

1 **Source differences in the components and cytotoxicity of PM<sub>2.5</sub> from**  
2 **automobile exhaust, coal combustion, and biomass burning**  
3 **contributing to urban aerosol toxicity**

4 **Xiao-San Luo<sup>1,#,\*</sup>, Weijie Huang<sup>1,#</sup>, Guofeng Shen<sup>2</sup>, Yuting Pang<sup>1</sup>, Mingwei Tang<sup>1</sup>, Weijun Li<sup>3</sup>,**  
5 **Zhen Zhao<sup>1</sup>, Hanhan Li<sup>1</sup>, Yaqian Wei<sup>1</sup>, Longjiao Xie<sup>4</sup>, Tariq Mehmood<sup>5</sup>**

6 <sup>1</sup>International Center for Ecology, Meteorology, and Environment, School of Applied Meteorology, Nanjing University of  
7 Information Science & Technology, Nanjing 210044, China

8 <sup>2</sup>Laboratory of Earth Surface Processes, College of Urban and Environmental Sciences, Peking University, Beijing 100871,  
9 China

10 <sup>3</sup>Department of Atmospheric Sciences, School of Earth Sciences, Zhejiang University, Hangzhou 310027, China

11 <sup>4</sup>Health Science Center, Peking University, Beijing 100871, China

12 <sup>5</sup>College of Ecology and Environment, Hainan University, Haikou 570228, China

13 *Correspondence: Xiao-San Luo (xsluo@nuist.edu.cn)*

14 *#Authors contributed equally to this work*

15 **Abstract.** Although air quality guidelines generally use the atmospheric concentration of fine particulate matter (PM<sub>2.5</sub>) as the  
16 metric for air pollution evaluation and management treating all particles as equally toxic, it is inconsistent with the facts that  
17 particle toxicity are significantly related to their sources and chemical compositions. Therefore, judging the most harmful  
18 source and identifying the toxic component will be extremely helpful to optimize air quality standards and prioritize targeted  
19 PM<sub>2.5</sub> control strategies to more protect public health effectively. The combustions of fuels, including oil, coal, and biomass,  
20 are main anthropogenic sources of environmental PM<sub>2.5</sub>, however, their discrepant contributions to health risks of mixed  
21 ambient aerosol pollution dominated by respective emission intensity and unequal toxicity of chemical components are still  
22 unclear. In order to quantify the differences among these combustion primary emissions, ten types of PM<sub>2.5</sub> from each typical  
23 source group, i.e., vehicle exhaust, coal combustion, and plant biomass burning, were collected for comparative study with  
24 toxicological mechanisms. Totally thirty type individual combustion samples were inter-compared with representative urban  
25 ambient air PM<sub>2.5</sub> samples, which chemical characteristics and biological effects were investigated by component analysis  
26 (carbon, metals, soluble ions) and *in vitro* toxicity assays (cell viability, oxidative stress, inflammatory responses) of human  
27 lung adenocarcinoma epithelial cells (A549). Carbonaceous fractions were plenteous in automobile exhaust and biomass  
28 burning, while heavy metals were more plentiful in PM<sub>2.5</sub> from coal combustion and automobile exhaust. The overall ranking  
29 of mass-normalized cytotoxicity for source-specific PM<sub>2.5</sub> was automobile exhaust > coal combustion > plant biomass burning >  
30 ambient urban air, possibly with differential toxicity triggers, that the carbonaceous fractions (organic carbon, OC; elemental  
31 carbon, EC) and redox-active transition metals (V, Ni, Cr) assisted by water-soluble ions (Ca<sup>2+</sup>, Mg<sup>2+</sup>, F<sup>-</sup>, Cl<sup>-</sup>) might play

32 important roles in inducing cellular reactive organic species (ROS) production, causing oxidative stress and inflammation,  
33 resulting in cell injury and apoptosis, thus damage human health. Coupled with the source apportionment results of typical  
34 urban ambient air PM<sub>2.5</sub> in eastern China, reducing toxic PM<sub>2.5</sub> from these anthropogenic combustions will be greatly beneficial  
35 to public health, especially preferentially decreasing the diesel exhaust by strengthening emission standards, then lessening the  
36 coal combustion by replacement with low-ash clean coals, and depressing the crop straw burning emissions.

37

## 38 **1 Introduction**

39 As a mixture of multiple sources, ambient particulate matter (PM) arise from anthropogenic activities are continuously  
40 deteriorating the urban air quality, particularly in developing countries. Among these, fine PM with an aerodynamic diameter  
41 of less than 2.5  $\mu\text{m}$  (PM<sub>2.5</sub>) is recognized as a serious public health concern due to its long persistence in air, carcinogenicity  
42 and acute toxicity to humans (Al-Kindi et al., 2020). There were extensive epidemiological evidences that airborne PM can  
43 cause serious negative effects on human health, such as respiratory and cardiovascular diseases, genetic mutations, and  
44 developmental disorders (Chowdhury et al., 2022;Lelieveld et al., 2021;Smith, 2021;Clemens et al., 2017). Currently, either  
45 the world air quality guidelines or the national air quality standards use the mass concentration of PM<sub>2.5</sub> as the metric for PM<sub>2.5</sub>  
46 pollution evaluation and management, in which all particles are treated as equally toxic, however, it is inconsistent with the  
47 scientific facts that particle toxicity are significantly related to their sources and chemical compositions (Shiraiwa et al., 2017).  
48 Therefore, to identify which component(s) and source(s) of ambient PM are most harmful to health, will be very helpful to  
49 optimize air quality guidelines/standards and prioritize targeted PM control strategies to more effectively protect public health  
50 (Kelly and Fussell, 2020).

51 Besides natural sources like dust and sea spray, the vast majority of aerosols come from anthropogenic activities especially  
52 energy consumption, including the combustion of fossil fuels causing industrial emissions and automobile exhaust, and  
53 biomass burning (McDuffie et al., 2021;Wu et al., 2022). Finally, these diverse sources make the ambient air PM<sub>2.5</sub> become a  
54 complex mixture with multiple chemical components varying with time and space, which consisting mainly of sulfate, nitrate,  
55 ammonium, organic carbon (OC), elemental carbon (EC), mineral and trace metals (Bari and Kindzierski, 2016; Kelly and  
56 Fussell, 2020). The physiological mechanisms of PM-induced cell toxicity in respiratory system have been continuously  
57 investigated with some progresses (Kelly and Fussell, 2012, 2020; Shiraiwa et al., 2017; Mack et al., 2020; Li et al., 2022b),  
58 such as the metabolic activation, oxidative stress, inflammatory response, and apoptosis, focused on by current study. In brief,  
59 after inhalation and deposition onto the epithelium, redox-active materials in PM<sub>2.5</sub> can induce the release of reactive organic  
60 species (ROS), which cause oxidative stress (an imbalance between ROS and antioxidants, i.e., disequilibrium of the redox  
61 state of a cell) followed by inflammation and cell death. The ROS can mediate subsequent signaling pathways leading to  
62 biomolecule damage (e.g., DNA, lipid, and protein) and cellular injury, through mediating inflammatory responses including  
63 the release of pro-inflammatory cytokines like IL-6 and TNF- $\alpha$  by epithelial cells (Sabbir Ahmed et al., 2020; Landwehr et al.,

64 2021). For instance, oxidative stress could trigger the induction of pro-inflammatory transcription factors, such as nuclear  
65 factor (NF)- $\kappa$ B, via the mitogen-activated protein kinase (MAPK) signaling pathway. Components adsorbed on particle surface,  
66 such as redox-active metals (transition metals, Fe, Ni, V, Cr, Cu), organic compounds (polycyclic aromatic hydrocarbons,  
67 PAHs; quinones), or even carbonaceous core of particles, are responsible for oxidative stress (Cachon et al., 2014; Sabbir  
68 Ahmed et al., 2020). The non-redox active metals (Zn, Pb, Al) can also influence the toxic effects of transition metals by  
69 exacerbating or lessening the production of free radicals. The EC may not be a directly toxic component of PM<sub>2.5</sub> but rather  
70 operate as a universal carrier of combustion-derived chemicals (semi-volatile organic fractions, transition metals) of varying  
71 toxicity (Kelly and Fussell, 2020). Inorganic soluble sulphates and nitrates are acidic and can interact with and influence the  
72 solubility other compositions like metal bioavailability (Fang et al., 2017; Weber et al., 2016). However, which specific  
73 components and which particular sources are the most critical factors dominating the ambient aerosols' health risks, still leave  
74 puzzles unsolved.

75 Past studies performed in various countries have focused on physicochemical characterization or biological effects of  
76 ambient air PM<sub>2.5</sub> respectively (Weagle et al., 2018; Jia et al., 2017; Wang et al., 2020). For example, the source analysis of  
77 PM<sub>2.5</sub> by photochemical modelling (Bao et al., 2018), chemical composition of regional PM<sub>2.5</sub> (Chi et al., 2022), and the  
78 mechanism of PM<sub>2.5</sub> toxicity was independently reported recently (Jia et al., 2020). Because differences in particle composition,  
79 sources, and toxicity appear in different urban environments (Zhao et al., 2019; Borlaza et al., 2018), the source profiles of  
80 different emission inventories were applied to elucidate aerosol pollution characteristics and control strategies. For instance, it  
81 was found that straw burning during the harvest season is a major trigger of severe air pollution in many regions (Sahu et al.,  
82 2021). Aerosols from open biomass burning in the Amazon had a stronger ability to induce ROS than laboratory-generated  
83 secondary organic aerosols (Tuet et al., 2019). The particle composition of motor vehicle exhaust was related to automobile  
84 types with various fuels, engines, and loads (Lin et al., 2020). A strong catalytic reactivity of metals in PM emitted from diesel  
85 vehicles was observed by dithiothreitol (DTT) assay (Jesus et al., 2018). Sulfate is a major component of PM from Xi'an city,  
86 western China, mainly released from residential coal combustion activities (Dai et al., 2019). Traffic was suggested playing  
87 the most crucial role in enhancing the toxicity of fine particles (Park et al., 2018). Although there were emerging studies on  
88 particle emission from single source, quantitatively comparative studies on multi-source pollutants as well as the differential  
89 composition and unequal toxicity of various sources are still limited.

90 The main objective of current study was to compare the chemical components and corresponding mass-normalized  
91 toxicological effects of individual PM<sub>2.5</sub> from various combustion sources and their unequal contributions to ambient aerosol  
92 health risks. The aim is to provide detailed guidance on the targeting and precise control of specific anthropogenic sources  
93 with prominent risks based on their pivotal toxic components. Therefore, we collected both representative ambient PM<sub>2.5</sub>  
94 samples from urban air and abundant typical source PM<sub>2.5</sub> samples from automobile exhaust, coal combustion, and plant  
95 biomass burning. Their independent profiles of chemical compositions and *in vitro* cytotoxicity (cell viability, oxidative stress,  
96 and inflammatory responses) were investigated and intercompared, to assess the differences in source-to-receptor toxicity and

97 to infer the core toxic components and respective harmful contribution. The pivotal toxic components were identified based  
98 on the source-sink bi-directional composition-effect results, which were further used to assess the health toxicity contribution  
99 of various emission sources to ambient air PM<sub>2.5</sub>, supported by its source apportionment through positive matrix factorization  
100 (PMF) model. This study could advance the understanding to quantify the complex source contribution to high-risk PM<sub>2.5</sub>  
101 emission oriented to public health, which is imperative for precise prevention and control of atmospheric PM pollution.  
102

## 103 **2 Materials and methods**

### 104 **2.1 Collection of PM<sub>2.5</sub> samples from primary emissions of 30 typical combustion sources and from representative** 105 **ambient urban air**

106 As the main anthropogenic sources of the ambient air PM<sub>2.5</sub> pollution, totally 30 types of primary PM<sub>2.5</sub> samples emitted  
107 directly from automobile exhaust, coal combustion, and plant biomass burning were respectively collected as follows for both  
108 chemical and toxicological analyses.

109 Based on the classification of automobile fuel types as well as load and tailpipe emission standards provided by the 2019  
110 Annual Report on Environmental Management of Mobile Sources in China, a total of 10 types of vehicles were chosen for  
111 exhaust investigation. They were further categorized into 7 sub-groups, including small duty gasoline coaches (SDGCs), small  
112 duty diesel coaches (SDDCs), middle duty diesel coaches (MDDCs), heavy duty diesel coaches (HDDCs), light duty diesel  
113 vans (LDDVs), middle duty diesel vans (MDDVs), and heavy duty diesel vans (HDDVs). The detailed information of these  
114 representative local automobiles was showed in Table S1.

115 To cover all coal types consumed in the city, 10 representative types of coal were gathered for investigation. They were  
116 further classified into 4 sub-groups, including 2 types of honeycomb coal (HC), 3 types of anthracite coal (AC), and 2 types  
117 of bituminous coal (BC) mainly for restaurant or household use, and 3 types of industrial coal (IC) for coal-fired power plants  
118 and steel-smelting industry. The detailed characteristic analysis of these typical coals purchased from local market were  
119 showed in Table S2.

120 Considering the plant biomass combustion in rural areas surrounding the megacity, 10 representative types of agricultural  
121 and forestry solid wastes were gathered for investigation. Because of the high annual production of three staple food crops  
122 (rice, wheat, and corn) as well as soybean, peanut and rapeseed, their straws generated during harvest are often used as fuels  
123 in rural households. In addition, woods were also common fuels. Therefore, straws of rice, wheat, corn, soybean, peanut, rape,  
124 and sesame, corncob, branches of peach and pine, were selected as plant biomass fuels and further divided into 2 sub-groups,  
125 including 8 types of crop straw and 2 types of firewood. The detailed characteristic analysis of these typical plant biomass  
126 fuels collected from rural areas around Nanjing city were showed in Table S3.

127 The PM<sub>2.5</sub> samples directly emitted from these combustion sources were collected by dilution channel sampling method  
128 (Figure S1), using a 4-channel particulate matter dilution sampler (HY-805, Hengyuan Technology Development Co., CN).

129 Each sampling included 3 parallel channels of quartz microfiber filter (Figure S2) and 1 channel of Teflon membrane filter  
130 with diameters of 47 mm, through a size selector for PM<sub>2.5</sub> with a flow rate of 160 L min<sup>-1</sup>. Clean air was pumped for 10 min  
131 before and after each sample was collected. Before using, the blank quartz filters were incinerated by a muffle furnace at  
132 500 °C for 3 h to remove any possible organic matters, while Teflon filters were baked at 60 °C for 4 h. After being equilibrated  
133 in a constant temperature and humidity chamber for 24 h, the filters were weighed both before and after sampling for  
134 gravimetric measurements, then the mass of collected PM<sub>2.5</sub> could be calculated. The sampled filters were stored in a  
135 refrigerator at -20 °C before analysis. The quartz filter loaded PM<sub>2.5</sub> samples were used for carbon and ion analysis, and for  
136 toxicity tests, while the parallel Teflon filter loaded samples were used to determine metals.

137 As the actual mixture of various source particles in real environment, totally 16 representative ambient air PM<sub>2.5</sub> samples  
138 (each time lasting 23h) covering a year monthly were collected from December 2019 to October 2020 in an urban site  
139 surrounded by traffic, residential and commercial quarters of Nanjing city, Yangtze River Delta of eastern China, using a high-  
140 volume air sampler (800 L min<sup>-1</sup>) with quartz microfiber filters. Detailed procedures and sample information were described  
141 in previous paper (Li et al., 2022a), but the purpose of using these air samples in current study was to compare them with the  
142 specific source samples for evaluating the chemical and toxicological contributions of the combustion primary sources to  
143 environmental aerosols pollution.

## 144 **2.2 Chemical composition analysis**

145 All collected source and ambient PM<sub>2.5</sub> samples were conducted various component analysis (Li et al., 2023). For the  
146 concentrations of heavy metals in particulates, samples were digested by concentrated HNO<sub>3</sub>-HClO<sub>4</sub> acids with a progressive  
147 heating program and determined by inductively coupled plasma optical emission spectrometry (ICP-OES; Optima8000,  
148 PerkinElmer), with some elements at lower concentrations measured by ICP mass spectrometry (ICP-MS; NexIONTM300X,  
149 PerkinElmer). Blank filter, reagent blank, replicates, and standard reference material (NIST SRM 1648a, urban dust) were  
150 adopted for analytical quality control, with recoveries ranged 90-110 %. Carbonaceous species (OC and EC) in PM<sub>2.5</sub> were  
151 determined using a DRI-2001A OC/EC (Atmoslytic Inc., Calabasas, CA, USA). For the concentrations of water-soluble ions  
152 (WSIs), the main cations (Na<sup>+</sup>, K<sup>+</sup>, Mg<sup>2+</sup>, Ca<sup>2+</sup>, NH<sub>4</sub><sup>+</sup>) and anions (NO<sub>3</sub><sup>-</sup>, SO<sub>4</sub><sup>2-</sup>, Cl<sup>-</sup>, F<sup>-</sup>) in PM<sub>2.5</sub> were measured by ion  
153 chromatography (IC, Thermo Fisher Scientific, USA), using the Metrosep C6-150/4.0 column for cations and the Metrosep A  
154 Supp 5 150/4.0 column for anions, respectively.

## 155 **2.3 Preparing mass-normalized PM<sub>2.5</sub> suspension for cell exposure**

156 Totally 30 source and 16 ambient PM<sub>2.5</sub> samples were also performed cytotoxicity tests. In order to elute the particles  
157 completely from the quartz membranes, the PM<sub>2.5</sub>-loaded sample filter was cut into small pieces, immersed in ultrapure water  
158 and extracted six times (30 min for each) in an ultrasonic bath at 0 °C. Although the ultrasonication might impact the ROS  
159 (Miljevic et al., 2014), the inevitable systematical error was ignored in this study. The extract was then suction filtered through  
160 a 2.6 μm pore-size nylon membrane to remove possible quartz fragments, and the bulk filtrate was freeze-dried back to pure

161 PM<sub>2.5</sub> powder. Ultimately, based on particle mass, the gathered PM<sub>2.5</sub> was dispersed by sterile phosphate-buffered saline (PBS)  
162 to a concentration of 400 mg L<sup>-1</sup>, and then diluted to PM<sub>2.5</sub> suspension of 80 mg L<sup>-1</sup> with serum-free Dulbecco's modified eagle  
163 medium (DMEM) medium for following *in vitro* cell exposure (Li et al., 2022a).

#### 164 **2.4 Cell culture and cellular toxicity tests by *in vitro* PM<sub>2.5</sub> exposure**

165 Aerosol pollution can harm lung alveoli and epithelial cells, and the A549 human lung adenocarcinoma epithelial cell has long  
166 been used as a suitable epithelial alveolar model to investigate the interactions between PM and lung epithelial cells (Park et  
167 al., 2018; Li et al., 2022b). The A549 cells were cultured in RPMI-1640 medium (Gibco, USA) supplemented with 10% fetal  
168 bovine serum (FBS, Hyclone, USA) and 1% antibiotic penicillin-streptomycin (100 U mL<sup>-1</sup>) at 37 °C in a 5% CO<sub>2</sub> incubator.  
169 After PM<sub>2.5</sub> exposure, cell viability and the indicators reflecting oxidative damage and inflammatory responses were  
170 determined respectively. While the cell viability assay was helpful in determining PM<sub>2.5</sub> dose to cells, the endogenous ROS  
171 measurements revealed the status of cellular oxidative potential after PM<sub>2.5</sub> exposure followed by the relative effects of ROS  
172 on various stages of cellular toxicity like inflammatory responses (Gali et al., 2019). The cell viability (metabolic activity) was  
173 evaluated by mitochondrial activity and determined by the methyl-thiazol-tetrazolium (MTT) assay (Chen et al., 2019). After  
174 trypsin action, the density of cells in the logarithmic growth phase was adjusted to 1 × 10<sup>5</sup> mL<sup>-1</sup>. Cell suspensions were  
175 inoculated into 96-well plates (Costar, USA) at 100 μL per well. The blank control well (without medium and PM<sub>2.5</sub> suspension)  
176 and reagent control well (with medium but without PM<sub>2.5</sub> suspension) were set together. After incubation for 24 h and removing  
177 the cellular supernatant, various types of PM<sub>2.5</sub> suspension (concentration of 80 mg L<sup>-1</sup>) were added to 96-well plates and  
178 incubated for 24 h. Based on pre-experiments, the oxidative stress and inflammation response sensitively under this dose,  
179 while the cell viability can keep sufficient. Fresh medium and MTT reagent (Solarbio, Beijing, CN) were added to each well  
180 and the supernatant was discarded, then 100 μL of formazan lysate was added to each well. The optical density (OD) values  
181 were measured at 490 nm using a microplate reader (Thermo MULTISKAN FC, USA). Cell viability (%) = (OD<sub>treatment</sub> –  
182 OD<sub>blank control</sub>) / (OD<sub>reagent control</sub> – OD<sub>blank control</sub>). The levels of cellular ROS production causing oxidative stress in cells, pro-  
183 inflammatory cytokines including tumor necrosis factor-alpha (TNF-α) and interleukin-6 (IL-6) production for determining  
184 the expression of genes related to the inflammatory response in the supernatant were analyzed by enzyme-linked  
185 immunosorbent assay (ELISA) kits (Jiangsu Enzyme Biotechnology Co., Ltd., CN), and OD values were measured at 450 nm  
186 (Huang et al., 2020; Pang et al., 2020).

#### 187 **2.5 Data analysis**

188 The statistical analysis was performed by IBM SPSS statistics 24 and plotted by Origin 2020b software. Spearman correlation  
189 coefficients were produced by the correlation analysis. The variance was statistically significant when the statistical test level  
190 was p < 0.05, and extremely significant when p < 0.01. Statistical analyses were performed using Kruskal–Wallis test (Kruskal  
191 and Wallis, 1952).

192 The source apportionment of PM<sub>2.5</sub> mass in urban ambient air was conducted by the receptor model PMF (EPA PMF version  
193 5.0). Major constituents (OC, EC, Cu, Cr, Co, Ni, As, Pb, Mn, V, Na<sup>+</sup>, K<sup>+</sup>, Mg<sup>2+</sup>, Ca<sup>2+</sup>, NH<sub>4</sub><sup>+</sup>, Cl<sup>-</sup>, F<sup>-</sup>, NO<sub>3</sub><sup>-</sup>, and SO<sub>4</sub><sup>2-</sup>) were  
194 selected as input data, and a four-factor solution was chosen as the optimal solution based on an assessment of the  
195 interpretability of the source profiles and the seasonal variability of the source contributions.

196

### 197 **3 Results**

#### 198 **3.1 Contributions of combustion primary sources to urban ambient air PM<sub>2.5</sub>**

199 As shown in Figure S3, although have been significantly improved with the national air quality in recent years, the daily PM<sub>2.5</sub>  
200 concentrations of representative city Nanjing still exceeded the healthy guidelines obviously, with higher urban air PM<sub>2.5</sub>  
201 pollution level in the cold season<sup>23</sup>. Four major sources of the ambient PM<sub>2.5</sub> were produced by the PMF model, including  
202 secondary aerosols, and primary particles of automobile exhaust, coal combustion, and plant biomass burning, which account  
203 for 34%, 27.7%, 25.2%, and 13.1% of total PM<sub>2.5</sub> mass concentration, respectively. Their source profiles and proportions were  
204 showed in Figure 1. Therefore, although the contribution of secondary aerosols cannot be ignored, the main anthropogenic  
205 sources of urban air PM<sub>2.5</sub> were primary emissions (66%) from the various fuel combustions.

#### 206 **3.2 Chemical compositions of different PM<sub>2.5</sub> from 30 combustion sources and from representative urban ambient air**

207 Typical chemical components including carbonaceous fractions, heavy metals and WSIs of all PM<sub>2.5</sub> samples from both  
208 ambient urban air and 30 representative combustion primary sources (covering different categories of automobile exhaust, coal  
209 combustion, and plant biomass burning) were analyzed and compared with each other.

210 According to the comparisons of PM<sub>2.5</sub> bound carbonaceous fractions (Figure 2), automobile and biomass sourced PM<sub>2.5</sub>  
211 contained significantly higher total carbon (TC) content than coal combustion and ambient air, while the OC/EC ratio trend  
212 was ambient air > coal combustion > biomass burning > automobile exhaust sources. It indicated that the carbon content of  
213 ambient PM<sub>2.5</sub> mixture was lower and dominated by OC than that of combustion primary sources. Figures S4-S7 showed the  
214 detailed carbon fraction characteristics (contents and ratio) of PM<sub>2.5</sub> from each specific source. Carbonaceous fractions in  
215 automobile exhaust PM<sub>2.5</sub> were high but the difference between OC and EC content was small. Depending on the diverse  
216 automobile fuels, loads and tailpipe emission standards, the concentrations of carbon fractions in exhaust PM<sub>2.5</sub> varied widely  
217 with vehicle categories. The carbonaceous portion of PM<sub>2.5</sub> gradually declines as emission regulations rise, and EC likewise  
218 declines dramatically (Figure S4). However, such differences among coal types were less, except the bituminous coal with  
219 extreme high OC (Figure S5). The carbonaceous fraction of PM<sub>2.5</sub> from plant biomass burning differed in raw material species  
220 that tree branches source PM<sub>2.5</sub> generally contained higher carbon contents than those from crop straws (Figure S6).

221 Based on the grouped (Figure 3) and individual (Figures S8-S11) distributions of the measured heavy metals in various  
222 PM<sub>2.5</sub>, the V concentrations of combustion sources were generally higher while Co and Mn were lower than ambient urban air.

223 Coal combustion emissions carried highest levels of Pb and were enriched in Cu and As (Figure S9), while biomass burning  
224 were rich in Cr and Ni (Figure S10). However, automobile exhausts were enriched in most heavy metals, especially Cu, and  
225 Cr, Ni, V, Mn (Figure S8). Heavy metals from different types of automobile exhausts with the same emission standard varies  
226 greatly. Anthracite and industrial coal combustions contain similar heavy metals much more than bituminous coal. Generally,  
227 Pb, V, Mn, As, and Cu in branches source PM<sub>2.5</sub> were higher than straws, while Cr, Ni, and Co were dominant and higher in  
228 straw burning emissions. A special discovery was that corn cob burning PM<sub>2.5</sub> carried more heavy metals than corn straw and  
229 was the biomass with the highest emission levels of heavy metals. Correspondingly, ambient air PM<sub>2.5</sub> were also rich in most  
230 metals, especially Mn, Pb, and Ni, Cu, Cr. Therefore, coal combustion sources might contribute most Pb to urban ambient air,  
231 and contribute significant Cu and As with automobile exhaust emissions, while plant biomass burning and automobile sources  
232 contribute the Cr and Ni. Besides natural dust, automobile exhaust should be the main anthropogenic source of airborne Mn.  
233 Considering the PMF source apportionments of ambient aerosols, automobile exhaust should be the main source of Cr in urban  
234 air PM<sub>2.5</sub>, and also the source for Cu together with coal combustion.

235 According to the comparisons of water-soluble cation and anion concentrations in various PM<sub>2.5</sub> (Figure 4), coal  
236 combustions contained highest SO<sub>4</sub><sup>2-</sup> and NH<sub>4</sub><sup>+</sup>, automobile exhausts had highest contents of NO<sub>3</sub><sup>-</sup>, Na<sup>+</sup> and Ca<sup>2+</sup>, while plant  
237 biomass burning sources contained highest K<sup>+</sup> and Cl<sup>-</sup>, but Mg<sup>2+</sup> was the lowest for all sources. However, the urban ambient  
238 air PM<sub>2.5</sub> contained highest NO<sub>3</sub><sup>-</sup> and were also dominated by SO<sub>4</sub><sup>2-</sup> and NH<sub>4</sub><sup>+</sup>, for which NO<sub>3</sub><sup>-</sup> should be mainly contributed  
239 by secondary aerosols and automobile primary source, SO<sub>4</sub><sup>2-</sup> and NH<sub>4</sub><sup>+</sup> should be significantly from coal combustions. Besides  
240 NO<sub>3</sub><sup>-</sup>, Na<sup>+</sup> and Ca<sup>2+</sup>, automobile source PM<sub>2.5</sub> also had the highest F<sup>-</sup> and Mg<sup>2+</sup> concentrations than other sources. The detailed  
241 concentration distributions of WSIs in PM<sub>2.5</sub> from each specific source were provided in Figures S12-S14. The WSIs levels  
242 vary widely with specific source categories. PM<sub>2.5</sub> from LDDVs-2 had the lowest amount of WSIs compared to the other  
243 automobile exhausts (Figure S12). Similar to the metal composition, bituminous coal also had the lowest WSIs among all coals  
244 (Figure S13). Compared to branches, PM<sub>2.5</sub> from burning crop straws had much greater levels of K<sup>+</sup>, Cl<sup>-</sup>, SO<sub>4</sub><sup>2-</sup> and less levels  
245 of F<sup>-</sup>, NO<sub>3</sub><sup>-</sup> (Figure S14).

246 To summarize, the overall concentrations of measured TC, cumulated heavy metals and WSIs in PM<sub>2.5</sub> from each source  
247 type were showed in Figure 5. Among all source emission and environmental receptor samples, the cumulated heavy metals  
248 from coal combustion was highest and automobile exhaust was higher than ambient PM<sub>2.5</sub>, the overall carbon contents from  
249 automobile exhaust and biomass burning were both higher than ambient PM<sub>2.5</sub>, while only the cumulated soluble ions in PM<sub>2.5</sub>  
250 from primary source of coal combustion was equivalent to the ambient aerosols. In a word, chemical compositions of PM<sub>2.5</sub>  
251 distributed much diversely and varied significantly with the specific source types of combustion emissions.

### 252 **3.3 Cell viability, oxidative stress and inflammation levels exposed to various mass-normalized PM<sub>2.5</sub>**

253 Multiple toxicological endpoints (cell viability, oxidative stress, and inflammation) that facilitate identifying the specific  
254 particle triggering ROS and inflammatory responses resulting in cell death were evaluated for source-specific PM<sub>2.5</sub>. After 24



255 h exposure to the same dose of different PM<sub>2.5</sub> obtained from specific emission sources, the A549 lung cells also showed varied  
256 toxicological responses (Figure 6). The survival rate of cells exposed to automobile exhaust PM<sub>2.5</sub> was much lower than  
257 ambient air PM<sub>2.5</sub> (Figure 6.1). Automobile exhaust PM<sub>2.5</sub> induced the highest ROS production in cells higher than biomass  
258 burning and both sources were also much higher than ambient PM<sub>2.5</sub> (Figure 6.2). Coal combustion induced the highest cellular  
259 IL-6 production followed by automobile exhaust that was also higher than ambient air PM<sub>2.5</sub>, while the PM<sub>2.5</sub> from automobile  
260 exhaust and biomass burning induced similarly higher cellular production of TNF- $\alpha$  than ambient PM<sub>2.5</sub> (Figure 6.3, 6.4).  
261 These results suggested that, combustion primary emission PM<sub>2.5</sub> had stronger ability to induce oxidative stress and  
262 inflammatory injury in lung cells than ambient air PM<sub>2.5</sub>, thus resulted in the higher probability of apoptosis induction (Victor  
263 and Gottlieb, 2002; Wang et al., 2013). Generally, the mass-normalized PM<sub>2.5</sub> from primary source of automobile exhaust  
264 posed the strongest overall toxicity. Therefore, to protect public health by controlling PM<sub>2.5</sub> pollution, the anthropogenic  
265 combustions were key target sources, especially the most toxic automobile PM<sub>2.5</sub> should be reduced preferentially.

### 266 **3.4 Correlations between various PM<sub>2.5</sub> components and toxicity endpoints**

267 Spearman correlation coefficients between chemical compositions and cellular toxicological response indicators were applied  
268 to screen the key components of all PM<sub>2.5</sub> involved in cell injury (Figure 7). It was found that, the degrees of correlations  
269 varied with the toxicological mechanisms of different airborne chemicals. Based on the overall PM<sub>2.5</sub> samples from various  
270 sources, the pro-inflammatory cytokine IL-6 showed significantly strong positive correlations with some heavy metals (As,  
271 Pb, V, Cu), while TNF- $\alpha$  and oxidative stress (ROS) had similar significantly positive correlations with aerosol components  
272 of carbon fractions (EC, OC) and transition metals (V, Cr, Ni). The TNF- $\alpha$  also showed positive correlation with water soluble  
273 Cl<sup>-</sup> and K<sup>+</sup>, and ROS correlated with F<sup>-</sup>, Ca<sup>2+</sup> and Mg<sup>2+</sup>.

274

## 275 **4 Discussion**

### 276 **4.1 New chemical markers for source apportionments of ambient air PM<sub>2.5</sub>**

277 Combustion emissions are key anthropogenic sources contributing to urban air PM<sub>2.5</sub>, through both primary and secondary  
278 aerosols, which were 66% and 34% calculated by PMF model, respectively (Figure 1). The high concentrations of chemical  
279 markers are usually used in source analysis, such as ammonium sulfate and nitrate for secondary aerosols which are originated  
280 mainly from the gaseous precursors (e.g., NH<sub>3</sub>, SO<sub>2</sub> and NO<sub>x</sub>) (Mahilang et al., 2021), the EC, Cu, Mn, and Ni for vehicle  
281 exhaust (Srivastava et al., 2021), the As, Pb, OC, EC, SO<sub>4</sub><sup>2-</sup> and relatively low NO<sub>3</sub><sup>-</sup>/SO<sub>4</sub><sup>2-</sup> ratios for coal combustion (Dai et  
282 al., 2020), soluble K<sup>+</sup> and Cl<sup>-</sup> for plant burning (Jain et al., 2020). The detailed chemical species of these specific source  
283 emission PM<sub>2.5</sub> samples also supported the results. Moreover, low OC/EC ratio of high TC content, high NO<sub>3</sub><sup>-</sup>, F<sup>-</sup>, Na<sup>+</sup>, Ca<sup>2+</sup>  
284 and Mg<sup>2+</sup>, V and Mn of automobile exhaust; Pb and As, SO<sub>4</sub><sup>2-</sup> and NH<sub>4</sub><sup>+</sup> of coal combustion; soluble K<sup>+</sup> and Cl<sup>-</sup>, and high  
285 OC/EC ratio of high TC for plant biomass burning found in current study (Figures 2-5), could also be corresponding potential

286 aerosol source markers. The principal aim of this paper was to assess and contrast the chemical composition and potential  
287 harmfulness of PM arising from diverse anthropogenic sources, thus natural sources, like fugitive soil dust, were not included  
288 in the source examination.

#### 289 **4.2 Common PM<sub>2.5</sub> components related to specific combustion sources**

290 Generally, the automobile exhaust PM<sub>2.5</sub> had high TC content and low OC/EC value with considerable EC content (Figure 2),  
291 varying with specific vehicle types (Figure S4). The contents of the carbon fractions from diesel vehicles were higher than  
292 gasoline exhausts, and the OC/EC ratios of diesel exhausts were much lower than gasoline vehicles, owing to both considerable  
293 contents of EC and OC from diesel vehicle emission PM<sub>2.5</sub>. Some diesel vehicles showed higher EC emissions with age, so  
294 exhaust cleaning devices for them are suggested. In addition, the amounts of OC and EC in exhausts gradually decreased with  
295 the strengthened emission standards they met (Wong et al., 2020). In PM<sub>2.5</sub> samples obtained from coal combustion (Figure  
296 S5), the TC contents of bituminous coals were significantly higher than that of honeycomb coals, anthracite coals, and  
297 industrial coals, because bituminous coals contain higher volatile fraction, which is composed of organic matter. Therefore,  
298 besides the way of combustion and the use of combustion stoves, the coal quality related to different coal types and origins  
299 determine the carbonaceous fractions of the PM emitted by coal combustion (Zhang et al., 2022). In the PM<sub>2.5</sub> samples from  
300 plant biomass combustion (Figure S6), OC contents were generally higher than EC contents, except that pine branches  
301 contained higher EC and rapeseed straw had considerable contents of EC and OC. Dominated by OC (Figure S7), the  
302 concentrations of carbonaceous fractions in urban ambient air samples varied seasonally (Flores et al., 2020; Xu et al., 2019).  
303 Combining the TC contents and OC/EC ratios, carbonaceous components in ambient PM<sub>2.5</sub> mainly originate from semi-volatile  
304 organic compounds (SVOCs) (Wang et al., 2018) and combustion primary emissions for OC (Kang et al., 2018), and  
305 automobile exhaust for EC (Barraza et al., 2017). Thus, to control ambient carbon aerosol pollution, besides reducing the  
306 precursor emissions of secondary organic aerosols (SOA), controlling primary aerosols especially EC from diesel vehicles  
307 were key measures.

308 Airborne redox-active metals are usually linked with the oxidation stress of PM<sub>2.5</sub>. Different types of automobiles emitted  
309 diverse metal contents (Figure S8). Metal elements in automobile exhaust are primarily contributed by fuels, lubricants, and  
310 engine component abrasion. Because Mn is a common antidetonator that delays and prevents the oxidation of hydrocarbons  
311 and increases the octane number, which not only increases the thermal efficiency of the engine but also improves the emission  
312 performance of the vehicle (Cheung et al., 2010), the Mn content was greater in gasoline vehicle exhausts than in diesel  
313 vehicles. Although there are multi-sources of traffic Pb emissions such as fuel combustion and brake wear (Wang et al.,  
314 2019; Panko et al., 2019), the automobile exhaust Pb content of gasoline vehicles were greater than diesel vehicles owing to  
315 oil combustion. Moreover, for the same vehicle type (LDDVs-1 and 2; HDDVs-1 and 2; SDGCs-1 and 2), the stricter the  
316 emission standard required, the lower the exhaust metal contents. The metal contents in the PM<sub>2.5</sub> of trucks was higher than  
317 that of passenger cars (Wu et al., 2016). In the combustion PM<sub>2.5</sub> of 10 coal types (Figure S9), Pb contents were the highest

318 than other heavy metals, similar to available findings (Zhang et al., 2020). The PM<sub>2.5</sub> metals from bituminous coal were  
319 significantly lower than other coal types, because indicated by the coal quality analysis, bituminous coal has a low ash content  
320 which is mainly derived from non-combustible minerals in coal. These findings suggested that coal maturity might be an  
321 important factor influencing the metal composition of particulates emitted from coal combustion (Shen et al., 2021;Zhang et  
322 al., 2021). Heavy metal contents in biomass burned PM<sub>2.5</sub> varied much widely with raw plant types (Figure S10), although  
323 dominated by Cr and Ni. Different plant species and even different plant parts differ significantly in their ability to uptake and  
324 accumulate metals from soil (Zhao et al., 2020). Moreover, because of the high enrichment factors of some metals for crop  
325 straws (Zhang et al., 2016;Sun et al., 2019), they also released more Cr, Ni, and Co during burning than fuelwoods. Total metal  
326 emissions were highest in corn cob but lowest in peanut straw burning PM<sub>2.5</sub>. The heavy metals enriched in urban ambient air  
327 PM<sub>2.5</sub> demonstrated a seasonal pattern (Chen et al., 2018;Hsu et al., 2016) (Figure S11). Contents of V, Co, and As were  
328 relatively low and are less affected by seasonal changes. Accordingly, supported by the metal profiles of anthropogenic  
329 combustion sources and ambient aerosols, to control the environmental airborne heavy metal pollution, the Pb, Cu and As  
330 from honeycomb, anthracite and industrial coal combustion, Cu from vehicle exhausts and especially V from light duty diesel  
331 van with the CN.III emission standard and Mn from gasoline vehicles, Cr and Ni from biomass especially crop straws burning,  
332 should be key targets.

333 Epidemiological studies have also shown the mortality closely related to the WSIs such as sulfate and nitrate in aerosols  
334 (Ostro et al., 2009;Liang et al., 2022). Among the WSIs contents of various automobile exhaust PM<sub>2.5</sub> (Figure S12), NO<sub>3</sub><sup>-</sup> and  
335 Ca<sup>2+</sup> were the most abundant anion and cation, respectively. The high NO<sub>3</sub><sup>-</sup> in the automobile PM<sub>2.5</sub> may be due to NO<sub>x</sub>  
336 production during high-temperature combustion (Hao et al., 2019), while the high Ca<sup>2+</sup> content should be related to additives  
337 in automobile fuels and calcium-based lubricants (Yang et al., 2019). Moreover, the exhaust WSIs decreased with the  
338 strengthened automobile emission standards required. Coal combustion PM<sub>2.5</sub> contained relatively higher SO<sub>4</sub><sup>2-</sup> and NH<sub>4</sub><sup>+</sup>  
339 concentrations followed by Cl<sup>-</sup> than other WSIs species (Figure S13). Among various coal types, industrial coals emitted  
340 highest SO<sub>4</sub><sup>2-</sup> followed by honeycomb and industrial coal with also high NH<sub>4</sub><sup>+</sup>, but bituminous coals emitted low WSIs which  
341 were mainly NO<sub>3</sub><sup>-</sup>, F<sup>-</sup> and Na<sup>+</sup>, Ca<sup>2+</sup>. The WSIs emission factors of honeycomb coal were generally higher than those of lump  
342 coal (Yan et al., 2020). For biomass combustion emissions (Figure S14), Cl<sup>-</sup> and K<sup>+</sup> were dominant WSIs in PM<sub>2.5</sub> from straw-  
343 type fuels (Tao et al., 2016;Sillapapiromsuk et al., 2013), but fuelwood-type combustion emitted high NO<sub>3</sub><sup>-</sup>. Plant species  
344 absolutely determine the emissions (Liao et al., 2021). Finally, there were also high levels of NO<sub>3</sub><sup>-</sup>, SO<sub>4</sub><sup>2-</sup>, and NH<sub>4</sub><sup>+</sup> in ambient  
345 air PM<sub>2.5</sub> (Zhang et al., 2019) (Figure S15). Consequently, implied by the WSIs species distributed in combustion primary  
346 sources and environmental PM<sub>2.5</sub>, to control the aerosols ions pollution, the NO<sub>3</sub><sup>-</sup> from vehicle exhausts and fuelwood burning;  
347 SO<sub>4</sub><sup>2-</sup> and NH<sub>4</sub><sup>+</sup> from honeycomb, anthracite and industrial coal combustion; Cl<sup>-</sup> and K<sup>+</sup> from biomass especially crop straw  
348 burning, should be principal targets, by stricter automobile emission standards or using clean coals.

#### 349 **4.3 PM<sub>2.5</sub> toxicity related to specific sources by pivotal chemical components**

350 The complexity of the sources and compositions of atmospheric PM<sub>2.5</sub> leads to different toxicological effects (Newman et  
351 al., 2020; Kelly, 2021). The toxicological effects of PM<sub>2.5</sub> are not comparable among different studies owing to distinct  
352 exposure concentrations, biological models, endpoints, and PM<sub>2.5</sub> generation methods (Park et al., 2018; Kelly and Fussell,  
353 2020). In this study, we employed same exposure conditions and biological endpoints, in order to obtain comparable toxicity  
354 data for PM<sub>2.5</sub> from different sources. Our mass-normalized results demonstrated that automobile exhaust PM<sub>2.5</sub> induced the  
355 highest lethality and cellular ROS and TNF- $\alpha$  production, coal combustion PM<sub>2.5</sub> induced the highest cellular IL-6 production,  
356 plant biomass burning PM<sub>2.5</sub> induced considerable cellular TNF- $\alpha$  and ROS production (Figure 6). Generally, various toxicities  
357 of combustion emission primary PM<sub>2.5</sub> were much greater than the urban ambient air PM<sub>2.5</sub> (Figure 6), owing to the higher  
358 concentrations of specific toxic components in PM<sub>2.5</sub> from these sources. The supplementary information had included  
359 exhaustive cytotoxicity indicators from each individual source (Figure S16-S19). While the survival rate of cell exposed to  
360 CN.III emission standard PM<sub>2.5</sub> was the lowest and the capacity to induce cells to produce ROS was the highest for CN.IV,  
361 automobile exhaust had a similar potential to cause cells to produce inflammatory cytokines (Figure S16). The capability to  
362 induce IL-6 production in cells was highest for industrial coal PM<sub>2.5</sub>, whereas bituminous coal had the highest survival rate of  
363 cells and TNF- $\alpha$  induction capacity (Figure S17). From the Figure S18 we can see that the PM<sub>2.5</sub> cytotoxicity of straws and  
364 branches burning was analogous, but it should be noted that the cell viability of various straw PM<sub>2.5</sub> differs significantly, that  
365 may be related to the raw fuel characteristics.

366 These possible mechanisms were implied by the overall relationships between the measured chemical components with  
367 cytotoxicity indicators of PM<sub>2.5</sub> from various specific sources (Figure 7). In general, both TNF- $\alpha$  and ROS were significantly  
368 positively correlated with carbonaceous fractions and redox-active transition metals (V, Cr, Ni), which were main contributors  
369 of automobile exhausts and biomass burning. The IL-6 was significantly positively correlated with some heavy metals (As and  
370 Pb, V and Cu), which were main contributors of coal combustion sources. Potential mechanisms include that, carbon fractions  
371 bound in PM<sub>2.5</sub> could be transformed into reactive metabolites and then induce ROS production in cells (Stevanovic et al.,  
372 2019), and the PM<sub>2.5</sub> bound transition metals could also induce ROS production through the Fenton reaction and disrupt the  
373 function of enzymes in cells (Verma et al., 2010; Sørensen et al., 2005; Zou et al., 2016). Oxidative stress can lead to  
374 inflammatory infiltration of neutrophils and stimulate immune cells to produce inflammatory cytokines, among which TNF- $\alpha$   
375 and IL-6 play important roles in the inflammation development (Xu et al., 2020). Ultimately, excessive production of ROS  
376 leads to dysfunctional endoplasmic reticulum responses and dysfunctional lipid metabolism in ROS bursts can result in cell  
377 membrane damage and even cell death (Piao et al., 2018; Zhao et al., 2004). There have been some related supporting reports.  
378 For instance, the OC and EC were significantly associated with biological responses of PM from vehicle emissions collected  
379 in tunnels (Niu et al., 2020). The polar or quinone fractions of PAHs in diesel engine exhaust particles significantly contributed  
380 to the heightened toxic response (Xia et al., 2004). The PM<sub>2.5</sub> generated from biomass burning contained a substantial  
381 concentration of carbonaceous components. In addition, Cr and Ni in PM<sub>10</sub> from straws were highly associated with ROS (Li  
382 et al., 2023). In current study, cellular ROS was also correlated with water soluble Ca<sup>2+</sup>, F<sup>-</sup>, and Mg<sup>2+</sup>, which were main

383 contributors of automobile exhaust PM<sub>2.5</sub>. The Ca<sup>2+</sup> controls the membrane potential and regulates mitochondrial adenosine  
384 triphosphate (ATP) production, and excessive Ca<sup>2+</sup> leads to energy loss and more ROS production (Madreiter-Sokolowski et  
385 al., 2020). Moreover, the TNF- $\alpha$  was also positively correlated with water soluble Cl<sup>-</sup> and K<sup>+</sup>, which were main contributors  
386 of plant burning PM<sub>2.5</sub>. Therefore, the accumulations of some organic matters with high carbonaceous content (OC, EC) in  
387 PM<sub>2.5</sub> typically from automobile exhausts and plant biomass burning, redox-active metals (V, Cr, Ni) and water-soluble anions  
388 (Cl<sup>-</sup>, F<sup>-</sup>) and cations (Ca<sup>2+</sup>, Mg<sup>2+</sup>) contributed by various combustions, might induce ROS production in cells, cause cellular  
389 damage through oxidative stress and inflammatory responses, impair cell viability and finally harm human health.

390 Considering the multi-endpoints measured and the PM<sub>2.5</sub> toxicity mechanisms mentioned above, based on the cell viability  
391 first, and then ROS followed by inflammatory markers, together with the significantly related toxic chemical composition  
392 contents (Park et al., 2018), we put forward a general sequence of overall mass-normalized toxicity for these combustion  
393 source PM<sub>2.5</sub> to managers. To improve the urban environmental air quality for best public health benefits by controlling  
394 aerosols pollution, considering the differential toxicity intensity of each chemical component and their contributions from  
395 various sources to ambient aerosols, preferential targets of specific primary PM<sub>2.5</sub> sources and bound pollutants to be controlled  
396 are suggested as following sequence: Reducing all anthropogenic combustions, especially decreasing the automobile exhaust  
397 PM<sub>2.5</sub> with high contents of EC, transition metals (V, Cu, Ni, Cr), and ions (Ca<sup>2+</sup>, Mg<sup>2+</sup>, F<sup>-</sup>, Na<sup>+</sup>) from diesel exhausts by  
398 strengthening the emission standards, then lessening the coal combustion with high heavy metals (As, Pb, Cu) by replacement  
399 with low-ash clean coals, and depressing the biomass burning with high OC, Ni, Cr, Cl<sup>-</sup> and K<sup>+</sup> from crop straw emissions.

#### 400 **4.4 Limitations and perspectives**

401 In current study, we selected A549 cell based on previous abundant experimental experiences and also because it has been  
402 used popularly in *in vitro* toxicology studies to elucidate the cellular and molecular mechanisms of PM involved in lung for  
403 many decades (Li et al., 2022b). However, recently the human normal bronchial epithelial cell BEAS-2B was preferred over  
404 the human lung adenocarcinoma epithelial cell A549. For instance, both cells were used in an aerosol study (Bonetta et al.,  
405 2017), results of which highlighted the higher sensitivity of BEAS-2B cells respect to A549 also in samples with low level of  
406 pollutants, because the PM<sub>0.5</sub> samples from Italian towns can induce genotoxicity in normal cells while cancer cells might be  
407 resistant to their adverse effects. Therefore, although our results are reasonable under the same exposure conditions, there were  
408 still potential limitations of A549 cells since they may be more resistant to exposure to external compounds, and the generally  
409 more sensitive BEAS-2B cells are suggested for future studies.

410 In toxicity assessments, cell vitality reflects the overall health of cells, encompassing factors such as cell membrane integrity,  
411 intracellular metabolic activity, and cell proliferation capacity. Decreased cellular vitality may be associated with cell damage,  
412 toxic effects, or cellular apoptosis. Inflammation markers are employed to assess the extent and nature of inflammatory  
413 reactions, including the production of cytokines and inflammatory mediators, as well as the activation status of inflammatory  
414 cells. Inflammation is a complex physiological response, typically delineated by the immune and inflammatory reactions of

415 the body to stimuli such as injury or infection. Alterations in inflammation markers can indicate the intensity and nature of the  
416 inflammatory response. In this study, multiple biological responses of epithelial cells to various PM<sub>2.5</sub> were evaluated,  
417 including that, cell viability evaluated the mitochondrial dehydrogenase activity of the living cells, excessive intracellular ROS  
418 formation induced by PM<sub>2.5</sub> was responsible for oxidative stress to the cells, cytokines IL-6 and TNF- $\alpha$  were determined for  
419 the effect of PM<sub>2.5</sub> on pro-inflammatory response in cells. In general, *in vitro* data can be used to rank various types of particles  
420 in terms of the toxic potential including possible carcinogenicity. Each marker will help to understand the hazard and toxicity  
421 of PM<sub>2.5</sub>. However, the toxicity of PM<sub>2.5</sub> may be the result of multiple components acting through disparate physiological  
422 mechanisms, with inconsistent relationships among endpoints (Park et al., 2018). For instance, in BEAS-2B cells, oxidative  
423 stress generated by H<sub>2</sub>O<sub>2</sub> exposure often results in cytotoxicity rather than by stimulating cytokine/chemokine responses,  
424 sometimes no correlation between oxidative damage and cytokine/chemokine responses. Moreover, TNF- $\alpha$  gene was not  
425 detected in BEAS-2B cells exposed to atmospheric PM collected from Benin, but the gene expression of other inflammatory  
426 cytokines (IL-1 $\beta$ , IL-6, and IL-8) were significantly induced, and decreasing cell viability was highly correlated with high  
427 secretion of all studied cytokines (Cachon et al., 2014). Therefore, in the present study, it was impossible to analyze all  
428 chemicals in PM<sub>2.5</sub> and determine all related toxicological endpoints, so unmeasured chemicals and endpoints might also play  
429 roles in the incongruous or unexplained results, and we also can't over-explain the mechanisms just based on statistical  
430 relations. To overcome these hurdles, standardization of toxicological studies (experimental methodologies) and reporting  
431 guidelines are necessary for tracking and comparing results.

432 This study ranked the unequal "toxic effects" based on the same mass concentration of PM<sub>2.5</sub> exposure in body lung fluid  
433 system, while the "health risks" usually relating to the inhalation exposure concentration of PM<sub>2.5</sub> in ambient air were not  
434 calculated and evaluated quantitatively. Moreover, non-linear concentration-response functions for various endpoints and  
435 different exposure concentrations might also limit using toxicological data straightforwardly to predict health effects  
436 (morbidity, mortality) in human populations, so drawing conclusions precisely quantifying/ranking the health risks of PM<sub>2.5</sub>  
437 from specific sources or of individual PM<sub>2.5</sub> components is still not an easy task (Kelly and Fussell, 2020). Therefore, coupled  
438 with source apportionment and exposure level of ambient aerosols pollution, toxicology combined with epidemiology studies  
439 are essential for linking these factors and understanding scientific mechanisms to reach conclusions.

440

## 441 **5 Conclusions**

442 In current study, we found that 2/3 mass of urban ambient air PM<sub>2.5</sub> in a representative megacity of eastern China originated  
443 from primary sources of anthropogenic combustions including coal, automobile, and biomass. Because of the significant  
444 differences in the chemical compositions, the diverse PM<sub>2.5</sub> from both mixed ambient air and directly from individual  
445 combustion sources showed much differential mass-normalized *in vitro* toxicity to the human lung epithelial cells, either for  
446 the environmental aerosol samples collected from different seasons, or for the primary emissions of PM<sub>2.5</sub> from various specific

447 source types. According to the comparative study and correlation analysis, the carbonaceous fractions (OC, EC) and redox-  
448 active heavy metals (V, Ni, Cr) assisted by water-soluble ions ( $\text{Ca}^{2+}$ ,  $\text{Mg}^{2+}$ ,  $\text{F}^-$ ,  $\text{Cl}^-$ ) might play important roles in inducing  
449 cellular ROS production, causing oxidative stress and inflammation, resulting in cell injury and apoptosis, thus damage human  
450 health. These toxic pollutants accumulated in specific-source  $\text{PM}_{2.5}$  varied by the emission types and raw fuel properties.  
451 Combined with chemical composition and general cytotoxicity rank, the preferential controlling targets of specific combustion  
452 sources should be automobile exhaust (diesel vehicles with emission standards inferior to CN.IV), coal combustion (high ash  
453 and high sulfur coals), and plant biomass burning (crop straws). Although showing the synthetic effects of mixed compositions  
454 and complex sources, besides preventing the secondary aerosols from combustions, preferentially targeted reductions of these  
455 primary sources of toxic  $\text{PM}_{2.5}$  direct emissions, would produce the greatest benefits for public health with improved ambient  
456 air quality. Overall, this paper provides a precise, oriented, effective, efficient, and economical composition-source-based  
457 strategies for urban aerosols pollution control. However, as a prospect, the detailed mechanisms for unequal toxicity of PM  
458 with complicated components from various sources and their quantitative contributions to the health effects of ambient air  
459  $\text{PM}_{2.5}$  mixture still need in-depth study.

#### 460 **Supplementary materials**

461 There are 19 figures (Figure S1-S19) and 3 tables (Table S1-S3) in the Supporting Information.

#### 462 **Data availability**

463 All raw data can be provided by the corresponding authors upon request.

#### 464 **Author contributions**

465 XSL conceived and supervised the study; WH, YP, MT, HL, and ZZ collected the samples; WH, YP, MT, WL, HL, ZZ, GS,  
466 and LX analyzed the chemical compositions; WH, YP, and MT performed the toxicity tests; WH, YP, MT, and XSL analyzed  
467 the data; WH and XSL wrote the manuscript draft; XSL, WH, GS, and TM reviewed and edited the manuscript.

#### 468 **Competing interests**

469 The authors declare that they have no conflict of interest.

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472

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## 729 **Captions of figures**

730 **Figure 1.** The PMF factor profiles of various components and source percentages of secondary aerosol, automobile exhaust,  
731 coal combustion, and plant biomass burning contributing to the urban ambient air PM<sub>2.5</sub>.

732 **Figure 2.** Carbon contents (mg kg<sup>-1</sup>) and ratio in PM<sub>2.5</sub> from various specific sources (n=10 for each combustion source and  
733 n=16 for urban ambient air).

734 **Figure 3.** Heavy metal contents (mg kg<sup>-1</sup>) in PM<sub>2.5</sub> from various specific sources (n=10 for each combustion source and n=16  
735 for urban ambient air).

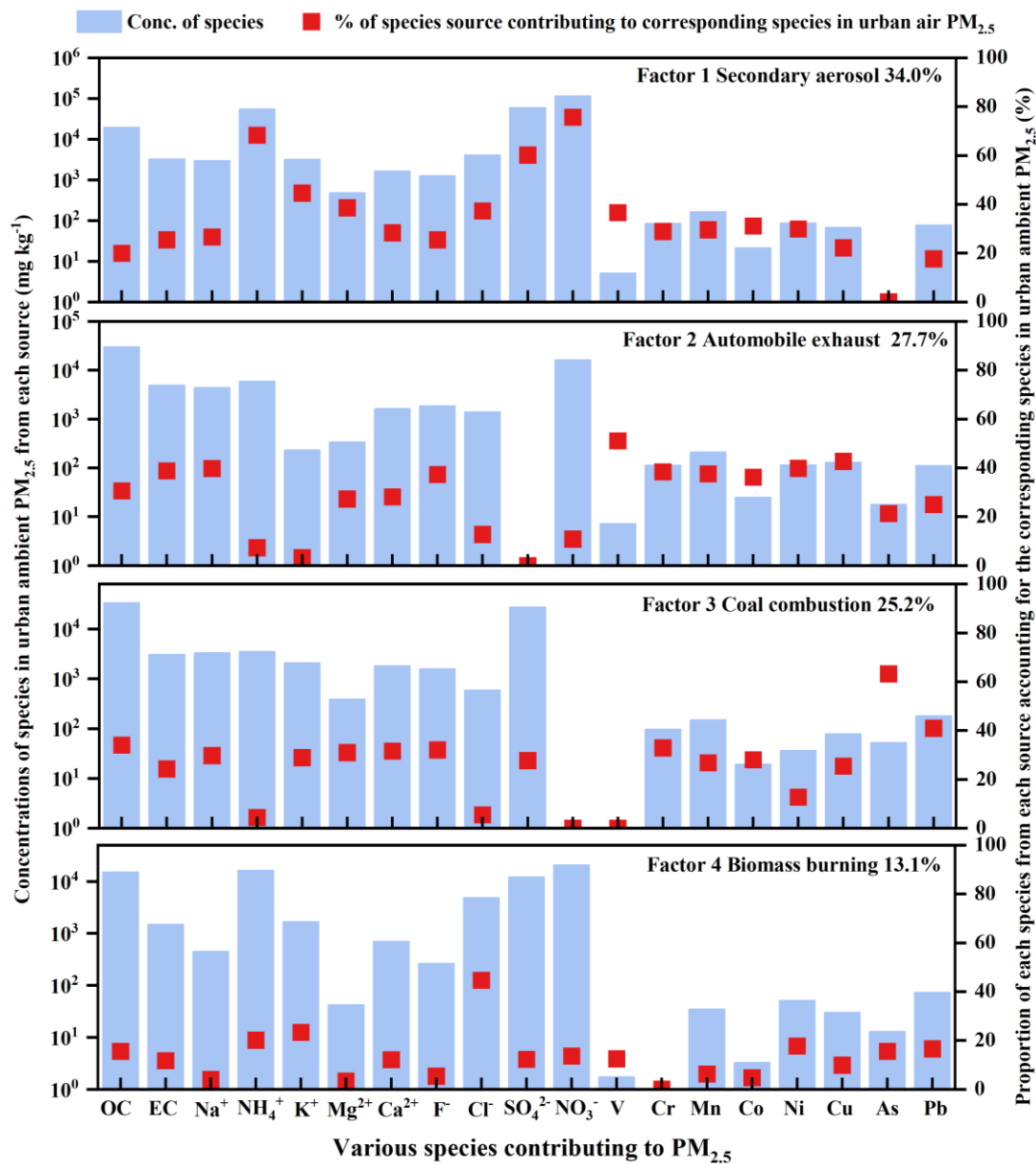
736 **Figure 4.** Water-soluble ion (WSI) contents (mg kg<sup>-1</sup>) in PM<sub>2.5</sub> from various specific sources (n=10 for each combustion  
737 source and n=16 for urban ambient air).

738 **Figure 5.** Cumulated typical measured components (mg kg<sup>-1</sup>) in PM<sub>2.5</sub> from various specific sources (n=10 for each  
739 combustion source and n=16 for urban ambient air).

740 **Figure 6.** Cell viability, oxidative stress and inflammation levels of human alveolar epithelial cell lines (A549) exposed to  
741 PM<sub>2.5</sub> suspension (80 mg L<sup>-1</sup>) from various specific sources (n=10 for each combustion source and n=16 for urban ambient  
742 air).

743 **Figure 7.** Overall correlations between typical cellular toxicological responses and chemical compositions of PM<sub>2.5</sub> from  
744 various sources (\*p < 0.05, #p < 0.01; n=46).

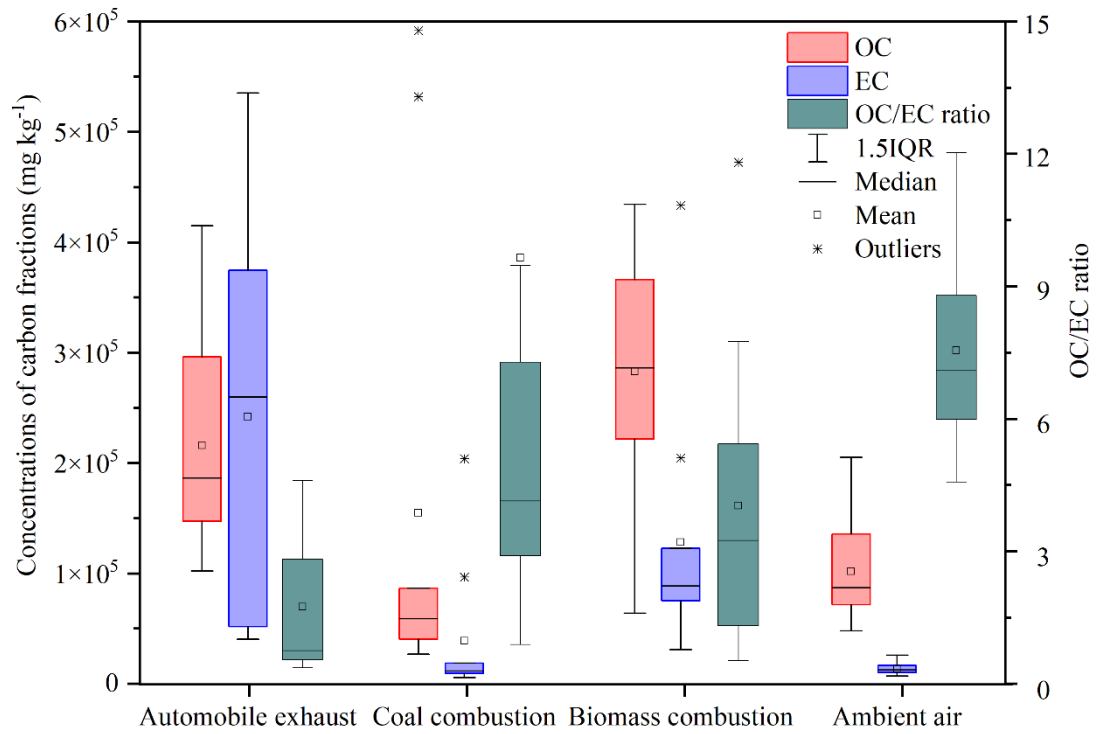
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 749 coal combustion, and plant biomass burning contributing to the urban ambient air  $PM_{2.5}$ .

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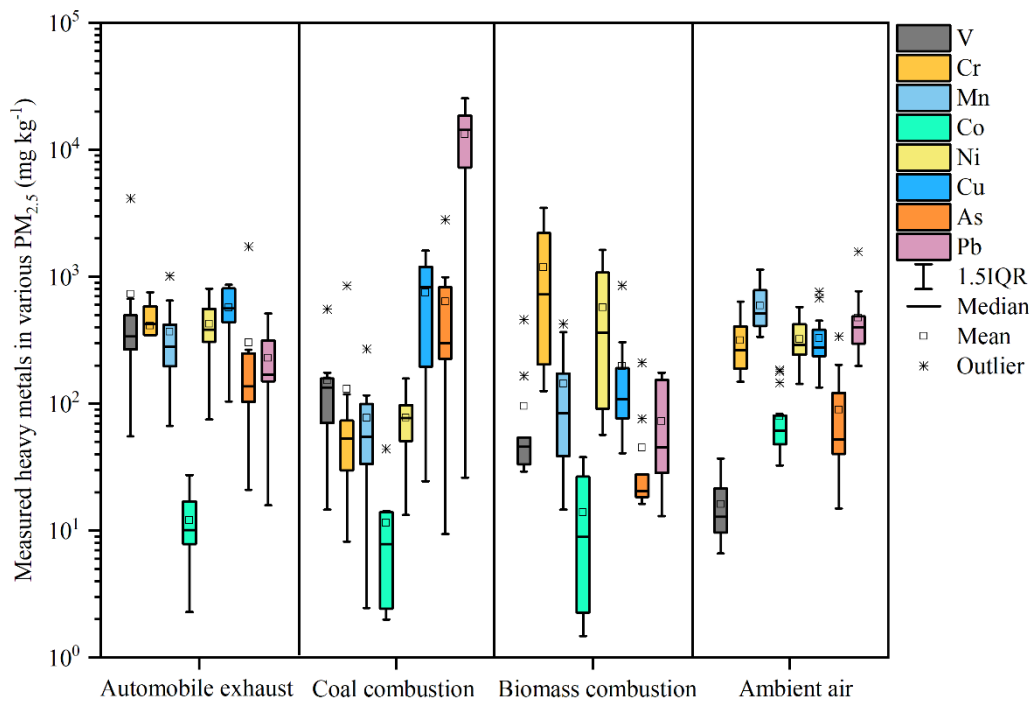


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753 **Figure 2.** Carbon contents (mg kg<sup>-1</sup>) and ratio in PM<sub>2.5</sub> from various specific sources (n=10 for each combustion source and  
 754 n=16 for urban ambient air).

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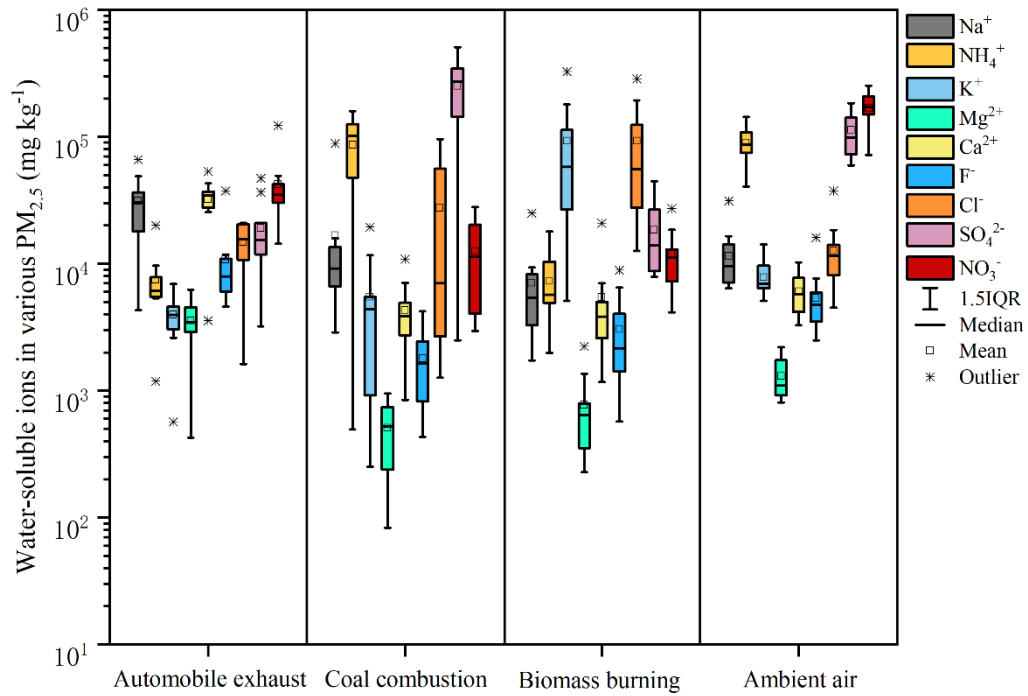


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757 **Figure 3.** Heavy metal contents (mg kg<sup>-1</sup>) in PM<sub>2.5</sub> from various specific sources (n=10 for each combustion source and

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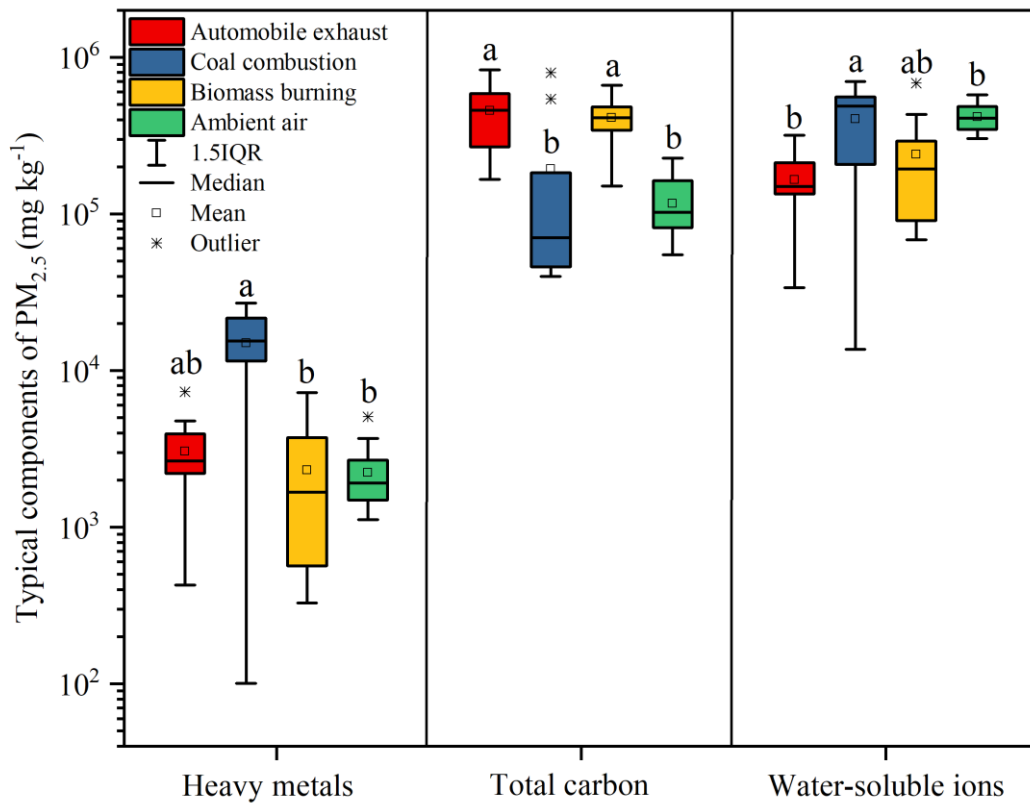
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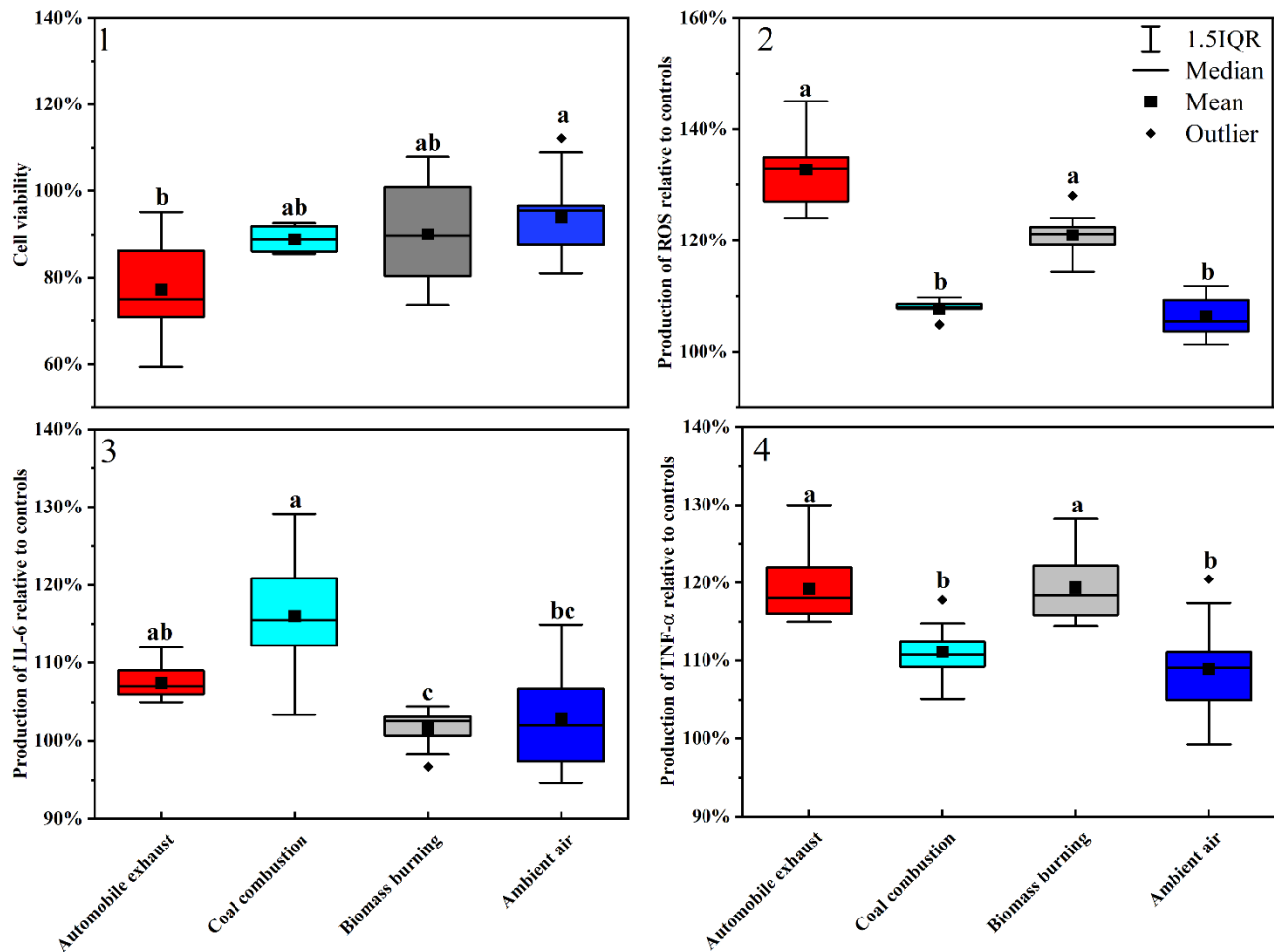
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766 **Figure 5.** Cumulated typical measured components (mg kg<sup>-1</sup>) in PM<sub>2.5</sub> from various specific sources (n=10 for each  
 767 combustion source and n=16 for urban ambient air). The letters a and b are significant groups classified by Kruskal–Wallis  
 768 test, p < 0.05.

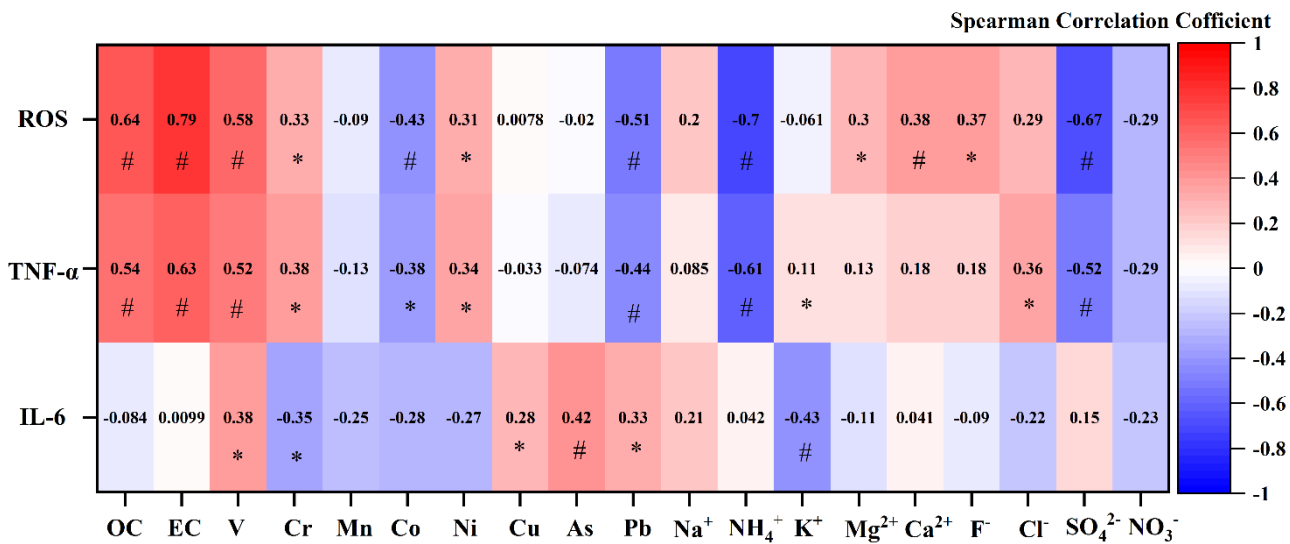


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