

1 **Source differences in the components and cytotoxicity of PM_{2.5} from**
2 **automobile exhaust, coal combustion, and biomass burning**
3 **contributing to urban aerosol toxicity**

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15 **Abstract.** Although air quality guidelines generally use the atmospheric concentration of fine particulate matter (PM_{2.5}) as the
16 metric for air pollution evaluation and management treating all particles as equally toxic, it is inconsistent with the facts that
17 particle toxicity are significantly related to their sources and chemical compositions. Therefore, judging the most harmful
18 source and identifying the toxic component will be extremely helpful to optimize air quality standards and prioritize targeted
19 PM_{2.5} control strategies to more protect public health effectively. The combustions of fuels, including oil, coal, and biomass,
20 are main anthropogenic sources of environmental PM_{2.5}, however, their discrepant contributions to health risks of mixed
21 ambient aerosol pollution dominated by respective emission intensity and unequal toxicity of chemical components are still
22 unclear. In order to quantify the differences among these combustion primary emissions, ten types of PM_{2.5} from each typical
23 source group, i.e., vehicle exhaust, coal combustion, and plant biomass burning, were collected for comparative study with
24 toxicological mechanisms. Totally thirty type individual combustion samples were inter-compared with representative urban
25 ambient air PM_{2.5} samples, which chemical characteristics and biological effects were investigated by component analysis
26 (carbon, metals, soluble ions) and *in vitro* toxicity assays (cell viability, oxidative stress, inflammatory responses) of human
27 lung adenocarcinoma epithelial cells (A549). Carbonaceous fractions were plenteous in automobile exhaust and biomass
28 burning, while heavy metals were more plentiful in PM_{2.5} from coal combustion and automobile exhaust. The overall ranking
29 of mass-normalized cytotoxicity for source-specific PM_{2.5} was automobile exhaust > coal combustion > plant biomass burning >
30 ambient urban air, possibly with differential toxicity triggers, that the carbonaceous fractions (organic carbon, OC; elemental
31 carbon, EC) and redox-active transition metals (V, Ni, Cr) assisted by water-soluble ions (Ca²⁺, Mg²⁺, F⁻, Cl⁻) might play

32 important roles in inducing cellular reactive organic species (ROS) production, causing oxidative stress and inflammation,
33 resulting in cell injury and apoptosis, thus damage human health. Coupled with the source apportionment results of typical
34 urban ambient air PM_{2.5} in eastern China, reducing toxic PM_{2.5} from these anthropogenic combustions will be greatly beneficial
35 to public health, especially preferentially decreasing the diesel exhaust by strengthening emission standards, then lessening the
36 coal combustion by replacement with low-ash clean coals, and depressing the crop straw burning emissions.

37

38 **1 Introduction**

39 As a mixture of multiple sources, ambient particulate matter (PM) arise from anthropogenic activities are continuously
40 deteriorating the urban air quality, particularly in developing countries. Among these, fine PM with an aerodynamic diameter
41 of less than 2.5 μm (PM_{2.5}) is recognized as a serious public health concern due to its long persistence in air, carcinogenicity
42 and acute toxicity to humans (Al-Kindi et al., 2020). There were extensive epidemiological evidences that airborne PM can
43 cause serious negative effects on human health, such as respiratory and cardiovascular diseases, genetic mutations, and
44 developmental disorders (Chowdhury et al., 2022;Lelieveld et al., 2021;Smith, 2021;Clemens et al., 2017). Currently, either
45 the world air quality guidelines or the national air quality standards use the mass concentration of PM_{2.5} as the metric for PM_{2.5}
46 pollution evaluation and management, in which all particles are treated as equally toxic, however, it is inconsistent with the
47 scientific facts that particle toxicity are significantly related to their sources and chemical compositions (Shiraiwa et al., 2017).
48 Therefore, to identify which component(s) and source(s) of ambient PM are most harmful to health, will be very helpful to
49 optimize air quality guidelines/standards and prioritize targeted PM control strategies to more effectively protect public health
50 (Kelly and Fussell, 2020).

51 Besides natural sources like dust and sea spray, the vast majority of aerosols come from anthropogenic activities especially
52 energy consumption, including the combustion of fossil fuels causing industrial emissions and automobile exhaust, and
53 biomass burning (McDuffie et al., 2021;Wu et al., 2022). Finally, these diverse sources make the ambient air PM_{2.5} become a
54 complex mixture with multiple chemical components varying with time and space, which consisting mainly of sulfate, nitrate,
55 ammonium, organic carbon (OC), elemental carbon (EC), mineral and trace metals (Bari and Kindzierski, 2016; Kelly and
56 Fussell, 2020). The physiological mechanisms of PM-induced cell toxicity in respiratory system have been continuously
57 investigated with some progresses (Kelly and Fussell, 2012, 2020; Shiraiwa et al., 2017; Mack et al., 2020; Li et al., 2022b),
58 such as the metabolic activation, oxidative stress, inflammatory response, and apoptosis, focused on by current study. In brief,
59 after inhalation and deposition onto the epithelium, redox-active materials in PM_{2.5} can induce the release of reactive organic
60 species (ROS), which cause oxidative stress (an imbalance between ROS and antioxidants, i.e., disequilibrium of the redox
61 state of a cell) followed by inflammation and cell death. The ROS can mediate subsequent signaling pathways leading to
62 biomolecule damage (e.g., DNA, lipid, and protein) and cellular injury, through mediating inflammatory responses including
63 the release of pro-inflammatory cytokines like IL-6 and TNF- α by epithelial cells (Sabbir Ahmed et al., 2020; Landwehr et al.,

64 2021). For instance, oxidative stress could trigger the induction of pro-inflammatory transcription factors, such as nuclear
65 factor (NF)- κ B, via the mitogen-activated protein kinase (MAPK) signaling pathway. Components adsorbed on particle surface,
66 such as redox-active metals (transition metals, Fe, Ni, V, Cr, Cu), organic compounds (polycyclic aromatic hydrocarbons,
67 PAHs; quinones), or even carbonaceous core of particles, are responsible for oxidative stress (Cachon et al., 2014; Sabbir
68 Ahmed et al., 2020). The non-redox active metals (Zn, Pb, Al) can also influence the toxic effects of transition metals by
69 exacerbating or lessening the production of free radicals. The EC may not be a directly toxic component of PM_{2.5} but rather
70 operate as a universal carrier of combustion-derived chemicals (semi-volatile organic fractions, transition metals) of varying
71 toxicity (Kelly and Fussell, 2020). Inorganic soluble sulphates and nitrates are acidic and can interact with and influence the
72 solubility other compositions like metal bioavailability (Fang et al., 2017; Weber et al., 2016). However, which specific
73 components and which particular sources are the most critical factors dominating the ambient aerosols' health risks, still leave
74 puzzles unsolved.

75 Past studies performed in various countries have focused on physicochemical characterization or biological effects of
76 ambient air PM_{2.5} respectively (Weagle et al., 2018; Jia et al., 2017; Wang et al., 2020). For example, the source analysis of
77 PM_{2.5} by photochemical modelling (Bao et al., 2018), chemical composition of regional PM_{2.5} (Chi et al., 2022), and the
78 mechanism of PM_{2.5} toxicity was independently reported recently (Jia et al., 2020). Because differences in particle composition,
79 sources, and toxicity appear in different urban environments (Zhao et al., 2019; Borlaza et al., 2018), the source profiles of
80 different emission inventories were applied to elucidate aerosol pollution characteristics and control strategies. For instance, it
81 was found that straw burning during the harvest season is a major trigger of severe air pollution in many regions (Sahu et al.,
82 2021). Aerosols from open biomass burning in the Amazon had a stronger ability to induce ROS than laboratory-generated
83 secondary organic aerosols (Tuet et al., 2019). The particle composition of motor vehicle exhaust was related to automobile
84 types with various fuels, engines, and loads (Lin et al., 2020). A strong catalytic reactivity of metals in PM emitted from diesel
85 vehicles was observed by dithiothreitol (DTT) assay (Jesus et al., 2018). Sulfate is a major component of PM from Xi'an city,
86 western China, mainly released from residential coal combustion activities (Dai et al., 2019). Traffic was suggested playing
87 the most crucial role in enhancing the toxicity of fine particles (Park et al., 2018). Although there were emerging studies on
88 particle emission from single source, quantitatively comparative studies on multi-source pollutants as well as the differential
89 composition and unequal toxicity of various sources are still limited.

90 The main objective of current study was to compare the chemical components and corresponding mass-normalized
91 toxicological effects of individual PM_{2.5} from various combustion sources and their unequal contributions to ambient aerosol
92 health risks. The aim is to provide detailed guidance on the targeting and precise control of specific anthropogenic sources
93 with prominent risks based on their pivotal toxic components. Therefore, we collected both representative ambient PM_{2.5}
94 samples from urban air and abundant typical source PM_{2.5} samples from automobile exhaust, coal combustion, and plant
95 biomass burning. Their independent profiles of chemical compositions and *in vitro* cytotoxicity (cell viability, oxidative stress,
96 and inflammatory responses) were investigated and intercompared, to assess the differences in source-to-receptor toxicity and

97 to infer the core toxic components and respective harmful contribution. The pivotal toxic components were identified based
98 on the source-sink bi-directional composition-effect results, which were further used to assess the health toxicity contribution
99 of various emission sources to ambient air PM_{2.5}, supported by its source apportionment through positive matrix factorization
100 (PMF) model. This study could advance the understanding to quantify the complex source contribution to high-risk PM_{2.5}
101 emission oriented to public health, which is imperative for precise prevention and control of atmospheric PM pollution.
102

103 **2 Materials and methods**

104 **2.1 Collection of PM_{2.5} samples from primary emissions of 30 typical combustion sources and from representative** 105 **ambient urban air**

106 As the main anthropogenic sources of the ambient air PM_{2.5} pollution, totally 30 types of primary PM_{2.5} samples emitted
107 directly from automobile exhaust, coal combustion, and plant biomass burning were respectively collected as follows for both
108 chemical and toxicological analyses.

109 Based on the classification of automobile fuel types as well as load and tailpipe emission standards provided by the 2019
110 Annual Report on Environmental Management of Mobile Sources in China, a total of 10 types of vehicles were chosen for
111 exhaust investigation. They were further categorized into 7 sub-groups, including small duty gasoline coaches (SDGCs), small
112 duty diesel coaches (SDDCs), middle duty diesel coaches (MDDCs), heavy duty diesel coaches (HDDCs), light duty diesel
113 vans (LDDVs), middle duty diesel vans (MDDVs), and heavy duty diesel vans (HDDVs). The detailed information of these
114 representative local automobiles was showed in Table S1.

115 To cover all coal types consumed in the city, 10 representative types of coal were gathered for investigation. They were
116 further classified into 4 sub-groups, including 2 types of honeycomb coal (HC), 3 types of anthracite coal (AC), and 2 types
117 of bituminous coal (BC) mainly for restaurant or household use, and 3 types of industrial coal (IC) for coal-fired power plants
118 and steel-smelting industry. The detailed characteristic analysis of these typical coals purchased from local market were
119 showed in Table S2.

120 Considering the plant biomass combustion in rural areas surrounding the megacity, 10 representative types of agricultural
121 and forestry solid wastes were gathered for investigation. Because of the high annual production of three staple food crops
122 (rice, wheat, and corn) as well as soybean, peanut and rapeseed, their straws generated during harvest are often used as fuels
123 in rural households. In addition, woods were also common fuels. Therefore, straws of rice, wheat, corn, soybean, peanut, rape,
124 and sesame, corncob, branches of peach and pine, were selected as plant biomass fuels and further divided into 2 sub-groups,
125 including 8 types of crop straw and 2 types of firewood. The detailed characteristic analysis of these typical plant biomass
126 fuels collected from rural areas around Nanjing city were showed in Table S3.

127 The PM_{2.5} samples directly emitted from these combustion sources were collected by dilution channel sampling method
128 (Figure S1), using a 4-channel particulate matter dilution sampler (HY-805, Hengyuan Technology Development Co., CN).

129 Each sampling included 3 parallel channels of quartz microfiber filter (Figure S2) and 1 channel of Teflon membrane filter
130 with diameters of 47 mm, through a size selector for PM_{2.5} with a flow rate of 160 L min⁻¹. Clean air was pumped for 10 min
131 before and after each sample was collected. Before using, the blank quartz filters were incinerated by a muffle furnace at
132 500 °C for 3 h to remove any possible organic matters, while Teflon filters were baked at 60 °C for 4 h. After being equilibrated
133 in a constant temperature and humidity chamber for 24 h, the filters were weighed both before and after sampling for
134 gravimetric measurements, then the mass of collected PM_{2.5} could be calculated. The sampled filters were stored in a
135 refrigerator at -20 °C before analysis. The quartz filter loaded PM_{2.5} samples were used for carbon and ion analysis, and for
136 toxicity tests, while the parallel Teflon filter loaded samples were used to determine metals.

137 As the actual mixture of various source particles in real environment, totally 16 representative ambient air PM_{2.5} samples
138 (each time lasting 23h) covering a year monthly were collected from December 2019 to October 2020 in an urban site
139 surrounded by traffic, residential and commercial quarters of Nanjing city, Yangtze River Delta of eastern China, using a high-
140 volume air sampler (800 L min⁻¹) with quartz microfiber filters. Detailed procedures and sample information were described
141 in previous paper (Li et al., 2022a), but the purpose of using these air samples in current study was to compare them with the
142 specific source samples for evaluating the chemical and toxicological contributions of the combustion primary sources to
143 environmental aerosols pollution.

144 **2.2 Chemical composition analysis**

145 All collected source and ambient PM_{2.5} samples were conducted various component analysis (Li et al., 2023). For the
146 concentrations of heavy metals in particulates, samples were digested by concentrated HNO₃-HClO₄ acids with a progressive
147 heating program and determined by inductively coupled plasma optical emission spectrometry (ICP-OES; Optima8000,
148 PerkinElmer), with some elements at lower concentrations measured by ICP mass spectrometry (ICP-MS; NexIONTM300X,
149 PerkinElmer). Blank filter, reagent blank, replicates, and standard reference material (NIST SRM 1648a, urban dust) were
150 adopted for analytical quality control, with recoveries ranged 90-110 %. Carbonaceous species (OC and EC) in PM_{2.5} were
151 determined using a DRI-2001A OC/EC (Atmoslytic Inc., Calabasas, CA, USA). For the concentrations of water-soluble ions
152 (WSIs), the main cations (Na⁺, K⁺, Mg²⁺, Ca²⁺, NH₄⁺) and anions (NO₃⁻, SO₄²⁻, Cl⁻, F⁻) in PM_{2.5} were measured by ion
153 chromatography (IC, Thermo Fisher Scientific, USA), using the Metrosep C6-150/4.0 column for cations and the Metrosep A
154 Supp 5 150/4.0 column for anions, respectively.

155 **2.3 Preparing mass-normalized PM_{2.5} suspension for cell exposure**

156 Totally 30 source and 16 ambient PM_{2.5} samples were also performed cytotoxicity tests. In order to elute the particles
157 completely from the quartz membranes, the PM_{2.5}-loaded sample filter was cut into small pieces, immersed in ultrapure water
158 and extracted six times (30 min for each) in an ultrasonic bath at 0 °C. Although the ultrasonication might impact the ROS
159 (Miljevic et al., 2014), the inevitable systematical error was ignored in this study. The extract was then suction filtered through
160 a 2.6 μm pore-size nylon membrane to remove possible quartz fragments, and the bulk filtrate was freeze-dried back to pure

161 PM_{2.5} powder. Ultimately, based on particle mass, the gathered PM_{2.5} was dispersed by sterile phosphate-buffered saline (PBS)
162 to a concentration of 400 mg L⁻¹, and then diluted to PM_{2.5} suspension of 80 mg L⁻¹ with serum-free Dulbecco's modified eagle
163 medium (DMEM) medium for following *in vitro* cell exposure (Li et al., 2022a).

164 **2.4 Cell culture and cellular toxicity tests by *in vitro* PM_{2.5} exposure**

165 Aerosol pollution can harm lung alveoli and epithelial cells, and the A549 human lung adenocarcinoma epithelial cell has long
166 been used as a suitable epithelial alveolar model to investigate the interactions between PM and lung epithelial cells (Park et
167 al., 2018; Li et al., 2022b). The A549 cells were cultured in RPMI-1640 medium (Gibco, USA) supplemented with 10% fetal
168 bovine serum (FBS, Hyclone, USA) and 1% antibiotic penicillin-streptomycin (100 U mL⁻¹) at 37 °C in a 5% CO₂ incubator.
169 After PM_{2.5} exposure, cell viability and the indicators reflecting oxidative damage and inflammatory responses were
170 determined respectively. While the cell viability assay was helpful in determining PM_{2.5} dose to cells, the endogenous ROS
171 measurements revealed the status of cellular oxidative potential after PM_{2.5} exposure followed by the relative effects of ROS
172 on various stages of cellular toxicity like inflammatory responses (Gali et al., 2019). The cell viability (metabolic activity) was
173 evaluated by mitochondrial activity and determined by the methyl-thiazol-tetrazolium (MTT) assay (Chen et al., 2019). After
174 trypsin action, the density of cells in the logarithmic growth phase was adjusted to 1 × 10⁵ mL⁻¹. Cell suspensions were
175 inoculated into 96-well plates (Costar, USA) at 100 μL per well. The blank control well (without medium and PM_{2.5} suspension)
176 and reagent control well (with medium but without PM_{2.5} suspension) were set together. After incubation for 24 h and removing
177 the cellular supernatant, various types of PM_{2.5} suspension (concentration of 80 mg L⁻¹) were added to 96-well plates and
178 incubated for 24 h. Based on pre-experiments, the oxidative stress and inflammation response sensitively under this dose,
179 while the cell viability can keep sufficient. Fresh medium and MTT reagent (Solarbio, Beijing, CN) were added to each well
180 and the supernatant was discarded, then 100 μL of formazan lysate was added to each well. The optical density (OD) values
181 were measured at 490 nm using a microplate reader (Thermo MULTISKAN FC, USA). Cell viability (%) = (OD_{treatment} –
182 OD_{blank control}) / (OD_{reagent control} – OD_{blank control}). The levels of cellular ROS production causing oxidative stress in cells, pro-
183 inflammatory cytokines including tumor necrosis factor-alpha (TNF-α) and interleukin-6 (IL-6) production for determining
184 the expression of genes related to the inflammatory response in the supernatant were analyzed by enzyme-linked
185 immunosorbent assay (ELISA) kits (Jiangsu Enzyme Biotechnology Co., Ltd., CN), and OD values were measured at 450 nm
186 (Huang et al., 2020; Pang et al., 2020).

187 **2.5 Data analysis**

188 The statistical analysis was performed by IBM SPSS statistics 24 and plotted by Origin 2020b software. Spearman correlation
189 coefficients were produced by the correlation analysis. The variance was statistically significant when the statistical test level
190 was p < 0.05, and extremely significant when p < 0.01. Statistical analyses were performed using Kruskal–Wallis test (Kruskal
191 and Wallis, 1952).

192 The source apportionment of PM_{2.5} mass in urban ambient air was conducted by the receptor model PMF (EPA PMF version
193 5.0). Major constituents (OC, EC, Cu, Cr, Co, Ni, As, Pb, Mn, V, Na⁺, K⁺, Mg²⁺, Ca²⁺, NH₄⁺, Cl⁻, F⁻, NO₃⁻, and SO₄²⁻) were
194 selected as input data, and a four-factor solution was chosen as the optimal solution based on an assessment of the
195 interpretability of the source profiles and the seasonal variability of the source contributions.

196

197 **3 Results**

198 **3.1 Contributions of combustion primary sources to urban ambient air PM_{2.5}**

199 As shown in Figure S3, although have been significantly improved with the national air quality in recent years, the daily PM_{2.5}
200 concentrations of representative city Nanjing still exceeded the healthy guidelines obviously, with higher urban air PM_{2.5}
201 pollution level in the cold season²³. Four major sources of the ambient PM_{2.5} were produced by the PMF model, including
202 secondary aerosols, and primary particles of automobile exhaust, coal combustion, and plant biomass burning, which account
203 for 34%, 27.7%, 25.2%, and 13.1% of total PM_{2.5} mass concentration, respectively. Their source profiles and proportions were
204 showed in Figure 1. Therefore, although the contribution of secondary aerosols cannot be ignored, the main anthropogenic
205 sources of urban air PM_{2.5} were primary emissions (66%) from the various fuel combustions.

206 **3.2 Chemical compositions of different PM_{2.5} from 30 combustion sources and from representative urban ambient air**

207 Typical chemical components including carbonaceous fractions, heavy metals and WSIs of all PM_{2.5} samples from both
208 ambient urban air and 30 representative combustion primary sources (covering different categories of automobile exhaust, coal
209 combustion, and plant biomass burning) were analyzed and compared with each other.

210 According to the comparisons of PM_{2.5} bound carbonaceous fractions (Figure 2), automobile and biomass sourced PM_{2.5}
211 contained significantly higher total carbon (TC) content than coal combustion and ambient air, while the OC/EC ratio trend
212 was ambient air > coal combustion > biomass burning > automobile exhaust sources. It indicated that the carbon content of
213 ambient PM_{2.5} mixture was lower and dominated by OC than that of combustion primary sources. Figures S4-S7 showed the
214 detailed carbon fraction characteristics (contents and ratio) of PM_{2.5} from each specific source. Carbonaceous fractions in
215 automobile exhaust PM_{2.5} were high but the difference between OC and EC content was small. Depending on the diverse
216 automobile fuels, loads and tailpipe emission standards, the concentrations of carbon fractions in exhaust PM_{2.5} varied widely
217 with vehicle categories. The carbonaceous portion of PM_{2.5} gradually declines as emission regulations rise, and EC likewise
218 declines dramatically (Figure S4). However, such differences among coal types were less, except the bituminous coal with
219 extreme high OC (Figure S5). The carbonaceous fraction of PM_{2.5} from plant biomass burning differed in raw material species
220 that tree branches source PM_{2.5} generally contained higher carbon contents than those from crop straws (Figure S6).

221 Based on the grouped (Figure 3) and individual (Figures S8-S11) distributions of the measured heavy metals in various
222 PM_{2.5}, the V concentrations of combustion sources were generally higher while Co and Mn were lower than ambient urban air.

223 Coal combustion emissions carried highest levels of Pb and were enriched in Cu and As (Figure S9), while biomass burning
224 were rich in Cr and Ni (Figure S10). However, automobile exhausts were enriched in most heavy metals, especially Cu, and
225 Cr, Ni, V, Mn (Figure S8). Heavy metals from different types of automobile exhausts with the same emission standard varies
226 greatly. Anthracite and industrial coal combustions contain similar heavy metals much more than bituminous coal. Generally,
227 Pb, V, Mn, As, and Cu in branches source PM_{2.5} were higher than straws, while Cr, Ni, and Co were dominant and higher in
228 straw burning emissions. A special discovery was that corn cob burning PM_{2.5} carried more heavy metals than corn straw and
229 was the biomass with the highest emission levels of heavy metals. Correspondingly, ambient air PM_{2.5} were also rich in most
230 metals, especially Mn, Pb, and Ni, Cu, Cr. Therefore, coal combustion sources might contribute most Pb to urban ambient air,
231 and contribute significant Cu and As with automobile exhaust emissions, while plant biomass burning and automobile sources
232 contribute the Cr and Ni. Besides natural dust, automobile exhaust should be the main anthropogenic source of airborne Mn.
233 Considering the PMF source apportionments of ambient aerosols, automobile exhaust should be the main source of Cr in urban
234 air PM_{2.5}, and also the source for Cu together with coal combustion.

235 According to the comparisons of water-soluble cation and anion concentrations in various PM_{2.5} (Figure 4), coal
236 combustions contained highest SO₄²⁻ and NH₄⁺, automobile exhausts had highest contents of NO₃⁻, Na⁺ and Ca²⁺, while plant
237 biomass burning sources contained highest K⁺ and Cl⁻, but Mg²⁺ was the lowest for all sources. However, the urban ambient
238 air PM_{2.5} contained highest NO₃⁻ and were also dominated by SO₄²⁻ and NH₄⁺, for which NO₃⁻ should be mainly contributed
239 by secondary aerosols and automobile primary source, SO₄²⁻ and NH₄⁺ should be significantly from coal combustions. Besides
240 NO₃⁻, Na⁺ and Ca²⁺, automobile source PM_{2.5} also had the highest F⁻ and Mg²⁺ concentrations than other sources. The detailed
241 concentration distributions of WSIs in PM_{2.5} from each specific source were provided in Figures S12-S14. The WSIs levels
242 vary widely with specific source categories. PM_{2.5} from LDDVs-2 had the lowest amount of WSIs compared to the other
243 automobile exhausts (Figure S12). Similar to the metal composition, bituminous coal also had the lowest WSIs among all coals
244 (Figure S13). Compared to branches, PM_{2.5} from burning crop straws had much greater levels of K⁺, Cl⁻, SO₄²⁻ and less levels
245 of F⁻, NO₃⁻ (Figure S14).

246 To summarize, the overall concentrations of measured TC, cumulated heavy metals and WSIs in PM_{2.5} from each source
247 type were showed in Figure 5. Among all source emission and environmental receptor samples, the cumulated heavy metals
248 from coal combustion was highest and automobile exhaust was higher than ambient PM_{2.5}, the overall carbon contents from
249 automobile exhaust and biomass burning were both higher than ambient PM_{2.5}, while only the cumulated soluble ions in PM_{2.5}
250 from primary source of coal combustion was equivalent to the ambient aerosols. In a word, chemical compositions of PM_{2.5}
251 distributed much diversely and varied significantly with the specific source types of combustion emissions.

252 **3.3 Cell viability, oxidative stress and inflammation levels exposed to various mass-normalized PM_{2.5}**

253 Multiple toxicological endpoints (cell viability, oxidative stress, and inflammation) that facilitate identifying the specific
254 particle triggering ROS and inflammatory responses resulting in cell death were evaluated for source-specific PM_{2.5}. After 24

255 h exposure to the same dose of different PM_{2.5} obtained from specific emission sources, the A549 lung cells also showed varied
256 toxicological responses (Figure 6). The survival rate of cells exposed to automobile exhaust PM_{2.5} was much lower than
257 ambient air PM_{2.5} (Figure 6.1). Automobile exhaust PM_{2.5} induced the highest ROS production in cells higher than biomass
258 burning and both sources were also much higher than ambient PM_{2.5} (Figure 6.2). Coal combustion induced the highest cellular
259 IL-6 production followed by automobile exhaust that was also higher than ambient air PM_{2.5}, while the PM_{2.5} from automobile
260 exhaust and biomass burning induced similarly higher cellular production of TNF- α than ambient PM_{2.5} (Figure 6.3, 6.4).
261 These results suggested that, combustion primary emission PM_{2.5} had stronger ability to induce oxidative stress and
262 inflammatory injury in lung cells than ambient air PM_{2.5}, thus resulted in the higher probability of apoptosis induction (Victor
263 and Gottlieb, 2002; Wang et al., 2013). Generally, the mass-normalized PM_{2.5} from primary source of automobile exhaust
264 posed the strongest overall toxicity. Therefore, to protect public health by controlling PM_{2.5} pollution, the anthropogenic
265 combustions were key target sources, especially the most toxic automobile PM_{2.5} should be reduced preferentially.

266 **3.4 Correlations between various PM_{2.5} components and toxicity endpoints**

267 Spearman correlation coefficients between chemical compositions and cellular toxicological response indicators were applied
268 to screen the key components of all PM_{2.5} involved in cell injury (Figure 7). It was found that, the degrees of correlations
269 varied with the toxicological mechanisms of different airborne chemicals. Based on the overall PM_{2.5} samples from various
270 sources, the pro-inflammatory cytokine IL-6 showed significantly strong positive correlations with some heavy metals (As,
271 Pb, V, Cu), while TNF- α and oxidative stress (ROS) had similar significantly positive correlations with aerosol components
272 of carbon fractions (EC, OC) and transition metals (V, Cr, Ni). The TNF- α also showed positive correlation with water soluble
273 Cl⁻ and K⁺, and ROS correlated with F⁻, Ca²⁺ and Mg²⁺.

274

275 **4 Discussion**

276 **4.1 New chemical markers for source apportionments of ambient air PM_{2.5}**

277 Combustion emissions are key anthropogenic sources contributing to urban air PM_{2.5}, through both primary and secondary
278 aerosols, which were 66% and 34% calculated by PMF model, respectively (Figure 1). The high concentrations of chemical
279 markers are usually used in source analysis, such as ammonium sulfate and nitrate for secondary aerosols which are originated
280 mainly from the gaseous precursors (e.g., NH₃, SO₂ and NO_x) (Mahilang et al., 2021), the EC, Cu, Mn, and Ni for vehicle
281 exhaust (Srivastava et al., 2021), the As, Pb, OC, EC, SO₄²⁻ and relatively low NO₃⁻/SO₄²⁻ ratios for coal combustion (Dai et
282 al., 2020), soluble K⁺ and Cl⁻ for plant burning (Jain et al., 2020). The detailed chemical species of these specific source
283 emission PM_{2.5} samples also supported the results. Moreover, low OC/EC ratio of high TC content, high NO₃⁻, F⁻, Na⁺, Ca²⁺
284 and Mg²⁺, V and Mn of automobile exhaust; Pb and As, SO₄²⁻ and NH₄⁺ of coal combustion; soluble K⁺ and Cl⁻, and high
285 OC/EC ratio of high TC for plant biomass burning found in current study (Figures 2-5), could also be corresponding potential

286 aerosol source markers. The principal aim of this paper was to assess and contrast the chemical composition and potential
287 harmfulness of PM arising from diverse anthropogenic sources, thus natural sources, like fugitive soil dust, were not included
288 in the source examination.

289 **4.2 Common PM_{2.5} components related to specific combustion sources**

290 Generally, the automobile exhaust PM_{2.5} had high TC content and low OC/EC value with considerable EC content (Figure 2),
291 varying with specific vehicle types (Figure S4). The contents of the carbon fractions from diesel vehicles were higher than
292 gasoline exhausts, and the OC/EC ratios of diesel exhausts were much lower than gasoline vehicles, owing to both considerable
293 contents of EC and OC from diesel vehicle emission PM_{2.5}. Some diesel vehicles showed higher EC emissions with age, so
294 exhaust cleaning devices for them are suggested. In addition, the amounts of OC and EC in exhausts gradually decreased with
295 the strengthened emission standards they met (Wong et al., 2020). In PM_{2.5} samples obtained from coal combustion (Figure
296 S5), the TC contents of bituminous coals were significantly higher than that of honeycomb coals, anthracite coals, and
297 industrial coals, because bituminous coals contain higher volatile fraction, which is composed of organic matter. Therefore,
298 besides the way of combustion and the use of combustion stoves, the coal quality related to different coal types and origins
299 determine the carbonaceous fractions of the PM emitted by coal combustion (Zhang et al., 2022). In the PM_{2.5} samples from
300 plant biomass combustion (Figure S6), OC contents were generally higher than EC contents, except that pine branches
301 contained higher EC and rapeseed straw had considerable contents of EC and OC. Dominated by OC (Figure S7), the
302 concentrations of carbonaceous fractions in urban ambient air samples varied seasonally (Flores et al., 2020; Xu et al., 2019).
303 Combining the TC contents and OC/EC ratios, carbonaceous components in ambient PM_{2.5} mainly originate from semi-volatile
304 organic compounds (SVOCs) (Wang et al., 2018) and combustion primary emissions for OC (Kang et al., 2018), and
305 automobile exhaust for EC (Barraza et al., 2017). Thus, to control ambient carbon aerosol pollution, besides reducing the
306 precursor emissions of secondary organic aerosols (SOA), controlling primary aerosols especially EC from diesel vehicles
307 were key measures.

308 Airborne redox-active metals are usually linked with the oxidation stress of PM_{2.5}. Different types of automobiles emitted
309 diverse metal contents (Figure S8). Metal elements in automobile exhaust are primarily contributed by fuels, lubricants, and
310 engine component abrasion. Because Mn is a common antidetonator that delays and prevents the oxidation of hydrocarbons
311 and increases the octane number, which not only increases the thermal efficiency of the engine but also improves the emission
312 performance of the vehicle (Cheung et al., 2010), the Mn content was greater in gasoline vehicle exhausts than in diesel
313 vehicles. Although there are multi-sources of traffic Pb emissions such as fuel combustion and brake wear (Wang et al.,
314 2019; Panko et al., 2019), the automobile exhaust Pb content of gasoline vehicles were greater than diesel vehicles owing to
315 oil combustion. Moreover, for the same vehicle type (LDDVs-1 and 2; HDDVs-1 and 2; SDGCs-1 and 2), the stricter the
316 emission standard required, the lower the exhaust metal contents. The metal contents in the PM_{2.5} of trucks was higher than
317 that of passenger cars (Wu et al., 2016). In the combustion PM_{2.5} of 10 coal types (Figure S9), Pb contents were the highest

318 than other heavy metals, similar to available findings (Zhang et al., 2020). The PM_{2.5} metals from bituminous coal were
319 significantly lower than other coal types, because indicated by the coal quality analysis, bituminous coal has a low ash content
320 which is mainly derived from non-combustible minerals in coal. These findings suggested that coal maturity might be an
321 important factor influencing the metal composition of particulates emitted from coal combustion (Shen et al., 2021;Zhang et
322 al., 2021). Heavy metal contents in biomass burned PM_{2.5} varied much widely with raw plant types (Figure S10), although
323 dominated by Cr and Ni. Different plant species and even different plant parts differ significantly in their ability to uptake and
324 accumulate metals from soil (Zhao et al., 2020). Moreover, because of the high enrichment factors of some metals for crop
325 straws (Zhang et al., 2016;Sun et al., 2019), they also released more Cr, Ni, and Co during burning than fuelwoods. Total metal
326 emissions were highest in corn cob but lowest in peanut straw burning PM_{2.5}. The heavy metals enriched in urban ambient air
327 PM_{2.5} demonstrated a seasonal pattern (Chen et al., 2018;Hsu et al., 2016) (Figure S11). Contents of V, Co, and As were
328 relatively low and are less affected by seasonal changes. Accordingly, supported by the metal profiles of anthropogenic
329 combustion sources and ambient aerosols, to control the environmental airborne heavy metal pollution, the Pb, Cu and As
330 from honeycomb, anthracite and industrial coal combustion, Cu from vehicle exhausts and especially V from light duty diesel
331 van with the CN.III emission standard and Mn from gasoline vehicles, Cr and Ni from biomass especially crop straws burning,
332 should be key targets.

333 Epidemiological studies have also shown the mortality closely related to the WSIs such as sulfate and nitrate in aerosols
334 (Ostro et al., 2009;Liang et al., 2022). Among the WSIs contents of various automobile exhaust PM_{2.5} (Figure S12), NO₃⁻ and
335 Ca²⁺ were the most abundant anion and cation, respectively. The high NO₃⁻ in the automobile PM_{2.5} may be due to NO_x
336 production during high-temperature combustion (Hao et al., 2019), while the high Ca²⁺ content should be related to additives
337 in automobile fuels and calcium-based lubricants (Yang et al., 2019). Moreover, the exhaust WSIs decreased with the
338 strengthened automobile emission standards required. Coal combustion PM_{2.5} contained relatively higher SO₄²⁻ and NH₄⁺
339 concentrations followed by Cl⁻ than other WSIs species (Figure S13). Among various coal types, industrial coals emitted
340 highest SO₄²⁻ followed by honeycomb and industrial coal with also high NH₄⁺, but bituminous coals emitted low WSIs which
341 were mainly NO₃⁻, F⁻ and Na⁺, Ca²⁺. The WSIs emission factors of honeycomb coal were generally higher than those of lump
342 coal (Yan et al., 2020). For biomass combustion emissions (Figure S14), Cl⁻ and K⁺ were dominant WSIs in PM_{2.5} from straw-
343 type fuels (Tao et al., 2016;Sillapapiromsuk et al., 2013), but fuelwood-type combustion emitted high NO₃⁻. Plant species
344 absolutely determine the emissions (Liao et al., 2021). Finally, there were also high levels of NO₃⁻, SO₄²⁻, and NH₄⁺ in ambient
345 air PM_{2.5} (Zhang et al., 2019) (Figure S15). Consequently, implied by the WSIs species distributed in combustion primary
346 sources and environmental PM_{2.5}, to control the aerosols ions pollution, the NO₃⁻ from vehicle exhausts and fuelwood burning;
347 SO₄²⁻ and NH₄⁺ from honeycomb, anthracite and industrial coal combustion; Cl⁻ and K⁺ from biomass especially crop straw
348 burning, should be principal targets, by stricter automobile emission standards or using clean coals.

349 **4.3 PM_{2.5} toxicity related to specific sources by pivotal chemical components**

350 The complexity of the sources and compositions of atmospheric PM_{2.5} leads to different toxicological effects (Newman et
351 al., 2020; Kelly, 2021). The toxicological effects of PM_{2.5} are not comparable among different studies owing to distinct
352 exposure concentrations, biological models, endpoints, and PM_{2.5} generation methods (Park et al., 2018; Kelly and Fussell,
353 2020). In this study, we employed same exposure conditions and biological endpoints, in order to obtain comparable toxicity
354 data for PM_{2.5} from different sources. Our mass-normalized results demonstrated that automobile exhaust PM_{2.5} induced the
355 highest lethality and cellular ROS and TNF- α production, coal combustion PM_{2.5} induced the highest cellular IL-6 production,
356 plant biomass burning PM_{2.5} induced considerable cellular TNF- α and ROS production (Figure 6). Generally, various toxicities
357 of combustion emission primary PM_{2.5} were much greater than the urban ambient air PM_{2.5} (Figure 6), owing to the higher
358 concentrations of specific toxic components in PM_{2.5} from these sources. The supplementary information had included
359 exhaustive cytotoxicity indicators from each individual source (Figure S16-S19). While the survival rate of cell exposed to
360 CN.III emission standard PM_{2.5} was the lowest and the capacity to induce cells to produce ROS was the highest for CN.IV,
361 automobile exhaust had a similar potential to cause cells to produce inflammatory cytokines (Figure S16). The capability to
362 induce IL-6 production in cells was highest for industrial coal PM_{2.5}, whereas bituminous coal had the highest survival rate of
363 cells and TNF- α induction capacity (Figure S17). From the Figure S18 we can see that the PM_{2.5} cytotoxicity of straws and
364 branches burning was analogous, but it should be noted that the cell viability of various straw PM_{2.5} differs significantly, that
365 may be related to the raw fuel characteristics.

366 These possible mechanisms were implied by the overall relationships between the measured chemical components with
367 cytotoxicity indicators of PM_{2.5} from various specific sources (Figure 7). In general, both TNF- α and ROS were significantly
368 positively correlated with carbonaceous fractions and redox-active transition metals (V, Cr, Ni), which were main contributors
369 of automobile exhausts and biomass burning. The IL-6 was significantly positively correlated with some heavy metals (As and
370 Pb, V and Cu), which were main contributors of coal combustion sources. Potential mechanisms include that, carbon fractions
371 bound in PM_{2.5} could be transformed into reactive metabolites and then induce ROS production in cells (Stevanovic et al.,
372 2019), and the PM_{2.5} bound transition metals could also induce ROS production through the Fenton reaction and disrupt the
373 function of enzymes in cells (Verma et al., 2010; Sørensen et al., 2005; Zou et al., 2016). Oxidative stress can lead to
374 inflammatory infiltration of neutrophils and stimulate immune cells to produce inflammatory cytokines, among which TNF- α
375 and IL-6 play important roles in the inflammation development (Xu et al., 2020). Ultimately, excessive production of ROS
376 leads to dysfunctional endoplasmic reticulum responses and dysfunctional lipid metabolism in ROS bursts can result in cell
377 membrane damage and even cell death (Piao et al., 2018; Zhao et al., 2004). There have been some related supporting reports.
378 For instance, the OC and EC were significantly associated with biological responses of PM from vehicle emissions collected
379 in tunnels (Niu et al., 2020). The polar or quinone fractions of PAHs in diesel engine exhaust particles significantly contributed
380 to the heightened toxic response (Xia et al., 2004). The PM_{2.5} generated from biomass burning contained a substantial
381 concentration of carbonaceous components. In addition, Cr and Ni in PM₁₀ from straws were highly associated with ROS (Li
382 et al., 2023). In current study, cellular ROS was also correlated with water soluble Ca²⁺, F⁻, and Mg²⁺, which were main

383 contributors of automobile exhaust PM_{2.5}. The Ca²⁺ controls the membrane potential and regulates mitochondrial adenosine
384 triphosphate (ATP) production, and excessive Ca²⁺ leads to energy loss and more ROS production (Madreiter-Sokolowski et
385 al., 2020). Moreover, the TNF- α was also positively correlated with water soluble Cl⁻ and K⁺, which were main contributors
386 of plant burning PM_{2.5}. Therefore, the accumulations of some organic matters with high carbonaceous content (OC, EC) in
387 PM_{2.5} typically from automobile exhausts and plant biomass burning, redox-active metals (V, Cr, Ni) and water-soluble anions
388 (Cl⁻, F⁻) and cations (Ca²⁺, Mg²⁺) contributed by various combustions, might induce ROS production in cells, cause cellular
389 damage through oxidative stress and inflammatory responses, impair cell viability and finally harm human health.

390 Considering the multi-endpoints measured and the PM_{2.5} toxicity mechanisms mentioned above, based on the cell viability
391 first, and then ROS followed by inflammatory markers, together with the significantly related toxic chemical composition
392 contents (Park et al., 2018), we put forward a general sequence of overall mass-normalized toxicity for these combustion
393 source PM_{2.5} to managers. To improve the urban environmental air quality for best public health benefits by controlling
394 aerosols pollution, considering the differential toxicity intensity of each chemical component and their contributions from
395 various sources to ambient aerosols, preferential targets of specific primary PM_{2.5} sources and bound pollutants to be controlled
396 are suggested as following sequence: Reducing all anthropogenic combustions, especially decreasing the automobile exhaust
397 PM_{2.5} with high contents of EC, transition metals (V, Cu, Ni, Cr), and ions (Ca²⁺, Mg²⁺, F⁻, Na⁺) from diesel exhausts by
398 strengthening the emission standards, then lessening the coal combustion with high heavy metals (As, Pb, Cu) by replacement
399 with low-ash clean coals, and depressing the biomass burning with high OC, Ni, Cr, Cl⁻ and K⁺ from crop straw emissions.

400 **4.4 Limitations and perspectives**

401 In current study, we selected A549 cell based on previous abundant experimental experiences and also because it has been
402 used popularly in *in vitro* toxicology studies to elucidate the cellular and molecular mechanisms of PM involved in lung for
403 many decades (Li et al., 2022b). However, recently the human normal bronchial epithelial cell BEAS-2B was preferred over
404 the human lung adenocarcinoma epithelial cell A549. For instance, both cells were used in an aerosol study (Bonetta et al.,
405 2017), results of which highlighted the higher sensitivity of BEAS-2B cells respect to A549 also in samples with low level of
406 pollutants, because the PM_{0.5} samples from Italian towns can induce genotoxicity in normal cells while cancer cells might be
407 resistant to their adverse effects. Therefore, although our results are reasonable under the same exposure conditions, there were
408 still potential limitations of A549 cells since they may be more resistant to exposure to external compounds, and the generally
409 more sensitive BEAS-2B cells are suggested for future studies.

410 In toxicity assessments, cell vitality reflects the overall health of cells, encompassing factors such as cell membrane integrity,
411 intracellular metabolic activity, and cell proliferation capacity. Decreased cellular vitality may be associated with cell damage,
412 toxic effects, or cellular apoptosis. Inflammation markers are employed to assess the extent and nature of inflammatory
413 reactions, including the production of cytokines and inflammatory mediators, as well as the activation status of inflammatory
414 cells. Inflammation is a complex physiological response, typically delineated by the immune and inflammatory reactions of

415 the body to stimuli such as injury or infection. Alterations in inflammation markers can indicate the intensity and nature of the
416 inflammatory response. In this study, multiple biological responses of epithelial cells to various PM_{2.5} were evaluated,
417 including that, cell viability evaluated the mitochondrial dehydrogenase activity of the living cells, excessive intracellular ROS
418 formation induced by PM_{2.5} was responsible for oxidative stress to the cells, cytokines IL-6 and TNF- α were determined for
419 the effect of PM_{2.5} on pro-inflammatory response in cells. In general, *in vitro* data can be used to rank various types of particles
420 in terms of the toxic potential including possible carcinogenicity. Each marker will help to understand the hazard and toxicity
421 of PM_{2.5}. However, the toxicity of PM_{2.5} may be the result of multiple components acting through disparate physiological
422 mechanisms, with inconsistent relationships among endpoints (Park et al., 2018). For instance, in BEAS-2B cells, oxidative
423 stress generated by H₂O₂ exposure often results in cytotoxicity rather than by stimulating cytokine/chemokine responses,
424 sometimes no correlation between oxidative damage and cytokine/chemokine responses. Moreover, TNF- α gene was not
425 detected in BEAS-2B cells exposed to atmospheric PM collected from Benin, but the gene expression of other inflammatory
426 cytokines (IL-1 β , IL-6, and IL-8) were significantly induced, and decreasing cell viability was highly correlated with high
427 secretion of all studied cytokines (Cachon et al., 2014). Therefore, in the present study, it was impossible to analyze all
428 chemicals in PM_{2.5} and determine all related toxicological endpoints, so unmeasured chemicals and endpoints might also play
429 roles in the incongruous or unexplained results, and we also can't over-explain the mechanisms just based on statistical
430 relations. To overcome these hurdles, standardization of toxicological studies (experimental methodologies) and reporting
431 guidelines are necessary for tracking and comparing results.

432 This study ranked the unequal "toxic effects" based on the same mass concentration of PM_{2.5} exposure in body lung fluid
433 system, while the "health risks" usually relating to the inhalation exposure concentration of PM_{2.5} in ambient air were not
434 calculated and evaluated quantitatively. Moreover, non-linear concentration-response functions for various endpoints and
435 different exposure concentrations might also limit using toxicological data straightforwardly to predict health effects
436 (morbidity, mortality) in human populations, so drawing conclusions precisely quantifying/ranking the health risks of PM_{2.5}
437 from specific sources or of individual PM_{2.5} components is still not an easy task (Kelly and Fussell, 2020). Therefore, coupled
438 with source apportionment and exposure level of ambient aerosols pollution, toxicology combined with epidemiology studies
439 are essential for linking these factors and understanding scientific mechanisms to reach conclusions.

440

441 **5 Conclusions**

442 In current study, we found that 2/3 mass of urban ambient air PM_{2.5} in a representative megacity of eastern China originated
443 from primary sources of anthropogenic combustions including coal, automobile, and biomass. Because of the significant
444 differences in the chemical compositions, the diverse PM_{2.5} from both mixed ambient air and directly from individual
445 combustion sources showed much differential mass-normalized *in vitro* toxicity to the human lung epithelial cells, either for
446 the environmental aerosol samples collected from different seasons, or for the primary emissions of PM_{2.5} from various specific

447 source types. According to the comparative study and correlation analysis, the carbonaceous fractions (OC, EC) and redox-
448 active heavy metals (V, Ni, Cr) assisted by water-soluble ions (Ca^{2+} , Mg^{2+} , F^- , Cl^-) might play important roles in inducing
449 cellular ROS production, causing oxidative stress and inflammation, resulting in cell injury and apoptosis, thus damage human
450 health. These toxic pollutants accumulated in specific-source $\text{PM}_{2.5}$ varied by the emission types and raw fuel properties.
451 Combined with chemical composition and general cytotoxicity rank, the preferential controlling targets of specific combustion
452 sources should be automobile exhaust (diesel vehicles with emission standards inferior to CN.IV), coal combustion (high ash
453 and high sulfur coals), and plant biomass burning (crop straws). Although showing the synthetic effects of mixed compositions
454 and complex sources, besides preventing the secondary aerosols from combustions, preferentially targeted reductions of these
455 primary sources of toxic $\text{PM}_{2.5}$ direct emissions, would produce the greatest benefits for public health with improved ambient
456 air quality. Overall, this paper provides a precise, oriented, effective, efficient, and economical composition-source-based
457 strategies for urban aerosols pollution control. However, as a prospect, the detailed mechanisms for unequal toxicity of PM
458 with complicated components from various sources and their quantitative contributions to the health effects of ambient air
459 $\text{PM}_{2.5}$ mixture still need in-depth study.

460 **Supplementary materials**

461 There are 19 figures (Figure S1-S19) and 3 tables (Table S1-S3) in the Supporting Information.

462 **Data availability**

463 All raw data can be provided by the corresponding authors upon request.

464 **Author contributions**

465 XSL conceived and supervised the study; WH, YP, MT, HL, and ZZ collected the samples; WH, YP, MT, WL, HL, ZZ, GS,
466 and LX analyzed the chemical compositions; WH, YP, and MT performed the toxicity tests; WH, YP, MT, and XSL analyzed
467 the data; WH and XSL wrote the manuscript draft; XSL, WH, GS, and TM reviewed and edited the manuscript.

468 **Competing interests**

469 The authors declare that they have no conflict of interest.

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729 **Captions of figures**

730 **Figure 1.** The PMF factor profiles of various components and source percentages of secondary aerosol, automobile exhaust,
731 coal combustion, and plant biomass burning contributing to the urban ambient air PM_{2.5}.

732 **Figure 2.** Carbon contents (mg kg⁻¹) and ratio in PM_{2.5} from various specific sources (n=10 for each combustion source and
733 n=16 for urban ambient air).

734 **Figure 3.** Heavy metal contents (mg kg⁻¹) in PM_{2.5} from various specific sources (n=10 for each combustion source and n=16
735 for urban ambient air).

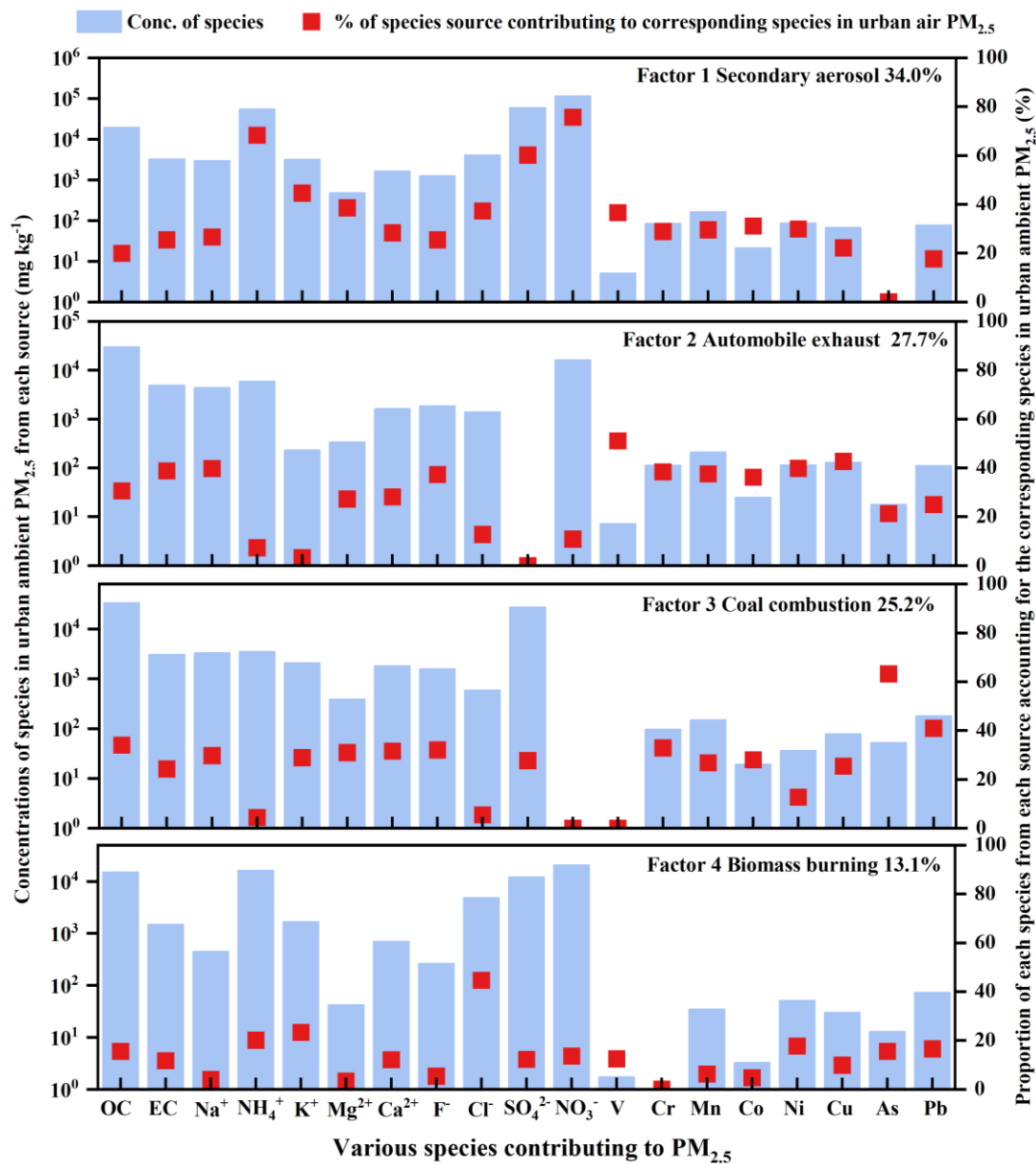
736 **Figure 4.** Water-soluble ion (WSI) contents (mg kg⁻¹) in PM_{2.5} from various specific sources (n=10 for each combustion
737 source and n=16 for urban ambient air).

738 **Figure 5.** Cumulated typical measured components (mg kg⁻¹) in PM_{2.5} from various specific sources (n=10 for each
739 combustion source and n=16 for urban ambient air).

740 **Figure 6.** Cell viability, oxidative stress and inflammation levels of human alveolar epithelial cell lines (A549) exposed to
741 PM_{2.5} suspension (80 mg L⁻¹) from various specific sources (n=10 for each combustion source and n=16 for urban ambient
742 air).

743 **Figure 7.** Overall correlations between typical cellular toxicological responses and chemical compositions of PM_{2.5} from
744 various sources (*p < 0.05, #p < 0.01; n=46).

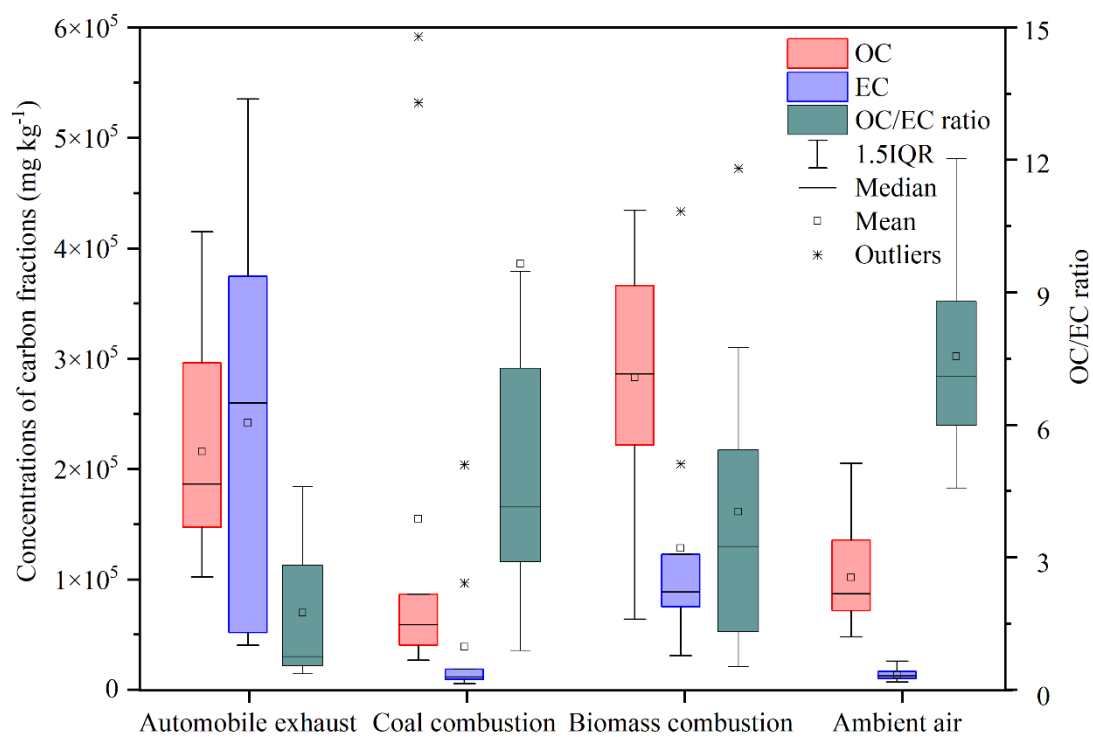
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748 **Figure 1.** The PMF factor profiles of various components and source percentages of secondary aerosol, automobile exhaust,
 749 coal combustion, and plant biomass burning contributing to the urban ambient air PM_{2.5}.

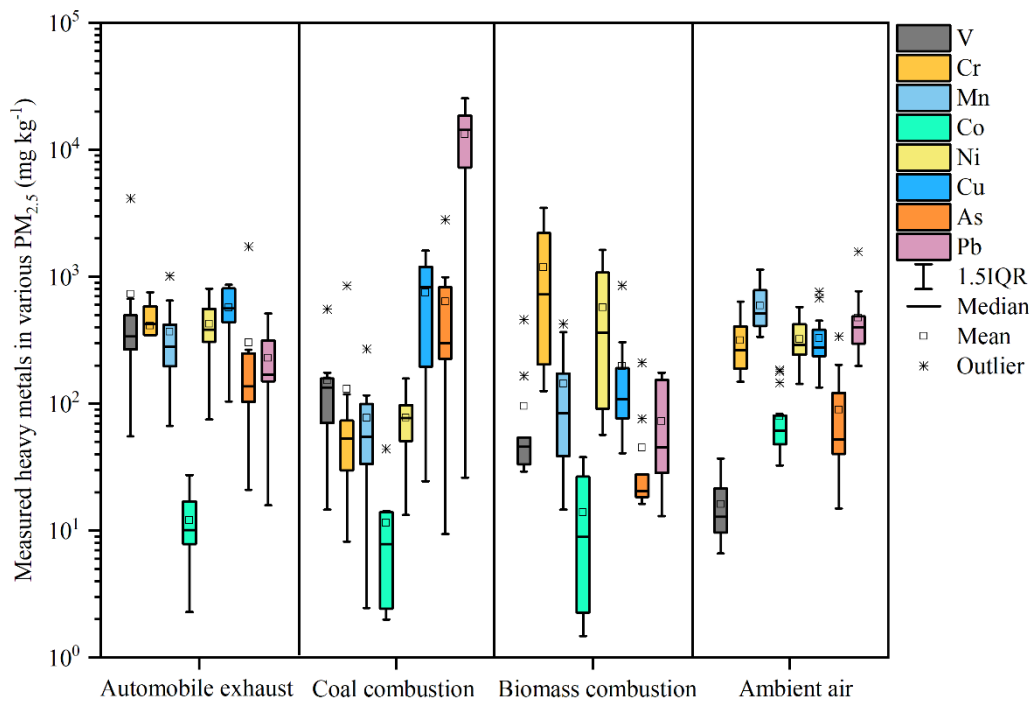
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753 **Figure 2.** Carbon contents (mg kg⁻¹) and ratio in PM_{2.5} from various specific sources (n=10 for each combustion source and
 754 n=16 for urban ambient air).

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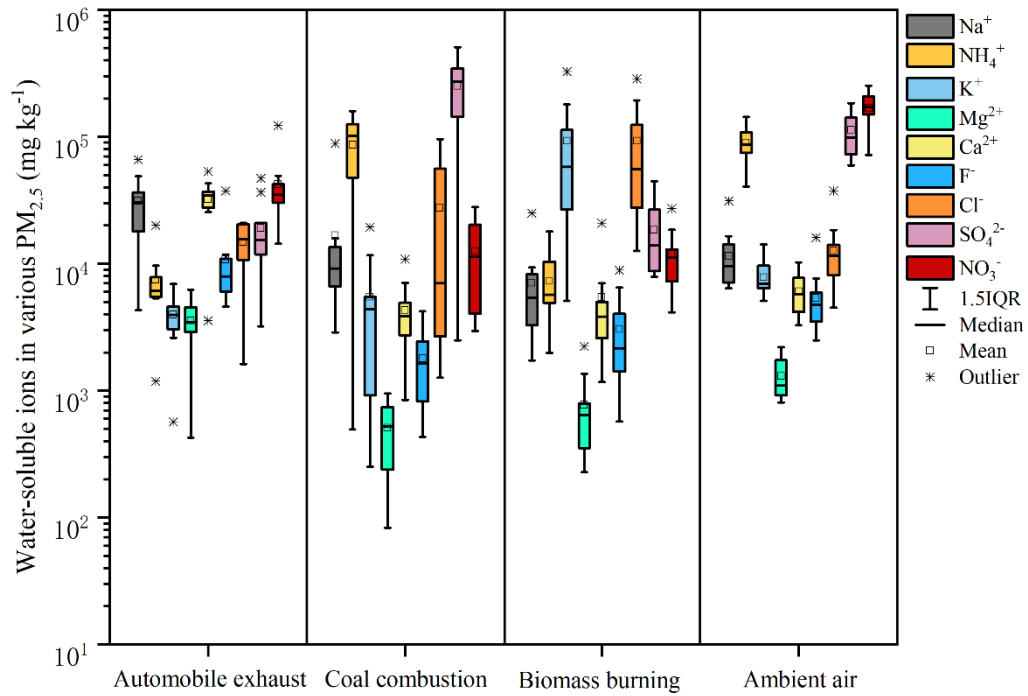


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757 **Figure 3.** Heavy metal contents (mg kg⁻¹) in PM_{2.5} from various specific sources (n=10 for each combustion source and

758 n=16 for urban ambient air).

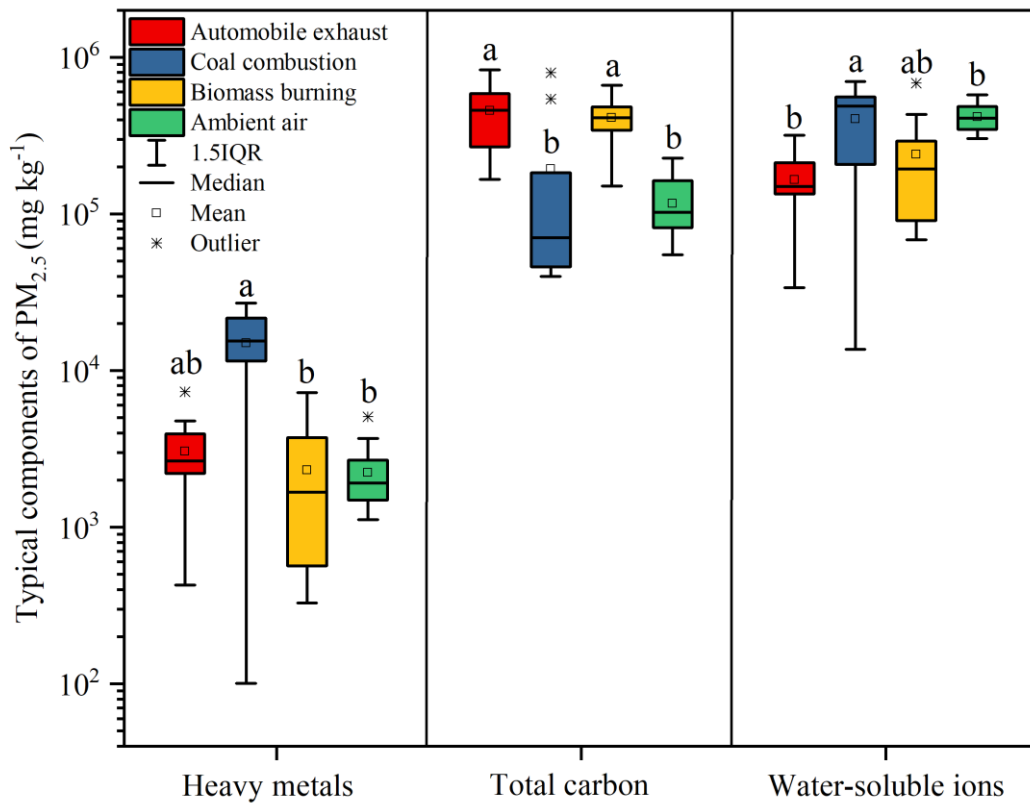
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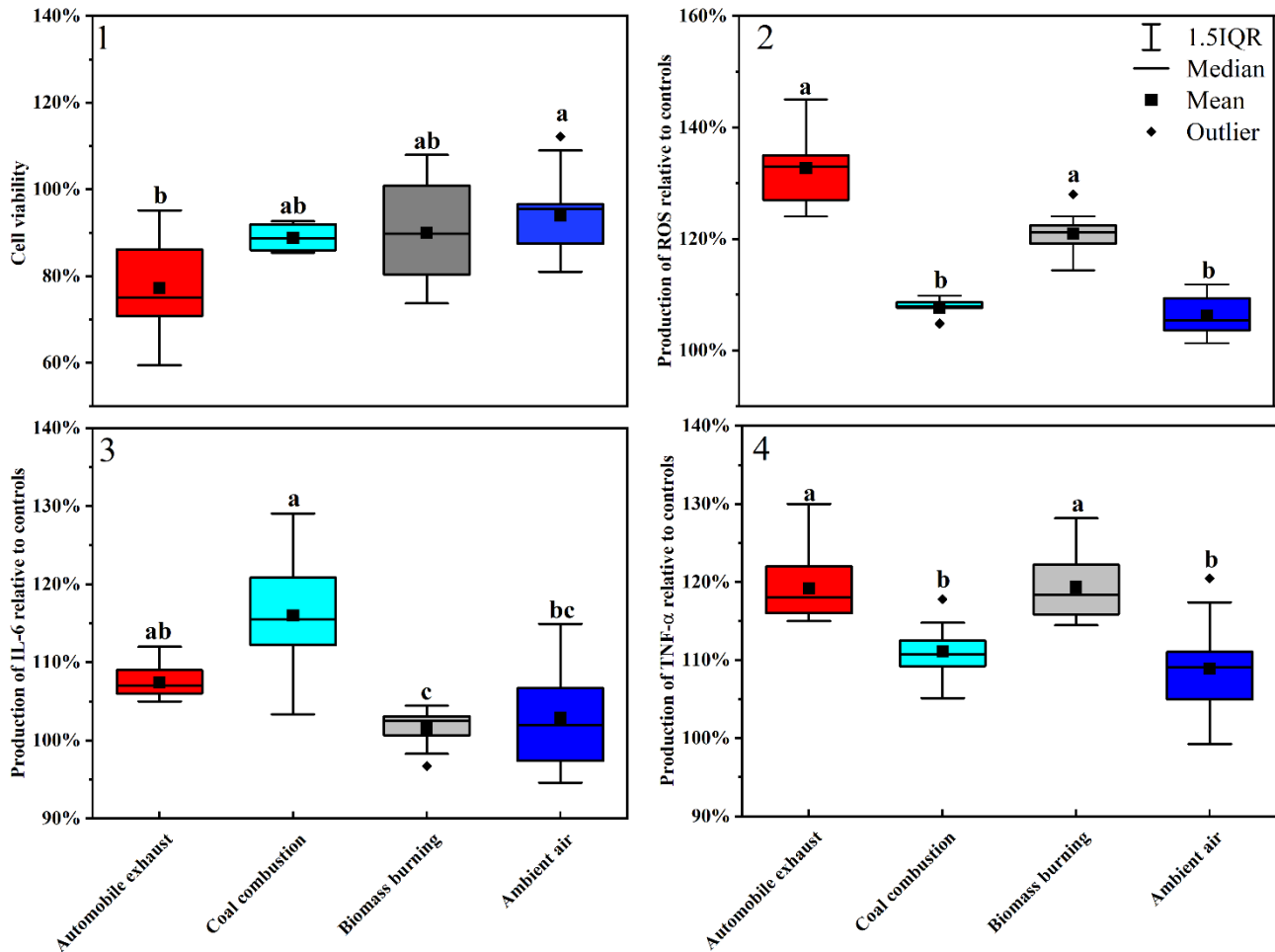
761 **Figure 4.** Water-soluble ion (WSI) contents ($mg\ kg^{-1}$) in $PM_{2.5}$ from various specific sources (n=10 for each combustion
762 source and n=16 for urban ambient air).

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766 **Figure 5.** Cumulated typical measured components (mg kg⁻¹) in PM_{2.5} from various specific sources (n=10 for each
 767 combustion source and n=16 for urban ambient air). The letters a and b are significant groups classified by Kruskal–Wallis
 768 test, p < 0.05.

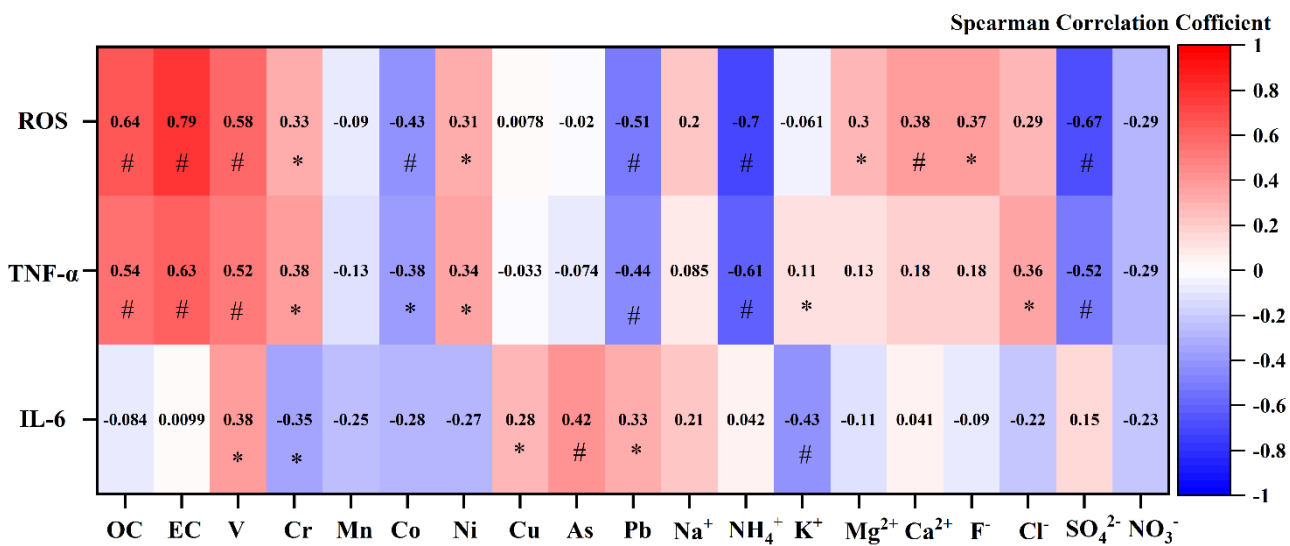


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771 **Figure 6.** Cell viability, oxidative stress and inflammation levels of human alveolar epithelial cell lines (A549) exposed to
 772 PM_{2.5} suspension (80 mg L⁻¹) from various specific sources (n=10 for each combustion source and n=16 for urban ambient
 773 air). The letters a, b and c are significant groups classified by Kruskal–Wallis test, p < 0.05.

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777 **Figure 7.** Overall correlations between typical cellular toxicological responses and chemical compositions of PM_{2.5} from

778 various sources (*p < 0.05, #p < 0.01; n=46).