

1 Substantially positive contributions of new particle formation to Cloud Condensation  
2 Nuclei under low supersaturation in China based on numerical model improvements

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## Abstract

New particle formation (NPF) and subsequent particle growth are important sources of condensation nuclei (CN) and cloud condensation nuclei (CCN). While many observations have shown positive contributions of NPF to CCN at low supersaturation, negative NPF contributions were often simulated in polluted environments. Using the observations in a coastal city of Qingdao, Beijing and Gucheng in North China, we thoroughly evaluate the simulated number concentrations of CN and CCN using a NPF-explicit parameterization embedded in WRF-Chem model. For CN, the initial simulation shows large biases of particle number concentrations at 10–40 nm and 40–100 nm. By adjusting the process of gas-particle partitioning, including the mass accommodation coefficient of sulfuric acid, the phase changes of primary organic aerosol emissions and the condensational amount of nitric acid, the improvement of the particle growth process yields a substantially reduced overestimates of CN. Regarding CCN, SOA formed from the oxidation of semi-volatile and intermediate volatility organic vapors (SI-SOA) yield is an important contributor. At default settings, the SI-SOA yield is too high without considering the differences in precursor oxidation rates. Lowering the SI-SOA yield under linear- $\text{H}_2\text{SO}_4$  nucleation scheme results in much improved CCN simulations compared to observations. On the basis of the bias-corrected model, we find substantially positive contributions of NPF to CCN at low supersaturation ( $\sim 0.2\%$ ) over the broad areas of China, primarily due to competing effects of increasing particle hygroscopicity, a result of reductions in SI-SOA amount, surpassing that of particle size decreases. The bias-corrected model is robustly applicable to other schemes, such as quadratic- $\text{H}_2\text{SO}_4$  nucleation scheme, in terms of CN and CCN, though the dependence of CCN on SI-SOA yield is diminished likely due to changes in particle composition. This study highlights the potentially much larger NPF contributions to CCN on a regional and even global basis.

## 66 **1. Introduction**

67 New particle formation (NPF) is a process in which gaseous vapors nucleate and  
68 form critical molecular clusters, followed by subsequent growth to larger sizes through  
69 condensation and coagulation (Kulmala et al., 2004; Kulmala et al., 2013; Lee et al.,  
70 2019). Newly formed particles could effectively grow into the size of cloud  
71 condensation nuclei (CCN) under certain supersaturation (SS), which exerts an impact  
72 on the cloud microphysical process and global radiation balance (Merikanto et al., 2009;  
73 Gordon et al., 2017; Kerminen et al., 2018; Ren et al., 2021). In addition, the efficient  
74 nucleation and explosive growth of particles may contribute to the formation of haze  
75 (Guo et al., 2014), affecting air quality and human health (Yuan et al., 2015; Chu et al.,  
76 2019; Kulmala et al., 2021).

77 The overestimate of condensation nuclei (CN) in numerical models is commonly  
78 seen, despite the attempt to rectify the bias (Matsui et al., 2013; Arghavani et al., 2022).  
79 It is a common way to reduce the nucleation rate which may reduce the particle number  
80 concentration in proportion (Matsui et al., 2013). For instance, in the study of NPF in  
81 East Asia in the spring of 2009, even after lowering the nucleation rate in a regional  
82 model of WRF-Chem applied in their study, the reduced number concentration of  
83 particles at 10–130 nm remained to be overestimated (Matsui et al., 2013). Using the  
84 same regional model and a similar method to reduce the nucleation rate, Arghavani et  
85 al. (2022) found particle number concentration at 10–100 nm was still overestimated  
86 by nearly one order of magnitude, despite the effectiveness to reduce the overestimates  
87 for the smaller particles such as 2.5–10 nm. In addition to the rate of NPF, the growth  
88 process of particles also has a crucial effect on particle number concentration and size  
89 distribution. In this process, the condensation of some chemical species such as sulfuric  
90 acid, nitrate and organic gases on particles plays a major role in particle growth (Yao et  
91 al., 2018; Lee et al., 2019; Li et al., 2022), and the uncertainty of their condensation  
92 amount may lead to the bias of CN simulation.

93 In addition to CN, there are large discrepancies in the predicted CCN between the  
94 numerical models and observational results. Furthermore, as an important source of  
95 CCN (Merikanto et al., 2009), the contribution of nucleation to CCN quantified by

96 numerical models is also highly uncertain. For example, in terms of predicting CCN,  
97 Fanourgakis et al. (2019) evaluated the CCN concentrations simulated by 16 global  
98 aerosol–climate and chemistry transport models with observations at 9 sites in Europe  
99 and Japan from 2011 to 2015, and found that all models underestimated CCN  
100 concentrations with a mean normalized mean bias of -36% at low supersaturation  
101 ( $SS=0.2\%$ ). WRF-Chem models also tend to underestimate the contribution of NPF on  
102 CCN, especially at low supersaturation. The continuous observation of CCN  
103 concentrations throughout the year (July 2008–June 2009) carried out in Hyytiälä,  
104 Finland, showed that under low SS, nucleation enhanced the CCN by 106% and 110%  
105 at  $SS=0.1\%$  and  $0.2\%$  respectively (Sihto et al., 2011). Observations acquired in Beijing  
106 from July 12 to September 25, 2008, also suggested that nucleation significantly  
107 increases CCN at all supersaturations, even when supersaturation is low (i.e.,  $0.07\%$   
108 and  $0.26\%$ ). Thus, the occurrence of NPF enhanced CCN by a factor of 1.7 and 2.2,  
109 respectively (Yue et al., 2011).

110 However, previous numerical experiments behave oppositely. For instance, Matsui  
111 et al. (2011) quantified the contribution of nucleation to CCN using WRF-chem in  
112 Beijing in August and September 2006 and found reduced CCN under low SS, e.g.,  
113 when  $SS=0.02\%$ , the concentration of CCN is reduced by up to  $\sim 50\%$ . They attributed  
114 this to the fact that the small particles produced by nucleation may inhibit the growth  
115 of the preexisting particles (Matsui et al., 2011). Similarly, Dong et al. (2019) conducted  
116 NPF simulations with the WRF-Chem for the summer of 2008 focusing on the Midwest  
117 of the United States, and found that the nucleation resulted in decreased CCN at low  
118 supersaturation ( $SS=0.1\%$ ). Besides, a study carried out for East Asia in 2009 also  
119 indicated that at low supersaturation (e.g.  $SS=0.1\%$ ), nucleation has little impact on  
120 CCN (Matsui et al., 2013). The contrasting effects of nucleation on CCN at low  
121 supersaturations in model and observations is not explained in these previous studies.

122 At the stage of particle growth, secondary organic aerosol (SOA) formed by  
123 atmospheric oxidation of organic vapors is a major contributor to particle growth to  
124 CCN-related sizes (Liu and Matsui, 2022; Qiao et al., 2021). SOA formed by multi-  
125 generational gas-phase oxidation of semi-volatile and intermediate volatility organic

126 compounds (S/IVOC) is called SI-SOA (Jimenez et al., 2009; Zhang et al., 2007). Zhao  
127 et al. (2016) made a comprehensive assessment of the roles of various SOA precursors  
128 in SOA formation in real atmosphere in China in 2010, and the results demonstrated  
129 that evaporated POA and IVOC (i.e. S/IVOC) made a significant contribution to SOA,  
130 contributing up to 82% to the average SOA concentration in eastern China. However,  
131 the effect of SI-SOA on CCN has not been fully studied.

132 In this paper, WRF-Chem was applied to simulate the effect of the NPF on CCN  
133 in China in February 2017. The simulated results from the WRF-Chem model are firstly  
134 compared with observations in Qingdao, Beijing and Gucheng, exhibiting large biases  
135 in CN. This is followed by an improvement through a few processes. At the end, the  
136 impact of SI-SOA yield and nucleation on CCN is investigated.

## 137 **2. Data and methods**

### 138 **2.1 Observations**

139 The measurements used in this study were carried out over the sampling site from  
140 February 5 to 24, 2017 at the campus of Ocean University of China (36°09'37"N,  
141 120°29'44"E) in Qingdao, which is surrounded by residential buildings and is situated  
142 about 10 km away from the city center. A fast mobility particle sizer (FMPS, TSI Model  
143 3091) was applied to measure the aerosol particle size distribution for the size range of  
144 5.6 nm to 560 nm (Liu et al., 2014b). The bulk CCN concentration is measured by a  
145 cloud condensation nuclei counter at three different supersaturations (0.2%, 0.4% and  
146 0.6%) and each supersaturation lasts for 20 minutes. More information about the CCN  
147 measurement can be found in Li et al. (2015). The urban site in Beijing is located on  
148 the roof of the building of the Chinese Academy of Meteorological Sciences (CAMS,  
149 39°95'N, 116°33'E) in the campus of the China Meteorological Administration, close  
150 to the main road with heavy traffic. The rural site is Gucheng (GC, 39°08'N, 115°40'E),  
151 located in Hebei Province, surrounded by farmland, and is a representative station of  
152 the severity of air pollution in Beijing Tianjin Hebei region. The particle number size  
153 distribution of these two sites in the range of 4–850 nm is measured by a Tandem  
154 Scanning Mobility Particle Sizer (TSMPS), and more information about the  
155 instruments can be found in Shen et al. (2018).

156 **2.2 Model configurations**

157 WRF-Chem version 3.9 is used to simulate NPF events, with the main physical  
 158 and chemical parameterization settings summarized in Table 1. The spatial resolution  
 159 is 36 km by 36 km with 35 vertical layers and a model top at 50 hPa. The regional  
 160 model simulations at a higher spatial resolution may be desirable in future when urban  
 161 pollution is focused. A continuous run from February 1 to 25, 2017, was conducted,  
 162 with the first five-day results as the spin-up and discarded in the analysis.

163 Table 1 WRF-Chem model configurations used in this work

Model configuration	
Microphysics	Morrison 2-moment microphysics scheme (Morrison et al., 2009)
Planetary Boundary Layer (PBL)	YSU boundary layer scheme (Hong et al., 2006)
Longwave and Shortwave Radiation	RRTMG longwave and shortwave radiation (Iacono et al., 2008)
Land model	Unified Noah Land Surface scheme (Chen and Dudhia, 2000; Tewari et al., 2016)
Cumulus	Grell-3D cumulus parameterization scheme (Grell, 1993)
Aerosol module	MOSAIC module (Zaveri et al., 2008; Matsui et al., 2011)
Gas-phase Chemistry	SAPRC-99 gas-phase chemistry scheme (Carter, 2000)

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 165 The meteorological initial and boundary conditions are driven by Climate Forecast  
 166 System model version 2 (CFSv2; (Saha et al., 2014)) reanalysis developed by National  
 167 Centre for Environmental Prediction (NCEP). The initial and boundary chemical  
 168 conditions of WRF-Chem are provided by Community Atmosphere Model with  
 169 Chemistry (CAM-Chem; (Buchholz et al., 2019)). Anthropogenic emissions for the  
 170 year of 2017 are obtained from the Multiresolution Emission Inventory for China  
 171 (MEIC, <http://www.meicmodel.org/>) emission dataset (Li et al., 2017; Zheng et al.,  
 172 2018).

173 The Model for Simulating Aerosol Interactions and Chemistry (MOSAIC) was  
 174 used to delineate dynamic gas-particle mass transfer to represent the condensation  
 175 growth of aerosol (Zaveri et al., 2008). The gas-particle partitioning of gas species on  
 176 particles is regulated by the mass transfer rate, which is related to mass accommodation  
 177 coefficient ( $\alpha$ ), a parameter involved in the model representing the probability of gas  
 178 molecules entering the bulk liquid phase (Pöschl et al., 1998). The original setting of  $\alpha$   
 179 for all condensing species for all size bins  $a$  in MOSAIC is 0.1 (Zaveri et al., 2008). In  
 180 the default release of WRF-Chem, MOSAIC was implemented in the sectional  
 181 framework with aerosol size distributions divided into 4 or 8 size bins spanning 39 nm  
 182 to 10  $\mu\text{m}$  in diameter. To explicitly express the nucleation and the growth of newly  
 183 formed particles, the aerosol size range in the MOSAIC module was extended from 1  
 184 nm to 10  $\mu\text{m}$ , with the number of aerosol size bins increased to 20 (Matsui et al., 2011;  
 185 Matsui et al., 2013; Lupascu et al., 2015; Lai et al., 2022). The calculation method of  
 186 CCN concentration in the WRF-chem model is referred to the study of Matsui et al.  
 187 (2011). Based on Köhler theory, CCN concentrations under the three given  
 188 supersaturations of 0.2%, 0.4% and 0.6% were calculated. The critical supersaturation  
 189 ( $S_c$ ) of each size bin in the WRF-chem model was calculated by the following formula:

$$190 \quad S_c = \sqrt{\frac{4 \times a^3}{27 \times r^3 \times \kappa}} \quad (1)$$

$$191 \quad a = \frac{2 \times \sigma}{R_v \times T \times \rho_\omega} \quad (2)$$

192 Where  $\alpha$  (m) is the coefficient of the Kelvin effect,  $\kappa$  is the volume-averaged  
 193 hygroscopicity, calculated using these values in Table 1,  $r$  (m) is the dry diameter,  $\sigma$  is  
 194 droplet surface tension over water ( $0.076 \text{ N m}^{-1}$ ),  $R_v$  is the gas constant for water vapor  
 195 ( $461.6 \text{ J K}^{-1}\text{kg}^{-1}$ ),  $T$  (K) is the air temperature, and  $\rho_\omega$  is the density of water ( $1000 \text{ kg}$   
 196  $\text{m}^{-3}$ ).

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Table 2 Hygroscopicity Parameters ( $\kappa$ ) in the WRF-Chem Model

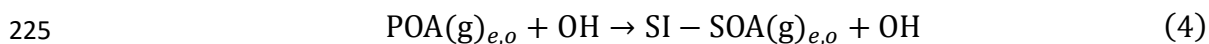
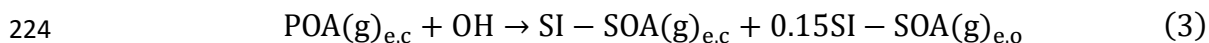
Species	Hygroscopicity ( $\kappa$ )
Sulfate	0.5
Ammonium	0.5
Nitrate	0.5
Black carbon	$10^{-6}$
Primary organic aerosol	0.14
Other inorganics	0.14
Sodium	1.16
Chloride	1.16

201

202 The chemical aging process of organic aerosols (OA) is modeled by the volatility  
 203 basis set (VBS) approach, which was widely used in air quality models to represent  
 204 complex mixtures of thousands of organic species (Donahue et al., 2006; Shrivastava  
 205 et al., 2011; Chrit et al., 2018). The VBS method classifies compounds according to the  
 206 effective saturation concentration ( $c^*$ ), which represents the proportion of the  
 207 component in the gas phase to the particle phase (Donahue et al., 2006), and species  
 208 with higher  $c^*$  values are more volatile. The oxidation of highly volatile precursors to  
 209 form relatively low volatile components represents the aging process of OA. OA  
 210 consists of directly emitted primary organic aerosols and photochemically produced  
 211 secondary organic aerosols (SOA) (Shrivastava et al., 2011). In this study, the  
 212 simplified 2-species VBS mechanism was applied to the simulation of SOA, during  
 213 which primary organic aerosol was represented by two species based on volatility with  
 214 effective saturation concentration  $c^*$  values (at 298 K and 1 atm) of  $10^{-2}$  and  $10^5 \mu\text{g}$   
 215  $\text{m}^{-3}$  (Shrivastava et al., 2011). Primary organic aerosols with  $c^*$  of  $10^5 \mu\text{g m}^{-3}$  refers to  
 216 S/IVOC, which is in the gas phase under most atmospheric conditions due to its high  
 217 volatility, while for those primary organic matters with  $c^*$  of  $10^{-2} \mu\text{g m}^{-3}$ , is treated as  
 218 gas phase as well in the original model. The SOA formed by photochemical oxidation  
 219 of S/IVOC precursors is called SI-SOA and the SOA formed by oxidation of VOC  
 220 precursors is named V-SOA. In the simplified 2-species VBS mechanism, SI-SOA ( $c^*$



221 of  $10^{-2}\mu\text{g m}^{-3}$ ) is formed by the oxidation reaction of S/IVOC precursors ( $c^*$  of  $10^5\mu\text{g}$   
222  $\text{m}^{-3}$ ) and OH with an oxidation rate constant of  $4 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$ . The equations  
223 for controlling the oxidation of S/IVOC precursors are as follows:



226 where POA(g) denotes primary organic aerosols with  $c^*$  of  $10^5 \mu\text{g m}^{-3}$ , which reacts  
227 with OH to form SI-SOA(g) with  $c^*$  of  $10^{-2} \mu\text{g m}^{-3}$ . Subscripts  $c$  and  $o$  represent the  
228 non-oxygen and oxygen parts respectively of given species and  $e$  is either the biomass  
229 or anthropogenic emission sector. In addition, SVOC and IVOC emissions  
230 corresponding to both anthropogenic and biomass burning emissions are derived based  
231 on constant emission ratio of S/IVOC to POA (Shrivastava et al., 2011). A detailed  
232 description of 2-species VBS mechanism can be found in Shrivastava et al. (2011).

### 233 **2.3 Model sensitivity formulations**

234 Three sets of sensitivity tests are designed and listed in Table3. The purposes of  
235 the three sets of experiments are as follows: (1) Adjust the condensation growth process  
236 of ultrafine particles in WRF-Chem model (Base, MAC, PEP, NOCD, RACD, with  
237 details in Table 3).; (2) Explore the effect of SI-SOA yield on CCN (Low\_Yield and  
238 High\_Yield); (3) Study the effect of nucleation process on CCN under the change of  
239 SI-SOA yield (Low\_Yield and High\_Yield and their corresponding cases without  
240 nucleation parameterization, i.e., Low\_nucoeff and High\_nucoeff). Each scenario will be  
241 explained in conjunctions with the results.

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Table 3 The sensitivity tests involved in this study

Purposes	Simulation scenarios	Description
Adjust the condensation growth process of ultrafine particles	Base	Simulation with the default setting with nucleation coefficient set as $2 \times 10^{-6} \text{ s}^{-1}$ , the same as Lai et al. (2022)
	Mass accommodation coefficient (MAC)	It is the same as Base except that the mass adjustment coefficient ( $\alpha$ ) of gaseous sulfuric acid is adjusted from 0.1 to 0.65.
	POA emission phase (PEP)	It is the same as MAC except that the phase of POA is changed from gas phase to particle phase.
	No condensation (NOCD)	It is the same as PEP except that no $\text{NH}_4\text{NO}_3$ condenses on particles below 40 nm.
	Ratio method for condensation (RACD)	It is the same as PEP except that the condensation of $\text{NH}_4\text{NO}_3$ on particles below 40 nm is reduced according to the ratio of acid particles to total particles reported in Wang et al. (2014).
Explore the effect of SI-SOA yield on CCN (Explore the effect of nucleation process on CCN under the change of SI-SOA yield)	High_Yield	Simulation with high oxidation rate of SI-SOA formation with reaction rate constant of $5 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$
	Low_Yield	Simulation with low oxidation rate of SI-SOA formation with reaction rate constant of $2 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$

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Explore the effect of nucleation process on CCN under the change of SI-SOA yield	High_NUCOFF  Low_NUCOFF	Simulations without nucleation parameterizations based on High_Yield  Simulations without nucleation parameterizations based on Low_Yield
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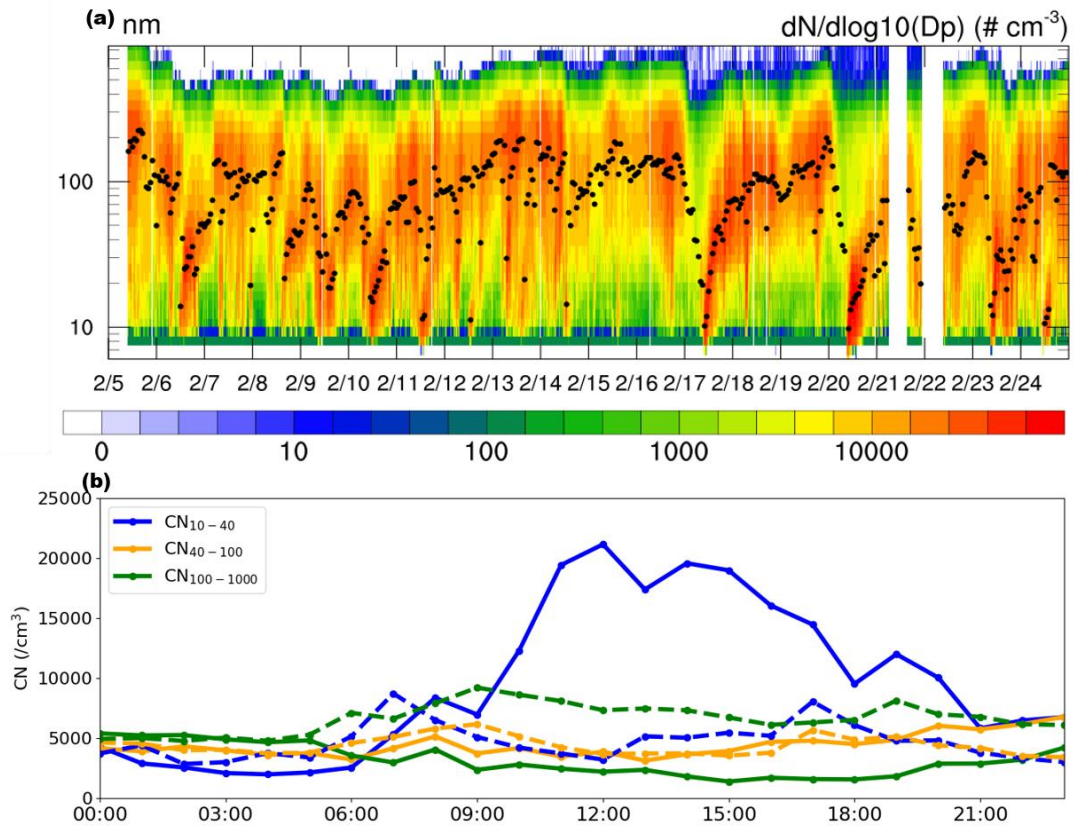
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255 **3. Results**

256 **3.1 Observational analysis**

257       Based on the criteria (Dal Maso et al., 2005; Kulmala et al., 2012), NPF is defined  
258 as an event with the emergence of a nucleation mode with particle diameters smaller  
259 than 25 nm, lasting for 2 hours or more, followed in general by a continuous particle  
260 growth. Six NPF events were identified in February 2017 in Qingdao, on the days of 6,  
261 9, 10, 17, 20 and 23 (Fig. 1a), yielding a frequency of ~30% and displaying a typical  
262 banana-shaped growth of particles in the particle number size distribution. Compared  
263 to a few other studies on NPF frequency in Qingdao, the results in this study are to a  
264 large extent consistent with that in the fall of 2012–2013 (30%; (Zhu et al., 2019)),  
265 slightly higher than that in summer 2016 (22%; (Zhu et al., 2019)) and lower than that  
266 in spring of 2010 (41%; (Liu et al., 2014c)). The higher frequency in spring in Qingdao  
267 is consistent with the observational results at different stations in the Northern  
268 Hemisphere in Nieminen et al. (2018).



269

270 Fig. 1. Distribution of particle number concentration. (a) Temporal evolution of particle  
 271 size distributions (colored shading) and geometric median diameter (GMD; dots in  
 272 black) in Qingdao on February 5-24, 2017. (b) The mean diurnal variation of  $CN_{10-40}$   
 273 (blue),  $CN_{40-100}$  (orange) and  $CN_{100-1000}$  (green) composited during the NPF (solid lines)  
 274 and non-NPF (dashed lines) days on February 5-24, 2017. All times are local times (LT)  
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276 During the six NPF events identified in February in Qingdao, the mean diurnal  
 277 cycle of  $CN_{10-40}$  (10–40 nm) particles exhibits triple peaks (solid blue in Fig. 1b), in the  
 278 morning (8:00 LT), noon (12:00–14:00 LT) and evening (19:00 LT), respectively. A  
 279 comparable three-peak feature was also observed in earlier years during 2016-2018 in  
 280 Qingdao (Zhu et al., 2021). The morning and evening peaks of  $CN_{10-40}$ , with values of  
 281  $\sim 5300 \text{ cm}^{-3}$  and  $\sim 12000 \text{ cm}^{-3}$ , respectively, are likely caused by the primary emissions  
 282 from traffic and cooking activities (Wu et al., 2021a; Wang et al., 2022; Cai et al., 2020).  
 283 The occurrence of NPF starts approximately at 9:00 am LT, accompanied by a  
 284 substantial increase in  $CN_{10-40}$  compared with non-NPF days (solid vs. dashed lines, in  
 285 blue), yielding a peak around noon ( $20000 \text{ cm}^{-3}$  during 12:00–14:00 LT). In addition,

286 larger particles (e.g., CN<sub>40–100</sub> and CN<sub>100–1000</sub>) displayed a slow or no increase in the  
287 afternoon.

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### 289 **3.2 Model improvement in particle number concentration simulations**

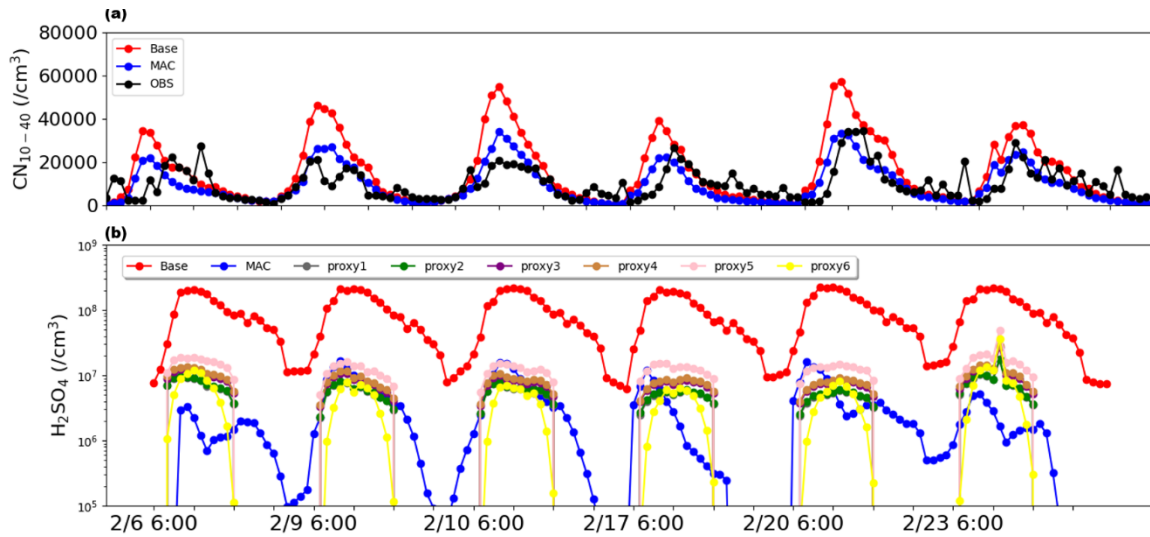
290 Particle number concentrations, primarily in two ranges of 10–40 nm and 40–100  
291 nm, are commonly simulated with large biases. In the smaller size range (10–40 nm),  
292 the particle number concentration is associated with NPF and particle growth. During  
293 NPF, despite differences among the formation mechanisms, H<sub>2</sub>SO<sub>4</sub> is considered the  
294 common species (Yu, 2005; Lovejoy et al., 2004), which often suffer large biases (Cai  
295 et al., 2016; Matsui et al., 2011). In the size range of 40–100 nm, the particle number  
296 concentration is primarily affected by the condensation growth of particles below 40  
297 nm, which is closely related to chemical components such as SOA and nitrate. Prior to  
298 the evaluation of particle number concentration, we first evaluate the compositions of  
299 PM<sub>2.5</sub> and criteria air pollutants including PM<sub>2.5</sub>, PM<sub>10</sub>, O<sub>3</sub>, SO<sub>2</sub>, CO, and NO<sub>2</sub>, showing  
300 relatively low biases compared to observations (section S1 and Fig. S1 and Fig. S2 of  
301 the supporting information).

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#### 303 **3.2.1 Bias correction of particle number concentration at 10–40 nm**

304 In this study, as shown in Fig. 2, comparisons of CN<sub>10–40</sub> between simulations (red  
305 line in Fig. 2a) and observations (black line in Fig. 2a) results of the six NPF events  
306 mentioned in the previous section in Qingdao in February 2017 indicate that model  
307 overestimates CN<sub>10–40</sub> with mean fractional bias of 48%. As one of the major processes  
308 affecting the particle number concentration of 10–40 nm, nucleation is governed by the  
309 particle nucleation rate of 1 nm particles (cm<sup>-3</sup> s<sup>-1</sup>), which is closely associated with the  
310 concentration of H<sub>2</sub>SO<sub>4</sub>. For instance, in a commonly applied activation mechanism,  
311 the nucleation rate calculated by  $J^* = K_{ACT} \times [H_2SO_4]$ . Note that K<sub>ACT</sub> is the nucleation  
312 coefficient considering the physical properties and chemical species of nucleation  
313 process under different environments, indicating that a lumped chemical species are  
314 included in the scheme reflected primarily in the nucleation coefficient k, set as  $2 \times 10$   
315  $^{-6} \text{ s}^{-1}$  based on previous studies (Sihto et al., 2006; Riipinen et al., 2007). Dong et al.

316 (2019) simulated NPF occurring in the summer of 2008 in the United States using the  
 317 NPF-explicit WRF-Chem based on the activation mechanism, which overestimated the  
 318 particle number concentration at 10–63 nm by nearly doubled, even when the  $K_{ACT}$   
 319 decreased by one order of magnitude (set at a very low value of  $10^{-7} \text{ s}^{-1}$ ). Therefore, it  
 320 is likely that the overestimation of particle number concentration in the smaller particle  
 321 size segment is probably due to the bias of simulated sulfuric acid.



322  
 323 Fig. 2. Time series of (a)  $CN_{10-40}$  on NPF days, where red and blue represent Base and  
 324 MAC simulation results respectively, and black represents observation results, and (b)  
 325 sulfuric acid gas concentration obtained by simulation and by proxies (dark grey: Eq.  
 326 5; green: Eq. 6; purple: Eq. 7; brown: Eq. 8; pink: Eq. 9; yellow: Eq. 10). All times are  
 327 in local times.

328

329 Measurement of sulfuric acid gases in the lower troposphere is challenging due to  
 330 the generally low ambient concentration of sulfuric acid ( $10^6-10^7 \text{ molecule cm}^{-3}$ ).  
 331 Different methods have been proposed to estimate ambient sulfuric acid concentrations  
 332 based on observations such as  $SO_2$  (Petäjä et al., 2009; Lu et al., 2019; Mikkonen et al.,  
 333 2011). For instance, Petäjä et al. (2009) proposed a linear method to approximate  
 334 observed  $H_2SO_4$  concentration in Hyytiälä, southern Finland. Moreover, a recent study  
 335 by Lu et al. (2019) proposed a nonlinear method to construct a number of proxies for  
 336 gaseous sulfuric acid concentration (Eq. 5–9), indicating that compared to the linear  
 337 method in Petäjä et al. (2009), the nonlinear relationship can provide more accurate

338 H<sub>2</sub>SO<sub>4</sub> concentration in Beijing during February–March 2018 period. In addition, we  
 339 also used another sulfuric acid nonlinear proxy (Eq. 10) based on long-term  
 340 observations in Germany, Finland, the United States, etc. (Mikkonen et al., 2011). In  
 341 this study, we adopt the above six nonlinear proxy methods (referred as proxy5 to  
 342 proxy10) to estimate H<sub>2</sub>SO<sub>4</sub> in Qingdao.

$$343 \quad [H_2SO_4] = 515.74 \times [SO_2]^{0.38} \times \text{Radiation}^{0.14} \times CS^{0.03} \quad (5)$$

$$344 \quad [H_2SO_4] = 280.05 \cdot \text{Radiation}^{0.14} [SO_2]^{0.40} \quad (6)$$

$$345 \quad [H_2SO_4] = 9.95 \times [SO_2]^{0.39} \times \text{Radiation}^{0.13} \times CS^{-0.01} \times [O_3]^{0.14} \quad (7)$$

$$346 \quad [H_2SO_4] = 14.38 \times [SO_2]^{0.38} \times \text{Radiation}^{0.13} \times [O_3]^{0.14} \quad (8)$$

$$347 \quad [H_2SO_4] = 0.0013 \times [SO_2]^{0.38} \times \text{Radiation}^{0.13} \times CS^{-0.17} \times ([O_3]^{0.14} + [NO_x]^{0.41}) \quad (9)$$

$$348 \quad [H_2SO_4] = 8.21 \times 10^{-3} \times [SO_2]^{0.62} \times \text{Radiation} \times (CS \times RH)^{-0.13} \quad (10)$$

349 where [SO<sub>2</sub>], [O<sub>3</sub>] and [NO<sub>x</sub>] (molecule cm<sup>-3</sup>) represents concentration of  
 350 observed SO<sub>2</sub>, O<sub>3</sub> and NO<sub>x</sub>, respectively. “Radiation” (W m<sup>-2</sup>) is global radiation. RH  
 351 (%) is the relative humidity, and CS (s<sup>-1</sup>) is the condensation sink, which is calculated  
 352 based on observed particle distribution.

353

354 The simulated H<sub>2</sub>SO<sub>4</sub> concentration from the Base simulation (dots in Fig. 2b) is  
 355 compared with observations obtained by proxies (see Fig. 2b), indicating that Base  
 356 simulations apparently overestimate by one order of magnitude compared to the H<sub>2</sub>SO<sub>4</sub>  
 357 estimated by proxies. The overestimation has been frequently reported previously, i.e.,  
 358 over Beijing (Matsui et al., 2011), which ascribes the bias to the overestimation of the  
 359 SO<sub>2</sub> concentration. In a more recent study, the sensitivity of H<sub>2</sub>SO<sub>4</sub> to SO<sub>2</sub> is tested, and  
 360 the result shows that even when SO<sub>2</sub> is reduced to an unrealistically low level, the  
 361 simulated H<sub>2</sub>SO<sub>4</sub> is still more than one order of magnitude higher than the observed  
 362 value (Lai et al., 2022), suggesting that the SO<sub>2</sub> concentration cannot fully explain the  
 363 overestimates.

364 In addition to the precursor of H<sub>2</sub>SO<sub>4</sub>, the mass accommodation coefficient ( $\alpha$ ),  
 365 representing the probability of impaction of a gaseous molecule on a liquid surface and  
 366 entering the bulk liquid phase, is another important factor affecting the concentration

367 of sulfuric acid gas. In the public release of WRF-Chem, mass accommodation  
368 coefficient is typically set to a low value of 0.1 for all gas species under different  
369 volatility during the condensation process, including H<sub>2</sub>SO<sub>4</sub> (Davidovits et al., 2004;  
370 Zaveri et al., 2008). Recent studies indicate that the low mass accommodation  
371 coefficient value may not be applicable to the low volatile gases, which tend to have a  
372 mean mass accommodation coefficient value of 0.7 and close to the unity (Krechmer et  
373 al., 2017). In fact, an earlier study has indicated based on experimental determination,  
374 the mass accommodation coefficient of H<sub>2</sub>SO<sub>4</sub> vapor in sulfuric acid aqueous solution  
375 was measured, and the best fit value was 0.65 (Pöschl et al., 1998). Accordingly, a  
376 sensitivity simulation was conducted by adjusting the mass accommodation coefficient  
377 of H<sub>2</sub>SO<sub>4</sub> from 0.1 to 0.65, referred to as MAC.

378 This simulation brought the H<sub>2</sub>SO<sub>4</sub> concentration (see Fig. 2b) much closer to the  
379 calculated results from proxies, and the corresponding biases reduced by approximately  
380 an order of magnitude. Notably, the MAC simulation decreases the overestimate of  
381 sulfuric acid gas concentration, resulting in a lower particle formation rate. The MAC  
382 simulation also significantly reduces overestimate of CN<sub>10–40</sub> (Fig. 2a), and mean  
383 fractional bias compared to observations decreases from 48% to 1%.

384

### 385 **3.2.2 Improvement of particle number concentration simulations at 40–100 nm**

386 The number concentration of particles in the 40–100 nm range is mainly affected  
387 by the coagulation and condensation processes. While the coagulation process tends to  
388 largely affect ultrafine particles below 10 nm than those with larger sizes (Wu et al.,  
389 2011), the condensation growth of particles during gas-particle partitioning at sizes of  
390 10–40 nm, to a large extent, governs the variations in number concentration of 40–100  
391 nm particles. The condensation process is primarily controlled by gas-particle  
392 partitioning of chemical species, which may change the chemical composition of  
393 particles, such as organic compounds and inorganics including sulfate, nitrate and  
394 ammonium.

395 Among the species contributing to the condensation growth of particles at 10–40  
396 nm, the organic compounds with  $c^*$  of  $10^{-2} \mu\text{g m}^{-3}$  play the dominant role (Pierce et al.,



397 2011). In the current model setting, the low volatile organic matter of  $10^{-2} \mu\text{g m}^{-3}$  comes  
398 from two gas-phase sources, including the direct emission of primary organic aerosol  
399 (POA) and SOA formed from S/IVOC (SI-SOA), conducive to condensation on  
400 particles. While the condensation of gaseous SOA is in general reasonable, the gas  
401 phase emissions of POA may be problematic. For instance, previous studies suggested  
402 that POA is in gas phase close to the emissions source. However, with rapid dilution  
403 and cooling in the atmosphere away from the source, most POA condenses to particle-  
404 phase (Roldin et al., 2011b; Roldin et al., 2011a; Shrivastava et al., 2008). Therefore,  
405 away from the emissions source POA, being in the particle phase, will not be involved  
406 in the growth of newly formed particles. Therefore, POA may not contribute to particle  
407 growth away from the emission sources, which caused different size distributions of  
408 POA compared to when it was emitted in the gas-phase (Fig. S3a vs. Fig. S3b). Emitting  
409 low volatility POA in the particle phase eliminates the unreasonable quasi-banana shape  
410 pattern exhibiting concomitant growth of newly formed particles with increasing mass  
411 concentration of POA.

412 The composition analysis (Fig. S3c) in the 10–40 nm particles mass from the  
413 model results indicates that organic compounds mentioned above only account for 21%  
414 of total mass (sulfates, nitrates, ammonium salts and organics) in this size range and the  
415 dominant species is nitrate which accounts for 51% of total mass, exhibiting  
416 inconsistencies with the previous studies which in general indicates a much smaller  
417 contribution of nitrate. For instance, Liu et al. (2014a) suggested that over North China  
418 Plain in summer 2009, organic matter accounted for 77% of particles around 30 nm,  
419 while the sum of  $\text{SO}_4^{2-}$ ,  $\text{NO}_3^-$  and  $\text{NH}_4^+$  only accounted for 18%. Recent observations  
420 conducted in Beijing also indicated that particles at 8–40 nm are mainly composed of  
421 organic matter (with mass fraction of  $\sim 80\%$ ) and sulfate (with mass fraction of  $\sim 13\%$ ),  
422 while nitrate content is very low (with mass fraction of  $\sim 3\%$ ) (Li et al., 2022). Another  
423 study showed that nitrate accounted for 7–8% at urban sites and 17% at rural sites for  
424 particles mass in 7–30 nm in the United States in 2007 (Bzdek et al., 2012). Therefore,  
425 the potentially too high modeled nitrate fraction in 10–40 nm in this study is tightly  
426 associated with the condensation process, with the specific reasons explained below.

427 The condensation of nitric acid on particles is highly constrained by the particle  
428 acidity. The acidity in smaller particles (i.e., 10–40 nm) tends to be higher than that in  
429 large particles, primarily due to the larger condensation of H<sub>2</sub>SO<sub>4</sub> (Lu et al., 2022), and  
430 particles with sizes greater than 40 nm have a much weaker acidity or are nearly neutral.  
431 For example, observed evidence has shown that acidic ultrafine particles account for a  
432 large proportion of ultrafine particles from 22 December 2010 to 15 January 2011 in  
433 Hong Kong, e.g., 65% for particles within 5.5–30 nm (Wang et al., 2014).

434 In the model, a particle is determined to be in solid phase when the ambient relative  
435 humidity is lower than the mutual deliquescence relative humidity of the particles  
436 (Zaveri et al., 2005; Zaveri et al., 2008), which is in general suitable for particles  
437 dominated by inorganics. In the study area, the results indicate that at most conditions  
438 relative humidity are relatively low and the particles are in solid phase, in which the  
439 condensation process is not affected by particle acidity and the condensation of nitric  
440 acid on particles is directly calculated based on the gas-particle equilibrium  
441 concentration (Zaveri et al., 2008). However, for particles below 40 nm, the main  
442 compositions are likely to be organic matter (Zhu et al., 2014; Ehn et al., 2014), which  
443 tends to be in liquid phase (Virtanen et al., 2011; Cheng et al., 2015), under which the  
444 condensation of nitric acid is strongly constrained by acidity. Therefore, the phase  
445 misrepresentation ignores the weakening effect of acidity on nitric acid condensation,  
446 resulting in too high nitrate therein.

447 To overcome this issue, we propose a ratio method for condensation (RACD) to  
448 partition the condensation of nitric acid on particles under 40 nm, by applying a ratio  
449 of the number concentration of non-acidic particles to ultrafine particles. The method  
450 is based on two assumptions, including: 1) little condensation of nitric acid on particles  
451 with strong acidity (Lu et al., 2022); 2) the condensation of nitric acid on particles is  
452 proportional to the ratio of the number concentration of non-acidic ultrafine particles to  
453 the total particles, despite the existence of uncertainties. Fig. S4 depicts the average  
454 particle number concentration and acid particle in the 1 to 40 nm range, calculated based  
455 on Wang et al. (2014). The ratio of non-acidic particles is 8% for particles below 10 nm,  
456 18% for particles at 10–15.8 nm, 30% for particles at 15.8–25.1nm, and 55% for

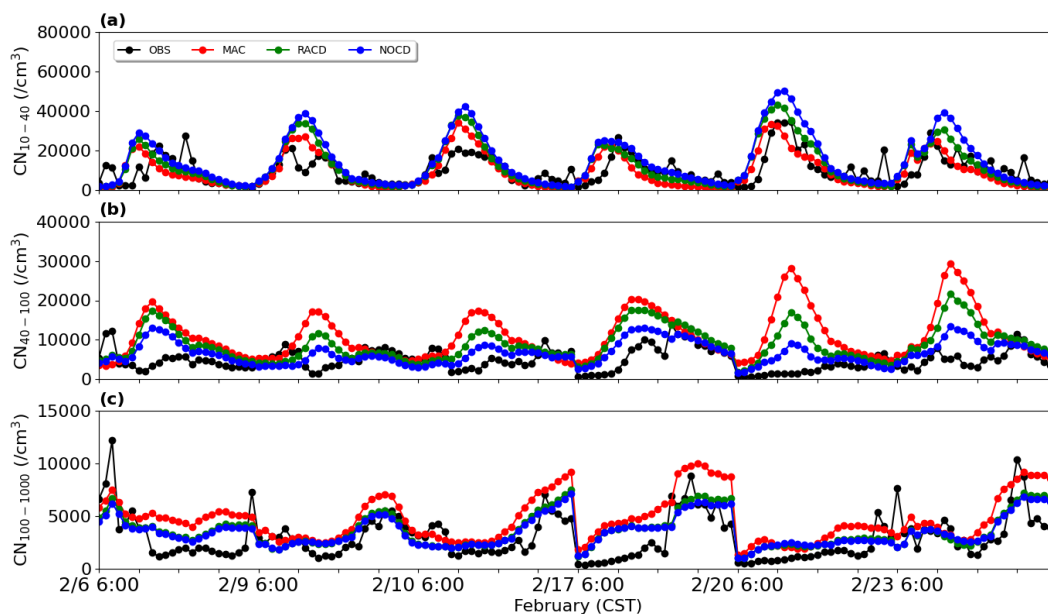
457 particles at 25.1–39.8 nm (Fig. S4). Note that the ratio is based on measurements  
458 acquired at a single site in Hong Kong, therefore more observational studies are needed  
459 to warrant the robustness of the method. Alternatively, the condensation of nitric acid  
460 on particles in bins from 1 nm to 40 nm is completely suppressed, referred to as NOCD.

461 The simulation results based on the two methods (RACD and NOCD) are shown in  
462 Fig. 3. Compared to MAC, RACD simulations reduce previously noted overestimation  
463 of particle number concentration in the 40–100 nm size range (Fig. 3b), with the mean  
464 fractional bias decreases from 83% to 63%. In addition to the amount of nitrate  
465 condensation during particle growth mentioned above, the overestimation of particle  
466 number concentrations in the 40–100 nm range may be attributed to nucleation process.  
467 More specifically, in the H<sub>2</sub>SO<sub>4</sub>-H<sub>2</sub>O binary nucleation mechanism used in this study,  
468 when the concentration of sulfuric acid gas is reduced (Section 3.2.1), the resulting  
469 decrease in nucleation rate leads to a slight decrease in particle number concentration  
470 at 40–100 nm relative to Base (mean fractional bias from 98% to 83%). Apart from that,  
471 it may also be related to the choice of nucleation parameterization scheme. For example,  
472 using a global chemical transport model GEOS-Chem with a nucleation mechanism in  
473 which formation rate is a function of the concentrations of sulfuric acid and low-  
474 volatility organics, Yu et al. (2015) overestimated the concentration of particles in the  
475 10–100 nm range by 161% at nine sites in the summer in North America. A possible  
476 explanation for this overestimation was given by the uncertainty of the predicted  
477 concentration of organic compounds involved in organics-mediated nucleation  
478 parameterization. After they switched to another scheme of the ion-mediated nucleation  
479 mechanism without organic matter, the number becomes 27% lower than the  
480 observations (Yu et al., 2015). The test based on different schemes is beyond the scope  
481 of the study, which is therefore not investigated.

482 Moreover, the overestimation of particles over 100 nm (CN<sub>100–1000</sub>; Fig. 3c), which  
483 have a strong influence on CCN, also decrease in the RACD simulation. Thus, the mean  
484 fractional bias decreases from 25% (MAC) to 1%. Note that the slight increase of CN<sub>10–</sub>  
485 40 through the application of RACD, can be linked to the decrease of nitrate  
486 condensation, and leads to weakened particle growth and enhanced particle number

487 concentration at 10–40 nm (Fig. 3a). The alternative method by completely removing  
 488 the nitrate condensation (NOCD) yields even better performance in particle number  
 489 concentration of 40–100 nm (mean fractional bias of 34%), indicating the feasibility by  
 490 reducing the nitrate condensation. The proportion of nitrate simulated by RACD is 23%,  
 491 closer to values reported in past observations (Bzdek et al., 2011; Bzdek et al., 2012),  
 492 while the nitrate (1%) in the scenario of NOCD seems to be too low. Considering the  
 493 limited observational information obtained based on previous studies, RACD is applied  
 494 in this study.

495 In addition to Qingdao, we evaluate the model performance over a few other sites,  
 496 including one site over urban Beijing and the other one over the rural area of Gucheng,  
 497 yielding consistent improvements in model simulations (Section S2; Fig. S6-S8).  
 498 Moreover, we select another empirical scheme, e.g., kinetics, and one classical  
 499 nucleation scheme, indicating the empirical scheme of activation scheme is in general  
 500 a good option in this study (Section S2; Fig. S9-S11; Table S1-3).



501  
 502 Fig. 3. The time series of (a)  $CN_{10-40}$ , (b)  $CN_{40-100}$  and (c)  $CN_{100-1000}$  on NPF days in  
 503 Qingdao on February 5-24 simulated from MAC (marked in red), NOCD (marked in  
 504 blue) and RACD (marked in green) as well as from observations (OBS) (marked in  
 505 black). All times are local time.

506  
 507

### 508 **3.3 Substantial contributions of SI-SOA to CCN**

509 Compared with the original model setting, after adjusting the growth process of  
510 ultrafine particles (RACD), the number concentration of particles tends to decrease,  
511 especially for particles above 40 nm. Ultrafine particles above 40 nm are important  
512 sources of CCN (Dusek et al., 2006), in this way, the number concentration of CCN  
513 also tends to decline. In addition, in the Base case, we found that the model  
514 overestimated  $CCN_{0.4\%}$  and  $CCN_{0.6\%}$ , with mean fractional bias being 64% and 87%,  
515 respectively. After adjusting the condensation growth process of ultrafine particles,  
516 under high supersaturation (i.e.,  $CCN_{0.4\%}$  and  $CCN_{0.6\%}$ ), the capability of the model in  
517 reproducing the CCN is improved. RACD reduces the overestimation of  $CCN_{0.4\%}$  and  
518  $CCN_{0.6\%}$ , with mean fractional bias reduced to 30% and 56%, respectively, although the  
519 overestimates still exist (Figs. S5b, c). However, for low supersaturation (i.e.,  $CCN_{0.2\%}$ ),  
520 the decrease of number concentration of CCN is too large, and mean fractional bias  
521 decreases from 7% to -45% (Fig. S5a), therefore the bias will be further adjusted later.

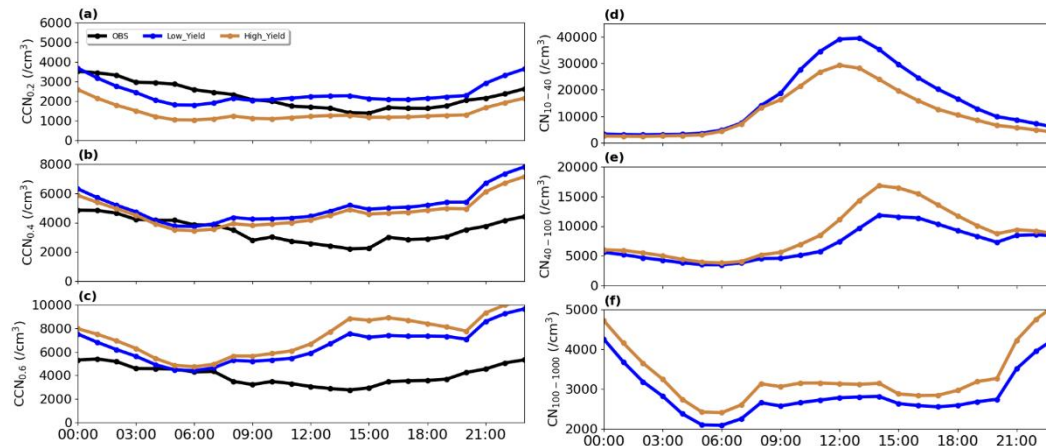
522 In addition to the growth process, the remaining overestimate of CCN under high  
523 SS and underestimate of CCN over low SS is likely to be influenced by the chemical  
524 compositions involved in the activation of ultrafine particles into CCN. Specifically,  
525 ultrafine particles can grow up to CCN size under certain SS (Pierce and Adams, 2007).  
526 This process is influenced by both particle size and hygroscopicity, and hygroscopicity  
527 is closely related to the chemical composition of particles (Petters and Kreidenweis,  
528 2007). In particular, inorganic compounds generally increase particle hygroscopicity,  
529 increasing CCN. SOA has dual effects on CCN since it decreases particle  
530 hygroscopicity but also promotes growth of particles, and these two effects are  
531 competitive with each other (Wu et al., 2015; Zaveri et al., 2021). Ultrafine particles  
532 must grow to a critical size to be activated into CCN (Dusek et al., 2006). SOA act as a  
533 major contributor in promoting the condensational growth of ultrafine particles to the  
534 critical size, facilitating particles activation into CCN. In contrast, SOA tends to reduce  
535 the hygroscopicity of particles, leading to a diminished ability of activation to CCN  
536 (Wu et al., 2015). These two competing effects work together and modulate the number

537 of CCN. Moreover, considering that SI-SOA is the main SOA component on ultrafine  
538 particles (Fig. S5d), the effect of SI-SOA on CCN is therefore explored in this study.

539 Considering SI-SOA is a product of S/IVOC oxidation, the oxidation rate of  
540 S/IVOC is tightly associated with CCN, which likely affects the bias of CCN. In the  
541 original model setup, the oxidation rate is set to be a constant of  $4 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1}$   
542  $\text{s}^{-1}$  for all S/IVOC. However, a recent study (Wu et al., 2021b) proposed that the  
543 oxidation rate can be as high as  $5 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$  such as for polycyclic aromatic  
544 hydrocarbons (PAHs), close to the original model value, but can be as low as half (i.e.,  
545  $2 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$ ) of the original modeling setting for S/IVOC species except  
546 PAHs (O-S/IVOCs). It is noteworthy that the oxidation rates of  $5 \times 10^{-11}$  and  $2 \times 10^{-11}$   
547 in general represent the upper and lower bounds (Zhao et al., 2016; Wu et al., 2021b).

548 To delve into how oxidation rates affect CCN, we set up a few numerical  
549 experiments (Table 3) to investigate the response of CCN to the oxidation rate of  
550 S/IVOC at three supersaturations (0.6%, 0.4%, 0.2%), including cases of High\_Yield  
551 and Low\_Yield. As it is shown in Fig. 4, decreasing the oxidation rate (Low\_Yield)  
552 leads to a reduction of  $\sim 10\%$  of CCN at high supersaturation (i.e., CCN<sub>0.6%</sub>) as  
553 compared to the High\_Yield simulation. This behaviour is a consequence of the  
554 decrease of particle number concentrations associated with Low\_Yield, particular of  
555 the particles close to the critical diameter (40–100 nm). In this case, the effect of particle  
556 size dominates the hygroscopicity. In contrast, at a lower supersaturation (CCN<sub>0.2%</sub>),  
557 CCN increases by 42% when the oxidation rate is switched from a high to a low value,  
558 which is due to the smaller fraction of SI-SOA contributing to particulate mass when  
559 the oxidation rate is low. In this case, relative to SOA, a larger fraction of other particle  
560 constituents such as inorganics, increase the volume weighted particle hygroscopicity  
561 (Dusek et al., 2006) which causes the increase of CCN number. This means that the  
562 effect of hygroscopicity on CCN surpasses the influence on particle size at low  
563 supersaturations. This conclusion is consistent with the observation conducted by Ma  
564 et al. (2016) in the North China Plain in 2013, which suggested that along with the  
565 decrease of SS, the particles that can be activated to CCN is more sensitive to changes  
566 of particle hygroscopicity. Similarly, based on observational data in northern China in

567 summer, Wang et al. (2023) found that CN in 2020 is lower than that in 2014 due to  
 568 particulate pollution control, however, the particles become more easily activated,  
 569 attributable to the larger extent of decrease in organic matters compared to inorganics,  
 570 leading to enhanced particle hygroscopicity and more conducive to activation.



571  
 572 Fig. 4. Average diurnal variation of (a)  $CCN_{0.2\%}$ , (b)  $CCN_{0.4\%}$  and (c)  $CCN_{0.6\%}$  and (d)  
 573  $CN_{10-40}$ , (e)  $CN_{40-100}$ , (f)  $CN_{100-1000}$  on NPF days in Qingdao on February 5-24, 2017,  
 574 in Low\_yield and High\_yield simulations, shown as blue and brown lines, and black  
 575 lines represent observation results.

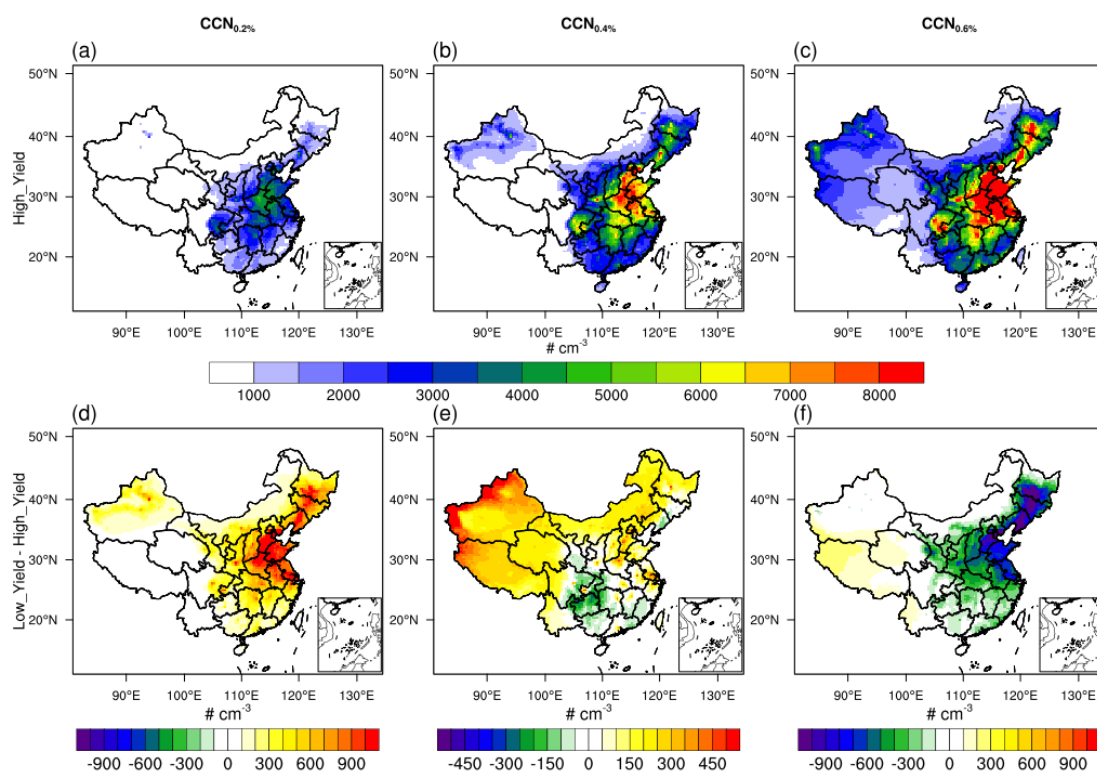
576

577 Furthermore, compared to the high yield of SI-SOA, the low SI-SOA yield results  
 578 in a high CCN concentration under low SS and low CCN concentration under high SS.  
 579 Therefore, both the underestimates of  $CCN_{0.2\%}$  (mean fractional bias of -45%) and  
 580 overestimates of  $CCN_{0.6\%}$  (mean fractional bias of 56%) mentioned above are improved,  
 581 with mean fractional bias of  $CCN_{0.2\%}$  and  $CCN_{0.6\%}$  reaching 7% and 43%, respectively  
 582 (Fig. 4a,c). This result suggests that the oxidation rate of S/IVOC is possibly closer to  
 583 the low value, which is understandable based on Wu et al. (2021b), who found that the  
 584 amount of O-S/IVOCs, which corresponds to a low oxidation rate, is in general much  
 585 larger (i.e., 20 times) than that of PAHs with a high oxidation rate.

586 In addition to the single site of Qingdao, we further explore the impact of SI-SOA  
 587 yield on CCN from a larger spatial coverage (Fig. 5). Consistent with the mechanism  
 588 revealed over Qingdao, even from a larger spatial perspective, a lower oxidation rate of  
 589 S/IVOC essentially enhances CCN at a lower SS (e.g.,  $CCN_{0.2\%}$ ; Fig. 5a) with the

590 highest increase over North China Plain area (Fig. 5a), and weakens CCN (i.e., by 10–  
591 20% over Beijing-Tianjin-Hebei) at a higher SS (Fig. 5c), particularly over the dense  
592 emission area (Fig. S12). It is worth noting that in the 2-species VBS mechanism used  
593 in our study, all S/IVOC in the inventory is calculated based on a constant emission  
594 ratio of S/IVOC to POA from all source categories (Shrivastava et al., 2011), which  
595 may miss part of S/IVOC due to different emission ratios of POA from different source  
596 (Chang et al., 2022). In addition, the simplified VBS mechanism used in our study does  
597 not take into account the multi-step oxidation of organic species, which may introduce  
598 some uncertainties. To be more specific, in the 2-species VBS mechanism, SI-SOA with  
599 effective saturation concentrations ( $c^*$ ) of  $10^{-2} \mu\text{g m}^{-3}$  is formed by the vapor phase  
600 oxidation of S/IVOC vapors with  $c^*$  of  $10^5 \mu\text{g m}^{-3}$ , reducing volatility by 7 orders of  
601 magnitude. The process of one-step oxidation does not mean to represent a physical  
602 process, but to parameterize the mean effect of a complex process of SOA formation  
603 (Shrivastava et al., 2011). However, in the real atmosphere, the gaseous VOCs often  
604 undergo multi-generational oxidation to form SOA (Garmash et al., 2020), during  
605 which the properties and composition of SOA change substantially. For instance, by  
606 adding the formation chemistry associated with multi-generational oxidation, Zhao et  
607 al. (2020) found improved simulations of vertical aerosol profile in the Amazon free  
608 troposphere compared to the simplified VBS mechanism.





609

610 Fig. 5. Spatial distributions of CCN concentrations at different supersaturations (SS),  
 611 (a) and (d) are  $CCN_{0.2\%}$ , (b) and (e) are  $CCN_{0.4\%}$ , and (c) and (f) are  $CCN_{0.6\%}$ . The top  
 612 panels exhibit the results from the High\_Yield simulation, and the bottom panels shows  
 613 the difference between the Low\_Yield and High\_Yield simulations.

614

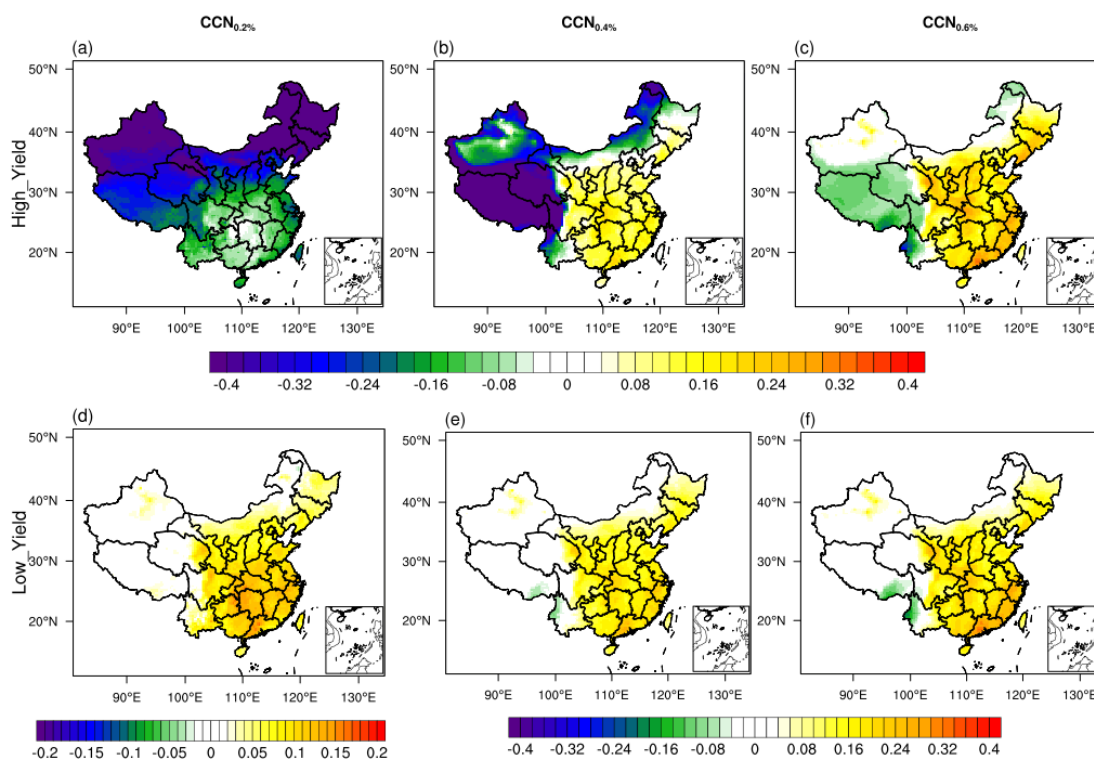
### 615 3.4 Contribution of nucleation to CCN under different SI-SOA yields

616 Considering the importance of nucleated particles on CCN (Yu et al., 2020;  
 617 Westervelt et al., 2013), we further investigate the influence of nucleation on CCN  
 618 under different SI-SOA yield conditions discussed above.

619 As shown in Fig. 6, in simulations close to the original model setting (High\_Yield),  
 620 when SS is low (i.e.,  $SS=0.2\%$ ), the nucleation process tends to reduce the CCN by  
 621  $\sim 10\text{--}50\%$ . In contrast, when the SS is high (0.6%), the nucleation results in a significant  
 622 increase in CCN in most regions of China. When the yield of SI-SOA is adjusted to a  
 623 lower level, the nucleation process has a positive contribution to CCN under both low  
 624 and high SS. Especially, when SS is low (0.2%), the sign reversal, i.e., from negative  
 625 (Fig. 6a) to positive (Fig. 6d) contributions of NPF to CCN along with the decrease of  
 626 SI-SOA yield, i.e., the increase is concentrated in the eastern China with an average of

627 10–20%. The primary mechanism lies in that along with the decrease of SI-SOA yield,  
 628 the smaller fraction of SI-SOA yields an increase in hygroscopicity, which surpasses  
 629 the suppression effect on particle growth due to reduced SI-SOA formation. In the real  
 630 atmosphere, when the supersaturation is usually low, e.g. about ~0.1% in polluted areas  
 631 (Kalkavouras et al., 2019; Hudson and Noble, 2014), CCN will likely reduce with  
 632 increasing oxidation rate of S/IVOC and corresponding SI-SOA formation.

633



634

635 Fig. 6. Spatial distribution of contribution of nucleation to CCN calculated by the ratio  
 636 of the difference between the parameterization with and without nucleation to the  
 637 parameterization with nucleation under different SI-SOA yields in China in February  
 638 2017. (a), (d) is CCN<sub>0.2%</sub>, (b), (e) is CCN<sub>0.4%</sub>, (c), (f) is CCN<sub>0.6%</sub>. The upper panel and  
 639 lower panel represent High\_Yield and Low\_Yield simulation respectively

640

641 In addition to the linear-H<sub>2</sub>SO<sub>4</sub> nucleation mechanism, one more empirical scheme  
 642 of kinetics nucleation is selected, which assumes that the nucleation rate is proportional  
 643 to the square of the concentration of sulfuric acid ( $J = K[H_2SO_4]^2$ ), to investigate the  
 644 effect of nucleation on CCN. Substantially positive contributions of nucleation to CCN

645 is found when the low SI-SOA yield is applied, consistent with what was shown based  
646 on the linear-H<sub>2</sub>SO<sub>4</sub> nucleation scheme (Fig. S13). However, nucleation contributes  
647 positively to CCN even when the SI-SOA yield is high in the quadratic-H<sub>2</sub>SO<sub>4</sub>  
648 nucleation scheme (e.g., kinetics nucleation scheme). When more sulfuric acid  
649 molecules participate in nucleation under this scheme than the linear-H<sub>2</sub>SO<sub>4</sub> nucleation  
650 scheme, the particles are more easily hygroscopically activated to CCN, which is  
651 equivalent to the effect of a reduction in organic components in the linear-H<sub>2</sub>SO<sub>4</sub>  
652 nucleation scheme (e.g., activation-type nucleation scheme). The results from this study  
653 show the importance of assessing the simulated effects of the nucleation scheme on not  
654 only the formation and growth process of particles but also climate factors such as CCN  
655 using observations.

656

## 657 **Conclusions and discussions**

658 In this study, WRF-Chem explicit-NPF simulations, with linear-H<sub>2</sub>SO<sub>4</sub> nucleation  
659 scheme (e.g., activation-type nucleation scheme), are used to investigate the observed  
660 wintertime NPF events and their contribution to CCN in China. Based on observations  
661 in a typical coastal city of Qingdao, as well as in the cities of Beijing and Gucheng over  
662 North China Plain, we identify high biases of the model simulated CN and CCN  
663 concentrations. Therefore, we updated and improved the parameterization setting on  
664 particle growth in the model, mainly including: (1) adjusting the mass accommodation  
665 coefficient ( $\alpha$ ) to from the default value of 0.1 to 0.65, an important parameter for  
666 sulfuric acid condensation; (2) proportionally reducing the condensation amount of  
667 nitric acid on particles below 40 nm, (3) changing the emitted low-volatility POA from  
668 gas to particle. Through these adjustments, the capability of the model in reproducing  
669 CN and CCN is substantially improved, leading to better agreement with the observed  
670 results, which significantly reduces the overestimation of CN<sub>10-40</sub> (mean fractional bias  
671 decreases from 48% to 1%) and CN<sub>40-100</sub> (mean fractional bias decreases from 98% to  
672 63%).

673

674 For CCN, due to the crucial role of SI-SOA in promoting the growth of ultrafine  
675 particles, on the basis of previous studies, we lower the oxidation rate of S/IVOC and  
676 hence the production rate of SI-SOA, which weakens the growth of particles to reach  
677 the critical size of CCN activation, but enhances particulate hygroscopicity favoring the  
678 activation to CCN. When the yield of SI-SOA is adjusted to the lower bound of  
679 literature value,  $CCN_{0.6\%}$  is reduced by  $\sim 10\%$  and is closer to observations. At low SS  
680 ( $CCN_{0.2\%}$ ), the decrease of SI-SOA yield has greater effects on the increase of particle  
681 hygroscopicity compared to the effect of the reduction of particle size due to the  
682 decrease of condensation growth. It results in an increase of CCN (as large as  $\sim 42\%$ ) in  
683 better agreement with observations. Under low SS conditions, common in the  
684 atmosphere, a 2.5-fold reduction in SI-SOA yield results in a substantial increase of  
685 CCN that switches from a negative contribution of new particle formation to CCN from  
686  $-50\% \sim -10\%$  to a positive contribution of  $10 \sim 20\%$ .

687 In addition to activation nucleation scheme, we have also tested a few other  
688 schemes such as the quadratic- $H_2SO_4$  nucleation scheme (e.g., kinetics nucleation  
689 scheme). Under this scheme, the bias-corrected method abovementioned is applicable  
690 to improving the simulations of concentrations of CN and CCN. It is noteworthy that  
691 the dependence of CCN on the SI-SOA yield is diminished, showing that under both  
692 high and low yields of SI-SOA, there are positive contributions of NPF to CCN. This  
693 is likely due to the increase in the amount of sulfuric acid involved in nucleation,  
694 making it more hygroscopic and easier to activate to CCN, and the high content of  
695 inorganic species makes them less sensitive to changes in SI-SOA yield, which deserves  
696 further investigation.

697

698 **Competing interests.** At least one of the (co-)authors is a member of the editorial board  
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700

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709 **References:**

- 710 Arghavani S, Rose C, Banson S, et al. 2022. The Effect of Using a New Parameterization of Nucleation  
711 in the WRF-Chem Model on New Particle Formation in a Passive Volcanic Plume. *Atmosphere*  
712 [J], 13(1): 15.
- 713 Buchholz R R, Emmons L K, Tilmes S 2019. The CESM2 Development Team., 2019. CESM2.1/CAM-  
714 Chem Instantaneous Output for Boundary Conditions. UCAR/NCAR - Atmospheric Chemistry  
715 Observations and Modeling Laboratory.
- 716 Bzdek B, Zordan C, Luther G, et al. 2011. Nanoparticle Chemical Composition During New Particle  
717 Formation. *Aerosol Science and Technology* [J], 45(1041-1048).
- 718 Bzdek B R, Zordan C A, Pennington M R, et al. 2012. Quantitative Assessment of the Sulfuric Acid  
719 Contribution to New Particle Growth. *Environmental Science & Technology* [J], 46(8): 4365-  
720 4373.
- 721 Cai C, Zhang X, Wang K, et al. 2016. Incorporation of new particle formation and early growth  
722 treatments into WRF/Chem: Model improvement, evaluation, and impacts of anthropogenic  
723 aerosols over East Asia. *Atmospheric Environment* [J], 124(262-284).
- 724 Cai J, Chu B, Yao L, et al. 2020. Size-segregated particle number and mass concentrations from different  
725 emission sources in urban Beijing. *Atmos. Chem. Phys.* [J], 20(21): 12721-12740.
- 726 Carter W 2000. Documentation of the SAPRC-99 Chemical Mechanism for VOC Reactivity Assessment.  
727 Final Report to California Air Resources Board [J].
- 728 Chang X, Zhao B, Zheng H, et al. 2022. Full-volatility emission framework corrects missing and  
729 underestimated secondary organic aerosol sources. *One Earth* [J], 5(403-412).
- 730 Chen F, Dudhia J 2000. Coupling an Advanced Land-Surface/Hydrology Model with the Penn  
731 State/NCAR MM5 Modeling System. 129(
- 732 Cheng Y, Su H, Koop T, et al. 2015. Size dependence of phase transitions in aerosol nanoparticles. *Nature*  
733 *Communications* [J], 6(1): 5923.
- 734 Chrit M, Sartelet K, Sciare J, et al. 2018. Modeling organic aerosol concentrations and properties during  
735 winter 2014 in the northwestern Mediterranean region. *Atmos. Chem. Phys.* [J], 18(24): 18079-  
736 18100.
- 737 Chu B, Kerminen V-M, Bianchi F, et al. 2019. Atmospheric new particle formation in China.  
738 *Atmospheric Chemistry and Physics* [J], 19(115-138).
- 739 Dal Maso M, Kulmala M, Riipinen I, et al. 2005. Formation and growth of fresh atmospheric aerosols:  
740 Eight years of aerosol size distribution data from SMEAR II, Hyytiälä, Finland. *Boreal*  
741 *Environment Research* [J], 10(323-336).
- 742 Davidovits P, Worsnop D R, Jayne J T, et al. 2004. Mass accommodation coefficient of water vapor on  
743 liquid water. *Geophysical Research Letters* [J], 31(22).

744 Donahue N M, Robinson A L, Stanier C O, et al. 2006. Coupled Partitioning, Dilution, and Chemical  
745 Aging of Semivolatile Organics. *Environmental Science & Technology* [J], 40(8): 2635-2643.

746 Dong C, Matsui H, Spak S, et al. 2019. Impacts of New Particle Formation on Short-term Meteorology  
747 and Air Quality as Determined by the NPF-explicit WRF-Chem in the Midwestern United States.  
748 *Aerosol and Air Quality Research* [J], 19(2): 204-220.

749 Dusek U, Frank G P, Hildebrandt L, et al. 2006. Size Matters More Than Chemistry for Cloud-Nucleating  
750 Ability of Aerosol Particles. *Science* [J], 312(5778): 1375-1378.

751 Ehn M, Thornton J, Kleist E, et al. 2014. A large source of low-volatility secondary organic aerosol.  
752 *Nature* [J], 506(476-479).

753 Fanourgakis G S, Kanakidou M, Nenes A, et al. 2019. Evaluation of global simulations of aerosol particle  
754 and cloud condensation nuclei number, with implications for cloud droplet formation. *Atmos.*  
755 *Chem. Phys.* [J], 19(13): 8591-8617.

756 Gordon H, Kirkby J, Baltensperger U, et al. 2017. Causes and importance of new particle formation in  
757 the present-day and preindustrial atmospheres. 122(16): 8739-8760.

758 Grell G A 1993. Prognostic Evaluation of Assumptions Used by Cumulus Parameterizations. *Monthly*  
759 *Weather Review* [J], 121(3): 764-787.

760 Guo S, Hu M, Zamora M L, et al. 2014. Elucidating severe urban haze formation in China. *Proceedings*  
761 *of the National Academy of Sciences* [J], 111(49): 17373-17378.

762 Hong S-Y, Noh Y, Dudhia J 2006. A New Vertical Diffusion Package with an Explicit Treatment of  
763 Entrainment Processes. *Monthly Weather Review - MON WEATHER REV* [J], 134(

764 Hudson J G, Noble S 2014. CCN and Vertical Velocity Influences on Droplet Concentrations and  
765 Supersaturations in Clean and Polluted Stratus Clouds. *Journal of the Atmospheric Sciences* [J],  
766 71(1): 312-331.

767 Iacono M, Delamere J, Mlawer E, et al. 2008. Radiative Forcing by Long-Lived Greenhouse Gases:  
768 Calculations with the AER Radiative Transfer Models. *Journal of Geophysical Research* [J],  
769 113(

770 Jimenez J L, Canagaratna M R, Donahue N M, et al. 2009. Evolution of Organic Aerosols in the  
771 Atmosphere. *Science* [J], 326(5959): 1525-1529.

772 Kalkavouras P, Bougiatioti A, Kalivitis N, et al. 2019. Regional new particle formation as modulators of  
773 cloud condensation nuclei and cloud droplet number in the eastern Mediterranean. *Atmos. Chem.*  
774 *Phys.* [J], 19(9): 6185-6203.

775 Kerminen V-M, Chen X, Vakkari V, et al. 2018. Atmospheric new particle formation and growth: Review  
776 of field observations. *Environmental Research Letters* [J], 13(

777 Krechmer J E, Day D A, Ziemann P J, et al. 2017. Direct Measurements of Gas/Particle Partitioning and  
778 Mass Accommodation Coefficients in Environmental Chambers. *Environ Sci Technol* [J],  
779 51(20): 11867-11875.

780 Kulmala M, Dada L, Daellenbach K R, et al. 2021. Is reducing new particle formation a plausible solution  
781 to mitigate particulate air pollution in Beijing and other Chinese megacities? *Faraday*  
782 *Discussions* [J], 226(0): 334-347.

783 Kulmala M, L L, Lehtinen K, et al. 2004. Initial steps of aerosol growth. *Atmospheric Chemistry and*  
784 *Physics* [J], 4(

785 Kulmala M, Petäjä T, Ehn M, et al. 2013. Chemistry of Atmospheric Nucleation: On the Recent Advances  
786 on Precursor Characterization and Atmospheric Cluster Composition in Connection with  
787 Atmospheric New Particle Formation. *Annual review of physical chemistry* [J], 65(

788 Kulmala M, Petäjä T, Nieminen T, et al. 2012. Measurement of the nucleation of atmospheric aerosol  
789 particles. *Nature Protocols* [J], 7(9): 1651-1667.

790 Lai S, Hai S, Gao Y, et al. 2022. The striking effect of vertical mixing in the planetary boundary layer on  
791 new particle formation in the Yangtze River Delta. *Science of The Total Environment* [J],  
792 829(154607).

793 Lee S-H, Gordon H, Yu H, et al. 2019. New Particle Formation in the Atmosphere: From Molecular  
794 Clusters to Global Climate. *Journal of Geophysical Research: Atmospheres* [J], 124(  
795 Li K, Zhu Y, Gao H, et al. 2015. A comparative study of cloud condensation nuclei measured between  
796 non-heating and heating periods at a suburb site of Qingdao in the North China. *Atmospheric*  
797 *Environment* [J], 112(40-53).

798 Li M, Liu H, Geng G, et al. 2017. Anthropogenic emission inventories in China: a review. *National*  
799 *Science Review* [J], 4(6): 834-866.

800 Li X, Li Y, Cai R, et al. 2022. Insufficient Condensable Organic Vapors Lead to Slow Growth of New  
801 Particles in an Urban Environment. *Environmental Science & Technology* [J], 56(14): 9936-  
802 9946.

803 Liu H J, Zhao C S, Nekat B, et al. 2014a. Aerosol hygroscopicity derived from size-segregated chemical  
804 composition and its parameterization in the North China Plain. *Atmos. Chem. Phys.* [J], 14(5):  
805 2525-2539.

806 Liu M, Matsui H 2022. Secondary Organic Aerosol Formation Regulates Cloud Condensation Nuclei in  
807 the Global Remote Troposphere. *Geophysical Research Letters* [J], 49(18): e2022GL100543.

808 Liu X, Zhu Y, Zheng M, et al. 2014b. Production and growth of new particles during two cruise  
809 campaigns in the marginal seas of China. *Atmospheric Chemistry and Physics* [J], 14(  
810 Liu X H, Zhu Y J, Zheng M, et al. 2014c. Production and growth of new particles during two cruise  
811 campaigns in the marginal seas of China. *Atmos. Chem. Phys.* [J], 14(15): 7941-7951.

812 Lovejoy E R, Curtius J, Froyd K D 2004. Atmospheric ion-induced nucleation of sulfuric acid and water.  
813 *Journal of Geophysical Research: Atmospheres* [J], 109(D8).

814 Lu H, Wang G, Guo H 2022. Ambient acidic ultrafine particles in different land-use areas in two  
815 representative Chinese cities. *Science of The Total Environment* [J], 830(154774).

816 Lu Y, Yan C, Fu Y, et al. 2019. A proxy for atmospheric daytime gaseous sulfuric acid concentration in  
817 urban Beijing. *Atmos. Chem. Phys.* [J], 19(3): 1971-1983.

818 Lupascu A, Easter R, Zaveri R, et al. 2015. Modeling particle nucleation and growth over northern  
819 California during the 2010 CARES campaign. *Atmos. Chem. Phys.* [J], 15(21): 12283-12313.

820 Ma N, Zhao C, Tao J, et al. 2016. Variation of CCN activity during new particle formation events in the  
821 North China Plain. *Atmos. Chem. Phys.* [J], 16(13): 8593-8607.

822 Matsui H, Koike M, Kondo Y, et al. 2011. Impact of new particle formation on the concentrations of  
823 aerosols and cloud condensation nuclei around Beijing. *Journal of Geophysical Research:*  
824 *Atmospheres* [J], 116(D19).

825 Matsui H, Koike M, Takegawa N, et al. 2013. Spatial and temporal variations of new particle formation  
826 in East Asia using an NPF-explicit WRF-chem model: North-south contrast in new particle  
827 formation frequency. 118(20): 11,647-611,663.

828 Merikanto J, Spracklen D V, Mann G W, et al. 2009. Impact of nucleation on global CCN. *Atmos. Chem.*  
829 *Phys.* [J], 9(21): 8601-8616.

830 Mikkonen S, Romakkaniemi S, Smith J N, et al. 2011. A statistical proxy for sulphuric acid concentration.  
831 *Atmos. Chem. Phys.* [J], 11(21): 11319-11334.

832 Morrison H, Thompson G, Tatarskii V 2009. Impact of Cloud Microphysics on the Development of  
833 Trailing Stratiform Precipitation in a Simulated Squall Line: Comparison of One and Two-  
834 Moment Schemes. *Monthly Weather Review - MON WEATHER REV* [J], 137(991-1007.

835 Nieminen T, Kerminen V M, Petäjä T, et al. 2018. Global analysis of continental boundary layer new  
836 particle formation based on long-term measurements. *Atmos. Chem. Phys.* [J], 18(19): 14737-  
837 14756.

838 Petäjä T, Mauldin I R L, Kosciuch E, et al. 2009. Sulfuric acid and OH concentrations in a boreal forest  
839 site. *Atmos. Chem. Phys.* [J], 9(19): 7435-7448.

840 Pierce J, Riipinen I, Kulmala M, et al. 2011. Quantification of the volatility of secondary organic  
841 compounds in ultrafine particles during nucleation events. *Atmospheric Chemistry and Physics  
842 Discussions* [J], 11(14495-14539.

843 Pöschl U, Canagaratna M, Jayne J T, et al. 1998. Mass Accommodation Coefficient of H<sub>2</sub>SO<sub>4</sub> Vapor on  
844 Aqueous Sulfuric Acid Surfaces and Gaseous Diffusion Coefficient of H<sub>2</sub>SO<sub>4</sub> in N<sub>2</sub>/H<sub>2</sub>O. *The  
845 Journal of Physical Chemistry A* [J], 102(49): 10082-10089.

846 Qiao X, Yan C, Li X, et al. 2021. Contribution of Atmospheric Oxygenated Organic Compounds to  
847 Particle Growth in an Urban Environment. *Environmental Science & Technology* [J], XXXX(  
848 Ren J, Chen L, Fan T, et al. 2021. The NPF Effect on CCN Number Concentrations: A Review and Re-  
849 Evaluation of Observations From 35 Sites Worldwide. *Geophysical Research Letters* [J], 48(19):  
850 e2021GL095190.

851 Riipinen I, Sihto S L, Kulmala M, et al. 2007. Connections between atmospheric sulphuric acid and new  
852 particle formation during QUEST III&ndash;IV campaigns in Heidelberg and Hyytiälä. *Atmos.  
853 Chem. Phys.* [J], 7(8): 1899-1914.

854 Roldin P, Swietlicki E, Massling A, et al. 2011a. Aerosol ageing in an urban plume – implication for  
855 climate. *Atmos. Chem. Phys.* [J], 11(12): 5897-5915.

856 Roldin P, Swietlicki E, Schurgers G, et al. 2011b. Development and evaluation of the aerosol dynamics  
857 and gas phase chemistry model ADCHEM. *Atmos. Chem. Phys.* [J], 11(12): 5867-5896.

858 Saha S, Moorthi S, Wu X, et al. 2014. The NCEP Climate Forecast System Version 2. *Journal of Climate*  
859 [J], 27(6): 2185-2208.

860 Shen X, Sun J, Zhang X, et al. 2018. Comparison of Submicron Particles at a Rural and an Urban Site in  
861 the North China Plain during the December 2016 Heavy Pollution Episodes. *Journal of  
862 Meteorological Research* [J], 32(26-37.

863 Shrivastava M, Fast J, Easter R, et al. 2011. Modeling organic aerosols in a megacity: comparison of  
864 simple and complex representations of the volatility basis set approach. *Atmospheric Chemistry  
865 and Physics* [J], 11(6639-6662.

866 Shrivastava M K, Lane T E, Donahue N M, et al. 2008. Effects of gas particle partitioning and aging of  
867 primary emissions on urban and regional organic aerosol concentrations. 113(D18).

868 Sihto S L, Kulmala M, Kerminen V M, et al. 2006. Atmospheric sulphuric acid and aerosol formation:  
869 implications from atmospheric measurements for nucleation and early growth mechanisms.  
870 *Atmos. Chem. Phys.* [J], 6(12): 4079-4091.

871 Sihto S L, Mikkilä J, Vanhanen J, et al. 2011. Seasonal variation of CCN concentrations and aerosol  
872 activation properties in boreal forest. *Atmos. Chem. Phys.* [J], 11(24): 13269-13285.

873 Tewari M, Wang W, Dudhia J, et al. 2016. Implementation and verification of the united NOAA land  
874 surface model in the WRF model [M].



875 Virtanen A, Kannosto J, Kuuluvainen H, et al. 2011. Bounce behavior of freshly nucleated biogenic  
876 secondary organic aerosol particles. *Atmos. Chem. Phys.* [J], 11(16): 8759-8766.

877 Wang D-W, Guo H, Chan C K 2014. Diffusion Sampler for Measurement of Acidic Ultrafine Particles  
878 in the Atmosphere. *Aerosol Science and Technology* [J], 48(12): 1236-1246.

879 Wang J, Li M, Li L, et al. 2022. Particle number size distribution and new particle formation in Xiamen,  
880 the coastal city of Southeast China in wintertime. *Science of The Total Environment* [J],  
881 826(154208).

882 Wang Y, Wang Y, Song X, et al. 2023. The impact of particulate pollution control on aerosol  
883 hygroscopicity and CCN activity in North China. *Environmental Research Letters* [J],  
884 18(074028).

885 Westervelt D M, Pierce J R, Riipinen I, et al. 2013. Formation and growth of nucleated particles into  
886 cloud condensation nuclei: model–measurement comparison. *Atmos. Chem. Phys.* [J], 13(15):  
887 7645-7663.

888 Wu H, Li Z, Jiang M, et al. 2021a. Contributions of traffic emissions and new particle formation to the  
889 ultrafine particle size distribution in the megacity of Beijing. *Atmospheric Environment* [J],  
890 262(118652).

891 Wu L, Ling Z, Shao M, et al. 2021b. Roles of Semivolatile/Intermediate-Volatility Organic Compounds  
892 on SOA Formation Over China During a Pollution Episode: Sensitivity Analysis and  
893 Implications for Future Studies. *Journal of Geophysical Research: Atmospheres* [J], 126(8):  
894 e2020JD033999.

895 Wu Z, Hu M, Yue D, et al. 2011. Evolution of particle number size distribution in an urban atmosphere  
896 during episodes of heavy pollution and new particle formation. *Science China Earth Sciences*  
897 [J], 54(11): 1772.

898 Yao L, Garmash O, Bianchi F, et al. 2018. Atmospheric new particle formation from sulfuric acid and  
899 amines in a Chinese megacity. *Science* [J], 361(278-281).

900 Yu F 2005. Quasi-unary homogeneous nucleation of H<sub>2</sub>SO<sub>4</sub>-H<sub>2</sub>O. *The Journal of Chemical Physics* [J],  
901 122(7): 074501.

902 Yu F, Luo G, Nair A A, et al. 2020. Wintertime new particle formation and its contribution to cloud  
903 condensation nuclei in the Northeastern United States. *Atmos. Chem. Phys.* [J], 20(4): 2591-  
904 2601.

905 Yu F, Luo G, Pryor S C, et al. 2015. Spring and summer contrast in new particle formation over nine  
906 forest areas in North America. *Atmos. Chem. Phys.* [J], 15(24): 13993-14003.

907 Yuan Q, Li W, Zhou S, et al. 2015. Integrated evaluation of aerosols during haze-fog episodes at one  
908 regional background site in North China Plain. *Atmospheric Research* [J], 156(102-110).

909 Yue D L, Hu M, Zhang R Y, et al. 2011. Potential contribution of new particle formation to cloud  
910 condensation nuclei in Beijing. *Atmospheric Environment* [J], 45(33): 6070-6077.

911 Zaveri R A, Easter R C, Fast J D, et al. 2008. Model for Simulating Aerosol Interactions and Chemistry  
912 (MOSAIC). 113(D13).

913 Zaveri R A, Easter R C, Peters L K 2005. A computationally efficient Multicomponent Equilibrium  
914 Solver for Aerosols (MESA). *Journal of Geophysical Research: Atmospheres* [J], 110(D24).

915 Zhang Q, Jimenez J L, Canagaratna M R, et al. 2007. Ubiquity and dominance of oxygenated species in  
916 organic aerosols in anthropogenically-influenced Northern Hemisphere midlatitudes.  
917 *Geophysical Research Letters* [J], 34(13).

918 Zhao B, Shrivastava M, Donahue N M, et al. 2020. High concentration of ultrafine particles in the  
919 Amazon free troposphere produced by organic new particle formation. Proceedings of the  
920 National Academy of Sciences [J], 117(41): 25344-25351.

921 Zhao B, Wang S, Donahue N M, et al. 2016. Quantifying the effect of organic aerosol aging and  
922 intermediate-volatility emissions on regional-scale aerosol pollution in China. Scientific  
923 Reports [J], 6(1): 28815.

924 Zheng B, Tong D, Li M, et al. 2018. Trends in China's anthropogenic emissions since 2010 as the  
925 consequence of clean air actions. Atmos. Chem. Phys. [J], 18(19): 14095-14111.

926 Zhu Y, Li K, Shen Y, et al. 2019. New particle formation in the marine atmosphere during seven cruise  
927 campaigns. Atmos. Chem. Phys. [J], 19(1): 89-113.

928 Zhu Y, Sabaliauskas K, Liu X, et al. 2014. Comparative analysis of new particle formation events in less  
929 and severely polluted urban atmosphere. Atmospheric Environment [J], 98(655-664).

930 Zhu Y, Shen Y, Li K, et al. 2021. Investigation of Particle Number Concentrations and New Particle  
931 Formation With Largely Reduced Air Pollutant Emissions at a Coastal Semi-Urban Site in  
932 Northern China. Journal of Geophysical Research: Atmospheres [J], 126(17): e2021JD035419.

933

934