

1 **Validation and field application of a low-cost device to measure CO₂**
2 **and ET fluxes**

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15 **Abstract**

16 Mitigating the global climate crisis and its consequences, such as more frequent and severe droughts, is one of the major
17 challenges for future agriculture. Therefore, identifying land use systems and management practices that reduce greenhouse
18 gas emissions (GHG) and promote water use efficiency (WUE) is crucial. This however, requires accurate and precise
19 measurements of carbon dioxide (CO₂) fluxes and evapotranspiration (ET). Despite that, commercial systems to measure CO₂
20 and ET fluxes are expensive and thus, often exclude research in ecosystems within the Global South. This is especially true
21 for research and data of agroecosystems in these areas, which are to date still widely underrepresented. Here, we present a
22 newly developed, low-cost, non-dispersive infrared (NDIR)-based, CO₂ and ET flux measurement device (~200 Euro) that
23 provides reliable, accurate and precise CO₂ and ET flux measurements in conjunction with manual closed chambers. To
24 validate the system, laboratory and field validation experiments were performed, testing multiple different low-cost sensors.
25 We demonstrate that the system delivers accurate and precise CO₂ and ET flux measurements using the K30 FR NDIR (CO₂)
26 and SHT31 (RH) sensor. An additional field trial application demonstrated its longer-term stability (> 3 months) and ability to
27 obtain valid net ecosystem C balances (NECB) and WUE. This was the case, even though environmental conditions at the
28 field trial application site in Sub-Saharan Africa were rather challenging (e.g., extremely high temperatures, humidity and
29 intense rainfall). Consequently, the developed low-cost CO₂ and ET flux measurement device not only provides reasonable
30 results but might also help to democratize science and close current data gaps.

31 **1 Introduction**

32 The global climate crisis is one of the most critical problems of our time and identifying and implementing measures to mitigate
33 or adapt to its consequences, such as more frequent and severe drought, is a key challenge. Solving this challenge, requires
34 first and foremost a substantial reduction of anthropogenic greenhouse gas (GHG) emissions in all sectors (IPCC, 2019). While
35 agriculture is a significant contributor to these anthropogenic GHG emissions (FAO, 2020), it might also offer the potential to
36 mitigate the climate crisis by increasing soil carbon (C) sequestration (Lal et al., 2004). Specifically, land use systems and
37 management practices which not only promote a net C uptake but also an efficient water use are needed. They might help to
38 increase soil C stocks and crop productivity, reducing GHG emissions while simultaneously sustaining yield, despite
39 intensifying climate stressors, such as more frequent and severe droughts. Hence, it is crucial to evaluate land use systems
40 regarding their potential to sequester additional C and effectively utilize water. Common parameters used to assess both, are
41 the net ecosystem C balance (NECB; Smith et al., 2010), and the agronomic and ecosystem water use efficiency (WUE; Beer
42 et al., 2009). Their determination, however, requires accurate and precise measurement of carbon dioxide (CO₂) and
43 evapotranspiration (ET) fluxes (Chapin et al., 2006; Livingston and Hutchinson, 1995; Rosenstock et al., 2016; Xu et al.,
44 2019).

45

46 Measurement of CO₂ and ET fluxes are commonly performed using eddy covariance or chamber based systems (Baldocchi et
47 al., 1996; Smith et al., 2010; Wang et al., 2017; Yang et al., 2014), while especially the latter are well suited for direct treatment
48 comparisons (Dubbett et al., 2014; Hoffmann et al., 2018; Kübert et al., 2020). In case of a remote study site location or
49 limitations in power supply, particularly manual closed chamber measurements are used to measure the CO₂ exchange and ET
50 fluxes (Rochette and Hutchinson, 2015). However, the relatively high costs of needed measurement equipment (particularly
51 gas analyzers) strongly limits their accessibility and often exclude research in ecosystems within the Global South. This
52 resulted in a pronounced underrepresentation of regions, land use systems and management practices from subtropical and
53 tropical South America, South Asia, and Africa, even though the quantification of e.g., CO₂ fluxes in these regions might
54 reduce disparities in the global CO₂ budget (Canadell et al., 2011; Gurney et al., 2002; Kondo et al., 2015).

55

56 Recent efforts to solve this financial constraint focus on developing low-cost, yet reliable, measurement devices. This was
57 catalyzed by the growing availability of relatively inexpensive microcontrollers, which are increasingly utilized for scientific,
58 environmental research (Blackstock et al., 2019; Capri et al., 2021). An additional contribution came from the improvement
59 in accuracy and precision of low-cost relative humidity (RH) and especially non-dispersive infrared (NDIR) CO₂ sensors.
60 Evaluation of commercially-available NDIR CO₂ sensors (Keimel et al., 2019; Martin et al., 2017; Pandey et al., 2007; Yasuda
61 et al., 2012) showed that they have acceptable precision and accuracy in measuring CO₂ concentrations especially when proper
62 calibration methods are applied. Although low-cost NDIR CO₂ sensors are commonly used in air quality monitoring studies
63 (Araújo et al., 2020; Wastine et al., 2022), these sensors have also been applied in environmental research (Bastviken et al.,
64 2015; Brown et al., 2020). For example, multiple studies have demonstrated the applicability of using low-cost NDIR CO₂
65 sensors for reliable measurements of soil CO₂ efflux (Brändle and Kunert, 2019; Curcoll et al., 2022; Harmon et al., 2015) and
66 water crop use determination (Capri et al., 2021). However, in case of RH sensors, the inversely increased measurement
67 uncertainty of total water vapor concentration with decreasing RH (e.g. a typical low-cost RH sensor has a measurement
68 accuracy of 1-3 % in relative but not absolute humidity) might constitute a problem. Despite first studies showing the potential
69 of using low-cost sensors as an alternative to more expensive commercial counterparts, there is still little evidence that in situ
70 closed chamber CO₂ and ET flux measurements using both, are comparable in precision and accuracy.

71

72 Here, we present the hard- and software implementation, as well as laboratory and in situ validation of a newly, low-cost and
73 open-source CO₂ and ET flux measurement device. We hypothesise that by using the device in conjunction with a manual
74 closed chamber 1.) CO₂ and ET fluxes can be reliably and accurately measured; and that 2.) measured CO₂ and ET fluxes can
75 be used to obtain valid estimates of net ecosystem C balance (NECB) and WUE, even under challenging environmental
76 conditions such as extremely high air temperatures, humidity, and precipitation. To test these hypotheses, we first validated
77 the accuracy and precision of four different low-cost NDIR CO₂ sensors (K30 FR, SCD30, MH-Z14, and MH-Z19) under
78 controlled laboratory conditions. Afterwards, the NDIR sensors passing laboratory validation as well as two different RH

79 sensors were validated in field. During field validation, ET and CO₂ fluxes (ecosystem respiration (R_{eco}) and net ecosystem
80 exchange (NEE)), as well as temperature-dependent R_{eco} and photosynthetic active radiation (PAR)-dependent gross primary
81 production (GPP) parameters, were compared to the results obtained simultaneously with a reference infrared gas analyser
82 (IRGA; LI-850, LI-COR, USA). Finally, the ability of the developed low-cost CO₂ and ET flux measurement device to obtain
83 reliable NECB and WUE as well as its practicability and stability were tested. Therefore, multiple devices were used during a
84 field trial application in Northern Ghana to obtain seasonal CO₂ exchange and ET, as well as NECB and WUE for four different
85 fertilizer treatments in a maize cultivation.

86 **2 Material and Methods**

87 **2.1 Hard- and software implementation**

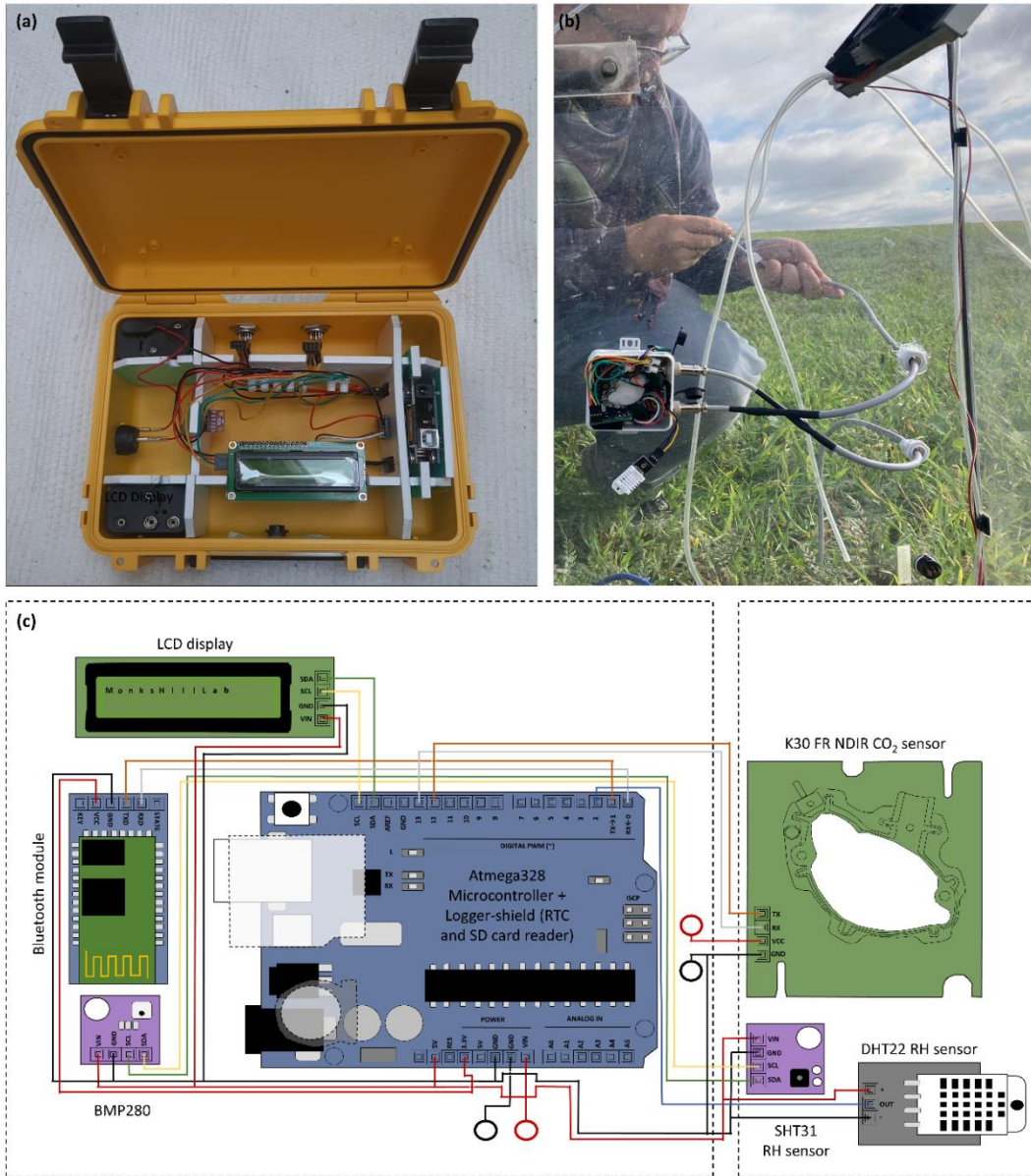
88 The developed, highly portable CO₂ and ET flux measurement device consists of a logger and sensor unit, both assembled out
89 of a combination of various low-cost, off-the-shelf components. A complete list of used components, distributors and prices is
90 given in Table 1. Figure 1 shows the assembled logger and attachable sensor unit, together with a schematic representation of
91 the wiring. The logger unit consists of an Arduino Uno like microcontroller (Atmega328, AZ-Delivery Vertriebs GmbH,
92 Germany) with attached Logger Shield module (AZ-Delivery Vertriebs GmbH, Germany) including an SD card reader and
93 SD card (2 GB) to store sensor readings and a real time clock (RTC) which helps to keep the time and date even when the
94 system is switched off. A BME280 air temperature (± 1 °C), air humidity (± 3 %) and air pressure sensor (± 1 hPa; Reichelt
95 electronics GmbH, Germany) as well as an LCD display (AZ-Delivery Vertriebs GmbH, Germany) and HC-05 Bluetooth
96 module are part of the logger unit and connected to the microcontroller. The logger unit is fitted into a weather and shock
97 resistant outdoor housing (B&W Outdoor Case Type 500, OVERHAUL MEDIA GmbH, Germany). It easily connects to end
98 user devices using the Bluetooth module, so data can be visualized inter-alia with a smartphone in real-time without the need
99 to open the weather and shock resistant outdoor housing. The external sensor unit consists of a NDIR-based CO₂ (0-10000
100 ppm, ± 30 ppm ± 3 % accuracy; K30 FR, Senseair AB, Sweden), an air humidity (RH) and air temperature sensor (SHT31, ± 2
101 % accuracy, Sensirion AG, Switzerland or DHT22, ± 2 to 5 % accuracy, Aosong Electronics Co., Ltd, China). Both sensors
102 were connected through a seven core cable to the logger unit using UART (K30 FR) and I2C (SHT31) data communication,
103 respectively. The power supply of the microcontroller is ensured by six rechargeable AA NiMH batteries (1.2 V; 2600 mAh)
104 in a 6×AA battery holder, which supply 7.2 V. Due to the power requirements of the external sensor unit (K30 FR and SHT31),
105 an additional 6×AA battery holder is attached to the housing directly. Software implementation was done using Arduino IDE
106 2.0.3.

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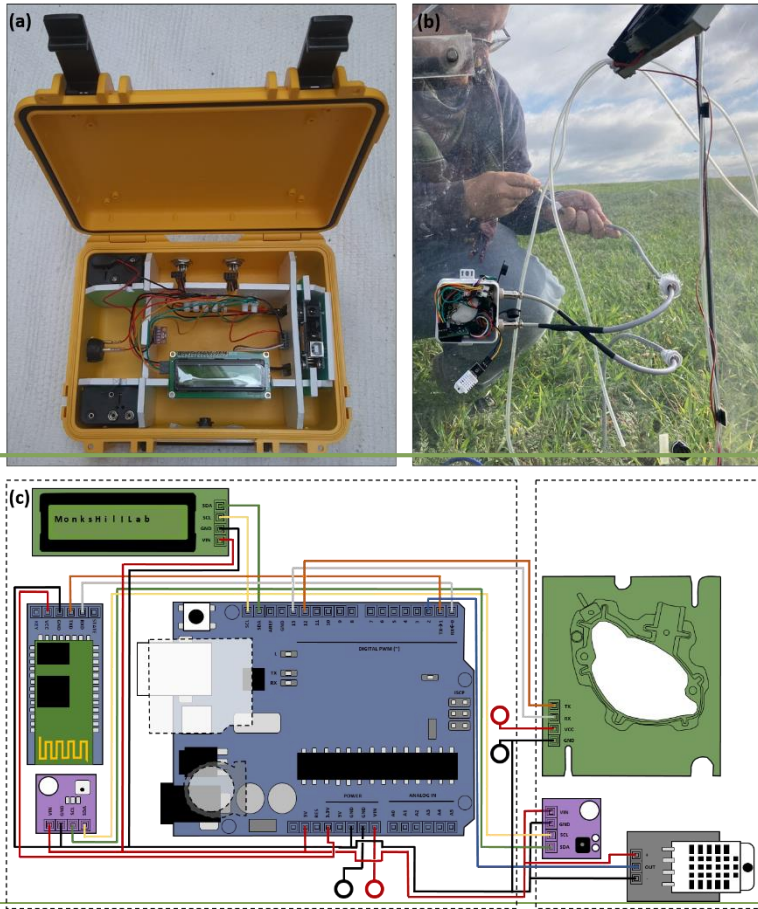
110 **Table 1:** Sensor components and cost (in Euro) at the time of writing, including weather and shock-proof housing and energy
 111 supply (rechargeable batteries). Components needed for optional semi-automatic mode are listed in addition.

COMPONENT	AMOUNT	DESCRIPTION	PRICE	DISTRIBUTOR
B&W OUTDOOR CASE TYP 500	1	Outdoor case for housing electrical components	28.75 Euro	www.profikoffer.de
PVC HARD FOAM PLATE	1	PVC 5 mm hard foam plate to create interior of housing for electronic components	1.5 Euro	www.amazon.de
LUSTER TERMINALS	12	Luster terminals for wiring electrical components within housing	0.6 Euro	www.amazon.de
0.2 MM² 24 AWG ELECTRICAL WIRE		Electrical wires for wiring electrical components within housing		www.amazon.de
7 PIN AVIATION CONNECTOR	2	Aviation connector to connect logger unit within weatherproof housing with passive NDIR sensor installed in the closed chamber to be attached	2.9 Euro	www.amazon.de
7 CORE RUBBER CABLE (1.5 M)	1	Cable to connect logger unit within weatherproof housing with passive NDIR sensor installed in the closed chamber to be attached	3.75 Euro	www.conrad.de
WS R13-112 AAAA ROCKER SWITCH	1	Rocker switch for switching on and off	1 Euro	www.reichelt.de
ATMEGA 328	1	Arduino Uno like microcontroller	5 Euro	www.az-delivery.de
DATALOGGER MODULE	1	Logger shield for Arduino UNO like microcontroller with SD card reader and RTC unit	4.6 Euro	www.az-delivery.de
HAMA CLASS 4, SD MEMORY CARD, 2 GB, 10 MB/S	1	SD memory card to save sensor readings	6 Euro	www.saturn.de
HC-05 BLUETOOTH WIRELESS RF-TRANSCEIVER-MODULE RS232	1	Bluetooth module for wireless communication	5.2 Euro	www.az-delivery.de
16x2 LCD OR OLED DISPLAY WITH I2C ADAPTER	1	LCD or OLED display for data visualization	3.7 Euro	www.az-delivery.de
BMP280	1	Air pressure, air humidity and air temperature sensor	1.7 Euro	www.reichelt.de
DHT22 OR SHT31 MODUL	1	Air temperature and air humidity sensor	6.4 Euro	www.az-delivery.de
SENSEAIR K30 FR (FAST RESPONSE)	1	CO ₂ measuring module with fast response time; Measuring range: 0 to 5000 ppm CO ₂ , operating range: 0 to 50 °C	85 Euro	www.driessen-kern.de
GOOBAY 11467 6x (4x) MIGNON (AA) BATTERY HOLDER	2 (1)	Battery holder for 6x NiMH rechargeable mignon (AA) batteries	4.6 Euro	www.conrad.de
CONRAD ENERGY HR06 MIGNON (AA)-AKKU NIMH 2600 MAH 1.2 V	12 (16)	NiMH rechargeable mignon (AA) batteries	38 Euro	www.conrad.de
4.5 V METAL BRUSH AIR PUMP	2	Air pump for flushing headspace of small chambers	9.45 Euro	www.berrybase.de
IRLZ44N MOSFET	1	Mosfet to control power supply to pumps	0.75 Euro	www.reichelt.de
COST OF OTHER NDIR SENSORS TESTED				
SENSIRION SCD30 MODULE	1	NDIR gas sensor for CO₂ (0-10000 ppm) integrated with humidity and temperature sensor in the same module	63.50 Euro	www.berrybase.de
MH-Z14 CO₂ SENSOR MODULE	1	NDIR gas sensor for accurately measuring the CO₂ concentration (0-10000 ppm)	55.60 Euro	www.kaufland.de
MH-Z19 CO₂ SENSOR MODULE	1	NDIR gas sensor for accurately measuring the CO₂ concentration (0-10000 ppm)	28.50 Euro	www.reichelt.de
TOTAL COST			199.7 Euro	

112



113 **Figure 1:** (a) Logger unit in weather and shock resistant housing, (b) external sensor unit attached to a transparent non-flow-
 114 through non-steady-state (NFT-NSS) closed chamber and (c) schematic representation of wiring.



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Figure 1: (a) Logger unit in weather and shock resistant housing, (b) external sensor unit attached to a transparent non-flow-through non-steady state (NFT-NSS) closed chamber and (c) schematic representation of wiring.

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2.2 Laboratory validation

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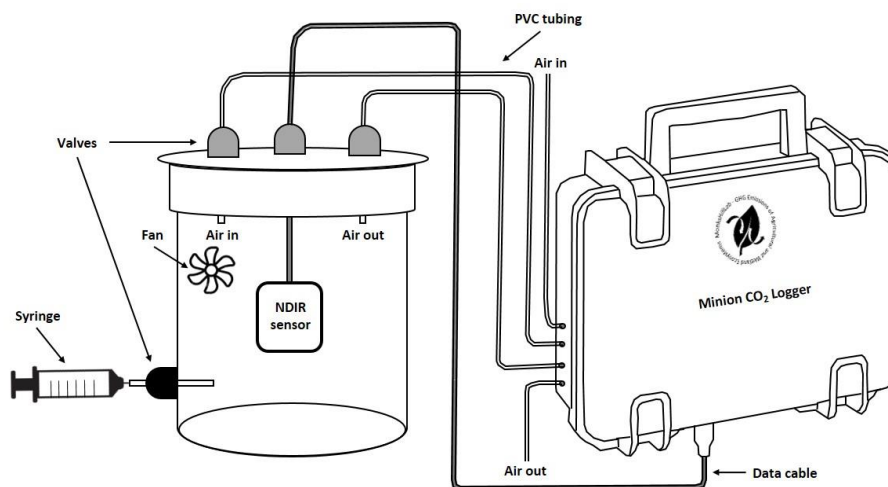
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To identify the NDIR sensor most suitable for in situ, dynamic closed chamber measurements, four different NDIR-based sensors were tested and validated regarding their precision and accuracy during a laboratory validation experiment. The sensors tested were 1.) MH-Z19 (Winsen Electronics Technology CO., LTD, China), 2.) MH-Z14 (Winsen Electronics Technology CO., LTD, China), 3.) SCD30 (Sensirion AG, Switzerland) and 4.) K30 FR (Senseair AB, Sweden). To identify the NDIR sensor most suitable for in situ, dynamic closed chamber measurements, four different NDIR-based sensors, namely 1.) MH-Z19 (Winsen Electronics Technology CO., LTD, China), 2.) MH-Z14 (Winsen Electronics Technology CO., LTD, China), 3.) SCD30 (Sensirion AG, Switzerland) and 4.) K30 FR (Senseair AB, Sweden) were tested and validated regarding their precision

126 and accuracy during a laboratory validation experiment. For this, sensors were placed separately into a sealed, ventilated,
127 cylindrical vessel (Fig. 2; V: 1425.5 cm³) and connected to the developed low-cost logger system.
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129
130 **Figure 2:** Experimental setup of the performed laboratory validation experiment for four different NDIR CO₂ sensors
131 connected to the developed low-cost CO₂ and ET flux measurement device (MH-Z14, MH-Z19, SCD30 and K30 FR).
132 Validation was performed through injecting distinct amounts of technical gas (Linde, Germany; 10000 ppm CO₂) into the air-
133 tight, sealed, cylindrical vessel.

134 All sensors were calibrated in ambient air prior to use according to manufacturer instructions. Afterwards different distinct
135 amounts (5 to 30 ml; in 5 ml steps; each step repeated five times) of a technical gas containing 10000 ppm CO₂ (Linde,
136 Germany) were injected into the sealed vessel using a syringe. In between injections, the vessel was flushed with ambient air
137 by two pumps (1.5 L min⁻¹) connected to the vessel (semi-automatic measurement mode of the developed device). Finally,
138 CO₂ concentration increases inside the vessel, measured in a 5 s interval by the NDIR-based sensors, from before to after
139 injection (Δ CO₂ in ppm) were compared against mixing-induced CO₂ concentration increases. Sensors that performed best in
140 terms of accuracy and precision were subsequently validated during the field validation experiment.

141 2.3 Field validation

142 Field validation of the low-cost CO₂ and ET flux measurement device was performed through parallel manual closed chamber
143 measurements using an infrared gas analyzer (IRGA; LI-850, LI-COR, USA) and NDIR sensors (CO₂) passing previous
144 laboratory validation, as well as two different RH sensors (ET). Measurements were conducted at the “PatchCrop”

145 experimental field, managed by the Leibniz Centre for Agricultural Landscape Research (Fig. 3; ZALF). “PatchCrop” features
146 multiple smaller patches (72 x 72 m), with diverse and site-specific crop rotations, aiming to create synergies and interactions
147 between fields.



148

149 **Figure 3:** Parallel opaque (R_{eco}) manual closed chamber measurements with a Li-COR 850 IRGA (LI-850, LI-COR, USA)
150 and the developed, low-cost CO_2 and ET flux measurement device at ZALF experimental field near the village of Tempelberg,
151 North-East Germany (52°26,827' N, 14°8492' E). The developed system was equipped with a K30 FR and SCD30 NDIR, as
152 well as SHT31 and DHT22 sensor.

153 The experimental field “PatchCrop” is located near the village of Tempelberg, Northeast Germany (52°26,827' N, 14°8492'
154 E). The temperate climate is characterized by a mean annual air temperature of 9.7°C and mean annual precipitation of 544
155 mm (ZALF weather station, 2010-2019). The medium loamy, sand textured soil can be classified as Luvisol (WRB). CO_2
156 exchange (NEE and R_{eco}) and ET measurements were conducted for a mixture of *Phacelia* and *Guizotia abyssinica* at three
157 repetitive plots, established at one of the patches through installing PVC collars (A: 0.5625 m²; 5 cm deep) in the beginning
158 of October 2022. Measurements started shortly after sunrise and lasted to late afternoon during two consecutive days, using a
159 dynamic, (non-)flow-through non-steady-state ((N)FT-NSS) manual closed chamber system. Used transparent (86 % light
160 transmission; NEE flux measurements) and opaque (R_{eco} flux measurements), cubic shaped PVC chambers had a total volume
161 of 0.296 m³ and were equipped with a fan for efficient headspace mixing. CO_2 and H_2O concentrations, as well as RH, during
162 chamber deployment were recorded in parallel using a LI-850 IRGA and the developed, low-cost measurement device,
163 equipped with a K30 FR, SCD30, SHT31 and DHT22 sensor, respectively. NEE, R_{eco} , and ET fluxes were measured by
164 alternately deploying the opaque and transparent chambers on the three pre-installed PVC frames. During individual 4 min

165 measurements, CO₂ and H₂O concentration changes in the chamber headspace, as well as RH, air temperature inside and
166 outside the chamber, soil temperature and humidity (TMS-4, TOMST, Czech Republic) as well as PAR (outside the chamber;
167 Skye, UK) were recorded at a 3 s (LI-850) and 5 s interval (NDIR and RH sensors). During individual 4 min measurements,
168 CO₂ and H₂O concentration, as well as RH, changes in the chamber headspace, air temperature inside and outside the chamber,
169 soil temperature and humidity (TMS 4, TOMST, Czech Republic) as well as PAR (outside the chamber; Skye, UK) were
170 recorded at a 3 s (LI 850) and 5 s interval (NDIR and RH sensors). To validate the low-cost CO₂ and ET flux measurement
171 device, measured R_{eco}, NEE, and ET fluxes, as well as the derived temperature and PAR dependency functions for R_{eco} and
172 GPP, respectively, were directly compared with results obtained in parallel with the LI-850. To validate the low-cost CO₂ and
173 ET flux measurement device, measured R_{eco}, NEE and ET fluxes, as well as derived temperature (R_{eco}) and PAR dependency
174 functions (GPP), were directly compared against results obtained in parallel with the LI 850.

175 **2.4 Field trial application**

176 The developed, low-cost measurement device has been tested for applicability and reliability under challenging environmental
177 conditions in an experimental field managed by the Council for Scientific and Industrial Research-Savanna Agricultural
178 Research Institute (Fig. 4; CSIR-SARI). The experimental field (21 × 54 m), located near the city of Nyankpala, Northern
179 Ghana (9°24'15.9" N, 01°00'12.1" W), featured a split-plot design (3 × 6 m; n=3) with the main plot assigned to tillage
180 practice (conventional vs. reduced tillage) and the subplot assigned to a factorial combination of organic and mineral fertilizers.
181 The tropical region around Nyankpala is characterized by a mean annual air temperature of 26 °C and a unimodal rainfall
182 pattern with a distinct rainy season from June to October followed by a dry season from November to May (Alua et al., 2018)
183 resulting in a mean annual precipitation of 1100 mm (CSIR-SARI weather station, 1995-2013). The soil is sandy loam textured
184 and classified as Acrisol (WRB). CO₂ exchange (NEE and R_{eco}) and ET measurements were conducted for maize (*Zea mays*)
185 from July to October 2022 at four out of the nine treatments with reduced tillage (bullock plough), namely: 1.) Fertisoil (5 t
186 ha⁻¹; commercial organic fertilizer in Northern Ghana; FT), 2.) farmyard manure (5 t ha⁻¹; FM), 3.) Fertisoil + NPK (5 t ha⁻¹ +
187 90-60-60 kg ha⁻¹; FT+MIN) and 4.) farmyard manure + NPK (5 t ha⁻¹ + 90-60-60 kg ha⁻¹; FM+MIN). Measurement campaigns
188 took place every two weeks from sunrise to late evening using a dynamic, NFT-NSS manual closed chamber system. Used
189 transparent (86 % light transmission; NEE flux measurements) and opaque (R_{eco} flux measurements), cubic shaped PVC
190 chambers had a total volume of 1.56 m³ and were equipped with a fan for efficient headspace mixing. CO₂ concentration and
191 RH changes during chamber deployment were recorded using the developed, low-cost measurement device, equipped with a
192 K30 FR and DHT22 sensor. During each measurement campaign, NEE, R_{eco}, and ET fluxes were measured by alternately
193 deploying the opaque and transparent chambers on pre-installed frames (A: 0.96 m²) at each of the measured plots.

194



Figure 4: Transparent (NEE) manual closed chamber measurement at CSIR-SARI experimental field, used for field trial application of the developed, low-cost CO₂ and ET flux measurement device, near the city of Nyankpala, Northern Ghana (9°24'15.9" N, 01°00'12.1" W).

2.5 Data processing

2.5.1 CO₂ and ET flux calculation, separation and gap-filling

For laboratory validation, the changes in CO₂ concentrations in the vessel, expressed as ΔCO_2 in ppm, were calculated as the mixing ratio of measured ambient air and injected technical gas CO₂ concentration (10000 ppm). These were compared with the ΔCO_2 obtained for the four different NDIR sensors as the difference in mean CO₂ concentrations measured for one minute right before and two minutes after injection. For the field validation, measured CO₂ and ET fluxes were calculated using a modular R script, described in detail by Hoffmann et al. (2015; CO₂) and Dahlmann et al. (2023; ET), respectively. Prior to CO₂ and ET flux calculation, underlying data was trimmed by removing the first and last 10 % of each chamber measurement dataset. This was conducted to eliminate data noise caused by turbulences and pressure fluctuations due to chamber deployment (Hoffmann et al., 2015), and to mitigate biases arising from the time needed to homogenize chamber headspace air (Vaidya et al., 2021). Prior to CO₂ and ET flux calculation a death band of 10 % was applied to the data of each chamber measurement. CO₂ concentrations measured using the LI-850 were additionally corrected for changes in water vapour during chamber measurements (Webb et al., 1980; McDermitt et al., 1993). Unlike the LI-850 which provided H₂O as mole fraction, used low-

212 cost RH sensors (DHT22 and SHT31) required additional post processing. RH measurements were converted into a mass
213 concentration following Hamel et al. (2015; Eq. 1):

$$215 \quad \text{H}_2\text{O} = \frac{\text{RH} \cdot e^s}{100 \cdot P} \quad (1)$$

216 where RH is the relative humidity, P is the gas pressure (Pa) and e^s is the saturated vapour pressure (Pa), calculated according
217 to Allen et al. (1998). Thereafter, CO₂ and ET fluxes were calculated based on the ideal gas law using a linear regression
218 approach (Eq. 2):

$$221 \quad f = \frac{MpV}{RTA} \cdot \frac{\Delta c}{\Delta t} \quad (2)$$

222 where M denotes the molar mass of the gas (g mol⁻¹), p denotes the ambient air pressure (Pa) and V denotes the chamber
223 volume (m³). Since plants accounted for < 0.1 % of the total chamber volume, a static chamber volume was assumed. R denotes
224 the gas constant (8.314 m³ Pa K⁻¹ mol⁻¹), T denotes temperature inside the chamber (K), A denotes the basal area (m²) and
225 $\Delta c/\Delta t$ denotes the linear CO₂ (e.g., Leiber-Sauheitl et al., 2014) and H₂O concentration change over time (e.g., Dahlmann et
226 al., 2023). The variables T and, more importantly, $\Delta c/\Delta t$, were obtained by applying a variable (window size 0.5 to 4 min)
227 moving window to each chamber measurement. ~~The variables T and, more importantly, $\Delta c/\Delta t$, were obtained by applying a~~
228 ~~variable moving window (0.5 to 3 min) to each chamber measurement.~~ Thus, resulting multiple ET and CO₂ fluxes per
229 measurement (based on generated variable moving window data subsets) were further evaluated according to the following
230 criteria: 1.) fulfilled prerequisites for applying a linear regression (normality (Lilliefors' adaption of the Kolmogorov-Smirnov
231 test), homoscedasticity (Breusch-Pagan test) and linearity); 2.) regression slope ($p \leq 0.1$, t-test); 3.) range of within-chamber air
232 temperature not larger than ± 1.5 K and a PAR deviation (only transparent chamber measurements) not larger than ± 20 % of
233 the average to ensure stable environmental conditions within the chamber throughout the respective measurement window; 4.)
234 no outliers present ($\pm 6 \times \text{IQR}$). Calculated CO₂ and ET fluxes that did not meet all criteria were discarded. In cases where more
235 than one flux per measurement met all criteria, the CO₂ and ET flux with the steepest slope and closest in time to chamber
236 closure were chosen. ~~In cases where more than one flux per measurement met all criteria, the CO₂ and ET flux with steepest~~
237 ~~slope and closest to chamber deployment were chosen.~~ For field validation and field trial application CO₂ fluxes were
238 additionally separated into its flux components R_{eco}, GPP and NEE and gap-filled through deriving empirical models. In the
239 case of R_{eco}, a temperature-dependent Arrhenius-type function was used and fitted for air as well as soil temperatures measured
240 in different depths (Lloyd and Taylor, 1994; Eq. 3).

$$243 \quad R_{\text{eco}} = R_{\text{ref}} \cdot e^{E_0 \left[\frac{1}{T_{\text{ref}} - T_0} - \frac{1}{T - T_0} \right]} \quad (3)$$

244

245 where R_{ref} is the respiration rate at the reference temperature (T_{ref} ; 283.15 K), E_0 is an activation energy-like parameter, T_0 is
246 the starting temperature constant (227.13 K) and T is the mean air or soil temperature during the flux measurement. Out of the
247 four obtained R_{eco} models (one model for air temperature inside the chamber, one for air temperature outside the chamber; soil
248 temperature at 2 and 5 cm depth), the model with the lowest Akaike information criterion (AIC) was finally used. In case of
249 GPP a PAR dependent, rectangular hyperbolic light-response function, based on the Michaelis–Menten kinetic, was used
250 (Elsgaard et al., 2012; Hoffmann et al., 2015; Wang et al., 2013; Eq. 4). Since GPP cannot be measured directly, GPP fluxes
251 were calculated as the difference between measured NEE and modelled R_{eco} fluxes, using campaign specific, previously
252 derived parameters R_{ref} and T_0 .

253

$$254 \quad GPP = \frac{GP_{max} \cdot \alpha \cdot PAR}{\alpha \cdot PAR + GP_{max}} \quad (4)$$

255

256 where GP_{max} is the maximum rate of C fixation at infinite PAR ($\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$), α is the light use efficiency ($\mu\text{mol CO}_2$
257 μmol^{-1} photons) and PAR is the photon flux density (corrected for chamber light transmission) of the photosynthetically active
258 radiation (μmol^{-1} photons $\text{m}^{-2} \text{ s}^{-1}$). In cases where the rectangular hyperbolic light-response function did not result in
259 significant parameter estimates, a non-rectangular hyperbolic light-response function was used (Gilmanov et al., 2007; 2013).
260 R_{eco} and GPP parameter sets were evaluated and discarded in case of non-significant parameter estimates. If no fit or a non-
261 significant fit was achieved, averaged flux rates were used for R_{eco} and GPP instead. R_{eco} , GPP and NEE were modelled in half
262 hourly steps for the entire period based on continuously monitored temperature and PAR. For ET, campaign-wise average
263 daily ET fluxes (for nighttime ET fluxes measured before, for daytime ET fluxes measured after 8:00) were determined and
264 linearly interpolated between campaigns for the entire crop growth period.

265 **2.5.2 NECB and WUE**

266 NECB for the field trial application experiment was calculated as the sum of cumulated NEE, C output such as harvested
267 biomass C and C input due to organic fertilizer application (Eq. 5; Smith et al., 2010).

268

$$269 \quad NECB = NEE + C_{input} - C_{output} \quad (5)$$

270

271 Several minor NECB components have not been considered, such as, C input from seeding and methane emissions. However,
272 due to their relatively low magnitude (e.g., no methane emissions in mineral soil under aerobe conditions) their influence on
273 the NECB of our study is neglectable. Values for R_{eco} , GPP, NEE, harvested biomass C and NECB are given using the
274 atmospheric sign convention (Ceschia et al., 2010), where positive values indicate C losses from the plant-soil system and

275 negative values indicate C uptake. Thus, NECB refers to the total change in below-ground C. WUE was calculated as the
276 agricultural WUE (WUE_{agro} ; Eq. 6; Hatfield and Dold, 2019).

277

$$278 \quad WUE = \frac{DM}{ET} \quad (6)$$

279

280 where DM denotes harvested dry biomass in g m^{-2} and ET is cumulative evapotranspiration in mm.

281 2.5.3 Error calculation and statistical analysis

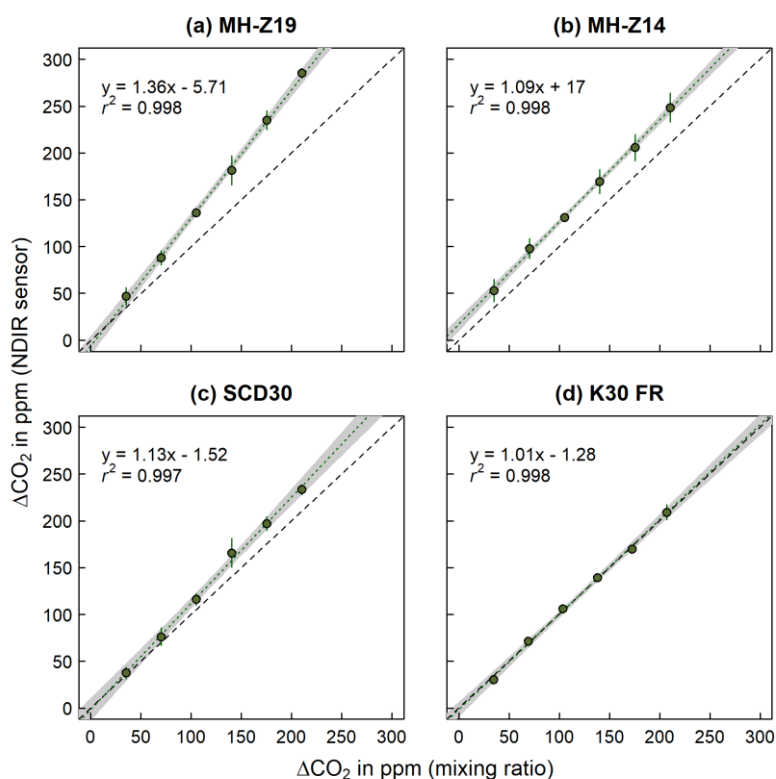
282 To test for normal distribution of the data obtained from laboratory and field validation measurements, Kolmogorov-Smirnov
283 test ($p < 0.05$) was performed. In case of normal distribution, significant differences between ΔCO_2 in ppm or R_{eco} , NEE, and
284 ET fluxes measured from low-cost sensors and mixing ratio ΔCO_2 or IRGA-based R_{eco} , NEE, and ET fluxes were determined
285 using one-sample t-test ($p < 0.05$). Error calculation for CO_2 fluxes, as well as crop season CO_2 exchange, were quantified using
286 a comprehensive error prediction algorithm described in detail by Hoffmann et al. (2015). The approach utilizes bootstrapping
287 alongside k-fold subsampling to estimate uncertainties for each flux measurement as well as subsequent R_{eco} and GPP
288 parametrization and final gap-filling. An adaptation of this approach was used to calculate errors in ET fluxes (Dahlmann et
289 al., 2023). Seasonal ET flux errors were then estimated based on $1.96 \times \text{SD}$ of daily average ET fluxes. ~~Error calculation for~~
290 ~~CO_2 and ET fluxes, as well as crop season CO_2 exchange and ET, were quantified using a comprehensive error prediction~~
291 ~~algorithm described in detail by Hoffmann et al. (2015).~~

292 3 Results and Discussion

293 3.1 Laboratory validation

294 Differences in accuracy and precision among the tested, four different low-cost NDIR sensors are shown in Fig. 5a-d as 1:1-
295 agreement plots between mixing ratio (calculated) and measured ΔCO_2 . While accuracy can be assessed as deviation from the
296 1:1-agreement line, precision is determined by the residual standard deviation (SD) and the coefficient of determination (r^2)
297 of the linear regression fitted on calculated versus measured ΔCO_2 . The K30 FR (Fig. 5d) showed the highest accuracy among
298 all tested NDIR sensors, reflecting well the increase in CO_2 concentration (ΔCO_2) derived through mixing ratio.
299 Correspondingly, no significant difference (one sample t-test, $p=0.80$) was found between calculated and measured ΔCO_2 . The
300 SCD30 (Fig. 5c), even though fairly accurate at lower, failed to reflect higher calculated ΔCO_2 values and generally tends to
301 overestimate triggered ΔCO_2 . Neither the MH-Z14 (Fig. 5b) nor the MH-Z19 (Fig. 5a) were sufficiently accurate and able to
302 reflect triggered ΔCO_2 . While the MH-Z14 showed a rather constant offset from the 1:1-agreement by 28 ppm, the MH-Z19
303 tends to increasingly overestimate higher ΔCO_2 values derived through mixing ratio. Hence, unlike the K30 FR, all other NDIR
304 sensors measured significantly higher ΔCO_2 when compared to mixing ratio ΔCO_2 (one sample t-test, $p < 0.01$). Unlike the

305 accuracy, overall precision and measurement repeatability among all four NDIR sensors was generally high and fairly
 306 comparable, showing a residual SD of 2.78 ppm, 4.23 ppm, 2.52 ppm and 3.58 ppm, respectively. Regarding the response time
 307 (defined as mean time from injection to measured initial CO₂ concentration increase), all four NDIR sensors differed
 308 substantially, with only 44 seconds for the K30 FR and more than 280 seconds for the MH-Z14. The same was true for the
 309 response strength (defined as the mean time from beginning to end of the injection triggered CO₂ concentration increase, which
 310 represents its steepness), with 61, 160 and 265 seconds for the K30 FR, SCD30 and MH-Z19 respectively. In case of the MH-
 311 Z14, response strength could not be evaluated, since no clear saturation after injection induced CO₂ concentration increase
 312 could be observed.



313
 314 **Figure 5:** 1:1-agreement between mixing ratio and measured ΔCO₂ in ppm from the four low-cost sensors tested (K30 FR,
 315 SCD30, MH-Z14 and MH-Z19). The dashed black line indicates the 1:1-agreement. The dotted green line shows the linear
 316 regression through the average ΔCO₂ for each injection step (n=5), calculated from the repetitive measurements per step. Error
 317 bars indicate ±1.96 SD. The grey shaded area represents the respective confidence band of the regression line.

318 While accuracy and precision are of course highly relevant, response time and response strength in particular play a key role
319 in determining the extent to which the tested NDIR sensors can be used for in situ NFT-NSS closed chamber measurements.
320 With a response time of almost 2 min and 5 min, respectively, as well as low response strength, MH-Z19 and MH-Z14 would
321 likely fail to correctly reflect ΔCO_2 during short-time (<4 min) closed chamber measurements, regardless of their low accuracy,
322 which makes them additionally unsuitable. Therefore, only the K30 FR (and to a much lower extent the SCD30) with its fast
323 response time and high response strength passed laboratory validation and met all necessary requirements for accurate and
324 precise in situ measurements of CO_2 exchange. Our findings, comparing accuracy and precision of four different NDIR sensors
325 during a laboratory setup, are in a good agreement with previous studies performing laboratory validation of single sensors.
326 Brändle and Kunert (2019), who compared the MH-Z14A NDIR sensor against a GFS-3000 (Heinz Walz GmbH, Germany)
327 during a laboratory validation observed a similar response time and a general measurement offset of approx. +40 ppm. Based
328 on this and an additionally conducted field validation, Brändle and Kunert (2019) also suggested that the MH-Z14A is not
329 suitable for short term measurements (<5 mins). Also findings of González Rivero et al. (2023), who tested the ability of the
330 SCD30 to reflect calibration gas concentrations and concluded an acceptable accuracy and response time, are in a good
331 agreement with results of the present study. The most widely tested NDIR sensors so far, however, are those of the K-Series
332 (as e.g., Ali et al., 2016; Blackstock et al., 2019; Brown et al., 2020; Mendes et al., 2015). Laboratory validation performed by
333 Blackstock et al. (2019) using K30 1 % sensor to measure a span of different CO_2 concentrations verified that it well reflects
334 CO_2 concentrations within the accuracy stated by the manufacturer. Similarly, laboratory tests performed by Mendes et al.
335 (2015) found that the K30 sensor has nearly perfect linear response against calibration gas CO_2 concentrations. Lastly, the
336 laboratory experiment by Ali et al. (2016) also highlighted the accuracy of the K30 1 % sensor when compared against
337 measurements of an SBA-5 CO_2 gas analyzer (PP Systems, USA). During their experiment both sensors showed a strong
338 correlation and no offset, when K30 1 % sensor self-calibration was used, highlighting the self-calibration capabilities of the
339 K-series sensors that contribute to their stable performance and high measurement repeatability with minimal maintenance
340 compared to other NDIR sensors.

341 **3.2 Field validation**

342 3.2.1. In situ ET flux validation

343 Two low-cost RH sensors (ET; SHT31 and DHT22) were tested in parallel with NDIR sensors passing the laboratory validation
344 (CO_2 ; K30 FR and SCD30) against LI-850 as reference. To avoid systematic impact of opaque chambers on plant transpiration
345 via stomatal closure upon darkening, in case of ET fluxes, only transparent chamber measurements were taken into account
346 (Larcher, 2003). Out of the 20 NEE measurements, 13 valid ET fluxes could be calculated in case of the LI-850. Compared to
347 that, 18 and 17 valid ET fluxes were obtained for the SHT31 and DHT22, respectively. Differences in accuracy and precision
348 for ET fluxes calculated based on RH measurements (Fig. 6c-d) compared to ET fluxes calculated based on LI-850 are shown

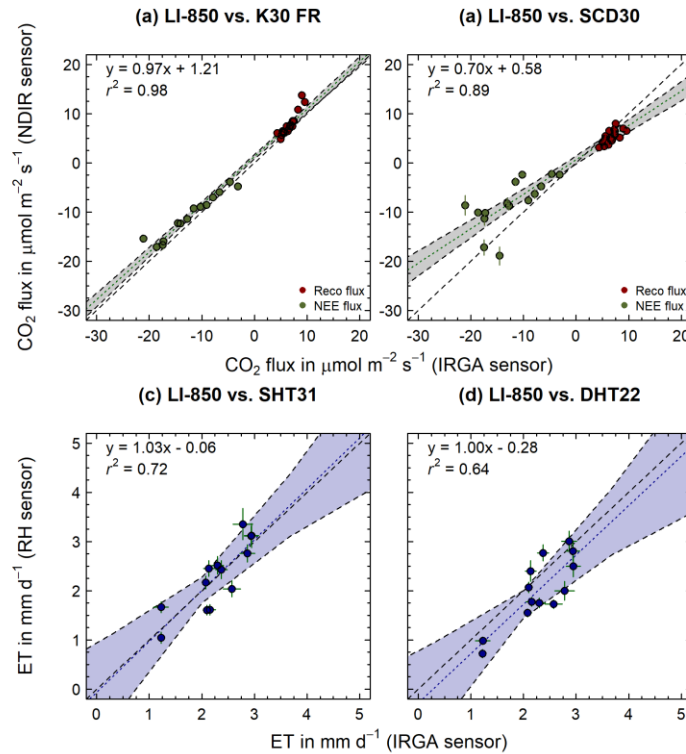
349 as 1:1-agreement plots in Fig. 6. No significant difference (mean diff. -0.01 mm d^{-1} ; one sample t-test, $p=0.89$) was found
350 between ET fluxes calculated from H_2O concentration and RH measurements, using the LI-850 and SHT31, respectively (Fig.
351 6c). Together with an r^2 of 0.72, this indicates a reasonable accuracy of SHT31 derived ET flux estimates. Compared to that,
352 ET fluxes, determined through RH measurements using the DHT22 (Fig. 6d), were significantly smaller (mean diff. 0.28 mm
353 d^{-1} ; one sample t-test, $p<0.05$) than LI-850 based ET fluxes and with an r^2 of 0.64, less accurate. This is consistent with sensor
354 accuracy for measuring relative humidity specified by their corresponding manufacturers, which are $\pm 2 \%$ accuracy for SHT31
355 and $\pm 2\text{-}5 \%$ accuracy for DHT22. Since these low-cost sensors were only capable of measuring at this level of accuracy, a
356 higher uncertainty at lower RH concentrations and consequently derived ET fluxes, might occur, even though not directly
357 detected within this study. The overall precision of SHT31 and DHT22 derived ET fluxes were fairly similar, but with a
358 residual SD of 0.36 and 0.39 mm d^{-1} , rather high.

359 3.2.2. In situ CO_2 flux validation

360 A total of 41 closed chamber measurements (R_{eco} : 21; NEE: 20) has been conducted during the two days field validation, using
361 the LI-850 as reference for both NDIR sensors passing the laboratory validation (CO_2 ; K30 FR and SCD30). While for the LI-
362 850, 41 valid CO_2 fluxes (R_{eco} : 21; NEE: 20) could be calculated, 35 (R_{eco} : 21; NEE: 14) and 36 (R_{eco} : 21; NEE: 15) valid
363 fluxes were obtained for K30 FR and SCD30, respectively. Differences in accuracy and precision for CO_2 fluxes calculated
364 based on NDIR (Fig. 6a-b) compared to CO_2 and ET fluxes calculated based on LI-850 are shown as 1:1-agreement plots in
365 Fig. 6. While the comparison between R_{eco} and NEE fluxes calculated from LI-850 and K30 FR measurements (Fig. 6a), was
366 in accordance with the laboratory validation and showed again the overall accuracy and precision of this NDIR sensor, a small
367 positive offset was found. Hence, CO_2 fluxes for the K30 FR were significantly higher (R_{eco} mean diff. $1.12 \mu\text{mol m}^{-2} \text{ s}^{-1}$; one
368 sample t-test, $p<0.05$) and less negative (NEE mean diff. $1.41 \mu\text{mol m}^{-2} \text{ s}^{-1}$; one sample t-test, $p<0.05$) when compared to LI-
369 850. No such systematic offset was found in case of the SCD30 (Fig. 6b), which showed significantly lower R_{eco} (mean diff. -
370 $1.33 \mu\text{mol m}^{-2} \text{ s}^{-1}$; one sample t-test, $p<0.05$) and much less negative NEE fluxes (mean diff. $-4.18 \mu\text{mol m}^{-2} \text{ s}^{-1}$; one sample t-
371 test, $p<0.05$) compared to LI-850. Since neither both NDIR sensors showed a similar offset, nor an overestimation was found
372 for the K30 FR during the laboratory validation already, it can be assumed that the detected offset in case of the K30 FR is
373 neither a direct result of microclimatic effects (e.g., increasing humidity), nor incorrect sensor readings. Instead, inter-alia
374 differences within the chamber headspace and the position of the NDIR sensor right below the chamber top, approx. 10 cm
375 above the LI-850 inlet and outlet, might help to explain it. Nonetheless, the NDIR sensor K30 FR still exhibited higher accuracy
376 than the SCD30 when validated against LI-850 flux measurements. The root mean squared error (RMSE), mean squared error
377 (MSE), and mean absolute error (MAE) obtained from the K30 FR (RMSE: $1.77 \mu\text{mol m}^{-2} \text{ s}^{-1}$; MSE: $3.16 \mu\text{mol m}^{-2} \text{ s}^{-1}$; MAE:
378 $1.34 \mu\text{mol m}^{-2} \text{ s}^{-1}$) were lower in comparison to SCD30 (RMSE: $3.97 \mu\text{mol m}^{-2} \text{ s}^{-1}$; MSE: $15.77 \mu\text{mol m}^{-2} \text{ s}^{-1}$; MAE: $2.80 \mu\text{mol}$
379 $\text{m}^{-2} \text{ s}^{-1}$). Compared to the K30 FR, especially NEE fluxes obtained by the SCD30, were also characterized by a very low
380 precision. The reason for this is certainly the lower CO_2 concentrations ($<400 \text{ ppm}$) in the NEE measurements, which are

381 clearly outside the measurement range specified by the manufacturer (400 to 10000 ppm). This also explains the decreasing
 382 precision with increased negative NEE fluxes obtained by SCD30, since these are likely related to CO₂ concentration
 383 measurements well below 400 ppm. The general underestimation of R_{eco} and NEE fluxes derived from SCD30, however, is
 384 probably a result of its rather long response time and lower response strength when compared to the K30 FR (see 3.1).

385



386

387 **Figure 6:** 1:1-agreement between (a-b) CO₂ (R_{eco}: dark red points; NEE: dark green points) and (c-d) ET fluxes measured with
 388 infrared gas analyzer (IRGA; LI-850, LI-COR, USA), and low-cost NDIR sensors (K30 FR and SCD30), as well as low-cost
 389 RH sensors (SHT 31 and DHT22), respectively. The dashed black line indicates the 1:1-agreement. The dotted green/blue line
 390 shows the linear regression through the measured CO₂/ET fluxes. The grey/blue shaded area represents the respective
 391 confidence band of the regression line. Error bars indicate calculated flux error (CI: 95%; $p < 0.05$).

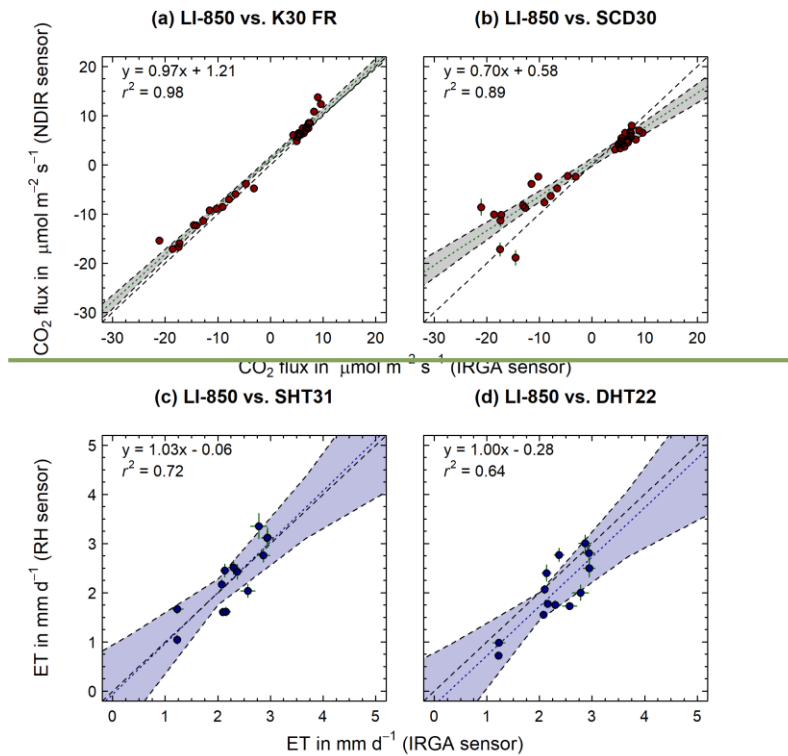


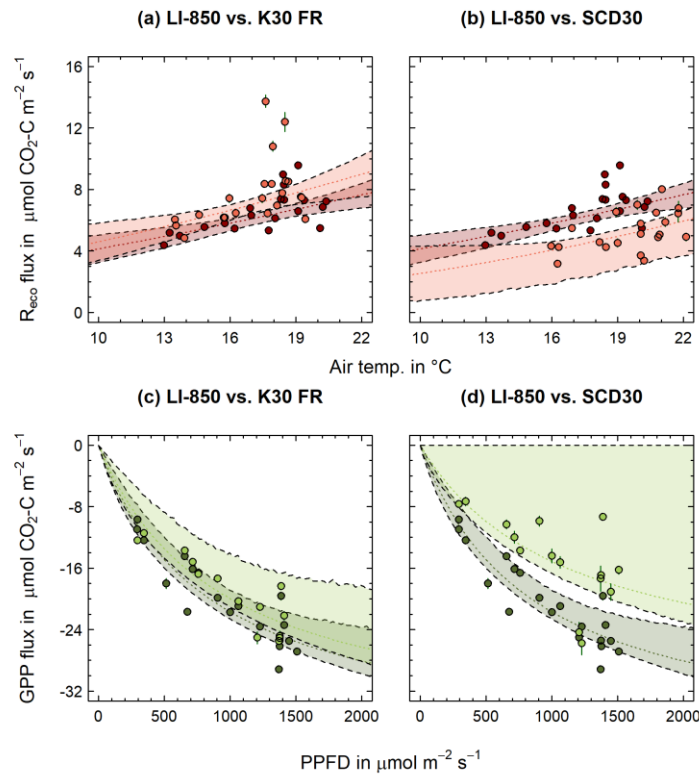
Figure 6: 1:1 agreement between (a-b) CO₂ and (c-d) ET fluxes measured with infrared gas analyzer (IRGA; LI-850, LI-COR, USA) and low cost NDIR sensors (K30 FR and SCD30). The dashed black line indicates the 1:1 agreement. The dotted green/blue line shows the linear regression through the measured CO₂/ET fluxes. The grey/blue shaded area represents the respective confidence band of the regression line. Error bars indicate calculated flux error (-).

3.2.3 Temperature- and PAR-dependency of measured CO₂ fluxes

Figure 7 shows temperature-dependent R_{eco} (Fig. 7a-b) and PAR-dependent GPP (Fig. 7c-d) parameter estimates for flux measurements performed with the LI-850 compared to K30 FR (Fig. 7a, 7c) and SCD30 (Fig. 7b, 7d), respectively. Since the R_{eco} and GPP parameters are based on the fluxes presented in Fig. 6, similar differences between LI-850, K30 FR and SCD30 could be obtained. With an R_{ref} and E_0 of 4.60 and 212.71, the K30 FR had similar, but slightly higher R_{eco} parameters (Fig. 7a) when compared to the LI-850 (R_{ref} : 4.14; E_0 : 195.01). This indicates not only in general higher R_{eco} fluxes but, more importantly, also a stronger increase of R_{eco} fluxes with rising temperature. In the case of the SCD30 (R_{ref} : 2.54; E_0 : 270.07), differences in R_{eco} parameters were, however, much more pronounced. The same tends to be true for obtained GPP parameters,

405 which were highly comparable for LI-850 (α : -0.048; GP_{\max} : -39.83) and K30 FR (α : -0.042; GP_{\max} : -38.42), but distinctly
406 different for SCD30 (α : -0.029; GP_{\max} : -31.83). As a result, the fitted K30 FR PAR dependency function was fully within the
407 confidence band of the LI-850 PAR dependency function. In summary, the K30 FR well represented R_{eco} and GPP fluxes
408 measured with the LI-850 and thereon based parameter estimates for R_{eco} and GPP. Unlike the K30 FR, the SCD30 was only
409 able to reflect LI-850 R_{eco} and GPP fluxes measured within the manufacture specified concentration range. Correspondingly,
410 accurate parameter estimates, especially with GPP, were not obtained. Our findings are further supported by studies that
411 compared the accuracy of K-series sensors against commercial sensor counterparts and its accuracy for field CO_2 flux
412 measurements (Curcoll et al., 2022). They integrated a K30 STA sensor into NFT-NSS chamber measurements and were able
413 to accurately measure CO_2 fluxes for a grassland ecosystem. Adding to that, the average CO_2 flux obtained during our study
414 using K30 FR ($0.4 \mu\text{mol m}^{-2} \text{s}^{-1}$) falls within the range of reported daily average NEE values (4 to $-6 \mu\text{mol m}^{-2} \text{s}^{-1}$) in the study
415 by Emmel et al. (2018) for a field site in Switzerland which was also covered with *Phacelia* cover crop. Based on the performed
416 field validation, the developed low-cost measurement device equipped with the K30 FR and SHT31 is likely to accurately
417 measure CO_2 and ET fluxes in situ, using NFT-NSS closed chambers.

418



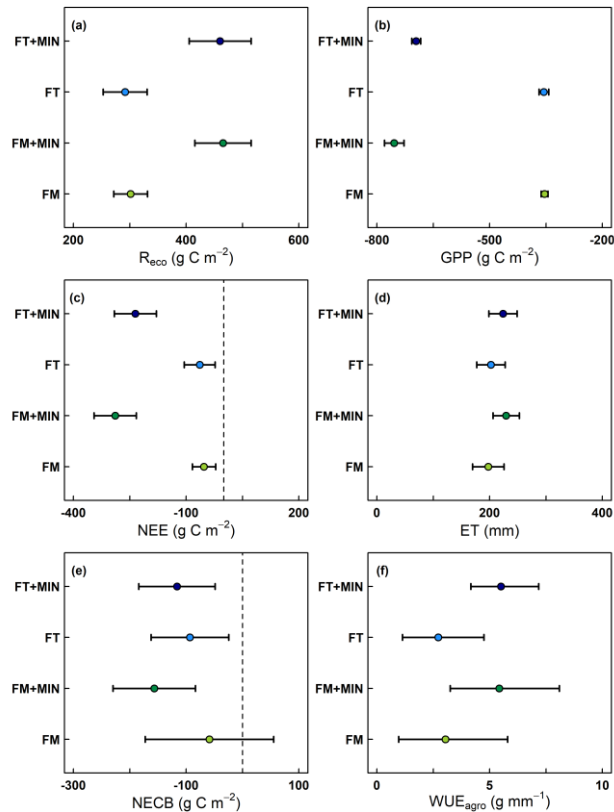
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420 **Figure 7:** Comparison of R_{eco} temperature dependency (dotted red lines) and GPP PAR dependency functions (dotted green
 421 lines) between LI-850 (dark red/green) and K30 FR and SCD30 (light red/green), respectively. Shaded red/green areas indicate
 422 confidence band around functions. Dots represent measured R_{eco} and derived GPP fluxes. Error bars indicate calculated flux
 423 error (CI: 95%; $p < 0.05$), ($\alpha = 0.9$).

424 3.3 Field trial application

425 During the measurement period, half-hourly air temperatures at the field site near Nyankpala, Northern Ghana, reached as high
 426 as 46°C , with daily average air temperatures ranging from 24°C to 32°C . Daily rainfall varied strongly between the rainy and
 427 dry season, with single heavy rain event of up to 115 mm d^{-1} . Consequently, average monthly air humidity was highest (65 to
 428 85 %) during the rainy season and as low as 23 % during the dry season. Irrespective of these harsh environmental conditions,
 429 the reliability of the developed low-cost measurement device could be proven during the field trial application. Periodically
 430 performed diurnal CO_2 measurement campaigns resulted in consistent R_{eco} and NEE fluxes, showing throughout the entire
 431 crop growth a clear light (PAR) dependency for derived GPP fluxes (data not shown). The maximum daily R_{eco} (3.9 g C m^{-2}

432 d^{-1}) and GPP ($-6.9 \text{ g C m}^{-2} d^{-1}$) fluxes derived for the non-mineral fertilized treatments, were well within the range ($4.0 \text{ g C m}^{-2} d^{-1}$ and $-7.0 \text{ g C m}^{-2} d^{-1}$) of EC derived maximum daily R_{eco} and GPP fluxes reported by Quansah et al. (2015), who measured a mixed fallow and cropping system in Northern Ghana, dominated by tall grasses. When adjusted for observation length, cumulative NEE, GPP and R_{eco} values obtained during the same study (27 g C m^{-2} , -195 g C m^{-2} and 222 g C m^{-2}) were found to be consistent with the average cumulative NEE, GPP and R_{eco} values obtained from the non-mineral fertilized treatments during our field trial application experiment ($-58 \pm 8 \text{ g C m}^{-2}$, $-355 \pm 1 \text{ g C m}^{-2}$ and $297 \pm 7 \text{ g C m}^{-2}$). (~~-58 g C m^{-2} , -355 g C m^{-2} and 297 g C m^{-2}~~). Also, EC measurements of an unfertilized cropland system (including maize) in Cameroon resulted with 218.5 g C m^{-2} in a comparable cumulative R_{eco} (Verchot et al., 2020). Regarding ET, the highest cumulative ET of our study (FM + MIN; $229 \pm 23 \text{ mm}$) (~~FM + MIN; 229 mm~~) was similar to the measured ET flux (238 mm) of a field site in Northern Benin, which was dominated by C4 plants (Mamadou et al., 2016). In general, obtained cumulative ET (Fig. 8d) for all four treatments were furthermore in a good agreement with ET obtained for Northern Ghana from average monthly actual evapotranspiration (FAO, 2019), corrected using phenology specific crop factors for grain maize (263 mm ; Brouwer and Heibloem, 1986). Cumulative R_{eco} and GPP fluxes recorded for the four different treatments well-reflected the difference in harvested biomass ($529 \pm 59 \text{ g C m}^{-2}$ for FT+MIN and $534 \pm 143 \text{ g C m}^{-2}$ for FM+MIN) (~~529 g C m^{-2} for FT+MIN and 267 g C m^{-2} for FM+MIN~~), with higher cumulative R_{eco} and GPP for higher crop biomass (Fig. 8a-b). Consequently, also NEE and thereon based NECB was higher for additionally, mineral fertilized treatments compared to non-mineral fertilized treatments, with differences between additionally, mineral and non-mineral fertilized treatments being more pronounced for FM when compared to FT (Fig. 8c and e). Similar tendencies were found for ET and thereon based WUE, with additionally, mineral fertilized treatments showing a higher ET and WUE compared to non-mineral fertilized treatments (Fig. 8d and f). This is in alignment with results reported by Mo et al. (2017) for maize in Kenya, where WUE increased with higher grain yield due to increasing mineral N fertilization. Besides the reliability of the developed low-cost measurement system, also its practicability was proved during the field trial application. Despite of the rather demanding environmental conditions, the system showed that it is uncomplicated and easy to operate even for untrained staff. After a short training session, even non-technical trained staff can conduct minor repairs of the system directly in the field. ~~It easily connects to end user devices using the Bluetooth module, so data can be visualized inter alia with a smartphone in real-time without the need to open the weather and shock resistant outdoor housing. After a short training session, even non-technical trained staff can conduct minor repairs of the system directly in the field. Its light weight and low power consumption with the 12 rechargeable NiMH batteries lasting for as long as eight hours, make the system especially suitable for in-situ closed chamber measurements in remote tropical areas. Compared to Li-ion batteries, the rechargeable NiMH batteries are furthermore relatively safe to use at high temperatures.~~ However, the missing user interface currently still prevents direct input of information, such as names of measurement location and soil temperatures, which made data post processing more tedious.



463

464 **Figure 8:** Cumulative (a-d) R_{eco} , GPP, NEE (g C m⁻²) and ET fluxes (mm) as well as thereon based estimates of (e-f) NECB
 465 (g C m⁻²) and WUE (g mm⁻¹) for the four different fertilizer treatments, namely: 1.) Fertisoil (5 t ha⁻¹; commercial organic
 466 fertilizer in Northern Ghana; FT), 2.) farmyard manure (5 t ha⁻¹; FM), 3.) Fertisoil + NPK (5 t ha⁻¹ + 90-60-60 kg ha⁻¹; FT+MIN)
 467 and 4.) farmyard manure + NPK (5 t ha⁻¹ + 90-60-60 kg ha⁻¹; FM+MIN). Error bars indicate calculated flux error (90% CI;
 468 $p < 0.1$).

469 4 Conclusions and implications for further use

470 Performed experiments showed that CO₂ and ET fluxes can be measured reliably and in a stable manner over time using
 471 inexpensive NDIR and RH sensors in conjunction with a manual closed chamber system. Out of the various low-cost CO₂ and
 472 RH sensors that were validated, the K30 FR and SHT31 proved to be the most accurate in measuring CO₂ and ET fluxes,
 473 respectively. Additionally, the developed low-cost measurement device was shown to be both practical and applicable to use
 474 even in environmentally challenging agroecosystems, as demonstrated by the field trial application in Northern Ghana, sub-

475 Saharan Africa. There within, seasonal CO₂ and ET fluxes turned out to be reliable and could be used to obtain valid NECB
476 and WUE estimates. Since the system developed is battery-powered (solar rechargeable), based on open-source technology
477 and all its components are low-cost, it can become easily accessible to a broad range of researchers. Its light-weight and low
478 power consumption with the 12 rechargeable NiMH batteries lasting for as long as eight hours, make the system especially
479 suitable for in situ closed chamber measurements in remote tropical areas. Compared to Li-ion batteries, the rechargeable
480 NiMH batteries are furthermore relatively safe to use at high temperatures. Since the system developed is battery powered
481 (solar rechargeable), based on open source technology and all its components are low cost, it can become easily accessible to
482 a broad range of researchers. This opens manifold potential applications, especially in the Global South, regarding the
483 evaluation and identification of various land use systems and management practices, in terms of their C sequestration potential,
484 water consumption and WUE. Therefore, the developed measurement device can be a valuable tool in evaluating and assessing
485 global C and water flux models, ultimately expanding the network for C budget and ET research that are both critical for
486 climate crisis adaptation and mitigation.

487 **5 Data and code availability**

488 The data and code referred to in this study are publicly accessible at <https://doi.org/10.4228/zalf-hdqh-br42>.

489 **6 Author contribution**

490 MH and RM conceptualized and developed the system and code. RM, DA and GS carried out the laboratory and field validation
491 experiments. MA conducted the field trial application. RM, MH, and MD wrote and prepared the manuscript with contributions
492 of all co-authors. All authors have reviewed and agreed to the final version of the manuscript.

493 **7 Competing interests**

494 The authors declare that they have no conflict of interest.

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504 application.

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