

# Source differences in the components and cytotoxicity of PM<sub>2.5</sub> from automobile exhaust, coal combustion, and biomass burning contributing to urban aerosol toxicity

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**Abstract.** Although air quality guidelines generally use the atmospheric concentration of fine particulate matter (PM<sub>2.5</sub>) as the metric for air pollution evaluation and management treating all particles as equally toxic, it is inconsistent with the facts that particle toxicity are significantly related to their sources and chemical compositions. Therefore, judging the most harmful source and identifying the toxic component will be extremely helpful to optimize air quality standards and prioritize targeted PM<sub>2.5</sub> control strategies to more protect public health effectively. The combustions of fuels, including oil, coal, and biomass, are main anthropogenic sources of environmental PM<sub>2.5</sub>, however, their discrepant contributions to health risks of mixed ambient aerosol pollution dominated by respective emission intensity and unequal toxicity of chemical components are still unclear. In order to quantify the differences among these combustion primary emissions, ten types of PM<sub>2.5</sub> from each typical source group, i.e., vehicle exhaust, coal combustion, and plant biomass burning, were collected for comparative study with toxicological mechanisms. Totally thirty type individual combustion samples were inter-compared with representative urban ambient air PM<sub>2.5</sub> samples, which chemical characteristics and biological effects were investigated by component analysis (carbon, metals, soluble ions) and in vitro toxicity assays (cell viability, oxidative stress, inflammatory responses) of human lung adenocarcinoma epithelial cells (A549). Carbonaceous fractions were plenteous in automobile exhaust and biomass burning, while heavy metals were more plentiful in PM<sub>2.5</sub> from coal combustion and automobile exhaust. The overall ranking of mass-normalized cytotoxicity for source-specific PM<sub>2.5</sub> was automobile exhaust > coal combustion > plant biomass burning > ambient urban air, possibly with differential toxicity triggers, that the carbonaceous fractions (organic carbon, OC; elemental carbon, EC) and redox-active transition metals (V, Ni, Cr) assisted by water-soluble ions (Ca<sup>2+</sup>, Mg<sup>2+</sup>, F, Cl) might play

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52 important roles in inducing cellular reactive organic species (ROS) production, causing oxidative stress and inflammation,  
53 resulting in cell injury and apoptosis, thus damage human health. Coupled with the source apportionment results of typical  
54 urban ambient air PM<sub>2.5</sub> in eastern China, reducing toxic PM<sub>2.5</sub> from these anthropogenic combustions will be greatly beneficial  
55 to public health, especially preferentially decreasing the diesel exhaust by strengthening emission standards, then lessening the  
56 coal combustion by replacement with low-ash clean coals, and depressing the crop straw burning emissions.

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## 58 1 Introduction

59 As a mixture of multiple sources, ambient particulate matter, (PM) arise from anthropogenic activities are continuously  
60 deteriorating the urban air quality, particularly in developing countries. Among these, fine PM<sub>2.5</sub> with an aerodynamic diameter  
61 of less than 2.5 μm (PM<sub>2.5</sub>) is recognized as a serious public health concern due to its long persistence in air, carcinogenicity  
62 and acute toxicity to humans (Al-Kindi et al., 2020). There were extensive epidemiological evidences that airborne PM<sub>2.5</sub> can  
63 cause serious negative effects on human health, such as respiratory and cardiovascular diseases, genetic mutations, and  
64 developmental disorders (Chowdhury et al., 2022; Lelieveld et al., 2021; Smith, 2021; Clemens et al., 2017). Currently, either  
65 the world air quality guidelines or the national air quality standards use the mass concentration of PM<sub>2.5</sub> as the metric for PM<sub>2.5</sub>  
66 pollution evaluation and management, in which all particles are treated as equally toxic, however, it is inconsistent with the  
67 scientific facts that particle toxicity are significantly related to their sources and chemical compositions (Shiraiwa et al., 2017).  
68 Therefore, to identify which component(s) and source(s) of ambient PM are most harmful to health, will be very helpful to  
69 optimize air quality guidelines/standards and prioritize targeted PM control strategies to more effectively protect public health  
70 (Kelly and Fussell, 2020).

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71 Besides natural sources like dust and sea spray, the vast majority of aerosols come from anthropogenic activities especially  
72 energy consumption, including the combustion of fossil fuels causing industrial emissions and automobile exhaust, and  
73 biomass burning (McDuffie et al., 2021; Wu et al., 2022). Finally, these diverse sources make the ambient air PM<sub>2.5</sub> become a  
74 complex mixture with multiple chemical components varying with time and space, which consisting mainly of sulfate, nitrate,  
75 ammonium, organic carbon (OC), elemental carbon (EC), mineral and trace metals (Bari and Kindzierski, 2016; Kelly and  
76 Fussell, 2020). The physiological mechanisms of PM-induced cell toxicity in respiratory system have been continuously  
77 investigated with some progresses (Kelly and Fussell, 2012, 2020; Shiraiwa et al., 2017; Mack et al., 2020; Li et al., 2022b),  
78 such as the metabolic activation, oxidative stress, inflammatory response, and apoptosis, focused on by current study. In brief,  
79 after inhalation and deposition onto the epithelium, redox-active materials in PM<sub>2.5</sub> can induce the release of reactive organic  
80 species (ROS), which cause oxidative stress (an imbalance between ROS and antioxidants, i.e., disequilibrium of the redox  
81 state of a cell) followed by inflammation and cell death. The ROS can mediate subsequent signaling pathways leading to  
82 biomolecule damage (e.g., DNA, lipid, and protein) and cellular injury, through mediating inflammatory responses including  
83 the release of pro-inflammatory cytokines like IL-6 and TNF-α by epithelial cells (Sabbir Ahmed et al., 2020; Landwehr et al.,

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2021). For instance, oxidative stress could trigger the induction of pro-inflammatory transcription factors, such as nuclear factor (NF)- $\kappa$ B, via the mitogen-activated protein kinase (MAPK) signaling pathway. Components adsorbed on particle surface, such as redox-active metals (transition metals, Fe, Ni, V, Cr, Cu), organic compounds (polycyclic aromatic hydrocarbons, PAHs; quinones), or even carbonaceous core of particles, are responsible for oxidative stress (Cachon et al., 2014; Sabbir Ahmed et al., 2020). The non-redox active metals (Zn, Pb, Al) can also influence the toxic effects of transition metals by exacerbating or lessening the production of free radicals. The EC may not be a directly toxic component of PM<sub>2.5</sub> but rather operate as a universal carrier of combustion-derived chemicals (semi-volatile organic fractions, transition metals) of varying toxicity (Kelly and Fussell, 2020). Inorganic soluble sulphates and nitrates are acidic and can interact with and influence the solubility other compositions like metal bioavailability (Fang et al., 2017; Weber et al., 2016). However, which specific components and which particular sources are the most critical factors dominating the ambient aerosols' health risks, still leave puzzles unsolved.

Past studies performed in various countries have focused on physicochemical characterization or biological effects of ambient air PM<sub>2.5</sub> respectively (Weagle et al., 2018; Jia et al., 2017; Wang et al., 2020). For example, the source analysis of PM<sub>2.5</sub> by photochemical modeling (Bao et al., 2018), chemical composition of regional PM<sub>2.5</sub> (Chi et al., 2022), and the mechanism of PM<sub>2.5</sub> toxicity was independently reported recently (Jia et al., 2020). Because differences in particle composition, sources, and toxicity appear in different urban environments (Zhao et al., 2019; Borlaza et al., 2018), the source profiles of different emission inventories were applied to elucidate aerosol pollution characteristics and control strategies. For instance, it was found that straw burning during the harvest season is a major trigger of severe air pollution in many regions (Sahu et al., 2021). Aerosols from open biomass burning in the Amazon had a stronger ability to induce ROS than laboratory-generated secondary organic aerosols (Tuet et al., 2019). The particle composition of motor vehicle exhaust was related to automobile types with various fuels, engines, and loads (Lin et al., 2020). A strong catalytic reactivity of metals in PM<sub>v</sub> emitted from diesel vehicles was observed by dithiothreitol (DTT) assay (Jesus et al., 2018). Sulfate is a major component of PM<sub>v</sub> from Xi'an city, western China, mainly released from residential coal combustion activities (Dai et al., 2019). Traffic was suggested playing the most crucial role in enhancing the toxicity of fine particles (Park et al., 2018). Although there were emerging studies on particle emission from single source, quantitatively comparative studies on multi-source pollutants as well as the differential composition and unequal toxicity of various sources are still limited.

The main objective of current study was to compare the chemical components and corresponding mass-normalized toxicological effects of individual PM<sub>2.5</sub> from various combustion sources and their unequal contributions to ambient aerosol health risks. The aim is to provide detailed guidance on the targeting and precise control of specific anthropogenic sources with prominent risks based on their pivotal toxic components. Therefore, we collected both representative ambient PM<sub>2.5</sub> samples from urban air and abundant typical source PM<sub>2.5</sub> samples from automobile exhaust, coal combustion, and plant biomass burning. Their independent profiles of chemical compositions and *in vitro* cytotoxicity (cell viability, oxidative stress, and inflammatory responses) were investigated and intercompared, to assess the differences in source-to-receptor toxicity and

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162 to infer the core toxic components and respective harmful contribution. The pivotal toxic components were identified based  
163 on the source-sink bi-directional composition-effect results, which were further used to assess the health toxicity contribution  
164 of various emission sources to ambient air PM<sub>2.5</sub>, supported by its source apportionment through positive matrix factorization  
165 (PMF) model. This study could advance the understanding to quantify the complex source contribution to high-risk PM<sub>2.5</sub>  
166 emission oriented to public health, which is imperative for precise prevention and control of atmospheric PM<sub>2.5</sub> pollution.  
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## 168 2 Materials and methods

### 169 2.1 Collection of PM<sub>2.5</sub> samples from primary emissions of 30 typical combustion sources and from representative 170 ambient urban air

171 As the main anthropogenic sources of the ambient air PM<sub>2.5</sub> pollution, totally 30 types of primary PM<sub>2.5</sub> samples emitted  
172 directly from automobile exhaust, coal combustion, and plant biomass burning were respectively collected as follows for both  
173 chemical and toxicological analyses.

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174 Based on the classification of automobile fuel types as well as load and tailpipe emission standards provided by the 2019  
175 Annual Report on Environmental Management of Mobile Sources in China, a total of 10 types of vehicles were chosen for  
176 exhaust investigation. They were further categorized into 7 sub-groups, including small duty gasoline coaches (SDGCs), small  
177 duty diesel coaches (SDDCs), middle duty diesel coaches (MDDCs), heavy duty diesel coaches (HDDCs), light duty diesel  
178 vans (LDDVs), middle duty diesel vans (MDDVs), and heavy duty diesel vans (HDDVs). The detailed information of these  
179 representative local automobiles was showed in Table S1.

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180 To cover all coal types consumed in the city, 10 representative types of coal were gathered for investigation. They were  
181 further classified into 4 sub-groups, including 2 types of honeycomb coal (HC), 3 types of anthracite coal (AC), and 2 types  
182 of bituminous coal (BC) mainly for restaurant or household use, and 3 types of industrial coal (IC) for coal-fired power plants  
183 and steel-smelting industry. The detailed characteristic analysis of these typical coals purchased from local market were  
184 showed in Table S2.

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185 Considering the plant biomass combustion in rural areas surrounding the megacity, 10 representative types of agricultural  
186 and forestry solid wastes were gathered for investigation. Because of the high annual production of three staple food crops  
187 (rice, wheat, and corn) as well as soybean, peanut and rapeseed, their straws generated during harvest are often used as fuels  
188 in rural households. In addition, woods were also common fuels. Therefore, straws of rice, wheat, corn, soybean, peanut, rape,  
189 and sesame, corncob, branches of peach and pine, were selected as plant biomass fuels and further divided into 2 sub-groups,  
190 including 8 types of crop straw and 2 types of firewood. The detailed characteristic analysis of these typical plant biomass  
191 fuels collected from rural areas around Nanjing city were showed in Table S3.

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192 The PM<sub>2.5</sub> samples directly emitted from these combustion sources were collected by dilution channel sampling method  
193 (Figure S1), using a 4-channel particulate matter dilution sampler (HY-805, Hengyuan Technology Development Co., CN).

209 Each sampling included 3 parallel channels of quartz microfiber filter (Figure S2) and 1 channel of Teflon membrane filter,  
210 with diameters of 47 mm, through a size selector for PM<sub>2.5</sub> with a flow rate of 160 L min<sup>-1</sup>. Clean air was pumped for 10 min  
211 before and after each sample was collected. Before using the blank quartz filters were incinerated by a muffle furnace at  
212 500 °C for 3 h to remove any possible organic matters, while Teflon filters were baked at 60 °C for 4 h. After being equilibrated  
213 in a constant temperature and humidity chamber for 24 h, the filters were weighed both before and after sampling for  
214 gravimetric measurements, then the mass of collected PM<sub>2.5</sub> could be calculated. The sampled filters were stored in a  
215 refrigerator at -20 °C before analysis. The quartz filter loaded PM<sub>2.5</sub> samples were used for carbon and ion analysis, and for  
216 toxicity tests, while the parallel Teflon filter loaded samples were used to determine metals.

217 As the actual mixture of various source particles in real environment, totally 16 representative ambient air PM<sub>2.5</sub> samples  
218 (each time lasting 23h) covering a year monthly were collected from December 2019 to October 2020 in an urban site  
219 surrounded by traffic, residential and commercial quarters of Nanjing city, Yangtze River Delta of eastern China, using a high-  
220 volume air sampler (800 L min<sup>-1</sup>) with quartz microfiber filters. Detailed procedures and sample information were described  
221 in previous paper (Li et al., 2022a), but the purpose of using these air samples in current study was to compare them with the  
222 specific source samples for evaluating the chemical and toxicological contributions of the combustion primary sources to  
223 environmental aerosols pollution.

## 224 2.2 Chemical composition analysis

225 All collected source and ambient PM<sub>2.5</sub> samples were conducted various component analysis (Li et al., 2023). For the  
226 concentrations of heavy metals in particulates, samples were digested by concentrated HNO<sub>3</sub>-HClO<sub>4</sub> acids with a progressive  
227 heating program and determined by inductively coupled plasma optical emission spectrometry (ICP-OES; Optima8000,  
228 PerkinElmer), with some elements at lower concentrations measured by ICP mass spectrometry (ICP-MS; NexIONTM300X,  
229 PerkinElmer). Blank filter, reagent blank, replicates, and standard reference material (NIST SRM 1648a, urban dust) were  
230 adopted for analytical quality control, with recoveries ranged 90-110 %. Carbonaceous species (OC and EC) in PM<sub>2.5</sub> were  
231 determined using a DRI-2001A OC/EC (Atmoslytic Inc., Calabasas, CA, USA). For the concentrations of water-soluble ions  
232 (WSIs), the main cations (Na<sup>+</sup>, K<sup>+</sup>, Mg<sup>2+</sup>, Ca<sup>2+</sup>, NH<sub>4</sub><sup>+</sup>) and anions (NO<sub>3</sub><sup>-</sup>, SO<sub>4</sub><sup>2-</sup>, Cl<sup>-</sup>, F<sup>-</sup>) in PM<sub>2.5</sub> were measured by ion  
233 chromatography (IC, Thermo Fisher Scientific, USA), using the Metrosep C6-150/4.0 column for cations and the Metrosep A  
234 Supp 5 150/4.0 column for anions, respectively.

## 235 2.3 Preparing mass-normalized PM<sub>2.5</sub> suspension for cell exposure

236 Totally 30 source and 16 ambient PM<sub>2.5</sub> samples were also performed cytotoxicity tests. In order to elute the particles  
237 completely from the quartz membranes, the PM<sub>2.5</sub>-loaded sample filter was cut into small pieces, immersed in ultrapure water  
238 and extracted six times (30 min for each) in an ultrasonic bath at 0 °C. Although the ultrasonication might impact the ROS  
239 (Miljevic et al., 2014), the inevitable systematical error was ignored in this study. The extract was then suction filtered through  
240 a 2.6 μm pore-size nylon membrane to remove possible quartz fragments, and the bulk filtrate was freeze-dried back to pure

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255 PM<sub>2.5</sub> powder. Ultimately, based on particle mass, the gathered PM<sub>2.5</sub> was dispersed by sterile phosphate-buffered saline (PBS)  
256 to a concentration of 400 mg L<sup>-1</sup>, and then diluted to PM<sub>2.5</sub> suspension of 80 mg L<sup>-1</sup> with serum-free Dulbecco's modified eagle  
257 medium (DMEM) medium for following *in vitro* cell exposure (Li et al., 2022a).

#### 258 2.4 Cell culture and cellular toxicity tests by *in vitro* PM<sub>2.5</sub> exposure

259 ~~Aerosol pollution can harm lung alveoli and epithelial cells, and the~~ A549 human lung adenocarcinoma epithelial cell ~~has long~~  
260 ~~been used as a suitable epithelial alveolar model to investigate the interactions between PM and lung epithelial cells (Park et~~  
261 ~~al., 2018; Li et al., 2022b).~~ The A549 cells were cultured in RPMI-1640 medium (Gibco, USA) supplemented with 10% fetal  
262 bovine serum (FBS, Hyclone, USA) and 1% antibiotic penicillin-streptomycin (100 U mL<sup>-1</sup>) at 37 °C in a 5% CO<sub>2</sub> incubator.  
263 ~~After PM<sub>2.5</sub> exposure, cell viability and the indicators reflecting oxidative damage and inflammatory responses were~~  
264 ~~determined respectively. While the cell viability assay was helpful in determining PM<sub>2.5</sub> dose to cells, the endogenous ROS~~  
265 ~~measurements revealed the status of cellular oxidative potential after PM<sub>2.5</sub> exposure followed by the relative effects of ROS~~  
266 ~~on various stages of cellular toxicity like inflammatory responses (Gali et al., 2019). The cell viability (metabolic activity) was~~  
267 ~~evaluated by mitochondrial activity and determined by the methyl-thiazol-tetrazolium (MTT) assay (Chen et al., 2019). After~~  
268 trypsin action, the density of cells in the logarithmic growth phase was adjusted to 1 × 10<sup>5</sup> mL<sup>-1</sup>. Cell suspensions were  
269 inoculated into 96-well plates (Costar, USA) at 100 μL per well. The blank control well (without medium and PM<sub>2.5</sub> suspension)  
270 and reagent control well (with medium but without PM<sub>2.5</sub> suspension) were set together. After incubation for 24 h ~~and removing~~  
271 ~~the cellular supernatant~~, various types of PM<sub>2.5</sub> suspension (concentration of 80 mg L<sup>-1</sup>) were added to 96-well plates and  
272 incubated for 24 h. ~~Based on pre-experiments, the oxidative stress and inflammation response sensitively under this dose,~~  
273 ~~while the cell viability can keep sufficient.~~ Fresh medium and MTT reagent (Solarbio, Beijing, CN) were added to each well  
274 and the supernatant was discarded, then 100 μL of formazan lysate was added to each well. The optical density (OD) values  
275 were measured at 490 nm using a microplate reader (Thermo MULTISKAN FC, USA). Cell viability (%) = (OD<sub>treatment</sub> –  
276 OD<sub>blank control</sub>) / (OD<sub>reagent control</sub> – OD<sub>blank control</sub>). The ~~levels of cellular ROS production causing oxidative stress in cells, pro-~~  
277 ~~inflammatory cytokines including tumor necrosis factor-alpha (TNF-α) and interleukin-6 (IL-6) production for determining~~  
278 ~~the expression of genes related to the inflammatory response~~ in the supernatant were analyzed by enzyme-linked  
279 immunosorbent assay (ELISA) kits (Jiangsu Enzyme Biotechnology Co., Ltd., CN), and OD values were measured at 450 nm  
280 (Huang et al., 2020; Pang et al., 2020).

#### 281 2.5 Data analysis

282 The statistical analysis was performed by IBM SPSS statistics 24 and plotted by Origin 2020b software. ~~Spearman~~ correlation  
283 coefficients were produced by the correlation analysis. The variance was statistically significant when the statistical test level  
284 was  $p < 0.05$ , and extremely significant when  $p < 0.01$ . ~~Statistical analyses were performed using Kruskal–Wallis test~~ (Kruskal  
285 and Wallis, 1952).

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298 The source apportionment of PM<sub>2.5</sub> mass in urban ambient air was conducted by the receptor model PMF (EPA PMF version  
299 5.0). Major constituents (OC, EC, Cu, Cr, Co, Ni, As, Pb, Mn, V, Na<sup>+</sup>, K<sup>+</sup>, Mg<sup>2+</sup>, Ca<sup>2+</sup>, NH<sub>4</sub><sup>+</sup>, Cl<sup>-</sup>, F<sup>-</sup>, NO<sub>3</sub><sup>-</sup>, and SO<sub>4</sub><sup>2-</sup>) were  
300 selected as input data, and a four-factor solution was chosen as the optimal solution based on an assessment of the  
301 interpretability of the source profiles and the seasonal variability of the source contributions.

302

### 303 3 Results

#### 304 3.1 Contributions of combustion primary sources to urban ambient air PM<sub>2.5</sub>

305 As shown in Figure S3, although have been significantly improved with the national air quality in recent years, the daily PM<sub>2.5</sub>  
306 concentrations of representative city Nanjing still exceeded the healthy guidelines obviously, with higher urban air PM<sub>2.5</sub>  
307 pollution level in the cold season<sup>23</sup>. Four major sources of the ambient PM<sub>2.5</sub> were produced by the PMF model, including  
308 secondary aerosols, and primary particles of automobile exhaust, coal combustion, and plant biomass burning, which account  
309 for 34%, 27.7%, 25.2%, and 13.1% of total PM<sub>2.5</sub> mass concentration, respectively. Their source profiles and proportions were  
310 showed in Figure 1. Therefore, although the contribution of secondary aerosols cannot be ignored, the main anthropogenic  
311 sources of urban air PM<sub>2.5</sub> were primary emissions (66%) from the various fuel combustions.

#### 312 3.2 Chemical compositions of different PM<sub>2.5</sub> from combustion sources and from representative urban ambient air

313 Typical chemical components including carbonaceous fractions, heavy metals and WSIs of all PM<sub>2.5</sub> samples from both  
314 ambient urban air and 30 representative combustion primary sources (covering different categories of automobile exhaust, coal  
315 combustion, and plant biomass burning) were analyzed and compared with each other.

316 According to the comparisons of PM<sub>2.5</sub> bound carbonaceous fractions (Figure 2), automobile and biomass sourced PM<sub>2.5</sub>  
317 contained significantly higher total carbon (TC) content than coal combustion and ambient air, while the OC/EC ratio trend  
318 was ambient air > coal combustion > biomass burning > automobile exhaust sources. It indicated that the carbon content of  
319 ambient PM<sub>2.5</sub> mixture was lower and dominated by OC than that of combustion primary sources. Figures S4-S7 showed the  
320 detailed carbon fraction characteristics (contents and ratio) of PM<sub>2.5</sub> from each specific source. Carbonaceous fractions in  
321 automobile exhaust PM<sub>2.5</sub> were high but the difference between OC and EC content was small. Depending on the diverse  
322 automobile fuels, loads and tailpipe emission standards, the concentrations of carbon fractions in exhaust PM<sub>2.5</sub> varied widely  
323 with vehicle categories. The carbonaceous portion of PM<sub>2.5</sub> gradually declines as emission regulations rise, and EC likewise  
324 declines dramatically (Figure S4). However, such differences among coal types were less, except the bituminous coal with  
325 extreme high OC (Figure S5). The carbonaceous fraction of PM<sub>2.5</sub> from plant biomass burning differed in raw material species  
326 that tree branches source PM<sub>2.5</sub> generally contained higher carbon contents than those from crop straws (Figure S6).

327 Based on the grouped (Figure 3) and individual (Figures S8-S11) distributions of the measured heavy metals in various  
328 PM<sub>2.5</sub>, the V concentrations of combustion sources were generally higher while Co and Mn were lower than ambient urban air.

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343 Coal combustion emissions carried highest levels of Pb and were enriched in Cu and As (Figure S9), while biomass burning  
344 were rich in Cr and Ni (Figure S10). However, automobile exhausts were enriched in most heavy metals, especially Cu, and  
345 Cr, Ni, V, Mn (Figure S8). Heavy metals from different types of automobile exhausts with the same emission standard varies  
346 greatly. Anthracite and industrial coal combustions contain similar heavy metals much more than bituminous coal. Generally,  
347 Pb, V, Mn, As, and Cu in branches source PM<sub>2.5</sub> were higher than straws, while Cr, Ni, and Co were dominant and higher in  
348 straw burning emissions. A special discovery was that corn cob burning PM<sub>2.5</sub> carried more heavy metals than corn straw and  
349 was the biomass with the highest emission levels of heavy metals. Correspondingly, ambient air PM<sub>2.5</sub> were also rich in most  
350 metals, especially Mn, Pb, and Ni, Cu, Cr. Therefore, coal combustion sources might contribute most Pb to urban ambient air,  
351 and contribute significant Cu and As with automobile exhaust emissions, while plant biomass burning and automobile sources  
352 contribute the Cr and Ni. Besides natural dust, automobile exhaust should be the main anthropogenic source of airborne Mn.  
353 Considering the PMF source apportionments of ambient aerosols, automobile exhaust should be the main source of Cr in urban  
354 air PM<sub>2.5</sub>, and also the source for Cu together with coal combustion.

355 According to the comparisons of water-soluble cation and anion concentrations in various PM<sub>2.5</sub> (Figure 4), coal  
356 combustions contained highest SO<sub>4</sub><sup>2-</sup> and NH<sub>4</sub><sup>+</sup>, automobile exhausts had highest contents of NO<sub>3</sub><sup>-</sup>, Na<sup>+</sup> and Ca<sup>2+</sup>, while plant  
357 biomass burning sources contained highest K<sup>+</sup> and Cl<sup>-</sup>, but Mg<sup>2+</sup> was the lowest for all sources. However, the urban ambient  
358 air PM<sub>2.5</sub> contained highest NO<sub>3</sub><sup>-</sup> and were also dominated by SO<sub>4</sub><sup>2-</sup> and NH<sub>4</sub><sup>+</sup>, for which NO<sub>3</sub><sup>-</sup> should be mainly contributed  
359 by secondary aerosols and automobile primary source, SO<sub>4</sub><sup>2-</sup> and NH<sub>4</sub><sup>+</sup> should be significantly from coal combustions. Besides  
360 NO<sub>3</sub><sup>-</sup>, Na<sup>+</sup> and Ca<sup>2+</sup>, automobile source PM<sub>2.5</sub> also had the highest F<sup>-</sup> and Mg<sup>2+</sup> concentrations than other sources. The detailed  
361 concentration distributions of WSIs in PM<sub>2.5</sub> from each specific source were provided in Figures S12-S14. The WSIs levels  
362 vary widely with specific source categories. PM<sub>2.5</sub> from LDDV<sub>2</sub>-2 had the lowest amount of WSIs compared to the other  
363 automobile exhausts (Figure S12). Similar to the metal composition, bituminous coal also had the lowest WSIs among all coals  
364 (Figure S13). Compared to branches, PM<sub>2.5</sub> from burning crop straws had much greater levels of K<sup>+</sup>, Cl<sup>-</sup>, SO<sub>4</sub><sup>2-</sup> and less levels  
365 of F<sup>-</sup>, NO<sub>3</sub><sup>-</sup> (Figure S14).

366 To summarize, the overall concentrations of measured TC, cumulated heavy metals and WSIs in PM<sub>2.5</sub> from each source  
367 type were showed in Figure 5. Among all source emission and environmental receptor samples, the cumulated heavy metals  
368 from coal combustion was highest and automobile exhaust was higher than ambient PM<sub>2.5</sub>, the overall carbon contents from  
369 automobile exhaust and biomass burning were both higher than ambient PM<sub>2.5</sub>, while only the cumulated soluble ions in PM<sub>2.5</sub>  
370 from primary source of coal combustion was equivalent to the ambient aerosols. In a word, chemical compositions of PM<sub>2.5</sub>  
371 distributed much diversely and varied significantly with the specific source types of combustion emissions.

### 372 3.3 Cell viability, oxidative stress and inflammation levels exposed to various mass-normalized PM<sub>2.5</sub>

373 Multiple toxicological endpoints (cell viability, oxidative stress, and inflammation) that facilitate identifying the specific  
374 particle triggering ROS and inflammatory responses resulting in cell death were evaluated for source-specific PM<sub>2.5</sub>. After 24

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382 h exposure to the same dose of different PM<sub>2.5</sub> obtained from specific emission sources, the A549 lung cells also showed varied  
383 toxicological responses (Figure 6). The survival rate of cells exposed to automobile exhaust PM<sub>2.5</sub> was much lower than  
384 ambient air PM<sub>2.5</sub> (Figure 6.1). Automobile exhaust PM<sub>2.5</sub> induced the highest ROS production in cells higher than biomass  
385 burning and both sources were also much higher than ambient PM<sub>2.5</sub> (Figure 6.2). Coal combustion induced the highest cellular  
386 IL-6 production followed by automobile exhaust that was also higher than ambient air PM<sub>2.5</sub>, while the PM<sub>2.5</sub> from automobile  
387 exhaust and biomass burning induced similarly higher cellular production of TNF- $\alpha$  than ambient PM<sub>2.5</sub> (Figure 6.3, 6.4).  
388 These results suggested that, combustion primary emission PM<sub>2.5</sub> had stronger ability to induce oxidative stress and  
389 inflammatory injury in lung cells than ambient air PM<sub>2.5</sub>, thus resulted in the higher probability of apoptosis induction (Victor  
390 and Gottlieb, 2002; Wang et al., 2013). Generally, the mass-normalized PM<sub>2.5</sub> from primary source of automobile exhaust  
391 posed the strongest overall toxicity. Therefore, to protect public health by controlling PM<sub>2.5</sub> pollution, the anthropogenic  
392 combustions were key target sources, especially the most toxic automobile PM<sub>2.5</sub> should be reduced preferentially. ▲

### 393 3.4 Correlations between various PM<sub>2.5</sub> components and toxicity endpoints

394 Spearman correlation coefficients between chemical compositions and cellular toxicological response indicators were applied  
395 to screen the key components of all PM<sub>2.5</sub> involved in cell injury (Figure 7). It was found that, the degrees of correlations  
396 varied with the toxicological mechanisms of different airborne chemicals. Based on the overall PM<sub>2.5</sub> samples from various  
397 sources, the pro-inflammatory cytokine IL-6 showed significantly strong positive correlations with some heavy metals (As,  
398 Pb, V, Cu), while TNF- $\alpha$  and oxidative stress (ROS) had similar significantly positive correlations with aerosol components  
399 of carbon fractions (EC, OC) and transition metals (V, Cr, Ni). The TNF- $\alpha$  also showed positive correlation with water soluble  
400 Cl<sup>-</sup> and K<sup>+</sup>, and ROS correlated with F<sup>-</sup>, Ca<sup>2+</sup> and Mg<sup>2+</sup>.

## 402 4 Discussion

### 403 4.1 New chemical markers for source apportionments of ambient air PM<sub>2.5</sub>

404 Combustion emissions are key anthropogenic sources contributing to urban air PM<sub>2.5</sub>, through both primary and secondary  
405 aerosols, which were 66% and 34% calculated by PMF model, respectively (Figure 1). The high concentrations of chemical  
406 markers are usually used in source analysis, such as ammonium sulfate and nitrate for secondary aerosols which are originated  
407 mainly from the gaseous precursors (e.g., NH<sub>3</sub>, SO<sub>2</sub> and NO<sub>x</sub>) (Mahilang et al., 2021), the EC, Cu, Mn, and Ni for vehicle  
408 exhaust (Srivastava et al., 2021), the As, Pb, OC, EC, SO<sub>4</sub><sup>2-</sup> and relatively low NO<sub>3</sub>/SO<sub>4</sub><sup>2-</sup> ratios for coal combustion (Dai et  
409 al., 2020), soluble K<sup>+</sup> and Cl<sup>-</sup> for plant burning (Jain et al., 2020). The detailed chemical species of these specific source  
410 emission PM<sub>2.5</sub> samples also supported the results. Moreover, low OC/EC ratio of high TC content, high NO<sub>3</sub><sup>-</sup>, F<sup>-</sup>, Na<sup>+</sup>, Ca<sup>2+</sup>  
411 and Mg<sup>2+</sup>, V and Mn of automobile exhaust; Pb and As, SO<sub>4</sub><sup>2-</sup> and NH<sub>4</sub><sup>+</sup> of coal combustion; soluble K<sup>+</sup> and Cl<sup>-</sup>, and high  
412 OC/EC ratio of high TC for plant biomass burning found in current study (Figures 2-5), could also be corresponding potential

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437 aerosol source markers. The principal aim of this paper was to assess and contrast the chemical composition and potential  
438 harmfulness of PM arising from diverse anthropogenic sources, thus natural sources, like fugitive soil dust, were not included  
439 in the source examination.

#### 440 4.2 Common PM<sub>2.5</sub> components related to specific combustion sources

441 Generally, the automobile exhaust PM<sub>2.5</sub> had high TC content and low OC/EC value with considerable EC content (Figure 2),  
442 varying with specific vehicle types (Figure S4). The contents of the carbon fractions from diesel vehicles were higher than  
443 gasoline exhausts, and the OC/EC ratios of diesel exhausts were much lower than gasoline vehicles, owing to both considerable  
444 contents of EC and OC from diesel vehicle emission PM<sub>2.5</sub>. Some diesel vehicles showed higher EC emissions with age, so  
445 exhaust cleaning devices for them are suggested. In addition, the amounts of OC and EC in exhausts gradually decreased with  
446 the strengthened emission standards they met (Wong et al., 2020). In PM<sub>2.5</sub> samples obtained from coal combustion (Figure  
447 S5), the TC contents of bituminous coals were significantly higher than that of honeycomb coals, anthracite coals, and  
448 industrial coals, because bituminous coals contain higher volatile fraction, which is composed of organic matter. Therefore,  
449 besides the way of combustion and the use of combustion stoves, the coal quality related to different coal types and origins  
450 determine the carbonaceous fractions of the PM<sub>2.5</sub> emitted by coal combustion (Zhang et al., 2022). In the PM<sub>2.5</sub> samples from  
451 plant biomass combustion (Figure S6), OC contents were generally higher than EC contents, except that pine branches  
452 contained higher EC and rapeseed straw had considerable contents of EC and OC. Dominated by OC (Figure S7), the  
453 concentrations of carbonaceous fractions in urban ambient air samples varied seasonally (Flores et al., 2020; Xu et al., 2019).  
454 Combining the TC contents and OC/EC ratios, carbonaceous components in ambient PM<sub>2.5</sub> mainly originate from semi-volatile  
455 organic compounds (SVOCs) (Wang et al., 2018) and combustion primary emissions for OC (Kang et al., 2018), and  
456 automobile exhaust for EC (Barraza et al., 2017). Thus, to control ambient carbon aerosol pollution, besides reducing the  
457 precursor emissions of secondary organic aerosols (SOA), controlling primary aerosols especially EC from diesel vehicles  
458 were key measures.

459 Airborne redox-active metals are usually linked with the oxidation stress of PM<sub>2.5</sub>. Different types of automobiles emitted  
460 diverse metal contents (Figure S8). Metal elements in automobile exhaust are primarily contributed by fuels, lubricants, and  
461 engine component abrasion. Because Mn is a common antidetonator that delays and prevents the oxidation of hydrocarbons  
462 and increases the octane number, which not only increases the thermal efficiency of the engine but also improves the emission  
463 performance of the vehicle (Cheung et al., 2010), the Mn content was greater in gasoline vehicle exhausts than in diesel  
464 vehicles. Although there are multi-sources of traffic Pb emissions such as fuel combustion and brake wear (Wang et al.,  
465 2019; Panko et al., 2019), the automobile exhaust Pb content of gasoline vehicles were greater than diesel vehicles owing to  
466 oil combustion. Moreover, for the same vehicle type (LDDVs-1 and 2; HDDVs-1 and 2; SDGCs-1 and 2), the stricter the  
467 emission standard required, the lower the exhaust metal contents. The metal contents in the PM<sub>2.5</sub> of trucks was higher than  
468 that of passenger cars (Wu et al., 2016). In the combustion PM<sub>2.5</sub> of 10 coal types (Figure S9), Pb contents were the highest

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485 than other heavy metals, similar to available findings (Zhang et al., 2020). The PM<sub>2.5</sub> metals from bituminous coal were  
486 significantly lower than other coal types, because indicated by the coal quality analysis, bituminous coal has a low ash content  
487 which is mainly derived from non-combustible minerals in coal. These findings suggested that coal maturity might be an  
488 important factor influencing the metal composition of particulates emitted from coal combustion (Shen et al., 2021;Zhang et  
489 al., 2021). Heavy metal contents in biomass burned PM<sub>2.5</sub> varied much widely with raw plant types (Figure S10), although  
490 dominated by Cr and Ni. Different plant species and even different plant parts differ significantly in their ability to uptake and  
491 accumulate metals from soil (Zhao et al., 2020). Moreover, because of the high enrichment factors of some metals for crop  
492 straws (Zhang et al., 2016;Sun et al., 2019), they also released more Cr, Ni, and Co during burning than fuelwoods. Total metal  
493 emissions were highest in corn cob but lowest in peanut straw burning PM<sub>2.5</sub>. The heavy metals enriched in urban ambient air  
494 PM<sub>2.5</sub> demonstrated a seasonal pattern (Chen et al., 2018;Hsu et al., 2016) (Figure S11). Contents of V, Co, and As were  
495 relatively low and are less affected by seasonal changes. Accordingly, supported by the metal profiles of anthropogenic  
496 combustion sources and ambient aerosols, to control the environmental airborne heavy metal pollution, the Pb, Cu and As  
497 from honeycomb, anthracite and industrial coal combustion, Cu from vehicle exhausts and especially V from light duty diesel  
498 van with the CN.III emission standard and Mn from gasoline vehicles, Cr and Ni from biomass especially crop straws burning,  
499 should be key targets.

500 Epidemiological studies have also shown the mortality closely related to the WSIs such as sulfate and nitrate in aerosols  
501 (Ostro et al., 2009;Liang et al., 2022). Among the WSIs contents of various automobile exhaust PM<sub>2.5</sub> (Figure S12), NO<sub>3</sub><sup>-</sup> and  
502 Ca<sup>2+</sup> were the most abundant anion and cation, respectively. The high NO<sub>3</sub><sup>-</sup> in the automobile PM<sub>2.5</sub> may be due to NO<sub>x</sub>  
503 production during high-temperature combustion (Hao et al., 2019), while the high Ca<sup>2+</sup> content should be related to additives  
504 in automobile fuels and calcium-based lubricants (Yang et al., 2019). Moreover, the exhaust WSIs decreased with the  
505 strengthened automobile emission standards required. Coal combustion PM<sub>2.5</sub> contained relatively higher SO<sub>4</sub><sup>2-</sup> and NH<sub>4</sub><sup>+</sup>  
506 concentrations followed by Cl<sup>-</sup> than other WSIs species (Figure S13). Among various coal types, industrial coals emitted  
507 highest SO<sub>4</sub><sup>2-</sup> followed by honeycomb and industrial coal with also high NH<sub>4</sub><sup>+</sup>, but bituminous coals emitted low WSIs which  
508 were mainly NO<sub>3</sub><sup>-</sup>, F<sup>-</sup> and Na<sup>+</sup>, Ca<sup>2+</sup>. The WSIs emission factors of honeycomb coal were generally higher than those of lump  
509 coal (Yan et al., 2020). For biomass combustion emissions (Figure S14), Cl<sup>-</sup> and K<sup>+</sup> were dominant WSIs in PM<sub>2.5</sub> from straw-  
510 type fuels (Tao et al., 2016;Sillapapiromsuk et al., 2013), but fuelwood-type combustion emitted high NO<sub>3</sub><sup>-</sup>. Plant species  
511 absolutely determine the emissions (Liao et al., 2021). Finally, there were also high levels of NO<sub>3</sub><sup>-</sup>, SO<sub>4</sub><sup>2-</sup>, and NH<sub>4</sub><sup>+</sup> in ambient  
512 air PM<sub>2.5</sub> (Zhang et al., 2019) (Figure S15). Consequently, implied by the WSIs species distributed in combustion primary  
513 sources and environmental PM<sub>2.5</sub>, to control the aerosols ions pollution, the NO<sub>3</sub><sup>-</sup> from vehicle exhausts and fuelwood burning;  
514 SO<sub>4</sub><sup>2-</sup> and NH<sub>4</sub><sup>+</sup> from honeycomb, anthracite and industrial coal combustion; Cl<sup>-</sup> and K<sup>+</sup> from biomass especially crop straw  
515 burning, should be principal targets, by stricter automobile emission standards or using clean coals.

#### 516 4.3 PM<sub>2.5</sub> toxicity related to specific sources by pivotal chemical components

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528 The complexity of the sources and compositions of atmospheric PM<sub>2.5</sub> leads to different toxicological effects (Newman et al., 2020; Kelly, 2021). The toxicological effects of PM<sub>2.5</sub> are not comparable among different studies owing to distinct exposure concentrations, biological models, endpoints, and PM<sub>2.5</sub> generation methods (Park et al., 2018; Kelly and Fussell, 2020). In this study, we employed same exposure conditions and biological endpoints, in order to obtain comparable toxicity data for PM<sub>2.5</sub> from different sources. Our mass-normalized results demonstrated that automobile exhaust PM<sub>2.5</sub> induced the highest lethality and cellular ROS and TNF- $\alpha$  production, coal combustion PM<sub>2.5</sub> induced the highest cellular IL-6 production, plant biomass burning PM<sub>2.5</sub> induced considerable cellular TNF- $\alpha$  and ROS production (Figure 6). Generally, various toxicities of combustion emission primary PM<sub>2.5</sub> were much greater than the urban ambient air PM<sub>2.5</sub> (Figure 6), owing to the higher concentrations of specific toxic components in PM<sub>2.5</sub> from these sources. The supplementary information had included exhaustive cytotoxicity indicators from each individual source (Figure S16-S19). While the survival rate of cell exposed to CN.III emission standard PM<sub>2.5</sub> was the lowest and the capacity to induce cells to produce ROS was the highest for CN.IV, automobile exhaust had a similar potential to cause cells to produce inflammatory cytokines (Figure S16). The capability to induce IL-6 production in cells was highest for industrial coal PM<sub>2.5</sub>, whereas bituminous coal had the highest survival rate of cells and TNF- $\alpha$  induction capacity (Figure S17). From the Figure S18 we can see that the PM<sub>2.5</sub> cytotoxicity of straws and branches burning was analogous, but it should be noted that the cell viability of various straw PM<sub>2.5</sub> differs significantly, that may be related to the raw fuel characteristics.

544 These possible mechanisms were implied by the overall relationships between the measured chemical components with cytotoxicity indicators of PM<sub>2.5</sub> from various specific sources (Figure 7). In general, both TNF- $\alpha$  and ROS were significantly positively correlated with carbonaceous fractions and redox-active transition metals (V, Cr, Ni), which were main contributors of automobile exhausts and biomass burning. The IL-6 was significantly positively correlated with some heavy metals (As and Pb, V and Cu), which were main contributors of coal combustion sources. Potential mechanisms include that, carbon fractions bound in PM<sub>2.5</sub> could be transformed into reactive metabolites and then induce ROS production in cells (Stevanovic et al., 2019), and the PM<sub>2.5</sub> bound transition metals could also induce ROS production through the Fenton reaction and disrupt the function of enzymes in cells (Verma et al., 2010; Sørensen et al., 2005; Zou et al., 2016). Oxidative stress can lead to inflammatory infiltration of neutrophils and stimulate immune cells to produce inflammatory cytokines, among which TNF- $\alpha$  and IL-6 play important roles in the inflammation development (Xu et al., 2020). Ultimately, excessive production of ROS leads to dysfunctional endoplasmic reticulum responses and dysfunctional lipid metabolism in ROS bursts can result in cell membrane damage and even cell death (Piao et al., 2018; Zhao et al., 2004). There have been some related supporting reports. For instance, the OC and EC were significantly associated with biological responses of PM<sub>2.5</sub> from vehicle emissions collected in tunnels (Niu et al., 2020). The polar or quinone fractions of PAHs in diesel engine exhaust particles significantly contributed to the heightened toxic response (Xia et al., 2004). The PM<sub>2.5</sub> generated from biomass burning contained a substantial concentration of carbonaceous components. In addition, Cr and Ni in PM<sub>10</sub> from straws were highly associated with ROS (Li et al., 2023). In current study, cellular ROS was also correlated with water soluble, Ca<sup>2+</sup>, F, and Mg<sup>2+</sup>, which were main

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841 contributors of automobile exhaust,  $PM_{2.5}$ . The  $Ca^{2+}$  controls the membrane potential and regulates mitochondrial adenosine  
842 triphosphate (ATP) production, and excessive  $Ca^{2+}$  leads to energy loss and more ROS production (Madreiter-Sokolowski et  
843 al., 2020). Moreover, the TNF- $\alpha$  was also positively correlated with water soluble  $Cl^-$  and  $K^+$ , which were main contributors  
844 of plant burning  $PM_{2.5}$ . Therefore, the accumulations of some organic matters with high carbonaceous content (OC, EC) in  
845  $PM_{2.5}$  typically from automobile exhausts and plant biomass burning, redox-active metals (V, Cr, Ni) and water-soluble anions  
846 ( $Cl^-$ ,  $F^-$ ) and cations ( $Ca^{2+}$ ,  $Mg^{2+}$ ) contributed by various combustions, might induce ROS production in cells, cause cellular  
847 damage through oxidative stress and inflammatory responses, impair cell viability and finally harm human health.

848 Considering the multi-endpoints measured and the  $PM_{2.5}$  toxicity mechanisms mentioned above, based on the cell viability  
849 first, and then ROS followed by inflammatory markers, together with the significantly related toxic chemical composition  
850 contents (Park et al., 2018), we put forward a general sequence of overall mass-normalized toxicity for these combustion  
851 source  $PM_{2.5}$  to managers. To improve the urban environmental air quality for best public health benefits by controlling  
852 aerosols pollution, considering the differential toxicity intensity of each chemical component and their contributions from  
853 various sources to ambient aerosols, preferential targets of specific primary  $PM_{2.5}$  sources and bound pollutants to be controlled  
854 are suggested as following sequence: Reducing all anthropogenic combustions, especially decreasing the automobile exhaust  
855  $PM_{2.5}$  with high contents of EC, transition metals (V, Cu, Ni, Cr), and ions ( $Ca^{2+}$ ,  $Mg^{2+}$ ,  $F^-$ ,  $Na^+$ ) from diesel exhausts by  
856 strengthening the emission standards, then lessening the coal combustion with high heavy metals (As, Pb, Cu) by replacement  
857 with low-ash clean coals, and depressing the biomass burning with high OC, Ni, Cr,  $Cl^-$  and  $K^+$  from crop straw emissions.

#### 858 4.4 Limitations and perspectives

859 In current study, we selected A549 cell based on previous abundant experimental experiences and also because it has been  
860 used popularly in *in vitro* toxicology studies to elucidate the cellular and molecular mechanisms of PM involved in lung for  
861 many decades (Li et al., 2022b). However, recently the human normal bronchial epithelial cell BEAS-2B was preferred over  
862 the human lung adenocarcinoma epithelial cell A549. For instance, both cells were used in an aerosol study (Bonetta et al.,  
863 2017), results of which highlighted the higher sensitivity of BEAS-2B cells respect to A549 also in samples with low level of  
864 pollutants, because the  $PM_{0.5}$  samples from Italian towns can induce genotoxicity in normal cells while cancer cells might be  
865 resistant to their adverse effects. Therefore, although our results are reasonable under the same exposure conditions, there were  
866 still potential limitations of A549 cells since they may be more resistant to exposure to external compounds, and the generally  
867 more sensitive BEAS-2B cells are suggested for future studies.

868 In toxicity assessments, cell vitality reflects the overall health of cells, encompassing factors such as cell membrane integrity,  
869 intracellular metabolic activity, and cell proliferation capacity. Decreased cellular vitality may be associated with cell damage,  
870 toxic effects, or cellular apoptosis. Inflammation markers are employed to assess the extent and nature of inflammatory  
871 reactions, including the production of cytokines and inflammatory mediators, as well as the activation status of inflammatory  
872 cells. Inflammation is a complex physiological response, typically delineated by the immune and inflammatory reactions of

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- 删除了: According to the source-based aerosols cytotoxicity comparisons (Figure 6), automobile exhaust  $PM_{2.5}$  induced the highest lethality and cellular ROS and TNF- $\alpha$  production, coal combustion  $PM_{2.5}$  induced the highest cellular IL-6 production, biomass burning  $PM_{2.5}$  induce considerable cellular TNF- $\alpha$  and ROS production. It has been reported that  $PM_{2.5}$  from coal combustion decreased cell viability, increased overall DNA methylation, and led to cellular DNA oxidative damage (Møller et al., 2014). Some studies proposed that the toxicity of PAHs in PMs from biomass burning was likely underestimated (Sarigiannis et al., 2015; Yin and Xu, 2018). The OC and EC were reported significantly associated with biological responses of PMs from vehicle emissions collected in tunnels (Niu et al., 2020).
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917 the body to stimuli such as injury or infection. Alterations in inflammation markers can indicate the intensity and nature of the  
 918 inflammatory response. In this study, multiple biological responses of epithelial cells to various PM<sub>2.5</sub> were evaluated,  
 919 including that cell viability evaluated the mitochondrial dehydrogenase activity of the living cells, excessive intracellular ROS  
 920 formation induced by PM<sub>2.5</sub> was responsible for oxidative stress to the cells, cytokines IL-6 and TNF-α were determined for  
 921 the effect of PM<sub>2.5</sub> on pro-inflammatory response in cells. In general, in vitro data can be used to rank various types of particles  
 922 in terms of the toxic potential including possible carcinogenicity. Each marker will help to understand the hazard and toxicity  
 923 of PM<sub>2.5</sub>. However, the toxicity of PM<sub>2.5</sub> may be the result of multiple components acting through disparate physiological  
 924 mechanisms, with inconsistent relationships among endpoints (Park et al., 2018). For instance, in BEAS-2B cells, oxidative  
 925 stress generated by H<sub>2</sub>O<sub>2</sub> exposure often results in cytotoxicity rather than by stimulating cytokine/chemokine responses,  
 926 sometimes no correlation between oxidative damage and cytokine/chemokine responses. Moreover, TNF-α gene was not  
 927 detected in BEAS-2B cells exposed to atmospheric PM collected from Benin, but the gene expression of other inflammatory  
 928 cytokines (IL-1β, IL-6, and IL-8) were significantly induced, and decreasing cell viability was highly correlated with high  
 929 secretion of all studied cytokines (Cachon et al., 2014). Therefore, in the present study, it was impossible to analyze all  
 930 chemicals in PM<sub>2.5</sub> and determine all related toxicological endpoints, so unmeasured chemicals and endpoints might also play  
 931 roles in the incongruous or unexplained results, and we also can't over-explain the mechanisms just based on statistical  
 932 relations. To overcome these hurdles, standardization of toxicological studies (experimental methodologies) and reporting  
 933 guidelines are necessary for tracking and comparing results.

934 This study ranked the unequal "toxic effects" based on the same mass concentration of PM<sub>2.5</sub> exposure in body lung fluid  
 935 system, while the "health risks" usually relating to the inhalation exposure concentration of PM<sub>2.5</sub> in ambient air were not  
 936 calculated and evaluated quantitatively. Moreover, non-linear concentration-response functions for various endpoints and  
 937 different exposure concentrations might also limit using toxicological data straightforwardly to predict health effects  
 938 (morbidity, mortality) in human populations, so drawing conclusions precisely quantifying/ranking the health risks of PM<sub>2.5</sub>  
 939 from specific sources or of individual PM<sub>2.5</sub> components is still not an easy task (Kelly and Fussell, 2020). Therefore, coupled  
 940 with source apportionment and exposure level of ambient aerosols pollution, toxicology combined with epidemiology studies  
 941 are essential for linking these factors and understanding scientific mechanisms to reach conclusions.

## 943 5 Conclusions

944 In current study, we found that 2/3 mass of urban ambient air PM<sub>2.5</sub> in a representative megacity of eastern China originated  
 945 from primary sources of anthropogenic combustions including coal, automobile, and biomass. Because of the significant  
 946 differences in the chemical compositions, the diverse PM<sub>2.5</sub> from both mixed ambient air and directly from individual  
 947 combustion sources showed much differential mass-normalized in vitro toxicity to the human lung epithelial cells, either for  
 948 the environmental aerosol samples collected from different seasons, or for the primary emissions of PM<sub>2.5</sub> from various specific

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956 source types. According to the comparative study and correlation analysis, the carbonaceous fractions (OC, EC) and redox-  
957 active heavy metals (V, Ni, Cr) assisted by water-soluble ions (Ca<sup>2+</sup>, Mg<sup>2+</sup>, F<sup>-</sup>, Cl<sup>-</sup>), might play important roles in inducing  
958 cellular ROS production, causing oxidative stress and inflammation, resulting in cell injury and apoptosis, thus damage human  
959 health. These toxic pollutants accumulated in ~~specific-source PM<sub>2.5</sub>~~ varied by the ~~emission types and raw~~ fuel properties.  
960 Combined with chemical composition and ~~general~~ cytotoxicity ~~rank~~, the preferential controlling targets of specific combustion  
961 sources should be automobile exhaust (diesel vehicles with emission standards inferior to CN.IV), coal combustion (high ash  
962 and high sulfur coals), and ~~plant~~ biomass burning (crop straws). Although showing the synthetic effects of mixed compositions  
963 and complex sources, besides preventing the secondary aerosols from combustions, ~~preferentially~~ targeted reductions of these  
964 primary sources of toxic PM<sub>2.5</sub> direct emissions, would produce the greatest benefits for public health with improved ambient  
965 air quality. Overall, this paper provides a precise, oriented, effective, efficient, and economical composition-source-based  
966 strategies for urban aerosols pollution control. However, as a prospect, the detailed mechanisms for ~~unequal~~ toxicity of PM<sub>2.5</sub>  
967 with complicated components from various sources and their quantitative contributions to the health effects of ambient air  
968 PM<sub>2.5</sub> mixture still need in-depth study.

#### 969 **Supplementary materials**

970 There are 19 figures (Figure S1-S19) and 3 tables (Table S1-S3) in the Supporting Information.

#### 971 **Data availability**

972 All raw data can be provided by the corresponding authors upon request.

#### 973 **Author contributions**

974 XSL conceived and supervised the study; WH, YP, MT, HL, and ZZ collected the samples; WH, YP, MT, WL, HL, ZZ, GS,  
975 and LX analyzed the chemical compositions; WH, YP, and MT performed the toxicity tests; WH, YP, MT, and XSL analyzed  
976 the data; WH and XSL wrote the manuscript draft; XSL, WH, GS, and TM reviewed and edited the manuscript.

#### 977 **Competing interests**

978 The authors declare that they have no conflict of interest.

#### 979 **Financial support**

980 This work was supported by the National Natural Science Foundation of China (NSFC 41977349, 41471418).

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## 1256 Captions of figures

1257 **Figure 1.** The PMF factor profiles of various components and source percentages of secondary aerosol, automobile exhaust,  
1258 coal combustion, and plant biomass burning contributing to the urban ambient air PM<sub>2.5</sub>.

1259 **Figure 2.** Carbon contents (mg kg<sup>-1</sup>) and ratio in PM<sub>2.5</sub> from various specific sources (n=10 for each combustion source and  
1260 n=16 for urban ambient air).

1261 **Figure 3.** Heavy metal contents (mg kg<sup>-1</sup>) in PM<sub>2.5</sub> from various specific sources (n=10 for each combustion source and n=16  
1262 for urban ambient air).

1263 **Figure 4.** Water-soluble ion (WSI) contents (mg kg<sup>-1</sup>) in PM<sub>2.5</sub> from various specific sources (n=10 for each combustion  
1264 source and n=16 for urban ambient air).

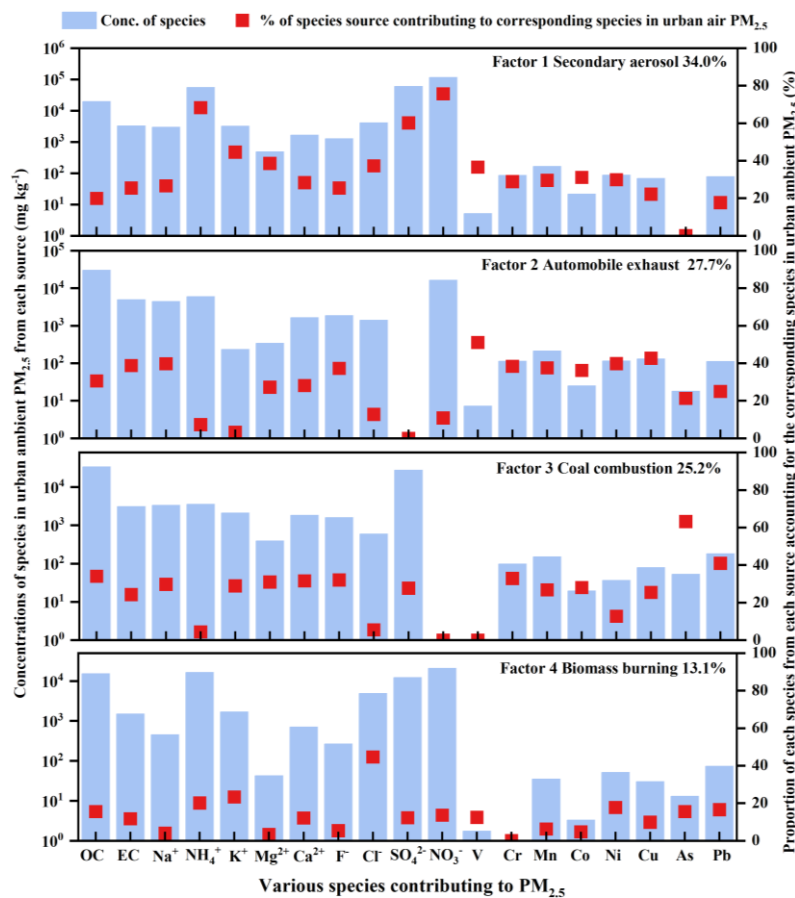
1265 **Figure 5.** Cumulated typical measured components (mg kg<sup>-1</sup>) in PM<sub>2.5</sub> from various specific sources (n=10 for each  
1266 combustion source and n=16 for urban ambient air).

1267 **Figure 6.** Cell viability, oxidative stress and inflammation levels of human alveolar epithelial cell lines (A549) exposed to  
1268 PM<sub>2.5</sub> suspension (80 mg L<sup>-1</sup>) from various specific sources (n=10 for each combustion source and n=16 for urban ambient  
1269 air).

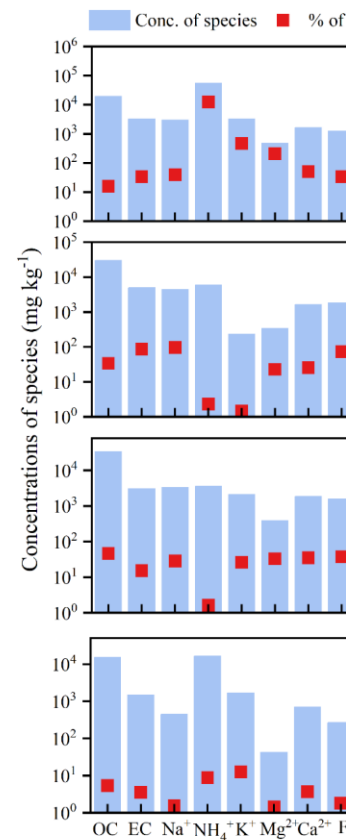
1270 **Figure 7.** Overall correlations between typical cellular toxicological responses and chemical compositions of PM<sub>2.5</sub> from  
1271 various sources (\*p < 0.05, #p < 0.01; n=46).

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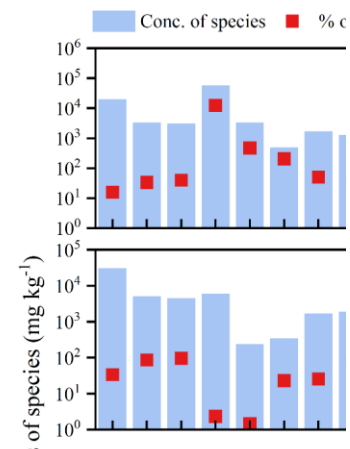
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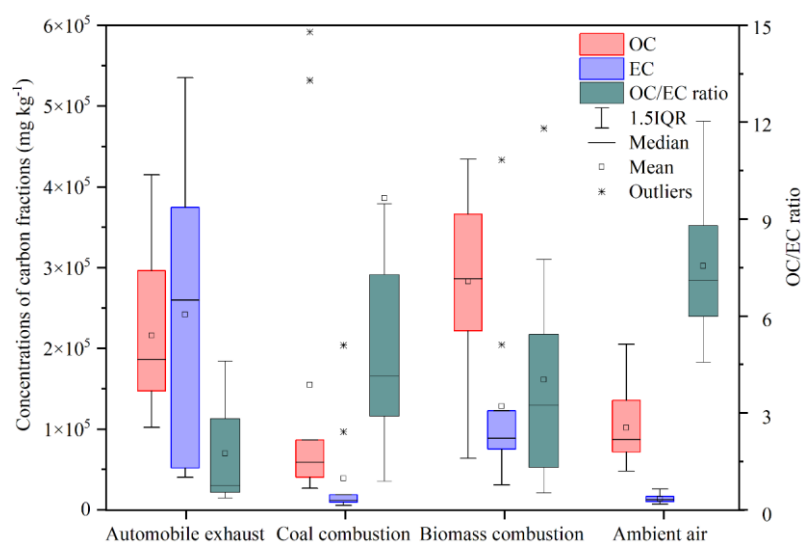


1276 **Figure 1.** The PMF factor profiles of various components and source percentages of secondary aerosol, automobile exhaust,  
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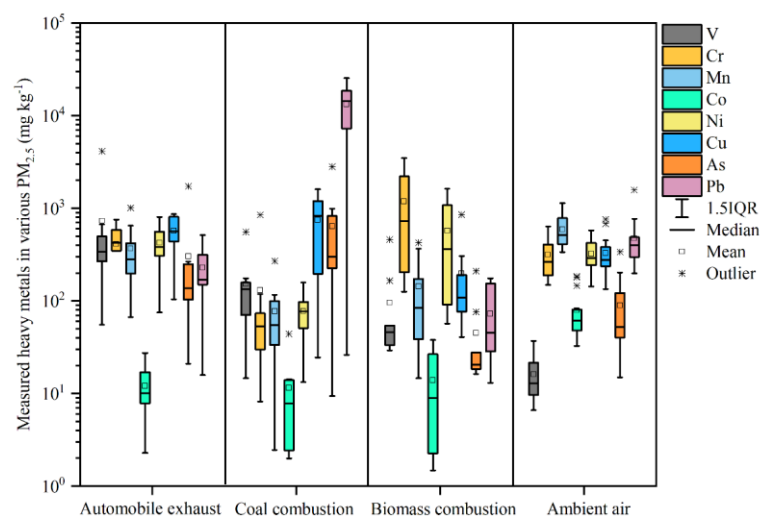


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1285 **Figure 2.** Carbon contents (mg kg<sup>-1</sup>) and ratio in PM<sub>2.5</sub> from various specific sources (n=10 for each combustion source and  
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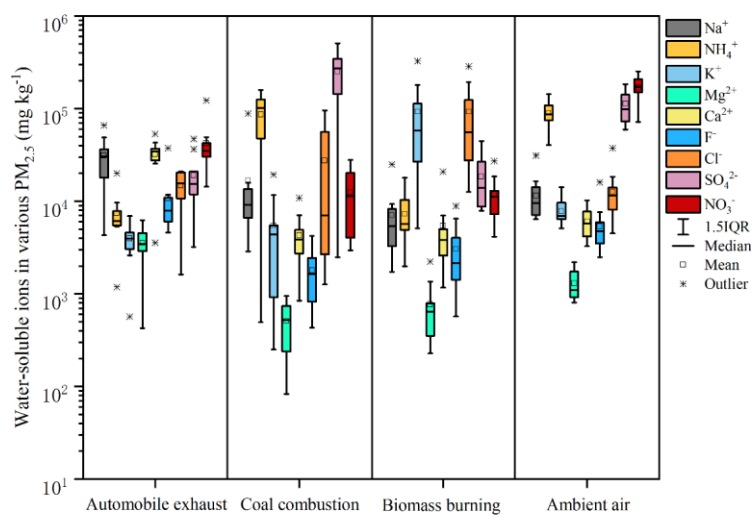




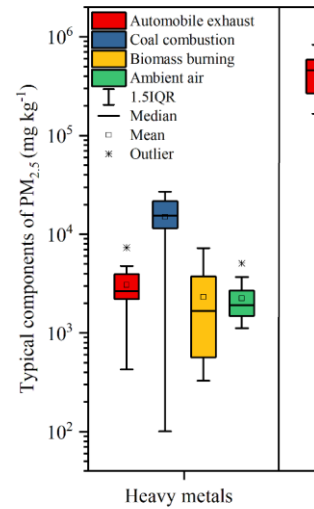
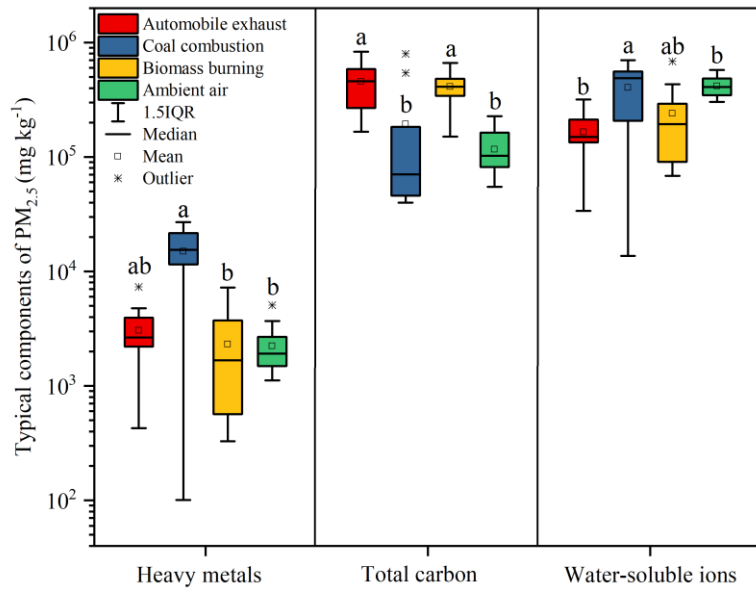
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1289 **Figure 3.** Heavy metal contents (mg kg<sup>-1</sup>) in PM<sub>2.5</sub> from various specific sources (n=10 for each combustion source and  
 1290 n=16 for urban ambient air).

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 1293 **Figure 4.** Water-soluble ion (WSI) contents ( $mg\ kg^{-1}$ ) in  $PM_{2.5}$  from various specific sources ( $n=10$  for each combustion  
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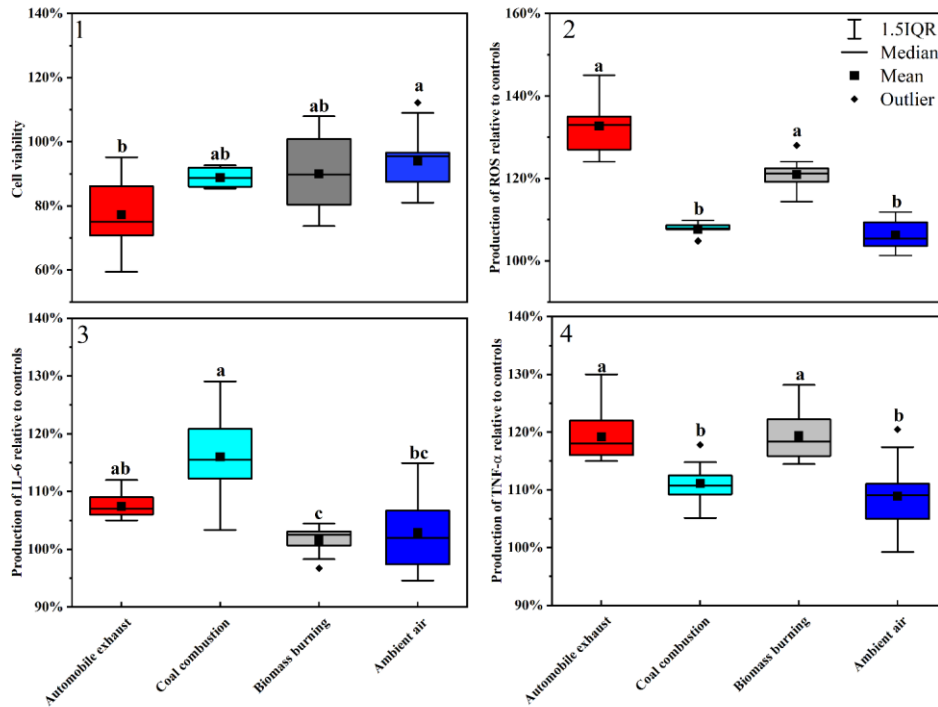
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298 **Figure 5.** Cumulated typical measured components (mg kg<sup>-1</sup>) in PM<sub>2.5</sub> from various specific sources (n=10 for each  
 299 combustion source and n=16 for urban ambient air). The letters a and b are significant groups classified by Kruskal–Wallis  
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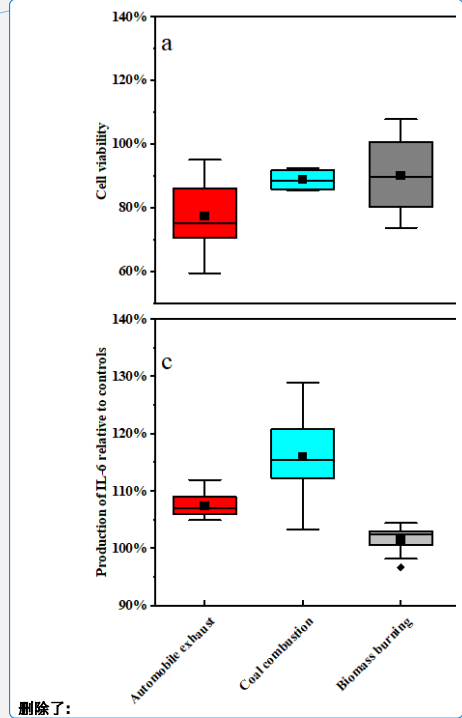


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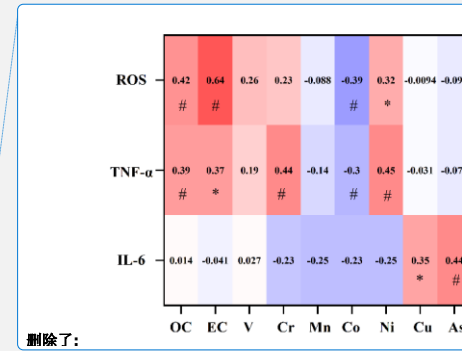
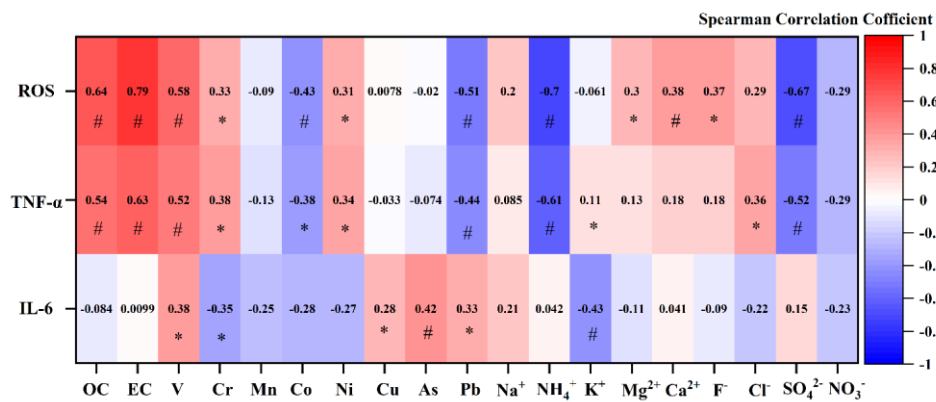
1305 **Figure 6.** Cell viability, oxidative stress and inflammation levels of human alveolar epithelial cell lines (A549) exposed to  
1306 PM<sub>2.5</sub> suspension (80 mg L<sup>-1</sup>) from various specific sources (n=10 for each combustion source and n=16 for urban ambient  
1307 air). The letters a, b and c are significant groups classified by Kruskal–Wallis test, p < 0.05.

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1312 **Figure 7.** Overall correlations between typical cellular toxicological responses and chemical compositions of PM<sub>2.5</sub> from  
 1313 various sources (\*p < 0.05, #p < 0.01; n=46).