources and from representative** 108 **ambient urban air**

109 Totally 30 types of primary PM_{2.5} samples emitted directly from automobile exhaust, coal combustion, and plant biomass
110 (domestic biofuel) burning were respectively collected as follows for both chemical and toxicological analyses.

111 A total of 10 types of vehicles were chosen for exhaust investigation. They were further categorized into 7 sub-groups,
112 including small duty gasoline coaches (SDGCs), small duty diesel coaches (SDDCs), middle duty diesel coaches (MDDCs),
113 heavy duty diesel coaches (HDDCs), light duty diesel vans (LDDVs), middle duty diesel vans (MDDVs), and heavy duty
114 diesel vans (HDDVs). The detailed information of these representative local automobiles was showed in Table S1.

115 To cover all coal types consumed in the city, 10 representative types of coal were gathered for investigation. They were
116 further classified into 4 sub-groups, including 2 types of honeycomb coal (HC), 3 types of anthracite coal (AC), and 2 types
117 of bituminous coal (BC) mainly for restaurant or household use, and 3 types of industrial coal (IC) for coal-fired power plants
118 and steel-smelting industry. The detailed characteristic to physical-chemical of these typical coals purchased from local market
119 were showed in Table S2.

120 Considering the plant biomass combustion in rural areas surrounding the megacity, 10 representative types of agricultural
121 and forestry solid wastes were gathered for investigation. Straws of rice, wheat, corn, soybean, peanut, rape, and sesame,
122 corncob, branches of peach and pine, were selected as plant biomass fuels and further divided into 2 sub-groups, including 8
123 types of crop straw and 2 types of firewood. The detailed characteristic analysis of these typical plant biomass fuels collected
124 from rural areas around Nanjing city were showed in Table S3.

125 The PM_{2.5} samples directly emitted from these combustion sources were collected by dilution channel sampling method
126 (Figure S1), using a 4-channel particulate matter dilution sampler (HY-805, Hengyuan Technology Development Co., CN).
127 Each sampling included 3 parallel channels of quartz microfiber filter (Figure S2) and 1 channel of Teflon membrane filter
128 with diameters of 47 mm, through a size selector for PM_{2.5} with a flow rate of 160 L min⁻¹ (each channel is 40 L/min). Clean

129 air was pumped for 10 min before and after each sample was collected. Before using, the blank quartz filters were incinerated
130 by a muffle furnace at 500 °C for 3 h to remove any possible organic matters, while Teflon filters were baked at 60 °C for 4 h.
131 After being equilibrated in a constant temperature and humidity chamber for 24 h, the filters were weighed both before and
132 after sampling for gravimetric measurements, then the mass of collected PM_{2.5} could be calculated. The sampled filters were
133 stored in a refrigerator at -20 °C before analysis. The quartz filter loaded PM_{2.5} samples were used for carbon and ion analysis,
134 and for toxicity tests, while the parallel Teflon filter loaded samples were used to determine metals.

135 As the actual mixture of various source particles in real environment, totally 16 ambient air PM_{2.5} samples (each time lasting
136 23h) covering a year monthly were collected from December 2019 to October 2020 in an urban site surrounded by traffic,
137 residential and commercial quarters of Nanjing city, Yangtze River Delta of eastern China, using a high-volume air sampler
138 (800 L min⁻¹) with quartz microfiber filters (Li et al., 2022a).

139 **2.2 Chemical composition analysis**

140 All collected source and ambient PM_{2.5} samples were conducted following component analysis (Li et al., 2023). For the
141 concentrations of heavy metals in particulates, samples were digested by concentrated HNO₃-HClO₄ acids with a progressive
142 heating program and determined by inductively coupled plasma optical emission spectrometry (ICP-OES; Optima8000,
143 PerkinElmer, for Cr, Mn, Ni and Pb), with elements (V, Co, As) at lower concentrations measured by ICP mass spectrometry
144 (ICP-MS; NexIONTM300X, PerkinElmer). Blank filter, reagent blank, replicates, and standard reference material (NIST SRM
145 1648a, urban dust) were adopted for analytical quality control, with recoveries ranged 90-110 %. Carbonaceous species (OC
146 and EC) in PM_{2.5} were determined using a DRI-2001A OC/EC (Atmoslytic Inc., Calabasas, CA, USA). For the concentrations
147 of water-soluble ions (WSIs), the main cations (Na⁺, K⁺, Mg²⁺, Ca²⁺, NH₄⁺) and anions (NO₃⁻, SO₄²⁻, Cl⁻, F⁻) in PM_{2.5} were
148 measured by ion chromatography (IC, Thermo Fisher Scientific, USA), using the Metrosep C6-150/4.0 column for cations and
149 the Metrosep A Supp 5 150/4.0 column for anions, respectively.

150 **2.3 Preparing mass-normalized PM_{2.5} suspension for cell exposure**

151 Totally 30 source and 16 ambient PM_{2.5} samples were also performed cytotoxicity tests. In order to elute the particles
152 completely from the quartz membranes, a whole PM_{2.5}-loaded sample filter was cut into small pieces, immersed in ultrapure
153 water and extracted six times (30 min for each) in an ultrasonic bath at 0 °C. Although the ultrasonication might impact the
154 ROS (Miljevic et al., 2014), the inevitable systematical error was ignored in this study. The extract was then suction filtered
155 through a 2.6 μm pore-size nylon membrane to remove possible quartz fragments, and the bulk filtrate was freeze-dried back
156 to pure PM_{2.5} powder. Ultimately, based on particle mass, the gathered PM_{2.5} was dispersed by sterile phosphate-buffered
157 saline (PBS) to a concentration of 400 mg L⁻¹, and then diluted to PM_{2.5} suspension of 80 mg L⁻¹ with serum-free Dulbecco's
158 modified eagle medium (DMEM) for following *in vitro* cell exposure (Li et al., 2022).

159 **2.4 Cell culture and cellular toxicity tests by *in vitro* PM_{2.5} exposure**

160 Aerosol pollution can harm lung alveoli and epithelial cells, and the A549 adenocarcinoma epithelial cell has long been used
161 as a suitable epithelial alveolar model (Park et al., 2018; Li et al., 2022b). The A549 cells were cultured in RPMI-1640 medium
162 (Gibco, USA) supplemented with 10% fetal bovine serum (FBS, Hyclone, USA) and 1% antibiotic penicillin-streptomycin
163 (100 U mL^{-1}) at $37 \text{ }^\circ\text{C}$ in a 5% CO_2 incubator. After $\text{PM}_{2.5}$ exposure, cell viability and the indicators reflecting oxidative
164 damage and inflammatory responses were determined respectively. While the cell viability assay was helpful in determining
165 $\text{PM}_{2.5}$ dose to cells, the endogenous ROS measurements revealed the status of cellular oxidative potential after $\text{PM}_{2.5}$ exposure
166 followed by the relative effects of ROS on various stages of cellular toxicity like inflammatory responses (Gali et al., 2019).
167 The cell viability (metabolic activity) was evaluated by mitochondrial activity and determined by the methyl-thiazol-
168 tetrazolium (MTT) assay (Chen et al., 2019). After trypsin action, the density of cells in the logarithmic growth phase was
169 adjusted to $1 \times 10^5 \text{ mL}^{-1}$. Cell suspensions were inoculated into 96-well plates (Costar, USA) at $100 \text{ }\mu\text{L}$ per well. The blank
170 control well (without medium and $\text{PM}_{2.5}$ suspension) and reagent control well (with medium but without $\text{PM}_{2.5}$ suspension)
171 were set together. After incubation for 24 h and removing the cellular supernatant, various types of $\text{PM}_{2.5}$ suspension
172 (concentration of 80 mg L^{-1}) were added to 96-well plates and incubated for 24 h. Based on pre-experiments, the oxidative
173 stress and inflammation response sensitively under this dose, while the cell viability can keep sufficient. Fresh medium and
174 MTT reagent (Solarbio, Beijing, CN) were added to each well and the supernatant was discarded, then $100 \text{ }\mu\text{L}$ of formazan
175 lysate was added to each well. The optical density (OD) values were measured at 490 nm using a microplate reader (Thermo
176 MULTISKAN FC, USA). Cell viability (%) = $(\text{OD}_{\text{treatment}} - \text{OD}_{\text{blank control}}) / (\text{OD}_{\text{reagent control}} - \text{OD}_{\text{blank control}})$. The levels of cellular
177 ROS production causing oxidative stress in cells, pro-inflammatory cytokines including tumor necrosis factor-alpha ($\text{TNF-}\alpha$)
178 and interleukin-6 (IL-6) production for determining the expression of genes related to the inflammatory response in the
179 supernatant were analyzed by enzyme-linked immunosorbent assay (ELISA) kits (Jiangsu Enzyme Biotechnology Co., Ltd.,
180 CN), and OD values were measured at 450 nm (Huang et al., 2020; Pang et al., 2020).

181 **2.5 Data analysis**

182 The statistical analysis was performed by IBM SPSS statistics 24 and plotted by Origin 2020b software. Spearman correlation
183 coefficients were produced by the correlation analysis. The variance was statistically significant when the statistical test level
184 was $p < 0.05$, and extremely significant when $p < 0.01$. Statistical analyses were performed using Kruskal-Wallis test (Kruskal
185 and Wallis, 1952).

186 The source apportionment of $\text{PM}_{2.5}$ mass in urban ambient air was conducted by the receptor models PMF (EPA PMF
187 version 5.0) and CMB (EPA CMB 8.0). All measured constituents (OC, EC, Cu, Cr, Co, Ni, As, Pb, Mn, V, Na^+ , K^+ , Mg^{2+} ,
188 Ca^{2+} , NH_4^+ , Cl^- , F^- , NO_3^- , and SO_4^{2-}) were selected as PMF model input data, and a four-factor solution was chosen as the
189 optimal solution based on an assessment of the interpretability of the source profiles and the seasonal variability of the source
190 contributions. Due to the high concentration of sulfate and nitrate in ambient $\text{PM}_{2.5}$, and being lack of specific actual source to
191 emit sulfate and nitrate, we added the virtual source profiles of secondary sources in CMB model (Table S4). The virtual source

192 profiles of secondary sources are represented by the proportion of sulfate, nitrate and ammonium in pure ammonium sulfate
193 and ammonium nitrate.

194

195 **3 Results**

196 **3.1 Contributions of combustion primary sources to urban ambient air PM_{2.5}**

197 As shown in Figure S3, although have been significantly improved with the national air quality in recent years, the estimated
198 annual PM_{2.5} concentrations of representative city Nanjing ($59.1 \pm 20.5 \mu\text{g m}^{-3}$) was 1.7 times higher than the China national
199 standard ($35 \mu\text{g m}^{-3}$) and 11.8 times higher than the WHO guidelines ($5 \mu\text{g m}^{-3}$). Urban air PM_{2.5} pollution levels in the cold
200 season were higher than the warm season. The similar source apportionment results from PMF and CMB models are illustrated
201 in Figure 1. Four major sources of the ambient PM_{2.5} were produced by the PMF model (Figure S4), including secondary
202 aerosols, and primary particles of automobile exhaust, coal combustion, and plant biomass burning, which account for 34.0%,
203 27.7%, 25.2%, and 13.1% of total PM_{2.5} mass concentration, respectively. The CMB model source profiles are shown in the
204 Table S4, and we normalized these source contribution of secondary aerosols (32.4%), and automobile exhaust (32.2%), coal
205 combustion (25.1%), plant biomass burning (10.3%). Therefore, although the contribution of secondary aerosols cannot be
206 ignored, the main anthropogenic sources of urban air PM_{2.5} were primary emissions from the various fuel combustions.

207 **3.2 Chemical compositions of different PM_{2.5} from 30 combustion sources and from representative urban ambient air**

208 Typical chemical components including carbonaceous fractions, heavy metals, and WSIs of all PM_{2.5} samples from both
209 ambient air and combustion sources were analyzed and compared with each other.

210 According to the comparisons of PM_{2.5} bound carbonaceous fractions (Figure 2), automobile and biomass sourced PM_{2.5}
211 contained significantly higher total carbon (TC) content than coal combustion and ambient air, while the OC/EC ratio trend
212 was ambient air > coal combustion > biomass burning > automobile exhaust sources. It indicated that the carbon content of
213 ambient PM_{2.5} mixture was lower and dominated by OC than that of combustion primary sources, implying that the OC in
214 ambient air may be aged or cleaned. The OC undergoes various chemical reactions in the atmosphere, such as oxidation by
215 ozone and hydroxyl radicals, resulting in degradation. Figures S4-S7 showed the detailed carbon fraction characteristics
216 (contents and ratio) of PM_{2.5} from each specific source. Carbonaceous fractions in automobile exhaust PM_{2.5} were high but the
217 difference between OC and EC content was small. Depending on the diverse automobile fuels, loads and tailpipe emission
218 standards, the concentrations of carbon fractions in exhaust PM_{2.5} varied widely with vehicle categories. The carbonaceous
219 portion of PM_{2.5} gradually declines as emission regulations rise, and EC likewise declines dramatically (Figure S5). However,
220 such differences among coal types were less, except the bituminous coal with extreme high OC (Figure S6). The carbonaceous
221 fraction of PM_{2.5} from domestic plant biomass burning differed in raw material species that tree branches source PM_{2.5}
222 generally contained higher carbon contents than those from crop straws (Figure S7).

223 Based on the grouped (Figure 3) and individual (Figures S9-S12) distributions of the measured heavy metals in various
224 PM_{2.5}, the V concentrations of combustion sources were generally higher while Co and Mn were lower than ambient urban air.
225 Coal combustion emissions carried highest levels of Pb and were enriched in Cu and As (Figure S10), while biomass burning
226 were rich in Cr and Ni (Figure S11). However, automobile exhausts were enriched in most heavy metals, especially Cu, and
227 Cr, Ni, V, Mn (Figure S9). Heavy metals from different types of automobile exhausts with the same emission standard varies
228 greatly. Anthracite and industrial coal combustions contain similar heavy metals much more than bituminous coal. Generally,
229 Pb, V, Mn, As, and Cu in branches source PM_{2.5} were higher than straws, while Cr, Ni, and Co were dominant and higher in
230 straw burning emissions. A special discovery was that corn cob burning PM_{2.5} carried more heavy metals than corn straw and
231 was the biomass with the highest emission levels of heavy metals. Correspondingly, ambient air PM_{2.5} were also rich in most
232 metals, especially Mn, Pb, and Ni, Cu, Cr. Therefore, coal combustion sources might contribute most Pb to urban ambient air,
233 and contribute significant Cu and As with automobile exhaust emissions, while plant biomass burning and automobile sources
234 contribute the Cr and Ni. Besides natural dust, automobile exhaust should be the main anthropogenic source of airborne Mn.
235 Considering the PMF source apportionments of ambient aerosols, automobile exhaust should be the main source of Cr in urban
236 air PM_{2.5}, and also the source for Cu together with coal combustion.

237 According to the comparisons of water-soluble cation and anion concentrations in various PM_{2.5} (Figure 4), coal
238 combustions contained highest SO₄²⁻ and NH₄⁺, automobile exhausts had highest contents of NO₃⁻, Na⁺ and Ca²⁺, while plant
239 biomass burning sources contained highest K⁺ and Cl⁻, but Mg²⁺ was the lowest for all sources. However, the urban ambient
240 air PM_{2.5} contained highest NO₃⁻ and were also dominated by SO₄²⁻ and NH₄⁺, for which NO₃⁻ should be mainly contributed
241 by secondary aerosols and automobile primary source, SO₄²⁻ and NH₄⁺ should be significantly from coal combustions. Besides
242 NO₃⁻, Na⁺ and Ca²⁺, automobile source PM_{2.5} also had the highest F⁻ and Mg²⁺ concentrations than other sources. The detailed
243 concentration distributions of WSIs in PM_{2.5} from each specific source were provided in Figures S12-S14. The WSIs levels
244 vary widely with specific source categories. PM_{2.5} from LDDVs-2 had the lowest amount of WSIs compared to the other
245 automobile exhausts (Figure S13). Similar to the metal composition, bituminous coal also had the lowest WSIs among all coals
246 (Figure S14). Compared to branches, PM_{2.5} from burning crop straws had much greater levels of K⁺, Cl⁻, SO₄²⁻ and less levels
247 of F⁻, NO₃⁻ (Figure S15).

248 To summarize, the overall concentrations of measured TC, cumulated heavy metals and WSIs in PM_{2.5} from each source
249 type were showed in Figure 5. Among all source emission and environmental receptor samples, the cumulated heavy metals
250 from coal combustion was highest and automobile exhaust was higher than ambient PM_{2.5}, the overall carbon contents from
251 automobile exhaust and biomass burning were both higher than ambient PM_{2.5}, while only the cumulated soluble ions in PM_{2.5}
252 from primary source of coal combustion was equivalent to the ambient aerosols. In a word, chemical compositions of PM_{2.5}
253 distributed much diversely and varied significantly with the specific source types of combustion emissions.

254 3.3 Cell viability, oxidative stress and inflammation levels exposed to various mass-normalized PM_{2.5}

255 Multiple toxicological endpoints (cell viability, oxidative stress, and inflammation) that facilitate identifying the specific
256 particle triggering ROS and inflammatory responses resulting in cell death were evaluated for source-specific PM_{2.5}. After 24
257 h exposure to the same dose of different PM_{2.5} obtained from specific emission sources, the A549 lung cells also showed varied
258 toxicological responses (Figure 6). The survival rate of cells exposed to automobile exhaust PM_{2.5} was much lower than
259 ambient air PM_{2.5} by 16.6% (Figure 6a). Automobile exhaust PM_{2.5} induced the highest ROS production in cells higher than
260 biomass burning, which was 26.4% and 14.8% higher than ambient PM_{2.5} (Figure 6b). Coal combustion induced the highest
261 cellular IL-6 production followed by automobile exhaust, which was 13.1% and 4.48% higher than ambient air PM_{2.5},
262 respectively; while the PM_{2.5} from automobile exhaust and biomass burning induced similarly 10.4% higher cellular production
263 of TNF- α than ambient PM_{2.5} (Figure 6c, 6d). These results suggested that, combustion primary emission PM_{2.5} had stronger
264 ability to induce oxidative stress and inflammatory injury in lung cells than ambient air PM_{2.5}, thus resulted in the higher
265 probability of apoptosis induction (Victor and Gottlieb, 2002; Wang et al., 2013). Generally, the mass-normalized PM_{2.5} from
266 primary source of automobile exhaust posed the strongest overall toxicity. Therefore, to protect public health by controlling
267 PM_{2.5} pollution, these anthropogenic combustions were key target sources, especially the most toxic automobile PM_{2.5} should
268 be reduced preferentially.

269 **3.4 Correlations between various PM_{2.5} components and toxicity endpoints**

270 Spearman correlation coefficients between chemical compositions and cellular toxicological response indicators were applied
271 to screen the key components of all PM_{2.5} involved in cell injury (Figure 7). It was found that, the degrees of correlations
272 varied with the toxicological mechanisms of different airborne chemicals. Based on the overall PM_{2.5} samples from various
273 sources, the pro-inflammatory cytokine IL-6 showed significantly strong positive correlations with some heavy metals (As,
274 Pb, V, Cu), while TNF- α and oxidative stress (ROS) had similar significantly positive correlations with aerosol components
275 of carbon fractions (EC, OC) and transition metals (V, Cr, Ni). The TNF- α also showed positive correlation with water soluble
276 Cl⁻ and K⁺, and ROS correlated with F⁻, Ca²⁺ and Mg²⁺.

277

278 **4 Discussion**

279 **4.1 Chemical markers for source apportionments of ambient air PM_{2.5}**

280 Combustion emissions are key anthropogenic sources contributing to urban air PM_{2.5}, through both primary and secondary
281 aerosols, which were 66% and 34% estimated by PMF model, 67.6% and 32.4% by CMB model, respectively (Figure 1).
282 Compared to the PMF results, the proportions of coal combustion and secondary sources in the CMB results show minimal
283 changes, while biomass contributions are slightly underestimated, and there is a slight increase in the proportion attributed to
284 vehicular emissions. The high concentrations of chemical markers are usually used in source analysis, such as ammonium
285 sulfate and nitrate for secondary aerosols which are originated mainly from the gaseous precursors (e.g., NH₃, SO₂ and NO_x)

286 (Mahilang et al., 2021), the EC, Cu, Mn, and Ni for vehicle exhaust (Srivastava et al., 2021), the As, Pb, OC, EC, SO_4^{2-} and
287 relatively low $\text{NO}_3^-/\text{SO}_4^{2-}$ ratios for coal combustion (Dai et al., 2020), soluble K^+ and Cl^- for plant burning (Jain et al., 2020).
288 The detailed chemical species of these specific source emission $\text{PM}_{2.5}$ samples also supported the results. Moreover, low
289 OC/EC ratio of high TC content, high NO_3^- , F^- , Na^+ , Ca^{2+} and Mg^{2+} , V and Mn of automobile exhaust; Pb and As, SO_4^{2-} and
290 NH_4^+ of coal combustion; soluble K^+ and Cl^- , and high OC/EC ratio of high TC for plant biomass burning found in current
291 study (Figures 2-5), could also be corresponding potential aerosol source markers.

292 **4.2 Common $\text{PM}_{2.5}$ components related to specific combustion sources**

293 Generally, the automobile exhaust $\text{PM}_{2.5}$ had high TC content and low OC/EC value with considerable EC content (Figure 2),
294 varying with specific vehicle types (Figure S5-8). The contents of the carbon fractions from diesel vehicles were 2.39 times
295 more than gasoline exhausts (Figure S5), and the OC/EC ratios of diesel exhausts were 37.3% of gasoline vehicles, owing to
296 both considerable contents of EC and OC from diesel vehicle emission $\text{PM}_{2.5}$. Some diesel vehicles showed higher EC
297 emissions with age, so exhaust cleaning devices for them are suggested. In addition, the amounts of OC and EC in exhausts
298 gradually decreased with the strengthened emission standards they met (Wong et al., 2020). In $\text{PM}_{2.5}$ samples obtained from
299 coal combustion (Figure S6), the TC contents of bituminous coals was 3.97, 6.41, and 11.6 times higher than that of honeycomb
300 coals, anthracite coals, and industrial coals, respectively, because bituminous coals contain higher volatile fraction. Emissions
301 of non-methane VOCs increase with the volatile content of the coal (He et al., 2022). The vast majority of organic aerosols
302 from bituminous coal are generated in the ignition and fierce combustion phases, which account for 99.9% of the entire
303 combustion process; while these two phases of anthracite coal generate only 77% of the entire process (Zhou et al., 2016).
304 Moreover, as the volatile matter in the coal decreases, the temperature at which weight loss begins and ends shift to higher
305 values, that may be due to the lower amount of aliphatic chains present. It has been reported that for bituminous maximum
306 weight loss happens in the range 490–600 °C, while in the case of anthracites coals it occurs between 750 and 870°C (De la
307 Puente et al., 1998). Therefore, besides the way of combustion and the use of combustion stoves, the coal quality related to
308 different coal types and origins determine the carbonaceous fractions of the PM emitted by coal combustion (Zhang et al.,
309 2022). In the $\text{PM}_{2.5}$ samples from plant biomass combustion (Figure S7), OC contents were 2.21 times higher than EC contents,
310 except that pine branches contained higher EC and rapeseed straw had considerable contents of EC and OC. The OC in ambient
311 $\text{PM}_{2.5}$ dominated the carbonaceous component (Figure S8), consistent with the North China Plain and Indo-Ganges Plain
312 (Flores et al., 2020; Xu et al., 2019). Combining the TC contents and OC/EC ratios, carbonaceous components in ambient $\text{PM}_{2.5}$
313 mainly originate from semi-volatile organic compounds (SVOCs) (Wang et al., 2018). Previous studies have reported that
314 carbonaceous aerosols are mainly originated from fossil fuel combustion in transportation, coal combustion in power plants
315 and industries, and biomass combustion (Kang et al., 2018; Zhang et al., 2015). Thus, to control ambient carbon aerosol
316 pollution, besides reducing the precursor emissions of secondary organic aerosols (SOA), controlling primary aerosols
317 especially EC from diesel vehicles might be effective measures.

318 Airborne redox-active metals are usually linked with the oxidation stress of PM_{2.5}. Different types of automobiles emitted
319 diverse metal contents (Figure S9). Metal elements in automobile exhaust are primarily contributed by fuels, lubricants, and
320 engine component abrasion. Because Mn is a common antidetonator that delays and prevents the oxidation of hydrocarbons
321 and increases the octane number, which not only increases the thermal efficiency of the engine but also improves the emission
322 performance of the vehicle (Cheung et al., 2010), the Mn content was greater in gasoline vehicle exhausts than in diesel
323 vehicles. Although there are multi-sources of traffic Pb emissions such as fuel combustion and brake wear (Wang et al.,
324 2019;Panko et al., 2019), the automobile exhaust Pb content of gasoline vehicles were greater than diesel vehicles owing to
325 oil combustion. Moreover, for the same vehicle type (LDDVs-1 and 2; HDDVs-1 and 2; SDGCs-1 and 2), the stricter the
326 emission standard required, the lower the exhaust metal contents. The metal contents in the PM_{2.5} of trucks was higher than
327 that of passenger cars (Wu et al., 2016). In the combustion PM_{2.5} of 10 coal types (Figure S10), Pb contents were the highest
328 than other heavy metals, similar to available findings (Zhang et al., 2020). The PM_{2.5} metals from bituminous coal were
329 significantly lower than other coal types, because indicated by the coal quality analysis, bituminous coal has a low ash content
330 which is mainly derived from non-combustible minerals in coal. These findings suggested that coal maturity might be an
331 important factor influencing the metal composition of particulates emitted from coal combustion (Shen et al., 2021;Zhang et
332 al., 2021). Heavy metal contents in biomass burned PM_{2.5} varied much widely with raw plant types (Figure S11), although
333 dominated by Cr and Ni. Different plant species and even different plant parts differ significantly in their ability to uptake and
334 accumulate metals from soil (Zhao et al., 2020). Moreover, because of the high enrichment factors of some metals for crop
335 straws (Zhang et al., 2016;Sun et al., 2019), they also released more Cr, Ni, and Co during burning than fuelwoods. Total metal
336 emissions were highest in corn cob but lowest in peanut straw burning PM_{2.5}. The heavy metals enriched in urban ambient air
337 PM_{2.5} showed slightly seasonal pattern (Figure S12), while contents of V, Co, and As were relatively low and less affected by
338 seasonal changes. Accordingly, supported by the metal profiles of anthropogenic combustion sources and ambient aerosols, to
339 control the environmental airborne heavy metal pollution, key targets might be the Pb, Cu and As from honeycomb, anthracite
340 and industrial coal combustion, Cu from vehicle exhausts and especially V from light duty diesel van with the CN.III emission
341 standard and Mn from gasoline vehicles, Cr and Ni from biomass especially crop straws burning.

342 Epidemiological studies have also shown the mortality closely related to the WSIs such as sulfate and nitrate in aerosols
343 (Ostro et al., 2009;Liang et al., 2022). Among the WSIs contents of various automobile exhaust PM_{2.5} (Figure S13), NO₃⁻ and
344 Ca²⁺ were the most abundant anion and cation, respectively. The high NO₃⁻ in the automobile PM_{2.5} may be due to NO_x
345 production during high-temperature combustion, while the high Ca²⁺ content should be related to additives in automobile fuels
346 and calcium-based lubricants (Hao et al., 2019;Yang et al., 2019). Moreover, the exhaust WSIs decreased with the strengthened
347 automobile emission standards required. Coal combustion PM_{2.5} contained relatively higher SO₄²⁻ and NH₄⁺ concentrations
348 followed by Cl⁻ than other WSIs species (Figure S14). Among various coal types, industrial coals emitted highest SO₄²⁻
349 followed by honeycomb and industrial coal with also high NH₄⁺, but bituminous coals emitted low WSIs which were mainly
350 NO₃⁻, F⁻ and Na⁺, Ca²⁺. The WSIs emission factors of honeycomb coal were generally higher than those of lump coal (Yan et

351 al., 2020). For biomass combustion emissions (Figure S15), Cl^- and K^+ were dominant WSIs in $\text{PM}_{2.5}$ from straw-type fuels
352 (Tao et al., 2016; Sillapapiromsuk et al., 2013), but fuelwood-type combustion emitted high NO_3^- . Plant species absolutely
353 determine the emissions (Liao et al., 2021). Finally, there were also high levels of NO_3^- , SO_4^{2-} , and NH_4^+ in ambient air $\text{PM}_{2.5}$
354 (Zhang et al., 2019) (Figure S16), even higher than the investigated combustion sources, so other sources like the secondary
355 aerosols may also contribute. Consequently, target combustion primary aerosols WSIs might include, the NO_3^- from vehicle
356 exhausts and fuelwood burning; SO_4^{2-} and NH_4^+ from honeycomb, anthracite and industrial coal combustion; Cl^- and K^+ from
357 biomass especially crop straw burning.

358 **4.3 $\text{PM}_{2.5}$ toxicity related to specific sources by pivotal chemical components**

359 The complexity of the sources and compositions of atmospheric $\text{PM}_{2.5}$ leads to different toxicological effects (Newman et al.,
360 2020; Kelly, 2021). The toxicological effects of $\text{PM}_{2.5}$ are not comparable among different studies owing to distinct exposure
361 concentrations, biological models, endpoints, and $\text{PM}_{2.5}$ generation methods (Park et al., 2018; Kelly and Fussell, 2020). In
362 this study, we employed same exposure conditions and biological endpoints, in order to obtain comparable toxicity data for
363 $\text{PM}_{2.5}$ from different sources. Our mass-normalized results demonstrated that automobile exhaust $\text{PM}_{2.5}$ induced the highest
364 lethality and cellular ROS and TNF- α production, coal combustion $\text{PM}_{2.5}$ induced the highest cellular IL-6 production, plant
365 biomass burning $\text{PM}_{2.5}$ induced considerable cellular TNF- α and ROS production (Figure 6). Generally, various toxicities of
366 combustion emission primary $\text{PM}_{2.5}$ were much greater than the urban ambient air $\text{PM}_{2.5}$ (Figure 6), owing to the higher
367 concentrations of specific toxic components in $\text{PM}_{2.5}$ from these sources. The supplementary information had included
368 exhaustive cytotoxicity indicators from each individual source (Figure S17-S20). While the survival rate of cell exposed to
369 CN.III emission standard $\text{PM}_{2.5}$ was the lowest and the capacity to induce cells to produce ROS was the highest for CN.IV,
370 automobile exhaust had a similar potential to cause cells to produce inflammatory cytokines (Figure S17). The capability to
371 induce IL-6 production in cells was highest for industrial coal $\text{PM}_{2.5}$, whereas bituminous coal had the highest survival rate of
372 cells and TNF- α induction capacity (Figure S18). From the Figure S19 we can see that the $\text{PM}_{2.5}$ cytotoxicity of straws and
373 branches burning was analogous, but it should be noted that the cell viability of various straw $\text{PM}_{2.5}$ differs significantly, that
374 may be related to the raw fuel characteristics.

375 These possible mechanisms were implied by the overall relationships between the measured chemical components with
376 cytotoxicity indicators of $\text{PM}_{2.5}$ from various specific sources (Figure 7). In general, both TNF- α and ROS were significantly
377 positively correlated with carbonaceous fractions and redox-active transition metals (V, Cr, Ni), which were main contributors
378 of automobile exhausts and biomass burning. The IL-6 was significantly positively correlated with some heavy metals (As and
379 Pb, V and Cu), which were main contributors of coal combustion sources. Potential mechanisms include that, carbon fractions
380 bound in $\text{PM}_{2.5}$ could be transformed into reactive metabolites and then induce ROS production in cells (Stevanovic et al.,
381 2019), and the $\text{PM}_{2.5}$ bound transition metals could also induce ROS production through the Fenton reaction and disrupt the
382 function of enzymes in cells (Verma et al., 2010; Sørensen et al., 2005). Oxidative stress can lead to inflammatory infiltration

383 of neutrophils and stimulate immune cells to produce inflammatory cytokines, among which TNF- α and IL-6 play important
384 roles in the inflammation development (Xu et al., 2020). Ultimately, excessive production of ROS leads to dysfunctional
385 endoplasmic reticulum responses and dysfunctional lipid metabolism in ROS bursts can result in cell membrane damage and
386 even cell death (Piao et al., 2018;Zhao et al., 2004). There have been some related supporting reports. For instance, the OC
387 and EC were significantly associated with biological responses of PM from vehicle emissions collected in tunnels (Niu et al.,
388 2020). The polar or quinone fractions of PAHs in diesel engine exhaust particles significantly contributed to the heightened
389 toxic response (Xia et al., 2004). The PM_{2.5} generated from biomass burning contained a substantial concentration of
390 carbonaceous components. In addition, Cr and Ni in PM₁₀ from straws were highly associated with ROS (Li et al., 2023). In
391 current study, cellular ROS was also correlated with water soluble Ca²⁺, F⁻, and Mg²⁺, which were main contributors of
392 automobile exhaust PM_{2.5}. The Ca²⁺ controls the membrane potential and regulates mitochondrial adenosine triphosphate (ATP)
393 production, and excessive Ca²⁺ leads to energy loss and more ROS production (Madreiter-Sokolowski et al., 2020). Moreover,
394 the TNF- α was also positively correlated with water soluble Cl⁻ and K⁺, which were main contributors of plant burning PM_{2.5}.
395 Therefore, the accumulations of some organic matters with high carbonaceous content (OC, EC) in PM_{2.5} typically from
396 automobile exhausts and plant biomass burning, redox-active metals (V, Cr, Ni) and water-soluble anions (Cl⁻, F⁻) and cations
397 (Ca²⁺, Mg²⁺) contributed by various combustions, might induce ROS production in cells, cause cellular damage through
398 oxidative stress and inflammatory responses, impair cell viability and finally harm human health.

399 Considering the multi-endpoints measured and the PM_{2.5} toxicity mechanisms mentioned above, based on the cell viability
400 first, and then ROS followed by inflammatory markers, together with the significantly related toxic chemical composition
401 contents (Park et al., 2018), we put forward a general sequence of overall mass-normalized toxicity for these combustion
402 source PM_{2.5} to managers. To improve the urban environmental air quality for better public health benefits by controlling
403 aerosols pollution, considering the differential toxicity intensity of each chemical component and their contributions from
404 various sources to ambient aerosols, preferential targets of specific primary PM_{2.5} sources and bound pollutants from
405 anthropogenic combustions are suggested as following sequence: reducing the automobile exhaust PM_{2.5} containing high
406 contents of EC, transition metals (V, Cu, Ni, Cr), and ions (Ca²⁺, Mg²⁺, F⁻, Na⁺) from diesel exhausts by strengthening the
407 emission standards and accelerating the phasing out of highly polluting vehicles; then lessening the coal combustion rich in
408 heavy metals (As, Pb, Cu) by replacement with low-ash clean coals; and depressing the biomass burning containing high OC,
409 Ni, Cr, Cl⁻ and K⁺ from rural crop straw emissions and promoting domestic cleaner energy such as natural gas.

410 **4.4 Limitations and perspectives**

411 In current study, we selected A549 cell based on previous abundant experimental experiences and also because it has been
412 used popularly in *in vitro* toxicology studies to elucidate the cellular and molecular mechanisms of PM involved in lung for
413 many decades (Li et al., 2022b). However, recently the human normal bronchial epithelial cell BEAS-2B was preferred over
414 the human lung adenocarcinoma epithelial cell A549. For instance, both cells were used in an aerosol study (Bonetta et al.,

415 2017), results of which highlighted the higher sensitivity of BEAS-2B cells respect to A549 also in samples with low level of
416 pollutants, because the PM_{0.5} samples from Italian towns can induce genotoxicity in normal cells while cancer cells might be
417 resistant to their adverse effects. Therefore, although our results are reasonable under the same exposure conditions, there were
418 still potential limitations of A549 cells since they may be more resistant to exposure to external compounds, and the generally
419 more sensitive BEAS-2B cells are suggested for future studies.

420 In toxicity assessments, cell vitality reflects the overall health of cells, encompassing factors such as cell membrane integrity,
421 intracellular metabolic activity, and cell proliferation capacity. Decreased cellular vitality may be associated with cell damage,
422 toxic effects, or cellular apoptosis. Inflammation markers are employed to assess the extent and nature of inflammatory
423 reactions, including the production of cytokines and inflammatory mediators, as well as the activation status of inflammatory
424 cells. Inflammation is a complex physiological response, typically delineated by the immune and inflammatory reactions of
425 the body to stimuli such as injury or infection. Alterations in inflammation markers can indicate the intensity and nature of the
426 inflammatory response. In this study, multiple biological responses of epithelial cells to various PM_{2.5} were evaluated,
427 including that, cell viability evaluated the mitochondrial dehydrogenase activity of the living cells, excessive intracellular ROS
428 formation induced by PM_{2.5} was responsible for oxidative stress to the cells, cytokines IL-6 and TNF- α were determined for
429 the effect of PM_{2.5} on pro-inflammatory response in cells. In general, in vitro data can be used to rank various types of particles
430 in terms of the toxic potential including possible carcinogenicity. Each marker will help to understand the hazard and toxicity
431 of PM_{2.5}. However, the toxicity of PM_{2.5} may be the result of multiple components acting through disparate physiological
432 mechanisms, with inconsistent relationships among endpoints (Park et al., 2018). For instance, in BEAS-2B cells, oxidative
433 stress generated by H₂O₂ exposure often results in cytotoxicity rather than by stimulating cytokine/chemokine responses,
434 sometimes no correlation between oxidative damage and cytokine/chemokine responses. Moreover, TNF- α gene was not
435 detected in BEAS-2B cells exposed to atmospheric PM collected from Benin, but the gene expression of other inflammatory
436 cytokines (IL-1 β , IL-6, and IL-8) were significantly induced, and decreasing cell viability was highly correlated with high
437 secretion of all studied cytokines (Cachon et al., 2014). Therefore, in the present study, it was impossible to analyze all
438 chemicals in PM_{2.5} and determine all related toxicological endpoints, so unmeasured chemicals and endpoints might also play
439 roles in the incongruous or unexplained results, and we also can't over-explain the mechanisms just based on statistical
440 relations. To overcome these hurdles, standardization of toxicological studies (experimental methodologies) and reporting
441 guidelines are necessary for tracking and comparing results.

442 This study ranked the unequal "toxic effects" based on the same mass concentration of PM_{2.5} exposure in body lung fluid
443 system, while the "health risks" usually relating to the inhalation exposure concentration of PM_{2.5} in ambient air were not
444 calculated and evaluated quantitatively. Moreover, non-linear concentration-response functions for various endpoints and
445 different exposure concentrations might also limit using toxicological data straightforwardly to predict health effects
446 (morbidity, mortality) in human populations, so drawing conclusions precisely quantifying/ranking the health risks of PM_{2.5}
447 from specific sources or of individual PM_{2.5} components is still not an easy task (Kelly and Fussell, 2020). Therefore, coupled

448 with source apportionment and exposure level of ambient aerosols pollution, toxicology combined with epidemiology studies
449 linking these factors and indicating scientific mechanisms would help to reach conclusions.

450 Moreover, the exact effective measures to control these specific key toxic components from the emissions of various
451 combustion sources indeed a challenge, but still need to be explored. The findings of this research provide a specific direction
452 for better air pollution control and public health. Besides the environmental technological methods of controlling toxic
453 components targeting source materials, combustion processes, and final emissions, the environmental management policies
454 are also beneficial to such aims, like the choice of fuel types, especially for the management of domestic biomass fuel burning.
455 For examples, potential solutions include promoting new green energy vehicles and low-ash clean coals, depressing the diesel
456 exhaust and rural crop straw burning emissions.

457

458 **5 Conclusions**

459 In current study, we found that 2/3 mass of urban ambient air PM_{2.5} in a typical megacity of eastern China originated from
460 primary sources of anthropogenic combustions including coal, automobile, and biomass. Because of the significant differences
461 in the chemical compositions, the diverse PM_{2.5} from both mixed ambient air and directly from individual combustion sources
462 showed much differential mass-normalized *in vitro* toxicity to the human lung epithelial cells, either for the environmental
463 aerosol samples collected from different seasons, or for the primary emissions of PM_{2.5} from various specific source types.
464 According to the comparative study and correlation analysis, the carbonaceous fractions (OC, EC) and redox-active heavy
465 metals (V, Ni, Cr) assisted by water-soluble ions (Ca²⁺, Mg²⁺, F⁻, Cl⁻) might play important roles in inducing cellular ROS
466 production, causing oxidative stress and inflammation, resulting in cell injury and apoptosis, thus damage human health. These
467 toxic pollutants accumulated in specific-source PM_{2.5} varied by the emission types and raw fuel properties. Combined with
468 chemical composition and general cytotoxicity rank, the preferential controlling targets of specific combustion sources might
469 be automobile exhaust (diesel vehicles with emission standards inferior to CN.IV), coal combustion (high ash and high sulfur
470 coals), and rural plant biomass burning (crop straws). Although showing the synthetic effects of mixed compositions and
471 complex sources, besides preventing the secondary aerosols from combustions, preferentially targeted reductions of toxic
472 PM_{2.5} direct emissions from these primary sources, would produce great benefits for public health with improved ambient air
473 quality. Overall, the chemical findings of our toxicological research could help to support the precise, oriented, effective,
474 efficient, and economical composition-source-based strategies for urban aerosols pollution control. However, as a prospect,
475 the detailed mechanisms for unequal toxicity of PM with complicated components from various sources and their quantitative
476 contributions to the health effects of ambient air PM_{2.5} mixture still need in-depth study.

477 **Supplementary materials**

478 There are 20 figures (Figure S1-S20) and 3 tables (Table S1-S4) in the Supporting Information.

479 **Data availability**

480 All raw data can be provided by the corresponding authors upon request.

481 **Author contributions**

482 XSL conceived and supervised the study; WH, YP, MT, HL, and ZZ collected the samples; WH, YP, MT, WL, HL, ZZ, GS,
483 and LX analyzed the chemical compositions; WH, YP, and MT performed the toxicity tests; WH, YP, MT, and XSL analyzed
484 the data; WH and XSL wrote the manuscript draft; XSL, WH, GS, and TM reviewed and edited the manuscript.

485 **Competing interests**

486 The authors declare that they have no conflict of interest.

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489

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711 **Captions of figures**

712 **Figure 1.** Source contributions (%) to the urban ambient air PM_{2.5} (models PMF vs CMB).

713 **Figure 2.** Carbon contents (mg kg⁻¹) and ratio in PM_{2.5} from various specific sources (n=10 for each combustion source and
714 n=16 for urban ambient air).

715 **Figure 3.** Heavy metal contents (mg kg⁻¹) in PM_{2.5} from various specific sources (n=10 for each combustion source and n=16
716 for urban ambient air).

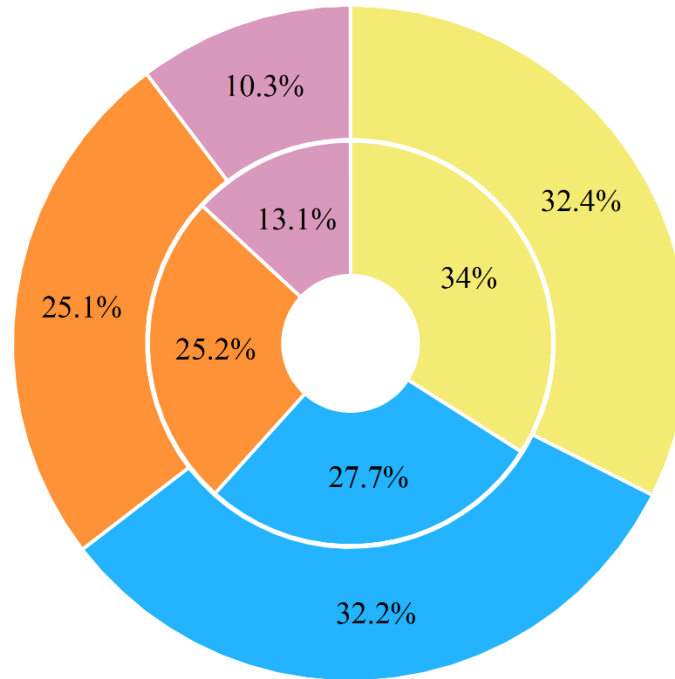
717 **Figure 4.** Water-soluble ion (WSI) contents (mg kg⁻¹) in PM_{2.5} from various specific sources (n=10 for each combustion
718 source and n=16 for urban ambient air).

719 **Figure 5.** Cumulated typical measured components (mg kg⁻¹) in PM_{2.5} from various specific sources (n=10 for each
720 combustion source and n=16 for urban ambient air).

721 **Figure 6.** Cell viability, oxidative stress and inflammation levels of human alveolar epithelial cell lines (A549) exposed to
722 PM_{2.5} suspension (80 mg L⁻¹) from various specific sources (n=10 for each combustion source and n=16 for urban ambient
723 air).

724 **Figure 7.** Overall correlations between typical cellular toxicological responses and chemical compositions of PM_{2.5} from
725 various sources (*p < 0.05, #p < 0.01; n=46).

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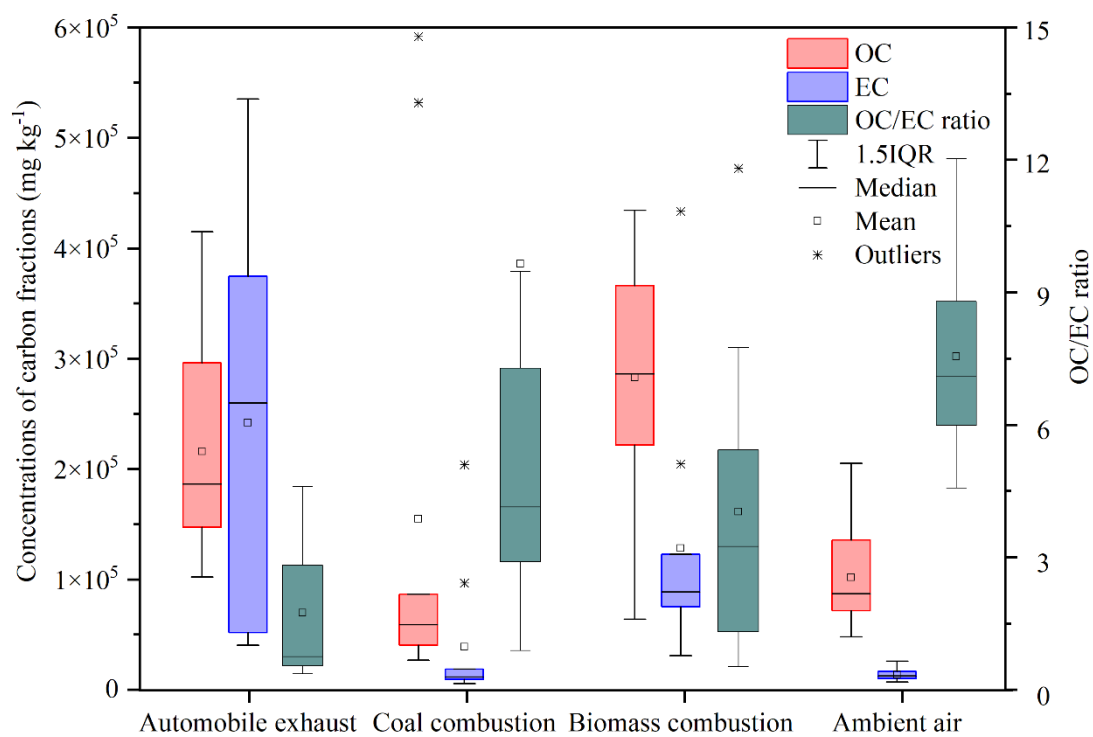
Inner ring - outer ring: PMF model -CMB model

Secondary aerosols Automobile exhaust Coal combustion Biomass burning

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729 **Figure 1.** Source contributions (%) to the urban ambient air PM_{2.5} (models PMF vs CMB).

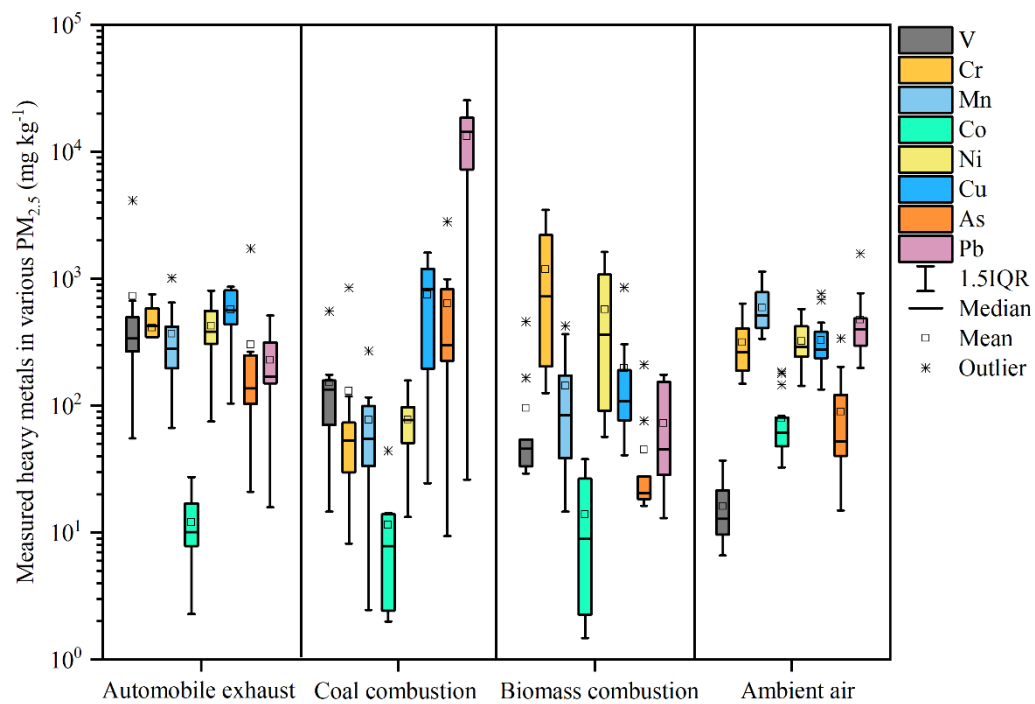
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733 **Figure 2.** Carbon contents (mg kg⁻¹) and ratio in PM_{2.5} from various specific sources (n=10 for each combustion source and
 734 n=16 for urban ambient air).

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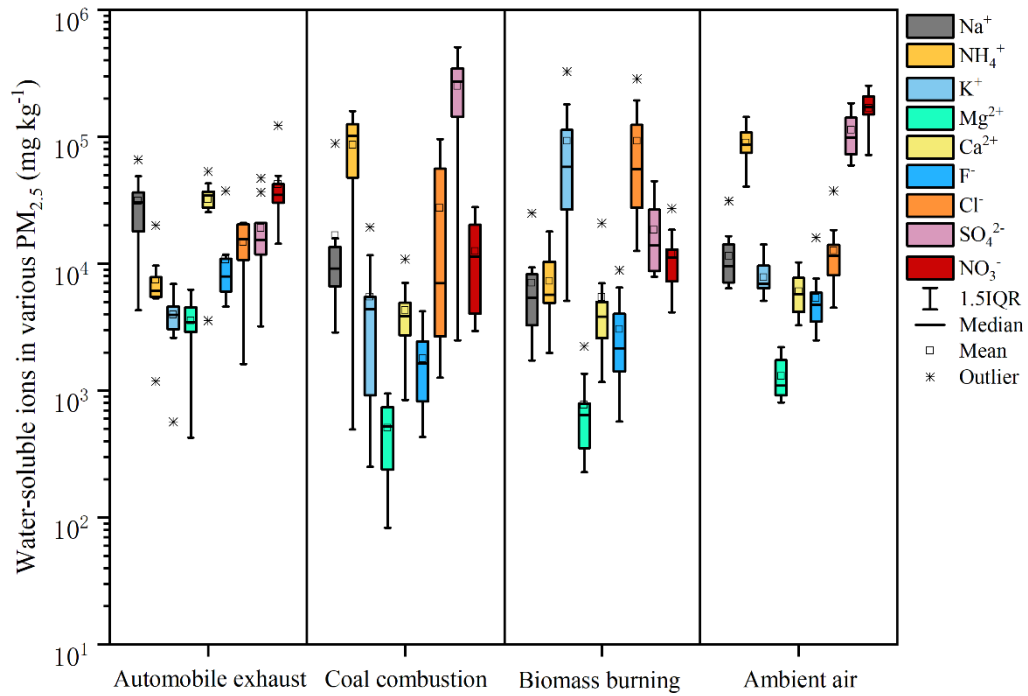


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737 **Figure 3.** Heavy metal contents ($mg\ kg^{-1}$) in $PM_{2.5}$ from various specific sources (n=10 for each combustion source and

738 n=16 for urban ambient air).

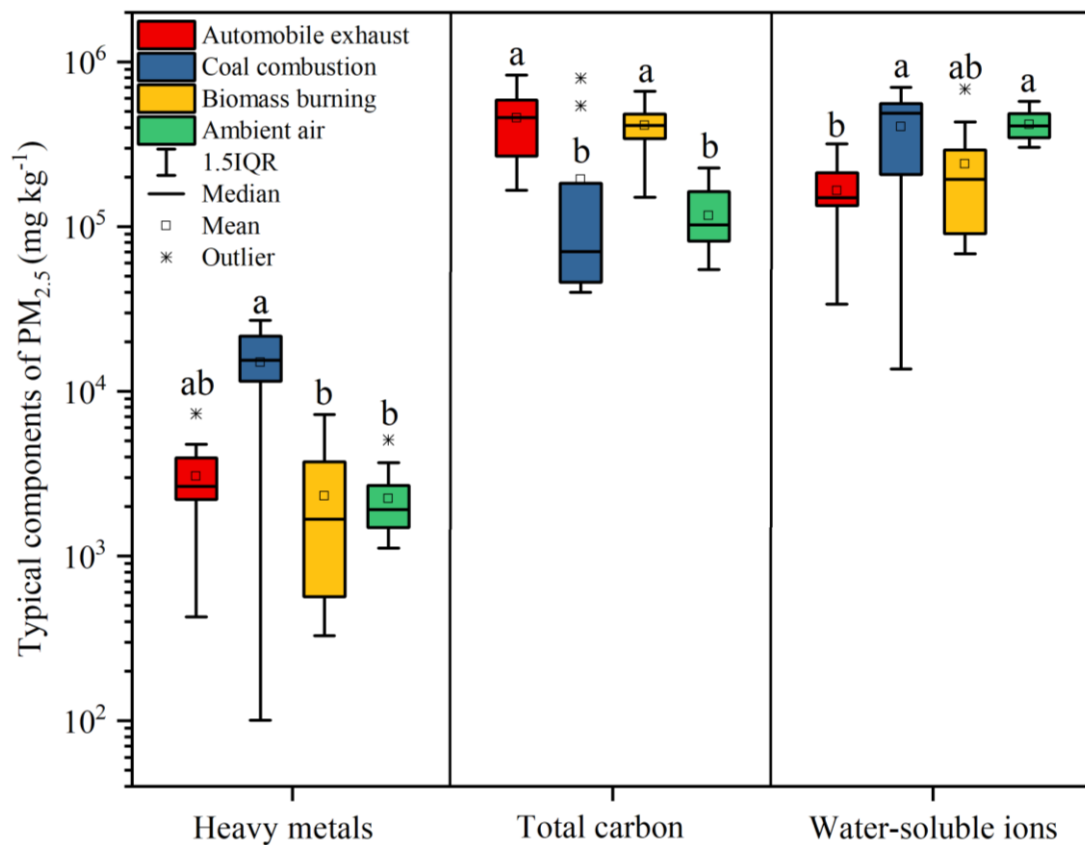
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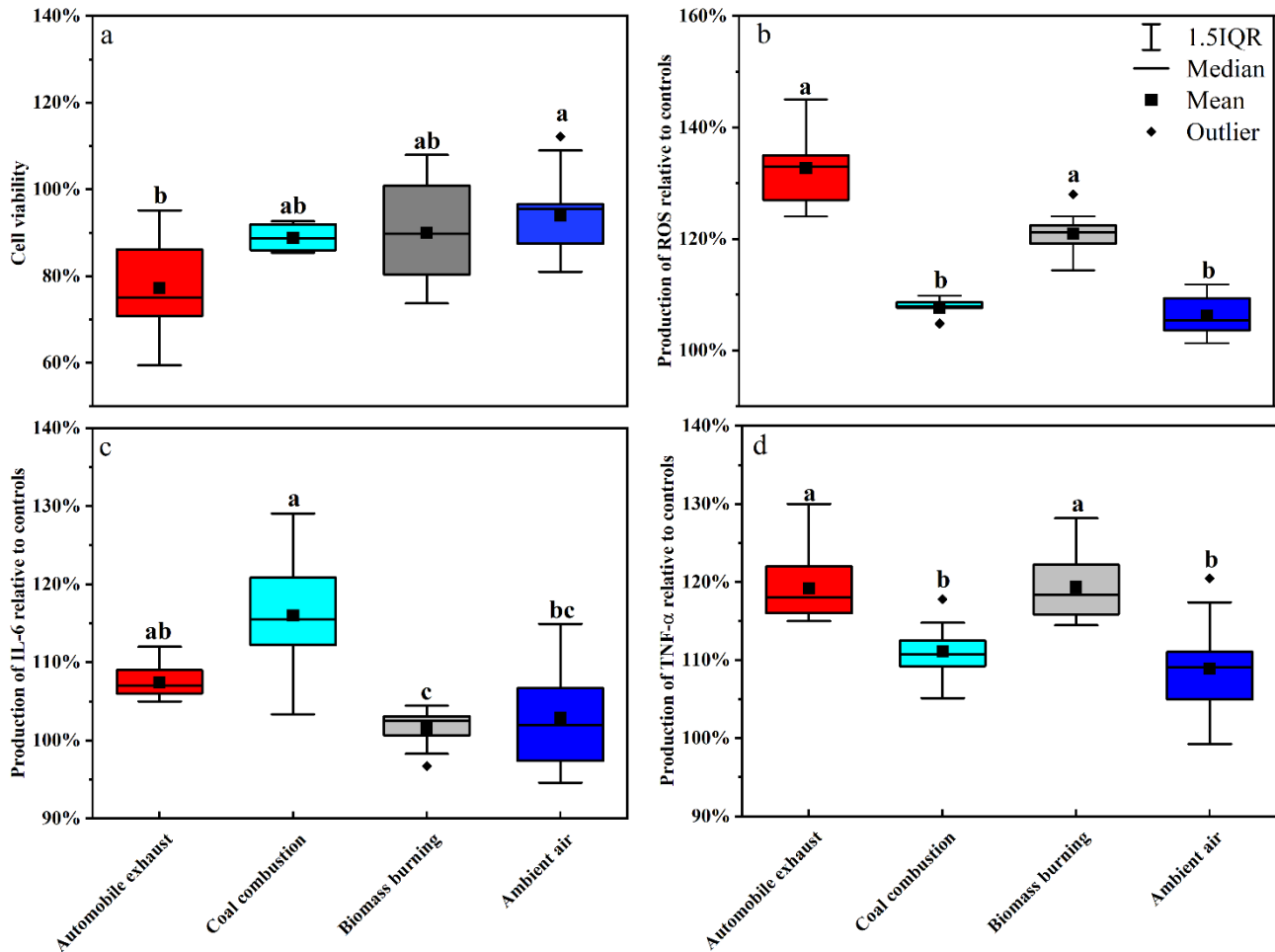
741 **Figure 4.** Water-soluble ion (WSI) contents (mg kg⁻¹) in PM_{2.5} from various specific sources (n=10 for each combustion
 742 source and n=16 for urban ambient air).

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745 **Figure 5.** Cumulated typical measured components (mg kg⁻¹) in PM_{2.5} from various specific sources (n=10 for each
 746 combustion source and n=16 for urban ambient air). Statistically significant differences between the groups are indicated by
 747 different letters (Kruskal-Wallis test, p < 0.05).

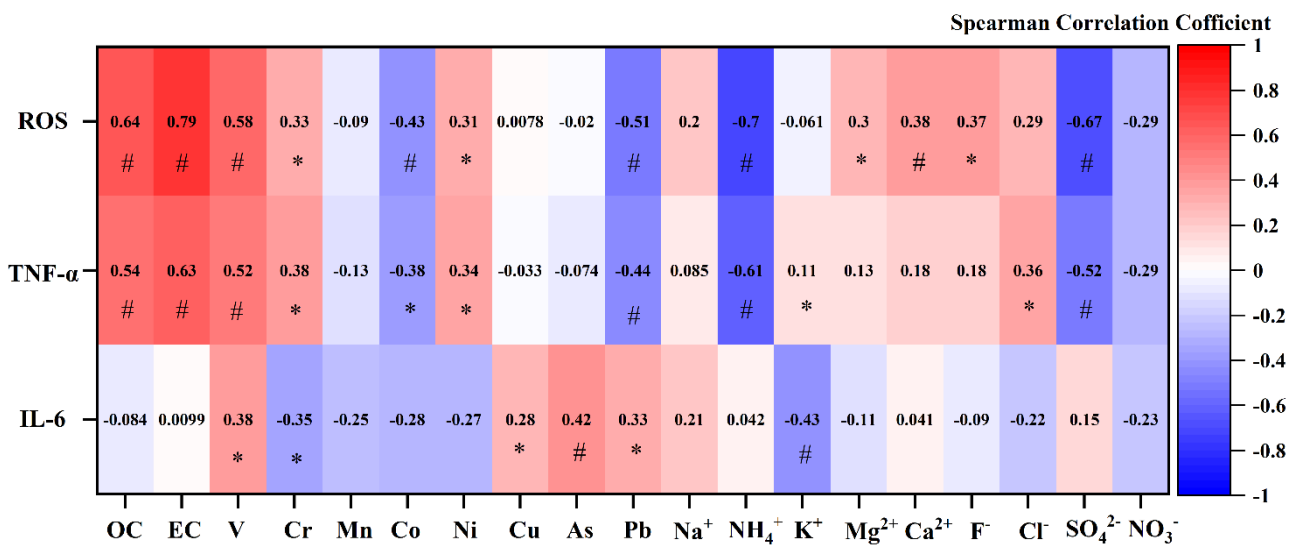


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750 **Figure 6.** Cell viability, oxidative stress and inflammation levels of human alveolar epithelial cell lines (A549) exposed to
 751 PM_{2.5} suspension (80 mg L⁻¹) from various specific sources (n=10 for each combustion source and n=16 for urban ambient
 752 air). Statistically significant differences between the groups are indicated by different letters (Kruskal-Wallis test, p < 0.05).

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756 **Figure 7.** Overall correlations between typical cellular toxicological responses and chemical compositions of PM_{2.5} from
 757 various sources (*p < 0.05, #p<0.01; n=46).

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