

1 **Anthropogenic amplification of biogenic secondary organic aerosol production**
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1 **Abstract**

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3 Biogenic secondary organic aerosols (SOA) contribute to a large fraction of fine aerosols globally,
4 impacting air quality and climate. The formation of biogenic SOA depends on not only emissions
5 of biogenic volatile organic compounds (BVOCs) but also anthropogenic pollutants including
6 primary organic aerosol, sulfur dioxide (SO_2), and nitrogen oxides (NO_x). However, the
7 anthropogenic impact on biogenic SOA production (AIBS) remains unclear. Here we use the
8 decadal trend and variability of observed OA in the southeast US, combined with a global
9 chemistry-climate model, to better constrain AIBS. We show that the reduction in SO_2 emissions
10 can only explain 40% of the decreasing decadal trend of OA in this region, constrained by the low
11 summertime month-to-month variability of surface OA. We hypothesize that the rest of OA
12 decreasing trend is largely due to reduction in NO_x emissions. By implementing a scheme for
13 monoterpene SOA with enhanced sensitivity to NO_x , our model can reproduce the decadal trend
14 and variability of OA in this region. Extending to centennial scale, our model shows that global
15 SOA production increases by 36% despite BVOC reductions from preindustrial period to present
16 day, largely amplified by AIBS. Our work suggests a strong coupling between anthropogenic and
17 biogenic emissions in biogenic SOA production that is missing from current climate models.

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1 **1. Introduction**

2 Terrestrial vegetation emits more than 1 Pg per year of BVOCs (Guenther et al., 2012), leading to
3 a major source of SOA in the atmosphere (Goldstein and Galbally, 2007). SOA exerts significant
4 impacts on climate, air quality and human welfare (Shrivastava et al., 2017; Pye et al., 2021), but
5 is not well represented in climate models. Global climate models differ largely in simulated SOA
6 burden, variability, and radiative effects (Tsigaridis et al., 2014) due to complexity associated with
7 emission of precursors, multiphase chemical and physical processes, aging, radiative properties,
8 and other processes (Shrivastava et al., 2017). Many climate models simply scale SOA yield with
9 BVOC precursors (Horowitz et al., 2020; Carslaw et al., 2013; Koch et al., 2011).

10
11 Current understanding of biogenic SOA formation has advanced far beyond this simple scaling of
12 BVOC emissions. SOA formation from BVOC oxidation is largely dependent on its oxidants
13 ($\text{OH}/\text{O}_3/\text{NO}_3$) and the yields show non-linear behavior under different NO_x conditions (Ng et al.,
14 2017; Presto et al., 2005). One advanced scheme is the Volatility Basis Set (VBS), in which
15 intermediate semivolatile products from the oxidation of BVOCs are grouped into volatility bins
16 and can reversibly condense onto pre-existing organic aerosols (Donahue et al., 2006; Pye et al.,
17 2010). VBS accounts for the dependence of SOA formation on atmospheric oxidants, NO_x -
18 dependent chemical regimes, POA and temperature. Some studies showed that VBS schemes
19 underestimated observations and that photochemical aging schemes with varying complexity may
20 improve simulation results in different regions and seasons (Zheng et al., 2015; Robinson et al.,
21 2007; Oak et al., 2022). Another pathway is through reactive uptake of smaller molecules onto
22 aqueous aerosols. Several isoprene oxidation products, such as epoxides (IEPOX) (Paulot et al.,
23 2009) and glyoxal (Liggio, 2005; Li et al., 2016), though often not directly condensable due to
24 their high equilibrium vapor pressure, can undergo aqueous phase reactions and oligomerize in the
25 condensed phase. The detailed mechanism is complicated by aerosol acidity, composition, and
26 coating (Shrivastava et al., 2017). These advancements highlight the role of anthropogenic
27 emissions modulating biogenic SOA formation through nitrogen oxides (NO_x), SO_2 and primary
28 organic aerosol (POA).

29
30 One major uncertainty is to what extent anthropogenic emissions modulate biogenic SOA
31 formation. In the southeast US (SEUS), a region largely covered by natural vegetation and also
32 heavily populated, organic aerosol shows a decreasing trend in the recent two decades (Kim et al.,
33 2015; Attwood et al., 2014), likely due to reductions in POA (Blanchard et al., 2016; Ridley et al.,
34 2018; Liu et al., 2023), sulfate and aerosol water (Christiansen et al., 2020; Ridley et al., 2018;
35 Marais et al., 2017; Malm et al., 2017; Blanchard et al., 2016; Liu et al., 2023) and NO_x (Zheng et
36 al., 2015; Xu et al., 2015; Pye et al., 2019). Several studies suggest that SO_2 largely modulates
37 SOA through reactive uptake of IEPOX (Pye et al., 2013; Marais et al., 2017), but the acidity-
38 catalyzed sulfate uptake mechanism appears to overestimate the trend of OA reduction in the
39 SEUS (Zheng et al., 2020). The role of NO_x remains unclear. While SOA yield generally decreases
40 with NO_x level due to fragmentation of large molecules (Kroll and Seinfeld, 2008), recent studies
41 show that NO_x can in fact increase SOA production (Zheng et al., 2015; Xu et al., 2015; Pye et al.,
42 2019; Pullinen et al., 2020). The combined effect of NO_x , SO_2 and POA can be significant (Carlton
43 et al., 2010; Hoyle et al., 2011), but remain unconstrained by ambient observations.

44
45 Here we use the decadal trend and variability of observed OA in the southeast US, combined with
46 other observational datasets and a global chemistry-climate model (GFDL AM4.1) (Horowitz et

1 al., 2020), to better constrain the anthropogenic impact on biogenic SOA production (AIBS). We
2 use three schemes (summarized in Table 1 and detailed in Methods), to investigate the AIBS from
3 decadal to centennial time scales.

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4 5 2. Methods

6 2.1 GFDL AM4.1

7 The Geophysical Fluid Dynamics Laboratory (GFDL)'s Atmospheric Model version 4.1 (AM4.1)
8 (Horowitz et al., 2020) is a three-dimensional global chemistry-climate model that includes
9 interactive simulation of stratospheric chemistry and tropospheric O₃-NO_x-CO-VOC and bulk
10 aerosol chemistry, allowing explicit treatment of aerosol reactive uptake of IEPOX and glyoxal
11 (Li et al., 2016, 2018; Mao et al., 2013). AM4.1 has 49 vertical levels from surface to 1Pa (~80km).
12 We conduct AM4.1 simulations at a horizontal resolution of 1°×1.25° latitude by longitude and a
13 main dynamical atmosphere time step of 30 minutes. Annual varying historical anthropogenic
14 emissions from pre-industrial era to present day (1849 to 2016) are from the Community Emissions
15 Data System (CEDS), (Hoesly et al., 2018) and the data set of van Marle et al. (2017), which are
16 developed in support of the Coupled Model Intercomparison Project Phase 6 (CMIP6). Global fire
17 emissions are based on Global Fire Emissions Database version 4 (GFED4), the Fire Modeling
18 Intercomparison Project (FireMIP), visibility-observations and Global Charcoal Database (GCD).
19 Biogenic isoprene and monoterpene emissions are calculated online by the Model of Emissions of
20 Gases and Aerosols from Nature version 2.1 (MEGAN2.1), using empirical functions of plant-
21 functional-type (PFT)-specific emission basal factors, leaf area index (LAI), temperature and light.
22 Dependence of soil moisture, O₃ and CO₂ are neglected due to large uncertainties. LAI values
23 follow an annual cycle of the year 1992 and PFTs are prescribed at the 1992 level. The gas-phase
24 and aerosol chemistry is detailed in Horowitz et al. (2020), in which heterogenous reactive uptake
25 of HO₂, HO₃, NO₂, N₂O₅, NO₃, SO₂, IEPOX and glyoxal onto aerosol surfaces are included. Dry
26 and wet deposition of gases are described in Paulot et al. (2016). More details could be found in
27 Horowitz et al. (2020) and Dunne et al. (2020). Radiative effects of SOA is calculated assuming
28 SOA is externally mixed from other aerosols (Horowitz et al., 2020), although ISOA is formed
29 through sulfate uptake in the chemistry module.

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31 We perform simulations for years 1998-2016 for present day (PD) and 1870-1888 for pre-
32 industrial period (PI). In each simulation, the first two years are discarded as spin-up. The
33 remaining 17 years are used for analysis. The PD simulations are nudged with reanalysis winds
34 from NCEP-DOE Reanalysis 2. The PI simulations are free running with no nudging. All
35 simulations are driven by observed or reconstructed sea surface temperature and sea-ice (Horowitz
36 et al., 2020). In the two PI simulations, we scale up the isoprene and monoterpene emission basal
37 factors by 35% to account for the higher natural vegetation cover at pre-industrial period than
38 today, equivalent to a 26% reduction of natural vegetation cover from PI to PD (Unger, 2014). We
39 apply this single scaling factor to BVOC emission basal factors as an idealized study instead of
40 using reconstructed land cover type and LAI to avoid uncertainties in historical vegetation
41 reconstructions.

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42 2.2 Modeling of SOA formation

43 In GFDL AM4.1, SOA is composed of anthropogenic SOA (ASOA), isoprene SOA (ISOA) and
44 monoterpene SOA (TSOA). ASOA is formed through the oxidation of C₄H₁₀ by OH in all
45 simulations. In the default "Simple" scheme, ISOA and TSOA are assumed to be produced with a

1 pseudo-emission equivalent to a 10% per-carbon yield of the interactively calculated isoprene and
2 monoterpane emissions, respectively. This 10% yield in the Simple scheme is consistent with
3 previous model versions GFDL AM3 and AM4.0 and within the range of estimates suggested by
4 other studies. For example, a chemical transport model GEOS-Chem assumed a 3% yield for
5 isoprene and a 10% yield of monoterpane emissions (Pye et al., 2010; Pai et al., 2020). However,
6 a study using a more complex scheme suggested a SOA yield from isoprene of 13% per carbon
7 (Bates and Jacob, 2019).

8
9 We implement a complex (CMPX) SOA scheme in GFDL AM4.1, in which isoprene and
10 monoterpenes are oxidized by OH, O₃ and NO₃ to form ISOA and TSOA. ISOA is computed
11 through the aqueous-phase uptake of IEPOX and glyoxal onto sulfate aerosol. The uptake
12 coefficients for IEPOX and glyoxal are set to 0.001, different than previous studies using higher
13 or acidity-dependent uptake coefficients (Marais et al., 2016; Lin et al., 2014a). This is supported
14 by the OA month-to-month variability (MMV) in summer and its decadal trend over SEUS, as a
15 previous model with acidity-dependent uptake coefficients shows too high of MMV and too much
16 OA in the early 2000s (Zheng et al., 2020). The uptake rate coefficients can be even lower due to
17 the effect of aerosol-phase state (Zhang et al., 2018b). To avoid uncertainties associated with
18 aerosol acidity, relative humidity, and coating effect, we here apply uptake coefficient of 0.001 for
19 both IEPOX and glyoxal. This leads to good agreement between our model and observation in
20 SEUS on both OA magnitude and summertime MMV (Figure 1, S1, and S2).

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22 In the updated “CMPX” scheme, TSOA is calculated by a 4-product Volatility Basis Set (VBS)
23 summarized in Table 1. Organic peroxy radicals (RO₂) formed from OH- and O₃-initiated
24 oxidation of monoterpane can react with NO under high-NO_x conditions and with HO₂ under
25 low-NO_x conditions. The low-NO_x pathway (RO₂+HO₂) has higher yields for SOA than the
26 high-NO_x pathway (RO₂+NO) (Pye et al., 2010; Zheng et al., 2015). The branching ratio
27 between the low- versus high-NO_x pathways are defined as:

$$29 \quad \beta_{NO} = \frac{k_{RO2+NO} * [NO]}{k_{RO2+NO} * [NO] + k_{RO2+HO2} * [HO_2]}$$

30
31 Where k_{RO2+NO} and k_{RO2+HO2} represent the reaction rate coefficients of RO₂+NO and RO₂+HO₂,
32 respectively. At nighttime, the NO₃-initiated oxidation of monoterpenes has a high yield of
33 organic nitrates and contributes a significant amount of SOA (Ng et al., 2017). The surrogate
34 TSOA products are implemented in addition to the original gas-phase monoterpane oxidation
35 chemistry in AM4.1, and the implementation does not doubt count reductions of OH, O₃ and
36 NO₃. There is little difference in the concentration of these gases between the CMPX and Simple
37 simulations. The gas-phase chemistry has been validated in Horowitz et al. (2020) and in Figure
38 S3 in which we show that summertime surface O₃ and NO₂ in SEUS well reproduce their
39 observed decreasing trend.

40
41 Such semi-empirical partitioning-based VBS schemes have been widely used in chemistry-
42 climate models and Earth system models (e.g. Zheng et al., 2015; Tilmes et al., 2019). Recent
43 research show that these schemes may underestimate SOA formation without considering further
44 aging processes, such as oligomerization in the organic phase and aqueous-phase reactions (Hu
45 et al., 2013; Yu et al., 2021; Oak et al., 2022 and references therein). One major recent identified

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1 explicit mechanism is the formation of monoterpene-derived highly oxygenated molecules
2 (HOMs) through the autoxidation of peroxy radicals (Crounse et al., 2013; Ehn et al., 2014; Pye
3 et al., 2019). Mechanistic schemes of monoterpene-derived SOA have been developed with
4 varying complexity at a cost of more tracers and reactions (Pye et al., 2019; Berkemeier et al.,
5 2020; Pullinen et al., 2020; Yu et al., 2021), which may not be mature for a global climate model
6 as part of an Earth system model considering large uncertainties associated with multi-phase
7 processes and increased computational cost. In this study, in addition to the semi-empirical VBS
8 scheme, we implement a simplified photochemical aging parameterization to the semivolatile
9 oxidation products of terpenes in the CMPX scheme (CMPX_ag) (Zheng et al., 2015), to account
10 for the decrease in volatility as a result of OH oxidation (Donahue et al., 2012). We apply a rate
11 constant of $k_{OH} = 4 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$ (Robinson et al., 2007), in line with recent estimates of
12 $2-4 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$ for terpene SOA (Donahue et al., 2012; Isaacman-VanWertz et al.,
13 2018). Including the aging scheme in CMPX does not increase computational cost notably. This
14 simplified aging scheme does not explicitly represent up-to-date knowledge of SOA chemistry
15 but similarly increases the SOA burden as well as the sensitivity of SOA to NO_x, improving the
16 model underestimate of SOA by the VBS scheme. The details of the Simple, CMPX, and
17 CMPX_ag schemes are summarized in Table 1.

19 **2.3 Observational datasets**

20 For model evaluation we use long-term measurements of organic aerosol (OA) or organic carbon
21 (OC). We do not use explicit SOA tracers in this study because it is not suitable to use short-term
22 observations to validate simulated results from a chemistry-climate model like AM4.1, in which
23 meteorology is not offline specified by reanalysis data but is free running in the dynamic core.
24 However, long-term (covering at least months to years) measurement of explicit SOA species is
25 not available. We use filter measurement of organic carbon from two surface aerosol measurement
26 networks in the US: IMPROVE (the Interagency Monitoring of Protected Visual Environments)
27 (Solomon et al., 2014) and SEARCH (the SouthEastern Aerosol Research and Characterization)
28 (Edgerton et al., 2005). IMPROVE and SEARCH report daily average organic carbon
29 measurements every 3 days. We focus on SEUS which is both heavily vegetated and populated.
30 We select 20 IMPROVE sites and 3 SEARCH rural sites within the SEUS region (29-37°N, 74-
31 96°W) and calculate monthly average of OA across these sites for each network (see site locations
32 in Figure S4). We apply a seasonal-dependent ratio to convert organic carbon to OA mass: 2.2 in
33 June-July-August, 1.8 in December-January-February and 1.9 in other months (Philip et al., 2014).
34 In Section 3.1, we calculate the absolute trend of a variable as the slope of the regression line of
35 the variable's value versus time, and we calculate the relative trend (represented by "m" in Figure
36 1) as the absolute trend divided by the variable's 2000-2016 average.

37 We also compare modeling results to OA measurement by Aerosol Chemical Speciation Monitor
38 (ACSM). We select 3 European sites from the ACTRIS (the Aerosol, Clouds and Trace Gases
39 Research Infrastructure) network (Crenn et al., 2015): Hyttila (Finland), Puy de Dome (France)
40 and Birkenes II (Norway); two sites from the ARM (Atmospheric Radiation Measurement)
41 network (Uin et al., 2019): Southern Great Plains (US) and Manacapuru, Amazonia (Brazil). These
42 sites are covered by natural vegetation and have more than a year's worth of data available. We
43 average the original hourly OA measurement to monthly mean data for these sites to compare with
44 modeling results.

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3. Results

3.1 Decadal trend of summertime OA in SEUS and its variability

The SEUS is a region heavily influenced by both biogenic and anthropogenic emissions (Mao et al., 2018). In the last two decades, organic aerosol shows a decreasing trend, resulting from reductions in anthropogenic pollutants including SO_2 and NO_x (Marais et al., 2017; Blanchard et al., 2016; Ridley et al., 2018). The CMPX and CMPX_ag scheme successfully reproduce the summertime surface OA concentrations measured from the IMPROVE and SEARCH networks at 4 and 5.5 $\mu\text{g}/\text{m}^3$, respectively (Figure 1a). The Simple scheme has a significant overestimate ($\sim 7 \mu\text{g}/\text{m}^3$).

We first examine the simulated decadal OA trend in the SEUS against filter-based measurements from IMPROVE and SEARCH networks. From 2000 to 2016, the measured summer OA declines by $-0.13 \mu\text{g}/\text{m}^3/\text{year}$ from SEARCH and by $-0.09 \mu\text{g}/\text{m}^3/\text{year}$ from IMPROVE, both at a reduction rate of $-2.3\%/\text{year}$ (Figure 1a). This decreasing trend is well reproduced by the CMPX_ag simulation with a decrease of $-0.11 \mu\text{g}/\text{m}^3$ (-2.0%) per year, and a smaller decrease of $-0.06 \mu\text{g}/\text{m}^3$ (-1.4%) per year, with the CMPX scheme. Considering the varying reduction trends among different sites (Figure S4), both the CMPX and CMPX_ag schemes well reproduce the SEUS OA trend in general. In contrast, the Simple scheme shows a slight increase ($+0.7\%/\text{year}$) in surface OA due to lack of AIBS and little change of POA in 2000–2016 in this region (Figure 1c).

We further examine the summertime month-to-month variability of surface OA. We find that both CMPX_ag and CMPX schemes can well reproduce the low summertime month-to-month variability of surface OA (standard deviation smaller than $2 \mu\text{g}/\text{m}^3$) constrained by IMPROVE and SEARCH measurements (Figure S2), using fixed uptake coefficients ($\gamma=0.001$) of IEPOX and glyoxal. This summertime month-to-month variability was found to be too high (standard deviation up to $5 \mu\text{g}/\text{m}^3$) in the early 2000s using an acidity-dependent IEPOX reactive uptake scheme (Marais et al., 2016, 2017), pointing to additional species besides SO_2 driving the decreasing OA trend.

One unique feature of the CMPX_ag simulation is the dominance of TSOA (Figure 1), mainly through enhanced sensitivity of TSOA production to NO_x . Such dominance of TSOA in this region is also supported by recent field observations (Xu et al., 2018; Zhang et al., 2018a). We find TSOA contributes to 60% of the surface OA trend in the CMPX_ag scheme, mainly through NO_x reduction. The NO_3 -initiated pathway contributes to the majority of surface TSOA decrease (Figure S5), resulting from the rapid decrease of NO_3 (Figure 1d) (Boyd et al., 2017; Rollins et al., 2012). Compared to the CMPX scheme, the dominant contribution of TSOA is largely due to the OH aging effect, which amplifies the SOA yield from all monoterpene oxidation channels. As a result, we find that NO_x reduction accounts for 60% of OA decrease in SEUS. This enhanced sensitivity to NO_x , resonates with recent developments on monoterpene-derived highly oxygenated organic molecules or autoxidation (Pye et al., 2019), highlighting the importance of NO_x in AIBS.

ISOA contributes to 40% of surface OA trend in the CMPX_ag scheme, mainly through SO_2 reduction. The decrease in surface ISOA, at $-0.05 \mu\text{g}/\text{m}^3/\text{year}$, is associated with the strong reduction in sulfate ($-7\%/\text{year}$). The rapidly decreasing sulfate, NO_x and O_3 in the model are

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1 consistent with observations over the SEUS (Figure S3) and previous studies (Zheng et al., 2020;
2 Wells et al., 2021; Simon et al., 2015). In contrast to Marais et al. (2017), we find that this
3 nondominant role of ISOA brings model into much better agreement with observations, especially
4 on the low summertime month-to-month variability of surface OA (standard deviation smaller than
5 $2 \mu\text{g/m}^3$) constrained by IMPROVE and SEARCH measurements (Figure S2) (Zheng et al., 2020).
6 The observed summertime month-to-month variability also implies a weaker dependence of OA
7 to sulfate aerosols in this region than as shown in Marais et al. (2017), highlighting the importance
8 of TSOA.
9

10 We find a similar trend of summer OA column concentration to the surface OA trend in the model.
11 The CMPX_ag simulation suggests a decreasing trend in summer OA column concentration,
12 driven by both TSOA ($-0.13 \text{ mg/m}^2/\text{year}$) and ISOA ($-0.12 \text{ mg/m}^2/\text{year}$) (Figure 1b). Similar to
13 the surface, the aging effect increases the column production of TSOA in CMPX_ag and its
14 sensitivity to changes in NO_x compared with the CMPX scheme.
15

16 **3.2 Present-day OA in vegetated regions and global budget**

17 We further evaluate the modeled surface OA against measurements by Aerosol Chemical
18 Speciation Monitor (ACSM) in other vegetated regions in the Amazon, Europe and US (Figure 2).
19 In the Amazon region, the CMPX_ag scheme successfully reproduces the high surface OA
20 concentration from August to November and low OA in other months (Figure 2c). The Simple
21 scheme greatly overestimates surface OA in all seasons because of its high SOA yield (10%) from
22 isoprene emissions. The CMPX scheme well reproduces the low OA concentrations from January
23 to July but only predicts half of observed OA in months with high OA concentrations. In the
24 European sites from the ACTRIS network (Figure 2d-f), all model simulations underestimate
25 measured OA. One possible reason is uncertainties associated with BVOC emissions and biogenic
26 SOA. Jiang et al. (2019) showed that MEGAN overestimates isoprene emission but underestimates
27 monoterpene emissions in Europe by a factor of 3. At the US Southern Great Plains site, from the
28 ARM network (Figure 2b), the CMPX_ag and CMPX schemes successfully capture the measured
29 OA seasonal variation but underestimate OA magnitude. In the SEUS compared to filter
30 measurements (Figure 2a), all simulations show lower OA in winter than observations, likely due
31 to an underestimate of wintertime emissions of POA (Tsigaridis et al., 2014; Liu et al., 2021). In
32 general, the updated CMPX_ag and CMPX schemes agree well with observations in the Amazon
33 and US where biogenic emissions are high. The good performance of the CMPX_ag scheme in the
34 Amazon, better than the CMPX scheme, gives us confidence that the traditional VBS in the CMPX
35 scheme may underestimate the contribution of TSOA and its sensitivity to NO_x .
36

37 Globally, the SOA burden from the Simple, CMPX and CMPX_ag schemes are 0.99, 0.50 and
38 1.05 Tg, respectively, and their SOA production rates are 82, 40 and 69 Tg/year (Figure 3), in
39 agreement with other global modeling studies. The AeroCom phase II model intercomparison
40 summarizes a median SOA source of 51 Tg/year with a range between 16 to 121 Tg/year
41 (Tsigaridis et al., 2014), although top-down methods indicate SOA source could be up to 50-380
42 Tg/year (Spracklen et al., 2011). Uncertainties associated with BVOC emissions contribute to the
43 wide spread of SOA estimate by global models. In GFDL AM4.1, annual isoprene and
44 monoterpene emissions are computed to be 505 ± 14 and 137 ± 5 Tg/year, respectively (Figure 3), in
45 line with previous estimates (Guenther et al., 2012).
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1 Detailed SOA budgets for the three schemes are summarized in Table 2. The CMPX and
2 CMPX_ag schemes have much less ISOA than the Simple scheme as the latter has high pseudo
3 emission of isoprene SOA, which is 10% in GFDL AM4.1 as compared to 3% used in other models
4 like GEOS-Chem (Pai et al., 2020; Henze and Seinfeld, 2006). ISOA (22.2 Tg/year) and TSOA
5 (14.4 Tg/year) in the CMPX scheme are consistent with previous estimate by GEOS-Chem (Pai et
6 al., 2020; Zheng et al., 2020). The CMPX_ag scheme has higher TSOA (44 Tg/year) than CMPX
7 and Simple due to the aging effect of semivolatile oxidation products from terpenes (Figure 3),
8 and is close to the high end of estimate (12.7-40 Tg/year) by AeroComII (Tsigaridis et al., 2014).
9 ASOA is often neglected by global models despite an estimate of 13.5 Tg/year suggesting ASOA
10 as a non-negligible source (Tsigaridis et al., 2014). In GFDL AM4.1, ASOA (3.3 Tg/year) only
11 considers oxidation of C₄H₁₀, which does not well represent all ASOA and warrants further
12 research.

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14 3.3 Centennial change in biogenic SOA and direct radiative forcing

15 We now extend our analysis of AIBS from the decadal scale to the centennial scale. To represent
16 the higher natural vegetation cover during PI, we scale up isoprene and monoterpene emission
17 basal factor in the PI simulations by 35%, equivalent to a 26% reduction of natural vegetation
18 cover from PI to PD (Unger, 2014). This simple scaling should be considered as an idealized study
19 to avoid uncertainties associated with historical vegetation reconstruction and the complex role of
20 CO₂ including both fertilization and inhibition effects. From 1870s to 2000s, the simulated
21 isoprene emissions decrease from 632±15 to 505±14 Tg/year (-20%) and monoterpene emissions
22 decrease from 161±5 to 137±5 Tg/year (-15%) (Figure 3a, maps in Figure S6), consistent with
23 previous studies (Heald and Spracklen, 2015).

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24 Despite the reduction in BVOC emissions from PI to PD, we show a significant increase of
25 biogenic SOA (Figure 3b, maps in Figure S7 and S8), resulting from increase in anthropogenic
26 emissions and amplified by AIBS. With an increase by 1.4, 7, and 4 for emissions of POA, SO₂
27 and NO_x, total SOA production increases by 36% and its burden increases by 42% (in the
28 CMPX_ag scheme). ASOA, ISOA and TSOA contribute 17%, 62%, and 21% to the changes in
29 total SOA production, respectively. In contrast, the Simple scheme shows a decrease of SOA
30 production following the reduction in BVOC emissions. The large increase of SOA from PI to PD
31 differs from previous estimates (Spracklen et al., 2011; Heald and Spracklen, 2015; Zhu et al.,
32 2019; Scott et al., 2017; Lin et al., 2014b; Heald and Geddes, 2016; Hoyle et al., 2009), largely
33 due to AIBS constrained by observations.

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35 The total PI-to-PD SOA rise is largely dominated by ISOA (62%), resulting from the strong
36 increase in anthropogenic SO₂ emissions and uptake of IEPOX and glyoxal onto sulfate aerosols.
37 The global burden of sulfate aerosol has doubled from 0.7 Tg at PI to 1.6 Tg at PD, with large
38 increase over the tropics, SEUS, and Eurasia (Figure S9). The increase in TSOA is due to both
39 increased NO_x emissions and POA emissions. In contrast to the decadal trend where β_{NO} barely
40 changes, the PI-to-PD increase of TSOA due to the change of NO_x is suppressed by the shift of
41 β_{NO} . The branching ratio β_{NO} increases from a global average of 0.32 at PI to 0.61 at PD (Figure
42 S10), indicating a shift from low-NO_x pathway (higher yields) to high-NO_x pathway (lower yields)
43 for the OH- and O₃-initiated oxidation. These competing effects lead to a net +10% change in
44 TSOA production and a +14% increase in burden from PI to PD. The PI-to-PD change in TSOA
45 in the CMPX scheme is small (-7% in production and +6% in burden). Increased POA provides

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1 more organic mass for monoterpene oxidation products to condense on, especially in central Africa
2 and central South America (Figure S9).

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4 The large increase of biogenic SOA leads to a cooling direct radiative forcing (DRF) from PI to
5 PD, opposed to the warming suggested by the Simple scheme. DRF is usually defined as the
6 difference between PI and PD direct radiative fluxes at top-of-atmosphere under all-sky conditions.
7 We show in Figure 4, the global instantaneous DRF at top-of-atmosphere of -(26–44) mW/m²,
8 comparable to that of POA (-98 mW/m²). In contrast, the Simple scheme shows a warming DRF
9 of +17 mW/m², largely due to lack of AIBS. The DRF of SOA in the updated schemes resides
10 within reported AeroComII estimates, which ranges from -210 to -10 mW/m², with a mean value
11 of -60 mW/m² and a median value of -20 mW/m² (Myhre et al., 2013). Due to this increase of
12 SOA burden, our results may also imply a large indirect radiative forcing from biogenic SOA that
13 is missing from previous work (Carslaw et al., 2013).

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14 4. Summary

15 Our work suggests a strong coupling between anthropogenic and biogenic emissions in biogenic
16 SOA production. Constrained by observations in SEUS, we show that the summertime OA
17 decreasing trend is likely driven by reduction in both NO_x and SO₂ emissions, through TSOA and
18 ISOA. First, in a previous study (Zheng et al., 2020) we prove that the scheme of acidity-catalyzed
19 aqueous ISOA formation (Marais et al., 2016) strongly overestimates summertime month-to-
20 month variability of surface OA, therefore in this study we use fixed uptake coefficients for
21 isoprene oxidation products to avoid uncertainties associated with acidity, relative humidity, and
22 coating effect. Second, both the CMPX and CMPX_ag schemes reproduce the observed OA
23 magnitude and the decadal trend in SEUS, in which SO₂ alone cannot explain this trend. The
24 CMPX_ag scheme shows a faster OA decrease and better agrees with long-term filter
25 measurement, which is largely driven by NO_x (60%). Third, the CMPX_ag scheme successfully
26 reproduces the observed OA magnitude and seasonal cycle in Amazon, outperforming the CMPX
27 and Simple schemes. Our results point to the importance role of NO_x on modulating biogenic SOA,
28 in line with recent understanding on autoxidation (Crounse et al., 2013; Ehn et al., 2014; Pye et
29 al., 2019), although further studies are warranted. For example, the CMPX_ag scheme with a
30 simplified aging parameterization does not mechanistically represent the most up-to-date
31 understanding of HOMs and organic nitrates (Takeuchi and Ng, 2019; Berkemeier et al., 2020;
32 Pullinen et al., 2020; Yu et al., 2021). Long-term measurement of ISOA and TSOA tracers across
33 different regions and seasons are needed to develop future mechanistic SOA schemes that are suit
34 for global climate models with minimal computational cost. In this study, the success of the
35 updated schemes in capturing the observed OA trend and month-to-month variability provides
36 confidence in model simulations over longer time scales.

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39 At centennial scale, atmospheric SOA mass increases significantly from PI to PD despite
40 reductions in BVOC emissions, posing a top-of atmosphere instantaneous radiative forcing of -
41 (26–44) mW/m². ISOA dominates the total SOA change as a result of a significant rise in global
42 sulfate aerosol from PI to PD, especially in the fast-developing regions like Africa, Middle East,
43 India, and China. POA increases greatly in central Africa and central South America as well as
44 India and east China, which enhances TSOA production. The significant increase in SOA due to
45 AIBS in these regions poses new challenges to meet the World Health Organization's
46 recommendation on annual fine particulate matter exposure (5 µg/m³) (Pai et al., 2022). Under

1 future scenarios with reduced emissions of SO_2 , NO_x and POA, the AIBS may indicate larger
2 reductions in SOA than current model predictions, but their relative importance cannot be linearly
3 extrapolated based on PI and PD simulations. Model simulations with future emission scenarios
4 are needed, which is beyond the scope of this study.

5
6 The updated SOA scheme in GFDL AM4.1 shows an advance in representing vegetation-
7 chemistry-climate interactions than the default model which assumes fixed yields of SOA from
8 biogenic hydrocarbons, although a variety of uncertainties still exist in the evaluation of SOA and
9 its climate impact. First, the model likely underestimates wintertime POA in the US, total OA in
10 Europe and anthropogenic SOA globally. Second, the model does not consider absorbing SOA or
11 brown carbon which could form from biomass burning and aging (Tsigaridis and Kanakidou,
12 2018). The model applies the same optical parameters for all SOA as hydrophilic POA. Third,
13 other than the uncertainties discussed above about acidity-dependency, formation of HOMs and
14 organic nitrates, other properties that influence the multiphase growth of SOA, including coating
15 and viscosity, are also not implemented in our model (Shrivastava et al., 2017). Finally, the model
16 does not consider nucleation of extremely low volatile compounds from BVOC oxidation, which
17 may increase SOA in pristine environments in the pre-industrial period, thus reducing the PI-to-
18 PD radiative forcing of SOA (Gordon et al., 2016; Zhu et al., 2019). These uncertainties warrant
19 further research in studies on anthropogenic-influenced SOA in climate models.
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Deleted: Second, we use a simplified aging parameter to represent the enhanced SOA sensitivity of NO_x . This should be improved by more explicit schemes on autoxidation (Pye et al., 2019) and organic nitrates(Takeuchi and Ng, 2019).

1 **Acknowledgements**
2 We acknowledge funding from NOAA grant NA18OAR4310114 and NASA grant
3 80NSSC21K0428. We thank Fabien Paulot and Songmiao Fan for internal GFDL review and
4 support from GFDL's Model Development Team, Modeling Systems Division, Operations group,
5 and the RDHPCS supercomputing resources. We also acknowledge the Electric Power Research
6 Institute (EPRI) and Southern Company for support of the SEARCH network and Atmospheric
7 Research & Analysis, Inc; the US Environmental Protection Agency (EPA) for support of the
8 IMPROVE network and Air Quality System; the European Union's Horizon 2020 research and
9 innovation programme under grant agreement No 654109 for support of the ACTRIS network;
10 and the Atmospheric Radiation Measurement (ARM) user facility, a U.S. Department of Energy
11 (DOE) office of science user facility managed by the Biological and Environmental Research
12 Program.

13

14 **Author Contributions**

15 Conceptualization: YZ, JM

16 Methodology: YZ, LWH, RM, DJP, VN, JM

17 Investigation: YZ, LWH, JM

18 Writing—original draft: YZ, JM

19 Writing—review & editing: YZ, LWH, RM, DJP, VN, JL, JM

20

21 **Data availability**

22 The IMPROVE filter OA and sulfate data is available at <http://views.cira.colostate.edu/iwdw/>. The
23 ACTRIS ACSM OA data is available at <https://actris.nilu.no/>. The ARM ACSM OA data is
24 available at <https://www.arm.gov/data/>. The EPA's AQS data is available at <https://aqs.epa.gov>.
25 Model outputs are available at <https://doi.org/10.6084/m9.figshare.21493986.v1>.

26

27 **Competing interests**

28 The authors declare no competing interests.

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