

1 Marked recent declines in boron in Baltic Sea cod otoliths – a bellwether  
2 of incipient acidification in a vast hypoxic system?  
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11 **Abstract:** Ocean acidification is spreading globally as a result of anthropogenic CO<sub>2</sub> emissions,  
12 but the Baltic Sea has until recently been thought to be relatively well-buffered by terrigenous  
13 inputs of alkalinity from its watershed. We discovered a 3- to 5-fold decline in boron (as B:Ca)  
14 in otoliths of eastern Baltic Sea cod (EBC) between the late 1990s and 2021. B: Examining a  
15 time series of EBC otoliths, we found varying levels of B:Ca since the 1980s, with the most  
16 recent years at an all-time low during this period. This trend correlates with declines in pH and  
17 dissolved oxygen, but not with changes in salinity. We examined possible physiological  
18 influences on B:Ca by including a collection of Icelandic cod as an out-group. Icelandic cod  
19 otoliths showed strongly positive correlations of B:Ca with physiologically regulated P:Ca; this  
20 was not the case for EBC. Finally, B:Ca in EBC otoliths is negatively correlated to some extent  
21 with Mn:Mg, a proposed proxy for hypoxia exposure. This negative relationship is hypothesized  
22 to reflect the dual phenomena of hypoxia and acidification as a result of decomposition of large  
23 algal blooms. Taken together, the otolith biomarkers Mn:Mg and B:Ca suggest a general increase  
24 in both hypoxia and acidification within the Baltic intermediate and deep waters in the last  
25 decade reflected in cod otoliths.

26 **Keywords:** Baltic Sea cod, otolith chemistry, boron, pH, hypoxia  
27  
28

## 29 **1 Introduction**

30 Ocean acidification, a known result of greenhouse gas emissions of CO<sub>2</sub> loading into the  
31 oceans, is highly variable in its distribution in the world's oceans (Doney et al., 2009). This is  
32 true in shallow coastal seas such as the brackish Baltic Sea, where in some locations surface  
33 pCO<sub>2</sub> values in winter greatly exceed other coastal regions (Reusch et al., 2018), but buffering  
34 can occur from river inputs bearing proton acceptors weathered from rock and soil (Müller et al.,  
35 2016; Kuliński et al., 2017). Alkalinity, defined as the sum of acid neutralizing compounds, is a  
36 complicated mix in the Baltic Sea (Kuliński et al., 2022). In the northern Baltic, i.e., the  
37 Bothnian Sea, total alkalinity ( $A_T$ ) is low as it is largely driven by  $A_T$ -poor riverine inputs from  
38 the Scandinavian Peninsula that is built mostly of granite. In the southern and central Baltic,  $A_T$   
39 is higher due to salt inputs from the North Sea, influence of  $A_T$ -rich continental rivers, and an  
40 additional contribution from internal cycling of carbon as algal blooms decompose at anoxic  
41 conditions (Kuliński et al., 2022). Despite buffering from alkalinity sources, pH is highly  
42 variable in space and time with its recent tendency to decline at greater depths in many parts of  
43 the Baltic Sea (Kulinski et al., 2022). This decline is associated with the ongoing eutrophication  
44 and higher vertical export of organic matter that leads also to worsening deoxygenation (Kuliński  
45 et al., 2022).

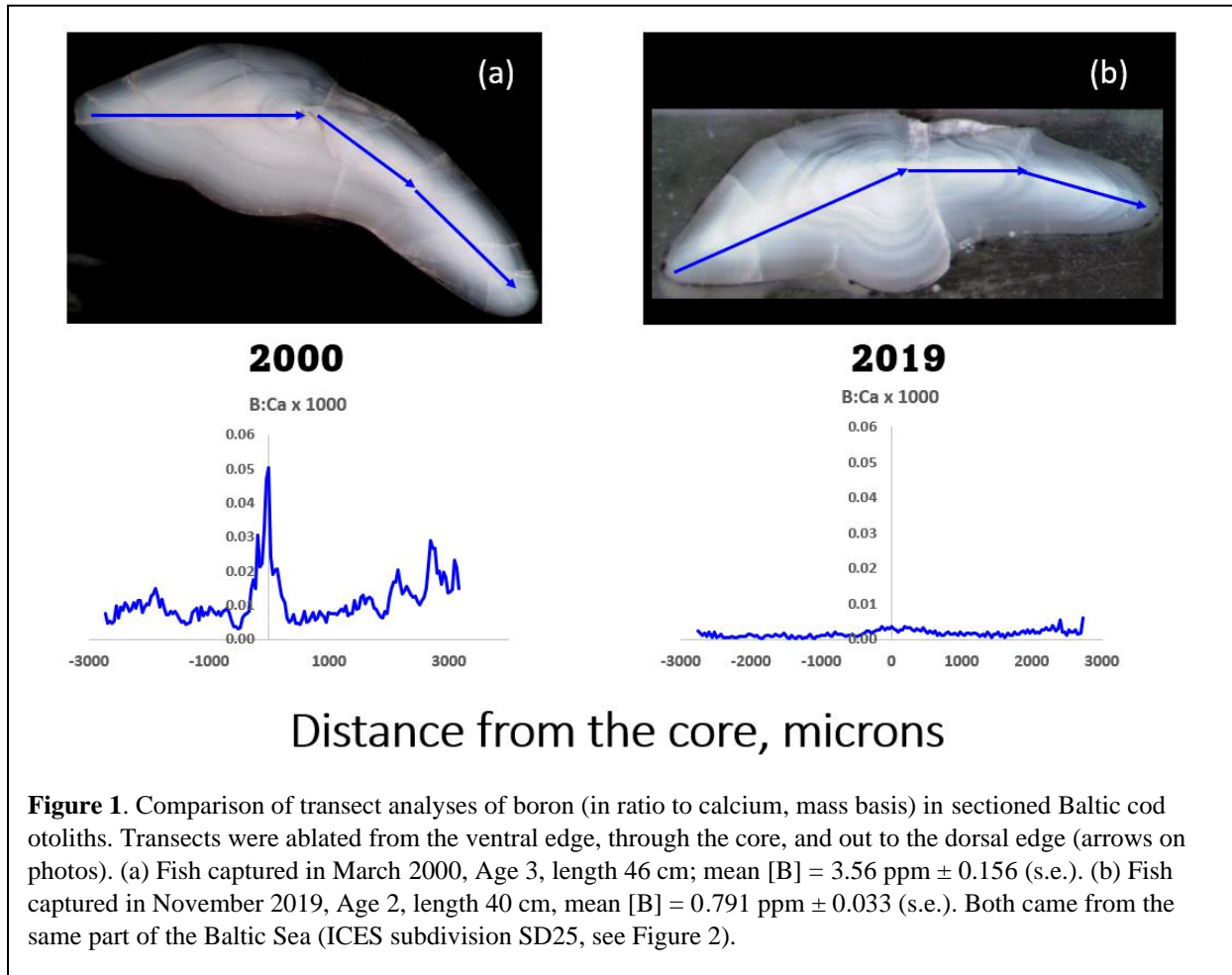
46 The population of the Eastern Baltic cod (*Gadus morhua*, hereafter referred to as EBC)  
47 has been severely impacted by a number of factors that include past overfishing, hypoxia,  
48 parasite infections, and seal predation (Eero et al., 2020), leading to reduced growth and age at  
49 maturity (Eero et al., 2016; ICES, 2019). Cod exposed to hypoxia were shown to have lower  
50 growth rates and worsened body condition, quantified by the Fulton K index of weight:length  
51 ratio (Casini et al., 2016; Limburg & Casini, 2018).

52 Recent studies that reported links between hypoxia exposure and body condition have  
53 relied on the measurement of trace elements in fish otoliths (ear-stones). Briefly, otoliths, the  
54 calcified structures that form part of the hearing/balance system in teleost fishes, are sectioned  
55 and then analyzed by ablating micro-transects along the major growth axis and analyzing by  
56 mass spectrometry (see Methods). Hypoxia exposure was quantified by means of a proxy  
57 developed from analyzing manganese concentrations in otoliths (Limburg et al., 2015), in ratio to  
58 magnesium which corrects for growth influences on Mn uptake, since both are affected by

59 growth but only manganese is affected directly by hypoxic conditions (Limburg et al., 2018;  
60 Limburg & Casini, 2018). The resulting elemental data are lifetime concentration histories, being  
61 incorporated at the time of exposure. The use of otolith manganese to track hypoxia exposure is  
62 one of the emerging biomarkers in fisheries ecology (Reis-Santos et al., 2022).

63         Recently, with acquisition of more sensitive instrumentation, we began to experiment  
64 with quantifying elements having sub-ppm concentrations in cod otoliths, including the trace  
65 element boron. Interestingly, preliminary results from this work showed elevated levels of boron  
66 in 2000 but not in 2019. Boron is noteworthy because it is an indicator of salinity. Specifically,  
67 Boron in seawater generally correlates with salinity (Kuliński et al., 2017) predominantly in the  
68 form of weak boric acid ( $\text{H}_3\text{BO}_3$ ) at standard seawater salinity (35 PSU) and pH of 8. However,  
69 the dissociated anion borate  $\text{B}(\text{OH})_4^-$  is positively, nonlinearly correlated with pH (Yu et al.,  
70 2007). Borate is the dominant form incorporated into carbonates such as limestones,  
71 foraminifera, and corals in proportion to pH (Yu et al., 2007; Hönisch et al., 2012; Levin et al.,  
72 2015) and is often confirmed by  $\delta^{11}\text{B}$  analysis, since it has a large, negative fractionation  
73 compared to boric acid. However, even bulk boron, as B:Ca ratios, has been found to correlate  
74 positively (and linearly) with pH in foraminifera (Yu et al., 2007), albeit in a species-specific  
75 manner.

76 We measured sub-ppm to low-ppm concentrations of B (ca. 0.5 – 5 ppm, see Methods  
77 section) in otoliths of EBC captured in 2000 that showed what appeared to be seasonal variations  
78 (example shown in Figure 1(a)), albeit not in synchrony with other elements that have proven  
79 useful to age cod (Heimbrand et al., 2020). It was therefore surprising when we began to analyze  
80 otoliths of EBC captured in 2019 and discovered greatly reduced concentrations (Figure 1(b)).  
81 The measurements were repeated for verification.



82 With such differences suggestive of change, we returned to an archive of otoliths used for  
83 another project (Limburg and Casini, 2018), as well as adding previously un-analyzed otoliths, to  
84 develop a time series of otolith B:Ca.

85 In this study, we explored the extent to which otolith B:Ca varied through time and  
86 whether its values were correlated with pH, salinity, or other otolith derived values that are  
87 proxies for environmental or physiological factors. In particular, otolith P:Ca is known to be

88 under physiological influence (Thomas et al., 2017; Heimbrand et al., 2020). In cod, higher  
89 values of P:Ca occur during the growing season and are thus hypothesized to be associated with  
90 growth and other activities. We selected this as well as our proxy for hypoxia exposure, otolith  
91 Mn:Mg, and posed the following questions:

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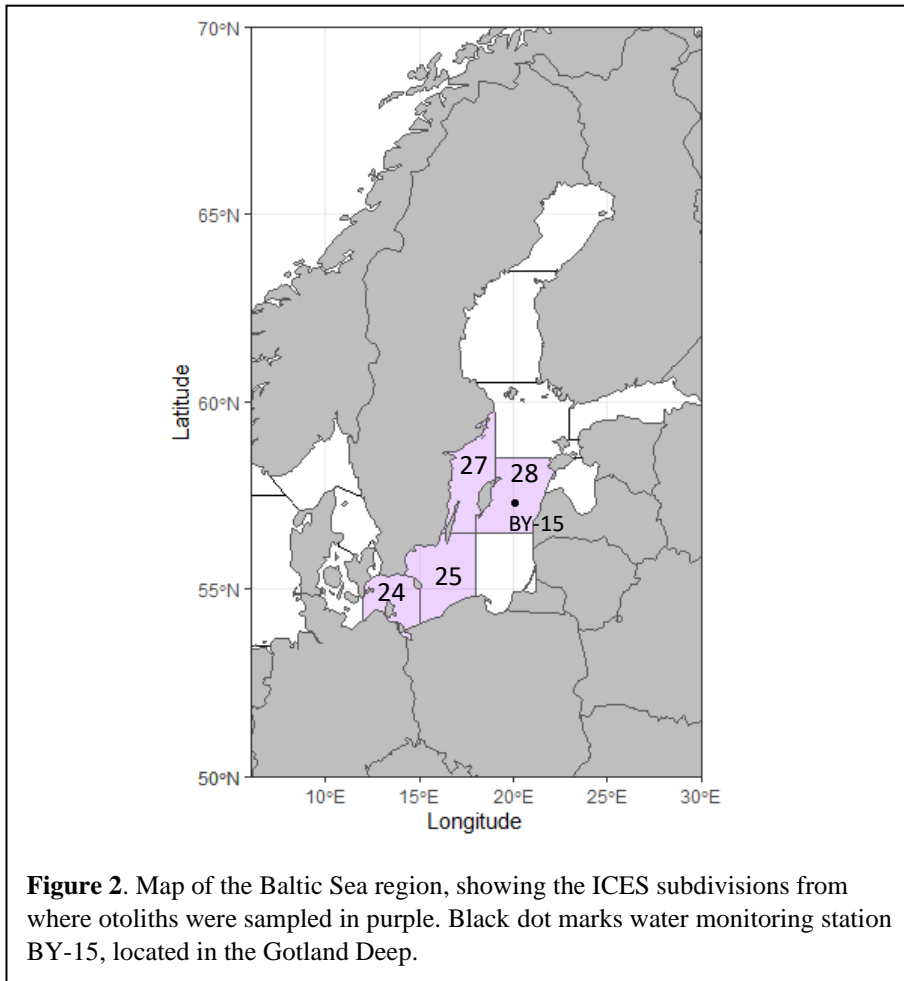
- 93 1. What are the temporal trends in otolith boron (as B:Ca ratios)?
- 94 2. How do B:Ca ratios correlate to trends in water chemistry, particularly salinity and pH?
- 95 3. If pH declines are associated with hypoxia, do we see a correlation of B:Ca with our hypoxia  
96 proxy, manganese:magnesium (Mn:Mg)?
- 97 4. If there are any physiological influences, do we see a correlation of B:Ca with elemental ratios  
98 known to be under physiological control, e.g., phosphorus:calcium (P:Ca)?

99

## 100 **2 Materials and Methods**

### 101 **2.1 Otolith collections and microchemical analysis**

102 Otoliths of Baltic cod (N = 156) were obtained from both fishery-independent and  
103 fishery-dependent surveys conducted by the Swedish Fisheries Board and its successor, the  
104 Department of Aquatic Resources, Swedish University of Agricultural Sciences. Fish were  
105 collected from ICES sub-divisions (SD) 24, 25, 27, and 28 (Figure 2), spanning the period 1988-



106 2021. To construct the time series, we selected previously analyzed (but not for B:Ca) otoliths that were comparable in age range to those collected in 2019-2021. All otoliths were measured, weighed, and subsequently embedded in epoxy (Struers EpoFix) and sectioned transversely to expose the core. They were smoothed with a succession of lapping films to 3 microns.

124 In addition, cod otoliths (N = 14) were obtained from fisheries-independent surveys from Iceland (ICES SD 5a), as an out-group for studies of possible physiological correlations. This part of the North Atlantic does not have hypoxia or acidification issues in the deep waters. The Icelandic otoliths were prepared and analyzed in similar fashion.

128 Baltic cod were aged by examining seasonal patterns in otolith Mg:Ca and P:Ca, following Heimbrand et al. (2020). Icelandic cod were also aged by otolith chemistry, although several of them had been previously aged visually by age readers in Iceland. For the latter otoliths, the seasonal chemical patterns produced age estimations that matched exactly with the visual estimates.

133 Chemical analyses were performed by laser ablation inductively coupled plasma mass  
134 spectrometry (LA-ICP-MS) at the Analytical and Technical Services group at SUNY ESF.  
135 Otoliths were ablated with a 192-nm laser ablation unit (Teledyne Cetac Excite 2) along transects  
136 as illustrated in Figure 1; the ablated material was transported via an Ar-He carrier gas mixture  
137 into a Thermo iCAP TQ plasma mass spectrometer where isotopes were quantified ( $^{11}\text{B}$ ,  $^{25}\text{Mg}$ ,  
138  $^{31}\text{P}$ ,  $^{43}\text{Ca}$ ,  $^{55}\text{Mn}$ ,  $^{63}\text{Cu}$ ,  $^{66}\text{Zn}$ ,  $^{88}\text{Sr}$ ,  $^{127}\text{I}$ ,  $^{138}\text{Ba}$ , and  $^{208}\text{Pb}$ ). Daily performance of the system was  
139 optimized with NIST 612 glass standard and U.S. Geological Survey standards MACS-3,  
140 MAPS-4, and MAPS-5 were used as standards for calibration and drift correction. The MAPS  
141 (abbreviation for Micro-Analytical Phosphate Standard) standards, while having good matching  
142 to the otolith matrix, were slightly softer and were therefore ablated at a higher energy (25%  
143 power,  $2.03\text{ J/cm}^2$  fluence) than the otoliths (20% power,  $1.62\text{ J/cm}^2$ ) to achieve an otolith  
144 calcium value around 38 weight percent. Transects were made with  $110\text{ }\mu\text{m}$ -diameter ablation  
145 spots at a speed of  $7\text{ }\mu\text{m/sec}$ . Although often edge-to-edge laser transects were made (to check  
146 for symmetry), only data collected on the dorsal (longer) axis were used in our analysis due to  
147 greater spatial resolution. Once data were worked up, annual mean values were parsed by the age  
148 determination technique in Heimbrand et al. (2020), i.e., averages within annual growth rings  
149 were calculated. These were matched to corresponding calendar years.

## 150 **2.2 Water data**

151 Water properties data (temperature (T, °C), salinity (S, psu), dissolved oxygen (DO,  
152 mL/L), pH, and total alkalinity ( $A_T$ , mmol/L)) were downloaded from the Swedish  
153 Meteorological and Hydrological Institute's database, SHARKWeb (<https://sharkweb.smhi.se/>).  
154 Station BY-15, Gotland Deep, was selected as being representative of the central Baltic (Figure  
155 2) and having one of the longest and most detailed timeseries. We selected depths that  
156 corresponded to where cod had been found in fishery-independent surveys (Casini et al., 2021).  
157 Between 1985-1995, we averaged water data from 30-60 meters, and from 1996 onward,  
158 between 40-75 meters. Annual mean water values were computed and matched to corresponding  
159 otolith chemistry data (annual means of Element:Ca data parsed to calendar years as described  
160 above).

161

## 162 **2.3 Analysis**

163 Visualization (graphs) and statistical analyses were performed in Excel, R (base R (R  
164 Core Team 2022) and ggplot2 (Wickham, 2016)), and JMP version 17.0 (JMP 2022). Statistical  
165 analyses were separated into examination of relationships of B:Ca values to potential  
166 environmental drivers (primarily pH and S, but also  $A_T$ , DO, and T) and to the internal variables  
167 P:Ca and Mn:Mg. For the environmental drivers, a PCA analysis was first performed to examine  
168 relationships among them. Thereafter, B:Ca was regressed on the first principal component  
169 eigenvector to test for its relationship to the main drivers. This was followed by mixed linear  
170 models of environmental drivers on B:Ca, with each fish's age (nested within fish ID) being  
171 treated as a random effect. This analysis took account of repeated measures within the same  
172 individual, and was intended to examine more closely the effect of individual drivers.

173 For the internal variables, the P:Ca and Mn:Mg transect data were regressed on  
174 corresponding B:Ca for each fish, and the slopes, intercepts and  $R^2$  values were calculated. These  
175 were examined for trends by decade (1980s, 1990s, 2000s, and 2010s) as well as compared with  
176 the Icelandic out-group. Positive correlations with P:Ca would suggest a physiological influence  
177 on B:Ca incorporation, whereas negative correlations with Mn:Mg would suggest a linkage  
178 between B:Ca as an indicator of pH and our indicator of hypoxia exposure (cf. Cavole, 2021).  
179 These were tested using ANOVA with slopes as the dependent variable and decades (for Baltic  
180 Sea) and Iceland as grouping variables.

181

## 182 **3 Results**

### 183 **3.1. Trends in water variables**

184 A principal components analysis of the water variables found that the first two  
185 components explained 92.9% of the variance (72.4% by PC1 and 20.5% by PC2). The analysis  
186 showed three groupings: (1) DO and pH that were opposite to (2) salinity and  $A_T$ , respectively;  
187 and (3) temperature was on a separate axis (Figure S1). Linear regression of B:Ca on the first  
188 principal component yielded a negative relation ( $B:Ca = 0.000485(PC1) + 0.00246$ ,  $R^2 = 0.21$ ,  $p$   
189  $< 0.0001$ ).

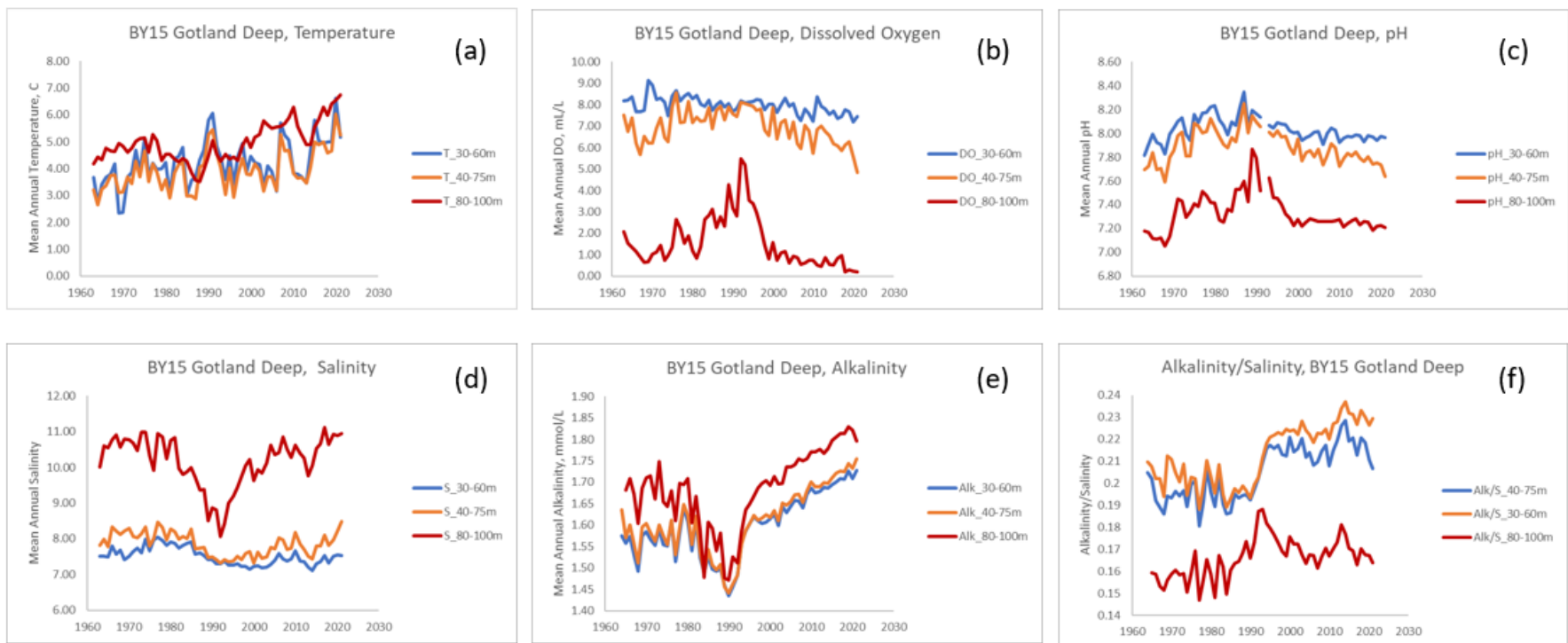
190 Examination of the long-term time series of data from Station BY-15 shows that while  
191 temperature generally increased (Figure 3(a)), the other water variables showed very nonlinear



192 patterns (Figures 3(b-e). Alkalinity is in part a function of salinity; dividing  $A_T$  by  $S$  produces a  
193 time series showing a dramatic state change around 1990 (Figure 3(f)), particularly in the water  
194 layers occupied by cod. Despite the increasing  $A_T$ , pH has declined more or less monotonically  
195 at midwater depths (40-75 m, Figure 3(c)).

196

### Station BY-15, Gotlandsdjupet 1963-2021



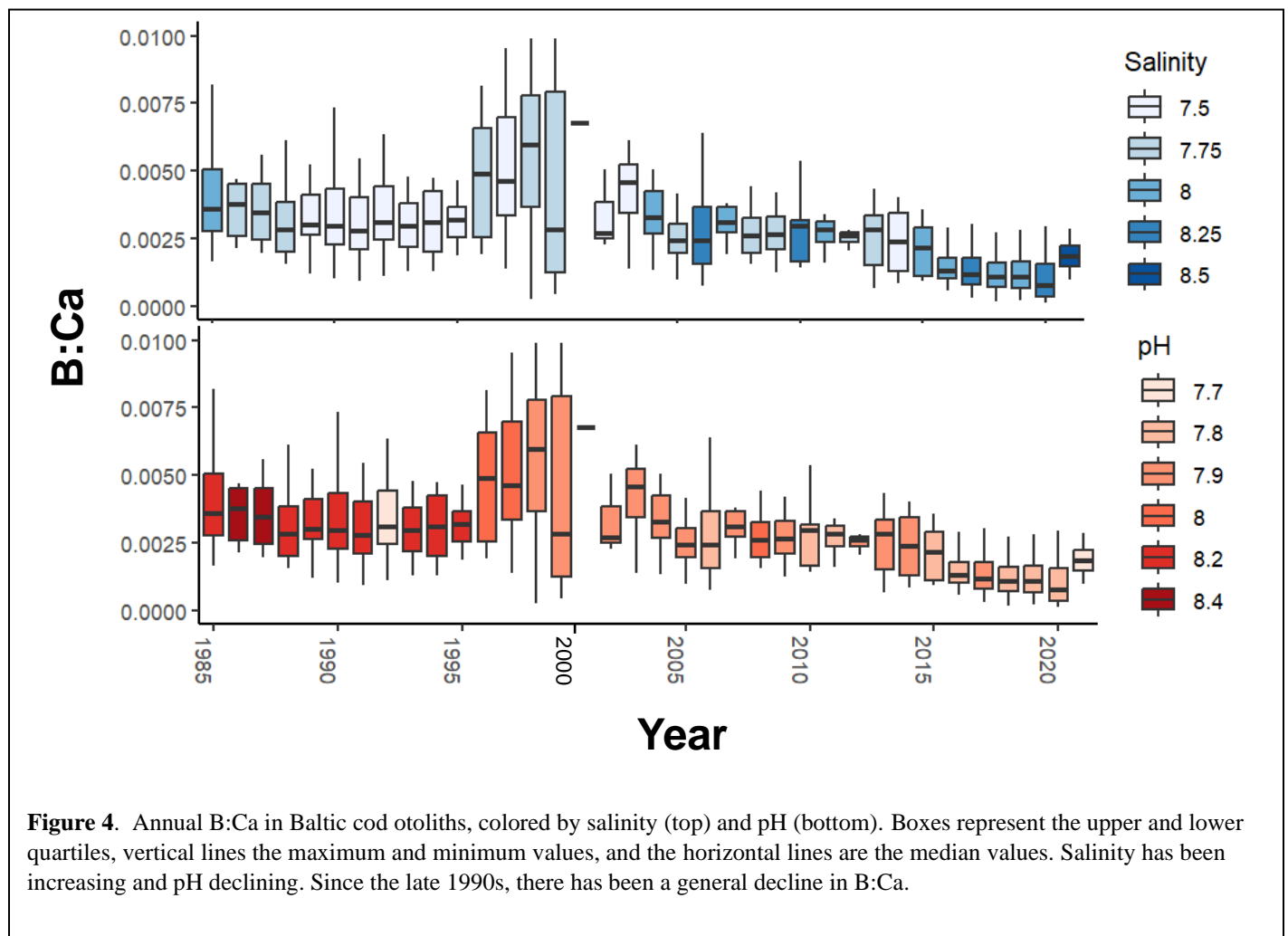
**Figure 3.** Time series of water data for three depth zones, Gotland Deep Station BY-15. (a) Temperature. (b) Dissolved oxygen. (c) pH. (d) Salinity. (e) Total alkalinity. (f) Alkalinity normalized to salinity.

199

### 200 3.2 Trends in otolith B:Ca with salinity and pH

201 Over the period 1985-2019, box plots of B:Ca by year show a pattern of increase toward  
202 a maximum, albeit with great variation, in the late 1990s, followed by gradual decline (Figure 4).  
203 Visual comparisons with salinity and pH show that neither explain all the variation in B:Ca, but  
204 since 2000, pH and B:Ca both trend downward whereas salinity has increased. The increase in  
205 B:Ca in 2021 is from age-0, which is elevated relative to older ages (cf. Figure 1(a)).

206



**Figure 4.** Annual B:Ca in Baltic cod otoliths, colored by salinity (top) and pH (bottom). Boxes represent the upper and lower quartiles, vertical lines the maximum and minimum values, and the horizontal lines are the median values. Salinity has been increasing and pH declining. Since the late 1990s, there has been a general decline in B:Ca.

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209 **3.3. Effects of water parameters on otolith B:Ca**

210 **3.3.1. Univariate effects of water parameters on otolith B:Ca**

211 Following the PCA, repeated measures regressions of otolith B:Ca on each of the five  
212 water variables was conducted to examine effect sizes and significance each variable alone.  
213 These regressions showed significant effects of each of the five water variables on otolith B:Ca  
214 (Table 1).  $A_T$  had the largest effect and was negative (Table 1). Dissolved oxygen and pH both  
215 had positive slopes, and temperature and salinity both were negative. All univariate analyses  
216 were highly significant. Ages of individual fish were not significant as measured by Wald's p-  
217 value in restricted maximum likelihood (REML) analysis.

**Table 1.** Repeated measures univariate regressions of otolith B:Ca on water variables, ordered by goodness of fit (-2 log likelihood criterion; lowest is best).

Variable	Parameter slope	R <sup>2</sup>	df	F ratio	p	-2 Log Likelihood
Total Alkalinity	-9.42E-03	0.24	647	158.8	< 0.0001	-6358
Dissolved Oxygen	1.00E-03	0.21	615	142.2	< 0.0001	-6341
Salinity	-3.00E-03	0.22	565	133.5	< 0.0001	-6336
Temperature	-8.30E-04	0.13	639	58.7	< 0.0001	-6269
pH	5.10E-03	0.15	633	109.4	< 0.0001	-6189

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221 **3.3.2. Two-factor tests of water parameters on otolith B:Ca**

222 Because of hypothesized positive relationships of salinity and pH on B:Ca, as well as the  
223 large effect size of  $A_T$ , we examined pairwise (two-factor) models of these independent variables  
224 on otolith B:Ca (Table 2). This allowed us to examine the relative strength and direction of  
225 trends (slope parameters, Table 2) on B:Ca when testing two independent variables together.  
226 Aside from an overall model using all the data, these were filtered by decade to examine finer  
227 scale trends.

**Table 2.** Repeated measures 2-factor models of effects on otolith B:Ca, by decade. Numbers are estimated coefficients. \*\* = significant at  $p < 0.01$ . \* = significant at  $p < 0.05$ . Negative 2 Log Likelihood is a metric of goodness-of-fit, with smaller numbers indicating better fit.

1. pH and

Alkalinity:

Parameter	1980s	1990s	2000s	2010s	Overall
pH	0.00128	-0.00774*	-0.00375	-0.00121	-0.00385*
Alkalinity	0.00873	-0.00538	-0.01840*	-0.02585**	-0.01585**
-2 Log Lik.	-399.2	-1231.8	-896	-3979.4	-6225.5

2. pH and Salinity:

Parameter	1980s	1990s	2000s	2010s	Overall
pH	0.00151	-0.00226	-0.00983**	0.00378*	0.00225**
Salinity	0.00322	0.00365	-0.00347**	-0.00049	-0.00208**
-2 Log Lik.	397.9	-1231.6	-894.8	-3948.2	-6201.6

3. Alkalinity and Salinity:

Parameter	1980s	1990s	2000s	2010s	Overall
Alkalinity	-0.007269	0.002042	-0.01253	-0.027822**	-0.006591**
Salinity	0.003768	0.004802**	-0.000171	0.000478	-0.001274**
-2 Log Lik.	-401.4	-1357.9	-891.9	-3977.6	-6354.8

228

229 In both pairwise tests with pH, this parameter had positive slopes in the 1980s and 2010s  
 230 but negative slopes in the 1990s and 2000s. Alkalinity had mostly a negative relation to otolith  
 231 B:Ca. Salinity had a positive effect in the 1980s and 1990s but became negative in the 2000s and  
 232 2010s. Salinity overall had a negative effect on B:Ca, in contrast to expectations. We note the  
 233 strong, negative effect of pH in the 1990s when paired with  $A_T$ , and the strong, positive effect of  
 234 pH in the 2010s when paired with salinity. Age, when treated as a random variable nested within  
 235 individual fish (by fish ID), was again not significant. However, in a separate analysis (Figure  
 236 S2), age (not treated as a random variable) and otolith B:Ca were inversely related in the 1980s,  
 237 1990s, and 2000s, and had a weakly positive trend in the 2010s.

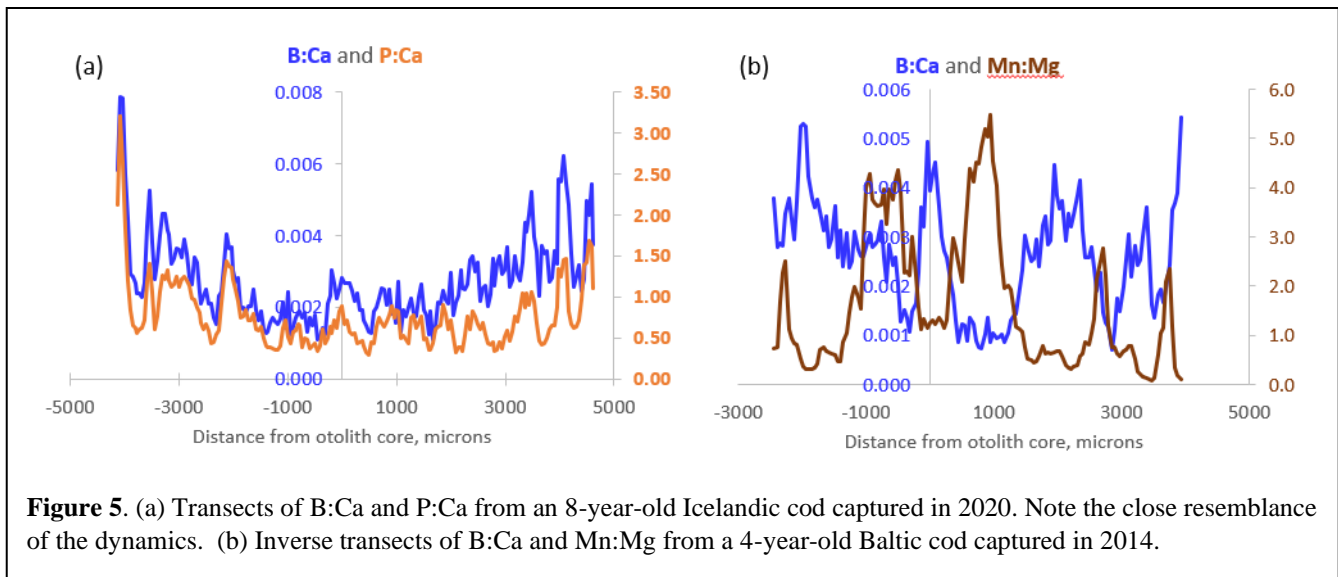
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241 **3.4. Internal parameters and otolith B:Ca**

242 During exploratory examination of the data, P:Ca was found to correlate highly with  
243 B:Ca in Icelandic cod otoliths (Figure 5(a)). Otolith Mn:Mg, on the other hand, is our proxy for  
244 hypoxia exposure (Limburg and Casini, 2018), and often correlated negatively with B:Ca (Figure  
245 5(b)). This is consistent with co-occurrence of reduced conditions in water (and presence of  
246 bioavailable ionic  $Mn^{2+}$ ) and high  $pCO_2$  and thus low pH, both being the effect of enhanced  
247 organic matter remineralization and/or poor ventilation of the water layers occupied by cod.

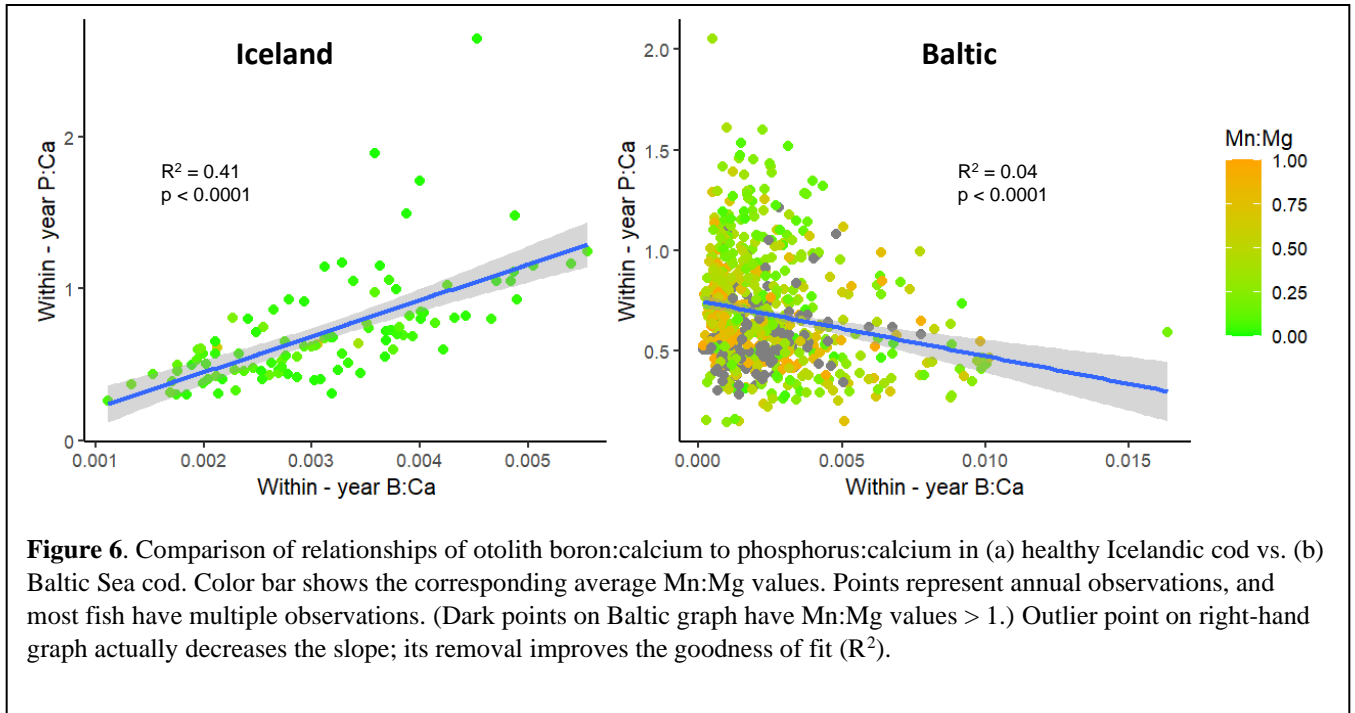


**Figure 5.** (a) Transects of B:Ca and P:Ca from an 8-year-old Icelandic cod captured in 2020. Note the close resemblance of the dynamics. (b) Inverse transects of B:Ca and Mn:Mg from a 4-year-old Baltic cod captured in 2014.

248

249 **3.4.1. P:Ca as a proxy for physiological status**

250 For this analysis, the Icelandic “out-group” of cod living in normoxic water were  
251 included. Annual average P:Ca was highly correlated with B:Ca in Icelandic cod otoliths, but  
252 inversely correlated in Baltic cod with a great deal of scatter (Figure 6). Although the overall  
253 trend for Baltic cod is negative, this hides trends that become more visible at a finer temporal  
254 scale. Thus, we examined the slopes and  $R^2$  values of individual fish regressions of P:Ca on B:Ca  
255 (Figure S3). The 2010s showed the strongest trends when slopes were plotted on  $R^2$  values, with  
256 Icelandic cod and the 2000s also trending positively, and weaker trends in the 1980s and 1990s  
257 (Figure S3(a)). Box plots of  $R^2$  values by decade and out-group showed wide variation (Figure  
258 S3(b)), and an analysis of variance found no statistical difference among Baltic fish by decade,  
259 but Icelandic otoliths had significantly higher  $R^2$  values ( $p < 0.001$ ).



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262

### 263 3.4.2. Mn:Mg as a proxy for hypoxia exposure

264 We tested the hypothesis that hypoxia (indexed by increased otolith Mn:Mg) was coupled  
 265 to acidification (indexed by declining B:Ca). We used a similar analysis of individual fish  
 266 regressions of Mn:Mg, our hypoxia exposure proxy, on B:Ca, as was performed with P:Ca (see  
 267 Methods). This analysis led to somewhat inconclusive results (Figure S4). The decades of the  
 268 1980s, 1990s, and 2000s all had mostly positive slopes of Mn:Mg with B:Ca, and slopes  
 269 increased with increasing  $R^2$  values. Only the 2010s displayed a pattern consistent with the  
 270 hypothesis. Iceland showed no relationship of slopes to goodness of fit. Further examination of  
 271 the data included plotting trends of these slopes against fish age (Figure S5). This revealed that  
 272 younger EBC otoliths tended to have positive slopes, with slopes becoming more negative with  
 273 age, for the decades of the 1990s-2010s; no trend was observed in the 1980s, and Icelandic cod  
 274 had increasingly positive slopes as they aged (Figure S6). For the 2010s, it suggests that as  
 275 Baltic cod age they encounter more hypoxia perhaps as they use deeper habitats, and with it, a

276 decline in B:Ca that may indicate lower environmental pH. Overall, this analysis suggests that  
277 the decade of the 2010s may have had evidence of acidification driven by the same mechanisms  
278 that produced hypoxia.

279

#### 280 **4. Discussion**

281 This study was initiated because of finding a large discrepancy in otolith B:Ca ratios  
282 between fish captured in 2000 and others captured in 2019 – 2021. Archived otoliths from  
283 previous analyses were then used to create a time series of data from 1985-2021. Otolith B:Ca  
284 ratios were analyzed in relation to long-term records of water chemistry in a central location in  
285 the Baltic Sea to represent trends in environmental conditions. Ratios were also examined against  
286 an elemental ratio (P:Ca) determined in other studies to be controlled physiologically, as well as  
287 a ratio (Mn:Mg) that serves as a proxy for hypoxia exposure (Limburg and Casini, 2018).

288 Boron has been included in a number of otolith chemistry studies, mostly as part of a  
289 multivariate suite of tracers to determine provenance. Martino et al. (2017) studied boron as part  
290 of a suite of trace elements that might be affected by experimentally elevated pCO<sub>2</sub>, but found no  
291 effect. Cavole (2021) performed the first studies of otolith B:Ca as a potential proxy for ocean  
292 acidification, using wild fishes across a natural pH gradient (7.4 to 8.1). Cavole et al. (2023),  
293 examining potential tracers of warming, hypoxia, and acidification in deep-sea fishes, found  
294 wide variation in B:Ca but no consistent trends.

295 In this study, univariate regressions of otolith B:Ca on five water variables all were  
296 highly significant, with dissolved oxygen and pH showing positive relationships while negative  
297 relationships were observed with alkalinity, temperature, and salinity (Table 1). These variables  
298 are interrelated through ecological processes: warming temperatures and salinity in combination  
299 with nutrients produces massive algal blooms in the Baltic Sea, which as they die off and decay,  
300 consume oxygen and change the pH and alkalinity (Kuliński et al., 2022). In that analysis, total  
301 alkalinity ( $A_T$ ) showed the strongest effect, followed by dissolved oxygen and salinity.

302 Further pairwise analysis of  $A_T$ , pH, and salinity by decade showed varying patterns of  
303 effect sizes and significance. Importantly, when testing pH and salinity together, pH had a  
304 significant, positive effect on B:Ca and salinity a significant, negative effect. Kuliński et al.



305 (2018), building on earlier studies, showed that dissolved boron is tightly coupled to salinity in  
306 the Baltic Sea, with an intercept  $> 0$  indicating watershed sources of B input. The decoupling of  
307 otolith B:Ca from salinity and strong relationship to pH and alkalinity in the 2010s suggests that  
308 environmental conditions moved into a different regime from the previous three decades.

309         Among the analyzed environmental variables (Figure 3), temperature shows a clear,  
310 positive trend starting from the 1980s. This is consistent with analyses performed by Lehmann et  
311 al. (2022) who found warming of 0.3-0.5 °C per decade in the whole water column in the central  
312 Baltic (station BY15) between 1979 and 2018. For other variables, trends are not that  
313 straightforward. For salinity, the variability reveals a stratification pattern typical for the central  
314 Baltic, with significantly higher S in the deep waters and a brackish surface water layer.  
315 However, in the 1980s and early 1990s there was a clear S drop in the deep waters being the  
316 result of a lower frequency of so-called Major Baltic Inflows that bring dense, saline water from  
317 the North Sea to the Baltic (Meier et al. 2022). This relaxed haline stratification in the central  
318 Baltic, which in turn led to better ventilation of deep waters and significantly raised O<sub>2</sub>  
319 availability (Figure 3 (b) and (d); Lehmann et al., 2022) increased water volume for cod  
320 spawning and feeding. However, the 1980s decade was also the time when the maximum  
321 nutrient loads from land occurred, propelling eutrophication in the central Baltic. High organic  
322 matter production (and its export to bottom waters) co-occurred with the re-enhancement of  
323 water column stratification at the turn of the 20<sup>th</sup> and 21<sup>st</sup> centuries (increase in bottom S, Figure  
324 3 (d)). Both these factors led to a significant drop in O<sub>2</sub> in deeper waters (Figure 3 (b); Kuliński  
325 et al. (2022)), which limited the optimal water volume for cod expansion – a negative factor, that  
326 also coincided with high fishing exploitation.

327         Furthermore, all these changes together with a rising pressure from atmospheric CO<sub>2</sub>  
328 increase made fundamental changes in the marine CO<sub>2</sub> system and seawater pH. First, the  
329 relatively low pH in the 1960s was increasing by the 1990s as an initial effect of eutrophication  
330 (Figure 3 (c)). Surprisingly, this trend was paralleled with a drop in A<sub>T</sub> (Figure 3 (e)), which is a  
331 counterintuitive phenomenon in marine CO<sub>2</sub> system studies. However, the relatively constant  
332 A<sub>T</sub>/S ratio (Figure 3 (f)) clearly suggests that this initial A<sub>T</sub> decrease was an effect of salinity  
333 decrease. Already in the 1990s we can observe a change in trends and a clear decoupling of A<sub>T</sub>  
334 from S as revealed by the marked A<sub>T</sub>/S increase, suggesting A<sub>T</sub> increase being higher than

335 observed S changes at that time (Figure 3 (d,e,f)). While in surface waters this  $A_T$  increase was  
336 likely due to climate-related (or  $CO_2$ -induced) enhancement of weathering on land and higher  
337 riverine  $A_T$  loads, the  $A_T$  increase observed in deep waters was probably an effect of  
338 hypoxic/anoxic  $A_T$  generation (Müller et al., 2016, Neumann et al., 2022). The latter is consistent  
339 with a large and sudden expansion of the hypoxic/anoxic areas in the central Baltic that occurred  
340 in the late 1990s (Meier et al., 2022) – a phenomenon that is also believed to be a root cause for a  
341 biogeochemical regime shift to the so-called vicious cycle – a self-supporting eutrophication  
342 mechanism (Kuliński et al., 2022). Due to these effects, pH after reaching its maximum in the  
343 late 1980s (effect of eutrophication), dropped somewhat and remained relatively constant in the  
344 21<sup>st</sup> century expressing in that way the mutually cancelling effects of ocean acidification (effect  
345 of rising atmospheric  $CO_2$ ) and hypoxia/anoxia/weathering-driven alkalization.

346 As is the case for some other trace elements (Hüssy et al., 2021), in this study boron  
347 uptake into otoliths appears to have some unexplained physiological regulation. This conclusion  
348 is based on close correspondence between B:Ca and P:Ca in otoliths from Icelandic cod that are  
349 not stressed by hypoxia (Figure 4(a), Figure 6(a)). In contrast to the Icelandic cod, the  
350 relationship of P:Ca to B:Ca in Baltic cod otoliths was highly variable, with a negative trend in  
351 the aggregate (Figure 6(b)). Regressions of P:Ca on B:Ca conducted on individual fish revealed  
352 mostly positive slopes, with highest slopes observed in the 2010s (Figure S3(a)), but lower, more  
353 variable  $R^2$  values (Figure S3(b)). Although not statistically different,  $R^2$  values were lowest in  
354 the 1980s and next-lowest in the 2010s. Compared to the Icelandic cod, all Baltic otoliths  
355 showed less coupling between P:Ca and B:Ca.

356 The  $CO_2^-$  and  $A_T$ -induced pH changes have definitely modified borates' availability in  
357 water and likely the B:Ca ratio in the Baltic cod's otoliths (Fig. 5). However, as shown by the  
358 strong correlation of B:Ca and the physiologically regulated P:Ca in the healthy, normoxic  
359 Icelandic cod's otoliths, the B:Ca ratio may also contain built-in information about the overall  
360 physiological condition of fish (and not only the pH-dependency). This suggests that a complex  
361 picture of B:Ca development in the EBC otoliths over the recent decades may be affected by  
362 multiple factors, not only by pH changes but also physiological stress caused by overall changes  
363 that have recently occurred in the Baltic (Eero et al., 2020). These include expansion of

364 hypoxia/anoxia, eutrophication, and biogeochemical regime shift. Together with overfishing,  
365 these factors led to an overall collapse of the EBC population in the Baltic.

366 Examination of B:Ca in relation to an otolith index of hypoxia exposure, Mn:Mg,  
367 revealed mostly positive slopes of individual fish regressions, with the notable exceptions of  
368 Baltic cod from the 2010s and Icelandic cod (Figure S4). Limburg and Casini (2018) adopted  
369 the ratio Mn:Mg as a hypoxia exposure proxy rather than Mn:Ca in part to correct for growth  
370 influences on Mn (see Limburg et al., 2015) as otolith Mg is related to metabolism and growth  
371 (Limburg et al., 2018). The positive relationships of Mn:Mg to B:Ca cannot be fully explained,  
372 but the inverse relationships observed in the 2010s is consistent with the hypothesis that cod  
373 were exposed to both hypoxia and lower pH (and also higher alkalinity). Some of the positive  
374 slopes were due to young ages of fish; we observed more coupling of Mn:Mg and B:Ca in young  
375 fish, and we hypothesize that at younger ages the growth effect on otolith Mn is not fully  
376 removed by dividing by Mg. Baltic cod had increasingly negative slopes of the Mn:Mg vs. B:Ca  
377 regressions as they aged (Figure S5), with the 2010s displaying the most negative slopes.

378

## 379 **5. Conclusions**

380 Baltic cod showed variable patterns of otolith B:Ca over a 35-year period, with maximum  
381 values in the late 1990s, and a gradual decline in B:Ca thereafter, to an all-time low in 2019-20.  
382 These trends are imperfectly related to water chemistry data, but showed strong correlations that  
383 varied by decade. Otolith B:Ca ratios and relationships in the 2010s were suggestive of  
384 environmental changes in pH and total alkalinity that could be coupled to deoxygenation.  
385 Physiological influences observed in a healthy Icelandic population were diminished in Baltic  
386 cod.

387 Undoubtedly, some of the relationships found could be due to the choice of samples, and  
388 a more focused study might select otoliths from areas where trends in acidification are clear.  
389 Future work could also examine otoliths from other species, as well as determine, if possible, the  
390 form of boron (borate or boric acid) taken up by otoliths. Complex interactions notwithstanding,  
391 we suggest that B:Ca in cod's otoliths can be a prospective variable in the palette of important  
392 and available tools to look into the environmental changes through the lens of otolith chemistry.

393

394 *Code/Data availability.* The codes and data are available upon request to the corresponding  
395 author.

396 *Author contributions:* KL designed and carried out the study. YH contributed data and figures.  
397 All authors contributed text.

398 *Competing interests.* The authors declare that they have no conflict of interest.

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401

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