



Errors associated with calculating the gross nitrification rates in forested catchments using the triple oxygen isotopic composition (Δ^{17} O) of stream nitrate

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Abstract

2 A novel method for quantifying the gross nitrification rate (GNR) in each forested catchment using the triple oxygen isotopic composition (Δ^{17} O) of stream nitrate eluted 3 from the catchment has been proposed and applied in several recent studies. However, 4 the equations used in the calculations include the approximation that the Δ^{17} O value 5 of nitrate metabolized through either assimilation or denitrification within the forested 6 soil is equal to the Δ^{17} O value of nitrate in the stream. The GNR estimated from the 7 Δ^{17} O value of stream nitrate was more than six times the actual GNR in our simulated 8 calculation for a forested catchment where the nitrate in the soil exhibited Δ^{17} O values 9 larger than those in the stream while showing a decreasing trend with increasing 10 11 depths until that of stream nitrate at the bottom. As most of the reported soil nitrate in forested catchments from past studies showed Δ^{17} O values higher than those of the 12 stream nitrate eluted from each catchment, we concluded that the GNR estimated 13 from the Δ^{17} O value of stream nitrate in the forested catchments was, to some extent, 14 15 an overestimate of the actual GNR. 16 17 1 Introduction 18 Nitrate (NO₃⁻) is one of the important nitrogen nutrients for primary production in 19 forested ecosystems. Nitrification is the microbial process that produces NO₃⁻ in each 20 forested ecosystem. Thus, quantifying the nitrification rate can assist in the evaluation 21 of the present and future states of each forest ecosystem. While the net nitrification rate can be estimated from the increase in NO₃⁻ concentration during a certain period, 22





the gross nitrification rate (GNR), which includes the net nitrification rate and the 23 metabolic rate of nitrate (e.g., assimilated by plants or decomposed through 24 denitrification), reflects the internal N cycling better than the net nitrification rate 25 (Bengtsson et al., 2003), especially in forested ecosystems, where the net nitrification 26 27 rate is negligible (Stark and Hart, 1997) while the metabolic rate is significant so that the GNR often exceeds the net nitrification rate by order of magnitude (Verchot et al., 28 29 2001). Recently, several studies have successfully estimated GNR in water environments 30 such as lakes, using the Δ^{17} O values of NO₃⁻ as a conserved tracer of the mixing ratio 31 between the atmospheric nitrate (NO₃⁻atm) deposited into the water environment and 32 33 the remineralized nitrate (NO₃ re) produced through nitrification therein (Tsunogai et al., 2011, 2018). Although NO₃ re always shows the Δ^{17} O values close to 0 ‰ because 34 its oxygen atoms derive from either terrestrial O2 or H2O through nitrification, 35 NO_{3-atm}^{-} always displays an anomalous enrichment in ¹⁷O with Δ^{17} O values being 36 approximately $+26 \pm 3$ ‰ in Japan (Tsunogai et al., 2010, 2016) because of oxygen 37 transfers from atmospheric ozone (Michalski et al., 2003; Nelson et al., 2018). 38 Additionally, Δ^{17} O is almost stable during "mass-dependent" isotope fractionation 39 processes (Michalski et al., 2004; Tsunogai et al., 2016). Therefore, regardless of the 40 partial metabolism through denitrification or assimilation after deposition in a water 41 column, Δ^{17} O can be used as a conserved tracer of NO₃ atm to calculate the mixing 42





- ratio of NO₃⁻_{atm} to total NO₃⁻ (NO₃⁻_{atm}/NO₃⁻_{total}) in a water column using the
- 44 following equation:

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$$[NO_3^-]/[NO_3^-] = [NO_3^-]/([NO_3^-] + [NO_3^-]) = \Delta^{17}O/\Delta^{17}O_{atm}$$
 (1)

- where $\Delta^{17}O_{atm}$ and $\Delta^{17}O$ denote the $\Delta^{17}O$ values of NO_{3-atm}^{-} and NO_{3}^{-} dissolved in
- each water environment, respectively. Using both the NO₃⁻_{atm}/NO₃⁻_{total} ratio estimated
- 48 from the Δ^{17} O value of NO₃⁻ in a lake water column and the deposition rate of
- 49 NO_{3 atm} into the lake, the GNR has been successfully estimated (Tsunogai et al., 2011,
- 50 2018).
- In addition to application in water environments, the Δ^{17} O method has also been
- 52 applied to forested catchments for GNR determination (Fang et al., 2015; Hattori et
- al., 2019; Huang et al., 2020). By using the deposition flux of NO₃ atm into the
- catchment as well as the elution flux of both unprocessed NO₃ atm and NO₃ re from the
- stream, which can be determined from the Δ^{17} O values of NO₃⁻ in stream water eluted
- 56 from the catchment, the GNR in each forested catchment has been estimated in a
- 57 manner similar to the estimation for the water environments (Fang et al., 2015).
- Contrary to water environments, where the Δ^{17} O values of NO₃⁻ within the water
- 59 layers are homogeneous and can be measured easily, it is often difficult to determine
- the Δ^{17} O values of the NO₃⁻ metabolized in soil layers. Consequently, past studies
- approximated the values to be equal to the Δ^{17} O value of stream NO₃⁻ eluted from
- each forested catchment without actual observation (Fang et al., 2015, Hattori et al.,
- 63 2019, Huang et al., 2020). However, such an approximation should be conducted with





- extreme caution, as the Δ^{17} O values of soil NO₃⁻ are not always equal to those of the
- 65 stream (Hattori et al., 2019, Rose, 2014, Osaka et al., 2010). To clarify the details of
- the approximation along with its impact on the final estimated GNR, we present an
- accurate relationship between the Δ^{17} O of soil NO₃⁻ and GNR, starting from the basic
- 68 isotope mass balance equations. Then, we present the GNR estimated for a forested
- catchment in which the Δ^{17} O values of NO₃⁻ in soil are measured. Finally, we
- compare the GNR estimated in this study with the GNR estimated from the Δ^{17} O
- values of stream NO₃⁻.

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2 Calculation

- 74 The total mass balance equation of NO₃⁻ including the GNR in each catchment can
- 75 be expressed as follows:

$$76 NO_{3-\text{deposition}} + GNR = NO_{3-\text{leaching}} + NO_{3-\text{uptake}} + GDR (2)$$

- 77 where NO₃ deposition, GNR, NO₃ leaching, NO₃ uptake, and GDR denote the deposition flux
- 78 of NO₃⁻ into each catchment, gross nitrification rate in each catchment, leaching flux
- 79 of NO₃⁻ from each catchment, uptake rate of NO₃⁻ in each catchment, and gross
- 80 denitrification rate in each catchment, respectively.
- The isotope mass balance for each Δ^{17} O value of NO₃⁻ in the catchment can also be
- 82 calculated using the same method:

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$$NO_3^-_{deposition} \times \Delta^{17}O(NO_3^-)_{atm} + GNR \times \Delta^{17}O(NO_3^-)_{nitrification} = NO_3^-_{leaching} \times \Delta^{17}O(NO_3^-)_{nitrification}$$

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$$_3$$
)_{stream} + NO_3 ⁻_{uptake} × Δ ¹⁷ $O(NO_3$ ⁻)_{uptake} + $GDR \times \Delta$ ¹⁷ $O(NO_3$ ⁻)_{denitrification} (3)





- where $\Delta^{17}O(NO_3^-)_{atm}$, $\Delta^{17}O(NO_3^-)_{nitrification}$, $\Delta^{17}O(NO_3^-)_{stream}$, $\Delta^{17}O(NO_3^-)_{uptake}$, and
- 86 $\Delta^{17}O(NO_3^-)_{denitrification}$ denote the $\Delta^{17}O$ value of NO_3^- atm deposited into each
- catchment, that of NO₃ re produced through nitrification, that of NO₃ eluted from
- each catchment, that of NO₃⁻ assimilated by plants and other organisms in each
- 89 catchment, and that of NO₃⁻ decomposed through denitrification in each catchment,
- 90 respectively.
- If the Δ^{17} O values of NO₃⁻ in the forested soil layers, where the NO₃⁻ was
- metabolized through either assimilation (by plants and other organisms) or
- denitrification, are equal to the Δ^{17} O value of NO₃⁻ in the stream, Eq. 4 can be
- 94 expressed as follows:

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$$\Delta^{17}O(NO_3^-)_{uptake} = \Delta^{17}O(NO_3^-)_{denitrification} = \Delta^{17}O(NO_3^-)_{stream}$$
 (4)

- 96 Consequently, by combining Eqs. 3 and 4, we could obtain the following
- 97 relationship:

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$$NO_{3-deposition} \times \Delta^{17}O(NO_{3-deposition} + GNR \times \Delta^{17}O(NO_{3-deposition} = (NO_{3-leaching} + NO_{3-uptak})$$

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$$_{\rm e} + \rm GDR) \times \Delta^{17}O(NO_3^-)_{\rm stream}$$
 (5)

- We could estimate the GNR using Eq. 6 obtained from Eqs. 2 and 5 because we can
- approximate the Δ^{17} O values of NO₃ re produced through nitrification
- $(\Delta^{17}O(NO_3^-)_{nitrification})$ to be 0 (Michalski et al., 2003; Tsunogai et al., 2010):

103 GNR =
$$NO_{3 \text{ deposition}}^{3} \times (\Delta^{17}O(NO_{3}^{-})_{atm} - \Delta^{17}O(NO_{3}^{-})_{stream})/\Delta^{17}O(NO_{3}^{-})_{stream}$$
 (6)





Eq. 6 corresponds to the equation used in previous studies for quantifying the GNR in each forested catchment (Eq. 4 in Fang et al., 2015; Eq. 8 in Hattori et al., 2019; Eq. 4 Huang et al., 2020).

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3 Results and Discussion

The Δ^{17} O values of NO₃⁻ in forested soil layers should be equal to those of NO₃⁻ in the stream, as presented in Eq. 4 to obtain Eq. 6. While the number of simultaneous observations of the oxygen isotopes of NO₃⁻ in both the soil and stream in a given forested catchment is limited (Hattori et al., 2019, Osaka et al., 2010, Rose, 2014), the limited observations show that the oxygen isotopic ratios of soil NO₃⁻ are mostly higher than those of stream NO₃⁻. For example, Hattori et al. (2019) reported that more than 60 % of the soil exhibited Δ^{17} O values significantly higher than those of stream NO_3^- determined simultaneously ($\Delta^{17}O = +1$ to +3 %). In addition, they found a decreasing Δ^{17} O trend in soil NO₃ with depth, declining from greater than +20 \% at the surface to less than +3 ‰ at depths of 25–90 cm from the surface. A similar decreasing trend in the vertical distribution had been found in $\delta^{18}O$ in another forested catchment, from greater than +35 \% at the surface soil to less than +10 \% at depths of 30–50 cm from the soil surface (Osaka et al., 2010). Besides, most of the soil NO₃⁻ also exhibited δ^{18} O values higher than those in the stream (Osaka et al., 2010). Furthermore, Rose (2014) monitored the horizontal distribution of Δ^{17} O of soil NO₃⁻ by randomly setting 15 tension-free lysimeters at depths of 0–10 cm in a 39 ha





forested catchment, reporting Δ^{17} O values significantly higher in soil NO₃⁻ (+9.1 ± 125 5.8 ‰ on average) than in the stream NO₃⁻ (+0.5 ‰ on average) eluted from the 126 127 forested catchment. As most of the root biomass is concentrated in the top 10 cm of the soil in forested catchments (Jackson et al., 1996), most uptake reactions should 128 occur in that top 10 cm of soil. Consequently, the significant difference in the Δ^{17} O 129 values between soil NO₃⁻ and stream NO₃⁻, particularly in the surface soil layers, 130 imply that the estimated GNRs in the forested catchment obtained from Eq. 6 were 131 132 inaccurate. To demonstrate the impact of the differences in Δ^{17} O between soil NO₃⁻ and stream 133 NO₃⁻ on the GNR, along with presenting the problems associated with the 134 135 approximation to obtain Eq. 6, we estimated the GNR for two simulated forested soils—one with NO₃⁻ showing a decreasing trend in Δ^{17} O down to the Δ^{17} O of stream 136 137 NO₃⁻ (heterogeneous soil) (Fig. 1a and 1b) and one with NO₃⁻ showing the same Δ^{17} O values as those of stream NO₃⁻ (homogeneous soil) (Fig. 2a and 2b). With 138 Hattori et al. (2019) reporting the NO₃⁻deposition as 7.0 kg of N ha⁻¹ y⁻¹, NO₃⁻leaching as 139 2.6 kg of N ha⁻¹ y⁻¹, $\Delta^{17}O(NO_3^-)_{atm}$ as +28.0 %, and $\Delta^{17}O(NO_3^-)_{stream}$ as +2.2 % in 140 their forested catchment study, we adopted the same values in our calculation. 141 142 We divided the soils in the heterogeneous forest soils into 10 layers in the vertical direction, simulating the soils observed by Hattori et al. (2019), in which the Δ^{17} O 143 values of NO₃⁻ gradually decreased with increasing depth, varying from +28.0 to 144 +2.2 ‰ with a rate of decrease of +2.58 ‰ for each step (Fig. 1b). Similarly, we 145





assumed a gradual decrease with increasing depth in the leaching flux of NO₃⁻, i.e., 146 from 7 to 2.6 kg of N ha⁻¹ y⁻¹ with a rate of decrease of 0.44 kg of N ha⁻¹ y⁻¹ for each 147 step (Fig. 1c). In the homogeneous forest soils, we also divided the forested soils into 148 10 layers in the vertical direction. The change with depth in the leaching flux of NO₃⁻ 149 was the same as that in the heterogeneous soils (Fig. 2c), whereas the Δ^{17} O values of 150 NO₃⁻ were constant at +2.2 ‰ in the soil layers (Fig. 2b). 151 152 Applying the total mass balance and isotope mass balance of NO₃⁻ shown in Eqs. 2 153 and 3 to each layer, we estimated both the GNR (Figs. 1e and 2e) and total metabolic rate of NO₃⁻ (GDR + uptake) (Figs. 1d and 2d) in each layer assuming the following: 154 (1) Δ^{17} O values of NO₃⁻ are constant in each layer, (2) vertical flow of NO₃⁻ in the 155 156 soil layers proceed downward from the surface to the water layer with a uniform residence time in each layer, and (3) the GNR and metabolic rate of NO₃⁻ (GDR + 157 uptake) is zero in the water layer (layers beyond the no. 10 soil layer). Then, by 158 integrating the GNR determined for each layer, we estimated the total GNR in each 159 forested catchment. 160 Although the GNR estimated for the catchment with the homogeneous Δ^{17} O values 161 in soil NO₃⁻ was 83.6 kg of N ha⁻¹ y⁻¹, exactly equal to that estimated by Hattori et al. 162 (2019) using Eq. 6 (Fig. 2e), the total GNR was a much smaller 13.0 kg of N ha⁻¹ y⁻¹, 163 simulated for the catchment with the heterogeneous Δ^{17} O values in soil NO₃⁻ (Fig. 164 1e). Consequently, we conclude the following: (1) past studies estimating the GNR 165 using Eq. 6 approximated the Δ^{17} O value of soil NO₃⁻ was homogeneous and always 166





equal to that of stream NO_3^- mathematically and (2) the differences between the $\Delta^{17}O$ 167 values of the soil NO₃⁻ metabolized in a forested catchment and that of stream NO₃⁻ 168 resulted in a significant deviation in the GNR estimated using Eq. 6 from the actual 169 GNR. 170 Note that the linear variation in the leaching flux and Δ^{17} O values of soil NO₃⁻ used 171 in the simulated calculations (Fig. 1) is just one of many possible variations in the 172 forested catchments. It is impossible to decide whether the linear variation was 173 realistic until the downward water flux, along with the concentration and Δ^{17} O values 174 of NO₃⁻, is determined for each soil layer. However, the simultaneous observations of 175 the oxygen isotopes of soil NO₃⁻ and stream NO₃⁻ (Hattori et al., 2019; Osaka et al., 176 177 2010; Nakagawa et al., 2018; Rose, 2014) imply that the approximation of the Δ^{17} O values of the soil NO₃⁻ metabolized through assimilation or denitrification to be 178 always equal to the Δ^{17} O value of stream NO₃⁻, shown in Fig. 2b, is unrealistic. 179 By combining the mass balance and isotope mass balance shown in Eqs. 2 and 3, 180 Eq. 7 can be obtained to accurately estimate the GNR: 181 $GNR = NO_3^{-}_{leaching} - NO_3^{-}_{deposition} + (NO_3^{-}_{deposition} \times \Delta^{17}O(NO_3^{-})_{atm} -$ 182 $NO_3^{-}_{leaching} \times \Delta^{17}O(NO_3^{-})_{stream}) / \Delta^{17}O(NO_3^{-})_{soil}$ (7) 183 where $\Delta^{17}O(NO_3^-)_{soil}$ denotes the $\Delta^{17}O$ values of NO_3^- in forested soil, from which 184 the NO₃⁻ was metabolized through either assimilation or denitrification. As most of 185 the soil NO_3^- measured to date exhibit $\Delta^{17}O$ values higher than those of the stream 186 NO₃⁻ eluted from each catchment (Hattori et al., 2019, Rose, 2014), the GNR 187





estimated from stream NO₃⁻ using Eq. 6 is higher than the GNR estimated from soil 188 NO₃⁻ using Eq. 7, to some extent. In other words, the GNR estimated from Eq. 6 189 overestimated the GNR in each forested catchment to some extent. 190 191 If we estimated the downward water flux at each soil layer, together with the NO₃⁻ concentration and Δ¹⁷O value of NO₃⁻ in each soil layer using a tension-free lysimeter 192 (Inoue et al., 2021), we could estimate the vertical change in the leaching flux of 193 NO_3^- for each soil layer along with the $\Delta^{17}O$ value of soil NO_3^- in each layer. Then, 194 applying Eq. (7) in each layer, we can more accurately estimate the GNR for the 195 forested catchment by integrating the GNR estimated for each soil layer together with 196 a more accurate metabolic rate of NO₃⁻ (GDR + uptake) of the forested catchment. 197 198 However, without such an observation of the distribution of the Δ^{17} O value of NO₃⁻, it is difficult to assume that the Δ^{17} O values of soil NO₃⁻ are always equal to those of 199 stream NO₃⁻; thus, the GNR should be reported with errors in which the possible 200 variations in the Δ^{17} O values of soil NO₃⁻ are considered. 201 202 4 Conclusion 203 Past studies have proposed the Δ^{17} O method to determine the GNR in each forested 204 catchment. The equations used in the calculation presuppose that the Δ^{17} O values of 205 NO₃⁻ in forested soils are homogeneous and equal to those of NO₃⁻ in the stream; 206 however, in reality, the values are often heterogeneous and do not always equal to 207 those corresponding to the stream. It is essential to clarify/verify the Δ^{17} O values of 208





 NO_3^- in the forested soils and stream before applying the stream $NO_3^- \Delta^{17}O$ values to 209 estimate the GNR. 210 211 212 Data availability. All data are presented in the Supplement. 213 Author contributions. WD, UT, and FN designed the study. WD and UT performed 214 data analysis and wrote the paper. 215 216 217 Competing interests. The authors declare that they have no conflict of interest. 218 219 Acknowledgments We thank two anonymous referees, Dr. Joel Bostic, and Dr. Lucy Rose, for their 220 valuable remarks on an earlier version of this paper. We are grateful to the members 221 222 of the Biogeochemistry Group, Nagoya University, for their valuable support throughout this study. This work was supported by a Grant-in-Aid for Scientific 223 Research from the Ministry of Education, Culture, Sports, Science, and Technology of 224 Japan under grant numbers 22H00561, 17H00780, 22K19846, the Grant-in-Aid for 225 JSPS Fellows under grant number 23KJ1088, the Yanmar Environmental 226 Sustainability Support Association, and the river fund of the river foundation, Japan. 227 Weitian Ding would like to take this opportunity to thank the "Nagoya University 228 Interdisciplinary Frontier Fellowship" supported by Nagoya University and JST, the 229





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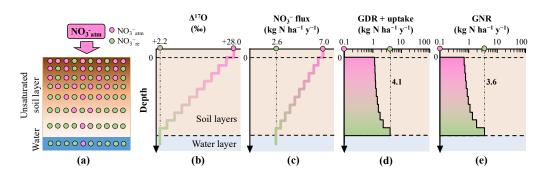


Figure. 1. Distribution of NO_3^- atm in the simulated forested soil where the distribution of the $\Delta^{17}O$ values of NO_3^- is heterogeneous (a). Vertical distribution of the following parameters in the forested soil: the simulated $\Delta^{17}O$ values of NO_3^- (b), simulated leaching flux of NO_3^- (c), estimated NO_3^- consumption rate (GDR + uptake) (d), and estimated GNR (e).

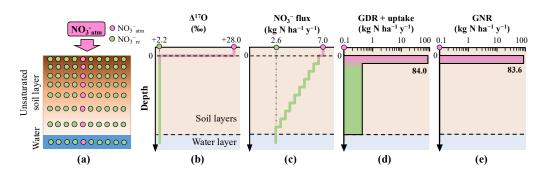


Figure. 2. Distribution of NO₃⁻ atm in the simulated forested soil where the distribution of the Δ^{17} O values of NO₃⁻ is homogeneous (a). Vertical distribution of the following parameters in the forested soil: the simulated Δ^{17} O values of NO₃⁻ (b), simulated leaching flux of NO₃⁻ (c), estimated NO₃⁻ consumption rate (GDR + uptake) (d), and estimated GNR (e).