

1 Substantially positive contributions of new particle formation to Cloud Condensation  
2 Nuclei under low supersaturation in China based on numerical model improvements

3

4 Chupeng Zhang<sup>1#</sup>, Shangfei Hai<sup>2,11#</sup>, Yang Gao<sup>1\*</sup>, Yuhang Wang<sup>3\*</sup>, Shaoqing Zhang<sup>4,2</sup>,  
5 Lifang Sheng<sup>2</sup>, Bin Zhao<sup>5</sup>, Shuxiao Wang<sup>5</sup>, Jingkun Jiang<sup>5</sup>, Xin Huang<sup>6</sup>, [Shen Xiaojing<sup>7</sup>](#),  
6 [Sun Junying<sup>7</sup>](#), Aura [Lupaseu<sup>7</sup>Lupascu<sup>8</sup>](#), Manish [Shrivastava<sup>8</sup>Shrivastava<sup>9</sup>](#), Jerome D.  
7 [Fast<sup>8</sup>Fast<sup>9</sup>](#), Wenxuan Cheng<sup>1</sup>, Xiuwen Guo<sup>1</sup>, Ming Chu<sup>1</sup>, Nan [Ma<sup>9</sup>Ma<sup>10</sup>](#), Juan  
8 [Hong<sup>9</sup>Hong<sup>10</sup>](#), Qiaoqiao [Wang<sup>9</sup>Wang<sup>10</sup>](#), Xiaohong Yao<sup>1</sup> and Huiwang Gao<sup>1</sup>

9

10 <sup>1</sup>Frontiers Science Center for Deep Ocean Multispheres and Earth System, and Key Laboratory of  
11 Marine Environmental Science and Ecology, Ministry of Education, Ocean University of China,  
12 and Laoshan Laboratory, Qingdao, 266100, China

13 <sup>2</sup>College of Oceanic and Atmospheric Sciences, Ocean University of China, Qingdao, 266100,  
14 China

15 <sup>3</sup>School of Earth and Atmospheric Sciences, Georgia Institute of Technology, Atlanta, GA, 30332,  
16 USA

17 <sup>4</sup>Frontiers Science Center for Deep Ocean Multispheres and Earth System, and Key Laboratory of  
18 Physical Oceanography, Ocean University of China, and Laoshan Laboratory, Qingdao, 266100,  
19 China

20 <sup>5</sup>State Key Joint Laboratory of Environment Simulation and Pollution Control, School of  
21 Environment, Tsinghua University, Beijing, 100084 China, and State Environmental Protection  
22 Key Laboratory of Sources and Control of Air Pollution Complex, Beijing 100084, China

23 <sup>6</sup>School of Atmospheric Sciences, Nanjing University, Nanjing, 210023, China

24 <sup>7</sup>[Institute<sup>7</sup>State Key Laboratory of Severe Weather & Key Laboratory of Atmospheric Chemistry  
25 of CMA, Chinese Academy of Meteorological Sciences, Beijing, 100081, China](#)

26 <sup>8</sup>[Institute](#) for Advanced Sustainability Studies, Potsdam D-14467, Germany

27 <sup>8</sup>[Atmospheric<sup>9</sup>Atmospheric](#) Sciences and Global Change Division, Pacific Northwest National  
28 Laboratory, Richland, WA, 99354, USA

29 <sup>9</sup>[Institute<sup>10</sup>Institute](#) for Environmental and Climate Research, Jinan University, Guangzhou,  
30 510000, China

31

32 [11CMA Earth System Modeling and Prediction Center, China Meteorological Administration,  
33 Beijing 100081, China](#)

34

#Authors contributed equally to this study.

35

\*To whom correspondence to: yanggao@ouc.edu.cn, yuhang.wang@eas.gatech.edu

36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65

---

### Abstract

New particle formation (NPF) and subsequent particle growth are important sources of condensation nuclei (CN) and cloud condensation nuclei (CCN). While ~~a number~~ of many observations have shown positive contributions of NPF to CCN at low supersaturation, negative NPF contributions were often simulated ~~in polluted~~ environment. Using the observations in a typical coastal city of Qingdao, we thoroughly evaluate the simulated number concentrations of CN and CCN using a NPF-explicit parameterization embedded in WRF-Chem model. In terms of CN, the initial simulation shows large biases of particle number concentrations at 10–40 nm (CN<sub>10–40</sub>) and 40–100 nm (CN<sub>40–100</sub>). By adjusting the process of gas-particle partitioning, including mass accommodation coefficient of sulfuric acid, the phase changes of primary organic aerosol emissions and the condensational amount of nitric acid, the concomitant improvement of the particle growth process yields a substantial reduction of overestimates of CN<sub>10–40</sub> and CN<sub>40–100</sub>. Regarding CCN, SOA formed from the oxidation of semi-volatile and intermediate volatility organic vapors (SI-SOA) yield is an important contributor. In the original WRF-Chem model with 20 size bins setting, the yield of SI-SOA is too high without considering the differences in oxidation rates of the precursors. Lowering the SI-SOA yield results in much improved simulations of the observed CCN concentrations. On the basis of the bias-corrected model, we find substantial positive contributions of NPF to CCN at low supersaturation (~0.2%) in Qingdao and over the broad areas of China, primarily due to the competing effects of increasing particle hygroscopicity surpassing that of particle size decrease. This study highlights the potentially much larger NPF contributions to CCN on a regional and even global basis.

66  
67  
68  
69  
70

## 71 **1. Introduction**

72 New particle formation (NPF) is a process in which gaseous vapors nucleate and  
73 form critical molecular clusters, followed by subsequent growth to larger sizes through  
74 condensation and coagulation (~~Kulmala et al., 2004; Kulmala et al., 2013; Lee et al.,~~  
75 ~~2019~~)([Kulmala et al., 2004; Kulmala et al., 2013; Lee et al., 2019](#)). Newly formed  
76 particles could effectively grow into the size of cloud condensation nuclei (CCN) under  
77 certain supersaturation (SS), which exerts an impact on the cloud microphysical process  
78 and global radiation balance (~~Merikanto et al., 2009; Gordon et al., 2017; Kerminen et~~  
79 ~~al., 2018; Ren et al., 2021~~)([Merikanto et al., 2009; Gordon et al., 2017; Kerminen et al.,](#)  
80 [2018; Ren et al., 2021](#)). In addition, the ~~high-efficiency~~[efficient](#) nucleation and  
81 explosive growth of particles may contribute to the formation of haze (~~Guo et al.,~~  
82 ~~2014~~)([Guo et al., 2014](#)), affecting air quality and human health (~~Yuan et al., 2015~~[Yuan](#)  
83 ~~et al., 2015; Chu et al., 2019~~[Chu et al., 2019](#); Kulmala et al., 2021).

84 The overestimate of condensation nuclei (CN) in numerical models are commonly  
85 seen, despite the attempt to ~~correctify~~[rectify](#) the bias (~~Matsui et al., 2013~~)([Matsui et al.,](#)  
86 [2013](#); Arghavani et al., 2022). It is a common way to reduce the nucleation rate which  
87 may reduce the particle number concentration in proportion (~~Matsui et al.,~~  
88 ~~2013~~)([Matsui et al., 2013](#)). For instance, in the study of NPF in East Asia in the spring  
89 of 2009, even after lowering the nucleation rate in a regional model of WRF-Chem  
90 applied in their study, the reduced number concentration of particles at 10–130 nm  
91 remained to be overestimated (~~Matsui et al., 2013~~)([Matsui et al., 2013](#)). Using the same  
92 regional model and a similar method to reduce the nucleation rate, Arghavani et al.  
93 (2022) found particle number concentration at 10–100 nm was still overestimated by  
94 nearly one order of magnitude, despite the effectiveness to reduce the overestimates for  
95 the smaller particles such as 2.5–10 nm. In addition to the rate of NPF, the growth

---

96 process of particles also has a crucial effect on particle number concentration and size  
97 distribution. In this process, the condensation of some chemical species such as sulfuric  
98 acid, nitrate and organic gases on particles plays a major role in particle growth (~~Yao et  
99 al., 2018; Lee et al., 2019; Li et al., 2022~~)([Yao et al., 2018; Lee et al., 2019; Li et al.,  
100 2022](#)), and the uncertainty of their condensation amount may lead to the bias of CN  
101 simulation.

102 In addition to CN, there are large discrepancies in the predicted CCN between the  
103 numerical models and observational results. Furthermore, as an important source of  
104 CCN (~~Merikanto et al., 2009~~)([Merikanto et al., 2009](#)), the contribution of nucleation to  
105 CCN quantified by numerical models is also highly uncertain. For example, in terms of  
106 predicting CCN, ~~Fanourgakis et al. (2019)~~[Fanourgakis et al. \(2019\)](#) evaluated the CCN  
107 concentrations simulated by 16 global aerosol–climate and chemistry transport models  
108 with observations at 9 sites in Europe and Japan from 2011 to 2015, and found that all  
109 models underestimated CCN concentrations with a mean normalized mean bias (NMB)  
110 of -36% at low supersaturation (SS=0.2%). ~~Models~~[WRF-Chem models](#) also tend to  
111 underestimate the contribution of NPF on CCN, especially at low supersaturation. The  
112 continuous observation of CCN concentrations throughout the year (July 2008–June  
113 2009) carried out in Hyytiälä, Finland, showed that under low SS, nucleation enhanced  
114 the CCN by 106% and 110% at SS=0.1% and 0.2% respectively (~~Sihto et al.,  
115 2011~~)([Sihto et al., 2011](#)). Observations acquired in Beijing from July 12 to September  
116 25, 2008, also suggested that nucleation significantly increases CCN at all  
117 supersaturations, even when supersaturation is low (i.e., 0.07% and 0.26%). Thus, the  
118 occurrence of NPF enhanced CCN by a factor of 1.7 and 2.2, respectively (~~Yue et al.,  
119 2011~~)([Yue et al., 2011](#)).

120 However, previous numerical experiments behave oppositely. ~~For instance, Matsui  
121 et al. (2011) quantified the contribution of nucleation to CCN using WRF-chem in  
122 Beijing in August and September 2006 and found reduced CCN under low SS, e.g.,  
123 when SS=0.02%, the concentration of CCN is reduced by up to ~50%. They attributed  
124 this to the fact that the small particles produced by nucleation may inhibit the growth  
125 of the preexisting particles~~[For instance, Matsui et al. \(2011\) quantified the contribution](#)

---

126 [of nucleation to CCN using WRF-chem in Beijing in August and September 2006 and](#)  
127 [found reduced CCN under low SS, e.g., when SS=0.02%, the concentration of CCN is](#)  
128 [reduced by up to ~50%. They attributed this to the fact that the small particles produced](#)  
129 [by nucleation may inhibit the growth of the preexisting particles](#) (~~Matsui et al.,~~  
130 ~~2011~~)(Matsui et al., 2011). Similarly, ~~Dong et al. (2019)~~Dong et al. (2019) conducted  
131 NPF simulations with the WRF-Chem for the summer of 2008 focusing on the Midwest  
132 of the United States, -and found that the nucleation resulted in decreased CCN at low  
133 supersaturation (SS=0.1%). Besides, a study carried out for East Asia in 2009 also  
134 indicated that at low supersaturation (e.g. SS=0.1%), nucleation has little impact on  
135 CCN (~~Matsui et al., 2013~~)(Matsui et al., 2013). The contrasting effects of nucleation on  
136 CCN at low supersaturations in model and observations is not explained in these  
137 previous studies.

138 At the stage of particle growth, secondary organic aerosol (SOA) formed by  
139 atmospheric oxidation of organic vapors is a major contributor to particle growth to  
140 CCN-related sizes (~~Liu and Matsui, 2022; Qiao et al., 2021~~)(Liu and Matsui, 2022;  
141 [Qiao et al., 2021](#)). SOA formed by multi-generational gas-phase oxidation of semi-  
142 volatile and intermediate volatility organic compounds (S/IVOC) is called SI-SOA  
143 (~~Jimenez et al., 2009; Zhang et al., 2007~~)(Jimenez et al., 2009; Zhang et al., 2007).-Zhao  
144 ~~et al. (2016)~~. [Zhao et al. \(2016\)](#) made a comprehensive assessment of the roles of  
145 various SOA precursors in SOA formation in real atmosphere in China in 2010, and the  
146 results demonstrated that evaporated POA and IVOC (i.e. S/IVOC) made a significant  
147 contribution to SOA, contributing up to 82% to the average SOA concentration in  
148 eastern China. However, the effect of SI-SOA on CCN has not been fully studied.

149 In this paper, WRF-Chem was applied to simulate the effect of the NPF on CCN  
150 in Qingdao in February 2017. The simulated results from the WRF-Chem model are  
151 firstly compared with observations in Qingdao, exhibiting large biases in CN. This is  
152 followed by an improvement through a few processes. At the end, the impact of SI-  
153 SOA yield and nucleation on CCN is investigated.

## 154 **2. Data and methods**

### 155 **2.1 Observations**

---

156 The measurements used in this study were carried out over the sampling site from  
157 February 5 to 24, 2017 at the campus of Ocean University of China (36°09'37"N,  
158 120°29'44"E ) in Qingdao, which is surrounded by residential buildings and is situated  
159 about 10 km away from the city center. A fast mobility particle sizer (FMPS, TSI Model  
160 3091) was applied to measure the aerosol particle size distribution for the size range of  
161 5.6 nm to 560 nm ([Liu et al., 2014b](#))([Liu et al., 2014b](#)).~~The size-resolved CCN activity~~  
162 ~~is measured by a cloud condensation nuclei counter (CCNc) at three different~~  
163 ~~supersaturations (0.2%, 0.4% and 0.6%) and each supersaturation lasts for 5 minutes.~~  
164 ~~More information about the CCN measurement can be found in Li et al. (2015).~~ The  
165 bulk CCN concentration is measured by a cloud condensation nuclei counter at three  
166 different supersaturations (0.2%, 0.4% and 0.6%) and each supersaturation lasts for 20  
167 minutes. More information about the CCN measurement can be found in Li et al. (2015).  
168 The urban site in Beijing is located on the roof of the building of the Chinese Academy  
169 of Meteorological Sciences (CAMS, 39°95'N, 116°33'E) in the campus of the China  
170 Meteorological Administration, close to the main road with heavy traffic. The rural site  
171 is Gucheng (GC, 39°08'N, 115°40'E), located in Hebei Province, surrounded by  
172 farmland, and is a representative station of the severity of air pollution in Beijing Tianjin  
173 Hebei region. The particle number size distribution of these two sites in the range of 4–  
174 850 nm is measured by a Tandem Scanning Mobility Particle Sizer (TSMPS), and more  
175 information about the instruments can be found in Shen et al. (2018).

## 176 2.2 Model configurations

177 WRF-Chem version 3.9 is used to simulate NPF events, with the main physical  
178 and chemical parameterization settings summarized in Table 1. The spatial resolution  
179 is 36 km by 36 km with 35 vertical layers and a model top at 50 hPa. The regional  
180 model simulations at a higher spatial resolution may be desirable in future when urban  
181 pollution is focused. A continuous run from February 1 to 25, 2017, was conducted,  
182 with the first five-day results as the spin-up and discarded in the analysis.

183

184

Table 1 WRF-Chem model configurations used in this work

Model configuration	
Microphysics	Morrison 2-moment microphysics scheme ( <del>Morrison et al., 2009</del> )(Morrison et al., 2009)
Planetary Boundary Layer (PBL)	YSU boundary layer scheme ( <del>Hong et al., 2006</del> )(Hong et al., 2006)
Longwave and Shortwave Radiation	RRTMG longwave and shortwave radiation ( <del>Iacono et al., 2008</del> )(Iacono et al., 2008)
Land model	Unified Noah Land Surface scheme ( <del>Chen and Dudhia, 2000; Tewari et al., 2016</del> )(Chen and Dudhia, 2000; Tewari et al., 2016)
Cumulus	Grell-3D cumulus parameterization scheme ( <del>Grell, 1993</del> )(Grell, 1993)
Aerosol module	MOSAIC module ( <del>Zaveri et al., 2008; Matsui et al., 2011</del> )(Zaveri et al., 2008; Matsui et al., 2011)
Gas-phase Chemistry	SAPRC-99 gas-phase chemistry scheme ( <del>Carter, 2000</del> )(Carter, 2000)

185

186 The meteorological initial and boundary conditions are driven by Climate Forecast  
 187 System model version 2 (CFSv2; (~~Saha et al., 2014~~)(Saha et al., 2014)) reanalysis  
 188 developed by National Centre for Environmental Prediction (NCEP). The initial and  
 189 boundary chemical conditions of WRF-Chem are provided by Community Atmosphere  
 190 Model with Chemistry (CAM-Chem; (~~Buchholz et al., 2019~~)(Buchholz et al., 2019)).  
 191 Anthropogenic emissions for the year of 2017 are obtained from the Multiresolution  
 192 Emission Inventory for China (MEIC, <http://www.meicmodel.org/>) emission dataset  
 193 (~~Li et al., 2017; Zheng et al., 2018~~)(Li et al., 2017; Zheng et al., 2018).

194 The Model for Simulating Aerosol Interactions and Chemistry (MOSAIC) was  
 195 used to delineate dynamic gas-particle mass transfer to represent the condensation  
 196 growth of aerosol (~~Zaveri et al., 2008~~)(Zaveri et al., 2008). The gas-particle partitioning  
 197 of gas species on particles is regulated by the mass transfer rate, which is related to  
 198 mass accommodation coefficient ( $\alpha$ ), a parameter involved in the model representing

199 the probability of gas molecules entering the bulk liquid phase (Pöschl et al.,  
 200 1998)(Pöschl et al., 1998). The original setting of  $\alpha$  for all condensing species for all  
 201 size bins  $a$  in MOSAIC is 0.1 (Zaveri et al., 2008)(Zaveri et al., 2008). In the default  
 202 release of WRF-Chem, MOSAIC was implemented in the sectional framework with  
 203 aerosol size distributions divided into 4 or 8 size bins spanning 39 nm to 10  $\mu\text{m}$  in  
 204 diameter. To explicitly express the nucleation and the growth of newly formed particles,  
 205 the aerosol size range in the MOSAIC module was extended from 1 nm to 10  $\mu\text{m}$ , with  
 206 the number of aerosol size bins increased to 20 (Matsui et al., 2011; Matsui et al., 2013;  
 207 Lupascu et al., 2015; Lai et al., 2022)(Matsui et al., 2011; Matsui et al., 2013; Lupascu  
 208 et al., 2015; Lai et al., 2022)-. The calculation method of CCN concentration in the  
 209 WRF-chem model is referred to the study of Matsui et al. (2011). Based on Köhler  
 210 theory, CCN concentrations under the three given supersaturations of 0.2%, 0.4% and  
 211 0.6% were calculated. The critical supersaturation ( $S_c$ ) of each size bin in the WRF-  
 212 chem model was calculated by the following formula:

$$S_c = \sqrt{\frac{4 \times a^3}{27 \times r^3 \times \kappa}} \quad (1)$$

$$a = \frac{2 \times \sigma}{R_v \times T \times \rho_w} \quad (2)$$

215 Where  $\alpha$  (m) is the coefficient of the Kelvin effect,  $\kappa$  is the volume-averaged  
 216 hygroscopicity, calculated using these values in Table 1,  $r$  (m) is the dry diameter,  $\sigma$  is  
 217 droplet surface tension over water ( $0.076 \text{ N m}^{-1}$ ),  $R_v$  is the gas constant for water vapor  
 218 ( $461.6 \text{ J K}^{-1} \text{ kg}^{-1}$ ),  $T$  (K) is the air temperature, and  $\rho_w$  is the density of water ( $1000 \text{ kg}$   
 219  $\text{m}^{-3}$ ).

220 Table 2 Hygroscopicity Parameters ( $\kappa$ ) in the WRF-Chem Model

<u>Species</u>	<u>Hygroscopicity (<math>\kappa</math>)</u>
<u>Sulfate</u>	<u>0.5</u>
<u>Ammonium</u>	<u>0.5</u>
<u>Nitrate</u>	<u>0.5</u>
<u>Black carbon</u>	<u><math>10^{-6}</math></u>
<u>Primary organic aerosol</u>	<u>0.14</u>



<u>Other inorganics</u>	<u>0.14</u>
<u>Sodium</u>	<u>1.16</u>
<u>Chloride</u>	<u>1.16</u>

221

222 The chemical aging process of organic aerosols (OA) is modeled by the volatility  
 223 basis set (VBS) approach, which was widely used in air quality models to represent  
 224 complex mixtures of thousands of organic species ([Donahue et al., 2006](#); [Shrivastava](#)  
 225 [et al., 2011](#); [Chrit et al., 2018](#))([Donahue et al., 2006](#); [Shrivastava et al., 2011](#); [Chrit et](#)  
 226 [al., 2018](#)). The VBS method classifies compounds according to the effective saturation  
 227 concentration ( $C_c^*$ ), which represents the proportion of the ~~component~~  
 228 the gas phase to the particle phase ([Donahue et al., 2006](#))([Donahue et al., 2006](#)), and  
 229 species with higher  $C_c^*$  values are more volatile. The oxidation of highly volatile  
 230 precursors to form relatively low volatile components represents the aging process of  
 231 OA. OA consists of directly emitted primary organic aerosols and photochemically  
 232 produced secondary organic aerosols (SOA) ([Shrivastava et al., 2011](#))([Shrivastava et](#)  
 233 [al., 2011](#)). In this study, the simplified 2-species VBS mechanism was applied to the  
 234 simulation of SOA, during which primary organic aerosol was represented by two  
 235 species based on volatility with effective saturation concentration  $C_c^*$  values (at 298 K  
 236 and 1 atm) of  $10^{-2}$  and  $10^5 \mu\text{g m}^{-3}$  ([Shrivastava et al., 2011](#))([Shrivastava et al., 2011](#)).  
 237 Primary organic aerosols with  $C_c^*$  of  $10^5 \mu\text{g m}^{-3}$  refers to S/IVOC, which is in the gas  
 238 phase under most atmospheric conditions due to its high volatility, while for those  
 239 primary organic ~~aerosols~~ with  $C_c^*$  of  $10^{-2} \mu\text{g m}^{-3}$ , is treated as gas phase as well  
 240 in the original model. The SOA formed by photochemical oxidation of S/IVOC  
 241 precursors is called SI-SOA and the SOA formed by oxidation of VOC precursors is  
 242 named V-SOA. In the simplified 2-species VBS mechanism, SI-SOA ( ~~$C_c^*$  of  $10^{-2} \mu\text{g m}^{-3}$~~ )  
 243 ~~is formed by the oxidation reaction of S/IVOC precursors ( $C_c^*$  of  $10^5 \mu\text{g m}^{-3}$ ) and OH~~  
 244 ~~with an oxidation rate constant of  $4 \times 10^{-11} \text{cm}^3 \text{molec}^{-1} \text{s}^{-1}$ . A detailed description of 2-~~  
 245 ~~species VBS mechanism can be found in [Shrivastava et al. \(2011\)](#).~~  $C_c^*$  of  $10^{-2} \mu\text{g m}^{-3}$  is  
 246 ~~formed by the oxidation reaction of S/IVOC precursors ( $C_c^*$  of  $10^5 \mu\text{g m}^{-3}$ ) and OH with~~

an oxidation rate constant of  $4 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$ . The equations for controlling the oxidation of S/IVOC precursors are as follows:



where POA(g) denotes primary organic aerosols with  $c^*$  of  $10^5 \mu\text{g m}^{-3}$ , which reacts with OH to form SI-SOA(g) with  $c^*$  of  $10^{-2} \mu\text{g m}^{-3}$ . Subscripts  $c$  and  $o$  represent the non-oxygen and oxygen parts respectively of given species and  $e$  is either the biomass or anthropogenic emission sector. In addition, SVOC and IVOC emissions corresponding to both anthropogenic and biomass burning emissions are derived based on constant emission ratio of S/IVOC to POA (Shrivastava et al., 2011). A detailed description of 2-species VBS mechanism can be found in Shrivastava et al. (2011).

### 2.3 Model sensitivity formulations

Three sets of sensitivity tests are designed and listed in ~~Table 2~~Table 3. The purposes of the three sets of experiments are as follows: (1) Adjust the condensation growth process of ultrafine particles in WRF-Chem model (Base, MAC, POA, NOCD, RACD, with details in Table 23).; (2) Explore the effect of SI-SOA yield on CCN (Low-Yield and High-Yield); (3) Study the effect of nucleation process on CCN under the change of SI-SOA yield (Low-Yield and High-Yield and their corresponding cases without nucleation parameterization, i.e., Low\_nuoff and High\_nuoff). Each scenario will be explained in conjunctions with the results.

Table 2-3 The sensitivity tests involved in this study

Purposes	<u>Simulation scenarios</u>	<u>Simulation scenarios</u>	Description
Adjust the condensation growth process of ultrafine particles		Base	<del>Simulation with the default setting with nucleation coefficient set as <math>2 \times 10^{-6} \text{ s}^{-1}</math>, the same as Lai et al. (2022)</del> <u>Simulation with the default setting with nucleation coefficient set as <math>2 \times 10^{-6} \text{ s}^{-1}</math>, the same as Lai et al. (2022)</u>
		Mass accommodation coefficient (MAC)	It is the same as Base except that the mass adjustment coefficient ( $\alpha$ ) of gaseous sulfuric acid is adjusted from 0.1 to 0.65.
		POA emission phase (PEP)	It is the same as MAC except that the phase of POA is changed from gas phase to particle phase.
		No condensation (NOCD)	It is the same as PEP except that no $\text{NH}_4\text{NO}_3$ condenses on particles below 40 nm.
		Ratio method for condensation (RACD)	<del>It is the same as PEP except that the condensation of <math>\text{NH}_4\text{NO}_3</math> on particles</del>

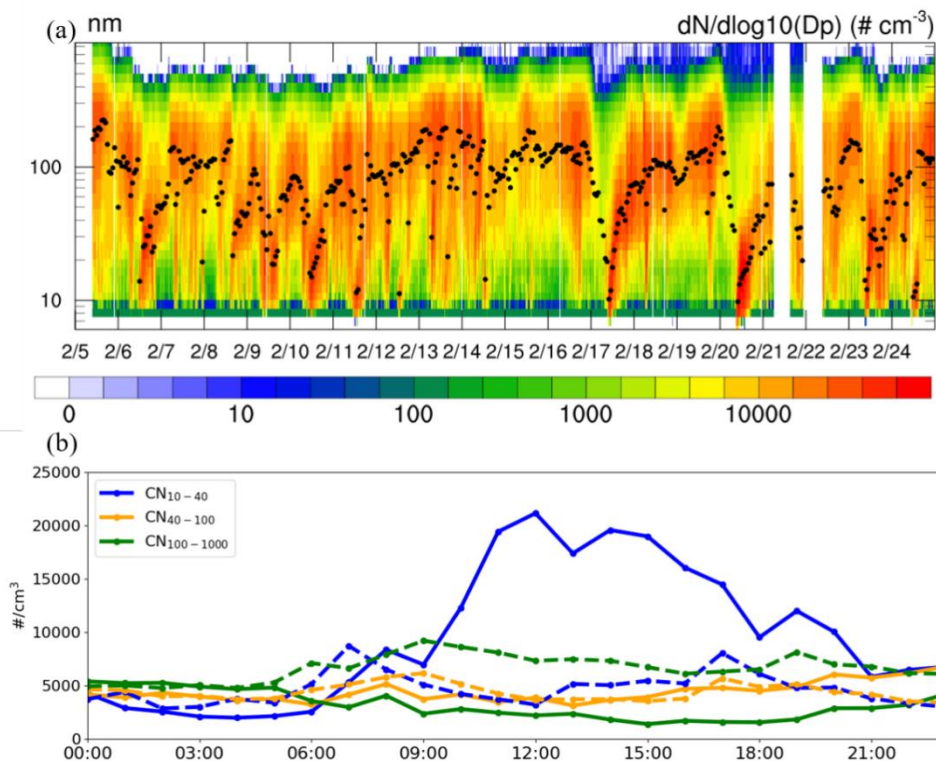
<p>Explore the effect of SI-SOA yield on CCN (Explore the effect of nucleation process on CCN under the change of SI-SOA yield)</p>	<p>High-<u>Yield</u></p> <p>Low-Yield</p>	<p><u>below 40 nm is reduced according to the ratio of acid particles to total particles reported in Wang et al. (2014). It is the same as PEP except that the condensation of NH<sub>4</sub>NO<sub>3</sub> on particles below 40 nm is reduced according to the ratio of acid particles to total particles reported in Wang et al. (2014).</u></p> <p>Simulation with high oxidation rate of SI-SOA formation with reaction rate constant of <math>5 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}</math></p> <p>Simulation with low oxidation rate of SI-SOA formation with reaction rate constant of <math>2 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}</math></p>
<p>Explore the effect of nucleation process on CCN under the change of SI-SOA yield</p>	<p>High_NUCOFF</p> <p>Low_NUCOFF</p>	<p>Simulations without nucleation parameterizations based on High-<u>Yield</u></p> <p>Simulations without nucleation parameterizations based on Low-Yield</p>

280

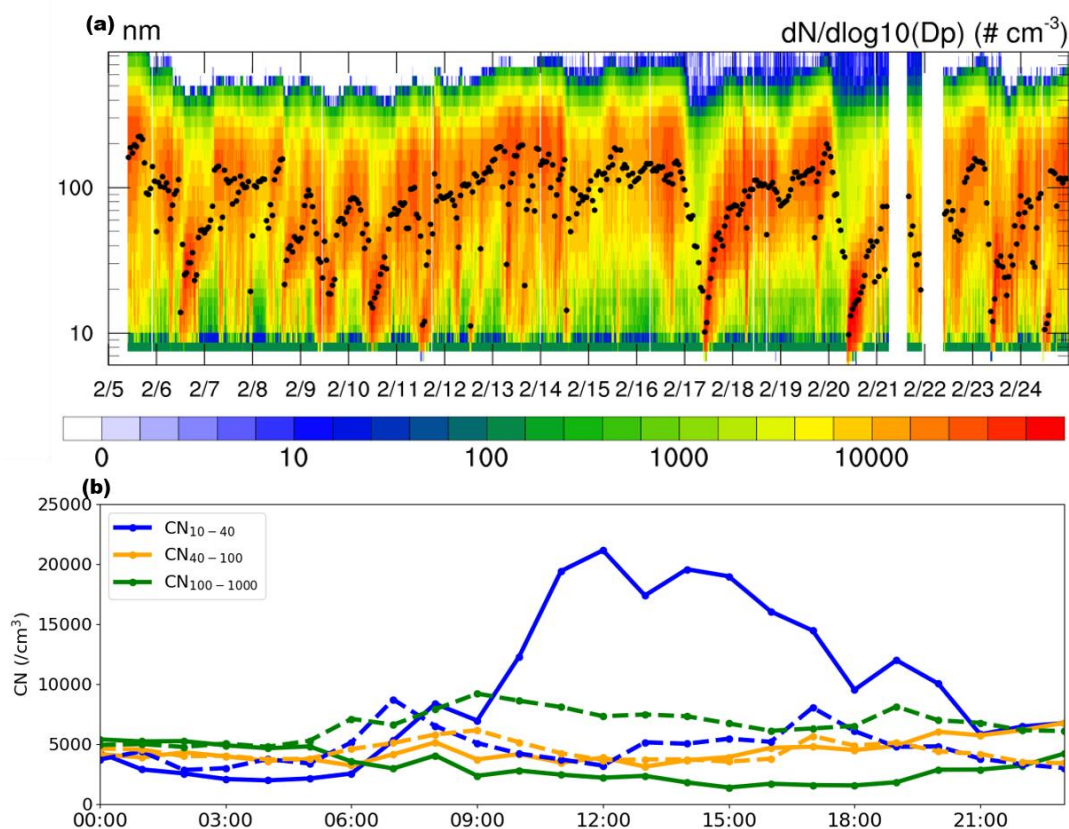
### 281 3. Results

#### 282 3.1 Observational analysis

283 Based on the criteria (Dal Maso et al., 2005; Kulmala et al., 2012)(Dal Maso et al.,  
284 2005; Kulmala et al., 2012), NPF is defined as an event with the emergence of a  
285 nucleation mode with particle diameters smaller than 25 nm, lasting for 2 hours or more,  
286 followed in general by a continuous particle growth. Six NPF events were identified in  
287 February 2017 in Qingdao, on the days of 6, 9, 10, 17, 20 and 23 (Fig. 1a), yielding a  
288 frequency of ~30% and displaying a typical banana-shaped growth of particles in the  
289 particle number size distribution. Compared to a few other studies on NPF frequency  
290 in Qingdao, the results in this study are to a large extent consistent with that in the fall  
291 of 2012–2013 (30%; (Zhu et al., 2019)(Zhu et al., 2019)), slightly higher than that in  
292 summer 2016 (22%; (Zhu et al., 2019)(Zhu et al., 2019)) and lower than that in spring  
293 of 2010 (41%; (Liu et al., 2014c)(Liu et al., 2014c)). The higher frequency in spring in  
294 Qingdao is consistent with the observational results at different stations in the Northern  
295 Hemisphere in Nieminen et al. (2018).



296



297

298 Fig. 1. Distribution of particle number concentration. (a) Temporal evolution of particle  
 299 size distributions (colored shading) and geometric median diameter (GMD; dots in  
 300 black) in Qingdao on February 5-24, 2017. (b) The mean diurnal variation of CN<sub>10-40</sub>  
 301 (blue), CN<sub>40-100</sub> (orange) and CN<sub>100-1000</sub> (green) composited during the NPF (solid lines)  
 302 and non-NPF (dashed lines) days on February 5-24, 2017. All times are local times (LT)  
 303

304 During the six NPF events identified in February in Qingdao, the mean diurnal  
 305 cycle of CN<sub>10-40</sub> (10–40 nm) particles exhibits triple peaks (solid blue in Fig. 1b), in the  
 306 morning (8:00 LT), noon (12:00–14:00 LT) and evening (19:00 LT), respectively. A  
 307 comparable three-peak feature was also observed in earlier years during 2016-2018 in  
 308 Qingdao (Zhu et al., 2021). (Zhu et al., 2021). The morning and evening peaks of CN<sub>10-</sub>  
 309 40, with values of  $\sim 5300 \text{ cm}^{-3}$  and  $\sim 12000 \text{ cm}^{-3}$ , respectively, are likely caused by the  
 310 primary emissions from traffic and cooking activities (Wu et al., 2021a; Wang et al.,  
 311 2022; Cai et al., 2020)(Wu et al., 2021a; Wang et al., 2022; Cai et al., 2020). The  
 312 occurrence of NPF starts approximately at 9:00 am LT, accompanied by a substantial  
 313 increase in CN<sub>10-40</sub> compared with non-NPF days (solid vs. dashed lines, in blue),

---

314 yielding a peak around noon ( $20000 \text{ cm}^{-3}$  during 12:00–14:00 LT). In addition, larger  
315 particles (e.g.,  $\text{CN}_{40-100}$  and  $\text{CN}_{100-1000}$ ) displayed a slow or no increase in the afternoon.

316

### 317 **3.2 Model improvement in particle number concentration simulations**

318 Particle number concentrations, primarily in two ranges of 10–40 nm and 40–100  
319 nm, are commonly simulated with large biases. In the smaller size range (10–40 nm),  
320 the particle number concentration is associated with NPF and [particlegrowth](#) growth.  
321 During NPF, despite differences among the formation mechanisms,  $\text{H}_2\text{SO}_4$  is  
322 considered the common species ([Yu, 2005; Lovejoy et al., 2004](#))([Yu, 2005; Lovejoy et](#)  
323 [al., 2004](#)), which often suffer large biases ([Cai et al., 2016; Matsui et al., 2011](#))([Cai et](#)  
324 [al., 2016; Matsui et al., 2011](#)). In the size range of 40–100 nm, the particle number  
325 concentration is primarily affected by the condensation growth of particles below 40  
326 nm, which is closely related to chemical components such as SOA and nitrate. [Prior to](#)  
327 [the evaluation of particle number concentration, we first evaluate the compositions of](#)  
328 [PM<sub>2.5</sub> and criteria air pollutants including PM<sub>2.5</sub>, PM<sub>10</sub>, O<sub>3</sub>, SO<sub>2</sub>, CO, and NO<sub>2</sub>, showing](#)  
329 [relatively low biases compared to observations \(section S1 and Fig. S1 and Fig. S2 of](#)  
330 [the supporting information\).](#)

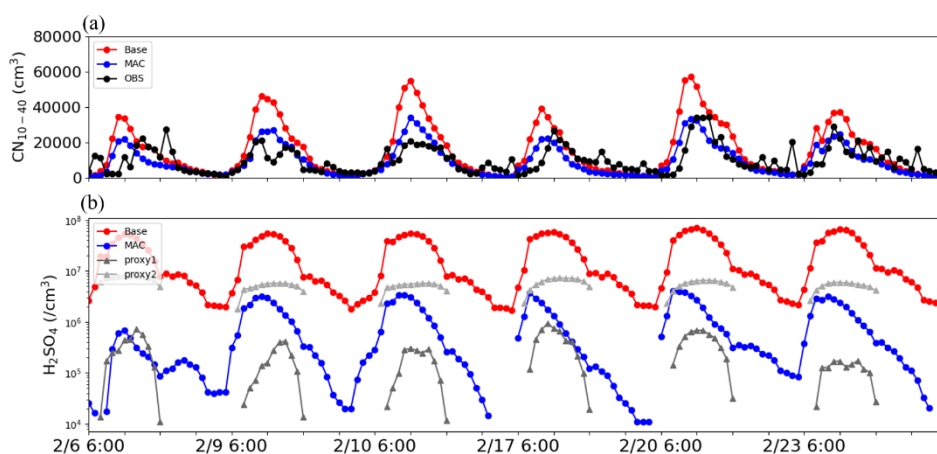
331

#### 332 **3.2.1 Bias correction of particle number concentration at 10–40 nm**

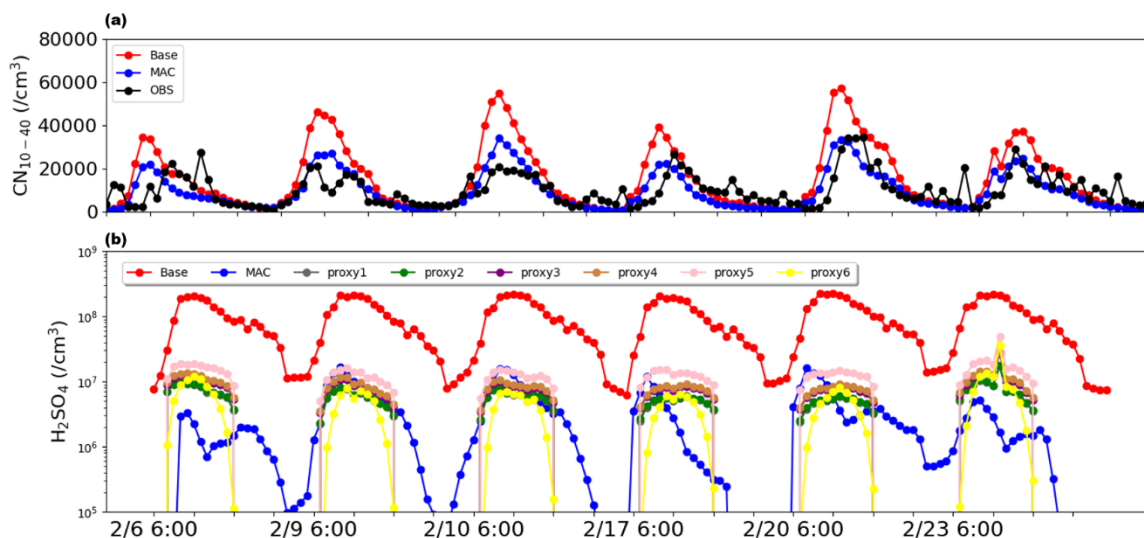
333 In this study, as shown in Fig. 2, comparisons of  $\text{CN}_{10-40}$  between simulations (red  
334 line in Fig. 2a) and observations (black line in Fig. 2a) results of the six NPF events  
335 mentioned in the previous section in Qingdao in February 2017 indicate that model  
336 overestimates  $\text{CN}_{10-40}$  with mean fractional bias (MFB) of 48%. As one of the major  
337 processes affecting the particle number concentration of 10–40 nm, nucleation is  
338 governed by the particle nucleation rate of 1 nm particles ( $\text{cm}^{-3} \text{ s}^{-1}$ ), which is closely  
339 associated with the concentration of  $\text{H}_2\text{SO}_4$ . For instance, in a commonly applied  
340 activation mechanism, the nucleation rate calculated by  $J^* = K_{\text{ACT}} \times [\text{H}_2\text{SO}_4]$ , [where](#).  
341 [Note that](#)  $K_{\text{ACT}}$  is the nucleation coefficient [considering the physical properties and](#)  
342 [chemical species of nucleation process under different environments, indicating that a](#)  
343 [lumped chemical species are included in the scheme reflected primarily in the](#)



344 nucleation coefficient  $k$ , set as  $2 \times 10^{-6} \text{ s}^{-1}$  based on previous studies (Sihto et al.,  
 345 2006; Riipinen et al., 2007)(Sihto et al., 2006; Riipinen et al., 2007). Dong et al. (2019).  
 346 Dong et al. (2019) simulated NPF occurring in the summer of 2008 in the United States  
 347 using the NFPNPF-explicit WRF-Chem based on the activation mechanism, which  
 348 overestimated the particle number concentration at 10–63 nm by nearly doubled, even  
 349 when the  $K_{\text{ACT}}$  decreased by one order of magnitude (set at a very low value of  $10^{-7} \text{ s}^{-1}$ ).  
 350 Therefore, it is likely that the overestimation of particle number concentration in  
 351 the smaller particle size segment is probably due to the bias of simulated sulfuric acid.



Figure



355 Fig. 2. Time series of (a)  $\text{CN}_{10-40}$  on NPF days, where red and blue represent Base and  
 356 MAC simulation results respectively, and black represents observation results, and (b)  
 357 sulfuric acid gas concentration obtained by simulation and by proxies (dark grey: Eq.



358 ~~1; light grey5; green: Eq. 26; purple: Eq. 7; brown: Eq. 8; pink: Eq. 9; yellow: Eq. 10).~~

359 All times are in local times.

360  
361 Measurement of sulfuric acid gases in the lower troposphere is challenging due to  
362 the generally low ambient concentration of sulfuric acid ( $10^6$ – $10^7$  molecule  $\text{cm}^{-3}$ ).  
363 Different methods have been proposed to estimate ambient sulfuric acid concentrations  
364 based on observations such as  $\text{SO}_2$  (Petäjä et al., 2009; Lu et al., 2019)(Petäjä et al.,  
365 2009; Lu et al., 2019; Mikkonen et al., 2011). For instance, Petäjä et al. (2009) proposed  
366 a linear method (Eq. 1) to approximate observed  $\text{H}_2\text{SO}_4$  concentration in Hyytiälä,  
367 southern Finland. Moreover, a recent study by Lu et al. (2019) proposed a nonlinear  
368 method (Eq. 2), indicating that compared to the linear method in Petäjä et al. (2009),  
369 the nonlinear relationship can provide more accurate  $\text{H}_2\text{SO}_4$  concentration in Beijing  
370 during February–March 2018 period. In this study, we adopt both of the above two proxy  
371 methods (referred as proxy1 and proxy2. For instance, Petäjä et al. (2009) proposed a  
372 linear method to approximate observed  $\text{H}_2\text{SO}_4$  concentration in Hyytiälä, southern  
373 Finland. Moreover, a recent study by Lu et al. (2019) proposed a nonlinear method to  
374 construct a number of proxies for gaseous sulfuric acid concentration (Eq. 5–9),  
375 indicating that compared to the linear method in Petäjä et al. (2009), the nonlinear  
376 relationship can provide more accurate  $\text{H}_2\text{SO}_4$  concentration in Beijing during  
377 February–March 2018 period. In addition, we also used another sulfuric acid nonlinear  
378 proxy (Eq. 10) based on long-term observations in Germany, Finland, the United States,  
379 etc. (Mikkonen et al., 2011). In this study, we adopt the above six nonlinear proxy  
380 methods (referred as proxy5 to proxy10) to estimate  $\text{H}_2\text{SO}_4$  in Qingdao.

$$381 \quad [H_2SO_4] = k \frac{[SO_2] \cdot UVB(Glob)}{CS} \quad (1)$$

$$382 \quad [H_2SO_4] = 280.05 \cdot UVB^{0.14} [SO_2]^{0.4} \quad (2)$$

$$383 \quad [H_2SO_4] = 515.74 \times [SO_2]^{0.38} \times \text{Radiation}^{0.14} \times CS^{0.03} \quad (5)$$

$$384 \quad [H_2SO_4] = 280.05 \cdot \text{Radiation}^{0.14} [SO_2]^{0.40} \quad (6)$$

$$385 \quad [H_2SO_4] = 9.95 \times [SO_2]^{0.39} \times \text{Radiation}^{0.13} \times CS^{-0.01} \times [O_3]^{0.14} \quad (7)$$

$$386 \quad [H_2SO_4] = 14.38 \times [SO_2]^{0.38} \times \text{Radiation}^{0.13} \times [O_3]^{0.14} \quad (8)$$

$$[H_2SO_4] = 0.0013 \times [SO_2]^{0.38} \times \text{Radiation}^{0.13} \times CS^{-0.17} \times ([O_3]^{0.14} + [NO_x]^{0.41}) \quad (9)$$

$$[H_2SO_4] = 8.21 \times 10^{-3} \times [SO_2]^{0.62} \times \text{Radiation} \times (CS \times RH)^{-0.13} \quad (10)$$

where  $k$  is an empirically derived factor with value of  $2.3 \times 10^{-9} \text{ m}^2 \text{ W}^{-1} \text{ s}^{-1}$ ;  $[SO_2]$ ,  $[O_3]$  and  $[NO_x]$  (molecule  $\text{cm}^{-3}$ ) represents the concentration of observed  $SO_2$ , UVB(Glob) and UVB ( $\text{W m}^{-2}$ ) are solar radiation in global  $O_3$  and UV-B range (280–320 nm) radiation  $NO_x$ , respectively, “Radiation” ( $\text{W m}^{-2}$ ) is global radiation. RH (%) is the relative humidity, and CS ( $\text{s}^{-1}$ ) is the condensation sink, which is calculated based on observed particle distribution.

The simulated  $H_2SO_4$  concentration from the Base simulation (dots in Fig. 2b) is compared with observations obtained by proxies (see Fig. 2b), indicating that Base simulations apparently overestimate by one order of magnitude compared to the higher  $H_2SO_4$  estimated by proxy 2 proxies. The overestimation has been frequently reported previously, i.e., over Beijing (Matsui et al., 2011)(Matsui et al., 2011), which ascribes the bias to the overestimation of the  $SO_2$  concentration. In a more recent study, the sensitivity of  $H_2SO_4$  to  $SO_2$  is tested, and the result shows that even when  $SO_2$  is reduced to an unrealistically low level, the simulated  $H_2SO_4$  is still more than one order of magnitude higher than the observed value (Lai et al., 2022), suggesting that the  $SO_2$  concentration cannot fully explain the overestimates.

In addition to the precursor of  $H_2SO_4$ , the mass accommodation coefficient ( $\alpha$ ), representing the probability of impaction of a gaseous molecule on a liquid surface and entering the bulk liquid phase, is another important factor affecting the concentration of sulfuric acid gas. In the public release of WRF-Chem, mass accommodation coefficient is typically set to a low value of 0.1 for all gas species under different volatility during the condensation process, including  $H_2SO_4$  (Davidovits et al., 2004; Zaveri et al., 2008)(Davidovits et al., 2004; Zaveri et al., 2008). Recent studies indicate that the low mass accommodation coefficient value may not be applicable to the low volatile gases, which tend to have a mean mass accommodation coefficient value of 0.7 and close to the unity (Krechmer et al., 2017)(Krechmer et al., 2017). In fact, an earlier study has indicated based on experimental determination, the mass accommodation coefficient of  $H_2SO_4$  vapor in sulfuric acid aqueous solution was

417 measured, and the best fit value was 0.65 (Pöschl et al., 1998)(Pöschl et al., 1998).  
418 Accordingly, a sensitivity simulation was conducted by adjusting the mass  
419 accommodation coefficient of H<sub>2</sub>SO<sub>4</sub> from 0.1 to 0.65, referred to as MAC.

420 This simulation brought the H<sub>2</sub>SO<sub>4</sub> concentration (see Fig. 2b) much closer to the  
421 calculated results from proxies, and the corresponding biases reduced by approximately  
422 an order of magnitude. Notably, the MAC simulation decreases the overestimate of  
423 sulfuric acid gas concentration, resulting in a lower particle formation rate. The MAC  
424 simulation also significantly reduces overestimate of CN<sub>10-40</sub> (Fig. 2b2a), and  
425 ~~MFB~~ mean fractional bias compared to observations decreases from 48% to 1%.

### 427 3.2.2 Improvement of particle number concentration simulations at 40–100 nm

428 The number concentration of particles in the 40–100 nm range is mainly affected  
429 by the coagulation and condensation processes. While the coagulation process tends to  
430 largely affect ultrafine particles below 10 nm than those with larger sizes (Wu et al.,  
431 2011)(Wu et al., 2011), the condensation growth of particles during gas-particle  
432 partitioning at sizes of 10–40 nm, to a large extent, governs the variations in number  
433 concentration of 40–100 nm particles. The condensation process is primarily controlled  
434 by gas-particle partitioning of chemical species, which may change the chemical  
435 composition of particles, such as organic compounds and inorganics including sulfate,  
436 nitrate and ammonium.

437 Among the species contributing to the condensation growth of particles at 10–40  
438 nm, the organic compounds with ~~with  $C^*$~~  of  $10^{-2}$   $\mu\text{g m}^{-3}$  ~~play~~ play the dominant role  
439 (Pierce et al., 2011)(Pierce et al., 2011). In the current model setting, ~~there is a total of~~  
440 ~~two volatility sets, including  $10^{-2}$   $\mu\text{g m}^{-3}$  and  $10^5$   $\mu\text{g m}^{-3}$ .~~ ~~The~~ the low volatile organic  
441 matter of  $10^{-2}$   $\mu\text{g m}^{-3}$  comes from two gas-phase sources, including the direct emission  
442 of primary organic aerosol (POA) and SOA formed from S/IVOC (SI-SOA), conducive  
443 to condensation on particles. While the condensation of gaseous SOA is in general  
444 reasonable, the gas phase emissions of POA may be problematic. For instance, previous  
445 studies suggested that POA is in gas phase close to the emissions source, ~~but~~. However,  
446 with rapid dilution and cooling in the atmosphere away from the source, most POA

---

447 condenses to particle-phase ([Roldin et al., 2011b](#); [Roldin et al., 2011a](#); [Shrivastava et](#)  
448 [al., 2008](#))([Roldin et al., 2011b](#); [Roldin et al., 2011a](#); [Shrivastava et al., 2008](#)). Therefore,  
449 away from the emissions source POA, being in the particle phase, will not be involved  
450 in the growth of newly formed particles. Therefore, ~~we assume that low volatility POA~~  
451 ~~is emitted in the~~ may not contribute to particle ~~phase rather than~~ growth away from the  
452 ~~gas-phase~~ emission sources, which caused different size distributions of POA compared  
453 to when it was emitted in the gas-phase (Fig. [S1a](#)[S3a](#) vs. Fig. [S1b](#)[S3b](#)). Emitting low  
454 volatility POA in the particle phase eliminates the unreasonable quasi-banana shape  
455 pattern exhibiting concomitant growth of newly formed particles with increasing mass  
456 concentration of POA.

457 The composition analysis (Fig. [S1e](#)[S3c](#)) in the 10–40 nm particles mass from the  
458 model results indicates that organic compounds mentioned above only account for 21%  
459 of total mass (sulfates, nitrates, ammonium salts and organics) in this size range and the  
460 dominant species is nitrate which accounts for 51% of total mass, exhibiting  
461 inconsistencies with the previous studies which in general indicates a much smaller  
462 contribution of nitrate. ~~For instance, Liu et al. (2014a)~~ For instance, Liu et al. (2014a)  
463 suggested that over North China Plain in summer 2009, organic matter accounted for  
464 77% of particles around 30 nm, while the sum of  $\text{SO}_4^{2-}$ ,  $\text{NO}_3^-$  and  $\text{NH}_4^+$  only accounted  
465 for 18%. Another study showed that nitrate accounted for 7–8% at urban sites and 17%  
466 at rural sites for particles mass in 7–30 nm in the United States in 2007 (~~Bzdek et al.,~~  
467 [2012](#))([Bzdek et al., 2012](#)). Therefore, the potentially too high modeled nitrate fraction  
468 in 10–40 nm in this study is tightly associated with the condensation process, with the  
469 specific reasons explained below.

470 The condensation of nitric acid on particles is highly constrained by the particle  
471 acidity. The acidity in smaller particles (i.e., 10–40 nm) tends to be higher than that in  
472 large particles, primarily due to the larger condensation of  $\text{H}_2\text{SO}_4$  (~~Lu et al., 2022~~)([Lu](#)  
473 [et al., 2022](#)), and particles with sizes greater than 40 nm have a much weaker acidity or  
474 are nearly neutral. For example, observed evidence has shown that acidic ultrafine  
475 particles account for a large proportion of ultrafine particles from 22 December 2010  
476 to 15 January 2011 in Hong Kong, e.g., 65% for particles within 5.5–30 nm (Wang et

---

477 al., 2014).

478 In the model, a particle is determined to be in solid phase when the ambient relative  
479 humidity is lower than the mutual deliquescence relative humidity of the particles  
480 ([Zaveri et al., 2005](#); [Zaveri et al., 2008](#))([Zaveri et al., 2005](#); [Zaveri et al., 2008](#)), which  
481 is in general suitable for particles dominated by inorganics. In the study area, the results  
482 indicate that at most conditions relative humidity are relatively low and the particles  
483 are in solid phase, in which the condensation process is not affected by particle acidity  
484 and the condensation of nitric acid on particles is directly calculated based on the gas-  
485 particle equilibrium concentration ([Zaveri et al., 2008](#))([Zaveri et al., 2008](#)). However,  
486 for particles below 40 nm, the main compositions are likely to be organic matter ([Zhu  
487 et al., 2014](#); [Ehn et al., 2014](#))([Zhu et al., 2014](#); [Ehn et al., 2014](#)), which tends to be in  
488 liquid phase ([Virtanen et al., 2011](#); [Cheng et al., 2015](#))([Virtanen et al., 2011](#); [Cheng et  
489 al., 2015](#)), under which the condensation of nitric acid is strongly constrained by acidity.  
490 Therefore, the phase misrepresentation ignores the weakening effect of acidity on nitric  
491 acid condensation, resulting in too high nitrate therein.

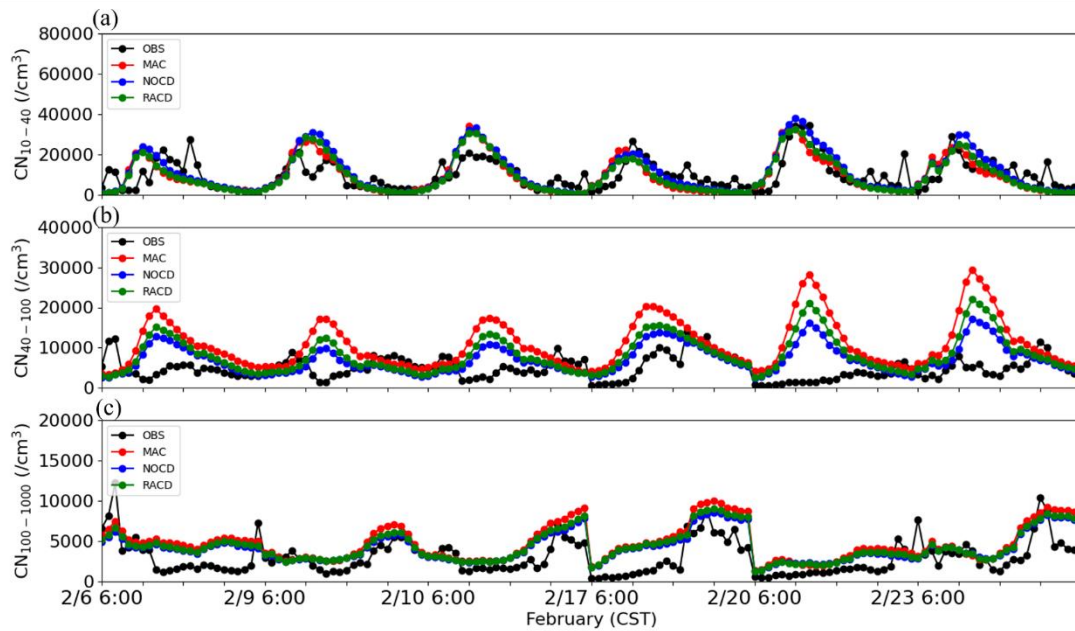
492 To overcome this issue, we propose a ratio method for condensation (RACD) to  
493 partition the condensation of nitric acid on particles under 40 nm, by applying a ratio  
494 of the number concentration of non-acidic particles to ultrafine particles. The method  
495 is based on two assumptions, including: 1) little condensation of nitric acid on particles  
496 with strong acidity ([Lu et al., 2022](#))([Lu et al., 2022](#)); 2) the condensation of nitric acid  
497 on particles is proportional to the ratio of the number concentration of non-acidic  
498 ultrafine particles to the total particles, despite the existence of uncertainties. -Fig. [S2S4](#)  
499 depicts the average particle number concentration and acid particle in the 1 to 40 nm  
500 range, calculated based on [Wang et al. \(2014\)](#)[Wang et al. \(2014\)](#). The ratio of non-acidic  
501 particles is 8% for particles below 10 nm, 18% for particles at 10–15.8nm8 nm, 30%  
502 for particles at 15.8–25.1nm, and 55% for particles at 25.1–39.8nm8 nm (Fig. [S2S4](#)).  
503 Note that the ratio is based on measurements acquired at a single site in Hong Kong,  
504 therefore more observational studies are needed to warrant the robustness of the method.  
505 Alternatively, the condensation of nitric acid on particles in bins from 1nm to 40 nm is  
506 completely suppressed, referred to as NOCD.

---

507 The simulation results based on the two methods (RACD and NOCD) are shown  
508 in Fig. 3. Compared to MAC, RACD simulations reduce previously noted  
509 overestimation of [PNCparticle number concentration](#) in the 40–100 nm size range (Fig.  
510 3b), with the [MFBmean fractional bias](#) decreases from 83% to 63%. In addition to the  
511 amount of nitrate condensation during particle growth mentioned above, the  
512 overestimation of particle number concentrations in the 40–100 nm range may be  
513 attributed to nucleation process. More specifically, in the H<sub>2</sub>SO<sub>4</sub>-H<sub>2</sub>O binary nucleation  
514 mechanism used in this study, when the concentration of sulfuric acid gas is reduced  
515 (Section 3.2.1), the resulting decrease in nucleation rate leads to a slight decrease in  
516 particle number concentration at 40–100 nm relative to Base ([MFBmean fractional bias](#)  
517 from 98% to 83%). Apart from that, it may also be related to the choice of nucleation  
518 parameterization scheme. For example, using a global chemical transport model GEOS-  
519 Chem, [Yu et al. \(2015\)](#) with a nucleation mechanism in which formation rate is a  
520 [function of the concentrations of sulfuric acid and low-volatility organics](#), [Yu et al.](#)  
521 [\(2015\)](#) overestimated the concentration of particles in the 10–100 nm range by 161%  
522 at nine sites in the summer in North America. A possible explanation for this  
523 overestimation was given by the uncertainty of the predicted concentration of organic  
524 compounds involved in organics-mediated nucleation parameterization. After they  
525 switched to another scheme of the ion-mediated nucleation mechanism without organic  
526 matter, the number becomes 27% lower than the observations ([Yu et al., 2015](#))([Yu et](#)  
527 [al., 2015](#)). The test based on different schemes is beyond the scope of the study, which  
528 is therefore not investigated.

529 Moreover, the overestimation of particles over 100 nm (CN<sub>100–1000</sub>; Fig. 3c), which  
530 have a strong influence on CCN, also decrease in the RACD simulation. Thus, the  
531 [MFBmean fractional bias](#) decreases from 25% (MAC) to 1%. Note that the slight  
532 increase of CN<sub>10–40</sub> through the application of RACD, can be linked to the decrease of  
533 nitrate condensation, and leads to weakened particle growth and enhanced particle  
534 number concentration at 10–40 nm (Fig. 3a). The alternative method by completely  
535 removing the nitrate condensation (NOCD) yields even better performance in particle  
536 number concentration of 40–100 nm ([MFBmean fractional bias](#) of 34%), indicating the

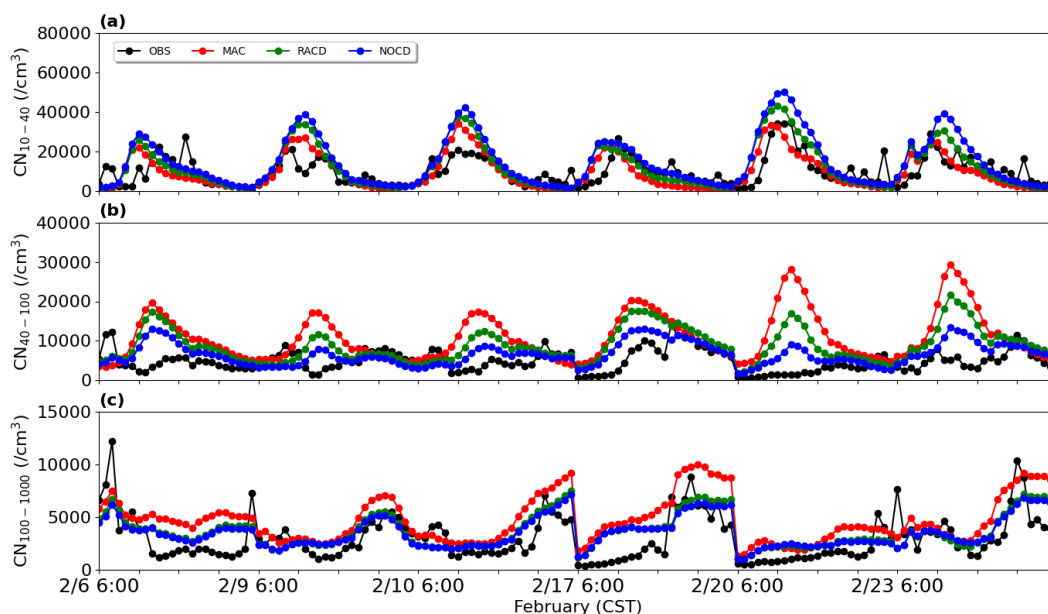
537 feasibility by reducing the nitrate condensation.- The proportion of nitrate simulated by  
538 RACD is 23%, closer to values reported in past observations (Bzdek et al., 2011; Bzdek  
539 et al., 2012)(Bzdek et al., 2011; Bzdek et al., 2012), while the nitrate (1%) in the  
540 scenario of NOCD seems to be too low. Considering the limited observational  
541 information obtained based on previous studies, RACD is applied in this study.



542

543 Figure 3In addition to Qingdao, we evaluate the model performance over a few  
544 other sites, including one site over urban Beijing and the other one over the rural area  
545 of Gucheng, yielding consistent improvements in model simulations (Section S2; Fig.  
546 S6-S8). Moreover, we select another empirical scheme, e.g., kinetics, and one classical  
547 nucleation scheme, indicating the empirical scheme of activation scheme is in general  
548 a good option in this study (Section S2; Fig. S9-S11; Table S1-3).





549

550 **Fig. 3.** The time series of (a)  $CN_{10-40}$ , (b)  $CN_{40-100}$  and (c)  $CN_{100-1000}$  on NPF days in  
 551 Qingdao on February 5-24 simulated from MAC (marked in red), NOCD (marked in  
 552 blue) and RACD (marked in green) as well as from observations (OBS) (marked in  
 553 black). All times are local time.

554

### 555 3.3 Substantial contributions of SI-SOA to CCN

556 Compared with the original model setting, after adjusting the growth process of  
 557 ultrafine particles (RACD), the number concentration of particles tend to decrease,  
 558 especially for particles above 40 nm. Ultrafine particles above 40 nm are important  
 559 sources of CCN (Dusek et al., 2006)(Dusek et al., 2006), in this way, the number  
 560 concentration of CCN also tends to decline. In addition, in the Base case, we found that  
 561 the model overestimated  $CCN_{0.4\%}$  and  $CCN_{0.6\%}$ , with **MFBmean fractional bias** being  
 562 64% and 87%, respectively. After adjusting the condensation growth process of  
 563 ultrafine particles, under high supersaturation (i.e.,  $CCN_{0.4\%}$  and  $CCN_{0.6\%}$ ), the  
 564 capability of the model in reproducing the CCN is improved. RACD reduces the  
 565 overestimation of  $CCN_{0.4\%}$  and  $CCN_{0.6\%}$ , with **MFBmean fractional bias** reduced to 30%  
 566 and 56%, respectively, although the overestimates still exist (Figs. S3bS5b, c). However,  
 567 for low supersaturation (i.e.,  $CCN_{0.2\%}$ ), the decrease of number concentration of CCN



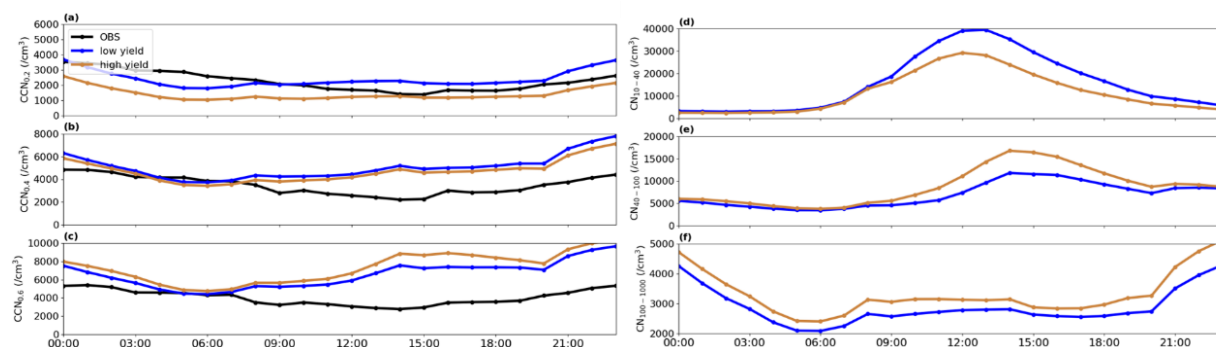
---

568 is too large, and [MFBmean fractional bias](#) decreases from 7% to -45% ([Figs. S3aFig.](#)  
569 [S5a](#)), therefore the bias will be further adjusted later.

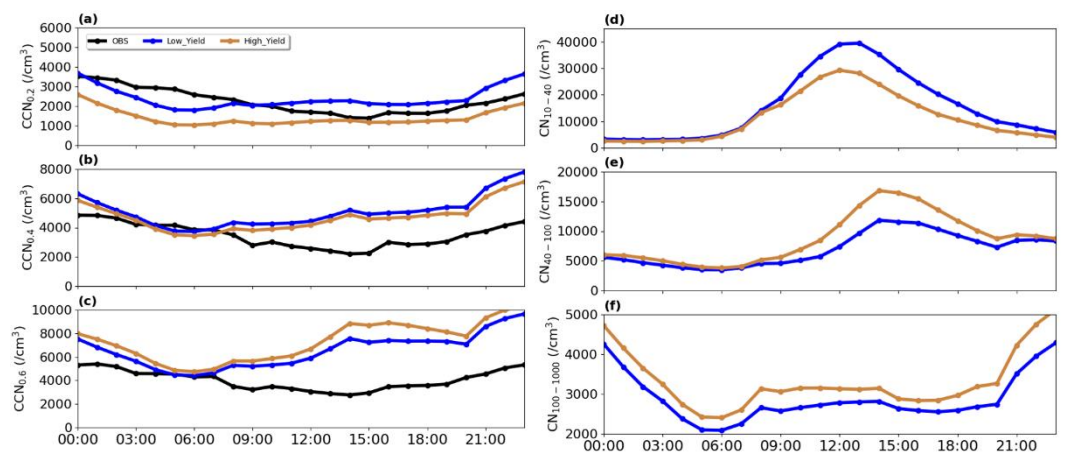
570 In addition to the growth process, the remaining overestimate of CCN under high  
571 SS and underestimate of CCN over low SS is likely to be influenced by the chemical  
572 compositions involved in the activation of ultrafine particles into CCN. Specifically,  
573 ultrafine particles can grow up to CCN size under certain SS (Pierce and Adams, 2007).  
574 This process is influenced by both particle size and hygroscopicity, and hygroscopicity  
575 is closely related to the chemical composition of particles (Petters and Kreidenweis,  
576 2007). In particular, inorganic compounds generally increase particle hygroscopicity,  
577 increasing CCN. SOA has dual effects on CCN since it decreases particle  
578 hygroscopicity but also promotes growth of particles, and these two effects are  
579 competitive with each other (Wu et al., 2015; Zaveri et al., 2021). Ultrafine particles  
580 must grow to a critical size to be activated into CCN ([Dusek et al., 2006](#))([Dusek et al.,](#)  
581 [2006](#)). SOA act as a major contributor in promoting the condensational growth of  
582 ultrafine particles to the critical size, facilitating particles activation into CCN. In  
583 contrast, SOA tends to reduce the hygroscopicity of particles, leading to a diminished  
584 ability of activation to CCN (Wu et al., 2015). These two competing effects work  
585 together and modulate the number of CCN. Moreover, considering that SI-SOA is the  
586 main SOA component on ultrafine particles (Fig. [S3dS5d](#)), the effect of SI-SOA on  
587 CCN is therefore explored in this study.

588 Considering SI-SOA is a product of S/IVOC oxidation, the oxidation rate of  
589 S/IVOC is tightly associated with CCN, which likely affects the bias of CCN. In the  
590 original model setup, the oxidation rate is set to be a constant of  $4 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1}$   
591  $\text{s}^{-1}$  for all S/IVOC. However, a recent study ([Wu et al., 2021b](#))([Wu et al., 2021b](#))  
592 proposed that the oxidation rate can be as high as  $5 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$  such as for  
593 polycyclic aromatic hydrocarbons (PAHs), close to the original model value, but can be  
594 as low as half (i.e.,  $2 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$ ) of the original modeling setting for S/IVOC  
595 species except PAHs (O-S/IVOCs). It is noteworthy that the oxidation rates of  $5 \times 10^{-$   
596  $11$  and  $2 \times 10^{-11}$  in general represent the upper and lower bounds ([Zhao et al., 2016; Wu](#)  
597 [et al., 2021b](#))([Zhao et al., 2016; Wu et al., 2021b](#)).

598 To delve into how oxidation rates affect CCN, we set up a few numerical  
 599 experiments (Table 2) to investigate the response of CCN to the oxidation rate of  
 600 S/IVOC at three supersaturations (0.6%, 0.4%, 0.2%), including cases of High\_Yield  
 601 and Low\_Yield. As it is shown in Fig. 4, decreasing the oxidation rate (Low\_Yield)  
 602 leads to a reduction of  $\sim 10\%$  of CCN at high supersaturation (i.e., CCN<sub>0.6%</sub>) as  
 603 compared to the High\_Yield simulation. This behaviour is a consequence of the  
 604 decrease of particle number concentrations associated with Low\_Yield, particular of  
 605 the particles close to the critical ~~diameter~~diameter (40–100 nm). In this case, the effect  
 606 of particle size dominates the hygroscopicity. In contrast, at a lower supersaturation  
 607 (CCN<sub>0.2%</sub>), CCN increases by 42% when the oxidation rate is switched from a high to  
 608 a low value, which is due to the smaller fraction of SI-SOA contributing to particulate  
 609 mass when the ~~oxidate~~oxidation rate is low. In this case, relative to SOA, a larger  
 610 fraction of other particle constituents such as inorganics, increase the volume weighted  
 611 particle hygroscopicity (Dusek et al., 2006)(Dusek et al., 2006) which causes the  
 612 increase of CCN number. This means that the effect of hygroscopicity on CCN  
 613 surpasses the influence on particle size at low supersaturations. This conclusion is  
 614 consistent with the observation conducted by Ma et al. (2016)Ma et al. (2016) in the  
 615 North China Plain in 2013, which suggested that along with the decrease of SS, the  
 616 particles that can be activated to CCN is more sensitive to changes of particle  
 617 hygroscopicity.



618  
619 Figure



620

621 **Fig. 4.** Average diurnal variation of (a)  $CCN_{0.2\%}$ , (b)  $CCN_{0.4\%}$  and (c)  $CCN_{0.6\%}$  and (d)  
 622  $CN_{10-40}$ , (e)  $CN_{40-100}$ , (f)  $CN_{100-1000}$  on NPF days in Qingdao on February 5-24, 2017,  
 623 in Low-yield and High-yield simulations, shown as blue and brown lines, and black  
 624 lines represent observation results.

625

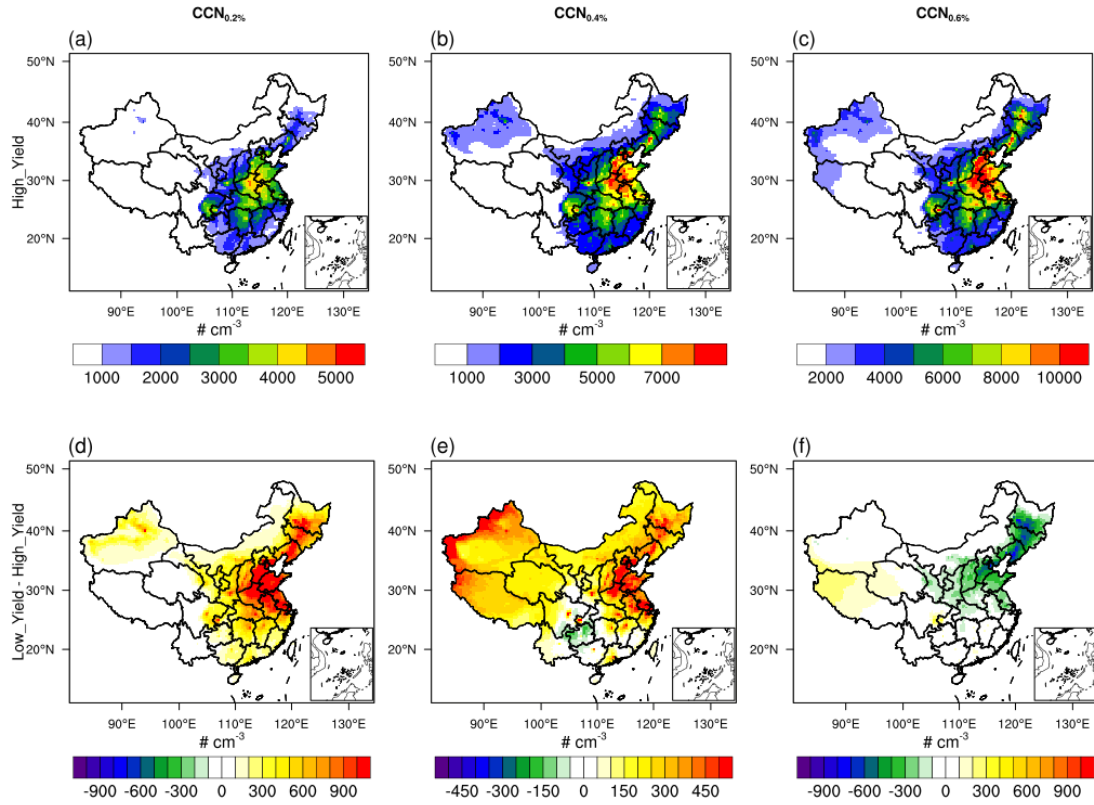
626 Furthermore, compared to the high yield of SI-SOA, the low SI-SOA yield results  
 627 in a high CCN concentration under low SS and low CCN concentration under high SS.  
 628 Therefore, both the underestimates of  $CCN_{0.2\%}$  (MFB<sub>mean fractional bias</sub> of -45%) and  
 629 overestimates of  $CCN_{0.6\%}$  (MFB<sub>mean fractional bias</sub> of 56%) mentioned above are  
 630 improved, with MFB<sub>mean fractional bias</sub> of  $CCN_{0.2\%}$  and  $CCN_{0.6\%}$  reaching 7% and  
 631 43%, respectively (FigsFig. 4a,c). This result suggests that the oxidation rate of S/IVOC  
 632 is possibly closer to the low value, which is understandable based on Wu et al.  
 633 (2021b)understandable based on Wu et al. (2021b), who found that the amount of O-  
 634 S/IVOCs, which corresponds to a low oxidation rate, is in general much larger (i.e., 20  
 635 times) than that of PAHs with a high oxidation rate.

636

In addition to the single site of Qingdao, we further explore the impact of SI-SOA  
 637 yield on CCN from a larger spatial coverage (Fig. 5). Consistent with the mechanism  
 638 revealed over Qingdao, even from a larger spatial perspective, a lower oxidation rate of  
 639 S/IVOC essentially enhances CCN at a lower SS (e.g.,  $CCN_{0.2\%}$ ; Fig. 5a) with the  
 640 highest increase over northNorth China Plain area (FigsFig. 5a), and weakens CCN (i.e.,  
 641 by 10–20% over Beijing-Tianjin-Hebei) at a higher SS (FigsFig. 5c), particularly over  
 642 the dense emission area (Fig. S4S12). It is worth noting that in the 2-species VBS

---

643 mechanism used in our study, all S/IVOC in the inventory is calculated based on a  
644 constant emission ratio of S/IVOC to POA from all source categories (~~Shrivastava et~~  
645 ~~al., 2011~~)([Shrivastava et al., 2011](#)), which may miss part of S/IVOC due to different  
646 emission ratios of POA from different source (~~Chang et al., 2022~~)([Chang et al., 2022](#)).  
647 In addition, the simplified VBS mechanism used in our study does not take into account  
648 the multi-step oxidation of organic species, which may introduce some uncertainties.  
649 To be more specific, in the 2-species VBS mechanism, SI-SOA with effective saturation  
650 concentrations ( $c^*$ ) of  $10^{-2} \mu\text{g m}^{-3}$  is formed by the vapor phase oxidation of S/IVOC  
651 vapors with  $c^*$  of  $10^5 \mu\text{g m}^{-3}$ , reducing volatility by 7 orders of magnitude. The process  
652 of one-step oxidation does not mean to represent a physical process, but to parameterize  
653 the mean effect of a complex process of SOA formation (Shrivastava et al., 2011).  
654 However, in the real atmosphere, the gaseous VOCs often undergo multi-generational  
655 oxidation to form SOA (Garmash et al., 2020), during which the properties and  
656 composition of SOA change substantially.



Figure

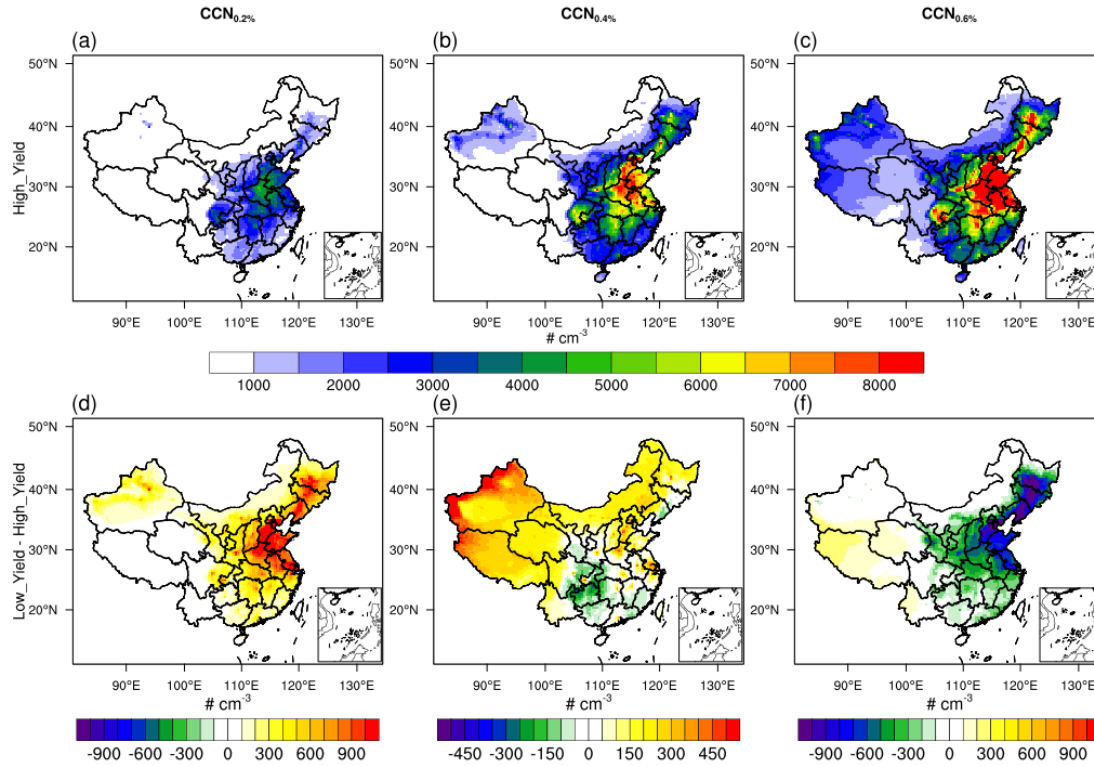


Fig. 5. Spatial distributions of CCN concentrations at different supersaturations (SS), (a) and (d) are  $CCN_{0.2\%}$ , (b) and (e) are  $CCN_{0.4\%}$ , and (c) and (f) are  $CCN_{0.6\%}$ . The top

---

662 panels exhibit the results from the High\_Yield simulation, and the bottom panels shows  
663 the difference between the Low\_Yield and High\_Yield simulations.

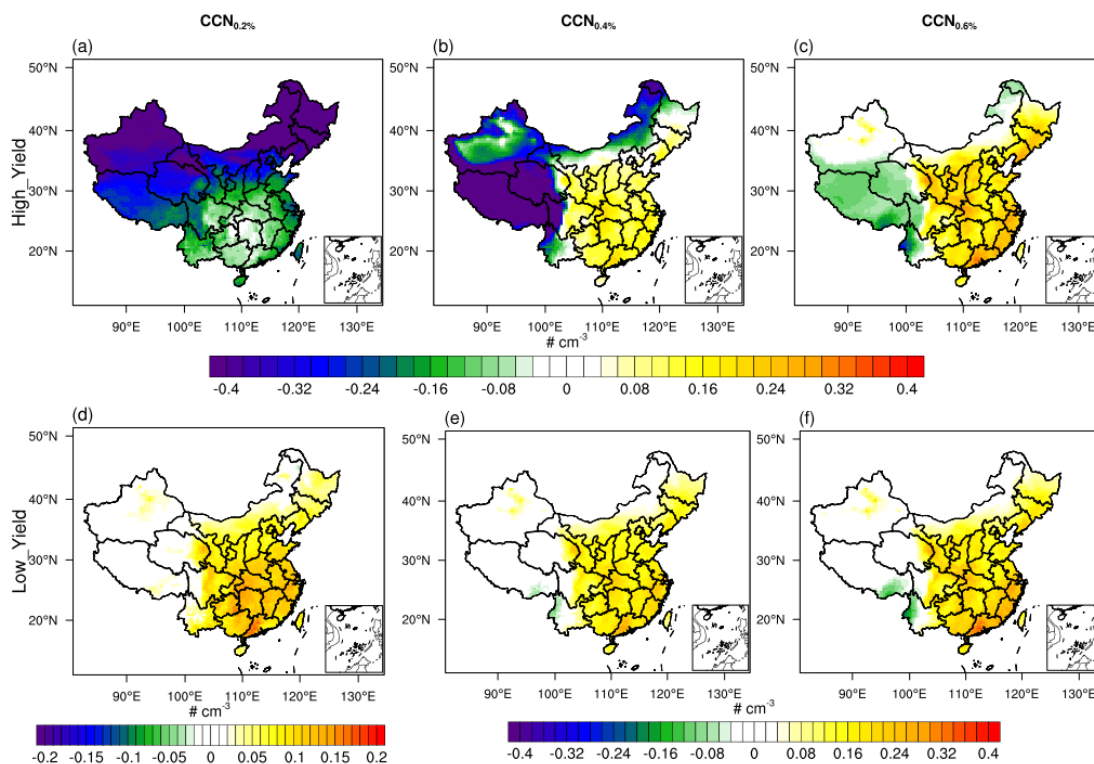
664

### 665 **3.4 Contribution of nucleation to CCN under different SI-SOA yields**

666 Considering the importance of nucleated particles on CCN ([Yu et al., 2020;](#)  
667 [Westervelt et al., 2013](#))([Yu et al., 2020;](#) [Westervelt et al., 2013](#)), we further investigate  
668 the influence of nucleation on CCN under different SI-SOA yield conditions discussed  
669 above.

670 As shown in Fig. 6, in simulations close to the original model setting (High\_Yield),  
671 when SS is low (i.e., SS=0.2%), the nucleation process tends to reduce the CCN by  
672 ~10–50%. In contrast, when the SS is high (0.6%), the nucleation results in a significant  
673 increase in CCN in most regions of China. When the yield of SI-SOA is adjusted to a  
674 lower level, the nucleation process has a positive contribution to CCN under both low  
675 and high SS. Especially, when SS is low (0.2%), the sign reversal, i.e., from negative  
676 (Fig. 6a) to positive (Fig. 6d) ~~contribution~~[contributions](#) of NPF to CCN along with the  
677 decrease of SI-SOA yield, i.e., -the increase is concentrated in the eastern China with  
678 an average of 10–20%.- The primary mechanism lies in that along with the decrease of  
679 SI-SOA yield, the smaller fraction of SI-SOA yields an increase in hygroscopicity,  
680 which surpasses the suppression effect on particle growth due to reduced SI-SOA  
681 formation. In the real atmosphere, when the supersaturation is usually low, e.g. about  
682 ~0.1% in polluted areas (~~[Kalkavouras et al., 2019;](#) [Hudson and Noble,](#)~~  
683 [2014](#))([Kalkavouras et al., 2019;](#) [Hudson and Noble, 2014](#)), CCN will likely reduce with  
684 increasing oxidation rate of S/IVOC and corresponding SI-SOA formation.

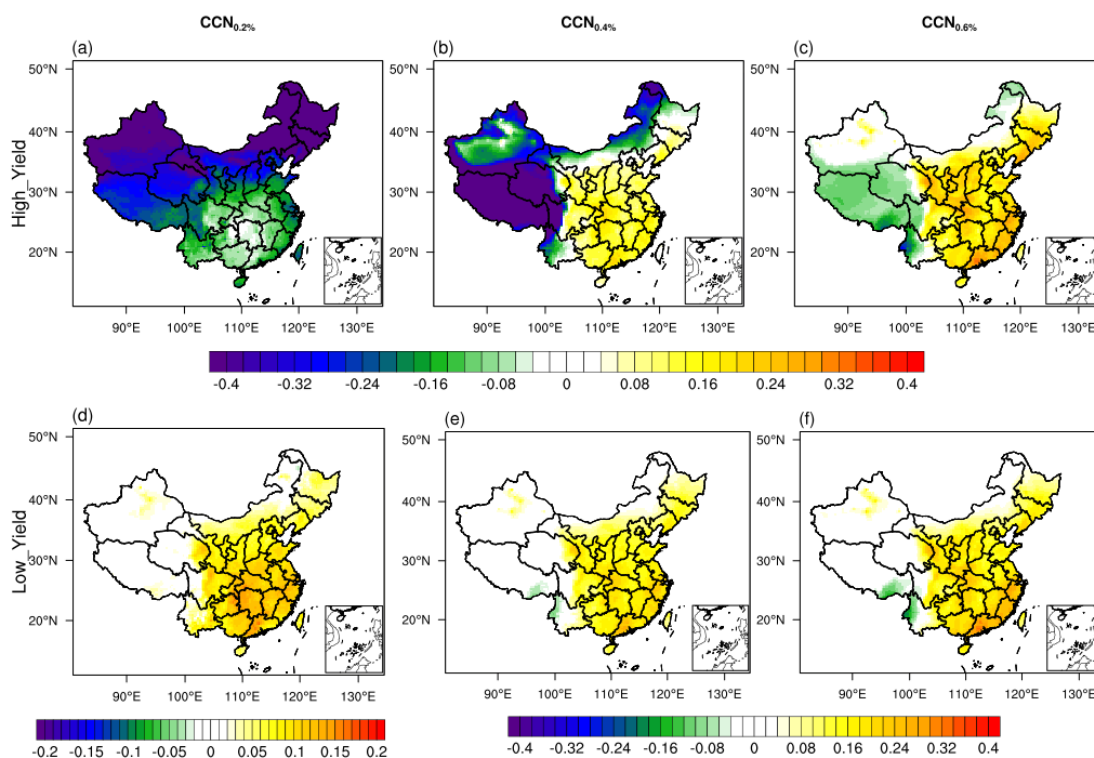




685

686

Figure



687

688 **Fig. 6.** Spatial distribution of contribution of nucleation to CCN calculated by the ratio  
 689 of the difference between the parameterization with and without nucleation to the  
 690 parameterization with nucleation under different SI-SOA yields in China in February

---

691 2017. (a), (d) is CCN<sub>0.2%</sub>,(b), (e) is CCN<sub>0.4%</sub>,(c), (f) is CCN<sub>0.6%</sub>.The upper panel and  
692 lower panel represent High\_Yield and Low\_Yield simulation respectively.

693  
694  
695 In addition to the linear-H<sub>2</sub>SO<sub>4</sub> nucleation mechanism, one more empirical scheme  
696 of kinetics nucleation is selected, which assumes that the nucleation rate is proportional  
697 to the square of the concentration of sulfuric acid ( $J = K[H_2SO_4]^2$ ), to investigate the  
698 effect of nucleation on CCN. Substantially positive contributions of nucleation to CCN  
699 is found when the low SI-SOA yield is applied, consistent with what was shown based  
700 on the linear-H<sub>2</sub>SO<sub>4</sub> nucleation scheme (Fig. S13). However, nucleation contributes  
701 positively to CCN even when the SI-SOA yield is high in the quadratic-H<sub>2</sub>SO<sub>4</sub>  
702 nucleation scheme (e.g., kinetics nucleation scheme). When more sulfuric acid  
703 molecules participate in nucleation under this scheme than the linear-H<sub>2</sub>SO<sub>4</sub> nucleation  
704 scheme, the particles are more easily hygroscopically activated to CCN, which is  
705 equivalent to the effect of a reduction in organic components in the linear-H<sub>2</sub>SO<sub>4</sub>  
706 nucleation scheme (e.g., activation-type nucleation scheme). The results from this study  
707 show the importance of assessing the simulated effects of the nucleation scheme on not  
708 only the formation and growth process of particles but also climate factors such as CCN  
709 using observations.

## 711 **Conclusions**

712 In this study, WRF-Chem explicit-NPF ~~simulations~~simulations are used to  
713 investigate the observed wintertime NPF events and their contribution to CCN in China.  
714 Based on observations in a typical coastal city of Qingdao, we identify high biases of  
715 the model simulated CN and CCN concentrations. Therefore, we updated and improved  
716 the parameterization setting on particle growth in the model, mainly including: (1)  
717 adjusting the mass accommodation coefficient ( $\alpha$ ) to from the default value of 0.1 to  
718 0.65, an important parameter for sulfuric acid condensation; (2) proportionally reducing  
719 the condensation amount of nitric acid on particles below 40 nm, (3) changing the  
720 emitted low-volatility POA from gas to particle. Through these adjustments, the



---

721 capability of the model in reproducing CN and CCN is substantially improved, leading  
722 to better agreement with the observed results, which significantly reduces the  
723 overestimation of CN<sub>10-40</sub> ([MFBmean fractional bias](#) decreases from 48% to 1%) and  
724 CN<sub>40-100</sub> ([MFBmean fractional bias](#) decreases from 98% to 63%).  
725

726 For CCN, due to the crucial role of SI-SOA in promoting the growth of ultrafine  
727 particles, on the basis of previous studies, we lower the oxidation rate of S/IVOC and  
728 hence the production rate of SI-SOA, which weakens the growth of particles to reach  
729 the critical size of CCN activation, but enhances particulate hygroscopicity favoring the  
730 activation to CCN. When the yield of SI-SOA is adjusted to the lower bound of  
731 literature value, CCN<sub>0.6%</sub> is reduced by ~10% and is closer to observations. At low SS  
732 (CCN<sub>0.2%</sub>), the decrease of SI-SOA yield has greater effects on the increase of particle  
733 hygroscopicity compared to the effect of the reduction of particle size due to the  
734 decrease of condensation growth. It results in an increase of CCN (as large as ~42%) in  
735 better agreement with observations. Under low SS conditions, common in the  
736 atmosphere, a 2.5-fold reduction in SI-SOA yield results in a substantial increase of  
737 CCN that switches from a negative ~~contribution~~[contribution](#) of new particle formation  
738 to CCN from -50%~-10% to a positive contribution of 10~20%. [The substantial  
739 contribution of new particle formation to CCN under low SS and SI-SOA is applicable  
740 to other mechanisms such as kinetics.](#)  
741

742 **Competing interests.** At least one of the (co-)authors is a member of the editorial board  
743 of Atmospheric Chemistry and Physics.  
744

745 **Acknowledgements.** This research was supported by grants from the National Natural  
746 Science Foundation of China (42122039) and Fundamental Research Funds for the  
747 Central Universities (202072001). Y.W. was supported by the National Science  
748 Foundation Atmospheric Chemistry Program. M.S. was supported by the U.S.  
749 Department of Energy (DOE) Office of Science, Office of Biological and

---

750 Environmental Research (BER) through the Early Career Research Program and the  
751 Atmospheric System Research (ASR) program.

752

753

754 **References:**

755 Arghavani S, Rose C, Banson S, et al. 2022. The Effect of Using a New Parameterization of Nucleation  
756 in the WRF-Chem Model on New Particle Formation in a Passive Volcanic Plume. *Atmosphere*  
757 [J], 13(1): 15.

758 [Bergman T, Laaksonen A, Korhonen H, et al. 2015. Geographical and diurnal features of amine-enhanced  
759 boundary layer nucleation. \*Journal of Geophysical Research: Atmospheres\* \[J\], 120\(18\): 9606-  
760 9624.](#)

761 Buchholz R R, Emmons L K, Tilmes S 2019. The CESM2 Development Team., 2019. CESM2.1/CAM-  
762 Chem Instantaneous Output for Boundary Conditions. UCAR/NCAR - Atmospheric Chemistry  
763 Observations and Modeling Laboratory.

764 Bzdek B, Zordan C, Luther G, et al. 2011. Nanoparticle Chemical Composition During New Particle  
765 Formation. *Aerosol Science and Technology* [J], 45(1041-1048).

766 Bzdek B R, Zordan C A, Pennington M R, et al. 2012. Quantitative Assessment of the Sulfuric Acid  
767 Contribution to New Particle Growth. *Environmental Science & Technology* [J], 46(8): 4365-  
768 4373.

769 Cai C, Zhang X, Wang K, et al. 2016. Incorporation of new particle formation and early growth  
770 treatments into WRF/Chem: Model improvement, evaluation, and impacts of anthropogenic  
771 aerosols over East Asia. *Atmospheric Environment* [J], 124(262-284).

772 Cai J, Chu B, Yao L, et al. 2020. Size-segregated particle number and mass concentrations from different  
773 emission sources in urban Beijing. *Atmos. Chem. Phys.* [J], 20(21): 12721-12740.

774 Carter W 2000. Documentation of the SAPRC-99 Chemical Mechanism for VOC Reactivity Assessment.  
775 Final Report to California Air Resources Board [J].

776 Chang X, Zhao B, Zheng H, et al. 2022. Full-volatility emission framework corrects missing and  
777 underestimated secondary organic aerosol sources. *One Earth* [J], 5(403-412).

778 [Chang Y, Gao Y, Lu Y, et al. 2021. Discovery of a Potent Source of Gaseous Amines in Urban China.  
779 \*Environmental Science & Technology Letters\* \[J\], 8\(9\): 725-731.](#)

780 Chen F, Dudhia J 2000. Coupling an Advanced Land-Surface/Hydrology Model with the Penn  
781 State/NCAR MM5 Modeling System. 129(

782 Cheng Y, Su H, Koop T, et al. 2015. Size dependence of phase transitions in aerosol nanoparticles. *Nature*  
783 *Communications* [J], 6(1): 5923.

784 Chrit M, Sartelet K, Sciare J, et al. 2018. Modeling organic aerosol concentrations and properties during  
785 winter 2014 in the northwestern Mediterranean region. *Atmos. Chem. Phys.* [J], 18(24): 18079-  
786 18100.

787 Chu B, Kerminen V-M, Bianchi F, et al. 2019. Atmospheric new particle formation in China.  
788 *Atmospheric Chemistry and Physics* [J], 19(115-138).

789 Dal Maso M, Kulmala M, Riipinen I, et al. 2005. Formation and growth of fresh atmospheric aerosols:  
790 Eight years of aerosol size distribution data from SMEAR II, Hyytiälä, Finland. *Boreal*  
791 *Environment Research* [J], 10(323-336).

---

792 Davidovits P, Worsnop D R, Jayne J T, et al. 2004. Mass accommodation coefficient of water vapor on  
793 liquid water. *Geophysical Research Letters* [J], 31(22).

794 Donahue N M, Robinson A L, Stanier C O, et al. 2006. Coupled Partitioning, Dilution, and Chemical  
795 Aging of Semivolatile Organics. *Environmental Science & Technology* [J], 40(8): 2635-2643.

796 Dong C, Matsui H, Spak S, et al. 2019. Impacts of New Particle Formation on Short-term Meteorology  
797 and Air Quality as Determined by the NPF-explicit WRF-Chem in the Midwestern United States.  
798 *Aerosol and Air Quality Research* [J], 19(2): 204-220.

799 Dusek U, Frank G P, Hildebrandt L, et al. 2006. Size Matters More Than Chemistry for Cloud-Nucleating  
800 Ability of Aerosol Particles. *Science* [J], 312(5778): 1375-1378.

801 Ehn M, Thornton J, Kleist E, et al. 2014. A large source of low-volatility secondary organic aerosol.  
802 *Nature* [J], 506(476-479).

803 Fanourgakis G S, Kanakidou M, Nenes A, et al. 2019. Evaluation of global simulations of aerosol particle  
804 and cloud condensation nuclei number, with implications for cloud droplet formation. *Atmos.*  
805 *Chem. Phys.* [J], 19(13): 8591-8617.

806 Gordon H, Kirkby J, Baltensperger U, et al. 2017. Causes and importance of new particle formation in  
807 the present-day and preindustrial atmospheres. *122(16): 8739-8760.*

808 Grell G A 1993. Prognostic Evaluation of Assumptions Used by Cumulus Parameterizations. *Monthly*  
809 *Weather Review* [J], 121(3): 764-787.

810 Guo S, Hu M, Zamora M L, et al. 2014. Elucidating severe urban haze formation in China. *Proceedings*  
811 *of the National Academy of Sciences* [J], 111(49): 17373-17378.

812 Hong S-Y, Noh Y, Dudhia J 2006. A New Vertical Diffusion Package with an Explicit Treatment of  
813 Entrainment Processes. *Monthly Weather Review - MON WEATHER REV* [J], 134(  
814 Hudson J G, Noble S 2014. CCN and Vertical Velocity Influences on Droplet Concentrations and  
815 Supersaturations in Clean and Polluted Stratus Clouds. *Journal of the Atmospheric Sciences* [J],  
816 71(1): 312-331.

817 Iacono M, Delamere J, Mlawer E, et al. 2008. Radiative Forcing by Long-Lived Greenhouse Gases:  
818 Calculations with the AER Radiative Transfer Models. *Journal of Geophysical Research* [J],  
819 113(  
820 Jimenez J L, Canagaratna M R, Donahue N M, et al. 2009. Evolution of Organic Aerosols in the  
821 Atmosphere. *Science* [J], 326(5959): 1525-1529.

822 Kalkavouras P, Bougiatioti A, Kalivitis N, et al. 2019. Regional new particle formation as modulators of  
823 cloud condensation nuclei and cloud droplet number in the eastern Mediterranean. *Atmos. Chem.*  
824 *Phys.* [J], 19(9): 6185-6203.

825 Kerminen V-M, Chen X, Vakkari V, et al. 2018. Atmospheric new particle formation and growth: Review  
826 of field observations. *Environmental Research Letters* [J], 13(  
827 Krechmer J E, Day D A, Ziemann P J, et al. 2017. Direct Measurements of Gas/Particle Partitioning and  
828 Mass Accommodation Coefficients in Environmental Chambers. *Environ Sci Technol* [J],  
829 51(20): 11867-11875.

830 Kulmala M, Dada L, Daellenbach K R, et al. 2021. Is reducing new particle formation a plausible solution  
831 to mitigate particulate air pollution in Beijing and other Chinese megacities? *Faraday*  
832 *Discussions* [J], 226(0): 334-347.

833 Kulmala M, L L, Lehtinen K, et al. 2004. Initial steps of aerosol growth. *Atmospheric Chemistry and*  
834 *Physics* [J], 4(

---

835 Kulmala M, Petäjä T, Ehn M, et al. 2013. Chemistry of Atmospheric Nucleation: On the Recent Advances  
836 on Precursor Characterization and Atmospheric Cluster Composition in Connection with  
837 Atmospheric New Particle Formation. *Annual review of physical chemistry* [J], 65(  
838 Kulmala M, Petäjä T, Nieminen T, et al. 2012. Measurement of the nucleation of atmospheric aerosol  
839 particles. *Nature Protocols* [J], 7(9): 1651-1667.

840 Lai S, Hai S, Gao Y, et al. 2022. The striking effect of vertical mixing in the planetary boundary layer on  
841 new particle formation in the Yangtze River Delta. *Science of The Total Environment* [J],  
842 829(154607).

843 Lee S-H, Gordon H, Yu H, et al. 2019. New Particle Formation in the Atmosphere: From Molecular  
844 Clusters to Global Climate. *Journal of Geophysical Research: Atmospheres* [J], 124(  
845 Li K, Zhu Y, Gao H, et al. 2015. A comparative study of cloud condensation nuclei measured between  
846 non-heating and heating periods at a suburb site of Qingdao in the North China. *Atmospheric*  
847 *Environment* [J], 112(40-53).

848 Li M, Liu H, Geng G, et al. 2017. Anthropogenic emission inventories in China: a review. *National*  
849 *Science Review* [J], 4(6): 834-866.

850 Li X, Li Y, Cai R, et al. 2022. Insufficient Condensable Organic Vapors Lead to Slow Growth of New  
851 Particles in an Urban Environment. *Environmental Science & Technology* [J], 56(14): 9936-  
852 9946.

853 Liu H J, Zhao C S, Nekat B, et al. 2014a. Aerosol hygroscopicity derived from size-segregated chemical  
854 composition and its parameterization in the North China Plain. *Atmos. Chem. Phys.* [J], 14(5):  
855 2525-2539.

856 Liu M, Matsui H 2022. Secondary Organic Aerosol Formation Regulates Cloud Condensation Nuclei in  
857 the Global Remote Troposphere. *Geophysical Research Letters* [J], 49(18): e2022GL100543.

858 Liu X, Zhu Y, Zheng M, et al. 2014b. Production and growth of new particles during two cruise  
859 campaigns in the marginal seas of China. *Atmospheric Chemistry and Physics* [J], 14(  
860 Liu X H, Zhu Y J, Zheng M, et al. 2014c. Production and growth of new particles during two cruise  
861 campaigns in the marginal seas of China. *Atmos. Chem. Phys.* [J], 14(15): 7941-7951.

862 Lovejoy E R, Curtius J, Froyd K D 2004. Atmospheric ion-induced nucleation of sulfuric acid and water.  
863 *Journal of Geophysical Research: Atmospheres* [J], 109(D8).

864 Lu H, Wang G, Guo H 2022. Ambient acidic ultrafine particles in different land-use areas in two  
865 representative Chinese cities. *Science of The Total Environment* [J], 830(154774).

866 Lu Y, Yan C, Fu Y, et al. 2019. A proxy for atmospheric daytime gaseous sulfuric acid concentration in  
867 urban Beijing. *Atmos. Chem. Phys.* [J], 19(3): 1971-1983.

868 Lupascu A, Easter R, Zaveri R, et al. 2015. Modeling particle nucleation and growth over northern  
869 California during the 2010 CARES campaign. *Atmos. Chem. Phys.* [J], 15(21): 12283-12313.

870 Ma N, Zhao C, Tao J, et al. 2016. Variation of CCN activity during new particle formation events in the  
871 North China Plain. *Atmos. Chem. Phys.* [J], 16(13): 8593-8607.

872 Matsui H, Koike M, Kondo Y, et al. 2011. Impact of new particle formation on the concentrations of  
873 aerosols and cloud condensation nuclei around Beijing. *Journal of Geophysical Research:*  
874 *Atmospheres* [J], 116(D19).

875 Matsui H, Koike M, Takegawa N, et al. 2013. Spatial and temporal variations of new particle formation  
876 in East Asia using an NPF-explicit WRF-chem model: North-south contrast in new particle  
877 formation frequency. 118(20): 11,647-611,663.

878 Merikanto J, Spracklen D V, Mann G W, et al. 2009. Impact of nucleation on global CCN. *Atmos. Chem.*  
879 *Phys. [J]*, 9(21): 8601-8616.

880 [Mikkonen S, Romakkaniemi S, Smith J N, et al. 2011. A statistical proxy for sulphuric acid concentration.](#)  
881 [Atmos. Chem. Phys. \[J\]](#), 11(21): 11319-11334.

882 Morrison H, Thompson G, Tatarskii V 2009. Impact of Cloud Microphysics on the Development of  
883 Trailing Stratiform Precipitation in a Simulated Squall Line: Comparison of One and Two-  
884 Moment Schemes. *Monthly Weather Review - MON WEATHER REV [J]*, 137(991-1007).

885 [Napari I, Noppel M, Vehkamäki H, et al. 2002. Parametrization of ternary nucleation rates for H2SO4-](#)  
886 [NH3-H2O vapors. Journal of Geophysical Research: Atmospheres \[J\]](#), 107(D19): AAC 6-1-  
887 [AAC 6-6.](#)

888 Nieminen T, Kerminen V M, Petäjä T, et al. 2018. Global analysis of continental boundary layer new  
889 particle formation based on long-term measurements. *Atmos. Chem. Phys. [J]*, 18(19): 14737-  
890 14756.

891 Petäjä T, Mauldin I R L, Kosciuch E, et al. 2009. Sulfuric acid and OH concentrations in a boreal forest  
892 site. *Atmos. Chem. Phys. [J]*, 9(19): 7435-7448.

893 Pierce J, Riipinen I, Kulmala M, et al. 2011. Quantification of the volatility of secondary organic  
894 compounds in ultrafine particles during nucleation events. *Atmospheric Chemistry and Physics*  
895 *Discussions [J]*, 11(14495-14539).

896 Pöschl U, Canagaratna M, Jayne J T, et al. 1998. Mass Accommodation Coefficient of H2SO4 Vapor on  
897 Aqueous Sulfuric Acid Surfaces and Gaseous Diffusion Coefficient of H2SO4 in N2/H2O. *The*  
898 *Journal of Physical Chemistry A [J]*, 102(49): 10082-10089.

899 Qiao X, Yan C, Li X, et al. 2021. Contribution of Atmospheric Oxygenated Organic Compounds to  
900 Particle Growth in an Urban Environment. *Environmental Science & Technology [J]*, XXXX(  
901 Ren J, Chen L, Fan T, et al. 2021. The NPF Effect on CCN Number Concentrations: A Review and Re-  
902 Evaluation of Observations From 35 Sites Worldwide. *Geophysical Research Letters [J]*, 48(19):  
903 e2021GL095190.

904 Riipinen I, Sihto S L, Kulmala M, et al. 2007. Connections between atmospheric sulphuric acid and new  
905 particle formation during QUEST III&ndash;IV campaigns in Heidelberg and Hyttiälä. *Atmos.*  
906 *Chem. Phys. [J]*, 7(8): 1899-1914.

907 Roldin P, Swietlicki E, Massling A, et al. 2011a. Aerosol ageing in an urban plume – implication for  
908 climate. *Atmos. Chem. Phys. [J]*, 11(12): 5897-5915.

909 Roldin P, Swietlicki E, Schurgers G, et al. 2011b. Development and evaluation of the aerosol dynamics  
910 and gas phase chemistry model ADCHEM. *Atmos. Chem. Phys. [J]*, 11(12): 5867-5896.

911 Saha S, Moorthi S, Wu X, et al. 2014. The NCEP Climate Forecast System Version 2. *Journal of Climate*  
912 *[J]*, 27(6): 2185-2208.

913 [Shen X, Sun J, Zhang X, et al. 2018. Comparison of Submicron Particles at a Rural and an Urban Site in](#)  
914 [the North China Plain during the December 2016 Heavy Pollution Episodes. Journal of](#)  
915 [Meteorological Research \[J\]](#), 32(26-37).

916 Shrivastava M, Fast J, Easter R, et al. 2011. Modeling organic aerosols in a megacity: comparison of  
917 simple and complex representations of the volatility basis set approach. *Atmospheric Chemistry*  
918 *and Physics [J]*, 11(6639-6662).

919 Shrivastava M K, Lane T E, Donahue N M, et al. 2008. Effects of gas particle partitioning and aging of  
920 primary emissions on urban and regional organic aerosol concentrations. 113(D18).

---

921 Sihto S L, Kulmala M, Kerminen V M, et al. 2006. Atmospheric sulphuric acid and aerosol formation:  
922 implications from atmospheric measurements for nucleation and early growth mechanisms.  
923 Atmos. Chem. Phys. [J], 6(12): 4079-4091.

924 Sihto S L, Mikkilä J, Vanhanen J, et al. 2011. Seasonal variation of CCN concentrations and aerosol  
925 activation properties in boreal forest. Atmos. Chem. Phys. [J], 11(24): 13269-13285.

926 Tewari M, Wang W, Dudhia J, et al. 2016. Implementation and verification of the united NOAA land  
927 surface model in the WRF model [M].

928 Virtanen A, Kannosto J, Kuuluvainen H, et al. 2011. Bounce behavior of freshly nucleated biogenic  
929 secondary organic aerosol particles. Atmos. Chem. Phys. [J], 11(16): 8759-8766.

930 Wang D-W, Guo H, Chan C K 2014. Diffusion Sampler for Measurement of Acidic Ultrafine Particles  
931 in the Atmosphere. Aerosol Science and Technology [J], 48(12): 1236-1246.

932 Wang J, Li M, Li L, et al. 2022. Particle number size distribution and new particle formation in Xiamen,  
933 the coastal city of Southeast China in wintertime. Science of The Total Environment [J],  
934 826(154208).

935 Westervelt D M, Pierce J R, Riipinen I, et al. 2013. Formation and growth of nucleated particles into  
936 cloud condensation nuclei: model-measurement comparison. Atmos. Chem. Phys. [J], 13(15):  
937 7645-7663.

938 Wu H, Li Z, Jiang M, et al. 2021a. Contributions of traffic emissions and new particle formation to the  
939 ultrafine particle size distribution in the megacity of Beijing. Atmospheric Environment [J],  
940 262(118652).

941 Wu L, Ling Z, Shao M, et al. 2021b. Roles of Semivolatile/Intermediate-Volatility Organic Compounds  
942 on SOA Formation Over China During a Pollution Episode: Sensitivity Analysis and  
943 Implications for Future Studies. Journal of Geophysical Research: Atmospheres [J], 126(8):  
944 e2020JD033999.

945 Wu Z, Hu M, Yue D, et al. 2011. Evolution of particle number size distribution in an urban atmosphere  
946 during episodes of heavy pollution and new particle formation. Science China Earth Sciences  
947 [J], 54(11): 1772.

948 Yao L, Garmash O, Bianchi F, et al. 2018. Atmospheric new particle formation from sulfuric acid and  
949 amines in a Chinese megacity. Science [J], 361(278-281).

950 Yu F 2005. Quasi-unary homogeneous nucleation of H<sub>2</sub>SO<sub>4</sub>-H<sub>2</sub>O. The Journal of Chemical Physics [J],  
951 122(7): 074501.

952 Yu F, Luo G, Nair A A, et al. 2020. Wintertime new particle formation and its contribution to cloud  
953 condensation nuclei in the Northeastern United States. Atmos. Chem. Phys. [J], 20(4): 2591-  
954 2601.

955 Yu F, Luo G, Pryor S C, et al. 2015. Spring and summer contrast in new particle formation over nine  
956 forest areas in North America. Atmos. Chem. Phys. [J], 15(24): 13993-14003.

957 Yuan Q, Li W, Zhou S, et al. 2015. Integrated evaluation of aerosols during haze-fog episodes at one  
958 regional background site in North China Plain. Atmospheric Research [J], 156(102-110).

959 Yue D L, Hu M, Zhang R Y, et al. 2011. Potential contribution of new particle formation to cloud  
960 condensation nuclei in Beijing. Atmospheric Environment [J], 45(33): 6070-6077.

961 Zaveri R A, Easter R C, Fast J D, et al. 2008. Model for Simulating Aerosol Interactions and Chemistry  
962 (MOSAIC). 113(D13).

963 Zaveri R A, Easter R C, Peters L K 2005. A computationally efficient Multicomponent Equilibrium  
964 Solver for Aerosols (MESA). Journal of Geophysical Research: Atmospheres [J], 110(D24).

---

965 Zhang Q, Jimenez J L, Canagaratna M R, et al. 2007. Ubiquity and dominance of oxygenated species in  
966 organic aerosols in anthropogenically-influenced Northern Hemisphere midlatitudes.  
967 *Geophysical Research Letters* [J], 34(13).

968 Zhao B, Wang S, Donahue N M, et al. 2016. Quantifying the effect of organic aerosol aging and  
969 intermediate-volatility emissions on regional-scale aerosol pollution in China. *Scientific*  
970 *Reports* [J], 6(1): 28815.

971 Zheng B, Tong D, Li M, et al. 2018. Trends in China's anthropogenic emissions since 2010 as the  
972 consequence of clean air actions. *Atmos. Chem. Phys.* [J], 18(19): 14095-14111.

973 Zhu Y, Li K, Shen Y, et al. 2019. New particle formation in the marine atmosphere during seven cruise  
974 campaigns. *Atmos. Chem. Phys.* [J], 19(1): 89-113.

975 Zhu Y, Sabaliauskas K, Liu X, et al. 2014. Comparative analysis of new particle formation events in less  
976 and severely polluted urban atmosphere. *Atmospheric Environment* [J], 98(655-664).

977 Zhu Y, Shen Y, Li K, et al. 2021. Investigation of Particle Number Concentrations and New Particle  
978 Formation With Largely Reduced Air Pollutant Emissions at a Coastal Semi-Urban Site in  
979 Northern China. *Journal of Geophysical Research: Atmospheres* [J], 126(17): e2021JD035419.

980

981