Measurement Report: Atmospheric aging of combustion-derived particles: impact on stable free radical concentration and its ability to produce reactive oxygen species in aqueous media

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Abstract. Environmentally persistent free radicals (EPFR) are a pollutant found on fine atmospheric particulate matter (PM$_{2.5}$), particularly on PM$_{2.5}$ formed from combustion processes. EPFR are organic radicals that can endure in the environment for days to years. Interest in the toxicity of EPFR has increased significantly in recent years, as it has been shown to have substantial ability to form reactive oxygen species (ROS), but little is known about how its characteristics change as PM$_{2.5}$ ages in the atmosphere. Here, we exposed newly produced hexane-generated soot to simulated sunlight for 24 hours. Changes to the EPFR characteristics of the particles were measured by electron paramagnetic resonance (EPR) spectroscopy. The soot was then added to water and a second exposure to light was used to measure hydroxyl radical (OH) formation from both photo-aged and dark-aged soot. There were no changes to EPFR characteristics (spin concentration, g-factor, peak width, or lineshape) due to the exposure to simulated sunlight, however the soot's ability to form OH was greatly reduced by photo-aging. Photo-aged soot resulted in an almost 60% reduction in OH formation over soot which had been aged in the dark for the same amount of time.

1 Introduction

Fine atmospheric particulate matter, particles smaller than 2.5 µm in diameter (PM$_{2.5}$), are a well-documented environmental problem. Because of their small size, PM$_{2.5}$ can penetrate deep into the lungs where they have been shown to cause oxidative damage via the production of reactive oxygen species (ROS) (Feng et al., 2016; Vidrio et al., 2009)PM$_{2.5}$ has been linked to many cardiopulmonary health concerns such as asthma, high blood pressure, and COPD (He et al., 2022; Chan et al., 2008), as well as diabetes (Brook et al., 2012, 2013) and cancer (Guo et al., 2020; Katanoda et al., 2011). One of the major sources of PM$_{2.5}$ is combustion, from sources such as automobile exhaust, power plants, and wildfires. Climate change has already driven an increase in the number and severity of wildfires globally (Coop et al., 2022) and there is evidence that PM$_{2.5}$ from wildfires may be more hazardous than from other sources (Aguilera et al., 2021).

PM$_{2.5}$ can stay aloft for a very long time and can travel thousands of kilometers from its point source. During this time, it is subjected to various environmental conditions such as weather, sunlight, and interaction with atmospheric gasses like NO$_X$, SO$_X$, or reactive oxygen species (ROS, e.g. OH, O$_3$, O^{2-}), all of which can alter the chemical composition of PM$_{2.5}$ components, resulting in aging of the particles (Liu and Chan, 2022; Donahue et al., 2009).

Additionally, PM$_{2.5}$ has been shown to contain substantial concentrations of environmentally persistent free radicals (EPFR) (Chen et al., 2018; Runberg et al., 2020), a newly identified class of pollutant. Radicals are chemical species with at least one unpaired electron and typically are very reactive, with lifetimes generally on the scale of nanoseconds. EPFR, however, persist for days, weeks, or even years in the environment making them an important environmental contaminant (Gehling and Dellinger, 2013). EPFR are generally carbon- and oxygen-centered radicals and have been shown to form from the same combustion systems that result in PM$_{2.5}$ (Wang et al., 2019). Both PM$_{2.5}$ and EPFR have been reported to induce ROS formation (Tong et al., 2018; Arangio et al., 2016; Khachatryan et al., 2011), and it is possible that the EPFR component is a significant contributor of PM$_{2.5}$'s ability to form ROS.

40 To date, little is known about the impacts of atmospheric aging on EPFR bound to $PM_{2.5}$, such as its character, concentrations, or its ability to mediate ROS formation. One of the key components of $PM_{2.5}$ is soot, which contains EPFR, and as such it was chosen as a surrogate for environmental PM because its age and environmental encounters could be controlled.

In this study, the EPFR characteristics of fresh hexane-derived soot is compared to soot that has been photo aged. EPFR concentrations and characteristics are measured by electron paramagnetic resonance (EPR) spectroscopy, Fourier-transform

45 infrared (FT-IR) spectroscopy, and gas chromatography mass spectrometry (GC-MS) before and after illumination under a solar simulator. Additionally, the ability of both the photo-aged and non-aged soot to produce OH in aqueous media is studied, and the EPFR characteristics of the soot after OH formation is also investigated.

2 Materials and Methods

2.1 Photo-aging of combustion particles

50 **Stage I:** Hexane derived soot was produced as previously described (Runberg et al., 2020) by using a diffusion flame with a glass funnel inverted above the flame to accumulate the soot. The soot was collected from the funnel into a disposable aluminum baking dish (Reynolds KITCHENS® mini loaf pan) as soon as it was cool (Figure 1-I). The soot was transferred to the aluminum baking dishes only after it was fully cooled to prevent any interference from the foil on the EPFR structure.

Stage II: The bulk soot was divided roughly evenly between two additional baking dishes. One dish was placed under a xenon

55 lamp solar simulator (Oriel Sol 1A, Newport Solar Simulator equipped with an AM 1.5 Global filter) calibrated to one sun at the average North American zenith, and the other was kept in the dark by covering it with aluminum foil. The soot was aged in this manner for 24 hours (Figure 1-II). This process was completed twice, once to produce soot for photoreactions in stage III which included an OH probe, and once for reactions executed without an OH probe.

Stage III: After the soot had been aged for 24 hours, either with or without light, it was combined with ultra-pure water (18

60 MΩ) (Figure 1-III) and stirred overnight to create a well-mixed suspension, or slurry. Using a microbalance, 0.0225 ± 0.0005 g of soot was combined with 500 mL of MQ water for a final soot concentration of 45 ± 0.5 ppm (m/v) for slurries that were used for OH measurements. This is in line with concentrations of $PM_{2.5}$ obtained in previous work (Leresche et al., 2021). For slurries used for all other analyses, 118 ± 2 ppm (m/v) was used. Slurry concentrations for non-OH analyses were higher to allow for enough soot to be collected after the photoreactions for EPFR, FT-IR, and GC-MS analysis.

65 Samples of the prepared slurries were poured into two Teflon reaction vessels for the aqueous reaction in Stage III. Both reaction vessels were placed into water-cooled, jacketed glass beakers which were kept at 25℃. One was located under the solar simulator and the other was covered with aluminum foil to keep it dark. The slurries were stirred continuously for the 16-hour reaction time. For EPFR, FT-IR, and GC-MS analyses, the reactants were poured into clean baking dishes and dried overnight in an oven set to 50℃. Once the soot was dry, it was collected for further analysis. All samples were stored in a

70 freezer when not being analyzed, with storage times ranging from two to seven days (Table A1).

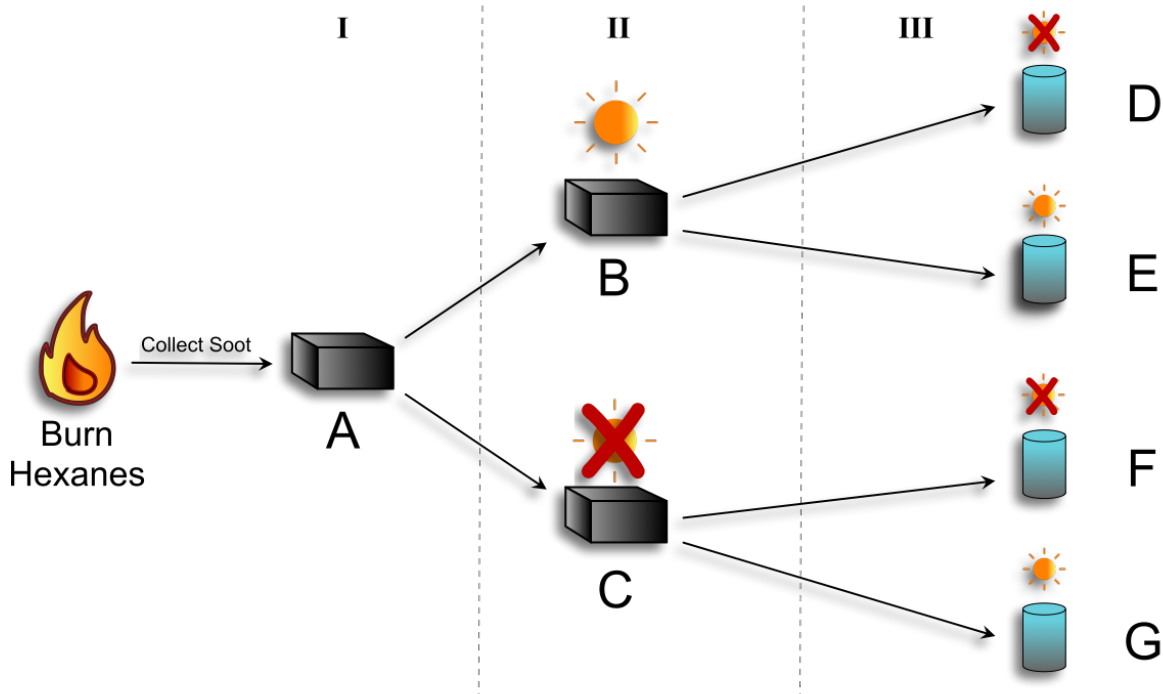



Figure 1. Schematic of photo-aging experimental setup. Hexanes were burned and resultant soot was collected (A). That soot was divided, and half was exposed to simulated sunlight for 24 hours (B) and half was kept in the dark (C). The soot from each was then added to water to make a slurry. The two soot slurries were then divided in half, and one was exposed to simulated sunlight for 16 hours (E and G) and the other was kept in the dark (D and F). Soot characterization was done at each stage (I, II, and III), via GC-MS, FT-IR and EPR. OH production was measured for stage III.

2.2 EPR Analysis

Samples of soot ranging between $0.4 – 0.9 \pm 0.1$ mg were collected in triplicate in 1.5 mm ID Pyrex closed-bottom capillary tubes after each stage of photo-aging for later EPR investigation; samples were capped with Critoseal and stored in a freezer until analysis. EPR analysis and EPFR characterization of soot samples were performed using previously described methods (Runberg et al., 2020). EPFR concentrations were measured using the spin-count feature available in the Bruker EMXnano Xenon software (see Table A1 for Q values and frequencies obtained). Spin counts were determined in triplicate for each sample and averaged. Lineshape evaluation, including g factors and Gaussian and Lorentzian line contributions, was performed using least squares analysis available in the EasySpin (easyspin.org) package available in MATLAB (Table A1). Triplicate soot samples were averaged together for final lineshape analysis.

2.3 Determination of [OH] via HPLC

Sodium benzoate (SBA) was used as an OH probe because it readily forms para-hydroxybenzoic acid (p-HBA) when exposed to OH radicals (Wu et al., 2017); p-HBA is relatively stable and can be easily quantified via high performance liquid

chromatography (HPLC). SBA (1.0 mM) was added to 45 ppm soot slurries just prior to beginning the photoreaction. This concentration was selected to be in substantial excess of the expected OH concentration, allowing the probe to capture the majority of the OH formed during the reaction. Aliquots of approximately 5 mL of soot slurry were collected before and after 16-hour photoreactions and filtered using 0.2μm, 25 mm diameter Whatman polypropylene filters. Filters were rinsed with a minimum of 15 mL ultrapure water prior to filtering. Filtrate was analyzed with no further modification via HPLC (Agilent 1100 with diode-array detector (DAD), and a C-18 Hydro-RP 250 × 4.6 mm column (Phenomenex)) using the method described previously (Runberg and Majestic, 2022). Each sample was run in triplicate and final p-HBA concentrations were averaged.

Fresh p-HBA standards were prepared in ultrapure water on the day of analysis, with a limit of quantification of 0.11 ± 0.06 μM, as determined by dividing the standard deviation of the response by the slope of the calibration curve and multiplying by 10 (International Committee on Harmonization, 2005). TotalOH production during the course of the reaction was determined by dividing the measured p-HBA concentration by the fraction of OH that forms p-HBA after 16 hours, previously determined to be 0.56 (Runberg and Majestic, 2022).

2.4 GC-MS Analysis

A solid-liquid organic extraction was done to obtain the organic-soluble fraction of soot for analysis via gas chromatography mass spectroscopy (GC-MS). Soot from each stage of the photo-aging process was extracted into acetonitrile (ACN) by submerging a small amount into an aliquot of ACN and allowing it to soak overnight. The ACN was then filtered using a 0.2 μm polypropylene Whatman filter to remove any remaining soot particulate before examination via an Agilent 7820A GC system coupled to a 5977B MSD. A previously described GC-MS method was used (Runberg and Majestic, 2022) and mass spectral analysis was preformed using Agilent MassHunter Qualitative Analysis Navigator (B.08.00) and NIST MS Search (version 2.3).

2.5 FT-IR

Fourier-transform infrared spectroscopy (FTIR) was done for all samples to assess any bulk changes to functional groups as the soot is aged. The system used was a D5 ATR Diamond, Thermo Scientific spectrograph. Analysis was performed by doing 30 scans per sample.

3 Results

3.1 EPFR Characterization

Figure 2 and Table A1 present the EPFR concentrations at various stages in the aging process. There was no significant difference in EPFR concentration at any phase of the soot aging process. Concentrations of EPFR ranged from (2.5 ± 0.3)

x 10^{16} spins mg^{-1} to (3.4 ± 0.6) x 10^{16} spins mg^{-1}. These concentrations that are consistent with previous soot and biochar studies (Runberg et al., 2020; Tian et al., 2009; Sigmund et al., 2021).

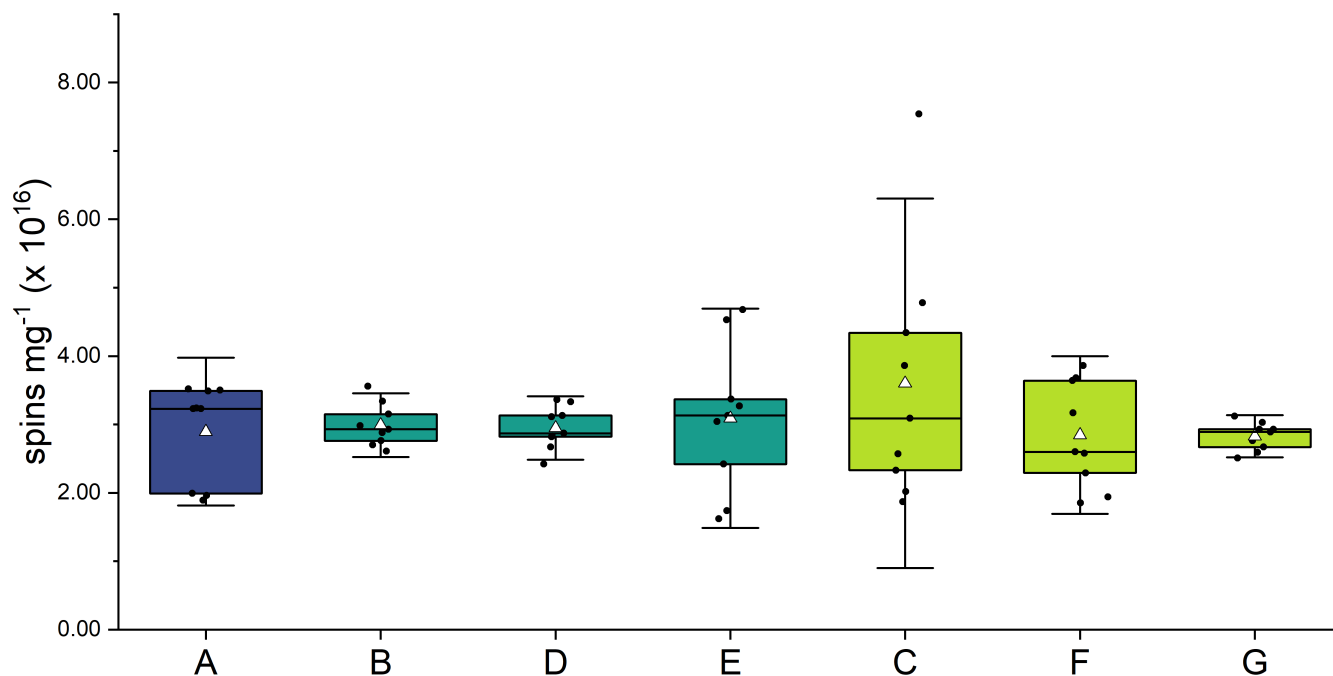



Figure 2 EPFR concentrations (spins per mg of soot) from soot samples at each step of the aging process (A-G). Three soot samples were collected from each phase of the experiment, and spins were calculated in triplicate for each sample. Average is shown as a white triangle.

One way to characterize an EPR signal is by determining the associated g-factor, a unitless proportionality constant determined by the ratio of the magnetic field and the applied microwave frequency. Equation 1 shows the relationship of g to the frequency (ν) and the magnetic field (B), where h is Planck's constant and β is the Bohr magnetron (Eaton et al., 2010), see Eq. (1):

$$g = \frac{h\upsilon}{\beta B},$$ (1)

Electrons are half-spin particles with two possible spin states, +½ and -½, with the difference in energy between them equal to ΔE. In EPR spectroscopy, the microwave frequency (ν) is usually held constant while the magnetic field (B) is modulated. When ΔE equals the applied microwave energy, resonance is achieved and absorption occurs, resulting in an EPR signal. The g-factor at the point of resonance can be used to identify the type of radical being measured (Eaton et al., 2010; EPR-Interpretation, 2022).

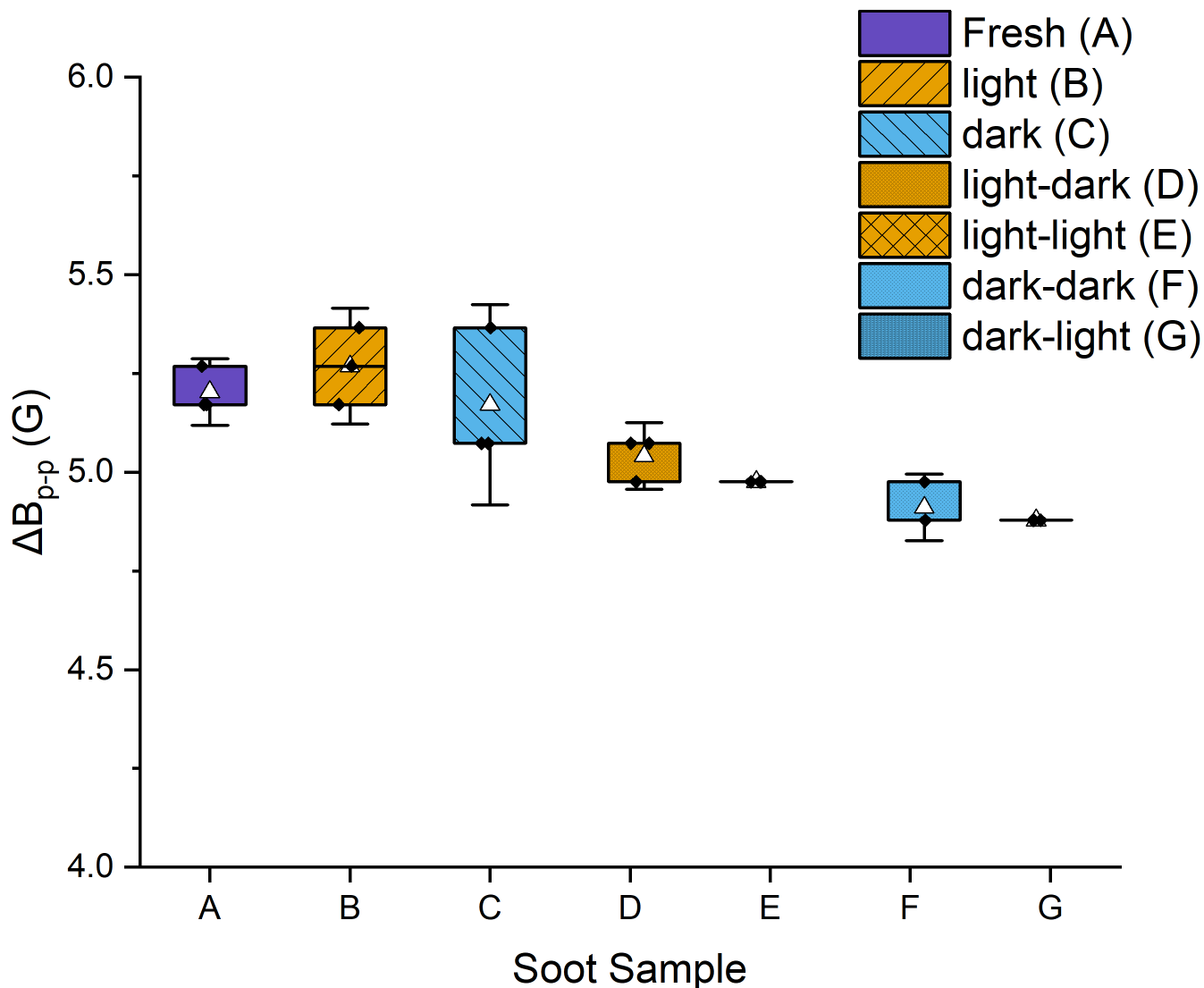
In the current study, there were no changes observed to the g-factor of the soot over the course of the aging process, which was between 2.0056-2.0057 for all samples (Table A1 and Figure A1) and is consistent with the g-factor of an organic radical (EPR-Interpretation, 2022). Because of the lack of hyperfine splitting in the EPR signal, it can be deduced that the soot is

comprised of a large number of organic radicals which cannot be uniquely defined. Changes in g-factor indicate a change in radical type, for example, a shift from predominantly carbon-centered to predominantly oxygen-centered radicals, which was

140   not observed in this study. However, this g-factor is notably higher than those reported in our previous study (Runberg et al., 2020), which were 2.0028 ± 0.0001. The soot for both studies was obtained in an identical manner, but the current study appears to have a much higher concentration of oxygen- or nitrogen-centered radicals. This is likely due to the use of a diffusion flame, which cannot control the air-to-fuel ratio, and oxygen or nitrogen were likely incorporated during combustion. While it is outside the scope of this study, it would be useful to control for this factor in future investigations.

145   A second parameter that can be used to characterize an EPFR signal is its peak width ($\Delta B_{p\text{-}p}$), measured as the distance along the x-axis (magnetic field, B) between the maximum and minimum absorbances (peaks) of the derivative EPR signal. Both the peak width and the line shape (discussed below) provide information about how the varying spin environments in a complex sample interact with one another. Changes to $\Delta B_{p\text{-}p}$ indicate changes to the composition of the bulk sample (Sorin and Vlasova, 1973).

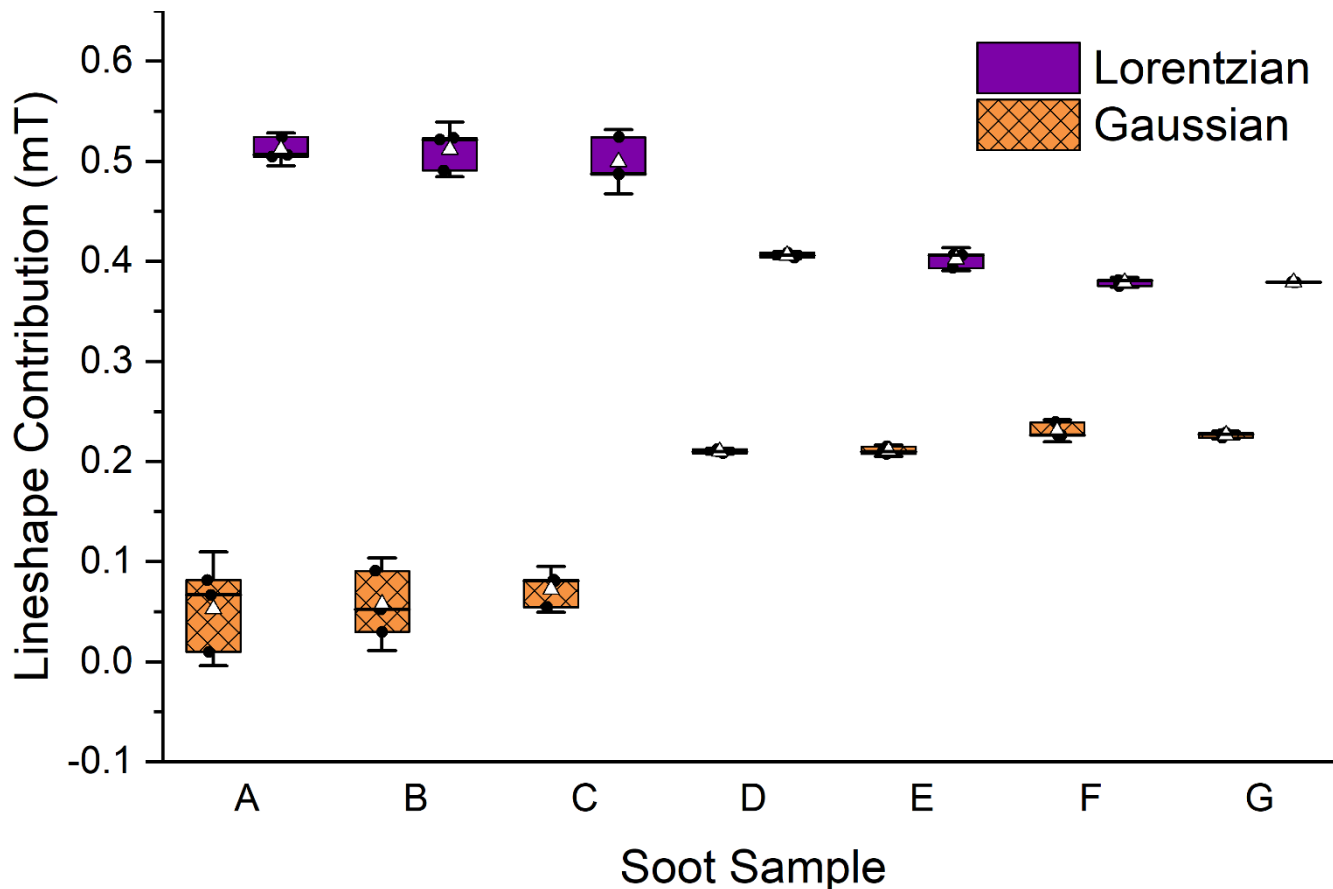
150   There was no difference in peak width between any of the dry samples (fresh, photo-aged, and dark-aged), indicating no impact to the spin environments from the photo-aging of dry soot. However, there were small but statistically significant differences between the dry and the wetted samples (t-tests done at 95% certainty, with $p < 0.05$) (Figure 3). The $\Delta B_{p\text{-}p}$ decreased for both aged soot samples (dark and light) upon submersion and agitation in water. For soot that had been photo-aged dry, $\Delta B_{p\text{-}p}$ decreased 4.4% in the dark and 5.7% when exposed to light. For soot that had been dark-aged dry, $\Delta B_{p\text{-}p}$ decreased 5.2% in the

155   dark and 5.8% when exposed to light. Additionally, when comparing the peak width of the wetted soot samples, soot that had been kept in the dark for dry-aging had $\Delta B_{p\text{-}p}$ 2.3% smaller than soot that had been dry-aged in light. This decrease in $\Delta B_{p\text{-}p}$ is likely an artifact of the increase in Gaussian character of the spectral lines for samples that had been submerged and agitated in water. This is discussed further in the next section.

160 **Figure 3 Box plot of peak-to-peak widths (ΔB_{p-p}) for each of the soot stages. Error bars are the standard deviation of three trials. The average is shown as a white triangle. Samples A, B, and C were dry soot, samples D-G had been wetted and then dried prior to analysis.**

Lineshape is another way to characterize an EPFR signal. EPR lines are complex functions comprised of the signals from each unpaired electron in the bulk sample. The bulk signal is modelled as a summation of each individual signal, using Gaussian and Lorentzian line functions. The more homogenous the sample, the more Lorentzian in character the lineshape will be. A more complex sample will result in more hyperfine coupling, which results in an increase in Gaussian character (Petrakis, 1967).

In this study, EPR lineshapes were seen to change in the wet portion of the aging process, with a significant decrease in Lorentzian character and an increase in Gaussian character (Figure 4) indicating a decrease in the homogeneity of the EPFR

170   environments in the soot and an increase in hyperfine interactions. Gaussian contribution increased from an average of (0.06 ± 0.03) mT across stages I and II to (0.22 ± 0.01) mT in stage III, an increase of 133%. Lorentzian contribution decreased 17% from a combined average of (0.51 ± 0.02) mT for dry soot to (0.40 ± 0.01) mT wetted soot. This is an indication that the variability of electron environments within the sample increased with exposure to the water, but not to the light. This is counter to our previous study which did not result in any lineshape changes due to exposure to water. However, in that study, the soot

175   was submerged and not agitated. So, the lineshape change seen here is likely due to the continuous agitation that occurred during the photoreaction, which broke up large soot fragments. An increase in soot density was observed between dry soot, and soot that had been wetted (see Table A1), supporting this assessment. Soot density was determined by dividing the mass of the soot by the height of the soot in the capillary tube.

180   **Figure 4 Changes in Gaussian and Lorentzian contributions to the EPR lineshape. Error bars are the standard deviation of three trials. The average is shown as a white triangle. Samples A, B, and C were dry soot, samples D-G had been wetted and then dried prior to analysis.**

3.2 Hydroxyl radical (OH) Formation

The ability of soot to produce OH was measured in aqueous conditions to mimic the conditions in a droplet of cloud water.
Analysis was separated by soot that had been previously photo-aged and soot that had been previously aged in the dark. No measurable OH was observed in either of the dark reactions (Figure 1, samples D and F), confirming that light is required for OH production. For soot slurries that were exposed to light, soot that had been photo-aged for 24 hours previously had the lowest OH production at 0.38 ± 0.02 μM. This is not significantly different ($p = 0.06$) than controls done in ultrapure water, which were (0.52 ± 0.06) μM. Soot that had been dark-aged for 24 hours prior to the photoreaction in water resulted in significantly higher OH production, at (0.89 ± 0.05) μM (Figure 5) ($p < 0.005$). This is consistent with a recent study reporting a decrease in OH formation in wildfire smoke plumes downfield from the point source, vs the same plume measured nearer the source (Akherati et al., 2022). This indicates that smoke which has been exposed to sunlight for a longer period of time (i.e. the downfield plume) results in lower OH production. While photoaging of airborne soot particles in a much more complex system such as a smoke plume is likely not the only reason for the decline, (e.g. there are multiple OH production pathways and the plume is diluted), the same pattern of diminished OH production is reported. A laboratory study showed reduced ROS activity in aged PM as well (Gehling et al., 2014). In that study, particulate matter was aged at room temperature in phosphate buffered saline solution to maintain a physiologically neutral pH, but no application of light was executed. The authors reported a decrease OH formation of about 11% after one day. This implies that other factors may play a role in the aging of soot, in addition to sunlight.

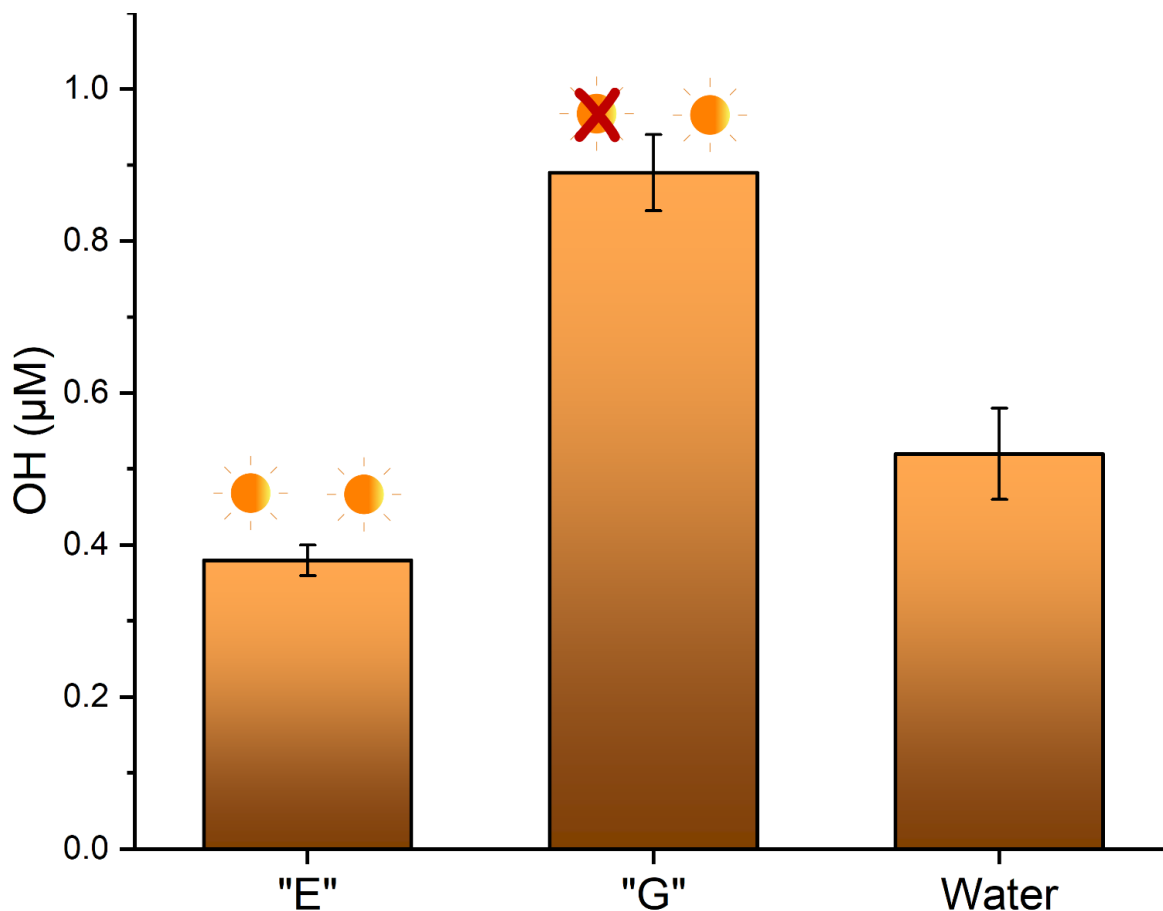



Figure 5 Production of OH measured after 16-hour lighted reactions of soot that had been previously aged for 24 hours in the light (E), aged for 24 hours in the dark (G) and an ultrapure water control.

3.3 GC-MS

GC-MS analysis was done to qualitatively determine if there was a change in the number of organic-soluble compounds in the soot at different stages of the aging process. In all dry soot samples (samples A, B, and C in Figure 1), only one compound was seen, at a retention time of 9.13 min. (Figure 6, top row). Analysis of the fragmentation pattern of this peak indicated no obvious nitrogen- or oxygen-containing fragments. For all samples that had been submerged in water (Samples D – G in Figure 1), there were several new peaks observed (Figure 6, middle and bottom rows). Mass spectral comparisons to the NIST MS database did not definitively identify any of the compounds, but all display a distinctly hydrocarbon-like fragmentation pattern (Figure A2).

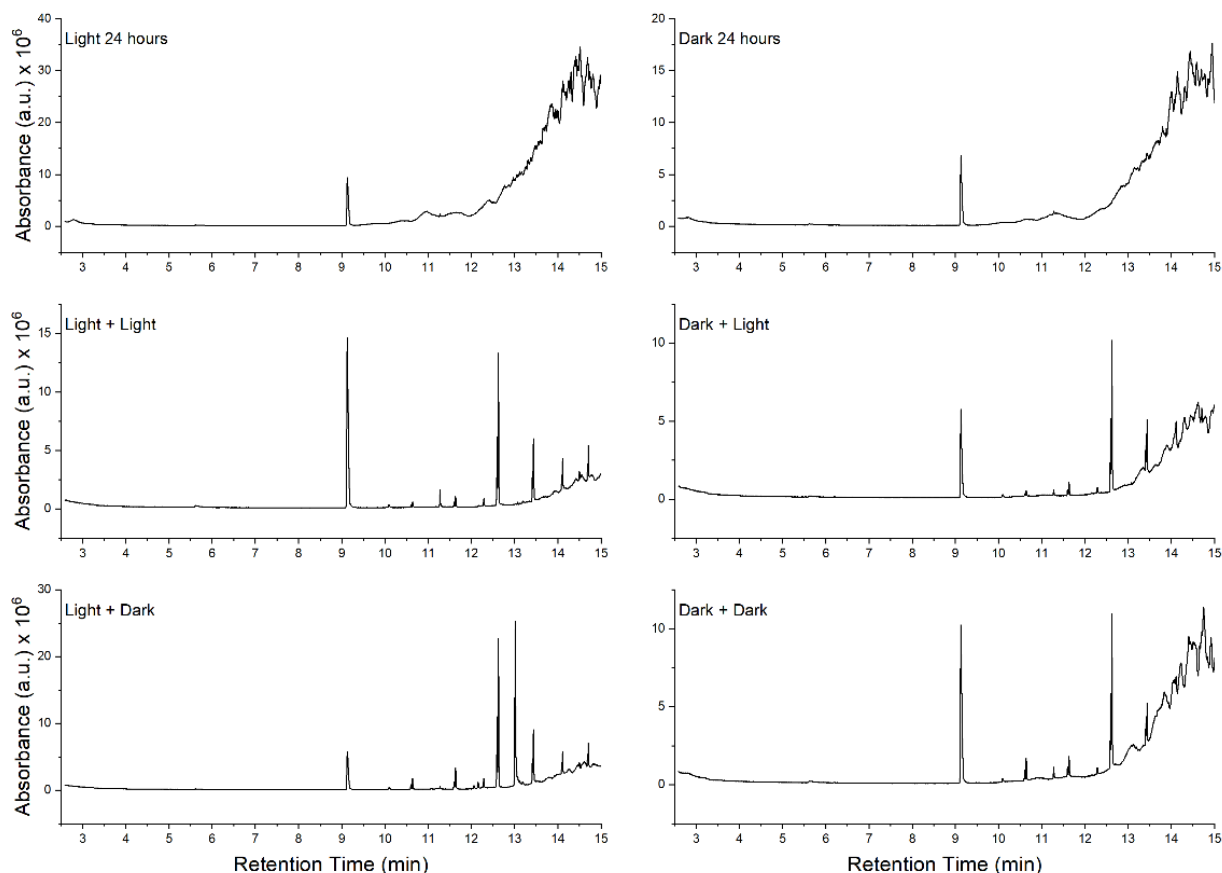
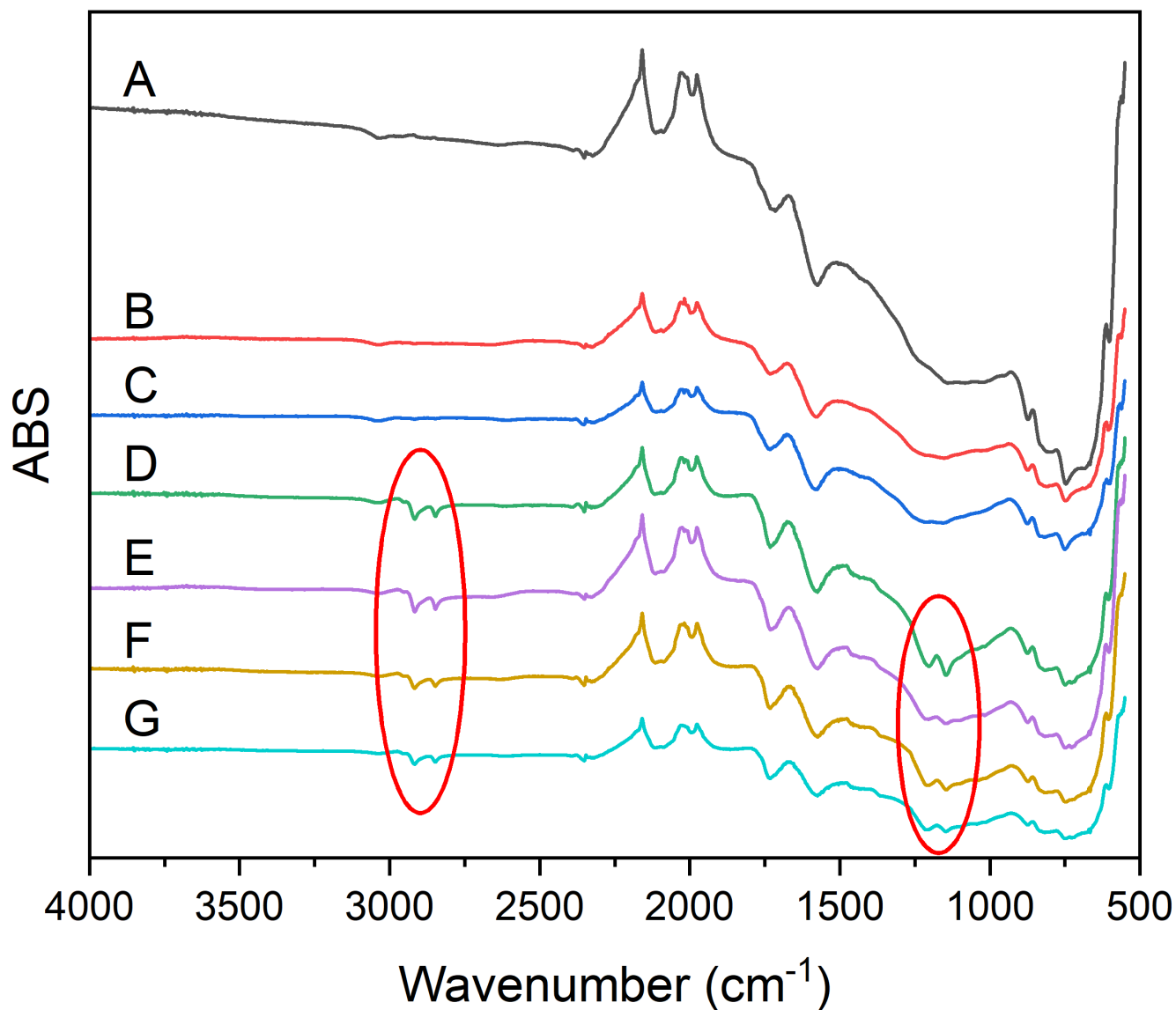



Figure 6 Gas chromatograms of soot at each step of the aging process. The left column contains chromatograms of soot extracts taken from the photo-aged pathway (samples B → E & D), and the right column contains extracts of soot that followed the dark-aged pathway (samples C → G & F). The first row is extracts from the soot after initial 24 hours of aging. The second row is soot that was photo-aged in water for 16 hours, and the bottom row is soot that was dark aged in water for 16 hours.

There is indication that oxygen is being incorporated into the new peaks observed in samples that had been wetted. For example, analysis of the fragmentation pattern on the peak which eluted at 13.44 min. suggests a McLafferty rearrangement (Burgers and Terlouw, 2016), resulting in an acylium cation with m/z = 43 and a McLafferty fragment at m/z = 57, implying the presence of at least one carbonyl group (Figure A3).

3.4 Fourier transform infrared (FTIR) spectroscopy

Fourier transform infrared (FTIR) spectroscopy was done for all soot samples to assess bulk changes in functional groups. All dry soot samples appear nearly identical, indicating no changes to functionality via dry aging. There is new peak formation in all samples that had been wetted, indicating that the change is likely due to the prolonged agitation in water rather than due to photoreactions of the soot or with any OH formed during the reaction. New peaks were observed at about 2915 cm^{-1} and 2850 cm^{-1}, consistent with an increase in C—H stretch, and at 1207 cm^{-1} and 1140 cm^{-1}, which indicate a C—O stretch (Figure 7).

Additionally, several peaks are observed in all samples: a peak at 1575 cm^{-1} suggests the presence of N—O stretching, a peak at 1730 cm^{-1} may be C=N, and a peak at 2090 cm^{-1} may be C=C=N, offering evidence of the nitro-components suggested by the g-factor of this soot, and was likely introduced during the combustion process.

230 **Figure 7 FTIR spectra of all soot samples. Fresh soot (A), dry soot after 24 hours of illumination (B), dry soot after 24 hours in the dark (C), soot that was photo-aged for 24 hours, followed by 16 hours in the dark, in water (D), soot that was photo-aged for 24 hours, followed by 16 hours in the light, in water €, soot that was dark-aged for 24 hours, followed by 16 hours in the dark, in water (F), soot that was dark-aged for 24 hours, followed by 16 hours in the light in water (G). Red circles indicate new peaks associated with incorporation of oxygen.**

4 Conclusions

EPFR characteristics of combustion particles, such as g-factor, line shape, peak width, and spin concentration, are unaffected by photoaging, however, soot's ability to generate OH when exposed to sunlight diminishes drastically after light exposure. This implies several possibilities: First, it may be something other than EPFR within the particles that drives the OH formation, or, second, that only EPFR on the surface of the particles are available for OH formation, while the bulk of the EPFR component is bound within the particles where they are not impacted by the light. Combustion particles that were not photoaged produced almost 60% more OH than photoaged soot did. It is possible that any changes to the surface EPFR due to exposure to light was muted by the larger concentration of EPFR found within the particles, resulting in no meaningful change to the EPFR concentration or characteristics of the exposed soot. The lack of changes to chemical composition of the soot upon irradiation, as seen in the GC-MS and UV-Vis analysis, support the hypothesis that only EPFR on the surface of the particle are available for OH formation.

Other environmental factors - specifically the chemical composition of natural cloud waters - may impact the way that combustion particles age in the environment, including some properties of EPFR, and future work should include this variable. Natural cloud waters are a complex matrix containing compounds such as sulfates, nitrates, metals, and organic matter. Previous work has indicated that natural cloud water has a much lower pH than pure water, and the lower pH of natural cloud water has been implicated in increased solubility of organic aerosols (Li et al., 2020), which may influence photo induced OH formation from soot.

In summary, EPFR characteristics of soot do not change upon photoaging, but the soot's ability to generate OH is greatly reduced. This has implications to human health, because it implies that freshly generated soot, as from a wildfire, is potentially more hazardous upon inhalation than is soot that has travelled further away from its point source. This could result in an increase in negative health outcomes for people living close to frequent wildfires. Due to global climate change, the frequency and intensity of wildfires is expected to increase, which means that more people will become exposed to freshly generated soot particles. Future studies should investigate the differences in EPFR characteristics and OH production using multiple fuel sources, including biofuels such as seen in wildfires.

Appendix A

Table A1 EPR data for each soot sample collected. Samples were collected in triplicate from each batch.

Batch	Q Value	Freq (GHz)	g factor	Gaussian (mT)	Lorentzian (mT)	Soot density (mg mm^{-1})	spins mg^{-1}		Freezer Storage (days)
A: Fresh Soot	4105	9.659	2.0057	0.0817	0.5066	0.052	3.24E+16	± 1.26E+14	7
	4104	9.658	2.0055	0.0668	0.5046	0.056	3.50E+16	± 3.48E+14	7

Sample								
	4105	9.659	2.0056	0.0099	0.5244	0.104	1.95E+16	± 1.58E+15	7
B: Dry photo-aged soot	4105	9.659	2.0055	0.0523	0.5232	0.076	2.93E+16	± 2.00E+14	6
	3518	9.658	2.0056	0.0907	0.4908	0.086	3.16E+16	± 2.65E+14	6
	4105	9.659	2.0056	0.0299	0.5216	0.081	2.85E+16	± 4.16E+15	6
C: Dry dark-aged soot	4104	9.659	2.0057	0.0818	0.4877	0.108	2.57E+16	± 2.89E+14	6
	4104	9.657	2.0057	0.0808	0.4867	0.100	3.11E+16	± 1.52E+16	6
	3518	9.659	2.0056	0.0548	0.5242	0.067	4.36E+16	± 1.72E+15	6
F: Dark-aged soot, dark OH	4105	9.660	2.0056	0.2084	0.4089	0.161	3.13E+16	± 7.81E+14	2
	4105	9.660	2.0055	0.2123	0.4059	0.180	2.87E+16	± 2.08E+14	2
	3519	9.660	2.0056	0.2100	0.4039	0.178	2.82E+16	± 1.13E+15	2
G: Dark-aged soot, light OH	4105	9.659	2.0056	0.2152	0.3934	0.200	2.42E+16	± 2.00E+14	2
	4105	9.660	2.0056	0.2101	0.4062	0.167	3.37E+16	± 1.15E+15	2
	4105	9.660	2.0056	0.2078	0.4073	0.180	3.03E+16	± 1.65E+15	2
D: Photo-aged soot, dark OH	4105	9.659	2.0055	0.2393	0.3754	0.220	2.58E+16	± 7.21E+14	2
	3518	9.659	2.0056	0.2266	0.3809	0.236	2.29E+16	± 5.96E+15	2
	4105	9.659	2.0056	0.2262	0.3813	0.244	2.72E+16	± 1.08E+15	2
E: Photo-aged soot, light OH	3519	9.661	2.0056	0.2238	0.3790	0.193	2.93E+16	± 5.48E+15	2
	3518	9.660	2.0056	0.2288	0.3792	0.207	2.92E+16	± 3.46E+14	2
	3518	9.658	2.0056	0.2274	0.3794	0.227	2.61E+16	± 7.64E+15	2

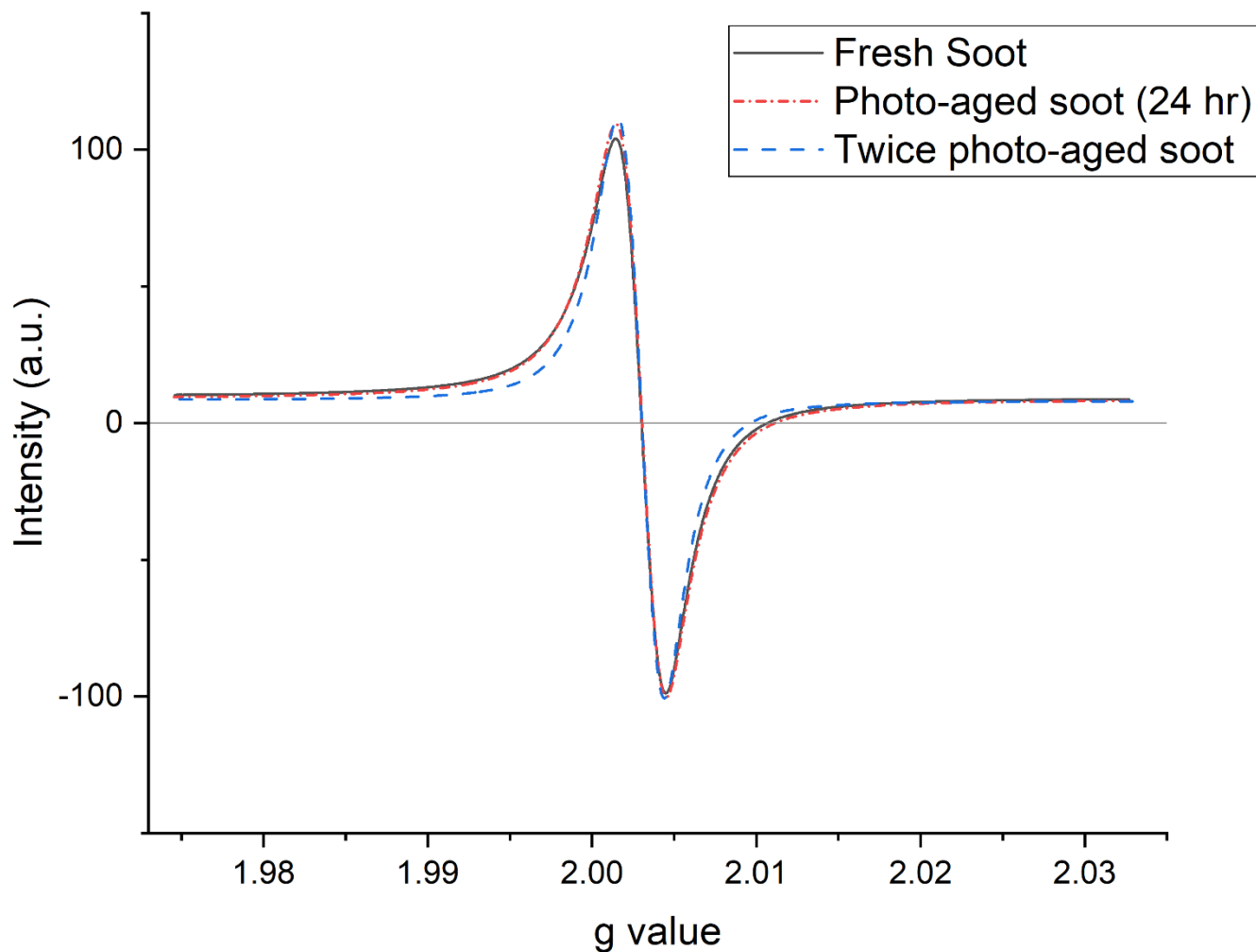



Figure A1 Representative EPR spectra of three soot samples at different stages of photo-aging, Fresh (black, solid line), soot aged for 24 hours under simulated sunlight (red dash-dot), and soot photo-aged twice, once dry, and once wet (blue, dash-dash). Signals have been baseline corrected and intensity normalized.

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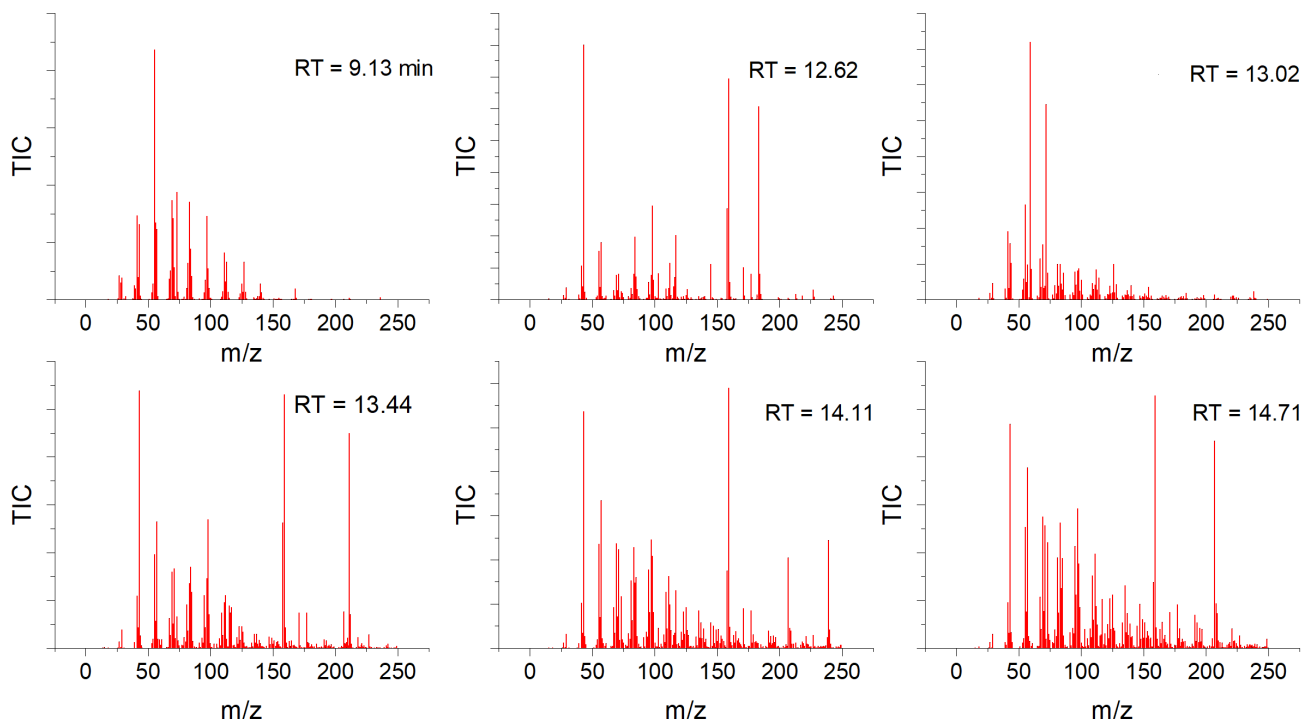
Figure A2 Mass Spectra of select peaks from the GC obtained from "Light->Dark" soot extracted into acetonitrile. Retention times refer to those seen in Figure 6.
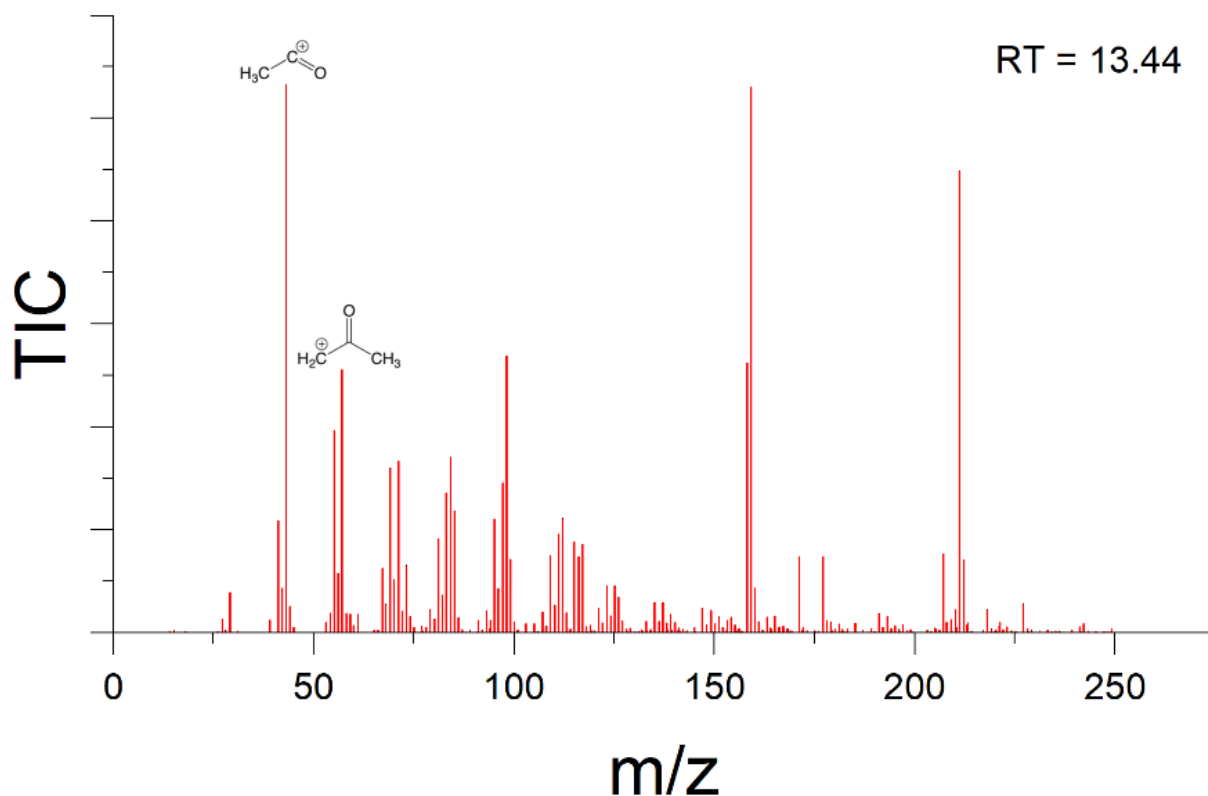
Figure A3 Mass spectrum of the peak seen at retention time 13.44 minutes.

Data Availability

Data described in this manuscript can be accessed at repository under data doi

275 (https://doi.org/10.56902/Chem.Data.2023.1)(Majestic and Runberg, 2023)(Majestic and Runberg, 2023)(Majestic and Runberg, 2023)

Author Contribution

Heather L. Runberg: Methodology, Software, Validation, Investigation, Writing - Original Draft, Data Curation, Visualization

Brian J. Majestic: Conceptualization, Resources, Writing - Review & Editing, Supervision, Project administration

280 **Competing interests**

The authors declare that they have no conflict of interest.

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285 ## References

Aguilera, R., Corringham, T., Gershunov, A., and Benmarhnia, T.: Wildfire smoke impacts respiratory health more than fine particles from other sources: observational evidence from Southern California, Nat Commun, 12, 1493, https://doi.org/10.1038/s41467-021-21708-0, 2021.

290 Akherati, A., He, Y., Garofalo, L. A., Hodshire, A. L., Farmer, D. K., Kreidenweis, S. M., Permar, W., Hu, L., Fischer, E. V., Jen, C. N., Goldstein, A. H., Levin, E. J. T., DeMott, P. J., Campos, T. L., Flocke, F., Reeves, J. M., Toohey, D. W., Pierce, J. R., and Jathar, S. H.: Dilution and photooxidation driven processes explain the evolution of organic aerosol in wildfire plumes, Environmental Science: Atmospheres, 2, 1000–1022, https://doi.org/10.1039/D1EA00082A, 2022.

Arangio, A. M., Tong, H., Socorro, J., Pöschl, U., and Shiraiwa, M.: Quantification of environmentally persistent free radicals

295 and reactive oxygen species in atmospheric aerosol particles, Atmos Chem Phys, 16, 13105–13119, https://doi.org/10.5194/acp-16-13105-2016, 2016.

Brook, R. D., Xu, X., Bard, R. L., Dvonch, J. T., Morishita, M., Kaciroti, N., Sun, Q., Harkema, J., and Rajagopalan, S.: Reduced metabolic insulin sensitivity following sub-acute exposures to low levels of ambient fine particulate matter air pollution, Science of the Total Environment, 448, 66–71, https://doi.org/10.1016/j.scitotenv.2012.07.034, 2012.

300 Brook, R. D., Cakmak, S., Turner, M. C., Brook, J. R., Crouse, D. L., Peters, P. A., Van Donkelaar, A., Villeneuve, P. J., Brion, O., Jerrett, M., Martin, R. V, Rajagopalan, S., Goldberg, M. S., Arden, C., and Burnett, R. T.: Long-Term Fine Particulate Matter Exposure and Mortality From Diabetes in Canada, Diabetes Care, 36, 3313–3320, https://doi.org/10.2337/dc12-2189, 2013.

Burgers, P. C. and Terlouw, J. K.: Ion structures in mass spectrometry, in: Encyclopedia of Spectroscopy and Spectrometry,

305 Elsevier, 322–329, https://doi.org/10.1016/B978-0-12-803224-4.00182-5, 2016.

Chan, C.-C., Chuang, K.-J., Chen, W.-J., Chang, W.-T., Lee, C.-T., and Peng, C.-M.: Increasing cardiopulmonary emergency visits by long-range transported Asian dust storms in Taiwan, Environ Res, 106, 393–400, https://doi.org/10.1016/j.envres.2007.09.006, 2008.

Chen, Q., Wang, M., Wang, Y., Zhang, L., Xue, J., Sun, H., and Mu, Z.: Rapid determination of environmentally persistent

310 free radicals (EPFRs) in atmospheric particles with a quartz sheet-based approach using electron paramagnetic resonance (EPR) spectroscopy, Atmos Environ, 184, 140–145, https://doi.org/10.1016/j.atmosenv.2018.04.046, 2018.

Coop, J. D., Parks, S. A., Stevens-Rumann, C. S., Ritter, S. M., Hoffman, C. M., and Varner, J. M.: Extreme fire spread events and area burned under recent and future climate in the western USA, Global Ecology and Biogeography, 00, 1–11, https://doi.org/10.1111/GEB.13496, 2022.

315 Donahue, N. M., Robinson, A. L., and Pandis, S. N.: Atmospheric organic particulate matter: From smoke to secondary organic aerosol, Atmospheric Environment journal, 43, 94–106, https://doi.org/10.1016/j.atmosenv.2008.09.055, 2009.

Eaton, G. R., Eaton, S. S., Barr, D. P., and Weber, R. T.: Quantitative EPR, Springer-Verlag/Wien, https://doi.org/10.1007/978-3-211-92948-3, 2010.

Feng, S., Gao, D., Liao, F., Zhou, F., and Wang, X.: The health effects of ambient PM 2.5 and potential mechanisms, Ecotoxicol
320 Environ Saf, 128, 67–74, https://doi.org/10.1016/j.ecoenv.2016.01.030, 2016.

Gehling, W. and Dellinger, B.: Environmentally persistent free radicals and their lifetimes in PM 2.5, Environ Sci Technol, 47, 8172–8178, https://doi.org/10.1021/es401767m, 2013.

Gehling, W., Khachatryan, L., and Dellinger, B.: Hydroxyl Radical Generation from Environmentally Persistent Free Radicals (EPFRs) in PM 2.5, Environ. Sci. Technol, 48, https://doi.org/10.1021/es401770y, 2014.

325 Guo, C., Chan, T.-C., Teng, Y.-C., Lin, C., Bo, Y., Chang, L.-Y., Lau, A. K. H., Tam, T., Wong, M. C. S., and Lao, Q.: Long-term exposure to ambient fine particles and gastrointestinal cancer mortality in Taiwan: A cohort study, Environ Int, 138, 105640, https://doi.org/10.1016/j.envint.2020.105640, 2020.

He, L., Norris, C., Cui, X., Li, Z., Barkjohn, K. K., Teng, Y., Fang, L., Lin, L., Wang, Q., Zhou, X., Hong, J., Li, F., Zhang, Y., Schauer, J. J., Black, M., Bergin, M. H., and Zhang, J.: Oral cavity response to air pollutant exposure and association with
330 pulmonary inflammation and symptoms in asthmatic children, Environ Res, 206, 112275, https://doi.org/10.1016/j.envres.2021.112275, 2022.

International Committee on Harmonization: Validation of Analytical Procedures: Text and Methodology, Q2(R1), in: ICH Harmonized Tripartite Guideline, 2005.

Katanoda, K., Sobue, T., Satoh, H., Tajima, K., Suzuki, T., Nakatsuka, H., Takezaki, T., Nakayama, T., Nitta, H., Tanabe, K.,
335 and Tominaga, S.: An Association Between Long-Term Exposure to Ambient Air Pollution and Mortality From Lung Cancer and Respiratory Diseases in Japan, J Epidemiol, 21, 132–143, https://doi.org/10.2188/jea.JE20100098, 2011.

Khachatryan, L., Vejerano, E., Lomnicki, S., and Dellinger, B.: Environmentally persistent free radicals (EPFRs). 1. Generation of reactive oxygen species in aqueous solutions, Environ Sci Technol, 45, 8559–8566, https://doi.org/10.1021/es201309c, 2011.

340 Leresche, F., Salazar, J. R., Pfotenhauer, D. J., Hannigan, M. P., Majestic, B. J., and Rosario-ortiz, F. L.: Photochemical Aging of Atmospheric Particulate Matter in the Aqueous Phase, Environ Sci Technol, https://doi.org/10.1021/acs.est.1c00978, 2021.

Li, T., Wang, Z., Wang, Y., Wu, C., Liang, Y., Xia, M., Yu, C., Yun, H., Wang, W., Wang, Y., Guo, J., Herrmann, H., and Wang, T.: Chemical characteristics of cloud water and the impacts on aerosol properties at a subtropical mountain site in Hong Kong SAR, Atmos Chem Phys, 20, 391–407, https://doi.org/10.5194/acp-20-391-2020, 2020.

345   Liu, Y. and Chan, C. K.: The oxidative potential of fresh and aged elemental carbon-containing airborne particles: a review †, Environ Sci Process Impacts, 24, 252, https://doi.org/10.1039/d1em00497b, 2022.

Majestic, B. and Runberg, H. L.: Data Set for "Atmospheric Aging of Combustion-Derived Particles: Impact on Stable Free Radical Concentration and Its Ability to Produce Reactive Oxygen Species in Aqueous Media," https://doi.org/https://doi.org/10.56902/Chem.Data.2023.1, 17 January 2023.

350   Petrakis, L.: Spectral Line Shapes Gaussian and Lorentzian functions in magnetic resonance, J Chem Educ, 44, 432, 1967.

Runberg, H. L. and Majestic, B. J.: Hydroxyl radical (OH) formation during the photooxidation of anthracene and its oxidized derivatives, Atmos Environ, 286, 119214, https://doi.org/10.1016/J.ATMOSENV.2022.119214, 2022.

Runberg, H. L., Mitchell, D. G., Eaton, S. S., Eaton, G. R., and Majestic, B. J.: Stability of environmentally persistent free radicals (EPFR) in atmospheric particulate matter and combustion particles, Atmos Environ, 240, 117809,

355   https://doi.org/10.1016/j.atmosenv.2020.117809, 2020.

Sigmund, G., Santín, C., Pignitter, M., Tepe, N., Doerr, S. H., and Hofmann, T.: Environmentally persistent free radicals are ubiquitous in wildfire charcoals and remain stable for years, Commun Earth Environ, 2, https://doi.org/10.1038/s43247-021-00138-2, 2021.

Sorin, L. A. and Vlasova, M. V.: EPR Line Shapes and Line Widths, in: Electron Spin Resonance of Paramagnetic Crystals,

360   Springer, Boston, 111–124, 1973.

Tian, L., Koshland, C. P., Yano, J., Yachandra, V. K., Yu, I. T. S., Lee, S. C., and Lucas, D.: Carbon-Centered Free Radicals in Particulate Matter Emissions from Wood and Coal Combustion, Energy & Fuels, 23, 2523–2526, https://doi.org/10.1021/ef8010096, 2009.

Tong, H., Lakey, P. S. J., Arangio, A. M., Socorro, J., Shen, F., Lucas, K., Brune, W. H., Pöschl, U., and Shiraiwa, M.: Reactive

365   Oxygen Species Formed by Secondary Organic Aerosols in Water and Surrogate Lung Fluid, Environ Sci Technol, 52, 11642–11651, https://doi.org/10.1021/acs.est.8b03695, 2018.

Vidrio, E., Phuah, C. H., Dillner, A. M., and Anastasio, C.: Generation of hydroxyl radicals from ambient fine particles in a surrogate lung fluid solution, Environ Sci Technol, 43, 922–927, https://doi.org/10.1021/es801653u, 2009.

Wang, Y., Li, S., Wang, M., Sun, H., Mu, Z., Zhang, L., Li, Y., and Chen, Q.: Source apportionment of environmentally

370   persistent free radicals (EPFRs) in PM2.5 over Xi'an, China, Science of the Total Environment, 689, 193–202, https://doi.org/10.1016/j.scitotenv.2019.06.424, 2019.

Wu, C., De Visscher, A., and Gates, I. D.: Reactions of hydroxyl radicals with benzoic acid and benzoate, RSC Adv, 7, 35776–35785, https://doi.org/10.1039/C7RA05488B, 2017.

EPR-Interpretation: https://chem.libretexts.org/@go/page/1792, last access: 2 May 2022.

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