

1 **Technical note: Intercomparison Study of the EC Radiocarbon Analysis**

2 **Methods Using Synthetic Known Samples**

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46 Abstract

47
48 The accurate identification of elemental carbon (EC) source in aerosol based on
49 radiocarbon (^{14}C) depends on the method of EC isolation. The lack of aerosol EC
50 reference materials with "true" ^{14}C values makes it impossible to evaluate the
51 accuracy of various methods for the analysis of ^{14}C -EC in aerosols. In this study, EC
52 isolation methods were evaluated by using samples of mixed biomass burning, vehicle
53 exhaust and coal-combustion. The results show that ^{14}C -EC was not only related to
54 the isolation method but also to the types and proportions of biomass sources in the
55 sample. The hydrolysis (Hypy) method, which can be used to isolate a highly
56 stable portion of EC_{Hypy} and avoid charring, is a more effective and stable approach
57 for the matrix-independent ^{14}C quantification of EC in aerosols. The ^{13}C - EC_{Hypy} and
58 non-fossil EC_{Hypy} values of SRM1649b were -24.9‰ and 11‰, respectively.
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62 1. Introduction

63
64 The elemental carbon (EC) or black carbon (BC) fraction of carbonaceous aerosols
65 (CAs) is derived from the incomplete combustion of fossil fuels or biomass and is
66 responsible for an overall warming effect of the Earth by either absorbing incoming
67 solar radiation in the atmosphere or reducing the albedo of surface materials (i.e.,
68 snow and ice) (Fuzzi et al., 2006; Schwarz et al., 2015; Szidat, 2009; Szidat et al.,
69 2004a; Szidat et al., 2009). The limited understanding of EC aerosol emissions results
70 in poorly constrained estimates of their contribution to anthropogenic climate
71 warming that globally may be second only to CO_2 and regionally, such as over East
72 Asia, the dominant driver of climate change (Chen et al., 2013). Therefore, detailed
73 knowledge of the sources of EC is necessary for the implementation of mitigation
74 strategies for EC reduction. Carbon isotope (^{14}C and ^{13}C) analysis is a powerful tool
75 for unambiguously distinguishing the carbon sources of EC (Currie, 2000; Szidat,
76 2009; Szidat et al., 2009; Gustafsson et al., 2009; Kirillova et al., 2013; Liu et al.,
77 2013; Zencak et al., 2007; Zhang et al., 2019b). Carbon isotope source apportionment
78 requires the physical isolation of organic carbon (OC) and EC, which is complicated
79 by the fact that there is no sharp boundary between OC and EC in carbonaceous
80 aerosols (Elmquist et al., 2006). Therefore, one of the large challenges of this method
81 is the isolation of EC for ^{14}C and ^{13}C analysis.
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83 Based on the thermal stability of EC, several methods for isolating OC and EC
84 from aerosols have been developed. An intercomparison of 9 laboratories for ^{14}C
85 analysis of carbonaceous aerosol samples was conducted in a previous study, and ^{14}C
86 analysis of EC revealed a large deviation of 28-79% between the approaches as a
87 consequence of the different isolation techniques (Szidat et al., 2013). Due to the
88 application of the same principle as aerosol OC and EC measurement, thermo-optical
89 isolation, also named oxygen-based OC-EC isolation, has gradually become the main
90 method for EC isolation in recent years. Although the ^{14}C results of EC between three
91 independent laboratory methods showed good agreement, the recovery of EC differed
92 greatly (Zenker et al., 2017). Recently, hydrolysis (Hypy) has been introduced as
93 an EC isolation method (Meredith et al., 2012; Zhang et al., 2019b). A comparison
94 study of ^{14}C -EC in aerosol samples isolated using the two-step heating method
95 (CTO-375), $\text{EC}_{\text{He/O}_2\text{-475}}$ method and Hypy method was also conducted (Zhang et al.,

2019b). However, the ^{14}C intercomparisons of all studies were mainly restricted to ambient filter samples or urban dust (SRM 1649a/b), for which the “true” ^{14}C activity of EC is not known. As the literature emphasizes (Dasari et al., 2022), even when methods give similar results, it may still be unclear whether the methods give accurate results. In the worst case, if the methods give different results, then it is impossible to determine which method (if any) gives an accurate value (Zenker et al., 2017). Therefore, the key to evaluating the accuracy of different isolation methods is to obtain suitable EC reference materials for the ^{14}C analysis of aerosols.

Carbonaceous aerosols are mainly composed of primary emissions from fossil fuel and biomass combustion and secondary organic compounds (Huang et al., 2014; Zhang et al., 2015). In general, secondary organic aerosols are relatively easy to isolate from EC using methods such as water or organic solvent extraction. However, it is difficult to isolate insoluble OC from EC in primary combustion products. Biomass burning, coal combustion and traffic emissions are the main primary sources of EC in aerosols (Bond et al., 2013). In this study, six samples were synthesized artificially by using biomass combustion (corn straw or pine wood), coal combustion and motor vehicle exhaust samples according to the relative content of fossil carbon and modern carbon in actual aerosols in this study. The theoretical calculated values of the EC contents and EC carbon isotopes in six synthetic samples were determined based on the measured isotopes of each source sample and the elemental carbon/total carbon (EC/TC) measured by using the thermal-optical transmittance (TOT) method. And the calculated ^{14}C value of EC can be taken as the “true” ^{14}C -EC value. Consequently, four EC isolation methods, including Hypy (Zhang et al., 2019b), CTO-375 (Liu et al., 2013), $\text{EC}_{\text{He/O}_2-475}$ (Liu et al., 2017) and EC_{LARA} (Zenker et al., 2017), were selected for EC isolation, and then the EC contents and EC carbon isotopes were compared to the corresponding theoretical calculated value of each synthetic sample. The accuracy of each isolation method was evaluated based on the recovery of the EC contents and carbon isotopes. Finally, the EC contents and EC carbon isotopes (^{14}C and ^{13}C) of urban dust (SRM 1649b) were determined by the isolation method with the best accuracy.

2. Materials and Methods

2.1. Sample collection

Corn straw (*Zea mays*, C4 plant, with a carbon isotope composition that differs significantly from fossil fuels), pine wood (*Pinus tabulaeformis* Carr. woody plant), one type of raw coal in chunks sourced from Yanzhou (YZ) in Shandong Province, and one type of gasoline truck exhaust were selected as the representative sources for biomass burning, coal combustion and vehicle exhaust. Corn straw and pine wood combustion products were collected through a sampling system. Coal was combusted in a high-efficiency stove, and $\text{PM}_{2.5}$ emissions were collected using a dilution sampling system. Vehicle exhaust particles were collected using the on-board emission measurement system. A description of the detailed sampling information was provided in a previous report (Zhang et al., 2019b).

2.2. EC isolation method

CTO-375 method: To achieve the complete removal of the OC from the $\text{EC}_{\text{CTO375}}$

146 fraction, the samples were treated by vaporizing the OC at 375 °C in a muffle furnace
147 in the presence of air using a shorter isolation time of 4 h (Liu et al., 2013).

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149 **EC_{He/O2-475} method:** The EC_{He/O2-475} fractions in the samples were purified in the
150 commercial OC-EC analyzer as follows: 120 s at 200 °C, 150 s at 300 °C, and 180 s at
151 475 °C in an oxidative atmosphere (10% oxygen, 90% helium), followed by 180 s at
152 650 °C in helium. Details of the handling methods were described in a previous report
153 (Liu et al., 2017).

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155 **EC_{LARA} method:** A punch of the water-extracted filter was treated with a
156 thermo-optical OC-EC analyzer using the first three steps of the “Swiss 4S protocol”
157 to remove all remaining water-insoluble OC, giving a residue that constituted the
158 EC_{LARA} sample (Zenker et al., 2017; Zhang et al., 2015).

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160 **Hydropyrolysis:** Each sample was mixed fully with ammonium
161 dioxodithiomolybdate [(NH₄)₂MoO₂S₂] as a catalyst to reach a nominal molybdenum
162 loading of more than 20% of sample carbon weight. The samples were first heated in
163 the reactor tube from ambient temperature to 250 °C at a rate of 300 °C min⁻¹ and then
164 from 300 °C to the final temperature (550 °C) at 8 °C min⁻¹; samples were then held
165 for 5 min under a hydrogen pressure of 15 MPa and a flow rate of 5.0 L/min. The
166 resulting residue was the EC_{Hypy} sample (Zhang et al., 2019b).

167 168 **2.3. OC/EC and carbon isotopes analysis**

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170 The OC and EC were analyzed by a laboratory OC/EC analyzer (Sunset Laboratory,
171 USA) using the NIOSH2 thermal protocol (Maenhaut et al., 2005; Salma et al., 2004).
172 Methods of ¹³C and ¹⁴C analysis for all samples were described in the SI (Liu et al.,
173 2013; Liu et al., 2017).

174 175 **2.4. Nuclear Magnetic Resonance (NMR) and Field Emission Scanning Electron 176 Microscopy (FESEM) analysis**

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178 NMR experiments were performed with an AVANCE III 400-MHz NMR
179 spectrometer (Bruker, Billerica, MA, USA). FESEM experiments were analyzed by
180 the field emission scanning electron microscope (Hitachi su8010, Hitachi, Japan).
181 Detailed experimental methods of NMR and FESEM analysis were described in the
182 SI (Zhang et al., 2019b; Chen et al., 2020).

183 184 **3. Results and discussion**

185 186 **3.1. Comparison of EC purified by different methods**

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188 The six synthetic known samples were made by biomass combustion (corn straw or
189 pine wood), coal combustion and motor vehicle exhaust (Table S1). The hybrid
190 samples were produced according to the proportioning principle, based on the
191 approximate proportions (Figure S1 and Table S2). The average deviations of carbon
192 content, ¹³C_{TC} and *f_M*(TC) of the hybrid samples between the theoretical values and
193 the test values were 0.30%, -0.12‰ and 0.03, respectively (Table S3), and there was
194 no significant difference (T-test, P=0.77, 0.96 and 0.49, respectively). These results
195 show that the samples were well mixed and were therefore suitable for the method

196 comparison experiments.

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198 Before comparing EC recovery rates, it is necessary to obtain relatively accurate
199 EC concentrations. The EC/TC ratios of the four combustion source samples were
200 analyzed by the TOT method (Table S4). The EC/TC results of the four combustion
201 source samples processed using the four isolation methods (CTO-375, EC_{He/O_2-475} ,
202 EC_{LARA} and Hypy) are listed in Table S4. The amount of EC obtained by the
203 CTO-375 method is obviously lower than the results of the other three methods. For
204 example, due to the high content of soot in the vehicle exhaust, the amount of EC can
205 reach about 20%, while the amount of EC in other source samples is less than 10%
206 (Hammes et al., 2007). It indicates that the CTO-375 method has obvious defects in
207 the quantitative analysis of EC content in aerosols. Therefore, this method is not
208 suitable for isolating EC to isotopes analysis. The EC/TC ratios of the coal
209 combustion and motor vehicle exhaust obtained by the other three methods are lower
210 than those of the TOT method, and the EC/TC ratios of the pine burning samples are
211 higher than those of the TOT method. Among the above four methods, the result
212 obtained by the Hypy method is the closest to the result of the TOT method. For the
213 corn straw combustion samples, the Hypy and EC_{LARA} methods are lower than the
214 TOT method, but the results obtained by the EC_{He/O_2-475} are higher than the TOT
215 method. On the one hand, this difference is the problem of the method itself. Each
216 method only isolates a specific part of the EC continuum, rather than all the
217 components of the EC continuum (Currie et al., 2002; de la Rosa et al., 2011; Schmidt
218 et al., 2001). On the other hand, the different isolation effect of the method is due to
219 the difference of the organic carbon structure in the source sample.

220

221 NMR spectroscopy is an essential tool for acquiring detailed structural
222 characterization results of the complex natural organic matter. The four combustion
223 source samples were characterized by using solid-state ^{13}C NMR (Figure S2 and
224 Table S5). The ^{13}C NMR results show that the average lower limit estimate for
225 organic oxygen (Kelemen et al., 2010) from biomass combustion is 37% higher than
226 that for fossil combustion. This finding indicates that there are more
227 oxygen-containing organic carbon components in biomass combustion samples. The
228 more organic oxygen in the sample, the greater the sample's polarity, which
229 contributes to the increased fraction of water-soluble components in biomass
230 combustion compared to the fossil combustion samples. In the absence of oxygen,
231 using an aromatization process based on the cleavage of O-alkylated carbons might
232 overestimate the EC content analyzed by thermal-optical methods (Li et al., 2013).
233 The fraction of methyls in the aliphatics (FMA) (Chen et al., 2020) and nonprotonated
234 aromatics (Kelemen et al., 2010) in pine wood combustion are 22% and 12% higher,
235 respectively, than those in corn straw. These findings indicate that the water solubility
236 of pine wood combustion products is worse than that of corn straw, in addition to the
237 fact that the aromatic structure of pine wood combustion materials is denser. The
238 water-soluble components of the corn straw combustion products were high, such that
239 OC charring has a greatly influenced EC isolation when using the EC_{He/O_2-475} method,
240 while minimally impacting EC isolation following the water extraction EC_{LARA}
241 methods (Zhang et al., 2012).

242

243 By using FESEM in this study (Figure S3), it was observed that the pine wood
244 samples contained a coke structure that was more condensed than the structure of the
245 corn straw samples. In addition, a large number of soot structures were observed in

246 the fossil source samples, indicating that fossil source samples are more condensed
247 than biomass samples.

248

249 In general, due to the differences between the isolation technologies, the EC/TC
250 ratios of the six synthetic known samples (Table S4) revealed a deviation of
251 approximately -99%~+125% between the calculated EC/TC ratios and the EC/TC
252 ratios isolated by the four methods. The deviations isolated by the Hypy, EC_{LARA},
253 EC_{He/O2-475} and CTO-375 methods were approximately -8%~+31%, -39%~+121%,
254 -62%~+125% and -93%~-99%, respectively. The result obtained by the Hypy method
255 is closest to the TOT method, and the average value of the ratio is 1.1, which shows
256 the advantages of stability and reliability in purifying EC.

257

258 **3.2. Comparison of carbon isotopes in EC**

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260 The different EC recoveries of the source samples may lead to different carbon
261 isotope results in the EC from the synthetic known samples. The theoretical EC/TC
262 ratios and ¹³C_{EC} values of each hybrid sample were calculated according to the EC/TC
263 (TOT) ratios and ¹³C_{TC} values in the combustion source and the relative proportion of
264 each source, respectively (Table S3). In the calculation, it is assumed that the ¹³C
265 value in the EC is very close with that in the TC in each source sample. The ¹³C
266 results of the EC from the six synthetic known samples processed by the four isolation
267 methods are shown in Table 1. No significant difference of ¹³C was observed between
268 the results of the four isolation methods and the corresponding theoretical calculation
269 values of all samples (T-test, P>0.05), except for the samples (S1, S2 and S3)
270 containing corn straw combustion material isolated using the EC_{He/O2-475} method,
271 which presented much higher ¹³C values. This finding indicated that the EC isolated
272 using the EC_{He/O2-475} method contained more biomass carbon. The reason for this
273 result may be that the organic carbon of combusted corn straw is charring during the
274 EC isolation process. However, this phenomenon was not observed in the samples
275 containing pine wood combustion, which may be the reason that the ¹³C value of the
276 pine wood combustion sample is close to that of coal combustion.

277

278 The ¹⁴C results of the EC are shown in Table 1 and Figure 1. Fraction of Modern
279 (*f_M*) is used to express the proportion of biomass burning. The theoretical values of
280 *f_M*(EC) in the hybrid samples were obtained according to the proportion of fossil
281 carbon and modern carbon in each sample in accordance with the EC/TC (TOT) ratios
282 (Figure 1). The results show that the *f_M* values obtained by different isolation methods
283 are quite different, and are generally affected by the ratios of combustion source
284 sample EC recovered by different methods. Due to the low recovery rate of EC by
285 CTO-375 method, the *f_M*(EC_{CTO-375}) value is irregular. Generally, the *f_M*(EC_{CTO-375})
286 value obtained by this method is more than twice the theoretical value. For the
287 EC_{He/O2-475} method, due to the influence of biomass burning OC charring the
288 *f_M*(EC_{He/O2-475}) value obtained by this method deviates greatly from the theoretical
289 value. On the whole, *f_M* values obtained by the Hypy and EC_{LARA} methods are
290 relatively close to the theoretical values, but the two methods have their own
291 advantages in the two different sets of samples. For the combustion of herbaceous
292 plants, the Hypy method has a low EC recovery rate for such source samples,
293 resulting in a small *f_M* value. For the combustion of woody combustion, due to the

294 higher EC recovery rate by the Hypy method, the $f_M(\text{EC}_{\text{Hypy}})$ value is slightly higher
295 than that of samples containing herbaceous plants. For the EC_{LARA} method, in the first
296 group of samples with corn straw combustion, the EC recovery rate of each source
297 sample is lower than the theoretical value, so that the $f_M(\text{EC}_{\text{LARA}})$ value obtained by
298 this method is the closest to the theoretical value. However, in the second group of
299 samples containing woody combustion, the $f_M(\text{EC}_{\text{LARA}})$ value obtained by this method
300 was significantly higher than the theoretical value, due to the higher recovery rate of
301 woody combustion EC by the EC_{LARA} method. The results show that the f_M value of
302 EC was not only related to the isolation method but also to the types and proportions
303 of biomass sources in the sample.

304
305 The above results show that the type of biomass combustion affects the efficiency
306 of the isolation method to purify black carbon and the accuracy of the radiocarbon test
307 results. Charcoal, harvest residues and wood materials are the most common biomass
308 fuels used as energy sources (Anenberg et al., 2013). From a global perspective, the
309 types of biomass fuels are complex, but generally can be divided into two categories,
310 herbaceous burning and woody plant burning, the ratio of the two types of biomass is
311 about 58:42 (Table S6). Different regions have different proportions. According to the
312 literature (Bond et al., 2004; Stevens et al., 2017; Streets et al., 2003; Zhang et al.,
313 2019a), developed countries, such as Europe and North America, have a relatively
314 high proportion of woody plants, while developing countries, such as Africa and Asia,
315 have a relatively high proportion of herbaceous plants (Table S6). According to the
316 recovery rate of different types of biomass combustion EC by different methods, the
317 deviations that may be caused by the results of testing ^{14}C in different regions are
318 estimated. The results are listed in Table S6. It can be seen that on a global scale, the
319 f_M value obtained by the Hypy method is the closest to the theoretical value. Therefore,
320 the Hypy method is an effective and stable approach for matrix-independent ^{14}C
321 quantification of EC avoiding charring in aerosols.

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323 3.3. Carbon isotopes of EC in SRM 1649b

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325 SRM 1649a/b, urban dust, was used to check the quality of EC or EC isotope
326 measurement method (Currie et al., 2002; Liu et al., 2013; Szidat et al., 2004b; Reddy
327 et al., 2002; Heal et al., 2011). SRM 1649b was prepared from the same particulate
328 material that was issued in 1982 as SRM 1649 and re-issued in 1999 as SRM 1649a,
329 and the only difference is that the bulk material was sieved to a smaller particle size
330 fraction ⁴⁷. Therefore, the f_M and $\delta^{13}\text{C}$ of TC in SRM 1649b obtained in this study was
331 consistent with SRM 1649a reported by Szidat et al (Szidat et al., 2004b). EC/TC
332 ratios, $f_M\text{-EC}$ and $\delta^{13}\text{C}\text{-EC}$ of SRM1649b isolated using Hypy method in this study
333 and archived data from the literature using different isolation methods are listed in
334 Table 2. The EC/TC ratios varied from 7.5% to 46% determined by the different
335 analyzing methods. However, the value of about ~28% obtained in this study is in
336 good agreement with that obtained by Hypy method (Meredith et al., 2012) and by
337 TOT method (Currie et al., 2002). The f_M and $\delta^{13}\text{C}$ of EC in SRM 1649a ranged from
338 0.038 to 0.153 and from -26.55‰ to -24.8‰ respectively reported by the previous
339 research using the different isolation methods except Hypy (Currie et al., 2002). The
340 corresponding values of 0.11 and -24.9‰ treated by Hypy method in this study are
341 just in the range of archived data. Therefore, the values of ^{13}C and ^{14}C of EC_{Hypy} in
342 SRM 1649b provide a definite and comparable reference for the future research
343 methods.

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4. Conclusions

Carbon isotope (^{14}C and ^{13}C) analysis is a powerful tool for distinguishing the carbon sources in carbonaceous aerosols. As addressed in this work, one main challenge of this method is the isolation of EC or BC for carbon isotope analysis. In this study, six synthetic known samples were collected, including biomass combustion (corn straw or pine wood) and coal combustion products and motor vehicle exhaust. These samples were then used to evaluate four EC isolation methods, which included the Hypy, CTO-375, $\text{EC}_{\text{He/O}_2\text{-475}}$ and EC_{LARA} methods. The results demonstrated the Hypy method was in good agreement with the thermo-optical (TOT) method for the quantification of EC. And the $\text{EC } f_M$ values depended not only on the isolation method but also on the types and proportions of the biomass sources in the samples. The Hypy method is the most appropriate EC isolation method of the four methods reported here, followed by the EC_{LARA} method. The Hypy method, which can be used to isolate a highly stable portion of EC_{Hypy} and avoid charring, is a more effective and stable approach for the matrix-independent ^{14}C quantification of EC in aerosols. The EC_{Hypy} of SRM1649b sample was isolated by the Hypy method. The results indicated that the $^{13}\text{C}\text{-EC}_{\text{Hypy}}$ and non-fossil EC_{Hypy} values of SRM1649b were -24.9% and 11% , respectively. These two isotope values was able to provide a valuable reference for other EC isolation methods.

Supporting Information

Supporting Information may be found in the online version of this article.

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Author contributions

GZ led the study. JL and XZ designed the study, developed the analysis protocols, and wrote the initial manuscript. XZ, SZ, JL, PD, SG, CT, YC and PP provided data, provided comments on the analysis, and contributed and reviewed the final manuscript.

Competing interests

The authors declare that they have no conflict of interest.

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Table 1. The $\delta^{13}\text{C}$ and ^{14}C results of EC from the six synthetic known samples by four isolation methods.

Sample	$\delta^{13}\text{C}\text{‰}(\text{EC})$ (calculated)	$f_M(\text{EC})$ (calculated)	$\delta^{13}\text{C}\text{‰}(\text{EC}_{\text{Hypy}})$ ($\pm 0.5 \text{‰}$)	$f_M(\text{EC}_{\text{Hypy}})$	$\delta^{13}\text{C}\text{‰}(\text{EC}_{\text{CTO-375}})$ ($\pm 0.5 \text{‰}$)	$f_M(\text{EC}_{\text{CTO-375}})$	$\delta^{13}\text{C}\text{‰}(\text{EC}_{\text{He/O2-475}})$ ($\pm 0.5 \text{‰}$)	$f_M(\text{EC}_{\text{He/O2-475}})$	$\delta^{13}\text{C}\text{‰}(\text{EC}_{\text{LARA}})$ ($\pm 0.5 \text{‰}$)	$f_M(\text{EC}_{\text{LARA}})$
S1	-22.98	0.2192	-23.40	0.0946 \pm 0.0013	-22.26	0.4027 \pm 0.0145	-19.69	0.3378 \pm 0.0025	-23.55	0.2140 \pm 0.0023
S2	-22.77	0.2795	-23.42	0.1375 \pm 0.0015	-22.42	0.5332 \pm 0.0172	-17.00	0.3809 \pm 0.0038	-23.77	0.2437 \pm 0.0024
S3	-22.43	0.3201	-23.31	0.1694 \pm 0.0016	-22.60	0.7368 \pm 0.0169	-18.69	0.4547 \pm 0.0029	-24.11	0.2590 \pm 0.0025
S4	-25.16	0.2403	-24.33	0.3145 \pm 0.0045	-25.83	0.4019 \pm 0.0076	-24.36	0.5005 \pm 0.0036	-24.28	0.5101 \pm 0.0030
S5	-25.52	0.3053	-24.55	0.3382 \pm 0.0049	-25.85	0.4779 \pm 0.0119	-24.57	0.6487 \pm 0.0036	-24.54	0.6655 \pm 0.0041
S6	-25.59	0.3483	-24.57	0.3707 \pm 0.0063	-25.76	0.5966 \pm 0.0098	-24.84	0.7226 \pm 0.0033	-24.68	0.7315 \pm 0.0034

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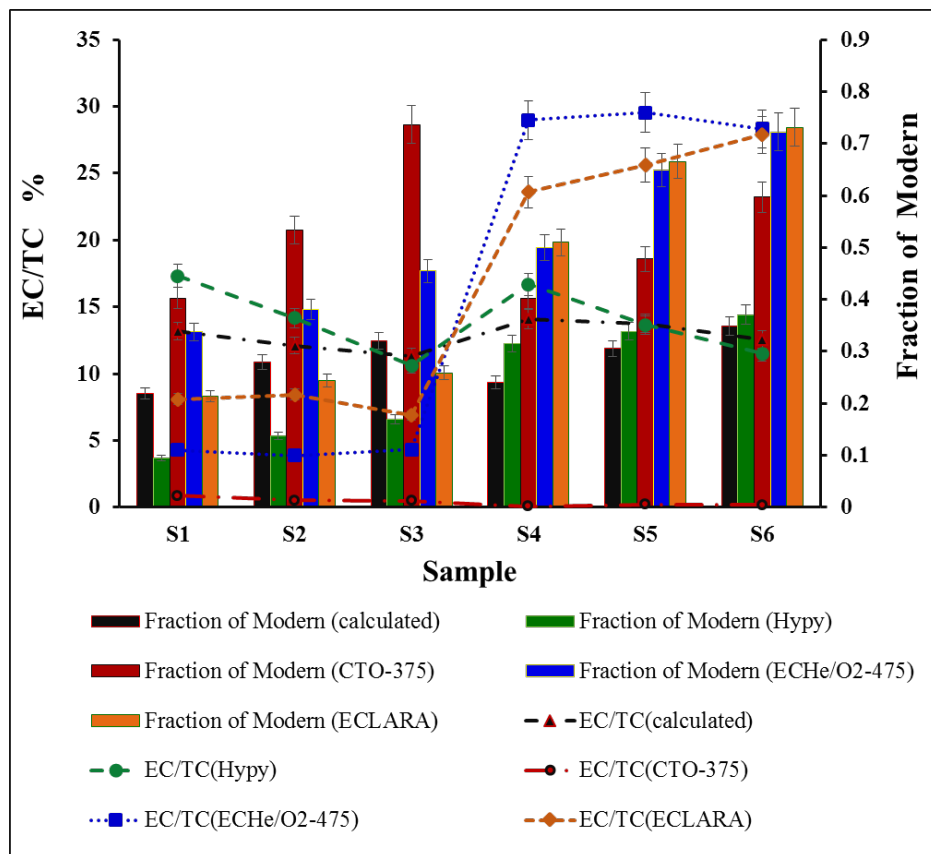
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560 **Table 2.** ^{14}C and ^{13}C analysis results in SRM 1649 a/b.

Sample	f_M	M_C (μg)	$\delta^{13}\text{C}$ (‰ vs VPDB)	EC/TC	reference
1649b TC	0.525±0.002 (n = 1)	220	-25.6±0.5 (n = 2)	0.275± 0.050	this work
1649a TC	0.522±0.018 (n = 5)	12–87	-25.5±0.6 (n = 2)	0.280 ± 0.080	ref (Szidat et al., 2004)
1649a TC	0.510±0.011 (n = 3)	not given	-25.3±0.3 (n = 2)	0.080±0.010	ref (Reddy et al., 2002)
1649a TC	0.610±0.040	not given	not given	not given	ref (Currie et al., 2002)
1649a TC	0.505±0.003	not given	not given	0.458± 0.025	ref (Currie et al., 2002)
1649a TC	0.517±0.004	not given	not given	not given	ref (Currie et al., 2002)
1649b EC	0.108 ± 0.002 (n = 1)	270	-24.9±0.5 (n = 2)	0.275±0.050	this work
1649b EC	0.112±0.080* (n = 1)	~460	-24.9±0.5 (n = 2)	0.275±0.050	this work
1649a EC	0.066±0.020 (n = 4)	37–70	-24.8±0.5 (n = 4)	0.280 ±0.080	ref (Szidat et al., 2004)
1649a EC	0.065±0.014 (n = 3)	~2800	-26.55± 0.04	0.080±0.010	ref (Reddy et al., 2002)
1649a EC	0.140±0.050 (n = 1)	459	not given	~0.280	ref (Liu et al., 2013)
1649a EC	0.150±0.080	not given	not given	not given	ref (Heal et al., 2011)
1649a EC	0.065±0.003 (n = 1)	not given	not given	0.077±0.002	ref (Currie et al., 2002)
1649a EC	0.153±0.002	not given	not given	0.458± 0.025	ref (Currie et al., 2002)
1649a EC	0.038±0.012	not given	not given	0.109± 0.005	ref (Currie et al., 2002)

* The same 1649b EC obtained by the Hypy method was sent to BETA for ^{14}C analysis.

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566 **Figure 1.** The EC/TC% and ^{14}C results of EC from the six synthetic known samples by four isolation methods.

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