

36 Substantially positive contributions of new particle formation to Cloud Condensation
37 Nuclei under low supersaturation in China based on numerical model improvements

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Abstract

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New particle formation (NPF) and subsequent particle growth are important sources of condensation nuclei (CN) and cloud condensation nuclei (CCN). While many observations have shown positive contributions of NPF to CCN at low supersaturation, negative NPF contributions were often simulated in polluted ~~environment~~environments.

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Using the observations in a ~~typical~~ coastal city of Qingdao, Beijing and Gucheng in North China,

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we thoroughly evaluate the simulated number concentrations of CN and CCN using a NPF-explicit parameterization embedded in WRF-Chem model. ~~In terms~~

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~~of~~For CN, the initial simulation shows large biases of particle number concentrations

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at 10–40 nm (~~CN₁₀₋₄₀~~) and 40–100 nm (~~CN₄₀₋₁₀₀~~). By adjusting the process of gas-

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particle partitioning, including the mass accommodation coefficient of sulfuric acid, the

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phase changes of primary organic aerosol emissions and the condensational amount of

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nitric acid, the ~~concomitant~~ improvement of the particle growth process yields a

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~~substantial reduction of~~substantially reduced overestimates of ~~CN₁₀₋₄₀ and CN₄₀₋₁₀₀~~CN.

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Regarding CCN, SOA formed from the oxidation of semi-volatile and intermediate

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volatility organic vapors (SI-SOA) yield is an important contributor. ~~In~~At default

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settings, the ~~original WRF-Chem model with 20 size bins setting, the yield of~~ SI-SOA

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yield is too high without considering the differences in precursor oxidation rates ~~of the~~

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~~precursors~~. Lowering the SI-SOA yield under linear-H₂SO₄ nucleation scheme results

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in much improved CCN simulations ~~of the observed CCN concentrations compared to~~

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observations. On the basis of the bias-corrected model, we find ~~substantial~~substantially

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positive contributions of NPF to CCN at low supersaturation (~0.2%) ~~in Qingdao and~~

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over the broad areas of China, primarily due to ~~the~~ competing effects of increasing

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particle hygroscopicity, a result of reductions in SI-SOA amount, surpassing that of

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particle size ~~decrease~~decreases. The bias-corrected model is robustly applicable to other

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schemes, such as quadratic-H₂SO₄ nucleation scheme, in terms of CN and CCN, though

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the dependence of CCN on SI-SOA yield is diminished likely due to changes in particle

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composition. This study highlights the potentially much larger NPF contributions to

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CCN on a regional and even global basis.

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1. Introduction

New particle formation (NPF) is a process in which gaseous vapors nucleate and form critical molecular clusters, followed by subsequent growth to larger sizes through condensation and coagulation (~~Kulmala et al., 2004; Kulmala et al., 2013; Lee et al., 2019~~)(Kulmala et al., 2004; Kulmala et al., 2013; Lee et al., 2019). Newly formed particles could effectively grow into the size of cloud condensation nuclei (CCN) under certain supersaturation (SS), which exerts an impact on the cloud microphysical process and global radiation balance (~~Merikanto et al., 2009; Gordon et al., 2017; Kerminen et al., 2018; Ren et al., 2021~~)(Merikanto et al., 2009; Gordon et al., 2017; Kerminen et al., 2018; Ren et al., 2021). In addition, the efficient nucleation and explosive growth of particles may contribute to the formation of haze (~~Guo et al., 2014~~)(Guo et al., 2014), affecting air quality and human health (~~Yuan et al., 2015; Chu et al., 2019; Kulmala et al., 2021~~)(Yuan et al., 2015; Chu et al., 2019; Kulmala et al., 2021).

The overestimate of condensation nuclei (CN) in numerical models ~~are~~is commonly seen, despite the attempt to rectify the bias (~~Matsui et al., 2013; Arghavani et al., 2022~~)(Matsui et al., 2013; Arghavani et al., 2022). It is a common way to reduce the nucleation rate which may reduce the particle number concentration in proportion (~~Matsui et al., 2013~~)(Matsui et al., 2013). For instance, in the study of NPF in East Asia in the spring of 2009, even after lowering the nucleation rate in a regional model of WRF-Chem applied in their study, the reduced number concentration of particles at 10–130 nm remained to be overestimated (~~Matsui et al., 2013~~)(Matsui et al., 2013). Using the same regional model and a similar method to reduce the nucleation rate, ~~Arghavani et al. (2022)~~Arghavani et al. (2022) found particle number concentration at 10–100 nm was still overestimated by nearly one order of magnitude, despite the effectiveness to

131 reduce the overestimates for the smaller particles such as 2.5–10 nm. In addition to the
132 rate of NPF, the growth process of particles also has a crucial effect on particle number
133 concentration and size distribution. In this process, the condensation of some chemical
134 species such as sulfuric acid, nitrate and organic gases on particles plays a major role
135 in particle growth (~~Yao et al., 2018; Lee et al., 2019; Li et al., 2022~~)([Yao et al., 2018;](#)
136 [Lee et al., 2019; Li et al., 2022](#)), and the uncertainty of their condensation amount may
137 lead to the bias of CN simulation.

138 In addition to CN, there are large discrepancies in the predicted CCN between the
139 numerical models and observational results. Furthermore, as an important source of
140 CCN (~~Merikanto et al., 2009~~)([Merikanto et al., 2009](#)), the contribution of nucleation to
141 CCN quantified by numerical models is also highly uncertain. For example, in terms of
142 predicting CCN, ~~Fanourgakis et al. (2019)~~[Fanourgakis et al. \(2019\)](#) evaluated the CCN
143 concentrations simulated by 16 global aerosol–climate and chemistry transport models
144 with observations at 9 sites in Europe and Japan from 2011 to 2015, and found that all
145 models underestimated CCN concentrations with a mean normalized mean bias of -36%
146 at low supersaturation (SS=0.2%). WRF-Chem models also tend to underestimate the
147 contribution of NPF on CCN, especially at low supersaturation. The continuous
148 observation of CCN concentrations throughout the year (July 2008–June 2009) carried
149 out in Hyytiälä, Finland, showed that under low SS, nucleation enhanced the CCN by
150 106% and 110% at SS=0.1% and 0.2% respectively (~~Sihto et al., 2011~~)([Sihto et al.,](#)
151 [2011](#)). Observations acquired in Beijing from July 12 to September 25, 2008, also
152 suggested that nucleation significantly increases CCN at all supersaturations, even
153 when supersaturation is low (i.e., 0.07% and 0.26%). Thus, the occurrence of NPF
154 enhanced CCN by a factor of 1.7 and 2.2, respectively (~~Yue et al., 2011~~)([Yue et al.,](#)
155 [2011](#)).

156 ~~However, previous numerical experiments behave oppositely. For instance, Matsui~~
157 ~~et al. (2011) quantified the contribution of nucleation to CCN using WRF-chem in~~
158 ~~Beijing in August and September 2006 and found reduced CCN under low SS, e.g.,~~
159 ~~when SS=0.02%, the concentration of CCN is reduced by up to ~50%. They attributed~~
160 ~~this to the fact that the small particles produced by nucleation may inhibit the growth~~

161 of the preexisting particles (Matsui et al., 2011). Similarly, Dong et al. (2019) conducted
162 NPF simulations with the WRF-Chem for the summer of 2008 focusing on the Midwest
163 of the United States, and found that the nucleation resulted in decreased CCN at low
164 supersaturation ($SS=0.1\%$). Besides, a study carried out for East Asia in 2009 also
165 indicated that at low supersaturation (e.g. $SS=0.1\%$), nucleation has little impact on
166 CCN (Matsui et al., 2013). The contrasting effects of nucleation on CCN at low
167 supersaturations in model and observations is not explained in these previous studies.

168 At the stage of particle growth, secondary organic aerosol (SOA) formed by
169 atmospheric oxidation of organic vapors is a major contributor to particle growth to
170 CCN-related sizes. However, previous numerical experiments behave oppositely. For
171 instance, Matsui et al. (2011) quantified the contribution of nucleation to CCN using
172 WRF-chem in Beijing in August and September 2006 and found reduced CCN under
173 low SS, e.g., when $SS=0.02\%$, the concentration of CCN is reduced by up to $\sim 50\%$.
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175 inhibit the growth of the preexisting particles (Matsui et al., 2011). Similarly, Dong et
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182 studies.

183 At the stage of particle growth, secondary organic aerosol (SOA) formed by
184 atmospheric oxidation of organic vapors is a major contributor to particle growth to
185 CCN-related sizes (Liu and Matsui, 2022; Qiao et al., 2021)(Liu and Matsui, 2022;
186 Qiao et al., 2021). SOA formed by multi-generational gas-phase oxidation of semi-
187 volatile and intermediate volatility organic compounds (S/IVOC) is called SI-SOA
188 (Jimenez et al., 2009; Zhang et al., 2007)(Jimenez et al., 2009; Zhang et al., 2007). Zhao
189 et al. (2016). Zhao et al. (2016) made a comprehensive assessment of the roles of
190 various SOA precursors in SOA formation in real atmosphere in China in 2010, and the

191 results demonstrated that evaporated POA and IVOC (i.e. S/IVOC) made a significant
192 contribution to SOA, contributing up to 82% to the average SOA concentration in
193 eastern China. However, the effect of SI-SOA on CCN has not been fully studied.

194 In this paper, WRF-Chem was applied to simulate the effect of the NPF on CCN
195 in [QingdaoChina](#) in February 2017. The simulated results from the WRF-Chem model
196 are firstly compared with observations in Qingdao, [Beijing and Gucheng](#), exhibiting
197 large biases in CN. This is followed by an improvement through a few processes. At
198 the end, the impact of SI-SOA yield and nucleation on CCN is investigated.

199 **2. Data and methods**

200 **2.1 Observations**

201 The measurements used in this study were carried out over the sampling site from
202 February 5 to 24, 2017 at the campus of Ocean University of China (36°09'37"N,
203 120°29'44"E-) in Qingdao, which is surrounded by residential buildings and is situated
204 about 10 km away from the city center. A fast mobility particle sizer (FMPS, TSI Model
205 3091) was applied to measure the aerosol particle size distribution for the size range of
206 5.6 nm to 560 nm ([Liu et al., 2014b](#))([Liu et al., 2014b](#)). The bulk CCN concentration is
207 measured by a cloud condensation nuclei counter at three different supersaturations
208 (0.2%, 0.4% and 0.6%) and each supersaturation lasts for 20 minutes. More information
209 about the CCN measurement can be found in [Li et al. \(2015\)](#)[Li et al. \(2015\)](#). The urban
210 site in Beijing is located on the roof of the building of the Chinese Academy of
211 Meteorological Sciences (CAMS, 39°95'N, 116°33'E) in the campus of the China
212 Meteorological Administration, close to the main road with heavy traffic. The rural site
213 is Gucheng (GC, 39°08'N, 115°40'E), located in Hebei Province, surrounded by
214 farmland, and is a representative station of the severity of air pollution in Beijing Tianjin
215 Hebei region. The particle number size distribution of these two sites in the range of 4–
216 850 nm is measured by a Tandem Scanning Mobility Particle Sizer (TSMPS), and more
217 information about the instruments can be found in [Shen et al. \(2018\)](#)[Shen et al. \(2018\)](#).

218 **2.2 Model configurations**

219 WRF-Chem version 3.9 is used to simulate NPF events, with the main physical
220 and chemical parameterization settings summarized in Table 1. The spatial resolution

221 is 36 km by 36 km with 35 vertical layers and a model top at 50 hPa. The regional
 222 model simulations at a higher spatial resolution may be desirable in future when urban
 223 pollution is focused. A continuous run from February 1 to 25, 2017, was conducted,
 224 with the first five-day results as the spin-up and discarded in the analysis.

225 Table 1 WRF-Chem model configurations used in this work

	Model configuration
Microphysics	Morrison 2-moment microphysics scheme (Morrison et al., 2009) <u>Morrison 2-moment microphysics scheme (Morrison et al., 2009)</u>
Planetary Boundary Layer (PBL)	YSU boundary layer scheme (Hong et al., 2006) <u>YSU boundary layer scheme (Hong et al., 2006)</u>
Longwave and Shortwave Radiation	RRTMG longwave and shortwave radiation (Iacono et al., 2008) <u>(Iacono et al., 2008)</u>
Land model	Unified Noah Land Surface scheme (Chen and Dudhia, 2000; Tewari et al., 2016) <u>Unified Noah Land Surface scheme (Chen and Dudhia, 2000; Tewari et al., 2016)</u>
Cumulus	Grell 3D cumulus parameterization scheme (Grell, 1993) <u>Grell-3D cumulus parameterization scheme (Grell, 1993)</u>
Aerosol module	MOSAIC module (Zaveri et al., 2008; Matsui et al., 2011) <u>(Zaveri et al., 2008; Matsui et al., 2011)</u>
Gas-phase Chemistry	SAPRC-99 gas-phase chemistry scheme (Carter, 2000) <u>SAPRC-99 gas-phase chemistry scheme (Carter, 2000)</u>

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227 The meteorological initial and boundary conditions are driven by Climate Forecast
 228 System model version 2 (CFSv2; ~~(Saha et al., 2014)~~(Saha et al., 2014)) reanalysis
 229 developed by National Centre for Environmental Prediction (NCEP). The initial and
 230 boundary chemical conditions of WRF-Chem are provided by Community Atmosphere
 231 Model with Chemistry (CAM-Chem; ~~(Buchholz et al., 2019)~~(Buchholz et al., 2019)).

232 Anthropogenic emissions for the year of 2017 are obtained from the Multiresolution
233 Emission Inventory for China (MEIC,
234 <http://www.meicmodel.org/>) emission dataset (Li et al.,
235 2017; Zheng et al., 2018)(Li et al., 2017; Zheng et al., 2018).

236 The Model for Simulating Aerosol Interactions and Chemistry (MOSAIC) was
237 used to delineate dynamic gas-particle mass transfer to represent the condensation
238 growth of aerosol (Zaveri et al., 2008)(Zaveri et al., 2008). The gas-particle partitioning
239 of gas species on particles is regulated by the mass transfer rate, which is related to
240 mass accommodation coefficient (α), a parameter involved in the model representing
241 the probability of gas molecules entering the bulk liquid phase (Pöschl et al.,
242 1998)(Pöschl et al., 1998). The original setting of α for all condensing species for all
243 size bins a in MOSAIC is 0.1 (Zaveri et al., 2008)(Zaveri et al., 2008). In the default
244 release of WRF-Chem, MOSAIC was implemented in the sectional framework with
245 aerosol size distributions divided into 4 or 8 size bins spanning 39 nm to 10 μm in
246 diameter. To explicitly express the nucleation and the growth of newly formed particles,
247 the aerosol size range in the MOSAIC module was extended from 1 nm to 10 μm , with
248 the number of aerosol size bins increased to 20 (Matsui et al., 2011; Matsui et al., 2013;
249 Lupascu et al., 2015; Lai et al., 2022)(Matsui et al., 2011; Matsui et al., 2013; Lupascu
250 et al., 2015; Lai et al., 2022). The calculation method of CCN concentration in the
251 WRF-chem model is referred to the study of Matsui et al. (2011)Matsui et al. (2011).
252 Based on Köhler theory, CCN concentrations under the three given supersaturations of
253 0.2%, 0.4% and 0.6% were calculated. The critical supersaturation (S_c) of each size bin
254 in the WRF-chem model was calculated by the following formula:

$$255 \quad S_c = \sqrt{\frac{4 \times a^3}{27 \times r^3 \times \kappa}} \quad (1)$$

$$256 \quad a = \frac{2 \times \sigma}{R_v \times T \times \rho_w} \quad (2)$$

257 Where α (m) is the coefficient of the Kelvin effect, κ is the volume-averaged
258 hygroscopicity, calculated using these values in Table 1, r (m) is the dry diameter, σ is
259 droplet surface tension over water (0.076 N m⁻¹), R_v is the gas constant for water vapor

260 (461.6 J K⁻¹kg⁻¹), T (K) is the air temperature, and ρ_w is the density of water (1000 kg
261 m⁻³).

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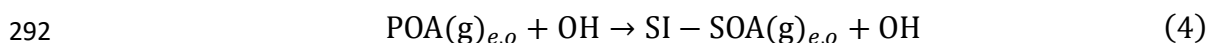
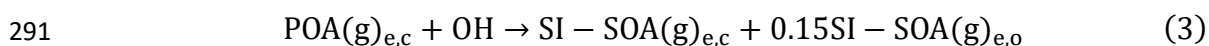
Table 2 Hygroscopicity Parameters (κ) in the WRF-Chem Model

Species	Hygroscopicity (κ)
Sulfate	0.5
Ammonium	0.5
Nitrate	0.5
Black carbon	10 ⁻⁶
Primary organic aerosol	0.14
Other inorganics	0.14
Sodium	1.16
Chloride	1.16

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267 The chemical aging process of organic aerosols (OA) is modeled by the volatility
268 basis set (VBS) approach, which was widely used in air quality models to represent
269 complex mixtures of thousands of organic species (Donahue et al., 2006; Shrivastava
270 et al., 2011; Chrit et al., 2018)(Donahue et al., 2006; Shrivastava et al., 2011; Chrit et
271 al., 2018). The VBS method classifies compounds according to the effective saturation
272 concentration (c^*), which represents the proportion of the component in the gas phase
273 to the particle phase (Donahue et al., 2006)(Donahue et al., 2006), and species with
274 higher c^* values are more volatile. The oxidation of highly volatile precursors to form
275 relatively low volatile components represents the aging process of OA. OA consists of
276 directly emitted primary organic aerosols and photochemically produced secondary
277 organic aerosols (SOA) (Shrivastava et al., 2011)(Shrivastava et al., 2011). In this study,
278 the simplified 2-species VBS mechanism was applied to the simulation of SOA, during
279 which primary organic aerosol was represented by two species based on volatility with
280 effective saturation concentration c^* values (at 298 K and 1 atm) of 10⁻² and 10⁵ μg

281 m^{-3} (~~Shrivastava et al., 2011~~)([Shrivastava et al., 2011](#)). Primary organic aerosols with
 282 c^* of $10^5 \mu\text{g m}^{-3}$ refers to S/IVOC, which is in the gas phase under most atmospheric
 283 conditions due to its high volatility, while for those primary organic matters with c^* of
 284 $10^{-2} \mu\text{g m}^{-3}$, is treated as gas phase as well in the original model. The SOA formed by
 285 photochemical oxidation of S/IVOC precursors is called SI-SOA and the SOA formed
 286 by oxidation of VOC precursors is named V-SOA. In the simplified 2-species VBS
 287 mechanism, SI-SOA (c^* of $10^{-2} \mu\text{g m}^{-3}$) is formed by the oxidation reaction of S/IVOC
 288 precursors (c^* of $10^5 \mu\text{g m}^{-3}$) and OH with an oxidation rate constant of $4 \times 10^{-11} \text{ cm}^3$
 289 $\text{molec}^{-1} \text{ s}^{-1}$. The equations for controlling the oxidation of S/IVOC precursors are as
 290 follows:



293 where POA(g) denotes primary organic aerosols with c^* of $10^5 \mu\text{g m}^{-3}$, which reacts
 294 with OH to form SI-SOA(g) with c^* of $10^{-2} \mu\text{g m}^{-3}$. Subscripts c and o represent the
 295 non-oxygen and oxygen parts respectively of given species and e is either the biomass
 296 or anthropogenic emission sector. In addition, SVOC and IVOC emissions
 297 corresponding to both anthropogenic and biomass burning emissions are derived based
 298 on constant emission ratio of S/IVOC to POA (~~Shrivastava et al., 2011~~). ~~A detailed~~
 299 ~~description of 2-species VBS mechanism can be found in Shrivastava et al.~~
 300 ~~(2011)~~([Shrivastava et al., 2011](#)). ~~A detailed description of 2-species VBS mechanism~~
 301 ~~can be found in Shrivastava et al. (2011).~~

302 2.3 Model sensitivity formulations

303 Three sets of sensitivity tests are designed and listed in Table3. The purposes of
 304 the three sets of experiments are as follows: (1) Adjust the condensation growth process
 305 of ultrafine particles in WRF-Chem model (Base, MAC, ~~POAPEP~~, NOCD, RACD,
 306 with details in Table 3).; (2) Explore the effect of SI-SOA yield on CCN (Low-Yield
 307 and High_Yield); (3) Study the effect of nucleation process on CCN under the change
 308 of SI-SOA yield (Low-Yield and High_Yield and their corresponding cases without
 309 nucleation parameterization, i.e., Low_nucoff and High_nucoff). Each scenario will be
 310 explained in conjunctions with the results.

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Table 3 The sensitivity tests involved in this study

Purposes	Simulation scenarios	Description
Adjust the condensation growth process of ultrafine particles	Base	Simulation with the default setting with nucleation coefficient set as $2 \times 10^{-6} \text{ s}^{-1}$, the same as Lai et al. (2022) Simulation with the default setting with nucleation coefficient set as $2 \times 10^{-6} \text{ s}^{-1}$, the same as Lai et al. (2022)
	Mass accommodation coefficient (MAC)	It is the same as Base except that the mass adjustment coefficient (α) of gaseous sulfuric acid is adjusted from 0.1 to 0.65.
	POA emission phase (PEP)	It is the same as MAC except that the phase of POA is changed from gas phase to particle phase.
	No condensation (NOCD)	It is the same as PEP except that no NH_4NO_3 condenses on particles below 40 nm.
Ratio method for condensation		It is the same as PEP except that the condensation of NH_4NO_3 on

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	(RACD)	particles below 40 nm is reduced according to the ratio of acid particles to total particles reported in Wang et al. (2014). It is the same as PEP except that the condensation of NH₄NO₃ on particles below 40 nm is reduced according to the ratio of acid particles to total particles reported in Wang et al. (2014).
Explore the effect of SI-SOA yield on CCN (Explore the effect of nucleation process on CCN under the change of SI-SOA yield)	High_Yield	Simulation with high oxidation rate of SI-SOA formation with reaction rate constant of $5 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$
	Low-_Yield	Simulation with low oxidation rate of SI-SOA formation with reaction rate constant of $2 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$
Explore the effect of nucleation process on CCN under the change of SI-SOA yield	High_NUCOFF	Simulations without nucleation parameterizations based on High_Yield
	Low_ NUCOFF	Simulations without nucleation parameterizations based on Low-_Yield

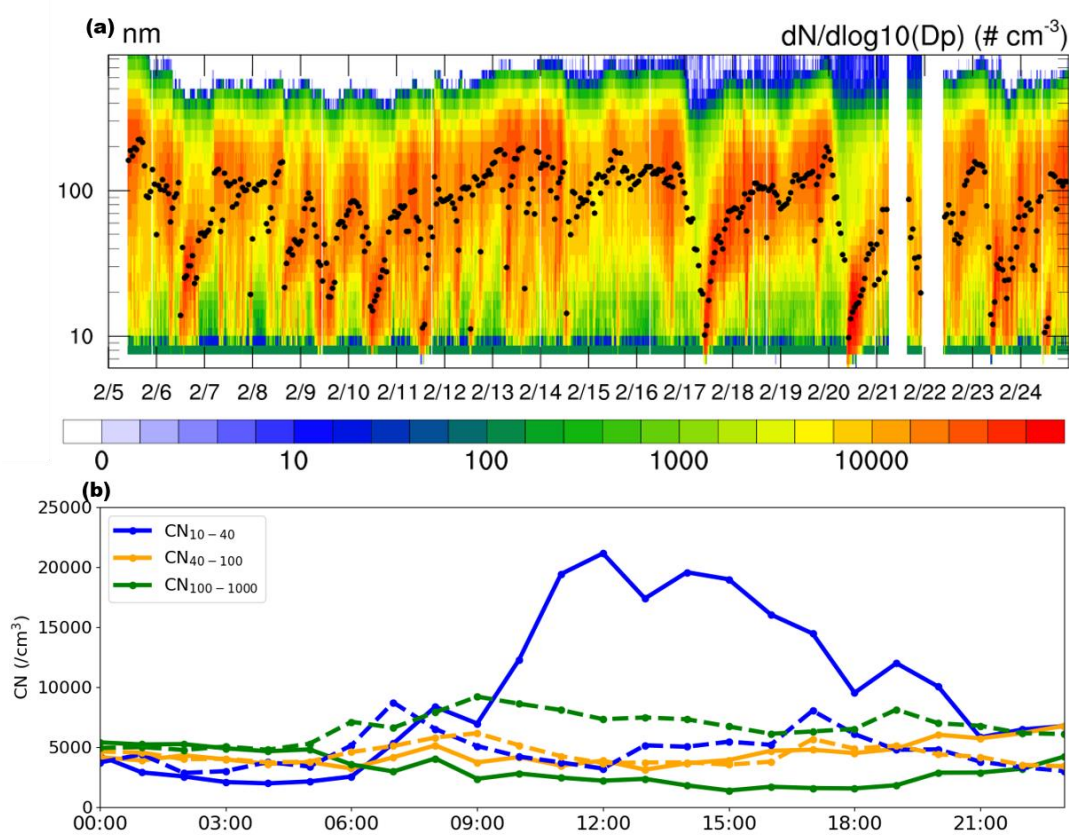
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336 3. Results

337 3.1 Observational analysis

338 Based on the criteria (~~Dal Maso et al., 2005; Kulmala et al., 2012~~)(Dal Maso et al.,
 339 2005; Kulmala et al., 2012), NPF is defined as an event with the emergence of a
 340 nucleation mode with particle diameters smaller than 25 nm, lasting for 2 hours or more,
 341 followed in general by a continuous particle growth. Six NPF events were identified in
 342 February 2017 in Qingdao, on the days of 6, 9, 10, 17, 20 and 23 (Fig. 1a), yielding a

343 frequency of ~30% and displaying a typical banana-shaped growth of particles in the
 344 particle number size distribution. Compared to a few other studies on NPF frequency
 345 in Qingdao, the results in this study are to a large extent consistent with that in the fall
 346 of 2012–2013 (30%; (Zhu et al., 2019)(Zhu et al., 2019)), slightly higher than that in
 347 summer 2016 (22%; (Zhu et al., 2019)(Zhu et al., 2019)) and lower than that in spring
 348 of 2010 (41%; (Liu et al., 2014c)(Liu et al., 2014c)). The higher frequency in spring in
 349 Qingdao is consistent with the observational results at different stations in the Northern
 350 Hemisphere in Nieminen et al. (2018)Nieminen et al. (2018).



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 352 Fig. 1. Distribution of particle number concentration. (a) Temporal evolution of particle
 353 size distributions (colored shading) and geometric median diameter (GMD; dots in
 354 black) in Qingdao on February 5-24, 2017. (b) The mean diurnal variation of CN₁₀₋₄₀
 355 (blue), CN₄₀₋₁₀₀ (orange) and CN₁₀₀₋₁₀₀₀ (green) composited during the NPF (solid lines)
 356 and non-NPF (dashed lines) days on February 5-24, 2017. All times are local times (LT)
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358 During the six NPF events identified in February in Qingdao, the mean diurnal
 359 cycle of CN₁₀₋₄₀ (10–40 nm) particles exhibits triple peaks (solid blue in Fig. 1b), in the

360 morning (8:00 LT), noon (12:00–14:00 LT) and evening (19:00 LT), respectively. A
361 comparable three-peak feature was also observed in earlier years during 2016–2018 in
362 Qingdao ([Zhu et al., 2021](#))([Zhu et al., 2021](#)). The morning and evening peaks of CN_{10–}
363 ₄₀, with values of ~5300 cm⁻³ and ~12000 cm⁻³, respectively, are likely caused by the
364 primary emissions from traffic and cooking activities ([Wu et al., 2021a](#); [Wang et al.,](#)
365 [2022](#); [Cai et al., 2020](#))([Wu et al., 2021a](#); [Wang et al., 2022](#); [Cai et al., 2020](#)). The
366 occurrence of NPF starts approximately at 9:00 am LT, accompanied by a substantial
367 increase in CN_{10–40} compared with non-NPF days (solid vs. dashed lines, in blue),
368 yielding a peak around noon (20000 cm⁻³ during 12:00–14:00 LT). In addition, larger
369 particles (e.g., CN_{40–100} and CN_{100–1000}) displayed a slow or no increase in the afternoon.

370

371 **3.2 Model improvement in particle number concentration simulations**

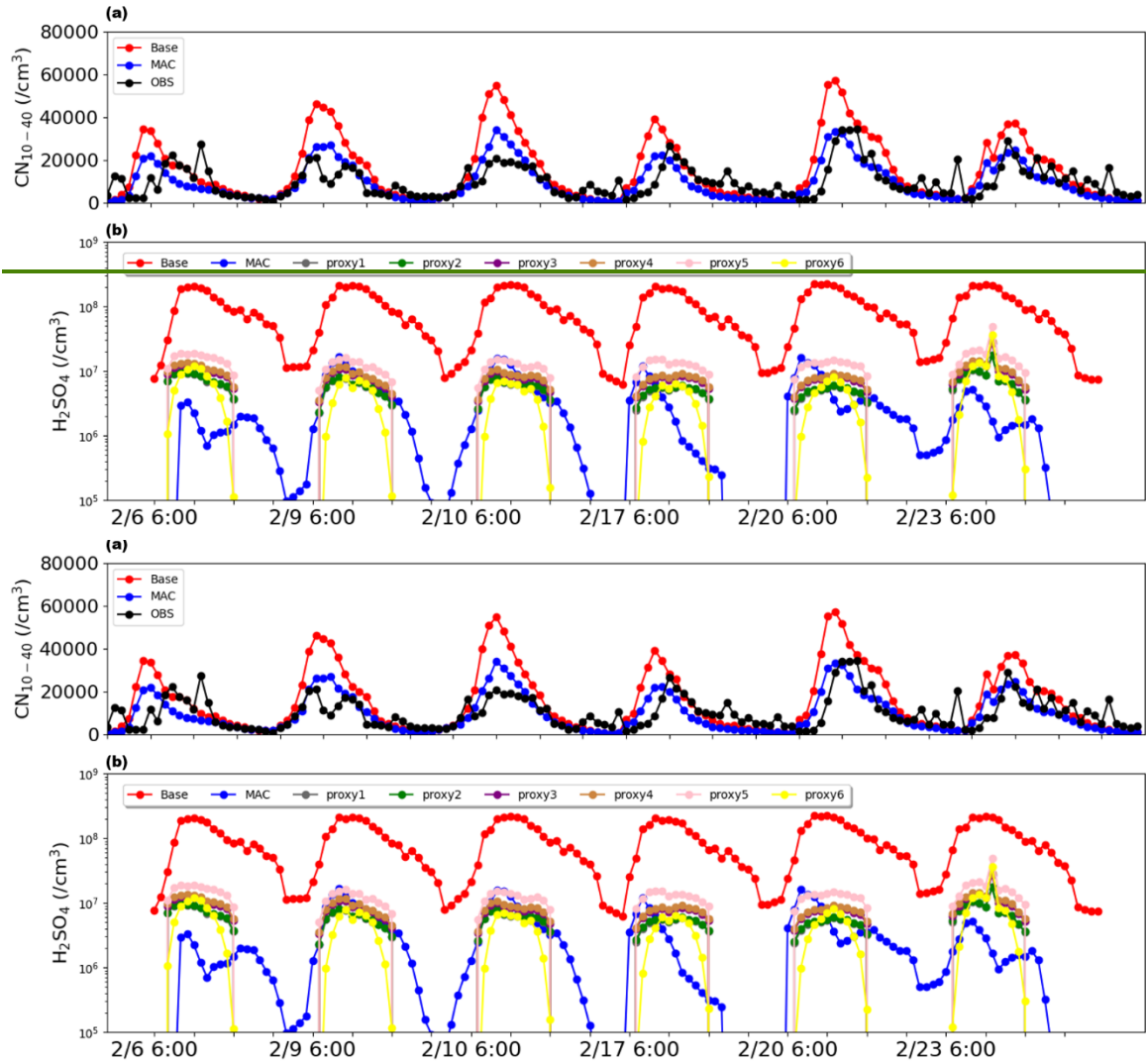
372 Particle number concentrations, primarily in two ranges of 10–40 nm and 40–100
373 nm, are commonly simulated with large biases. In the smaller size range (10–40 nm),
374 the particle number concentration is associated with NPF and particle growth. During
375 NPF, despite differences among the formation mechanisms, H₂SO₄ is considered the
376 common species ([Yu, 2005](#); [Lovejoy et al., 2004](#))([Yu, 2005](#); [Lovejoy et al., 2004](#)),
377 which often suffer large biases ([Cai et al., 2016](#); [Matsui et al., 2011](#))([Cai et al., 2016](#);
378 [Matsui et al., 2011](#)). In the size range of 40–100 nm, the particle number concentration
379 is primarily affected by the condensation growth of particles below 40 nm, which is
380 closely related to chemical components such as SOA and nitrate. Prior to the evaluation
381 of particle number concentration, we first evaluate the compositions of PM_{2.5} and
382 criteria air pollutants including PM_{2.5}, PM₁₀, O₃, SO₂, CO, and NO₂, showing relatively
383 low biases compared to observations (section S1 and Fig. S1 and Fig. S2 of the
384 supporting information).

385

386 **3.2.1 Bias correction of particle number concentration at 10–40 nm**

387 In this study, as shown in Fig. 2, comparisons of CN_{10–40} between simulations (red
388 line in Fig. 2a) and observations (black line in Fig. 2a) results of the six NPF events
389 mentioned in the previous section in Qingdao in February 2017 indicate that model

390 overestimates CN_{10-40} with mean fractional bias of 48%. As one of the major processes
391 affecting the particle number concentration of 10–40 nm, nucleation is governed by the
392 particle nucleation rate of 1 nm particles ($\text{cm}^{-3} \text{s}^{-1}$), which is closely associated with the
393 concentration of H_2SO_4 . For instance, in a commonly applied activation mechanism,
394 the nucleation rate calculated by $J^* = K_{\text{ACT}} \times [\text{H}_2\text{SO}_4]$. Note that K_{ACT} is the nucleation
395 coefficient considering the physical properties and chemical species of nucleation
396 process under different environments, indicating that a lumped chemical species are
397 included in the scheme reflected primarily in the nucleation coefficient k , set as 2×10
398 $^{-6} \text{s}^{-1}$ based on previous studies (~~Sihto et al., 2006; Riipinen et al., 2007~~)([Sihto et al.,](#)
399 [2006; Riipinen et al., 2007](#))-~~Dong et al. (2019)~~. [Dong et al. \(2019\)](#) simulated NPF
400 occurring in the summer of 2008 in the United States using the NPF-explicit WRF-
401 Chem based on the activation mechanism, which overestimated the particle number
402 concentration at 10–63 nm by nearly doubled, even when the K_{ACT} decreased by one
403 order of magnitude (set at a very low value of 10^{-7}s^{-1}). Therefore, it is likely that the
404 overestimation of particle number concentration in the smaller particle size segment is
405 probably due to the bias of simulated sulfuric acid.



408 Fig. 2. Time series of (a) CN_{10-40} on NPF days, where red and blue represent Base and
 409 MAC simulation results respectively, and black represents observation results, and (b)
 410 sulfuric acid gas concentration obtained by simulation and by proxies (dark grey: Eq.
 411 5; green: Eq. 6; purple: Eq. 7; brown: Eq. 8; pink: Eq. 9; yellow: Eq. 10). All times are
 412 in local times.

413
 414 Measurement of sulfuric acid gases in the lower troposphere is challenging due to
 415 the generally low ambient concentration of sulfuric acid (10^6-10^7 molecule cm^{-3}).
 416 Different methods have been proposed to estimate ambient sulfuric acid concentrations
 417 based on observations such as SO_2 (Petäjä et al., 2009; Lu et al., 2019; Mikkonen et al.,
 418 2011)(Petäjä et al., 2009; Lu et al., 2019; Mikkonen et al., 2011). For instance, Petäjä
 419 et al. (2009). For instance, Petäjä et al. (2009) proposed a linear method to approximate

420 observed H₂SO₄ concentration in ~~Hyttiala, southern Finland. Moreover, a recent study~~
 421 ~~by Lu et al. (2019)~~Hyttialä, southern Finland. Moreover, a recent study by Lu et al.
 422 (2019) proposed a nonlinear method to construct a number of proxies for gaseous
 423 sulfuric acid concentration (Eq. 5–9), indicating that compared to the linear method in
 424 ~~Petäjä et al. (2009)~~Petäjä et al. (2009), the nonlinear relationship can provide more
 425 accurate H₂SO₄ concentration in Beijing during February–March 2018 period. In
 426 addition, we also used another sulfuric acid nonlinear proxy (Eq. 10) based on long-
 427 term observations in Germany, Finland, the United States, etc. ~~(Mikkonen et al.,~~
 428 ~~2011)~~(Mikkonen et al., 2011). In this study, we adopt the above six nonlinear proxy
 429 methods (referred as proxy5 to proxy10) to estimate H₂SO₄ in Qingdao.

$$430 \quad [H_2SO_4] = 515.74 \times [SO_2]^{0.38} \times \text{Radiation}^{0.14} \times CS^{0.03} \quad (5)$$

$$431 \quad [H_2SO_4] = 280.05 \cdot \text{Radiation}^{0.14} [SO_2]^{0.40} \quad (6)$$

$$432 \quad [H_2SO_4] = 9.95 \times [SO_2]^{0.39} \times \text{Radiation}^{0.13} \times CS^{-0.01} \times [O_3]^{0.14} \quad (7)$$

$$433 \quad [H_2SO_4] = 14.38 \times [SO_2]^{0.38} \times \text{Radiation}^{0.13} \times [O_3]^{0.14} \quad (8)$$

$$434 \quad [H_2SO_4] = 0.0013 \times [SO_2]^{0.38} \times \text{Radiation}^{0.13} \times CS^{-0.17} \times ([O_3]^{0.14} + [NO_x]^{0.41}) \quad (9)$$

$$435 \quad [H_2SO_4] = 8.21 \times 10^{-3} \times [SO_2]^{0.62} \times \text{Radiation} \times (CS \times RH)^{-0.13} \quad (10)$$

436 where [SO₂], [O₃] and [NO_x] (molecule cm⁻³) represents concentration of
 437 observed SO₂, O₃ and NO_x, respectively. “Radiation” (W m⁻²) is global radiation. RH
 438 (%) is the relative humidity, and CS (s⁻¹) is the condensation sink, which is calculated
 439 based on observed particle distribution.

440
 441 The simulated H₂SO₄ concentration from the Base simulation (dots in Fig. 2b) is
 442 compared with observations obtained by proxies (see Fig. 2b), indicating that Base
 443 simulations apparently overestimate by one order of magnitude compared to the H₂SO₄
 444 estimated by proxies. The overestimation has been frequently reported previously, i.e.,
 445 over Beijing ~~(Matsui et al., 2011)~~(Matsui et al., 2011), which ascribes the bias to the
 446 overestimation of the SO₂ concentration. In a more recent study, the sensitivity of
 447 H₂SO₄ to SO₂ is tested, and the result shows that even when SO₂ is reduced to an
 448 unrealistically low level, the simulated H₂SO₄ is still more than one order of magnitude

449 higher than the observed value (Lai et al., 2022), suggesting that the SO₂ concentration
450 cannot fully explain the overestimates.

451 In addition to the precursor of H₂SO₄, the mass accommodation coefficient (α),
452 representing the probability of impaction of a gaseous molecule on a liquid surface and
453 entering the bulk liquid phase, is another important factor affecting the concentration
454 of sulfuric acid gas. In the public release of WRF-Chem, mass accommodation
455 coefficient is typically set to a low value of 0.1 for all gas species under different
456 volatility during the condensation process, including H₂SO₄ (Davidovits et al., 2004;
457 Zaveri et al., 2008)(Davidovits et al., 2004; Zaveri et al., 2008). Recent studies indicate
458 that the low mass accommodation coefficient value may not be applicable to the low
459 volatile gases, which tend to have a mean mass accommodation coefficient value of 0.7
460 and close to the unity (Krechmer et al., 2017)(Krechmer et al., 2017). In fact, an earlier
461 study has indicated based on experimental determination, the mass accommodation
462 coefficient of H₂SO₄ vapor in sulfuric acid aqueous solution was measured, and the best
463 fit value was 0.65 (Pöschl et al., 1998)(Pöschl et al., 1998). Accordingly, a sensitivity
464 simulation was conducted by adjusting the mass accommodation coefficient of H₂SO₄
465 from 0.1 to 0.65, referred to as MAC.

466 This simulation brought the H₂SO₄ concentration (see Fig. 2b) much closer to the
467 calculated results from proxies, and the corresponding biases reduced by approximately
468 an order of magnitude. Notably, the MAC simulation decreases the overestimate of
469 sulfuric acid gas concentration, resulting in a lower particle formation rate. The MAC
470 simulation also significantly reduces overestimate of CN_{10–40} (Fig. 2a), and mean
471 fractional bias compared to observations decreases from 48% to 1%.

472

473 3.2.2 Improvement of particle number concentration simulations at 40–100 nm

474 The number concentration of particles in the 40–100 nm range is mainly affected
475 by the coagulation and condensation processes. While the coagulation process tends to
476 largely affect ultrafine particles below 10 nm than those with larger sizes (Wu et al.,
477 2011)(Wu et al., 2011), the condensation growth of particles during gas-particle
478 partitioning at sizes of 10–40 nm, to a large extent, governs the variations in number

479 concentration of 40–100 nm particles. The condensation process is primarily controlled
480 by gas-particle partitioning of chemical species, which may change the chemical
481 composition of particles, such as organic compounds and inorganics including sulfate,
482 nitrate and ammonium.

483 Among the species contributing to the condensation growth of particles at 10–40
484 nm, the organic compounds with c^* of $10^{-2} \mu\text{g m}^{-3}$ play the dominant role ([Pierce et al.,
2011](#))([Pierce et al., 2011](#)). In the current model setting, the low volatile organic matter
485 of $10^{-2} \mu\text{g m}^{-3}$ comes from two gas-phase sources, including the direct emission of
486 primary organic aerosol (POA) and SOA formed from S/IVOC (SI-SOA), conducive
487 to condensation on particles. While the condensation of gaseous SOA is in general
488 reasonable, the gas phase emissions of POA may be problematic. For instance, previous
489 studies suggested that POA is in gas phase close to the emissions source. However, with
490 rapid dilution and cooling in the atmosphere away from the source, most POA
491 condenses to particle-phase ([Roldin et al., 2011b; Roldin et al., 2011a; Shrivastava et
al., 2008](#))([Roldin et al., 2011b; Roldin et al., 2011a; Shrivastava et al., 2008](#)). Therefore,
492 away from the emissions source POA, being in the particle phase, will not be involved
493 in the growth of newly formed particles. Therefore, POA may not contribute to particle
494 growth away from the emission sources, which caused different size distributions of
495 POA compared to when it was emitted in the gas-phase (Fig. S3a vs. Fig. S3b). Emitting
496 low volatility POA in the particle phase eliminates the unreasonable quasi-banana shape
497 pattern exhibiting concomitant growth of newly formed particles with increasing mass
498 concentration of POA.

501 The composition analysis (Fig. S3c) in the 10–40 nm particles mass from the
502 model results indicates that organic compounds mentioned above only account for 21%
503 of total mass (sulfates, nitrates, ammonium salts and organics) in this size range and the
504 dominant species is nitrate which accounts for 51% of total mass, exhibiting
505 inconsistencies with the previous studies which in general indicates a much smaller
506 contribution of nitrate. For instance, [Liu et al. \(2014a\)](#)[Liu et al. \(2014a\)](#) suggested that
507 over North China Plain in summer 2009, organic matter accounted for 77% of particles
508 around 30 nm, while the sum of SO_4^{2-} , NO_3^- and NH_4^+ only accounted for 18%. [Recent](#)

509 observations conducted in Beijing also indicated that particles at 8–40 nm are mainly
510 composed of organic matter (with mass fraction of ~80%) and sulfate (with mass
511 fraction of ~13%), while nitrate content is very low (with mass fraction of ~3%) (Li et
512 al., 2022). Another study showed that nitrate accounted for 7–8% at urban sites and 17%
513 at rural sites for particles mass in 7–30 nm in the United States in 2007 (~~Bzdek et al.,~~
514 ~~2012~~)(Bzdek et al., 2012). Therefore, the potentially too high modeled nitrate fraction
515 in 10–40 nm in this study is tightly associated with the condensation process, with the
516 specific reasons explained below.

517 The condensation of nitric acid on particles is highly constrained by the particle
518 acidity. The acidity in smaller particles (i.e., 10–40 nm) tends to be higher than that in
519 large particles, primarily due to the larger condensation of H₂SO₄ (~~Lu et al., 2022~~)(Lu
520 ~~et al., 2022~~), and particles with sizes greater than 40 nm have a much weaker acidity or
521 are nearly neutral. For example, observed evidence has shown that acidic ultrafine
522 particles account for a large proportion of ultrafine particles from 22 December 2010
523 to 15 January 2011 in Hong Kong, e.g., 65% for particles within 5.5–30 nm (~~Wang et~~
524 ~~al., 2014~~)(Wang et al., 2014).

525 In the model, a particle is determined to be in solid phase when the ambient relative
526 humidity is lower than the mutual deliquescence relative humidity of the particles
527 (~~Zaveri et al., 2005; Zaveri et al., 2008~~)(Zaveri et al., 2005; Zaveri et al., 2008), which
528 is in general suitable for particles dominated by inorganics. In the study area, the results
529 indicate that at most conditions relative humidity are relatively low and the particles
530 are in solid phase, in which the condensation process is not affected by particle acidity
531 and the condensation of nitric acid on particles is directly calculated based on the gas-
532 particle equilibrium concentration (~~Zaveri et al., 2008~~)(Zaveri et al., 2008). However,
533 for particles below 40 nm, the main compositions are likely to be organic matter (~~Zhu~~
534 ~~et al., 2014; Ehn et al., 2014~~)(Zhu et al., 2014; Ehn et al., 2014), which tends to be in
535 liquid phase (~~Virtanen et al., 2011; Cheng et al., 2015~~)(Virtanen et al., 2011; Cheng et
536 ~~al., 2015~~), under which the condensation of nitric acid is strongly constrained by acidity.
537 Therefore, the phase misrepresentation ignores the weakening effect of acidity on nitric
538 acid condensation, resulting in too high nitrate therein.

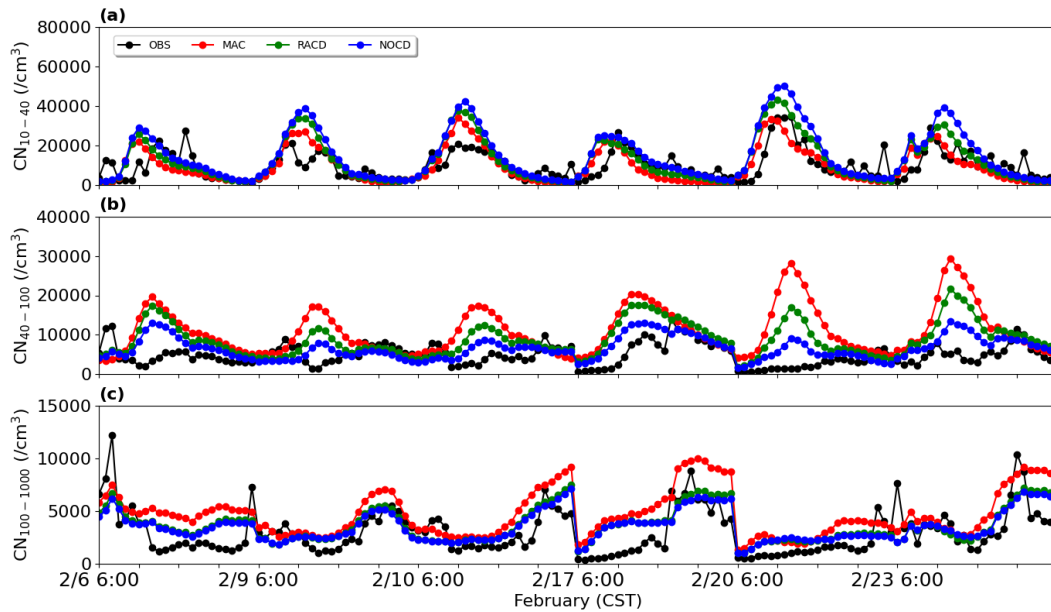
539 To overcome this issue, we propose a ratio method for condensation (RACD) to
540 partition the condensation of nitric acid on particles under 40 nm, by applying a ratio
541 of the number concentration of non-acidic particles to ultrafine particles. The method
542 is based on two assumptions, including: 1) little condensation of nitric acid on particles
543 with strong acidity ([Lu et al., 2022](#))([Lu et al., 2022](#)); 2) the condensation of nitric acid
544 on particles is proportional to the ratio of the number concentration of non-acidic
545 ultrafine particles to the total particles, despite the existence of uncertainties. Fig. S4
546 depicts the average particle number concentration and acid particle in the 1 to 40 nm
547 range, calculated based on [Wang et al. \(2014\)](#)[Wang et al. \(2014\)](#). The ratio of non-acidic
548 particles is 8% for particles below 10 nm, 18% for particles at 10–15.8 nm, 30% for
549 particles at 15.8–25.1nm, and 55% for particles at 25.1–39.8 nm (Fig. S4). Note that
550 the ratio is based on measurements acquired at a single site in Hong Kong, therefore
551 more observational studies are needed to warrant the robustness of the method.
552 Alternatively, the condensation of nitric acid on particles in bins from 1nm to 40 nm is
553 completely suppressed, referred to as NOCD.

554 The simulation results based on the two methods (RACD and NOCD) are shown in
555 Fig. 3. Compared to MAC, RACD simulations reduce previously noted overestimation
556 of particle number concentration in the 40–100 nm size range (Fig. 3b), with the mean
557 fractional bias decreases from 83% to 63%. In addition to the amount of nitrate
558 condensation during particle growth mentioned above, the overestimation of particle
559 number concentrations in the 40–100 nm range may be attributed to nucleation process.
560 More specifically, in the H₂SO₄-H₂O binary nucleation mechanism used in this study,
561 when the concentration of sulfuric acid gas is reduced (Section 3.2.1), the resulting
562 decrease in nucleation rate leads to a slight decrease in particle number concentration
563 at 40–100 nm relative to Base (mean fractional bias from 98% to 83%). Apart from that,
564 it may also be related to the choice of nucleation parameterization scheme. For example,
565 using a global chemical transport model GEOS-Chem with a nucleation mechanism in
566 which formation rate is a function of the concentrations of sulfuric acid and low-
567 volatility organics, [Yu et al. \(2015\)](#)[Yu et al. \(2015\)](#) overestimated the concentration of
568 particles in the 10–100 nm range by 161% at nine sites in the summer in North America.

569 A possible explanation for this overestimation was given by the uncertainty of the
570 predicted concentration of organic compounds involved in organics-mediated
571 nucleation parameterization. After they switched to another scheme of the ion-mediated
572 nucleation mechanism without organic matter, the number becomes 27% lower than the
573 observations (~~Yu et al., 2015~~)(Yu et al., 2015). The test based on different schemes is
574 beyond the scope of the study, which is therefore not investigated.

575 Moreover, the overestimation of particles over 100 nm (CN₁₀₀₋₁₀₀₀; Fig. 3c), which
576 have a strong influence on CCN, also decrease in the RACD simulation. Thus, the mean
577 fractional bias decreases from 25% (MAC) to 1%. Note that the slight increase of CN₁₀₋
578 ₄₀ through the application of RACD, can be linked to the decrease of nitrate
579 condensation, and leads to weakened particle growth and enhanced particle number
580 concentration at 10–40 nm (Fig. 3a). The alternative method by completely removing
581 the nitrate condensation (NOCD) yields even better performance in particle number
582 concentration of 40–100 nm (mean fractional bias of 34%), indicating the feasibility by
583 reducing the nitrate condensation. The proportion of nitrate simulated by RACD is 23%,
584 closer to values reported in past observations (~~Bzdek et al., 2011; Bzdek et al.,~~
585 ~~2012~~)(Bzdek et al., 2011; Bzdek et al., 2012), while the nitrate (1%) in the scenario of
586 NOCD seems to be too low. Considering the limited observational information obtained
587 based on previous studies, RACD is applied in this study.

588 In addition to Qingdao, we evaluate the model performance over a few other sites,
589 including one site over urban Beijing and the other one over the rural area of Gucheng,
590 yielding consistent improvements in model simulations (Section S2; Fig. S6-S8).
591 Moreover, we select another empirical scheme, e.g., kinetics, and one classical
592 nucleation scheme, indicating the empirical scheme of activation scheme is in general
593 a good option in this study (Section S2; Fig. S9-S11; Table S1-3).



594

595 Fig. 3. The time series of (a) CN_{10-40} , (b) CN_{40-100} and (c) $CN_{100-1000}$ on NPF days in
 596 Qingdao on February 5-24 simulated from MAC (marked in red), NOCD (marked in
 597 blue) and RACD (marked in green) as well as from observations (OBS) (marked in
 598 black). All times are local time.

599

600

601 3.3 Substantial contributions of SI-SOA to CCN

602 Compared with the original model setting, after adjusting the growth process of
 603 ultrafine particles (RACD), the number concentration of particles ~~tend~~tends to decrease,
 604 especially for particles above 40 nm. Ultrafine particles above 40 nm are important
 605 sources of CCN (~~Dusek et al., 2006~~)(Dusek et al., 2006), in this way, the number
 606 concentration of CCN also tends to decline. In addition, in the Base case, we found that
 607 the model overestimated $CCN_{0.4\%}$ and $CCN_{0.6\%}$, with mean fractional bias being 64%
 608 and 87%, respectively. After adjusting the condensation growth process of ultrafine
 609 particles, under high supersaturation (i.e., $CCN_{0.4\%}$ and $CCN_{0.6\%}$), the capability of the
 610 model in reproducing the CCN is improved. RACD reduces the overestimation of
 611 $CCN_{0.4\%}$ and $CCN_{0.6\%}$, with mean fractional bias reduced to 30% and 56%, respectively,
 612 although the overestimates still exist (Figs. S5b, c). However, for low supersaturation
 613 (i.e., $CCN_{0.2\%}$), the decrease of number concentration of CCN is too large, and mean

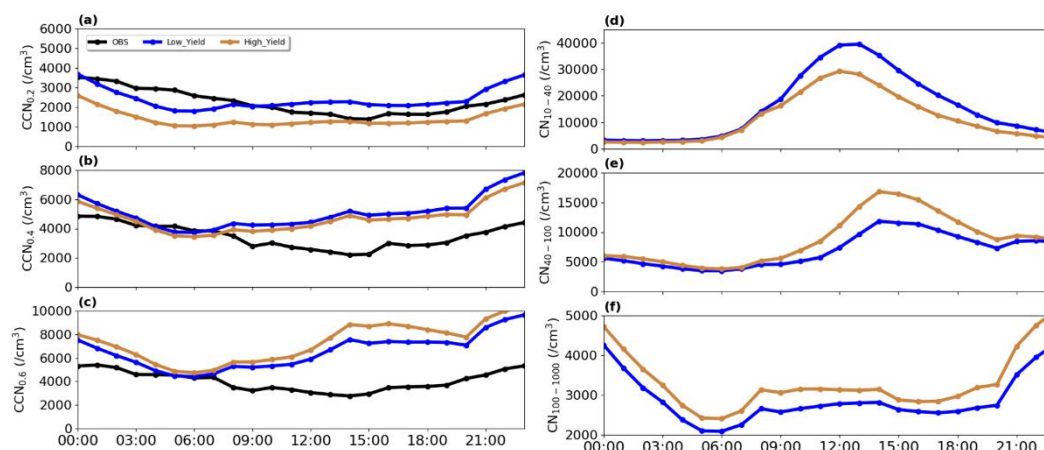
614 fractional bias decreases from 7% to -45% (Fig. S5a), therefore the bias will be further
615 adjusted later.

616 In addition to the growth process, the remaining overestimate of CCN under high
617 SS and underestimate of CCN over low SS is likely to be influenced by the chemical
618 compositions involved in the activation of ultrafine particles into CCN. Specifically,
619 ultrafine particles can grow up to CCN size under certain SS (Pierce and Adams, 2007).
620 This process is influenced by both particle size and hygroscopicity, and hygroscopicity
621 is closely related to the chemical composition of particles (Petters and Kreidenweis,
622 2007). In particular, inorganic compounds generally increase particle hygroscopicity,
623 increasing CCN. SOA has dual effects on CCN since it decreases particle
624 hygroscopicity but also promotes growth of particles, and these two effects are
625 competitive with each other (Wu et al., 2015; Zaveri et al., 2021). Ultrafine particles
626 must grow to a critical size to be activated into CCN (~~Dusek et al., 2006~~)(Dusek et al.,
627 2006). SOA act as a major contributor in promoting the condensational growth of
628 ultrafine particles to the critical size, facilitating particles activation into CCN. In
629 contrast, SOA tends to reduce the hygroscopicity of particles, leading to a diminished
630 ability of activation to CCN (Wu et al., 2015). These two competing effects work
631 together and modulate the number of CCN. Moreover, considering that SI-SOA is the
632 main SOA component on ultrafine particles (Fig. S5d), the effect of SI-SOA on CCN is
633 therefore explored in this study.

634 Considering SI-SOA is a product of S/IVOC oxidation, the oxidation rate of
635 S/IVOC is tightly associated with CCN, which likely affects the bias of CCN. In the
636 original model setup, the oxidation rate is set to be a constant of $4 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1}$
637 s^{-1} for all S/IVOC. ~~However, a recent study (Wu et al., 2021b)~~However, a recent study
638 (Wu et al., 2021b) proposed that the oxidation rate can be as high as $5 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1}$
639 s^{-1} such as for polycyclic aromatic hydrocarbons (PAHs), close to the original model
640 value, but can be as low as half (i.e., $2 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$) of the original modeling
641 setting for S/IVOC species except PAHs (O-S/IVOCs). It is noteworthy that the
642 oxidation rates of 5×10^{-11} and 2×10^{-11} in general represent the upper and lower bounds
643 (~~Zhao et al., 2016; Wu et al., 2021b~~)(Zhao et al., 2016; Wu et al., 2021b).

644 To delve into how oxidation rates affect CCN, we set up a few numerical
645 experiments (Table 23) to investigate the response of CCN to the oxidation rate of
646 S/IVOC at three supersaturations (0.6%, 0.4%, 0.2%), including cases of High_Yield
647 and Low_Yield. As it is shown in Fig. 4, decreasing the oxidation rate (Low_Yield)
648 leads to a reduction of ~10% of CCN at high supersaturation (i.e., CCN_{0.6%}) as
649 compared to the High_Yield simulation. This behaviour is a consequence of the
650 decrease of particle number concentrations associated with Low_Yield, particular of
651 the particles close to the critical diameter (40–100 nm). In this case, the effect of particle
652 size dominates the hygroscopicity. In contrast, at a lower supersaturation (CCN_{0.2%}),
653 CCN increases by 42% when the oxidation rate is switched from a high to a low value,
654 which is due to the smaller fraction of SI-SOA contributing to particulate mass when
655 the oxidation rate is low. In this case, relative to SOA, a larger fraction of other particle
656 constituents such as inorganics, increase the volume weighted particle hygroscopicity
657 (~~Dusek et al., 2006~~)(Dusek et al., 2006) which causes the increase of CCN number. This
658 means that the effect of hygroscopicity on CCN surpasses the influence on particle size
659 at low supersaturations. This conclusion is consistent with the observation conducted
660 by ~~Ma et al. (2016)~~Ma et al. (2016) in the North China Plain in 2013, which suggested
661 that along with the decrease of SS, the particles that can be activated to CCN is more
662 sensitive to changes of particle hygroscopicity. Similarly, based on observational data
663 in northern China in summer, Wang et al. (2023) found that CN in 2020 is lower than
664 that in 2014 due to particulate pollution control, however, the particles become more
665 easily activated, attributable to the larger extent of decrease in organic matters
666 compared to inorganics, leading to enhanced particle hygroscopicity and more
667 conducive to activation.

668



692
 693 Fig. 4. Average diurnal variation of (a) $CCN_{0.2\%}$, (b) $CCN_{0.4\%}$ and (c) $CCN_{0.6\%}$ and (d)
 694 CN_{10-40} , (e) CN_{40-100} , (f) $CN_{100-1000}$ on NPF days in Qingdao on February 5-24, 2017,
 695 in Low_yield and High_yield simulations, shown as blue and brown lines, and black
 696 lines represent observation results.

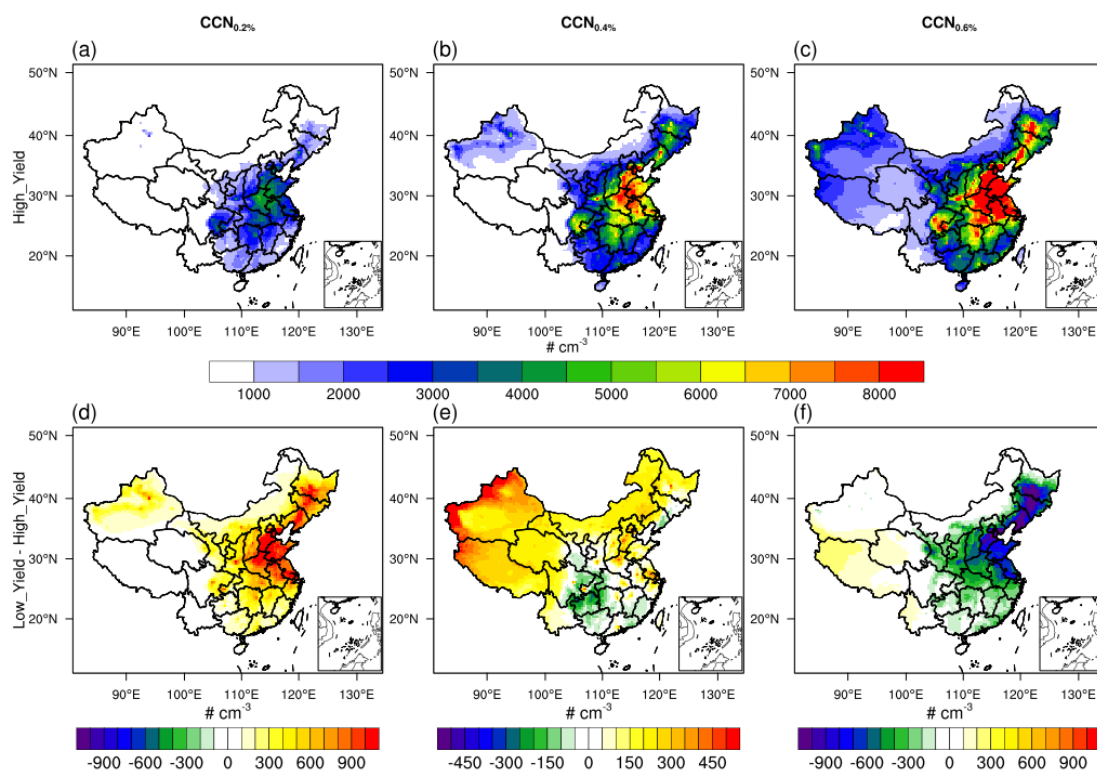
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 698 Furthermore, compared to the high yield of SI-SOA, the low SI-SOA yield results
 699 in a high CCN concentration under low SS and low CCN concentration under high SS.
 700 Therefore, both the underestimates of $CCN_{0.2\%}$ (mean fractional bias of -45%) and
 701 overestimates of $CCN_{0.6\%}$ (mean fractional bias of 56%) mentioned above are improved,
 702 with mean fractional bias of $CCN_{0.2\%}$ and $CCN_{0.6\%}$ reaching 7% and 43%, respectively
 703 (Fig. 4a,c). This result suggests that the oxidation rate of S/IVOC is possibly closer to
 704 the low value, which is understandable based on [Wu et al. \(2021b\)](#) [Wu et al. \(2021b\)](#),
 705 who found that the amount of O-S/IVOCs, which corresponds to a low oxidation rate,
 706 is in general much larger (i.e., 20 times) than that of PAHs with a high oxidation rate.

707 In addition to the single site of Qingdao, we further explore the impact of SI-SOA
 708 yield on CCN from a larger spatial coverage (Fig. 5). Consistent with the mechanism
 709 revealed over Qingdao, even from a larger spatial perspective, a lower oxidation rate of
 710 S/IVOC essentially enhances CCN at a lower SS (e.g., $CCN_{0.2\%}$; Fig. 5a) with the
 711 highest increase over North China Plain area (Fig. 5a), and weakens CCN (i.e., by 10–
 712 20% over Beijing-Tianjin-Hebei) at a higher SS (Fig. 5c), particularly over the dense
 713 emission area (Fig. S12). It is worth noting that in the 2-species VBS mechanism used
 714 in our study, all S/IVOC in the inventory is calculated based on a constant emission

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733 ratio of S/IVOC to POA from all source categories (Shrivastava et al.,
734 2011)(Shrivastava et al., 2011), which may miss part of S/IVOC due to different
735 emission ratios of POA from different source (Chang et al., 2022)(Chang et al., 2022).
736 In addition, the simplified VBS mechanism used in our study does not take into account
737 the multi-step oxidation of organic species, which may introduce some uncertainties.
738 To be more specific, in the 2-species VBS mechanism, SI-SOA with effective saturation
739 concentrations (c^*) of $10^{-2} \mu\text{g m}^{-3}$ is formed by the vapor phase oxidation of S/IVOC
740 vapors with c^* of $10^5 \mu\text{g m}^{-3}$, reducing volatility by 7 orders of magnitude. The process
741 of one-step oxidation does not mean to represent a physical process, but to parameterize
742 the mean effect of a complex process of SOA formation (Shrivastava et al., 2011).
743 However, in the real atmosphere, the gaseous VOCs often undergo multi-generational
744 oxidation to form SOA (Garmash et al., 2020), during which the properties and
745 composition of SOA change substantially. For instance, by adding the formation
746 chemistry associated with multi-generational oxidation, Zhao et al. (2020) found
747 improved simulations of vertical aerosol profile in the Amazon free troposphere
748 compared to the simplified VBS mechanism.



749

750 Fig. 5. Spatial distributions of CCN concentrations at different supersaturations (SS),

751 (a) and (d) are $CCN_{0.2\%}$, (b) and (e) are $CCN_{0.4\%}$, and (c) and (f) are $CCN_{0.6\%}$. The top
752 panels exhibit the results from the High_Yield simulation, and the bottom panels shows
753 the difference between the Low_Yield and High_Yield simulations.

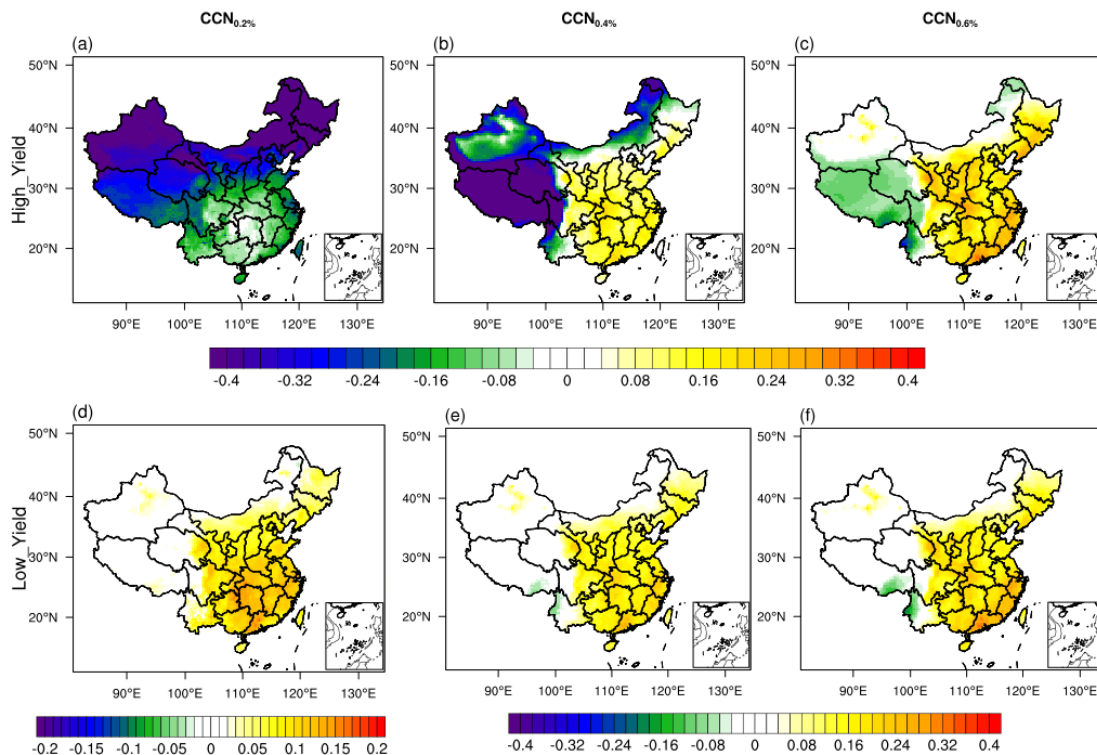
754

755 3.4 Contribution of nucleation to CCN under different SI-SOA yields

756 Considering the importance of nucleated particles on CCN ([Yu et al., 2020](#);
757 [Westervelt et al., 2013](#))([Yu et al., 2020](#); [Westervelt et al., 2013](#)), we further investigate
758 the influence of nucleation on CCN under different SI-SOA yield conditions discussed
759 above.

760 As shown in Fig. 6, in simulations close to the original model setting (High_Yield),
761 when SS is low (i.e., $SS=0.2\%$), the nucleation process tends to reduce the CCN by
762 $\sim 10\text{--}50\%$. In contrast, when the SS is high (0.6%), the nucleation results in a significant
763 increase in CCN in most regions of China. When the yield of SI-SOA is adjusted to a
764 lower level, the nucleation process has a positive contribution to CCN under both low
765 and high SS. Especially, when SS is low (0.2%), the sign reversal, i.e., from negative
766 (Fig. 6a) to positive (Fig. 6d) contributions of NPF to CCN along with the decrease of
767 SI-SOA yield, i.e., the increase is concentrated in the eastern China with an average of
768 $10\text{--}20\%$. The primary mechanism lies in that along with the decrease of SI-SOA yield,
769 the smaller fraction of SI-SOA yields an increase in hygroscopicity, which surpasses
770 the suppression effect on particle growth due to reduced SI-SOA formation. In the real
771 atmosphere, when the supersaturation is usually low, e.g. about $\sim 0.1\%$ in polluted areas
772 ([Kalkavouras et al., 2019](#); [Hudson and Noble, 2014](#))([Kalkavouras et al., 2019](#); [Hudson](#)
773 [and Noble, 2014](#)), CCN will likely reduce with increasing oxidation rate of S/IVOC
774 and corresponding SI-SOA formation.

775



776

777 Fig. 6. Spatial distribution of contribution of nucleation to CCN calculated by the ratio
 778 of the difference between the parameterization with and without nucleation to the
 779 parameterization with nucleation under different SI-SOA yields in China in February
 780 2017. (a), (d) is $CCN_{0.2\%}$, (b), (e) is $CCN_{0.4\%}$, (c), (f) is $CCN_{0.6\%}$. The upper panel
 781 and lower panel represent High_Yield and Low_Yield simulation respectively

782

783 In addition to the linear- H_2SO_4 nucleation mechanism, one more empirical scheme
 784 of kinetics nucleation is selected, which assumes that the nucleation rate is proportional
 785 to the square of the concentration of sulfuric acid ($J = K[H_2SO_4]^2$), to investigate the
 786 effect of nucleation on CCN. Substantially positive contributions of nucleation to CCN
 787 is found when the low SI-SOA yield is applied, consistent with what was shown based
 788 on the linear- H_2SO_4 nucleation scheme (Fig. S13). However, nucleation contributes
 789 positively to CCN even when the SI-SOA yield is high in the quadratic- H_2SO_4
 790 nucleation scheme (e.g., kinetics nucleation scheme). When more sulfuric acid
 791 molecules participate in nucleation under this scheme than the linear- H_2SO_4 nucleation
 792 scheme, the particles are more easily hygroscopically activated to CCN, which is
 793 equivalent to the effect of a reduction in organic components in the linear- H_2SO_4

794 nucleation scheme (e.g., activation-type nucleation scheme). The results from this study
795 show the importance of assessing the simulated effects of the nucleation scheme on not
796 only the formation and growth process of particles but also climate factors such as CCN
797 using observations.

798

799 **Conclusions and discussions**

800 In this study, WRF-Chem explicit-NPF simulations, with linear-H₂SO₄ nucleation
801 scheme (e.g., activation-type nucleation scheme), are used to investigate the observed
802 wintertime NPF events and their contribution to CCN in China. Based on observations
803 in a typical coastal city of Qingdao, as well as in the cities of Beijing and Gucheng over
804 North China Plain, we identify high biases of the model simulated CN and CCN
805 concentrations. Therefore, we updated and improved the parameterization setting on
806 particle growth in the model, mainly including: (1) adjusting the mass accommodation
807 coefficient (α) to from the default value of 0.1 to 0.65, an important parameter for
808 sulfuric acid condensation; (2) proportionally reducing the condensation amount of
809 nitric acid on particles below 40 nm, (3) changing the emitted low-volatility POA from
810 gas to particle. Through these adjustments, the capability of the model in reproducing
811 CN and CCN is substantially improved, leading to better agreement with the observed
812 results, which significantly reduces the overestimation of CN₁₀₋₄₀ (mean fractional bias
813 decreases from 48% to 1%) and CN₄₀₋₁₀₀ (mean fractional bias decreases from 98% to
814 63%).

815

816 For CCN, due to the crucial role of SI-SOA in promoting the growth of ultrafine
817 particles, on the basis of previous studies, we lower the oxidation rate of S/IVOC and
818 hence the production rate of SI-SOA, which weakens the growth of particles to reach
819 the critical size of CCN activation, but enhances particulate hygroscopicity favoring the
820 activation to CCN. When the yield of SI-SOA is adjusted to the lower bound of
821 literature value, CCN_{0.6%} is reduced by ~10% and is closer to observations. At low SS
822 (CCN_{0.2%}), the decrease of SI-SOA yield has greater effects on the increase of particle
823 hygroscopicity compared to the effect of the reduction of particle size due to the

824 decrease of condensation growth. It results in an increase of CCN (as large as ~42%) in
825 better agreement with observations. Under low SS conditions, common in the
826 atmosphere, a 2.5-fold reduction in SI-SOA yield results in a substantial increase of
827 CCN that switches from a negative contribution of new particle formation to CCN from
828 -50%~-10% to a positive contribution of 10~20%. ~~The substantial contribution of new
829 particle formation to CCN under low SS and SI-SOA is applicable to other mechanisms
830 such as kinetics.~~

831 In addition to activation nucleation scheme, we have also tested a few other
832 schemes such as the quadratic-H₂SO₄ nucleation scheme (e.g., kinetics nucleation
833 scheme). Under this scheme, the bias-corrected method abovementioned is applicable
834 to improving the simulations of concentrations of CN and CCN. It is noteworthy that
835 the dependence of CCN on the SI-SOA yield is diminished, showing that under both
836 high and low yields of SI-SOA, there are positive contributions of NPF to CCN. This
837 is likely due to the increase in the amount of sulfuric acid involved in nucleation,
838 making it more hygroscopic and easier to activate to CCN, and the high content of
839 inorganic species makes them less sensitive to changes in SI-SOA yield, which deserves
840 further investigation.

841

842 **Competing interests.** At least one of the (co-)authors is a member of the editorial board
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844

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852

853 **References:**

854 Arghavani S, Rose C, Banson S, et al. 2022. The Effect of Using a New Parameterization of Nucleation
855 in the WRF-Chem Model on New Particle Formation in a Passive Volcanic Plume. *Atmosphere*
856 [J], 13(1): 15.

857 ~~Bergman T, Laaksonen A, Korhonen H, et al. 2015. Geographical and diurnal features of amine-enhanced~~
858 ~~boundary layer nucleation. *Journal of Geophysical Research: Atmospheres* [J], 120(18): 9606-~~
859 ~~9624.~~

860 Buchholz R R, Emmons L K, Tilmes S 2019. The CESM2 Development Team., 2019. CESM2.1/CAM-
861 Chem Instantaneous Output for Boundary Conditions. UCAR/NCAR - Atmospheric Chemistry
862 Observations and Modeling Laboratory.

863 Bzdek B, Zordan C, Luther G, et al. 2011. Nanoparticle Chemical Composition During New Particle
864 Formation. *Aerosol Science and Technology* [J], 45(1041-1048).

865 Bzdek B R, Zordan C A, Pennington M R, et al. 2012. Quantitative Assessment of the Sulfuric Acid
866 Contribution to New Particle Growth. *Environmental Science & Technology* [J], 46(8): 4365-
867 4373.

868 Cai C, Zhang X, Wang K, et al. 2016. Incorporation of new particle formation and early growth
869 treatments into WRF/Chem: Model improvement, evaluation, and impacts of anthropogenic
870 aerosols over East Asia. *Atmospheric Environment* [J], 124(262-284).

871 Cai J, Chu B, Yao L, et al. 2020. Size-segregated particle number and mass concentrations from different
872 emission sources in urban Beijing. *Atmos. Chem. Phys.* [J], 20(21): 12721-12740.

873 Carter W 2000. Documentation of the SAPRC-99 Chemical Mechanism for VOC Reactivity Assessment.
874 Final Report to California Air Resources Board [J].

875 Chang X, Zhao B, Zheng H, et al. 2022. Full-volatility emission framework corrects missing and
876 underestimated secondary organic aerosol sources. *One Earth* [J], 5(403-412).

877 ~~Chang Y, Gao Y, Lu Y, et al. 2021. Discovery of a Potent Source of Gaseous Amines in Urban China.~~
878 ~~*Environmental Science & Technology Letters* [J], 8(9): 725-731.~~

879 Chen F, Dudhia J 2000. Coupling an Advanced Land-Surface/Hydrology Model with the Penn
880 State/NCAR MM5 Modeling System. 129(

881 Cheng Y, Su H, Koop T, et al. 2015. Size dependence of phase transitions in aerosol nanoparticles. *Nature*
882 *Communications* [J], 6(1): 5923.

883 Chrit M, Sartelet K, Sciare J, et al. 2018. Modeling organic aerosol concentrations and properties during
884 winter 2014 in the northwestern Mediterranean region. *Atmos. Chem. Phys.* [J], 18(24): 18079-
885 18100.

886 Chu B, Kerminen V-M, Bianchi F, et al. 2019. Atmospheric new particle formation in China.
887 *Atmospheric Chemistry and Physics* [J], 19(115-138).

888 Dal Maso M, Kulmala M, Riipinen I, et al. 2005. Formation and growth of fresh atmospheric aerosols:
889 Eight years of aerosol size distribution data from SMEAR II, Hyytiälä, Finland. *Boreal*
890 *Environment Research* [J], 10(323-336).

891 Davidovits P, Worsnop D R, Jayne J T, et al. 2004. Mass accommodation coefficient of water vapor on
892 liquid water. *Geophysical Research Letters* [J], 31(22).

893 Donahue N M, Robinson A L, Stanier C O, et al. 2006. Coupled Partitioning, Dilution, and Chemical
894 Aging of Semivolatile Organics. *Environmental Science & Technology* [J], 40(8): 2635-2643.

895 Dong C, Matsui H, Spak S, et al. 2019. Impacts of New Particle Formation on Short-term Meteorology
896 and Air Quality as Determined by the NPF-explicit WRF-Chem in the Midwestern United States.
897 *Aerosol and Air Quality Research* [J], 19(2): 204-220.

898 Dusek U, Frank G P, Hildebrandt L, et al. 2006. Size Matters More Than Chemistry for Cloud-Nucleating
899 Ability of Aerosol Particles. *Science [J]*, 312(5778): 1375-1378.

900 Ehn M, Thornton J, Kleist E, et al. 2014. A large source of low-volatility secondary organic aerosol.
901 *Nature [J]*, 506(476-479).

902 Fanourgakis G S, Kanakidou M, Nenes A, et al. 2019. Evaluation of global simulations of aerosol particle
903 and cloud condensation nuclei number, with implications for cloud droplet formation. *Atmos.*
904 *Chem. Phys. [J]*, 19(13): 8591-8617.

905 Gordon H, Kirkby J, Baltensperger U, et al. 2017. Causes and importance of new particle formation in
906 the present-day and preindustrial atmospheres. 122(16): 8739-8760.

907 Grell G A 1993. Prognostic Evaluation of Assumptions Used by Cumulus Parameterizations. *Monthly*
908 *Weather Review [J]*, 121(3): 764-787.

909 Guo S, Hu M, Zamora M L, et al. 2014. Elucidating severe urban haze formation in China. *Proceedings*
910 *of the National Academy of Sciences [J]*, 111(49): 17373-17378.

911 Hong S-Y, Noh Y, Dudhia J 2006. A New Vertical Diffusion Package with an Explicit Treatment of
912 Entrainment Processes. *Monthly Weather Review - MON WEATHER REV [J]*, 134(
913 Hudson J G, Noble S 2014. CCN and Vertical Velocity Influences on Droplet Concentrations and
914 Supersaturations in Clean and Polluted Stratus Clouds. *Journal of the Atmospheric Sciences [J]*,
915 71(1): 312-331.

916 Iacono M, Delamere J, Mlawer E, et al. 2008. Radiative Forcing by Long-Lived Greenhouse Gases:
917 Calculations with the AER Radiative Transfer Models. *Journal of Geophysical Research [J]*,
918 113(
919 Jimenez J L, Canagaratna M R, Donahue N M, et al. 2009. Evolution of Organic Aerosols in the
920 Atmosphere. *Science [J]*, 326(5959): 1525-1529.

921 Kalkavouras P, Bougiatioti A, Kalivitis N, et al. 2019. Regional new particle formation as modulators of
922 cloud condensation nuclei and cloud droplet number in the eastern Mediterranean. *Atmos. Chem.*
923 *Phys. [J]*, 19(9): 6185-6203.

924 Kerminen V-M, Chen X, Vakkari V, et al. 2018. Atmospheric new particle formation and growth: Review
925 of field observations. *Environmental Research Letters [J]*, 13(
926 Krechmer J E, Day D A, Ziemann P J, et al. 2017. Direct Measurements of Gas/Particle Partitioning and
927 Mass Accommodation Coefficients in Environmental Chambers. *Environ Sci Technol [J]*,
928 51(20): 11867-11875.

929 Kulmala M, Dada L, Daellenbach K R, et al. 2021. Is reducing new particle formation a plausible solution
930 to mitigate particulate air pollution in Beijing and other Chinese megacities? *Faraday*
931 *Discussions [J]*, 226(0): 334-347.

932 Kulmala M, L L, Lehtinen K, et al. 2004. Initial steps of aerosol growth. *Atmospheric Chemistry and*
933 *Physics [J]*, 4(
934 Kulmala M, Petäjä T, Ehn M, et al. 2013. Chemistry of Atmospheric Nucleation: On the Recent Advances
935 on Precursor Characterization and Atmospheric Cluster Composition in Connection with
936 Atmospheric New Particle Formation. *Annual review of physical chemistry [J]*, 65(
937 Kulmala M, Petäjä T, Nieminen T, et al. 2012. Measurement of the nucleation of atmospheric aerosol
938 particles. *Nature Protocols [J]*, 7(9): 1651-1667.

939 Lai S, Hai S, Gao Y, et al. 2022. The striking effect of vertical mixing in the planetary boundary layer on
940 new particle formation in the Yangtze River Delta. *Science of The Total Environment [J]*,
941 829(154607).

942 Lee S-H, Gordon H, Yu H, et al. 2019. New Particle Formation in the Atmosphere: From Molecular
943 Clusters to Global Climate. *Journal of Geophysical Research: Atmospheres* [J], 124(
944 Li K, Zhu Y, Gao H, et al. 2015. A comparative study of cloud condensation nuclei measured between
945 non-heating and heating periods at a suburb site of Qingdao in the North China. *Atmospheric*
946 *Environment* [J], 112(40-53).

947 Li M, Liu H, Geng G, et al. 2017. Anthropogenic emission inventories in China: a review. *National*
948 *Science Review* [J], 4(6): 834-866.

949 Li X, Li Y, Cai R, et al. 2022. Insufficient Condensable Organic Vapors Lead to Slow Growth of New
950 Particles in an Urban Environment. *Environmental Science & Technology* [J], 56(14): 9936-
951 9946.

952 Liu H J, Zhao C S, Nekat B, et al. 2014a. Aerosol hygroscopicity derived from size-segregated chemical
953 composition and its parameterization in the North China Plain. *Atmos. Chem. Phys.* [J], 14(5):
954 2525-2539.

955 Liu M, Matsui H 2022. Secondary Organic Aerosol Formation Regulates Cloud Condensation Nuclei in
956 the Global Remote Troposphere. *Geophysical Research Letters* [J], 49(18): e2022GL100543.

957 Liu X, Zhu Y, Zheng M, et al. 2014b. Production and growth of new particles during two cruise
958 campaigns in the marginal seas of China. *Atmospheric Chemistry and Physics* [J], 14(
959 Liu X H, Zhu Y J, Zheng M, et al. 2014c. Production and growth of new particles during two cruise
960 campaigns in the marginal seas of China. *Atmos. Chem. Phys.* [J], 14(15): 7941-7951.

961 Lovejoy E R, Curtius J, Froyd K D 2004. Atmospheric ion-induced nucleation of sulfuric acid and water.
962 *Journal of Geophysical Research: Atmospheres* [J], 109(D8).

963 Lu H, Wang G, Guo H 2022. Ambient acidic ultrafine particles in different land-use areas in two
964 representative Chinese cities. *Science of The Total Environment* [J], 830(154774).

965 Lu Y, Yan C, Fu Y, et al. 2019. A proxy for atmospheric daytime gaseous sulfuric acid concentration in
966 urban Beijing. *Atmos. Chem. Phys.* [J], 19(3): 1971-1983.

967 Lupascu A, Easter R, Zaveri R, et al. 2015. Modeling particle nucleation and growth over northern
968 California during the 2010 CARES campaign. *Atmos. Chem. Phys.* [J], 15(21): 12283-12313.

969 Ma N, Zhao C, Tao J, et al. 2016. Variation of CCN activity during new particle formation events in the
970 North China Plain. *Atmos. Chem. Phys.* [J], 16(13): 8593-8607.

971 Matsui H, Koike M, Kondo Y, et al. 2011. Impact of new particle formation on the concentrations of
972 aerosols and cloud condensation nuclei around Beijing. *Journal of Geophysical Research:*
973 *Atmospheres* [J], 116(D19).

974 Matsui H, Koike M, Takegawa N, et al. 2013. Spatial and temporal variations of new particle formation
975 in East Asia using an NPF-explicit WRF-chem model: North-south contrast in new particle
976 formation frequency. 118(20): 11,647-611,663.

977 Merikanto J, Spracklen D V, Mann G W, et al. 2009. Impact of nucleation on global CCN. *Atmos. Chem.*
978 *Phys.* [J], 9(21): 8601-8616.

979 Mikkonen S, Romakkaniemi S, Smith J N, et al. 2011. A statistical proxy for sulphuric acid concentration.
980 *Atmos. Chem. Phys.* [J], 11(21): 11319-11334.

981 Morrison H, Thompson G, Tatarskii V 2009. Impact of Cloud Microphysics on the Development of
982 Trailing Stratiform Precipitation in a Simulated Squall Line: Comparison of One and Two-
983 Moment Schemes. *Monthly Weather Review - MON WEATHER REV* [J], 137(991-1007).

984 ~~Napari I, Noppel M, Vehkamäki H, et al. 2002. Parametrization of ternary nucleation rates for H₂SO₄-~~
985 ~~NH₃-H₂O vapors. Journal of Geophysical Research: Atmospheres [J], 107(D19): AAC 6 1-~~
986 ~~AAC 6 6.~~

987 Nieminen T, Kerminen V M, Petäjä T, et al. 2018. Global analysis of continental boundary layer new
988 particle formation based on long-term measurements. Atmos. Chem. Phys. [J], 18(19): 14737-
989 14756.

990 Petäjä T, Mauldin I R L, Kosciuch E, et al. 2009. Sulfuric acid and OH concentrations in a boreal forest
991 site. Atmos. Chem. Phys. [J], 9(19): 7435-7448.

992 Pierce J, Riipinen I, Kulmala M, et al. 2011. Quantification of the volatility of secondary organic
993 compounds in ultrafine particles during nucleation events. Atmospheric Chemistry and Physics
994 Discussions [J], 11(14495-14539).

995 Pöschl U, Canagaratna M, Jayne J T, et al. 1998. Mass Accommodation Coefficient of H₂SO₄ Vapor on
996 Aqueous Sulfuric Acid Surfaces and Gaseous Diffusion Coefficient of H₂SO₄ in N₂/H₂O. The
997 Journal of Physical Chemistry A [J], 102(49): 10082-10089.

998 Qiao X, Yan C, Li X, et al. 2021. Contribution of Atmospheric Oxygenated Organic Compounds to
999 Particle Growth in an Urban Environment. Environmental Science & Technology [J], XXXX(
1000 Ren J, Chen L, Fan T, et al. 2021. The NPF Effect on CCN Number Concentrations: A Review and Re-
1001 Evaluation of Observations From 35 Sites Worldwide. Geophysical Research Letters [J], 48(19):
1002 e2021GL095190.

1003 Riipinen I, Sihto S L, Kulmala M, et al. 2007. Connections between atmospheric sulphuric acid and new
1004 particle formation during QUEST III–IV campaigns in Heidelberg and Hyytiälä. Atmos.
1005 Chem. Phys. [J], 7(8): 1899-1914.

1006 Roldin P, Swietlicki E, Massling A, et al. 2011a. Aerosol ageing in an urban plume – implication for
1007 climate. Atmos. Chem. Phys. [J], 11(12): 5897-5915.

1008 Roldin P, Swietlicki E, Schurgers G, et al. 2011b. Development and evaluation of the aerosol dynamics
1009 and gas phase chemistry model ADCHEM. Atmos. Chem. Phys. [J], 11(12): 5867-5896.

1010 Saha S, Moorthi S, Wu X, et al. 2014. The NCEP Climate Forecast System Version 2. Journal of Climate
1011 [J], 27(6): 2185-2208.

1012 Shen X, Sun J, Zhang X, et al. 2018. Comparison of Submicron Particles at a Rural and an Urban Site in
1013 the North China Plain during the December 2016 Heavy Pollution Episodes. Journal of
1014 Meteorological Research [J], 32(26-37).

1015 Shrivastava M, Fast J, Easter R, et al. 2011. Modeling organic aerosols in a megacity: comparison of
1016 simple and complex representations of the volatility basis set approach. Atmospheric Chemistry
1017 and Physics [J], 11(6639-6662).

1018 Shrivastava M K, Lane T E, Donahue N M, et al. 2008. Effects of gas particle partitioning and aging of
1019 primary emissions on urban and regional organic aerosol concentrations. 113(D18).

1020 Sihto S L, Kulmala M, Kerminen V M, et al. 2006. Atmospheric sulphuric acid and aerosol formation:
1021 implications from atmospheric measurements for nucleation and early growth mechanisms.
1022 Atmos. Chem. Phys. [J], 6(12): 4079-4091.

1023 Sihto S L, Mikkilä J, Vanhanen J, et al. 2011. Seasonal variation of CCN concentrations and aerosol
1024 activation properties in boreal forest. Atmos. Chem. Phys. [J], 11(24): 13269-13285.

1025 Tewari M, Wang W, Dudhia J, et al. 2016. Implementation and verification of the united NOAA land
1026 surface model in the WRF model [M].

1027 Virtanen A, Kannosto J, Kuuluvainen H, et al. 2011. Bounce behavior of freshly nucleated biogenic
1028 secondary organic aerosol particles. *Atmos. Chem. Phys.* [J], 11(16): 8759-8766.

1029 Wang D-W, Guo H, Chan C K 2014. Diffusion Sampler for Measurement of Acidic Ultrafine Particles
1030 in the Atmosphere. *Aerosol Science and Technology* [J], 48(12): 1236-1246.

1031 Wang J, Li M, Li L, et al. 2022. Particle number size distribution and new particle formation in Xiamen,
1032 the coastal city of Southeast China in wintertime. *Science of The Total Environment* [J],
1033 826(154208).

1034 [Wang Y, Wang Y, Song X, et al. 2023. The impact of particulate pollution control on aerosol
1035 hygroscopicity and CCN activity in North China. *Environmental Research Letters* \[J\],
1036 18\(074028\).](#)

1037 Westervelt D M, Pierce J R, Riipinen I, et al. 2013. Formation and growth of nucleated particles into
1038 cloud condensation nuclei: model–measurement comparison. *Atmos. Chem. Phys.* [J], 13(15):
1039 7645-7663.

1040 Wu H, Li Z, Jiang M, et al. 2021a. Contributions of traffic emissions and new particle formation to the
1041 ultrafine particle size distribution in the megacity of Beijing. *Atmospheric Environment* [J],
1042 262(118652).

1043 Wu L, Ling Z, Shao M, et al. 2021b. Roles of Semivolatile/Intermediate-Volatility Organic Compounds
1044 on SOA Formation Over China During a Pollution Episode: Sensitivity Analysis and
1045 Implications for Future Studies. *Journal of Geophysical Research: Atmospheres* [J], 126(8):
1046 e2020JD033999.

1047 Wu Z, Hu M, Yue D, et al. 2011. Evolution of particle number size distribution in an urban atmosphere
1048 during episodes of heavy pollution and new particle formation. *Science China Earth Sciences*
1049 [J], 54(11): 1772.

1050 Yao L, Garmash O, Bianchi F, et al. 2018. Atmospheric new particle formation from sulfuric acid and
1051 amines in a Chinese megacity. *Science* [J], 361(278-281).

1052 Yu F 2005. Quasi-unary homogeneous nucleation of H₂SO₄-H₂O. *The Journal of Chemical Physics* [J],
1053 122(7): 074501.

1054 Yu F, Luo G, Nair A A, et al. 2020. Wintertime new particle formation and its contribution to cloud
1055 condensation nuclei in the Northeastern United States. *Atmos. Chem. Phys.* [J], 20(4): 2591-
1056 2601.

1057 Yu F, Luo G, Pryor S C, et al. 2015. Spring and summer contrast in new particle formation over nine
1058 forest areas in North America. *Atmos. Chem. Phys.* [J], 15(24): 13993-14003.

1059 Yuan Q, Li W, Zhou S, et al. 2015. Integrated evaluation of aerosols during haze-fog episodes at one
1060 regional background site in North China Plain. *Atmospheric Research* [J], 156(102-110).

1061 Yue D L, Hu M, Zhang R Y, et al. 2011. Potential contribution of new particle formation to cloud
1062 condensation nuclei in Beijing. *Atmospheric Environment* [J], 45(33): 6070-6077.

1063 Zaveri R A, Easter R C, Fast J D, et al. 2008. Model for Simulating Aerosol Interactions and Chemistry
1064 (MOSAIC). 113(D13).

1065 Zaveri R A, Easter R C, Peters L K 2005. A computationally efficient Multicomponent Equilibrium
1066 Solver for Aerosols (MESA). *Journal of Geophysical Research: Atmospheres* [J], 110(D24).

1067 Zhang Q, Jimenez J L, Canagaratna M R, et al. 2007. Ubiquity and dominance of oxygenated species in
1068 organic aerosols in anthropogenically-influenced Northern Hemisphere midlatitudes.
1069 *Geophysical Research Letters* [J], 34(13).

-
- 1087 [Zhao B, Shrivastava M, Donahue N M, et al. 2020. High concentration of ultrafine particles in the](#)
1088 [Amazon free troposphere produced by organic new particle formation. Proceedings of the](#)
1089 [National Academy of Sciences \[J\], 117\(41\): 25344-25351.](#)
- 1090 Zhao B, Wang S, Donahue N M, et al. 2016. Quantifying the effect of organic aerosol aging and
1091 intermediate-volatility emissions on regional-scale aerosol pollution in China. Scientific
1092 Reports [J], 6(1): 28815.
- 1093 Zheng B, Tong D, Li M, et al. 2018. Trends in China's anthropogenic emissions since 2010 as the
1094 consequence of clean air actions. Atmos. Chem. Phys. [J], 18(19): 14095-14111.
- 1095 Zhu Y, Li K, Shen Y, et al. 2019. New particle formation in the marine atmosphere during seven cruise
1096 campaigns. Atmos. Chem. Phys. [J], 19(1): 89-113.
- 1097 Zhu Y, Sabaliauskas K, Liu X, et al. 2014. Comparative analysis of new particle formation events in less
1098 and severely polluted urban atmosphere. Atmospheric Environment [J], 98(655-664).
- 1099 Zhu Y, Shen Y, Li K, et al. 2021. Investigation of Particle Number Concentrations and New Particle
1100 Formation With Largely Reduced Air Pollutant Emissions at a Coastal Semi-Urban Site in
1101 Northern China. Journal of Geophysical Research: Atmospheres [J], 126(17): e2021JD035419.
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