Two-year measurements Higher absorption enhancement of Black Carbon properties black carbon in summer shown by two year measurements at the high-altitude mountain site of Pic du Midi Observatory in the French Pyrenees

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Abstract. Black Carbon containing particles Particles containing Black Carbon (BC) are strong lightabsorbers trongly absorb light, causing substantial radiative heating of the atmosphere. The climate-relevant properties of BC are poorly constrained in high-elevation high-altitude mountain regions, where numerous many complex interactions between BC, radiation, clouds and snow have important climate implications. This study presents two-year measurements of BC microphysical and optical properties at the research station of Pic du Midi (PDM) research station, a high-altitude observatory located at 2877 m above sea level in the French Pyrenees. Among the worldwide existing long-term monitoring sites , PDM has experiences limited influence of in the world, PDM is subject to limited influence from the planetary boundary layer (PBL), making it an appropriate a suitable site for characterizing free tropospheric the BC in the free troposphere (FT)BC.

The classification of the dominant aerosol type using the aerosol spectral optical properties of the aerosols-indicates that BC was the predominant absorption component of aerosols is the predominant aerosol absorption component at PDM and controlled the variation of controls the variation in Single Scattering Albedo (SSA) throughout the two years. Single-particle soot photometer (SP2) measurements showed of refractive BC (rBC) show a mean mass concentrations of BC (Mconcentration $(M_{BC;BC})$) of 35 ng m⁻³ and a relatively constant BC rBC core mass-equivalent diameter of around about 180 nm, which are typical values for remote mountain sites. Combining the $(M_{BC;BC})$ with in situ absorption measurements yielded a BC a rBC mass absorption coefficient (MAC_{BC;BC}) of $(M_{BC;BC})$ of $(M_{BC;BC})$ of $(M_{BC;BC})$ of $(M_{BC;BC})$ at $(M_{BC;BC})$ at $(M_{BC;BC})$ at $(M_{BC;BC})$ at $(M_{BC;BC})$ at $(M_{BC;BC})$ and $(M_{BC;BC})$ are $(M_{BC;BC})$ at $(M_{BC;BC})$ at $(M_{BC;BC})$ and $(M_{BC;BC})$ are sufficient reduction of the ratio $(M_{BC;BC})$ and $(M_{BC;BC})$ are sufficient reduction of $(M_{BC;BC})$. However we found that the wet removal process did not affect the size of $(M_{BC;BC})$ are sufficient reduction of $(M_{BC;BC})$. However we found that the wet removal process did not affect the size of $(M_{BC;BC})$ are sufficient reduction of $(M_{BC;BC})$ and $(M_{BC;BC})$ are sufficient reduction of $(M_{BC;BC})$ are sufficient reduction of $(M_{BC;BC})$ and $(M_{B$

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body of evidence suggests that biomass burning emissions effectively altered the concentration and optical properties of BC at PDM, leading to higher E_{abs} in summer compared to winter Combining the $\Delta M_{rBC}/\Delta CO$ ratio with air mass transport analysis, we observed additional sources from biomass burning in summer leading to an increase in M_{rBC} and E_{abs} . The diurnal pattern of E_{abs} in summer was opposite to that observed in winter with maximum values of 2.9 observed at noonmidday. We suggest that this daily variation results from photochemical processing driving BC may result from a photochemical process driving rBC mixing state rather than a change in BC emission sourcesources.

Such direct two-year observations of BC properties provide quantitative constraints for both regional and global climate models and have the potential to close the gap between model predicted and observed effects of BC on regional radiation budget and climate. The results demonstrates the complex influence of BC emission sources, transport pathways, atmospheric dynamics and chemical reactivity in driving the light absorption of BC.

1 Introduction

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Black Carbon (BC) is a light-absorbing carbonaceous aerosol produced by incomplete combustion of fossil fuelsand biomass. This includes, including anthropogenic emissions from traffic, residential heating and cooking, power plants, industries, but also natural emissions such as biomass burning (Bond et al., 2013; Bond and Bergstrom, 2006). Recent scientific assessments of the 6th IPCC (Intergovernmental Panel on Climate Change) report (Szopa et al., 2021b) estimates that BC is the most absorbing atmospheric aerosol with a best estimate of effective radiative forcing of around +0.107 W m⁻², thereby increasing the global mean surface air temperature by 0.063 °C for the period 1750–2019 (Szopa et al., 2021a). The contribution of BC to climate change is estimated to have highest uncertainty be among the highest uncertainties (~90%) in climate models, limiting their accuracy (Bellouin et al., 2020). The large uncertainty of BC direct radiative forcing due to BC-radiation interactions can be attributed, in addition to uncertainties in BC emissions and lifetime, to variations of its optical properties that are neglected by climate models (Matsui et al., 2018).

A crucial factor for in estimating the BC radiative effect is the mass absorption cross-section (MAC_{BC}), which is defined as the light absorption-equivalent cross-section of BC per unit of mass concentration (M_{BC}). The MAC_{BC} can be calculated either by dividing the measured absorption coefficient of BC by its mass concentration or by using Mie's Theory and the BC size distribution and coating thickness as input variables. Observations show that the BC radiative forcing is likely underestimated by about 10-40 around 10 to 40% in current climate models due to too low simulated MAC_{BC} (Bond et al., 2013; Boucher et al., 2016; Matsui et al., 2018; Myhre and Samset, 2015). In-situ measurements of MAC_{BC} have reported a wide range of values, going from 3.8 m² g⁻¹ to 58 m² g⁻¹ (Wei et al., 2020). Although such high variability can be attributed, in part, to the determination method of the MAC_{BC} based on M_{BC} and absorption measurement techniques, differences in MAC_{BC} values were found even for the same measurement technique.

Values of MAC_{BC} depend on BC microphysical and chemical properties, which are related to their emission sources (Schwarz et al., 2008) and the effects of aging processes during the transport in the atmosphere (Ko et al., 2020; Laborde et al., 2013; Sedlacek et al., 2022). Freshly emitted BC is made of porous, fractal-like aggregates of nanoparticles (Beeler and

Chakrabarty, 2022; China et al., 2013) that can become coated by condensation and/or coagulation with non-BC components (such as sulfate, nitrate, and organic components) during atmospheric aging (Fierce et al., 2020). Conversely this coating can be removed through evaporation and/or chemical processing via the production of more volatile substances (Sedlacek et al., 2022). Numerous studies have demonstrated that coating of BC with non-absorbing materials is accompanied by an enhancement of light absorption (E_{abs}) through the so-called lensing effect (Cappa et al., 2012; Denjean et al., 2020; Healy et al., 2015; Liu et al., 2015; McMeeking et al., 2014; Peng et al., 2016; Van de Hulst, 1957; Xie et al., 2019; Schwarz et al., 2006; Yus-Díez et al., 2022). However, most of these measurements were performed in the Planetary Boundary Layer (PBL) and over short periods from a few hours to as long as a season.

Both observations and model simulations pointed out an amplification of the warming rate by greenhouse gases and absorbing aerosols at high-mountain sites compared to PBL areas (Gao et al., 2018; Liu et al., 2009; Pepin et al., 2019; Rangwala, 2013). López-Moreno et al. (2014) found a positive trend between altitude and warming rate have shown by running several regional climate models that the occurrence of winter warm events in the Spanish Pyrenees, which could lead to the occurrence of warm events multiplied by two between 2021-2050 and even more until will gradually increase until 2080. This includes an increase in the number of warm days and nights and the number of snow/ice melting days at altitudes above 2000 m above sea level (asl). This so-called Elevation dependent warming Dependent Warming (EDW) has been reviewed by the Mountain Research Initiative EDW Working Group, 2015, who listed the possible mechanisms of behind this phenomenon (Pepin et al., 2015). Among the invoked reasons reasons given, BC is a potential driver of EDW by affecting both absorbing solar radiation in the troposphere and decreasing the surface albedo when deposited on the cryosphere, thereby accelerating snowmelt (Réveillet et al., 2022). In addition, BC was has been found to have a higher radiative effect when it is located above clouds rather than near the surface (Samset and Myhre, 2015; Sanroma et al., 2010). All these findings highlight the importance of studying BC at high altitude mountain sites, where its effects on climate could be even more significant.

This study presents two-year continuous measurements of BC and aerosol properties at the high-altitude long-term monitoring station Pic du Midi (PDM). Located at 2877 m above sea level (asl.) asl in the French Pyrenees, PDM has been early identified as a clean remote station (Marenco et al., 1994). By means of Using a backward particle dispersion model, Henne et al. (2010) found the influence of local anthropogenic emissions to be very limited at PDM, and classified the station in the "mostly remote" category. Collaud Coen et al. (2018) defined an "ABL-Topoindex" as a metrics of the atmospheric boundary layer influence for a mountain site. PDM was has been found to have a low ABL-Topoindex, similar to other Alpine high altitude stations. PDM is thus a suitable site to study both the background lower free troposphere (FT) over long timescales and injection of air masses from the PBL (Hulin et al., 2019; Tsamalis et al., 2014; Fu et al., 2016; Marusczak et al., 2017). In this article the instrumentation and methodology are presented in section 2. Section 3 presents the results of the measurement campaign including the meteorology and air mass origin, the

This paper aims to provide comprehensive picture of the seasonal and diurnal variability of rBC properties at PDM, and to explore the processes driving these properties. Specifically, in the indicated sections, the following questions are addressed:

1. What are the air mass transport pathways impacting PDM?

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- 2. What is the seasonal variability of aerosol optical properties and black carbon properties. In Section 4, the possible factors influencing the variability of E_{abs} are discussed. dominant aerosol types? What is the specific contribution of rBC to aerosol absorption?
- 3. How do the microphysical and optical properties of rBC vary on a seasonal and daily basis?
- 4. What are the roles of wet deposition, source and transport pathway in driving rBC absorption?

2 Methods

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2.1 Measurement site and observation period

Measurements were performed at the mountain research station Pic du Midi (PDM, 4242.9° 56'11" N, 0N, 0.1° 08'34" E, 2877 m. above sea level) asl) mountain research station in the French Pyrenees. This station is part of the Pyrenean Platform for Observation of the Atmosphere (P2OA)¹.

As shown in Fig. 1, the site is located 150 km east of the Atlantic coast. The high isolated summit is shifted about lies around 20 km north of the main Pyrenean crest (ridge of the Pyrenees (on the France-Spain border) and thus , closely dominates closely overlooks the French plain. Long-term monitoring of extensive numerous meteorological, gas and aerosol parameters have has been conducted for mostly two decades, notably through the Global Atmospheric Watch (GAW) program of the World Meteorological Organization (WMO), as well as the national research infrastructure ACTRIS-France. Results from the Hygroscopic properties of Black Carbon (h-BC) campaign performed from February 2019 to January 2021 at PDM (in addition to the routine measurements) are presented in this paper.

2.2 Instrumentation

2.2.1 Total inlet

All particle-measuring instruments were sampling air drawn sampled air taken in parallel from a Whole Air Inlet, which is utilized whole air inlet, located 2 m above the building roof. This inlet is used for the long-term observations and sucked air 2 m above the building roof. Air in mountainous sites and designed to maintain an isokinetic and laminar flow. The main flow rate was fixed at about 460 l min⁻¹. The splitter was fixed at the end of the stainless tube. The hat of the whole air inlet and the stainless tube were both thermo-regulated in order to avoid frost and gradually regulate the temperature of the samples air to the measurement room. The air was heated to ≈around 20°C in order to keep the relative humidity below 20%(Nessler et al., 2003) perform aerosol in-situ measurements at a relative humidity lower than 30 %. The instrumental room temperature was regulated at around 20°C. The annual cycle of the dew point temperature varied between about -10° and +5°C.

2.2.2 Black carbon measurements

¹http://p2oa.aeris-data.fr

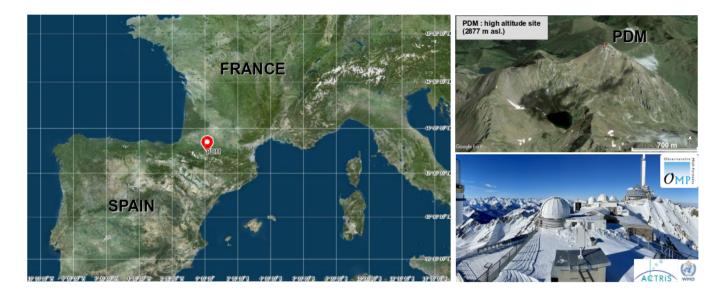


Figure 1. Geographical location of the Pic du Midi Observatory in the French Pyrenees (©IGN and ©Google Earth).

M_{BC} and BC BC can be measured by different methods which are based on different BC properties. Petzold et al. (2013) defined a specific nomenclature for BC according to the BC quantification method. Following the recommendation of the authors, BC quantified by laser-induced incandescence and thermal-optical analysis will be referred to as refractive black carbon (rBC) and elemental carbon (EC), respectively. More general discussion on BC without focusing on its measurement technique will be referred to as BC. The mass concentration of rBC and rBC size distribution were measured by a Single Particle Soot Photometer (SP2, Droplet Measurement Technology, BoulderLongmont, CO, USA). Its operating principles have been described in previous articles (Gao et al., 2007; Moteki and Kondo, 2007; Schwarz et al., 2006). In short, this instrument uses a laser-induced incandescence technique which quantifies BC-rBC mass in single particles. A continuous intracavity laser beam (Nd:YAG; λ =1064 nm) is used to heat BC-containing rBC-containing particles to their vaporization point. The measured incandescence signal of an individual BC-containing rBC-containing particle can be converted to a BC mass, which was calibrated using rBC mass, using a calibration curve obtained by recording the incandescence signal peak height of mobility size-selected fullerene soot particles (Alfa Aesar, lot #FS12S011) and assuming BC mass density of 1.8 g cm⁻³. This calibration was performed twice a year and did not evolve thorough change during the two-year of the measurement campaign.

The SP2 data were processed using a Python code following the method used in the SP2 Toolkit from the Paul Scherer Institute (Gysel et al., 2009). A comparison of M_{rBC} resulting from the SP2 Toolkit with our Python processing is presented in Text S1 in the Supplement. The SP2 used in this study measured BC rBC cores over a size range between 90 and 580 nm. The However, the observed size distributions showed an increase in Mthat an important fraction (around 12%) of M_{BCrBC} at diameters less below than 90 nm (Figure S1 is not measured by the SP2 (Fig. S2 in the Supplement). Because of these

small-mode particles below the SP2 detection window, the SP2 Mrange, the quantification of M_{BCrBC} measurements could be underestimated. To compensate the missing mass the observed $\frac{BC}{BC}$ size distributions have been fitted daily using the sum of three individual lognormal distribution to extrapolate $\frac{BC}{BC}$ size distribution in the range $\frac{10}{1}$ to 1000 nm. The position of the three modes were constrained in the following ranges: Mode $1:50 < \frac{dD_g}{dQ_g} < 100$ nm and $1.2 < \sigma_g < 3$; Mode $2:150 < \frac{dD_g}{dQ_g} < 250$ nm and $1.3 < \sigma_g < 2.9$; Mode $3:350 < \frac{dD_g}{dQ_g} < 500$ nm and $1 < \sigma_g < \frac{3}{3}$, with $\frac{D_g}{dQ_g}$ and $\frac{1}{3}$ the geometric mean diameter and the geometric standard deviation, respectively. Using the fitting procedure, a time-dependent missing mass correction was applied to the observed $\frac{1}{3}$ to calculate to overall $\frac{1}{3}$ the average missing mass correction factor applied over the campaign was 1.2 ± 1.1 (Mean value \pm STD). More details on the SP2 data procedure can be found in the Text S2 in the Supplement. The extent to which uncertainty in the uncertainty of this fitting procedure contribute contributes to the overall $\frac{1}{3}$ Merch was quantified by comparing the $\frac{1}{3}$ Calculated from the observed BC size distribution and the fit curve over the SP2 size range. An excellent match was obtained between the measured and fitted size distribution, resulting in differences by of less than 0.2 %. The combined uncertainty on the $\frac{1}{3}$ was estimated to be about 24.5 % by calculating the quadratic sum of the measurement uncertainties on sampling flow, anisokinetic sampling errors, and missing mass correction factor.

2.2.3 Aerosol properties

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A Scanning Mobility Particle Sizer (SMPS), combining a differential mobility analyzer (DMA, TSI model 3071, TSI Inc., Shoreview, USA) and a CPC (TSI model 3772, TSI Inc., Shoreview, USA) allows the determination of aerosols size distribution between 12.6 nm and 532.6 nm.

Aerosols Aerosol scattering coefficients (σ_{sca}) at three wavelengths (450 nm, 525 nm, 635 nm) were measured with an integrating nephelometer (model Aurora 3000, Ecotech Pty Ltd, Knoxfield, Australia). A calibration with carbon dioxide and filtered air was performed every three months. The instrument measures σ_{sca} in the angular range 10-170°, and the correction of Müller et al. (2011) was used to account for the angular truncation errors.

Aerosol absorption coefficients (σ_{ap}) were measured by a seven-wavelength aethalometer (model AE33, Magee Scientific Company, Berkeley, USA, measuring wavelengths: 370, 470, 520, 590, 660, 880, 950). This instrument measures light attenuation through a filter on which aerosol sample is deposited. The aethalometer filter loading effect was corrected online by the dual-spot manufacturer correction proposed by Drinovec et al. (2015). The multiple scattering artifact was corrected using a C value of 3.63, as obtained by Tinorua et al. (2023, in preparation) parameter used to correct the measured attenuation was set to 3.22, according to the value obtained at λ =880 nm by Yus-Díez et al. (2021) at the mountainous site of Montsec d'Ares located less than 200 km from the PDM. Uncertainty on the corrected σ_{ap} were was estimated to be 35 % (Zanatta et al., 2016). The detection limit of the aethalometer is 0.039 Mm⁻¹ (corresponding to an equivalent black carbon mass concentration of 0.005 μ g- μ g m⁻³ of M_{BC}, which corresponds to 0.0215 Mm⁻¹ in absorption. Values under this low limit were filtered out before the analysis).

2.2.4 Gas-phase measurements

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Two different instruments — have been deployed to measure carbon monoxide (CO) with a final 1-h time resolution time resolution of one hour: an IR-absorption analyser (TEI model 48CTL, TEI Thermo Environment Instruments, New Delhi, India) placed close to the aerosol instrumentation in order to detect pollution plumes produced locally at PDM ² (hourly CO concentrations above 200 ppb were filtered out), and a Cavity Ring Down Spectrometer (CRDSPicarro, model G2401, Picarro, Santa Clara, USA), located in an other building, used to measure the background carbon monoxide (CO) concentration and calculate ΔBC/ΔCO ratios(See Section 2.3).

A key issue in our study is the distinction between FT and PBL-influenced air masses. Optical properties of BC depends depend on its aging and transport pathways in the atmosphere, so that it is crucial to determine whether it has been transported over the BL PBL or in the FT. For this purpose, we routinely monitor the diurnal cycle of radon (222Rn) volumic activity (in mBq m⁻³) at PDM with a 1500-L high-sensitivity radon monitor manufactured by ANSTO (model D1500, ANSTO Australian Nuclear Science and Technology Organisation, Australia) (Whittlestone and Zahorowski, 1998). Radon is an inert radioactive gas emitted from ice-free soils with a half-life of 3.8 days, making it the most reliable tracer to discriminate between the FT and PBL- influenced air masses (Chambers et al., 2013).

2.3 Determination of intensive aerosol and BC properties

The spectral dependence of σ_{ap} was characterized by the Absorption Ångström Exponent (AAE_{aer,450-635}) as calculated between 450 and 635 nm as follows:

$$AAE_{aer,450-635} = \frac{-log\left(\frac{\sigma_{ap,450}}{\sigma_{ap,635}}\right)}{\left(\frac{log(450)}{log(635)}\right)} \frac{-log\left(\frac{\sigma_{ap,450}}{\sigma_{ap,635}}\right)}{\left(log\left(\frac{450}{635}\right)\right)}$$
(1)

For this calculation, $\sigma_{ap,470}$ and $\sigma_{ap,660}$ from the aethalometer were adjusted at the wavelengths of 450 and 635 nm measured by the nephelometer using the AAE calculated from the aethalometer between 370-470 nm and 590-660 nm. AAE_{aet,450-635} provides information about the chemical composition of atmospheric aerosols. Pure BC absorbs radiation across the whole solar spectrum with the same efficiency; thus, it is characterized by AAE_{aet,450-635} around 1 (Bond et al., 2013). Conversely light-absorbing organic particles known as brown carbon (BrC), as well as dust particles generally have an AAE_{aet,450-635} greater than 2 (Sun et al., 2007; Bergstrom et al., 2007; Schuster et al., 2016). Here the AAE_{aet,450-635} was calculated between the wavelengths of 450 and 635 nm, for which scattering coefficients are measured by the nephelometer. To do so, $\sigma_{ap,635}$ was $\sigma_{ap,635}$ which were not directly measured by the aethalometer, were first calculated using the intermediate calculation of AAE_{aet,370,470} and AAE_{aet,590-660}. Then, by using a rearranged form of equation (1) and replacing 470 by 450 and 660 by 635 $\sigma_{ap,450}$ and $\sigma_{ap,635}$ were derived from $\sigma_{ap,470}$ and $\sigma_{ap,635}$

²e.g. due to snow removal of the touristic platform

The wavelength dependence of σ_{sca} can be characterized by the Scattering Ångström Exponent (SAE_{aer,450-635}) calculated between 450 and 635 nm, as:

$$SAE_{aer,450-635} = \frac{-log\left(\frac{\sigma_{sca,450}}{\sigma_{sca,635}}\right)}{\left(\frac{log(450)}{log(635)}\right)} \frac{-log\left(\frac{\sigma_{sca,450}}{\sigma_{sca,635}}\right)}{\left(log\left(\frac{450}{635}\right)\right)}$$
(2)

The aerosols SAE_{aer,450,635} describes the relative contribution of fine and coarse mode particles (Clarke and Kapustin, 2010). Small values of SAE_{aer,450,635} indicate a higher contribution of large aerosol particles (e.g. dust and sea salt), while large values of SAE_{aer,450,635} indicate relatively smaller aerosol particles (Cappa et al., 2016). The aerosol Single Scattering Albedo (SSA_{aer, λ}) was calculated at the wavelengths of λ = 450, 525 and 635 nm using the following equation:

$$SSA_{aer,\lambda} = \frac{\sigma_{sca,\lambda}}{\sigma_{sca,\lambda} + \sigma_{ap,\lambda}} \tag{3}$$

For that purpose $SSA_{aer, \lambda}$ describes the relative importance of scattering and absorption to the total light extinction. Thus, it indicates the potential of aerosols to cool or warm the atmosphere. To calculate $SSA_{aer, \lambda}$, σ_{ap} was first calculated at the proper wavelengths λ using the AAE calculated at the closest wavelengths (AAE_{aer,370-470} to retrieve $\sigma_{ap,450}$, AAE_{aer,520-590} for the $\sigma_{ap,525}$, and AAE_{aer,590-660} for the $\sigma_{ap,635}$). $\triangle BC$

The ΔM_{rBC}/ΔCO emission ratio was calculated to provide information on the combustion sources, as well as on meteorological conditions BC wet deposition (Baumgardner et al., 2002). First, the background CO concentrations were estimated by taking the rolling 5th percentile of the values on a 14-day time window and then calculating a monthly mean (see fig \$2.83 in the Supplement) based on the method by of Kanaya et al. (2016). ΔCO was then calculated by subtracting the monthly background CO concentration to any measured hourly CO value. ΔBC was taken ΔM_{rBC} was considered to be equal to M_{BCrBC}, because we assume that the background BC is zero since the atmospheric lifetime of BC is known to be of a few days , which is much smaller than CO lifetime (1 or 2 months). (Park et al., 2005). By contrast CO lifetime is estimated at several days (Bey et al., 2001).

The Mass Absorption cross-section of BC (MACrBC (MAC_{BCrBC}) was determined as:

$$MAC_{\underline{BC}rBC} = \frac{\sigma_{ap,880}}{\underline{M_{BC}}} \frac{\sigma_{ap,880}}{\underline{M_{rBC}}}$$

$$(4)$$

M_{BCrBC} under (resp. over) below the 5th (resp. and above the 95th) percentile were filtered before MAC_{rBC} calculations to reduce the influence of outliers in statistical analyses. As shown in Tinorua et al. (2023, in preparation), the presence of dust can lead to strong overestimation of σ_{ap,880}. Therefore periods with dust In addition, we filtered out periods when dust were sampled at PDM were eliminated before the MACfor the calculation of MAC_{BCrBC} calculations following the method presented in section 3.2since Yus-Díez et al. (2021) observed significant biases in the multiple scattering correction of the aethalometer AE33 during such events.

The light-absorption enhancement factor E_{abs} can be determined as the MAC_{BCrBC} values value normalized by a reference value for pure, uncoated (bare) BC: rBC:

$$E_{abs} = \frac{MAC_{BC}}{MAC_{bare,BC}} \frac{MAC_{rBC}}{MAC_{bare,rBC}}$$
(5)

Three different methods are usually generally used to estimate MAC_{bare,BerBC}: the first one is to remove the coating of BC with a thermodenuder and measure the corresponding absorption (Cappa et al., 2012; Healy et al., 2015); the second one is to extrapolate measurements of MAC_{BerBC} as a function of the measured BC mixing ratio (Cappa et al., 2019); and the third one consists in calculating MAC_{bare,BerBC} from the measured BC rBC size distribution using Mie's theory and the mean geometric BC diameter (see fig. S3 in the Supplement) rBC diameter (Zanatta et al., 2018; Liu et al., 2017, see fig. S4 in the Supplement). Here we used this latest the latter method by assuming a BC rBC refractive index of $\frac{1.95}{1.95}$ - 0.79i at λ =880 nm (Bond and Bergstrom, 2006). The calculation of MAC_{bare,BC} using Mie's theory assume a simplified spherical assumption of rBC morphology. However rBC may exhibit complex morphologies whose optical behavior is imperfectly predicted by Mie's theory, introducing a bias in the retrieved MAC_{bare,BC} (Saleh et al., 2016). It might be considered that Mie's theory is suitable for estimating the absorption of highly aged rBC, which exhibit an internally mixed core-shell structure. China et al. (2015) used this method to calculate the E_{abs} of rBC in a high-altitude site of the Azores Islands because the large majority (70%) of these long-range transported particles were found highly compacted. Several studies found that Mie's scattering model captures basic optical properties of BC in biomass burning plumes (Liu et al., 2017; Denjean et al., 2020). Zanatta et al. (2018) calculated MAC_{rBC} of heavily coated rBC particles from the Arctic region using Mie's theory and found consistent results with direct measurements.

Time periods with high humidity (95%) or precipitations precipitation were filtered before analysis to avoid artifacts in the sampling inlet. Under precipitation some water droplet may indeed enter in the aerosol inlet and change both the inlet cut off diameter and the measured aerosol size distribution. This would bias all the measured aerosol properties. We also filtered periods where hourly CO concentrations exceeded 200 ppb in order to exclude local pollution events, e.g. due to snow removal of the touristic platform.

All aerosol and gas measurements were converted to standard temperature and pressure (273.15 K and 101.325_1013.25 hPa).

250 2.4 Identification of air mass origins

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The Hybrid Single Particle Lagrangian Integrated Trajectory Model (HYSPLIT) (Draxler and Hess, 1997) was used to calculate air masses backtrajectories. This model uses 3-hourly atmospheric data from the Global Data Assimilation System (GDAS) of the National Center for Environmental Prediction (NCEP) in a 1°×1° spatial resolution. More information can be found on https://www.ready.noaa.gov/index.php. One A backtrajectory was run every 24h going 72h back in time at 12h for the two-year period at 12:00 UTC for each day, going back 72 hours, for the two years of the campaign. Every backtrajectory arrived at the PDM altitude and coordinates. Precipitation rates along the back trajectories were also computed from the HYSPLIT

calculations, in order to classify days where when the air masses arriving at PDM encountered precipitation or not in the past 72h72 hours.

To discriminate FT and PBL-influenced air masses (hereafter referred as PBL/FT conditions), we followed the a methodology proposed initially by Griffiths et al. (2014), assuming that the diurnal radon increase, which is often typically observed at mountain sites during the daytime, is the result of transport of PBL air by thermal anabatic winds up to the summits. The method first consists in ranking each day the days of the sampling period by decreasing anabatic influence (no details on the iterative ranking process are given here, but can be found in Griffiths et al. (2014)). Then, a value called "anabatic radon" can be calculated for each day, which represents (in short) the average deviation of radon volumic activity above a nocturnal background (see again details details in the Supplement and in Griffiths et al. (2014)). Anabatic radon mostly decreases with increasing anabatic rank (Fig. S4 of the Supplement), at least up to a A threshold rank (here 282) corresponding to the absolute minimum of anabatic radon. After this rank, the radon variations are no more in phase with the diurnal thermal eycle, and may be due to any other causes than anabatic transport². For this reason, the threshold rank can be used to separate anabatically-influenced days from non influenced days. In the present study, when we needed to select hours with strong influence of see Fig. S5 and associated text in the Supplement) can then be determined to separate days with or without anabatic influence in the daytime.

In our study, it was necessary to select the observation hours strongly influenced by the boundary layer, we chose, among the 200 first. To do this, we selected the first 200 anabatic days in the ranking, and from these days, all the hours when the radon activity was higher than the daily median value. Conversely, hours without influence of PBL are selected among days after the rank greater than the median value for the current day.

We also needed an ensemble of observation hours with minimum influence of the PBL. In the latter case, we selected hours in the non-anabatic days (i.e. ranked after 282, when the radon activity was under the daily median value) with radon activity below the median value for the current day.

3 Results

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280 3.1 Meteorology and air mass classification

In the following, seasons are defined as follows: winter (December, January, February), spring (March, April, May), summer (June, July, August), and autumn (September, October, November). The meteorological conditions at PDM during the campaign were characterized by a strong seasonal trend of temperatures, with daily means ranging between -15 and +15°C (Figure 2Fig. S6 in the Supplement). The time series of relative humidity (RH) covered a wide range between 5% and 100% with an annual mean value of 71.2 %. Lower ambient RH was observed in summer compared to winter with median values of 67% and 78%, respectively. Irrespective of the season, the wind direction was dominated by westerly and south-westerly winds and a median speed of 7 m .s⁻¹.

²By construction, a value of "anabatic" radon can still be calculated but actually makes no more sense, explaining the random fluctuations after rank 282.

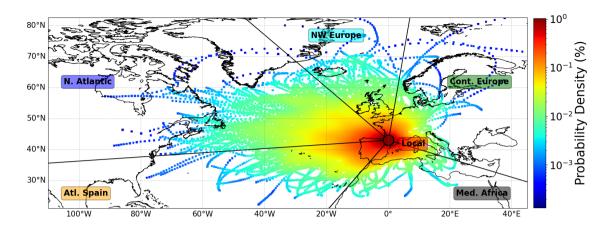


Figure 2. 72-h 72-hour Back trajectories of air masses measured at PDM over the measurement period 2019-2020. Geographical boundaries of the sectors used to classify the air mass back-trajectories are overlaid.

The backward trajectories performed with the HYSPLIT model on 72h-periods 72-hour time periods of time were assigned to six geographical zones, according to the position of their start starting point (shown in Figure 3Fig. 2): North-west Europe, Continental Europe, Med-Africa, Atlantic Spain, North Atlantic and Local (within a circular zone of 100 km radius around PDM). Transport The transport to PDM was generally westerly or southerlyfrom the west or south, from the Atlantic Oceantowards, North America or the Iberian Peninsula. It can also be noticed should also be noted that 99% of all atmospheric backward trajectories modelled to PDM reveal long-range transport (>100 km).

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295 conditions prevailing at the site following the methodology presented in Section 2.52.4. Details on the statistical results can be found in Table S1 in the Supplement. Over the campaign, 1149 hours were clearly identified as FT-influenced conditions, which represents 56% of the total classified hours. In winter, FT- and PBL-influenced FT and PBL conditions occurred roughly 74% and 26% of the analyzed time, respectively, against 48% and 52% for summer, respectively. These results are broadly in agreement with the previous study by Hulin et al. (2019) at PDM, which quantified around 47% of the days as PBL-influenced over a 10 years period. The PBL-influenced coditions occurred PBL conditions occurred mostly around 15:00 UTC (see Fig. S5 in the Supplement), consistent with the dynamics at mountain sites where plain-to-mountain winds and along-valley winds become the strongest in the afternoon (Whiteman, 2000).

Time series (left) and statistical distributions (median, 25th and 75th percentiles, right) of meteorological parameters measured at PDM in 2019-2020 with (a) the temperature and Hysplit air mass origin, (b) the relative humidity and (c) the wind direction and speed. Daily average data are shown.

3.2 Aerosol optical properties and classification of aerosol types

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Figure 4-3 presents daily time series and statistics of aerosol optical properties over the two-year measurement period. The average SSA \pm GSD (Geometric Standard Deviation) were 0.94 \pm 0.06, 0.94 \pm 0.07 and 0.95 \pm 0.08 at 450, 525 and 635 nm, respectively (Figure 4aFig. 3a). These values are in the range of those observed at mountain sites in Southern Europe (Bukowiecki et al., 2016; Laj et al., 2020; Pandolfi et al., 2014). The mean value \pm GSD of $\sigma_{ap,880}$ was 0.27 ± 0.25 Mm⁻¹, which falls in the range of 0.14 to 1.23 Mm⁻¹ obtained at Jungfraujoch and Montsec (Bukowiecki et al., 2016; Pandolfi et al., 2014). The average $\sigma_{sca} \pm \text{GSD}$ were 15.5 \pm 16.1 Mm⁻¹, 13.4 \pm 13.9 Mm⁻¹ and 12.2 \pm 12.9 Mm⁻¹ at 450, 525 and 635 nm, respectively (Fig. 3e). These weak values of σ_{abs} and σ_{sca} can be explained by the remote mountain site type, where almost no absorbing aerosols are locally emitted. There was a clear seasonality of SSA with a minimum around 0.93 during aerosol optical properties. SSA at the three wavelengths exhibited the lowest monthly mean values in spring-summer and higher values around 0.97 (0.94 \pm 0.02 at λ = 525 nm) and the highest in autumn-winter (0.99 \pm 0.01 at λ = 525 nm), as shown in Fig. 3a). Simultaneously, both SAE and $\sigma_{ap,880}$ increased by a factor 2 and 5, respectively, during the highest monthly mean SAE values were observed in spring-summer compared to autumn-winter, thus suggesting a higher influence of absorbing fine particles (1.23 ± 0.70) and reached a minimum in the winter (-0.25 ± 0.16) (Fig. 3b). This anticorrelation suggests a higher fraction of absorbing and fine particles relative to purely scattering and coarse particles at PDM during the warmest months (Figures 4b-d). Figure 4c shows a less pronounced AAE seasonal variation (1.13 ± 0.35), indicating a rather constant composition of absorbing aerosol particles spring-summer. Interestingly different trends can be observed between the summer and spring seasons. During spring 2019 the decrease of SSA is correlated with a slight enhancement of $\sigma_{ap,880}$ (Fig. 3d) and a decrease of σ_{sca} at all wavelengths. In summer the increase of $\sigma_{an,880}$ lead to values multiplied by a factor of four, while both SSA and SAE remained rather constant. All these parameters combined indicate a similar dominant aerosol type reaching PDM but with stronger contribution in summer. This is further confirmed by the simultaneous increase of M_{rBC} in summer shown in Fig. 5. This noteworthy annual seasonality of aerosol optical properties has previously been observed at other high mountain sites in Europe (Andrews et al., 2011; Collaud Coen et al., 2011; Laj et al., 2020; Pandolfi et al., 2018). It has been attributed to the seasonal variation of the continental boundary layer height, long-range transport events (e.g. Saharan dust outbreaks, coal burning from eastern Europe) and biomass burning both from forest fires in summer and domestic heating in winter.

A-The higher concentration of small and absorbing particles in summer at PDM could be attributed to a higher anthropogenic BC influence favored by strong vertical mixing and a higher PBL height, a higher occurrence of wildfires emitting large amounts of BC and Brown Carbon (BrC), or a lower precipitation rate.

In order to investigate these different hypotheses, a classification of the dominant aerosol type sampled at PDM was performed by using the spectral dependency of aerosol optical properties. Figure 5-4 shows AAE as a function of SAE, overlaid with the aerosol classification matrix from Cappa et al. (2016). Aerosol with the highest SSA values (violet points) tend to fall on the left-hand side of the plot with SAE values below 0, indicative of large particles such as marine sea salt, continental dust or highly processed/coated particles. The presence of large marine and dust aerosol is in line with backward trajectories showing a dominant origin of air masses coming from the Atlantic Ocean and Iberian Peninsula as far as North Africa (e.g. Figure

340 3Fig. 2). Dust being a strong light absorber, it is expected to lower the mean aerosol SSA. However Figure 5 Fig. 4 shows that SSA for dust-dominated aerosol (classified as having AAE values above 2) are quite similar as those for remote marine aerosol (classified as having AAE values below 1). Although Europe frequently experiences African dust events (Denjean et al., 2016; Dumont et al., 2023), our results indicate that these dust events were not absorbing enough to substantially lower the aerosol SSA at PDM. This is supported by previous estimates of SSA ranging between 0.90 –and 1.00 for dust particles transported in the Mediterranean region (Mallet et al., 2013; Denjean et al., 2016).

There was a natural clustering of the most light absorbing aerosols with SSA<0.9 (pink to yellow points) on the middle of the plot, with sections on the lower side with AAE between 0.5 and 1.5, which Cappa et al. (2016) defined as the sections dominated by BC or mixed with BC with large particlesthe sections mix of BC and large particles. The success of aerosol classification schemes is largely dependent on uncertainties in AAE attribution for each aerosol species. Although AAE = 1 is often considered for BC such as that in the classification by Cappa et al. (2016), observational and numerical estimates show a wide range of BC AAE from 0.6 to 1.3 (Kirchstetter et al., 2004; Liu et al., 2018) due to the variation of BC core size, coating thickness, composition and morphology (Liu et al., 2018; Zhang et al., 2018). Therefore it is possible that the large range of AAE values observed for the most light-absorbing aerosol were due to different microphysical and chemical properties of the sampled BC BC sampled at PDM.

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Interestingly almost none of the aerosol were classified as strong BrC and BC/BrC mixture (AAE and SAE values above 1.5), revealing a very low contribution of BrC to the aerosol absorption at PDM. An explanation could be the rapid BrC depletion within the first day after emission, by photobleaching or volatilization that has been observed in several studies (Forrister et al., 2015; Wong et al., 2019; Zeng et al., 2020). Altogether these results suggest that BC were was the predominant absorption component of aerosols at PDM and controlled the variation of SSA throughout the two observation years.

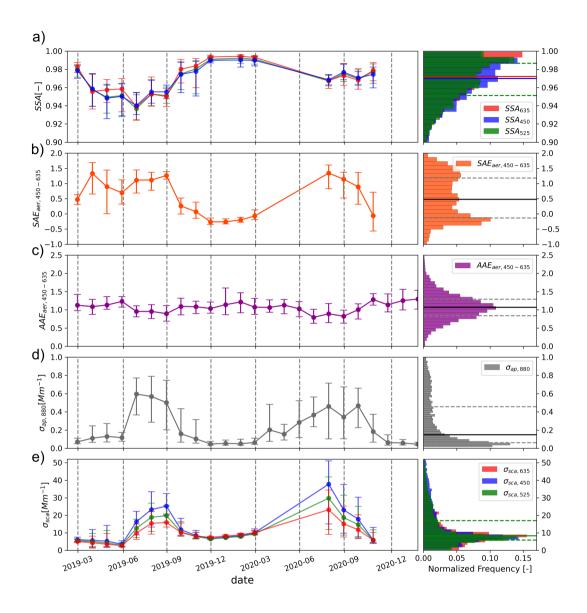


Figure 3. Time series (left) and statistical distributions (median, 25th and 75th percentiles, right) of aerosols optical properties measured at PDM in 2019-2020 with (a) Single Scattering Albedo at 450, 525 and 635 nm, (b) Scattering Angström Exponent at 450-635 nm, (c) Absorption Angström Exponent at 450-635 nm, (d) Absorption coefficient at 880 nm, (e) Scattering coefficients at 450, 525 and 635 nm. Daily average data are shown The dots and bars on the time series represent the median, the 25th and 75th percentiles, respectively, with a monthly frequency. Histograms was computed using a 1-day time frequency. Vertical dashed lines represent the seasons boundaries.

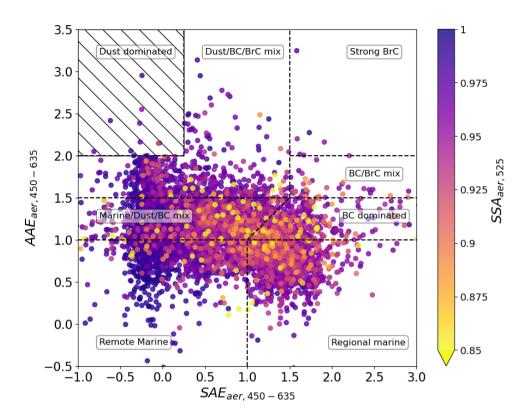


Figure 4. Hourly average Aerosols Absorption Angström Exponent vs. Scattering Angström Exponent calculated at 450-635 nm and colored as a function of the Single Scattering Albedo at 525 nm. The classification of aerosol type by Cappa et al. (2016) is also shown. The points in the dashed zone, representing dust events, were filtered before analyses of BC properties to avoid artifacts in the calculation of MAC_{BC-BC}.

360 3.3 BC rBC sources and properties Figure 6

rBC mass concentration

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Figure 5 shows the time series of the physical microphysical and optical properties of BCrBC. The mean M_{BCrBC} , shown in fig-Fig. 5a, was 34.8 \pm 35.7 (Mean \pm STD) ng m⁻³, which is a typical level for remote mountain sites. For instance, Sun et al. (2021) observed M_{BC} s-around 20 ng.m⁻³ from 9-years of measurements with a Multi-Angle Absorption Photometers (MAAP, model 5012, Thermo Scientific) at the Zugspitze-Schneefernerhaus station, Germany (2671 m a.s.l.asl). Motos et al. (2020) (Motos et al., 2020) measured Mmeasured M_{BCrBC} s-around 9 ng m⁻³ in summer at the JungfraujoghJungfraujoch, Switzerland (3580 m a.s.l.). Seasonal variations of M_{BC} (i.e. Fig. 6a) are similar to those of $\sigma_{ap,880}$ (i.e. Fig. 4d)with higher M_{BC} and $\sigma_{ap,880}$ in winter than in summer. BC asl). rBC represented around 7 \pm 5% of the total aerosol number concentration measured by the SMPS over the campaign. An increase of BC rBC number fraction by a factor 2.5 was found in summer (9 \pm 5%) compared to winter (4 \pm 3%). Simultaneously, $\sigma_{ap,880}$ increased by a factor 4 between winter and summer. Thus, it

confirms that BC rBC contributed to a significant part of the aerosol absorption at PDM.

Bivariate rBC emission sources

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Figure 6 shows bivariate polar plots obtained by combining hourly wind analysis and M_{BCrBC}s are shown in Figure 7 to investigate recent geographical origins of BC with 1-hour time resolution in winter and summer. Seasonal differences between the origin of highest MThe densities of M_{BCrBC} are thrown into relief, as well as the footprints of air masses backtrajectory densities plotted in Figure S5 in the Supplement. data weighted by M_{rBC} values, and normalized by the maximum M_{rBC} were plotted as a function of wind direction and speed. The darkest areas of the wind pattern are those where the highest M_{rBC} was measured with a high occurrence, whereas lightest zones exhibit lowest measured M_{rBC} and/or a little occurrence of measurements. Note that locally emitted pollution at the measurement station was filtered before the analysis, limiting local M_{rBC} contributions emitted from the PDM station (i.e. section 2.3).

In summer, the highest M_{BCrBC} was linked to winds values were mainly associated with moderate wind speeds (above 5 m s⁻¹) and from the west and south west, suggesting that regional transport from Atlantic Spain was an important source of BC (Fig7ba dominant regional transport (Fig. 6a). By contrast in winter (Fig7a. 6b), the highest M_{BCrBC} occurred mainly under more static atmospheric conditions (ie. for wind speeds under 10 below 5 m s⁻¹) and BC mostly came from the west and north-east highlighting different BC geographical sources. As a reminder, the locally emitted pollution at the measurement station was filtered before the analysis, limiting local Mno evident wind direction dependency. These results suggest that local-scale emissions could be a major contributor to M_{BCrBC} contributions emitted from the PDM stationin winter unlike summer. Further discussion on the role of PBL influence on M_{BCrBC} will be discussed in Section 4.2.

The $\triangle BC$ addressed in Section 3.4.2.

The $\Delta M_{rBC}/\Delta CO$ emission ratio, presented in Fig. 6b5b, shows a wide range of values from 0 to 10 ng m⁻³ ppbv⁻¹, with a mean value of 1.93 \pm 2.12 ng m⁻³ ppbv⁻¹. Summer values ratios were generally higher than winter emission ratios, which could reflect either lower BC rBC scavenging during transport or different emission sources of BC rBC between seasons. Indeed, Δ BC Δ M_{rBC}/ Δ CO emission ratio varies as a function of fuel types, combustion efficiencies and wet deposition by precipitations precipitation (Baumgardner et al., 2002; Taylor et al., 2014). This explains the high diversity of Δ BC Δ M_{rBC}/ Δ CO emission ratios obtained in the literature worldwide, going from 0.5 ng m⁻³ ppbv⁻¹ at the Jungfraujoch, Switzerland (Liu et al., 2010), to 9 ng m⁻³ ppbv⁻¹ in a biomass burning plume above Texas region during TexAQS 2006 campaign, USA (Spackman et al., 2010), if only studies using SP2 measurements are considered. Overall Δ BC Δ M_{rBC}/ Δ CO from fossil fuel tends to exhibit lower values to than those from biomass combustion (Guo et al., 2017; Pan et al., 2011; Zhu et al., 2019). To our knowledge the only available Δ BC Δ M_{rBC}/ Δ CO measurement measurements in Europe were performed during airborne measurements in the Cabauw industrial region, Netherlands, by McMeeking et al. (2010), who found very low values around 0.8 ng m⁻³ ppby⁻¹.

BC The high time variability of $\Delta M_{rBC}/\Delta CO$ reflects important differences between the scavenging processes impacting BC and/or the relative contribution of biomass burning and fossil fuel emissions in the production of BC measured at PDM.

rBC mass size distribution

rBC mass median core size diameter (D_{BCrBC,core}) was quite constant during the campaign with a mean geometric diameter of 179 ± 28 nm (Fig. 6e5c). An exception occurred in early December 2019, where we detected the presence of large BC rBC particles with D_{BCrBC,core} around 400 nm. However observations during this period may be the results of measurement uncertainties due to too low M_{BCrBC} (less than 10 ng m⁻³). The D_{BCrBC,core} values obtained at PDM are generally comparable to D_{BCrBC,core} that has been reported to range ranging from 180 to 225 nm for well-aged background BC rBC (Liu et al., 2010; McMeeking et al., 2010; Schwarz et al., 2010; Shiraiwa et al., 2008). However, our values are slightly higher than previous observations at Jungfraujoch by Motos et al. (2020) Motos et al. (2020) who reported D_{BCrBC,core} ranging from 130 and 150 nm in summer and winter. Instead of fitting the SP2 observations with a multimodal individual lognormal modes (e.g. Section 2.2.2), Motos et al. (2020) used a single-mode fit diameter approach which may bias the estimated SP2 D_{BCrBC,core} (Tinorua et al., in preparation).

rBC absorption

MAC(Zanatta et al., 2016).

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The ambient MAC_{BCrBC} was around 9.8.9.2 ± 2.7.3.7 m² g⁻¹ at λ=880 nm (Figure 6d)with systematically higher values in summer. This stands in the highest part of the range from 5.3 Fig. 5d). Several studies previously reported MAC_{BC} values between 8.9 to 9.5 m² g⁻¹ previously reported-13.1 for measurements at λ=637 nm in European mountain stations by Pandolfi et al. (2014), Yus-Díez et al. (2022) and Zanatta et al. (2016). However these (Pandolfi et al. 2014; Yus-Díez et al., 2022; Zanatta By using a AAE of unity, these values can be converted to MAC_{BC} between 6.4 and 9.5 m² g⁻¹ at λ=880 nm. These studies used different measurement techniquesand corrections factors, analysis method and correction factors from ours for estimating MAC_{BC} that eauses significant uncertaintymakes difficult the comparison of MAC_{BC} derived from different instruments. Pandolfi et al. (2014) performed a linear regression between σ_{ap.637} measured by a MAAP (Multi-Angle Absorption Photometer) and daily M_{EC} values from off-line filter-based measurements by a SUNSET OCEC Analyser. Yus-Díez et al. (2022) and Zanatta et al. (2016) retrieved MAC_{BC} with these instruments by calculating the ratio between the two parameters instead of a linear regression. Because of the absence of a standard method for quantifying M_{BC}, the absolute uncertainties on the literature MAC_{BC} obtained in the literature are very high ranging from ±30 to 70%, making difficult the comparison of

In terms of seasonality we found systematically higher values of MAC_{IBC} in summer (monthly mean \pm STD of 10.3 ± 3.3) compared to winter (8.3 \pm 3.8). Similar seasonal pattern was observed in Europe at Puy de Dôme (central France) and at Jungfraujoch (Swiss Alps) mountain sites (Sun et al., 2021; Motos et al., 2020; Zanatta et al., 2016). An opposite trend was observed at mountain sites affected by strong precipitation during monsoon such as the Tibetan Plateau and Himalayas regions where both MAC_{BC} and M_{BC} derived from different instruments (Zanatta et al., 2016). exhibit maximum values in winter or autumn (Zhao et al., 2017; Srivastava et al., 2022). The same seasonal pattern with elevated values in winter/autumn compared to summer/spring was observed at several rural and urban sites in the PBL, which was attributed to greater emissions

from residential heating combined to a lower PBL height (Zanatta et al., 2016; Kanaya et al., 2016; Yttri et al., 2007). However maximum of MAC_{BC} and M_{BC} in the PBL during cold periods is not a recurring observation even for a same measurement site. For instance Sun et al. (2022) showed in Beijing (China) that, due to the reduction of some predominant BC sources in winter consecutive to environmental policies, the annual cycle of M_{BC} changed over the years between 2012 and 2020.

Variations in MAC_{BCrBC} may exist for different reasons. We first addressed the question of whether the MAC_{BCrBC} depends on D_{BCrBC,core} in Figure S6-Fig. S8 in the Supplement. There was no clear correlation between MAC_{BCrBC} and D_{BCrBC,core}, which indicates that the variation in BC size was not the cause of the MAC_{BCrBC} variability. This is because D_{BCrBC,core} only varied within a relatively narrow range (percentiles the 25th and 75th percentiles around 164 and 195 nm) during the campaign. The observed MAC_{BCrBC} values were converted to equivalent E_{abs}, by dividing them by a reference MAC for pure uncoated BC (MAC_{bare,BC}). While values of MAC_{bare,BC} are reported in the literature, estimation of campaign-specific MAC_{bare,BC} allows for more robust determination of E_{abs} than using values from the literature since MAC_{bare,BC} is dependent of the size of uncoated BC (Bond and Bergstrom, 2006; Adachi et al., 2007, 2010; Adachi and Buseck, 2013; Cappa et al., 2012). Here MAC_{bare,BC} had an average value of 4.15 m² g⁻¹ with values ranging from 3.90 to 4.37 m² g⁻¹ considering the standard deviation of the mean D_{BCrBC,core}, which is in reasonable agreement with literature assessments (Liu et al., 2020b).

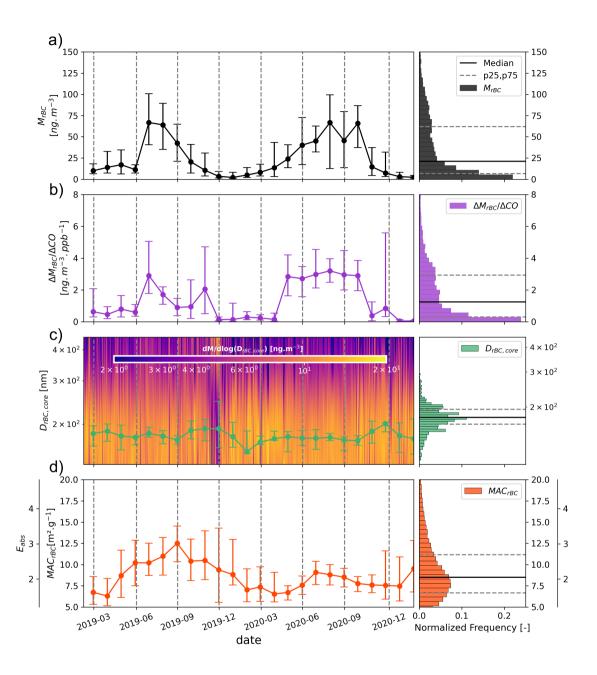


Figure 5. Time series (left) and statistical distributions (median, 25th and 75th percentiles, right) of BC-rBC properties measured at PDM in 2019-2020. (a) BC-rBC mass concentration, (b) ΔBCΔM_{rBC}/ΔCO emission ratio, (c) BC-rBC core mass size distribution with geometric diameter on in green solid line, (d) BC-rBC Mass Absorption Cross-Section and Absorption Enhancement at 880 nm. Daily average data are shown The dots and bars on the time series represent the median, the 25th and 75th percentiles, respectively, with a monthly frequency. Histograms was computed using a 1-day time frequency, as well as the colored background of the rBC core size distribution. Vertical dashed lines represent the seasons boundaries.

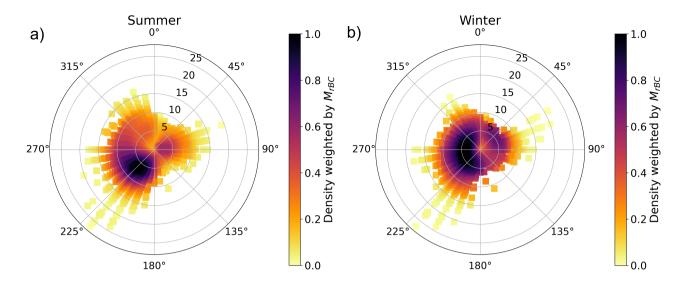


Figure 6. Bivariate polar plots of hourly mean BC concentration measured by the SP2 density of M_{rBC}, weighted by M_{rBC} values as a function of wind direction and speed in a) winter summer and b) summer winter. The colour scale shows the M_{rBC} density data weighted by M_{rBC}. The radial scale shows the wind speed, increases from the centre of the plot radially outwards. Both plots use hourly data. The weighted densities was normalised by their maxima.

The E_{abs} values derived from MAC_{BC} observations were significantly greater than unity with mean value of 2.4 2.2 ± 0.7

455 (Figure 6d0.9 (Fig. 5d)). Given the remote mountain location and presumable distance from fresh BC sources, it is expected that the BC rBC sources, rBC particles reaching PDM would be aged and relatively thickly coatedmay have undergone aging and have gained a consistent coating. Previous studies found an absorption enhancement of BC due to its coating with the aging time (Yus-Díez et al., 2022; Sedlacek et al., 2022; Peng et al., 2016). The most likely reason for the strong E_{abs} at PDM is a lensing effect due to the internal mixing of BC rBC with other particles that drives MAC_{BCrBC} variability, though we cannot eliminate changes in BC rBC morphology that can result from coating onto BCrBC. There was a significant seasonal trend in E_{abs} with higher values observed in summer, indicating that BC rBC reaching the PDM station has undergone longer aging processes during this season. These results are consistent with the measurements of Motos et al. (2020) at the Jungfraujoch, which also indicated a strong seasonality in BC rBC mixing state with larger coating in summer. Figure 8

Figure 7 further shows the diurnal variation of E_{abs} for every seasons. There was a remarkable notable opposite diurnal profile between seasons in E_{abs} with midday showing a minimum around 1.7 in winter, and a maximum around 2.9 in summer. Spring and autumn showed intermediate patterns with less regular E_{abs} throughout the day. These

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Taken together, these observations suggest that different sources and/or processes drove the seasonal contrast in BC rBC properties. The following section aims at investigating potential drivers of E_{abs} variations, including BC rBC wet scavenging, dominant BC rBC sources and transport pathways. Particular attention will be paid to winter and summer because these seasons differ greatly, whereas spring and autumn appear intermediate.

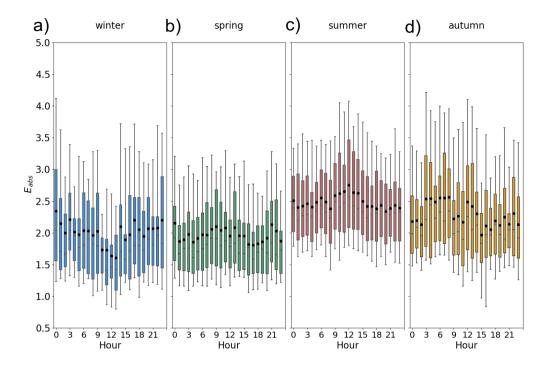


Figure 7. Diurnal cycles of E_{abs} for each season during 2019-2020 period. Winter includes Seasons are defined as follows: winter (December, Januaryand, February), spring eovers (March, Apriland, May), summer includes (June, Julyand, August), and months from autumn (Septemberto, October, Novemberare grouped together in autumn). Boxes, lines, black dots and whiskers indicate 25th percentile, 75th percentile, median, mean, 10th percentile and 90th percentile, respectively.

4 Investigation of factors influencing BC properties

3.1 The impact Investigation of BC wet scavenging on BC factors influencing rBC properties

3.1.1 The impact of wet scavenging on rBC properties

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We first investigated whether E_{abs} was modulated by a size-dependent BC_{TBC} wet scavenging process during precipitation along their transport pathway. This hypothesis is based on the fact that the removal of particles is favored for the largest and thickly coated BC_{TBC} because the activation of aerosols to cloud droplets is predominantly controlled by the particle size (Moteki et al., 2012; Motos et al., 2019; Ohata et al., 2016; Zhang et al., 2021). The wet removal of BC_{TBC} was investigated by performing a cluster analysis using $\Delta BC_{\Delta M_{TBC}}/\Delta CO$ data for which precipitation occurred or not along $\frac{72h}{72}$ -h back trajectories computed by the HYSPLIT model. Figure 9a shows significantly lower ΔBC Wet scavenging is expected to have a smaller impact on fresh rBC injected from the PBL than on rBC transported in the FT. Therefore, periods for which the site was under PBL influence were filtered.

Figure 8a shows median $\Delta M_{rBC}/\Delta CO$ of 2.1 ng m⁻³ ppbv⁻¹ for air masses affected by precipitationsprecipitation, against 0.7 ng m⁻³ ppbv⁻¹ without precipitation during the transport of the air masses. The reduction of $\Delta BC\Delta M_{rBC}/\Delta CO$ by $\sim 40\%$ a factor of three suggests that a significant removal process of BC rBC from the precipitation occurred long along the transport pathway, apart from vertical transport from the PBL. This result is confirmed by the dependence of $\Delta BC\Delta M_{rBC}/\Delta CO$ to RH in Figure 9bFig. 8b, where a sudden decline of $\Delta BC\Delta M_{rBC}/\Delta CO$ appeared for highest RH>80%. Figures 9, going from median $\Delta M_{rBC}/\Delta CO$ between 2.0 and 2.4 ng m⁻³ ppbv⁻¹ for RH<80% to a median $\Delta M_{rBC}/\Delta CO$ of \sim 0.4 ng m⁻³ ppbv⁻¹ above 80% of RH. This high rBC removal by wet deposition result is in line with measurements performed in regions at similar altitudes, such as Puy de Dôme, Mt. Nanling, and Mt. Sonnblick, where wet deposition represents 30 to 70 % of the BC removing processes in the troposphere (Yang et al., 2019).

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Figures 8 c-d show in contrast little influence of precipitation and RH on E_{abs} . Furthermore, the resultant BC after precipitations exhibited similar core size, as shown in Fig. S7 in the Supplement. Therefore we conclude that BC the rBC absorption enhancement, with a constant median E_{abs} value of around ~ 2.1 .

To better understand the negligible impact of rBC wet scavenging on MAC_{rBC}, we compared the measured rBC core size distribution of air masses affected or not by precipitation during their transport and under high RH conditions or not (Fig. 9). A two-fold lower M_{rBC} in precipitation conditions compared to that without precipitation provides additional evidence for the dominant role of wet scavenging for rBC. The same result appeared by comparing rBC core size distribution under wet or dry conditions. However wet scavenging did not significantly affect the size of BC-containing particles to drive changes in E_{abs}alter the modal diameter of rBC core size distribution.

This result contradicts contrasts with previous studies showing a decrease in BC rBC size due to wet scavenging (Kondo et al., 2016; Moteki et al., 2012; Taylor et al., 2014; Liu et al., 2020a). It could be explained by the size of BC rBC core sampled at PDM that was higher than the one described in these studies. Hoyle et al. (2016) evidenced at the Junfraujoch a threshold diameter of around 90 nm above which a particle activates to a droplet upon cloud formation. The majority of BC rBC sampled at PDM exhibited D_{BCrBC,core} above this critical diameter. Droplet In addition droplet activation of an aerosol particle occurs when the supersaturation of the surrounding water vapor exceeds a critical value of supersaturation. Thus, it is not likely for freshly emitted BC particles to act as cloud condensation nuclei due to their hydrophobic nature unless the water vapor supersaturation is higher than 2% (Wittbom et al., 2014), far beyond the actual supersaturation (0.1–0.6%) in ambient air. As the ambient supersaturation varies depending on the environment, it is difficult to conclude whether the insensitivity of BC rBC size distribution and E_{abs} to precipitation occurrence during the 72-h air mass history was solely due to the presence of large BC particles rBC particles at the sampling site or to a high ambient supersaturation in the precipitating clouds. Further measurements of simultaneous BC-rBC wet removal and effective supersaturation are needed to test these assumptions. two assumptions.

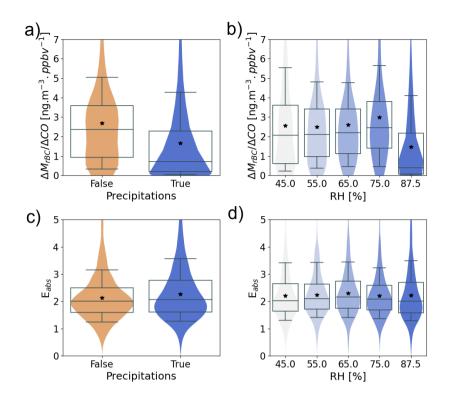


Figure 8. $\Delta_{tBC/\Delta CO}$ emission ratio and E_{abs} vs. a) and c) precipitation along the air mass back trajectory calculated with HYSPLIT model and b) and d) Relative Humidity measured at PDM. Violin plots represent the probability density function of each parameter. Statistics of the boxplots are the same as Fig. 7, PBL conditions were filtered before analysis.

3.2 The contrasted seasonal influence of FT and PBL on BC properties

Figure 10 shows the BC

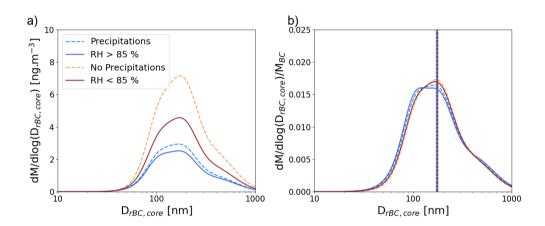


Figure 9. \triangle BC/ \triangle CO emission ratio and E_{abs} vs. a) and c) precipitations along the air mass back trajectory calculated with HYSPLIT model Mass size distributions and b) and d) Relative Humidity normalized mass size distributions of rBC core measured by the SP2 as a function of the presence or not of precipitation along the path of airmasses arriving at PDM, and whether the relative humidity was over 85 % or not. Violin plots represent The normalization was done with the probability density function of each parameter total rBC measured mass. Statistics of Vertical lines show the boxplots are geometric mean rBC core diameter colored by the same as Figcriteria described in the legend. 7.

515 3.1.1 The contrasted seasonal influence of FT and PBL on rBC properties

Figure 10 shows the rBC properties classified by FT- and PBL conditions FT and PBL conditions (methodology in Section 2.5) and by seasons. Air-As explained in Section 3.3, $\Delta M_{rBC/\Delta CO}$ ratio depends on the condition of combustion (fuel type, efficiency) and wet deposition by precipitation (Baumgardner et al., 2002; Taylor et al., 2014). We observed in Section 3.4.1 a large decrease of $\Delta M_{rBC/\Delta CO}$ when precipitation occurred during the transport of the air masses. In order to investigate the influence of rBC sources on rBC properties, precipitation events (air masses for which precipitation occurred along 72-h back trajectories computed by the HYSPLIT model where removed for this analysis) were removed in this section.

In winter 75 % of the M_{BCrBC} values were found under 105 ng m⁻³ under PBL influence, whereas under in PBL conditions, whereas in FT conditions they were under 45 ng m⁻³ (Figure 10aFig. 10a). Furthermore the diurnal cycle of M_{BCrBC} in winter PBL influenced conditions showed conditions showed an enhancement in the daytime (Figure S8-Fig. S9 in Supplement). This trend is consistent with intrusions of pollutants transported from PBL sources through convective mixing. During the night, the nocturnal depressed PBL in the valleys and the plain trapped the surface pollution below the mountain summits, and the subsiding cleaner air in the FT may have diluted the BC quantity at PDM lead to a decrease of rBC concentration at PDM by dilution. The higher Δ BC Δ M_{BC}/ Δ CO in PBL conditions than FT conditions PBL conditions than FT conditions may indicate additional sources from biomass combustion (Fig 10efrom the valley (Fig. 10c), which could be attributed to either residential wood heating or stubble-burning that is still a common practice in the Pyrenees (González-Olabarria et al., 2015). Figure 10b shows that 10b shows that PBL conditions were associated with lower E_{abs} was modulated by the atmospheric dynamic in winter with lower values in PBL-influenced conditions than FT-influenced abs values than FT conditions. Thereforethe significant decrease of E_{abs} observed at noon in winter (Figure 8a) might be the result of BC particles directly uplifted from the PBL, which have undergone shorter aging processes and less coating than BC transported into the FT, E_{abs} was strongly modulated by atmospheric dynamics in winter.

During summer vertical transport from the PBL occurred about half of the days analyzed in this study. Surprisingly these the thermally driven PBL injection did not significantly impact $M_{BC(Figure\ 10drBC)}$ measured at PDM (Figure 10d). This contrasts with our winter observations and most previous surface measurements at mountain sites, where the daytime PBL development had has been shown to enhance aerosol mass concentration (Herrmann et al., 2015; Venzac et al., 2009). ΔBC The summer M_{BC} values at PDM are twice as high as those observed in winter, which indicates a massive additional regional transport of rBC in the FT and a lower contribution of rBC from PBL injection. $\Delta M_{tBC}/\Delta CO$ in BL—and FT-conditions PBL and FT conditions were close to each other, with values around 2.8 ± 1.6 ng m⁻³ .ppbv⁻¹ (Mean \pm STD) and 3.3 ± 1.7 ng .m⁻³ ppbv⁻¹, respectively (Figure 10f), suggesting Fig. 10f). This result indicates that the FT has exhibited a significant background load in BC of rBC at the continental scale, which limits thus limiting the relative influence of PBL injection on $M_{BC,BC}$ during summer. The resulting E_{abs} was remarkably similar for PBL vs. FT air mass categories (Figure 10e). This is an evidence that the background FTmay be greatly influenced by biomass burning Fig. 10e). The high rBC loading transported in the FT coupled with the higher $\Delta M_{tBC}/\Delta CO$ observed in the summertime (Fig. 10d and e), could be due to a strong influence of biomass burning emissions on the background FT in Europe. For example (Petetin et al., 2018) have shown. The majority of

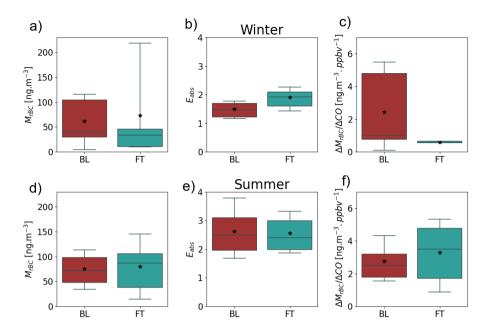


Figure 10. (a) BC rBC mass concentrations, (b) MAC_{BC} E_{abs} and (c) Δ BC Δ rBC/ Δ CO emission ratio as a function of the predominant influence at PDM in winter. The same for summer are given in d), e) and f). Red boxplots represents Boundary Layer PBL conditions and green boxes are Free Tropospheric FT conditions. Precipitation events were filtered before analyses.

trajectories reaching PDM in summer have crossed the Iberian Peninsula and, previously, North Africa and North America (Fig. S7 in the Supplement). In these regions large fire events frequently occur, which may explain the high concentrations of strongly absorbing rBC observed at PDM during summer. This hypothesis is supported by Petetin et al. (2018) who showed that biomass burning aerosol accounts for about 43 - 81% of the CO concentration in lower FT in summer using in situ in-situ airborne observations of CO from the IAGOS (In-service Aircraft for a Global Observing System) program. The ubiquitous presence of dilute biomass burning in the FT and its significant contribution to aerosol mass loading was also established using airborne measurements of ozone and precursor source tracers from the NASA Atmospheric Tomography mission (Bourgeois et al., 2021; Schill et al., 2020). These findings are consistent with the higher ΔBC/ΔCO observed in the summertime at PDM, indicating potential additional sources from biomass burning. Given that the majority of trajectories reaching PDM in summer traveled over the Iberian Peninsula (Fig. S5 in the Supplement), and prior to this, as far as North Africa and the North America where large fire events frequently occur, it is a possible explanation for stable concentrations of highly absorbing BC observed at PDM during summer. Additional measurements of the aerosol chemical composition and a precise source apportionment in particular of a tracer of biomass burning in the atmosphere such as levoglucosan should be performed at PDM to confirm this.

A question remains about the cause of the diurnal variation of E_{abs} in summer (Figure 8eFig. 7c). As shown in Figure S9Fig. 11, the E_{abs} increase was not temporally correlated with the wind direction change from West-South-West to South, as evidenced by the 2-h delay between the two events. Furthermore, while the E_{abs} increase occurs when $\triangle BC \triangle M_{rBC}/\Delta CO$ decreases, the E_{abs} drop in the afternoon was not accompanied by an increase in $\triangle BC \triangle M_{rBC}/\Delta CO$. Then increase of E_{abs} in the morning was most likely due to further aging and becoming heavily coated BC the appearance of heavily coated rBC rather than a change in BC rBC emission source. Several studies highlighted the major role of photochemical processing processes and extensive secondary aerosol generation to promote the light absorption enhancement of BC (Knox et al., 2009; Krasowsky et al., 2016; Liu et al., 2019; Wang et al., 2017; Xu et al., 2018; Yus-Díez et al., 2022). At PDM the enhanced E_{abs} at noon was accompanied by a strong elevation of particle number concentration in the diameter range 10-30 nm and a shift of the aerosol accumulation mode towards larger sizes, thus revealing both new particle formation and condensation onto preexisting which may be due to the condensation of gaseous species on aerosol particles (Figure S10 in the Supplement). This evidence imply the potential role of photochemical processing on BC in regulating the diurnal dynamics of E_{abs}Simultaneously, a strong elevation of particle number concentration in the diameter range 10-30 nm can be observed, revealing new particle formation most likely produced by photochemical reactions at this time of the day. It is thus possible that rBC particles became more coated via condensation of species produced by photochemical reactions at noon. However, it cannot be ruled out that the evolution of aerosol size distribution is a poor indicator of the rBC mixing state.

4 Summary and implications for climate models

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Continuous two-year measurements of BC refractive BC (rBC) properties and additional aerosol characteristics have been performed at the high-altitude mountain site Pic du Midi in the French Pyrenees. The classification of the dominant aerosol type using the spectral aerosol optical properties indicates that BC rBC is the predominant absorption component of aerosols at PDM and controls the variation of SSA throughout the two years. The lower SSA in summer (~0.93) than winter (in winter (~0.97) is correlated with a higher BC number fraction. AAE values around 1.13 ± 0.35 indicates a negligible rBC number fraction, whereas the influence of BrC to the aerosol absorption properties. SSA for dust-dominated aerosol were quite similar as those for remote marine aerosol, indicating that dust particles were not absorbing enough to substantially lower the aerosol SSA at PDM.

MAC strongly regulates BC radiative forcing, heating effect and interactions with snow. It is and dust was found negligible. One key parameter to constrain BC absorption and associated radiative forcing in climate models is the refractive index of BC, and in particular the resulting MAC_{BC}. It was not clear if BC at high-altitude mountain sites should have a thicker or thinner coating than in urban or plain sites or even should be coated at all. On the one hand, the longer BC lifetime and the low temperature in the FT free troposphere (FT) favor thicker coating due to enhanced condensation of low-volatility compounds in colder environment. On the other hand, the low concentrations of particles and gaseous precursors in the FT may limit the coating processes. Our two-year long observations show that the overall net effect is a strong absorption enhancement with a mean E_{abs} value of $\frac{2.4}{2.2} \pm \frac{0.7}{0.00}$.

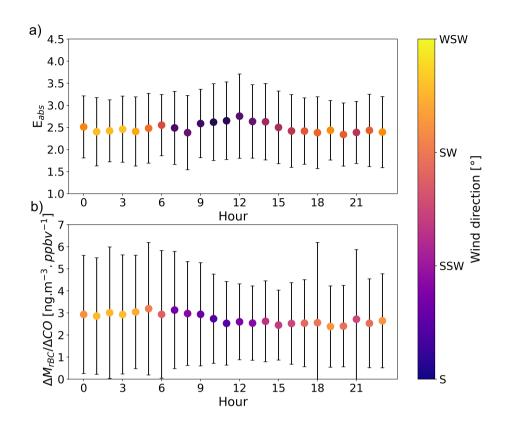


Figure 11. Hourly variation of (a) E_{abs} and (b) $\Delta M_{rBC}/\Delta CO$ values in summer. Dots represent mean values and whiskers are one standard deviation. Dots are colored as a function of the wind direction.

A significant reduction of $\Delta BC0.9$.

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The value of 7.5 m² g⁻¹ at λ = 550 nm of Bond and Bergstrom (2006) is the most common MAC_{BC} used in climate models. The recommendation was based on a compilation of experimental results for freshly generated BC at and near sources obtained earlier than the early-2000s. Nevertheless this value is largely under the MAC_{rBC} found in this study (9.2 m² g⁻¹ at λ = 550 nm, which can be converted to 14.7 m² g⁻¹ at λ = 550 nm assuming AAE=1). The review by Moteki (2023) has also come to a similar conclusion. The reasons behind this bias should be better understood, in the light of observations such as those provided in the present study.

This study has notably shown the high variability of rBC properties measured in a remote site, where they have undergone long-range transport and aging. Certain causes of the large variability in MAC_{rBC} have been eliminated and highlighted:

- Wet deposition is regarded as the main sink of BC, constraining its lifetime and size distribution, and thus its atmospheric concentration and optical properties. Our direct $\Delta M_{rBC}/\Delta CO$ when precipitation occurred along the air mass transport

suggests wet removal of BC. However the wet removal process has been found not to affect the size of BC-containing particles, resulting in unchanged E_{abs}. This contrasts with previous observations showing a preferential removal of BCwith large sizes (Kondo et al., 2016; Moteki et al., 2012; Taylor et al., 2014; Liu et al., 2020a). The difference measurements show the important role of wet deposition as a sink of rBC with around 67 % removed in the atmosphere by precipitation. However, we found a negligible impact of rBC wet removal process on both rBC size distribution and E_{abs}. This result may be due to either larger BC particles sampled at the combination of large rBC particles reaching PDM (D_{BCrBC,core} around 180 nm) or different meteorological conditions such as ambient and high critical supersaturation in precipitating clouds. The BC wet removal process was found to be one of the most misrepresented process in the representation of BC in models (Textor et al., 2006; Yu et al., 2019), leading to overestimated BC tropospheric concentrations and lifetime and in fine, a higher simulated radiative forcing (Samset et al., 2014; Schwarz et al., 2013). Substantial controversial and ambiguous issues in the wet scavenging processes of BC are apparent in current studies (Yang et al., 2019). Our results suggest that a bulk wet deposition parameterization (which does not account for particle size dependent scavenging) could realistically represent the actual BC wet scavenging at this site.

In addition, a large seasonal contrast in BC properties has been discovered, with higher M_{BC} and E

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- rBC core was found to have a mean D_{rBC,core} of 179 nm ± 28 nm, being reasonably independent of the season and day. There was no clear relationship between MAC_{rBC} and D_{rBC,core}, which indicates that the variation in rBC core size was not responsible for the MAC_{rBC} variability. Similar observations of rBC core size distribution in the atmosphere provided observational evidence of the stable distribution with a mode centered of around 200 nm approximately one day after emission (Liu et al., 2010; Schwarz et al., 2010; Shiraiwa et al., 2008). This self-similarity could greatly simplify the representation of MAC_{BC} in model simulations since a description of BC mixing state becomes the determinant factor of model performance when estimating BC optical properties and radiative forcing.
 - Different time scales of air movements and atmospheric processes affect MAC_{absin summer than winterBC} throughout the year. MAC_{IBC} values were found higher in summer (geometric mean of 10.3 m² g⁻¹), when the influence of regional-scale motions dominates the rBC load, than in winter (geometric mean of 8.3 m² g⁻¹), when the influence of local-scale motions outweighs the rBC load. There are three possible explanations for this. (i) The plumes traveling in the FT tend to have a longer lifetime providing sufficient time for rBC aging during transport. In winter this results in a strong diurnal variability of M_{BCrBC} (E_{abs}) with higher (lower) values in the middle of the day was linked to the injection of BC rBC originating from the PBL into a clean free troposphere. Many rural regions in the Pyrenees rely on wood burning for home heating during this season that may provide a significant source of BC at PDM. By contrast, during summer, the diurnal variation of M_{BC} was rather constant despite more frequent PBL conditions, implying that Mplanetary boundary layer (PBL). However the aging timescale can not be the only explanatory factor since thermally driven PBL injection did not significantly impact M_{BCRuctuations} were rather dominated by regional and long-range transport in the FT. Evidence suggests that biomass burning emissions from fires effective and E_{abs} was opposite to that in winter with maximum values of ~ 2.9 observed at noon. We suggest that this daily

variation results from photochemical processing driving BC mixing state rather than a change in BC emission source. Additional measurements of the chemical composition in both the gas- and particle-phases are required to confirm this. All these results on BC properties and notably their consequences on aerosol absorption are essential to better understand the role of BC in the climate system. In the light of the results obtained in the present study, BC absorption might not be correctly represented in many climate models. In particular, wet removal is an important process for climate modeling by moderating BC number and mass concentrations, lifetime and vertical distributions. Results here show that this wet removal of BC should be independent of the BC size and its mixing state, implying that the hydrophilicabs in summer and higher values have been observed in summer than in winter for similar FT conditions. (ii) The source of rBC emission was different between the winter and summer seasons. Combining ΔM_{rBC} /hydrophobic separation that could be done in some models may not be appropriate. Besides, our results also tend to show that it would be difficult for a climate model to correctly represent the MAC_{BC} and its variations without properly taking into account the diversity of sources and the different processes leading to its aging during transport Δ CO with air mass transport analysis, we observed additional sources from biomass burning in summertime leading to higher M_{rBC} and E_{abs}. (iii) Different aging processes occur between seasons, such as photochemical activity that could explain the observed amplification of light absorption by rBC around noon.

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The complexity and diversity of BC mixing states in the real atmosphere cannot be represented in climate models, and therefore these models generally use simplified schemes. A fixed e-folding timescale (1–3 days) is commonly used as the turn-over time for converting fresh BC particles into aged ones (Myhre et al., 2013). In addition, atmospheric models necessarily approximate the full complexity and diversity of BC composition, which can lead to mismatches with observed E_{abs} (Fierce et al., 2020). The findings presented here suggest that different dynamic processes governing rBC light absorption occur during the day and night, and between summer and winter. A parameterization of BC aging explicitly based on aerosol microphysical processes, in which the conversion rate is considered to vary depending on the environmental conditions (e.g., temperature, photochemical activity,...) and some key species (e.g., aerosol, coating precursors,...) may be required to adequately represent the true variability of MAC_{BC}.

Data availability. Aerosol microphysical and optical properties are freely available at http://ebas.nilu.no/ (NILU, 2018). CO data are available on the ICOS platform at https://www.icos-cp.eu/. rBC data are available upon request to the authors.

Author contributions. CD and ST designed the study, developed the analysis protocols, and wrote the initial manuscript. PN contributed to the data analysis. FG and VP provided data and methods to analyse them. TB, FG, EL, ST, VP and CD contributed to the measurement campaign. All authors reviewed the final manuscript.

Competing interests. The authors declare that they have no conflict of interest.

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