



Technical note: Intercomparison Study of the EC Radiocarbon Analysis Methods Using Synthetic Known Samples Xiangyun Zhang^{1,2,3}, Jun Li^{1,2,3}*, Sanyuan Zhu^{1,2,3}, Junwen Liu⁴, Ping Ding⁵, Shutao Gao^{1,2,3}, Chongguo Tian⁶, Yingjun Chen⁷, Ping'an Peng^{1,2,3}, Gan Zhang^{1,2,3*} ¹ State Key Laboratory of Organic Geochemistry and Guangdong province Key Laboratory of Environmental Protection and Resources Utilization, Guangzhou Institute of Geochemistry, Chinese Academy of Sciences, Guangzhou, 510640, China. ²CAS Center for Excellence in Deep Earth Science, Guangzhou, 510640, China. ³ Guangdong-Hong Kong-Macao Joint Laboratory for Environmental Pollution and Control, Guangzhou Institute of Geochemistry, Chinese Academy of Science, Guangzhou 510640, China. ⁴ Institute for Environmental and Climate Research, Jinan University, Guangzhou, 511443, China. ⁵ State Key Laboratory of Isotope Geochemistry, Guangzhou Institute of Geochemistry, Chinese Academy of Sciences, Guangzhou 510640, China. ⁶ Key Laboratory of Coastal Environmental Processes and Ecological Remediation, Yantai Institute of Coastal Zone Research, Chinese Academy of Sciences, Yantai, 264003, China. ⁷ Department of Environmental Science and Engineering, Fudan University, Shanghai, 200438, China. *Correspondence: Jun Li (junli@gig.ac.cn), Gan Zhang (zhanggan@gig.ac.cn)





46 Abstract

47

The accurate identification of elemental carbon (EC) source in aerosol based on 48 49 radiocarbon (14C) depends on the method of EC isolation. The lack of aerosol EC reference materials with "true" ¹⁴C values makes it impossible to evaluate the 50 accuracy of various methods for the analysis of ¹⁴C-EC in aerosols. In this study, EC 51 52 separation methods were evaluated by using samples of mixed biomass burning, vehicle exhaust and coal-combustion. The results show that ¹⁴C-EC was not only 53 54 related to the separation method but also to the types and proportions of biomass 55 sources in the sample. And the Hydropyrolysis (Hypy) method, which can be used to 56 separate a highly stable portion of EC_{Hypy} and avoid charring, is a more effective and stable approach for the matrix-independent ¹⁴C quantification of EC in aerosols. The 57 ¹³C-EC_{Hypy} and non-fossil EC_{Hypy} values of SRM1649b was 24.9‰ and 11%, 58 59 respectively.

60 61

62

63

64 65

1. Introduction

66 67

The elemental carbon (EC) or black carbon (BC) fraction of carbonaceous aerosols 68 69 (CAs) is derived from the incomplete combustion of fossil fuels or biomass and is 70 responsible for an overall warming effect of the Earth by either absorbing incoming 71 solar radiation in the atmosphere or reducing the albedo of surface materials (i.e., 72 snow and ice) (Fuzzi et al., 2006; Schwarz et al., 2015; Szidat, 2009; Szidat et al., 73 2004a; Szidat et al., 2009). The limited understanding of EC aerosol emissions causes 74 poorly constrained anthropogenic climate warming that globally may be second only 75 to CO₂ and regionally, such as over East Asia, the dominant driver of climate change (B. Chen et al., 2013). Therefore, detailed knowledge of the sources of EC is necessary 76 for the implementation of mitigation strategies for EC reduction. Carbon isotope (¹⁴C 77 78 and ¹³C) analysis is a powerful tool for unambiguously distinguishing the carbon 79 sources of EC (Currie, 2000; Szidat, 2009; Szidat et al., 2009; Gustafsson et al., 2009; Kirillova et al., 2013; D. Liu et al., 2013; Zencak et al., 2007; X. Y. Zhang et al., 80 2019b). Carbon isotope source apportionment requires the physical separation of OC 81 and EC, which is complicated by the fact that there is no sharp boundary between 82 organic carbon (OC) and EC in carbonaceous aerosols (Elmquist et al., 83 84 2006). Therefore, one of the large challenges of this method is the isolation of EC for ¹⁴C and ¹³C analysis. 85

86

Based on the thermal stability of EC, several methods for separating OC and EC 87 from aerosols have been developed. An intercomparison of 9 laboratories for ¹⁴C 88 analysis of carbonaceous aerosol samples was conducted in a previous study, and ¹⁴C 89 90 analysis of EC revealed a large deviation of 28-79% between the approaches as a consequence of the different separation techniques (Szidat et al., 2013). Due to the 91 92 same principle as aerosol OC and EC measurement, thermo-optical separation, also named oxygen-based OC-EC separation, has gradually become the main method for 93 EC separation in recent years. Although the ¹⁴C results of EC between three 94 independent laboratory methods showed good agreement, the recovery of EC differed 95 greatly (Zenker et al., 2017). Recently, hydropyrolysis (Hypy) has been introduced as 96





an EC separation method (Meredith et al., 2012; X.Y. Zhang et al., 2019b). A 97 comparison study of ¹⁴C-EC in aerosol samples separated using the two-step heating 98 method (CTO-375), $EC_{He/O2-475}$ method and Hypy method was also conducted (X.Y. 99 Zhang et al., 2019b). However, the ¹⁴C intercomparisons of all studies were mainly 100 restricted to ambient filter samples or urban dust (SRM 1649a/b), for which the "true" 101 ¹⁴C of EC is not known. As the literature emphasizes (Dasari et al., 2022), even when 102 103 methods give similar results, it may still be unclear whether the methods give accurate 104 results. In the worst case, if the methods give different results, then it is impossible to determine which method (if any) gives an accurate value (Zenker et al., 2017). 105 Therefore, the key to evaluating the accuracy of different separation methods is to 106 107 obtain suitable EC reference materials for the ¹⁴C analysis of aerosols.

108

109 Carbonaceous aerosols are mainly composed of primary emissions from fossil fuel 110 and biomass combustion and secondary organic compounds (Huang et al., 2014; Y.L. Zhang et al., 2015). In general, secondary organic aerosols are relatively easy to 111 112 separate from EC using methods such as water or organic solvent extraction. However, 113 it is difficult to separate insoluble OC from EC in primary combustion products. Biomass burning, coal combustion and traffic emissions are the main primary sources 114 of EC in aerosols (Bond et al., 2013). In this study, six samples were synthesized 115 artificially by using biomass combustion (corn straw or pine wood), coal combustion 116 117 and motor vehicle exhaust samples according to the relative content of fossil carbon and modern carbon in actual aerosols in this study. The theoretical calculated values 118 119 of the EC contents and EC carbon isotopes in six synthetic samples were determined based on the measured isotopes of each source sample and the EC/TC measured by 120 using the thermal-optical transmittance (TOT) method. And the calculated ¹⁴C value 121 of EC can be taken as the "true" ¹⁴C-EC value. Consequently, four EC separation 122 methods, including Hypy (X.Y. Zhang et al., 2019b), CTO-375 (D. Liu et al., 2013), 123 EC_{He/O2-475} (J.W. Liu et al., 2017) and EC_{LARA} (Zenker et al., 2017), were selected for 124 EC separation, and then the EC contents and EC carbon isotopes were compared to 125 the corresponding theoretical calculated value of each synthetic sample. The accuracy 126 of each separation method was evaluated based on the recovery of the EC contents 127 and carbon isotopes. Finally, the EC contents and EC carbon isotopes (14 C and 13 C) of 128 urban dust (SRM 1649b) were determined by the separation method with the best 129 130 accuracy.

131

132 **2. Materials and Methods**

- 133
- 134 2.1. Sample collection135

136 Corn straw (Zea mays, C4 plant, with carbon isotopes that differ significantly from fossil fuels), pine wood (Pinus tabulaeformis Carr. woody plant), one type of raw coal 137 138 in chunks sourced from Yanzhou (YZ) in Shandong Province, and one type of gasoline truck were selected as the representative sources for biomass burning, coal 139 140 combustion and vehicle exhaust. Corn straw and pine wood combustion products were collected through a sampling system. Coal was combusted in a high-efficiency 141 stove, and PM_{2.5} emissions were collected using a dilution sampling system. Vehicle 142 exhaust particles were collected using the on-board emission measurement system. A 143 description of the detailed sampling information was provided in a previous report.[X 144 145 *Y Zhang et al.*, 2019b]

146





147 **2.2. EC Purification Method**

148

149 **CTO-375 method:** To achieve the complete removal of the OC from the EC_{CTO375} 150 fraction, the samples were treated by vaporizing the OC at 375 °C in a muffle furnace 151 in the presence of air using a shorter separation time of 4 h (D. Liu et al., 2013).

152

EC_{He/02-475} **method:** The EC_{He/02-475} fractions in the samples were purified in the commercial OC-EC analyzer as follows: 120 s at 200 °C, 150 s at 300 °C, and 180 s at 475 °C in an oxidative atmosphere (10% oxygen, 90% helium), followed by 180 s at 650 °C in helium. Details of the handling methods were described in a previous report (J.W. Liu et al., 2017).

158

ECLARA method: A punch of the water-extracted filter was treated with a
thermo-optical OC-EC analyzer using the first three steps of the "Swiss 4S protocol"
to remove all remaining water-insoluble OC, giving a residue that constituted the
ECLARA sample (Zenker et al., 2017; Y.L. Zhang et al., 2015).

163

Hydropyrolysis: sample mixed fully with 164 Each was ammonium dioxydithiomolybdate [(NH₄)₂MoO₂S₂] as a catalyst to reach a nominal molybdenum 165 loading of 1% by weight. The samples were first heated in the reactor tube from 166 ambient temperature to 250 °C at a rate of 300 °C min⁻¹ and then from 300 °C to the 167 final temperature (550 °C) at 8 °C min⁻¹; samples were then held for 5 min under a 168 169 hydrogen pressure of 15 MPa and a flow rate of 5.0 L/min. The resulting residue was the EC_{Hypy} sample (X.Y. Zhang et al., 2019b). 170

171

172 2.3. OC/EC and carbon isotopes analysis

173

The OC and EC were analyzed by a laboratory OC/EC analyzer (Sunset Laboratory, USA) using the NIOSH2 thermal protocol (Maenhaut et al., 2005; Salma et al., 2004).
Methods of ¹³C and ¹⁴C analysisfor all samples were described in the SI (D. Liu et al., 2013; J.W. Liu et al., 2017).

178

180

179 **3. Results and discussion**

181 **3.1.** Comparison of EC purified by different methods

182

The six synthetic known samples were made by biomass combustion (corn straw or 183 pine wood), coal combustion and motor vehicle exhaust (Table S1). The hybrid 184 samples were produced according to the proportioning principle, based on the 185 186 approximate proportions (Figure S1 and Table S2). The average deviations of carbon content, ${}^{13}C_{TC}$ and $f_M(TC)$ of the hybrid samples between the theoretical values and 187 188 the test values were 0.30%, 0.12‰ and 0.03, respectively (Table S3), and there was no significant difference (T-test, P=0.77, 0.96 and 0.49, respectively). These results 189 190 show that the samples were well mixed and were sufficient to use for the method comparison experiments. 191

192

Before comparing EC recovery rates, it is necessary to obtain relatively accurate EC concentrations. The EC/TC ratios of the four combustion source samples were analyzed by the TOT method (Table S4). The EC/TC results of the four combustion source samples processed using the four separation methods (Hypy, CTO-375,





EC_{He/O2-475} and EC_{LARA}) are listed in Table S4. Compared with the TOT method, the 197 198 amount of EC obtained by the CTO-375 method is obviously lower than the results of the other three methods. For example, due to the high content of soot in the vehicle 199 exhaust, the amount of EC can reach about 20%, while the amount of EC in other 200 source samples is less than 10% (Hammes et al., 2007). It indicates that the CTO-375 201 202 method has obvious defects in the quantitative analysis of black carbon isotopes in 203 aerosols. The EC/TC ratios of the coal combustion and motor vehicle exhaust 204 obtained by the other three methods are lower than those of the TOT method, and the EC/TC ratios of the pine burning samples are higher than those of the TOT method. 205 Among the above four methods, the result obtained by the Hypy method is the closest 206 207 to the result of the TOT method. For the corn straw combustion samples, the Hypy 208 and ECLARA methods are lower than the TOT method, but the results obtained by the 209 EC_{He/O2-475} are higher than the TOT method. On the one hand, this difference is the 210 problem of the method itself. Each method only obtains a specific part of the EC 211 continuum, rather than all the components of the black carbon continuum (Currie et 212 al., 2002; de la Rosa et al., 2011; Schmidt et al., 2001). On the other hand, the 213 different separation effect of the method is due to the difference of the organic carbon structure in the source sample. 214

215

Nuclear Magnetic Resonance (NMR) spectroscopy is an essential tool for acquiring 216 detailed structural characterization results of the complex natural organic matter. The 217 four combustion source samples were characterized by using solid-state ¹³C NMR 218 (Figure S2 and Table S5). The ¹³C NMR results show that the average lower limit 219 estimate for organic oxygen (Kelemen et al., 2010) from biomass combustion is 37% 220 221 higher than that for fossil combustion. This finding indicates that there are more oxygen-containing organic carbon components in biomass combustion samples. The 222 223 more organic oxygen in the sample, the greater the sample's polarity, which contributes to the increased fraction of water-soluble components in biomass 224 225 combustion than in fossil combustion samples. In the absence of oxygen, using an aromatization process based on the cleavage of O-alkylated carbons might 226 overestimate the EC content analyzed by thermal-optical methods (Li et al., 2013). 227 The fraction of methyls in the aliphatics (FMA) (J. Chen et al., 2020) and 228 nonprotonated aromatics (Kelemen et al., 2010) in pine wood combustion are 22% 229 230 and 12% higher, respectively, than those in corn straw. These findings indicate that the 231 water solubility of pine wood combustion products is worse than that of corn straw, in addition to the fact that the aromatic structure of pine wood combustion materials is 232 233 denser. The water-soluble components of the corn straw combustion products were 234 high, such that OC charring has a greatly influenced EC separation when using the 235 EC_{He/O2-475} method, while minimally impacting EC separation following water 236 extraction EC_{LARA} methods (Y. L. Zhang et al., 2012).

237

By using field emission scanning electron microscopy (FESEM) in this study (Figure S3), it was observed that the pine wood samples contained a coke structure that was more compact than the structure of the corn straw samples. In addition, a large number of soot structures were observed in the fossil source samples, indicating that fossil source samples are more compact than biomass samples.

243

In general, due to the differences between the separation technologies, the EC/TC ratios of the six synthetic known samples (Table S4) revealed a deviation of approximately -98%~125% between the calculated EC/TC ratios and the EC/TC





ratios separated by the four methods. The deviations separated by the Hypy, EC_{LARA}, EC_{He/O2-475} and CTO-375 methods were approximately $-8\% \sim 31\%$, $-39\% \sim 121\%$, $-62\% \sim 125\%$ and $-93\% \sim -95\%$, respectively. The result obtained by the Hypy method is closest to the TOT method, and the average value of the ratio is 1.1, which shows the advantages of stability and reliability in purifying EC.

- 253 3.2. Comparison of carbon isotopes in EC
- 254

252

255 The different EC recoveries of the source samples may lead to different carbon isotope results in the EC from the synthetic known samples. The theoretical EC/TC 256 ratios and ¹³C_{EC} values of each hybrid sample were calculated according to the EC/TC 257 (TOT) ratios and ¹³C_{TC} values in the combustion source and the relative proportion of 258 each source, respectively (Table S3). In the calculation, it is assumed that the ¹³C 259 value in the EC is the same as that in the TC in each source sample. The ¹³C results of 260 the EC from the six synthetic known samples processed by the four separation 261 methods are shown in Table 1. No significant difference of ¹³C was observed between 262 the results of the four separation methods and the corresponding theoretical 263 calculation values of all samples (T-test, P>0.05), except for the samples (S1, S2 and 264 S3) containing corn straw combustion material separated using the EC_{He/O2-475} method, 265 which presented much higher ¹³C values. This finding indicated that the EC separated 266 using the $EC_{He/O2-475}$ method contained more biomass carbon. The reason for this 267 result may be that the organic carbon of combusted corn straw is charring during the 268 EC separation process. However, this phenomenon was not observed in the samples 269 containing pine wood combustion, which may be the reason that the ¹³C value of the 270 pine wood combustion sample is close to that of coal combustion. 271

272

The ¹⁴C results of the EC are shown in Table 1 and Figure 1. Fraction of Modern 273 274 $(f_{\rm M})$ is used to express the proportion of biomass burning. The theoretical values of 275 $f_{M}(EC)$ in the hybrid samples were obtained according to the proportion of fossil carbon and modern carbon in each sample in accordance with the EC/TC (TOT) ratios 276 (Figure 1). The results show that the $f_{\rm M}$ values obtained by different separation 277 methods are quite different, and are generally affected by the ratios of combustion 278 source sample EC recovered by different methods. Due to the low recovery rate of EC 279 280 by CTO-375 method, the $f_{\rm M}(\rm EC_{\rm CTO-375})$ value is irregular to follow. Generally, the $f_{\rm M}({\rm EC}_{{\rm CTO-375}})$ value obtained by this method is more than twice the theoretical value. 281 For the EC_{He/O2-475} method, due to the influence of biomass burning OC coking, the 282 $f_{\rm M}({\rm EC}_{\rm He/O2-475})$ value obtained by this method deviates greatly from the theoretical 283 value. On the whole, $f_{\rm M}$ values obtained by the Hypy and EC_{LARA} methods are 284 relatively close to the theoretical values, but the two methods have their own 285 advantages in the two different sets of samples. For the combustion of herbaceous 286 plants, the Hypy method has a low EC recovery rate for such source samples, 287 resulting in a small $f_{\rm M}$ value. For the combustion of woody combustion, due to the 288 higher EC recovery rate by the Hypy method, the $f_{M}(EC_{Hypy})$ value is slightly higher 289 than that of samples containing herbaceous plants. For the EC_{LARA} method, in the first 290 group of samples with corn straw combustion, the EC recovery rate of each source 291 292 sample is lower than the theoretical value, so that the $f_{\rm M}({\rm EC}_{\rm LARA})$ value obtained by 293 this method is the closest to the theoretical value. However, in the second group of 294 samples containing woody combustion, the $f_{\rm M}({\rm EC}_{\rm LARA})$ value obtained by this method





was significantly higher than the theoretical value, due to the higher recovery rate of woody combustion EC by the EC_{LARA} method. The results show that the f_M value of EC was not only related to the separation method but also to the types and proportions of biomass sources in the sample.

299

300 The above results show that the type of biomass combustion affects the efficiency 301 of the separation method to purify black carbon and the accuracy of the radiocarbon 302 test results. Charcoal, harvest residues and wood materials are the most common biomass fuels used as energy sources (Anenberg et al., 2013). From a global 303 304 perspective, the types of biomass fuels are complex, but generally can be divided into 305 two categories, herbaceous burning and woody plant burning, the ratio of the two 306 types of biomass is about 58:42 (Table S6). Different regions have different 307 proportions. According to the literature (Bond et al., 2004; Stevens et al., 2017; 308 Streets et al., 2003; W.M. Zhang et al., 2019a), developed countries, such as Europe 309 and North America, have a relatively high proportion of woody plants, while 310 developing countries, such as Africa and Asia, have a relatively high proportion of 311 herbaceous plants (Table S6). According to the recovery rate of different types of biomass combustion EC by different methods, the deviations that may be caused by 312 the results of testing ¹⁴C in different regions are estimated. The results are listed in 313 Table S6. It can be seen that on a global scale, the $f_{\rm M}$ value obtained by the Hypy 314 method is the closest to the theoretical value. Therefore, the Hypy method is an 315 effective and stable approach for matrix-independent ¹⁴C quantification of EC 316 317 avoiding charring in aerosols.

319 **3.3. Carbon isotopes of EC in SRM 1649b**

318 319 320

321 SRM 1649a/b, urban dust, was used to check the quality of EC or EC isotope 322 measurement method (D. Liu et al., 2013; Szidat et al., 2004b; Reddy et al., 2002; Heal et al., 2011). SRM 1649b was prepared from the same particulate material that 323 was issued in 1982 as SRM 1649 and re-issued in 1999 as SRM 1649a, and the only 324 difference is that the bulk material was sieved to a smaller particle size fraction ⁴⁷. 325 Therefore, the $f_{\rm M}$ and δ^{13} C of TC in SRM 1649b obtained in this study was consistent 326 with SRM 1649a reported by Szidat et al. (Szidat et al., 2004b). EC/TC ratios, $f_{\rm M}$ -EC 327 328 and δ^{13} C-EC of SRM1649b separated using Hypy method in this study and archived data of the literatures using different separation methods are listed in Table S7. The 329 330 EC/TC ratios varied from 7.5% to 46% determined by the different analyzing methods. However, the value of about ~28% obtained in this study is in good 331 agreement with that obtained by Hypy method (Meredith et al., 2012) and by TOT 332 method (Currie et al., 2002). The f_M and δ^{13} C of EC in SRM 1649a ranged from 0.038 333 to 0.153 and from -26.55‰ to -24.8 ‰ respectively reported by the previous research 334 using the different separation methods except Hypy (Currie et al., 2002). The 335 corresponding values of 0.11 and -24.9 ‰ treated by Hypy method in this study are 336 just in the range of archived data. Therefore, the values of ¹³C and ¹⁴C of EC_{Hypy} in 337 SRM 1649b provide a definite and comparable reference for the future research 338 methods. 339

340

341 **4.** Conclusions

342

343 Carbon isotope (${}^{14}C$ and ${}^{13}C$) analysis is a powerful tool for distinguishing the 344 carbon sources of carbonaceous aerosols. As addressed in this work, one main





challenge of this method is the isolation of EC or BC for carbon isotope analysis. In 345 346 this study, six synthetic known samples were collected, including biomass combustion (corn straw or pine wood) and coal combustion products and motor vehicle exhaust. 347 These samples were then used to evaluate four EC separation methods, which 348 included the Hypy, CTO-375, EC_{He/O2-475} and EC_{LARA} methods. The results 349 350 demonstrated the Hypy method was in good agreement with the thermo-optical (TOT) 351 method for the quantification of EC. And the EC f_M values depended not only on the 352 separation method but also on the types and proportions of the biomass sources in the samples. The Hypy method is the most appropriate EC separation method of the four 353 354 methods reported here, followed by the EC_{LARA} method. The Hypy method, which 355 can be used to separate a highly stable portion of EC_{Hypy} and avoid charring, is a more effective and stable approach for the matrix-independent ¹⁴C quantification of EC in 356 aerosols. The EC_{Hypy} of SRM1649b sample was separated by the Hypy method. The 357 results indicated that the ¹³C-EC_{Hypy} and non-fossil EC_{Hypy} values of SRM1649b was 358 359 24.9‰ and 11%, respectively. This result can provide participation value for other 360 separation methods.

361 362

363 Supporting Information

364 Supporting Information may be found in the online version of this article.

365 366

367 Acknowledgments

This work was supported by the National Natural Science Foundation of China (grant 368 nos. 42192510, 42030715 and 41977177), Guangdong Foundation for Program of 369 Technology Research (grant nos. 2019B121205006 370 Science and and 371 2020B1212060053) and Guangzhou Foundation for Program of Science and Technology Research (grant no. 202102080251). 372

374 Author contributions

GZ led the study. JL and XZ designed the study, developed the analysis protocols, and
wrote the initial manuscript. XZ, SZ, JL, PD, SG, CT, YC and PP provided data,
provided comments on the analysis, and contributed and reviewed the final
manuscript.

379

373

380 **Competing interests**

381 The authors declare that they have no conflict of interest.

382383 References

- Anenberg, S. C., Balakrishnan, K., Jetter, J., Masera, O., Mehta, S., Moss, J. and
 Ramanathan, V.: Cleaner Cooking Solutions to Achieve Health, Climate, and
 Economic Cobenefits, Environ. Sci. Technol., 47(9), 3944-3952,
 doi:10.1021/es304942e, 2013.
- Bond, T. C., et al.: Bounding the role of black carbon in the climate system: A
 scientific assessment, J. Geophys. Res.-Atmos., 118(11), 5380-5552,
 doi:10.1002/jgrd.50171, 2013.
- Bond, T. C., Streets, D. G., Yarber, K. F., Nelson, S. M., Woo, J. H. and Klimont, Z.: A
 technology-based global inventory of black and organic carbon emissions
 from combustion, J. Geophys. Res.-Atmos., 109(D14), 43,





394	doi:10.1029/2003jd003697, 2004.
395	Chen, B., et al.: Source Forensics of Black Carbon Aerosols from China, Environ. Sci.
396	Technol., 47(16), 9102-9108, doi:10.1021/es401599r, 2013.
397	Chen, J., Jia, W. L., Yu, C. L., Zhang, X. Y. and Peng, P. A.: Bound hydrocarbons and
398	structure of pyrobitumen rapidly formed by asphaltene cracking: Implications
399	for oil-source correlation, Org. Geochem., 146, 16,
400	doi:10.1016/j.orggeochem.2020.104053, 2020.
401	Currie, L. A.: Evolution and multidisciplinary frontiers of C-14 aerosol science,
402	Radiocarbon, 42(1), 115-126, 2000.
403	Currie, L. A., et al.: A critical evaluation of interlaboratory data on total, elemental,
404	and isotopic carbon in the carbonaceous particle reference material, NIST
405	SRM 1649a, J. Res. Natl. Inst. Stand. Technol., 107(3), 279-298,
406	doi:10.6028/jres.107.022, 2002.
407	Dasari, S., Widory, D.: Radiocarbon (14C) Analysis of Carbonaceous Aerosols:
408	Revisiting the Existing Analytical Techniques for Isolation of Black Carbon,
409	Front. Environ. Sci., 10, 907467, doi: 10.3389/fenvs.2022.907467, 2022.
410	de la Rosa, J. M., Garcia, L. S., de Andres, J. R., Gonzalez-Vila, F. J., Gonzalez-Perez,
411	J. A. and Knicker, H.: Contribution of black carbon in recent sediments of the
412	Gulf of Cadiz: Applicability of different quantification methodologies, Quat.
413	Int., 243(2), 264-272, doi:10.1016/j.quaint.2011.01.034, 2011.
414	Elmquist, M., Cornelissen, G., Kukulska, Z. and Gustafsson, O.: Distinct oxidative
415	stabilities of char versus soot black carbon: Implications for quantification and
416	environmental recalcitrance, Glob. Biogeochem. Cycle, 20(2), 11,
417	doi:10.1029/2005gb002629, 2006.
418	Fuzzi, S., et al.: Critical assessment of the current state of scientific knowledge,
419	terminology, and research needs concerning the role of organic aerosols in the
420	atmosphere, climate, and global change, Atmos. Chem. Phys., 6, 2017-2038,
421	doi:10.5194/acp-6-2017-2006, 2006.
422	Gustafsson, O., Krusa, M., Zencak, Z., Sheesley, R. J., Granat, L., Engstrom, E.,
423	Praveen, P. S., Rao, P. S. P., Leck, C. and Rodhe, H.: Brown Clouds over
424	South Asia: Biomass or Fossil Fuel Combustion?, Science, 323(5913),
425	495-498, doi:10.1126/science.1164857, 2009.
426	Hammes, K., et al.: Comparison of quantification methods to measure fire-derived
427	(black/elemental) carbon in soils and sediments using reference materials from
428	soil, water, sediment and the atmosphere, Glob. Biogeochem. Cycle, 21(3), 18,
429	doi:10.1029/2006gb002914, 2007.
430	Huang, R. J., et al.: High secondary aerosol contribution to particulate pollution
431	during haze events in China, Nature, $514(7521)$, $218-222$,
432	doi:10.1038/nature13774, 2014.
433	Kelemen, S. R., et al.: Characterization of solid bitumens originating from thermal
434	chemical alteration and thermochemical sulfate reduction, Geochim.
435	Cosmochini. Acta, 74(18), 5505-5552, doi:10.1010/j.gca.2010.06.013, 2010.
430 427	KIIIIIOva, E. N., Andersson, A., Sneesley, K. J., Krusa, M., Fraveen, P. S., Budnavant, K. Safai, D.D. Dao, D.S. D. and Custofsson, O. C. 12, and C. 14 based study
43/	K., Salai, F. D., Kao, F. S. F. and Gustalsson, O.: C-15- and C-14-based study





420	of comments and characteria and contains of contained while contains and contained and
438	of sources and atmospheric processing of water-soluble organic carbon $(WSOC)$ in Source Asian according to Complex Processing of Water-soluble organic carbon
439	(wSOC) in South Asian aerosois, J. Geophys. ResAtmos., $118(2)$, $614-626$,
440	doi:10.1002/jgrd.50130, 2013.
441	Li, X. M., Snen, Q. R., Zhang, D. Q., Mei, X. L., Ran, W., Xu, Y. C. and Yu, G. H.: D_{i} D_{i}
442	Functional Groups Determine Biochar Properties (pH and EC) as Studied by
443	Two-Dimensional C-13 NMR Correlation Spectroscopy, PLoS One, 8(6), 7,
444	doi:10.13/1/journal.pone.0065949, 2013.
445	Liu, D., Li, J., Zhang, Y. L., Xu, Y., Liu, X., Ding, P., Shen, C. D., Chen, Y. J., Tian, C.
446	G and Zhang, G: The Use of Levoglucosan and Radiocarbon for Source
447	Apportionment of PM2.5 Carbonaceous Aerosols at a Background Site in East
448	China, Environ. Sci. Technol., $4/(18)$, 10454-10461, doi:10.1021/es401250k,
449	
450	Liu, J. W., Li, J., Ding, P., Zhang, Y. L., Liu, D., Shen, C. D. and Zhang, G.:
451	Optimizing isolation protocol of organic carbon and elemental carbon for C-14
452	analysis using fine particulate samples, Atmos. Environ., 154, 9-19,
453	doi:10.1016/j.atmosenv.2017.01.027, 2017.
454	Maenhaut, W., Raes, N., Chi, X. G., Cafmeyer, J., Wang, W. and Salma, I.: Chemical
455	composition and mass closure for fine and coarse aerosols at a kerbside in
456	Budapest, Hungary, in spring 2002, X-Ray Spectrom., 34(4), 290-296,
457	doi:10.1002/xrs.820, 2005.
458	Meredith, W., Ascough, P. L., Bird, M. I., Large, D. J., Snape, C. E., Sun, Y. and
459	Tilston, E. L.: Assessment of hydropyrolysis as a method for the quantification
460	of black carbon using standard reference materials, Geochim. Cosmochim.
461	Acta, 97, 131-147, doi:10.1016/j.gca.2012.08.037, 2012.
462	Salma, I., Chi, X. G. and Maenhaut, W.: Elemental and organic carbon in urban
463	canyon and background environments in Budapest, Hungary, Atmos. Environ.,
464	38(1), 27-36, doi:10.1016/j.atmosenv.2003.09.047, 2004.
465	Schmidt, M. W. I., Skjemstad, J. O., Czimczik, C. I., Glaser, B., Prentice, K. M.,
466	Gelinas, Y. and Kuhlbusch, T. A. J.: Comparative analysis of black carbon in
467	soils, Glob. Biogeochem. Cycle, 15(1), 163-167, doi:10.1029/2000gb001284,
468	2001.
469	Schwarz, J. P., Holloway, J. S., Katich, J. M., McKeen, S., Kort, E. A., Smith, M. L.,
470	Ryerson, T. B., Sweeney, C. and Peischl, J.: Black Carbon Emissions from the
471	Bakken Oil and Gas Development Region, Environ. Sci. Technol. Lett., 2(10),
472	281-285, doi:10.1021/acs.estlett.5b00225, 2015.
473	Stevens, N., Lehmann, C. E. R., Murphy, B. P. and Durigan, G.: Savanna woody
474	encroachment is widespread across three continents, Glob. Change Biol.,
475	23(1), 235-244, doi:10.1111/gcb.13409, 2017.
476	Streets, D. G., Yarber, K. F., Woo, J. H. and Carmichael, G. R.: Biomass burning in
477	Asia: Annual and seasonal estimates and atmospheric emissions, Glob.
478	Biogeochem. Cycle, 17(4), 20, doi:10.1029/2003gb002040, 2003.
479	Szidat, S.: ATMOSPHERE Sources of Asian Haze, Science, 323(5913), 470-471,
480	doi:10.1126/science.1169407, 2009.
481	Szidat, S., et al.: INTERCOMPARISON OF C-14 ANALYSIS OF





482	CARBONACEOUS AEROSOLS: EXERCISE 2009, Radiocarbon, 55(2-3),
483	1496-1509, doi:10.2458/azu_js_rc.55.16314, 2013.
484	Szidat, S., et al.: Radiocarbon (C-14)-deduced biogenic and anthropogenic
485	contributions to organic carbon (OC) of urban aerosols from Zurich,
486	Switzerland, Atmos. Environ., 38(24), 4035-4044,
487	doi:10.1016/j.atmosenv.2004.03.066, 2004a.
488	Szidat, S., Jenk, T. M., Gaggeler, H. W., Synal, H. A., Hajdas, I., Bonani, G. and
489	Saurer, M.: THEODORE, a two-step heating system for the EC/OC
490	determination of radiocarbon (C-14) in the environment, Nucl. Instrum.
491	Methods Phys. Res. Sect. B-Beam Interact. Mater. Atoms, 223, 829-836,
492	doi:10.1016/j.nimb.2004.04.153, 2004b.
493	Szidat, S., M. Ruff, Perron, N., Wacker, L., Synal, H. A., Hallquist, M., Shannigrahi,
494	A. S., Yttri, K. E., Dye, C. and Simpson, D.: Fossil and non-fossil sources of
495	organic carbon (OC) and elemental carbon (EC) in Goteborg, Sweden, Atmos.
496	Chem. Phys., 9(5), 1521-1535, doi:10.5194/acp-9-1805-2009, 2009.
497	Zencak, Z., Elmquist, M. and Gustafsson, O.: Quantification and radiocarbon source
498	apportionment of black carbon in atmospheric aerosols using the CTO-375
499	method, Atmos. Environ., 41(36), 7895-7906,
500	doi:10.1016/j.atmosenv.2007.06.006, 2007.
501	Zenker, K., Vonwiller, M., Szidat, S., Calzolai, G., Giannoni, M., Bernardoni, V.,
502	Jedynska, A. D., Henzing, B., Meijer, H. A. J. and Dusek, U.: Evaluation and
503	Inter-Comparison of Oxygen-Based OC-EC Separation Methods for
504	Radiocarbon Analysis of Ambient Aerosol Particle Samples, Atmosphere,
505	8(11), 21, doi:10.3390/atmos8110226, 2017.
506	Zhang, W. M., Brandt, M., Penuelas, J., Guichard, F., Tong, X. Y., Tian, F. and
507	Fensholt, R.: Ecosystem structural changes controlled by altered rainfall
508	climatology in tropical savannas, Nat. Commun., 10, 7,
509	doi:10.1038/s41467-019-08602-6, 2019a.
510	Zhang, X. Y., et al.: Isolation and radiocarbon analysis of elemental carbon in
511	atmospheric aerosols using hydropyrolysis, Atmos. Environ., 198, 381-386,
512	doi:10.1016/j.atmosenv.2018.11.005, 2019b.
513	Zhang, Y. L., Perron, N., Ciobanu, V. G., Zotter, P., Minguillon, M. C., Wacker, L.,
514	Prevot, A. S. H., Baltensperger, U. and Szidat, S.: On the isolation of OC and
515	EC and the optimal strategy of radiocarbon-based source apportionment of
516	carbonaceous aerosols, Atmos. Chem. Phys., 12(22), 10841-10856,
517	doi:10.5194/acp-12-10841-2012, 2012.
518	Zhang, Y. L., Schnelle-Kreis, J., Abbaszade, G., Zimmermann, R., Zotter, P., Shen, R.
519	R., Schafer, K., Shao, L. Y., Prevot, A. S. H. and Szidat, S.: Source
520	Apportionment of Elemental Carbon in Beijing, China: Insights from
521	Radiocarbon and Organic Marker Measurements, Environ. Sci. Technol.,
522	49(14), 8408-8415, doi:10.1021/acs.est.5b01944, 2015.
523	NIST: Certificate of Analysis for Standard Reference Material 1649b, Urban Dust;
524	National Institute of Standards and Technology: Gaithersburg, MD, 2007.



Sample	δ ¹³ C‰(EC) (calculated)	f _M (EC) (calculated)	δ ¹³ C‰(EC _{Hypy}) (±0.5 ‰)	fM (ECHypy)	δ ¹³ C‰(ECcto-375) (±0.5 ‰)	fm (ECcto-375)	δ ¹³ C‰(EC _{He/O2-475}) (±0.5 ‰)	fm (ECHe/02-475)	δ ¹³ C‰(EC _{LARA}) (±0.5 ‰)	fm (ECLARA)
$\mathbf{S1}$	-22.98	0.2192	-23.40	0.0946 ± 0.0013	-22.26	0.4027 ± 0.0145	-19.69	0.3378 ± 0.0025	-23.55	0.2140 ± 0.0023
S2	-22.77	0.2795	-23.42	0.1375 ± 0.0015	-22.42	0.5332 ± 0.0172	-17.00	0.3809 ± 0.0038	-23.77	0.2437 ± 0.0024
S3	-22.43	0.3201	-23.31	0.1694 ± 0.0016	-22.60	0.7368 ± 0.0169	-18.69	0.4547 ± 0.0029	-24.11	0.2590 ± 0.0025
$\mathbf{S4}$	-25.16	0.2403	-24.33	0.3145 ± 0.0045	-25.83	0.4019 ± 0.0076	-24.36	0.5005 ± 0.0036	-24.28	0.5101 ± 0.0030
S5	-25.52	0.3053	-24.55	0.3382 ± 0.0049	-25.85	0.4779 ± 0.0119	-24.57	0.6487 ± 0.0036	-24.54	0.6655 ± 0.0041
$\mathbf{S6}$	-25.59	0.3483	-24.57	0.3707 ± 0.0063	-25.76	0.5966 ± 0.0098	-24.84	0.7226 ± 0.0033	-24.68	0.7315 ± 0.0034











13

541 542