

Formation and Loss of Light Absorbance by Phenolic Aqueous SOA by $\bullet\text{OH}$ and an Organic Triplet Excited State

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Abstract. Brown carbon (BrC) is an important component of biomass burning (BB) emissions that impacts Earth's radiation budget. BB directly emits primary BrC as well as gaseous phenolic compounds (ArOH), which react in the gas and aqueous phases with oxidants - such as hydroxyl radical ($\bullet\text{OH}$) and organic triplet excited states ($^3\text{C}^*$) - to form light-absorbing secondary organic aerosol (SOA). These reactions in atmospheric aqueous phases, such as cloud/fog drops and aerosol liquid water (ALW), form aqueous SOA (aqSOA), i.e., low-volatility, high molecular weight products. While these are important routes of aqSOA formation, the light absorption and lifetimes of the BrC formed are poorly characterized. To study these aspects, we monitored the formation and loss of light absorption by aqSOA produced by reactions of six highly substituted phenols with $\bullet\text{OH}$ and $^3\text{C}^*$. While the parent phenols absorb very little tropospheric sunlight, they are oxidized to aqSOA that can absorb significant amounts of sunlight. The extent of light absorption by the aqSOA depends on both the ArOH precursor and oxidant: more light-absorbing aqSOA is formed from more highly substituted phenols and from triplet reactions rather than $\bullet\text{OH}$. Under laboratory conditions, ~~continued reactions with $\bullet\text{OH}$~~ extended reaction times in $\bullet\text{OH}$ experiments diminish sunlight absorption by aqSOA on timescales of hours, while ~~continued reactions with~~ extended reaction times in $^3\text{C}^*$ experiments reduce light absorption much more slowly. Estimated lifetimes of light-absorbing phenolic aqSOA range from 3 to ~~43~~17 hours in cloud/fog drops, where $\bullet\text{OH}$ is the major sink, and from 0.~~57~~ to 8 hours in ALW, where triplet excited states are the major sink.

30 **1 Introduction**

Organic aerosols (OA) account for 20 to 90% of airborne particulate mass (Jimenez et al., 2009; Kanakidou et al., 2005), with significant effects on climate (Heald et al., 2008; Shrivastava et al., 2017) and human health (Aguilera et al., 2021; Kim et al., 2015; Zhou et al., 2021). Types of OA include primary OA (POA), which is directly emitted from sources such as biomass burning and fossil fuel combustion (Andreae, 2019; Bond et al., 2004), and secondary OA (SOA), which is formed from atmospheric aging in the gas or condensed phases (Graber and Rudich, 2006; Jimenez et al., 2009; Powelson et al., 2014; Volkamer et al., 2006). SOA usually accounts for a dominant fraction of the OA budget (Hallquist et al., 2009; Robinson et al., 2007). Biomass burning (BB) is a significant source of POA as well as reactive organic gases that form SOA (Akagi et al., 2011; Andreae, 2019; Bond et al., 2004; Hallquist et al., 2009), which together affect air quality, atmospheric composition, and climate (Garofalo et al., 2019; Kleinman et al., 2020; O'Dell et al., 2021).

40 While most OA scatters sunlight, brown carbon (BrC, i.e., light-absorbing organic carbon) absorbs sunlight and possibly causes warming (Jo et al., 2016; Reid et al., 2005; Zhang et al., 2020). Compared to black carbon, which strongly absorbs sunlight across the UV-vis spectrum, light absorption by BrC is strongly wavelength dependent, increasing from the visible to the UV (Hecobian et al., 2010; Laskin et al., 2015). BB is an important source of BrC that strongly absorbs light (Di Lorenzo et al., 2017; Kirchstetter and Thatcher, 2012; Palm et al., 2020; Zhong and Jang, 2014). POA and SOA formed from smog chamber experiments of biomass burning emissions show evidence of light absorption by BrC, especially in the near-UV region (Saleh et al., 2013). The absorption of BrC is affected by atmospheric oxidation, exposure to sunlight, and pH, as well as brown carbon source (Grieshop et al., 2009; Hennigan et al., 2011; Liu et al., 2021; Ortega et al., 2013). Also, atmospheric aging can both increase light absorption by forming chromophores but also photo-bleach chromophoric molecules (Fleming et al., 2020; Hems et al., 2020, 2021; Schnitzler et al., 2022; Zhao et al., 2015; Zhong and Jang, 2014).

One important class of reactive organic gases potentially important for secondary BrC formation are phenols, which we abbreviate as ArOH (Andreae, 2019; Chang and Thompson, 2010). ArOH are abundant emissions from BB, with high emissions of simple phenols and smaller amounts of highly substituted ArOH (Andreae, 2019; Schauer et al., 2001). ArOH typically have strong absorbance peaks under 300 nm that fall exponentially at longer wavelengths (Kaeswurm et al., 2021), although phenols with carbonyl or nitro substituents can absorb sunlight (Smith et al., 2016; Wang et al., 2022). Phenols rapidly undergo photochemical transformations both in the gas and aqueous phases with various oxidants to produce low volatility compounds (Berndt and Böge, 2003; Gurol and Nekouinaini, 1984; Ma et al., 2021; Sun et al., 2010; Yee et al., 2013; Yu et al., 2014). ArOH from BB have a wide range of Henry's law constants (K_H). Highly substituted ArOH with modest to high K_H values ($10^3 - 10^8 \text{ M atm}^{-1}$) will have a higher tendency to partition into the aqueous phase (McFall et al., 2020), where they can be oxidized to form large, low volatility products (Smith et al., 2016; Yu et al., 2016).

65 Atmospheric aqueous oxidants that rapidly oxidize ArOH during the daytime include the hydroxyl radical ($\bullet\text{OH}$) and organic triplet excited states of brown carbon ($^3\text{C}^*$) (Smith et al., 2015). Oxidation reactions of both simple and highly substituted phenols react rapidly with $\bullet\text{OH}$ and $^3\text{C}^*$ are fast, with second-order rate constants of roughly $10^9 \text{ M}^{-1} \text{ s}^{-1}$ and $10^8 \text{ M}^{-1} \text{ s}^{-1}$ at pH 5, respectively (Arciva et al., 2022; Buxton et al., 1988; Ma et al., 2021; Smith et al., 2015). Additionally, both the $\bullet\text{OH}$ and $^3\text{C}^*$ reactions efficiently form aqSOA, with average ($\pm 1 \sigma$) mass yields of $(82 \pm 12)\%$ and $(83 \pm 14)\%$, respectively (Arciva et al., 2022; Ma et al., 2021). SOA mass yields for ArOH oxidation in the aqueous phase are often higher than the parallel gas-phase reactions (Arciva et al., 2022; Berndt and Böge, 2003; Coeur-Tourneur et al., 2010; Ma et al., 2021), a result of more efficient oligomerization and functionalization reactions unique to in the aqueous phase (Yu et al., 2016), as opposed to the more common fragmentation in the gas phase. Additionally, aqueous reactions produce oligomers, i.e., low-volatility, conjugated products that can significantly enhance light absorption (Chang and Thompson, 2010; Huang et al., 2018; Li et al., 2014; Yu et al., 2016).

75 While ArOH the processing of ArOH is a significant source of light-absorbing aqSOA, little is known about how phenol-derived brown carbon is photobleached with continued reaction. In part, this likely depends on which oxidant produced the aqSOA, as reaction products can be quite different. For example, initial oxidation of ArOH by $\bullet\text{OH}$ and $^3\text{C}^*$ forms aqSOA that absorbs light above 300 nm (Jiang et al., 2023b; Li et al., 2022). ArOH reactions with $\bullet\text{OH}$ initially produce abundant hydroxylated-ArOH analogues, while $^3\text{C}^*$ oxidation of ArOH dominantly forms oligomers (Yu et al., 2016). (Misovich et al., 2021; Yu et al., 2016). However, the extended aging processes for phenolic brown carbon may vary since chromophores of different chemical composition react differently to photooxidation (Li et al., 2022). In cloud and fog conditions, extended $\bullet\text{OH}$ reactions transform phenolic aqSOA chromophores more rapidly compared to reactions with $^3\text{C}^*$ (Jiang et al., 2023b). However, in more concentrated conditions of particle water, the concentrations of $\bullet\text{OH}$ and $^3\text{C}^*$ are different than under the dilute conditions of cloud and fog drops (Ma et al., 2023, 2024), which likely affects the lifetimes of phenolic brown carbon.

85 In this study, we monitored the formation and loss of light absorption by aqueous SOA formed by reactions of six highly substituted BB phenols with $\bullet\text{OH}$ and $^3\text{C}^*$. We determined mass absorption coefficients (MAC) for the parent ArOH and for the evolving aqSOA as a function of reaction time. We also calculated the rate of sunlight absorption by aqSOA throughout the course of each reaction and the lifetimes of this absorbance in our experiments as well as for estimated cloud and particle water conditions.

2 Experimental

2.1 Chemicals

Chemicals were used as received. Hydrogen peroxide (H_2O_2 , 30% solution in water), 3,4-dimethoxybenzaldehyde (DMB, 99%), 4-hydroxy-3-methoxyphenylacetone (guaiaacyl acetone, GA, 96%), 4-(hydroxymethyl)-2-methoxyphenol (vanillyl

95 alcohol, VAL, $\geq 98\%$), 2-(4-hydroxyphenyl)ethanol (tyrosol, TYR, 98%), 4-hydroxy-3,5-dimethoxybenzoic acid (syngic acid, SyrAcid, $\geq 95\%$), 3-(4-hydroxy-3-methoxyphenyl)prop-2-enoic acid (*trans*-ferulic acid, FA, 99%), and 2-nitrobenzaldehyde (2NB, 98%) were purchased from Sigma-Aldrich. Upon illumination, *trans*-FA photoisomerizes to a 41:58 mixture of the *cis*- and *trans*- isomers; our experiments were performed on this mixture. (3,5-dimethoxy-4-hydroxyphenyl)acetone (syngyl acetone, SA, 82%) was synthesized by Carbosynth LLC. Sulfuric acid (trace metal grade) and acetonitrile (Optima LC-MS grade) were from Fisher Scientific. Chemical solutions were prepared in air-saturated ultrapure Milli-Q water (≥ 18.2 M Ω cm) from a Millipore Advantage A10 system with an upstream Barnstead activated carbon filter.

2.2 Reaction Solutions

105 Fresh air-saturated solutions were prepared daily containing 50 or 100 μ M ArOH and either 5 or 10 mM H₂O₂ (as an \bullet OH precursor) or 5 or 10 μ M DMB (as a 3 C* precursor) (Table S1). Solutions were adjusted to pH 5 using sulfuric acid and \sim 20 mL were transferred to an airtight 5 cm quartz cuvette. Solutions were illuminated at 20 $^{\circ}$ C with constant stirring in a solar simulator equipped with a 1000 W Xenon lamp with three downstream optical filters: a water filter, an AM1.0 air mass filter (AM1D-3L, Sciencetech), and a 295 nm long-pass filter (20CGA-295, Thorlabs); see Figure S1 for a typical photon flux.

110 During illumination we periodically removed aliquots of sample and measured the ArOH concentration using high-performance liquid chromatography (HPLC: Shimadzu LC-20AB pump, Thermo Scientific Accucore XL C₁₈ column (50 \times 3 mm, 4 μ m bead), and Shimadzu SPD-M20A photodiode array detector). HPLC conditions (eluent, flow rate, detection wavelengths) are described elsewhere (Arciva et al., 2022). We sampled to approximately three phenol half-lives, i.e., to $\sim 3t_{1/2}$, when roughly 12% of the initial phenol remained. We report times for each aliquot removal in Table S1 and show phenol decay kinetics in Figure S2. We measured dark controls in a separate temperature-controlled dark chamber at 20 $^{\circ}$ C with constant stirring; there was no phenol loss in the dark. Additionally, three phenols (SA, SyrAcid, and FA) undergo direct photodegradation, which contributed to 6 to 35% of measured ArOH loss (Figure S3).

120 At each time point we also measured light absorption by removing the sample cell from the illumination system and taking a UV/Vis spectrum in a Shimadzu UV-2501PC spectrophotometer baselined with Milli-Q water adjusted to pH 5 with sulfuric acid. Absorbance measurements were baseline corrected and represent the absorbance of reaction mixtures (i.e., parent ArOH, oxidant precursor, and aqSOA formed). Molar absorption coefficients for the six phenols are in Table S2 and shown graphically in Figure [S3S4](#). Absorbance of the mixtures during reaction are shown in Figure [S4S5](#) for the \bullet OH reactions and Figure [S5S6](#) for the 3 C* reactions.

125 2.3 Oxidant Concentrations

Using measured ArOH decay kinetics, we estimated the steady-state concentration of oxidant in each experiment. We first determined the pseudo-first-order decay rate constant (k'_{light}) for phenol loss by $\bullet\text{OH}$ or $^3\text{C}^*$ as the negative value of the slope of a plot $\ln\left(\frac{[\text{ArOH}]_t}{[\text{ArOH}]_0}\right)$ versus reaction time, where $[\text{ArOH}]_t$ and $[\text{ArOH}]_0$ are the phenol concentrations at times t and zero (Figure S2). Next, we normalized k'_{light} to sunlight conditions at midday on the winter solstice at Davis, and corrected for any direct photodegradation, using:

$$k'_{\text{ArOH}} = \left[\left(\frac{k'_{\text{light}}}{j_{2\text{NB},\text{exp}}} \right) \times j_{2\text{NB},\text{win}} \right] - j_{\text{ArOH}} \quad (1)$$

$$k'_{\text{ArOH}} = \left[\left(\frac{k'_{\text{light}}}{j_{2\text{NB},\text{exp}}} \right) \times j_{2\text{NB},\text{win}} \right] - j_{\text{ArOH}} \quad (1)$$

135 In this equation, k'_{ArOH} is the normalized rate constant for phenol loss, $j_{2\text{NB}}$ is the first-order decay rate constant for loss of 2NB (our actinometer) on the day of the phenol experiment, $j_{2\text{NB},\text{win}}$ is the value measured at midday on the winter solstice in Davis, 0.0070 s^{-1} (Anastasio and McGregor, 2001), and j_{ArOH} is the rate constant for direct photodegradation of ArOH. Rate constants are summarized in Table S3 and S4. We then estimated the concentration of oxidant in each solution by dividing k'_{ArOH} by the second-order rate constant for phenol reacting with oxidant ($k_{\text{ArOH},\text{OH}^\bullet}$, $\text{M}^{-1}\text{s}^{-1}$), $\bullet\text{OH}$ or $^3\text{C}^*$, which are from our previous works (Arciva et al., 2022; Ma et al., 2021). Since triplet solutions can also form $\bullet\text{OH}$ and singlet molecular oxygen ($^1\text{O}_2^*$), we estimated the potential contributions of these secondary oxidants to phenol loss. As described in Section S1 of the supplement, these oxidants appear to be negligible sinks for ArOH in our triplet experiments.

2.4 Mass absorption coefficients, Rate of Sunlight Absorption by aqSOA, and Lifetime of aqSOA Light Absorbance

145 Mass absorption coefficients (MAC; $\text{m}^2 \text{ g}^{-1}$) for the aqSOA at each sampling time were determined at each wavelength λ using:

$$\text{MAC}_{\text{aqSOA},\lambda} (\text{m}^2 \text{ g}^{-1}) = \frac{2.303 \times \text{Abs}_{\text{aqSOA},\lambda} \times 10^3 \times 10^{-4}}{l \times (Y_{\text{SOA}} \times \Delta[\text{ArOH}])} \quad (12)$$

150 where 2.303 converts the absorbance from base-10 to base-e, $\text{Abs}_{\text{aqSOA},\lambda}$ is the absorbance of the solution at wavelength λ , corrected to remove the absorbances from the oxidant precursor and remaining parent phenol (see Section S1S2); l is the path length of our cuvette (5 cm); Y_{SOA} is the aqSOA mass yield during oxidation of ArOH with $\bullet\text{OH}$ or $^3\text{C}^*$ (Arciva et al., 2022; Ma et al., 2021), which, which were determined using high-resolution aerosol mass spectrometry (AMS) and are summarized in Table S5; and $\Delta[\text{ArOH}]$ is the decrease in phenol mass concentration (in g L^{-1}) between times zero and t . The factor of 10^3

155 converts from L to cm³ while the factor of 10⁻⁴ converts from cm² to m². The MAC values for each of the six highly substituted
parent phenols were calculated as described in Section S1S2 of the supplement.

We calculated the overall rate of sunlight absorption by aqSOA (R_{abs}) at a given time point by multiplying the corresponding
MAC_{aqSOA} by the modelled actinic flux and summing from 280 to 800 nm:

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$$R_{\text{abs}} \text{ (mol photon g}^{-1} \text{ s}^{-1}) = \sum \left(\frac{\text{MAC}_{\text{aqSOA}} \times I_{\lambda} \times 10^4 \times \Delta\lambda}{N_A} \right). \quad (23)$$

Here, I_{λ} is the actinic flux (photons cm⁻² s⁻¹ nm⁻¹) at midday in Davis on the winter solstice from the Tropospheric Ultraviolet
165 and Visible (TUV) Radiation Model version 5.3 (Figure S1), $\Delta\lambda$ is the interval between TUV wavelengths (1 nm), N_A is
Avogadro's number, and the factor of 10⁴ converts from m² to cm². We calculated the first-order rate constant for loss of
sunlight absorption by aqSOA during continued reaction as the negative of the slope of a plot of the natural log of R_{abs} versus
reaction time. These rate constants were normalized to Davis wintertime solstice conditions to determine k'_{Rabs} , the pseudo-
first-order decay rate constant for the loss of sunlight absorption by aqSOA. From this we calculated the lifetime for aqSOA
170 sunlight absorption under our experimental conditions using Eq (34):

$$\tau_{\text{exp}} = \frac{1}{k'_{\text{Rabs}}}. \quad (3)$$

$$\tau_{\text{BrC,exp}} = \frac{1}{k'_{\text{Rabs}}}. \quad (4)$$

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This is the timescale for the loss of sunlight absorption during continued illumination, which we refer to as the lifetime of light
absorption by brown carbon or, more simply, the lifetime of brown carbon. To extrapolate these calculated lifetimes of
aqSOA BrC light absorbance from laboratory to ambient oxidant conditions, we first calculated values of $\frac{[\text{OH}]_{\text{ambient}}}{[\text{OH}]_{\text{exp}}}$, the value
of $\frac{[\text{OH}]_{\text{ambient}}}{[\text{OH}]_{\text{exp}}}$, which is the ambient-to-lab ratio of oxidant (³C* or •OH) concentration, and the analogous ratio for ³C*. We

180 then multiplied k'_{Rabs} by this ratio to determine $k'_{\text{BrC,ambient}}$, the pseudo-first-order rate constant for loss of BrC absorbance
due to that oxidant under ambient conditions. Aqueous concentrations of •OH and ³C* were taken from Ma et al. (2023a) Ma
et al. (2024) based on measurements of extracts from four different types of particles (winter-spring particles, summer-fall,
fresh wildfire, and aged wildfire) from Davis, California, and including mass transport of gas-phase •OH. We calculated light-
absorption BrC lifetimes for two conditions: cloud/fog drops and ALW, with PM mass/water mass ratios of 3 × 10⁻⁵ and 1 μg-
185 PM/μg-water, respectively. Aqueous Average aqueous oxidant concentrations (including mass transport of gas-phase •OH)

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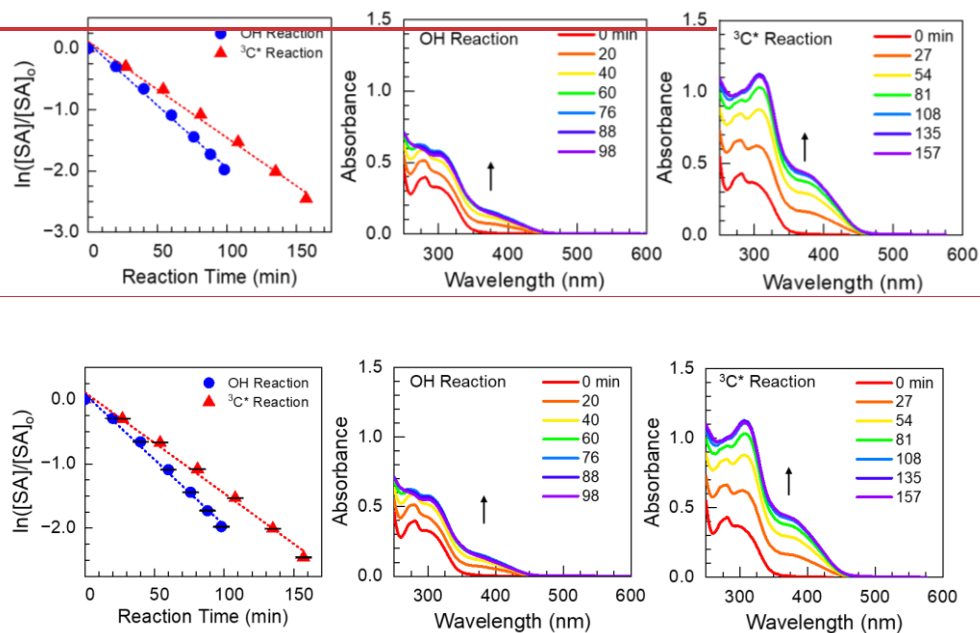
16) were: $[^{\bullet}\text{OH}] = 1.7.6 (\pm 4.3) \times 10^{-14.5}$ M for clouds and fog drops and $5.56.8 (\pm 1.9) \times 10^{-15}$ M for ALW; $[^3\text{C}^*] = 3.76 (\pm 2.6) \times 10^{-14}$ M for cloud/fog drops and $5.8.2 (\pm 3.9) \times 10^{-13}$ M for ALW. We also calculated the overall ambient lifetime of brown carbon with respect to both $^{\bullet}\text{OH}$ and $^3\text{C}^*$ under these two conditions using Eq (5);

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$$\tau_{\text{BrC,ambient}} = \frac{1}{k'_{\text{Rabs,OH}} + k'_{\text{Rabs,}^3\text{C}^*}} \quad (5)$$

3 Results and Discussion



195 **Figure 1.** Left panel: Decay kinetics for the loss of syringyl acetone (SA) reacting with $^{\bullet}\text{OH}$ (blue circles) or $^3\text{C}^*$ (red triangles). Error bars represent one standard deviation, estimated from the average relative standard deviations from the corresponding dark controls. Middle and right panels: Absorbance (in a 5 cm cell) of the reaction mixtures (i.e., oxidant precursor, starting phenol, and products) at various reaction times. Arrows represent the time trends in absorbance.

3.1 Aqueous Phenol Oxidation by $^{\bullet}\text{OH}$ and $^3\text{C}^*$

200 In each experiment, we measured both the aqueous oxidation of ArOH and changes in light absorption by the reaction mixture. As illustrated in Figure 1, the photooxidation of each phenol follows pseudo-first-order kinetics, with rate constants for phenol

loss (k'_{ArOH}) in the range of $(1.2 - 4.1) \times 10^{-4} \text{ s}^{-1}$ for $\bullet\text{OH}$ experiments (Table S3) and $(0.19 - 2.4) \times 10^{-4} \text{ s}^{-1}$ for $^3\text{C}^*$ experiments (Table S4). Differences in kinetics were largely driven by starting conditions, such as initial reactant and oxidant concentrations, and small changes in lamp intensity (see j_{2NB} values in Tables S3 and S4). We used k'_{ArOH} values to estimate the steady-state concentration of oxidant in each solution (Section 2.3). As shown in Table 1, concentrations of $\bullet\text{OH}$ were in the range of $(6.5 \text{ to } 18) \times 10^{-15} \text{ M}$, similar to the typical $[\bullet\text{OH}]$ range in cloud/fog drops and ALW (Kaur et al., 2019; Ma et al., 2023a2023). For $^3\text{C}^*$, our experimental concentrations were $(1.2 - 11) \times 10^{-14} \text{ M}$, in the range of cloud/fog drops but ~~10–100 times~~ lower than the ~~typical~~ average ALW value of roughly $10^{-12} \text{ M}^{13} \text{ M}$ described in section 2.4 (Ma et al., 2023a2023).

Each of the parent phenols has an absorption peak between 260 to 290 nm, but only FA, SyrAcid, and ~~(to a much smaller extent)~~ SA have any significant absorbance above 300 nm (Figure S3). These three phenols ~~can~~ undergo direct photodegradation (Ma et al., 2021), which ~~is a minor sink but~~ accounted for 6 – 35% of ArOH loss (Figure S3) and might be a source of light-absorbing products in our experiments. The amount of light absorption by the reaction mixture after illumination varies significantly between different solutions (Figures S4S5 and S5S6). General trends of absorption by phenolic reaction mixtures throughout the course of illumination vary, but there is always more absorption at longer wavelengths compared to the starting phenol. Oxidant type also impacts light absorbance by the aqSOA: after the initial time point, continued $\bullet\text{OH}$ oxidation generally decreases absorbance below wavelengths of 300 nm but increases absorbance above 300 nm (Figures 1 and S4S5), while $^3\text{C}^*$ oxidation mostly increases absorbance across all wavelengths (Figures 1 and S5S6).

Table 1. Oxidant concentrations, pseudo-first-order rate constants for the loss of BrC light absorption (k'_{Rabs}), the total rate of sunlight absorption at the first illumination time point (total $R_{abs,t1}$) for each reaction system, and the fraction of light absorption from 280 to 400 nm.

Phenol	$\bullet\text{OH}$ Reactions				$^3\text{C}^*$ Reactions			
	$[\bullet\text{OH}]_{\text{exp}}$ (10^{-15} M)	k'_{Rabs} (10^{-3} min^{-1})	Total $R_{abs,t1}$ ($10^{-4} \text{ mol photon g}^{-1} \text{ s}^{-1}$)	Fraction of $R_{abs,t1}$ from wavelengths < 400 nm	$[^3\text{C}^*]_{\text{exp}}$ (10^{-14} M)	k'_{Rabs} (10^{-3} min^{-1})	Total $R_{abs,t1}$ ($10^{-4} \text{ mol photon g}^{-1} \text{ s}^{-1}$)	Fraction of $R_{abs,t1}$ from wavelengths < 400 nm
Tyrosol (TYR)	8.9	1.0	0.85	0.23	4.0	0.095	0.62	0.37
Guaiacyl acetone (GA)	8.2	3.6	2.0	0.48	2.9	0.50	0.67	0.83
Vanillyl alcohol (VAL)	8.4	4.3	3.1	0.25	1.2	0.17	6.7	0.16
Ferulic acid (FA)	6.5	2.4	2.4	0.67	4.6	0.0053	0.79	0.74

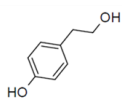
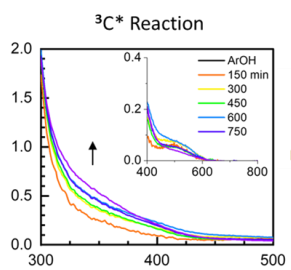
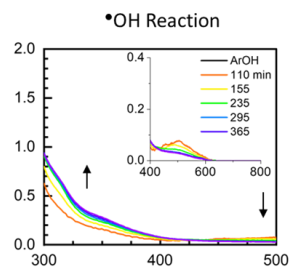
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Syringic Acid (SyrAcid)	18	9.4	0.70	0.63	11	3.9	2.4	0.64
Syringyl acetone (SA)	17	6.9	4.7	0.69	8.7	2.8	8.2	0.55

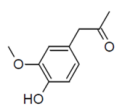
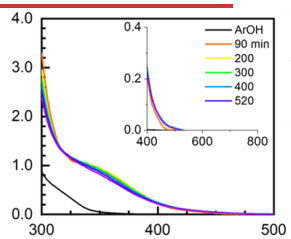
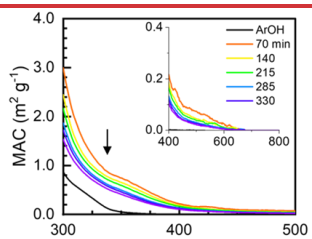
3.2 Mass Absorption Coefficients (MAC)

225 To compare light absorption by aqSOA as a function of phenol, oxidant, and illumination time, we determined mass absorption coefficients (MAC) for each experiment. aqSOA MAC values were determined by correcting measured absorbance values for the contributions from both the remaining starting phenol and the oxidant precursor (Section S4S2). Generally, across tropospherically relevant wavelengths (i.e., above 300 nm), the MAC curves of aqSOA are highest around 300 nm (Figures 2 and S6S7) and tail at longer wavelengths. The MAC values of the starting phenols (MAC_{ArOH}) are typically low compared to values of the resulting aqSOA (MAC_{aqSOA}), especially for the first oxidation time point ($MAC_{aqSOA,t1}$) for both $\bullet OH$ and $^3C^*$ (Table S6). The one exception is FA, which has a high parent absorbance that decreases during oxidation. After the first illumination time point, continued reaction ~~of the other phenols can increase or decrease the~~generally decreases light absorbance by the aqSOA, ~~depending on the oxidant, phenol,~~but there are some exceptions (e.g., TYR in Figure 2 and ~~wavelength-VAL with $^3C^*$ in Figure S7).~~

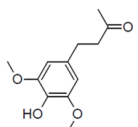
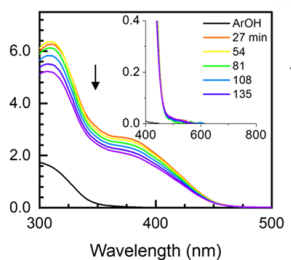
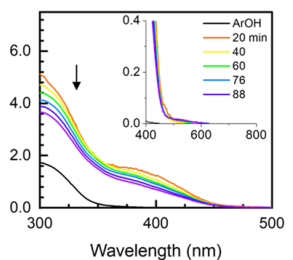
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tyrosol
(TYR)



guaiacyl acetone
(GA)



syringyl acetone
(SA)

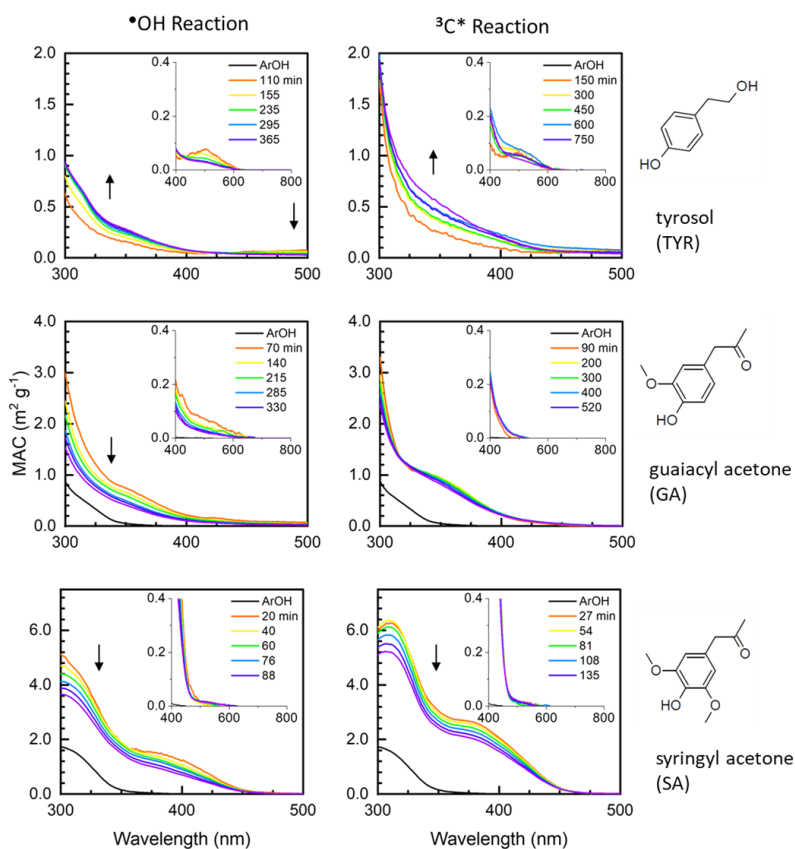


Figure 2. Mass absorption coefficients for aqSOA formed via reactions with $\bullet\text{OH}$ (left column) and $^3\text{C}^*$ (right column) for tyrosol (top plots), guaiacyl acetone (middle plots) and syringyl acetone (bottom plots). For a given phenol, each colored line represents a different illumination time (see legend for times). Arrows represent the time trends of aqSOA MAC values after the initial illumination time point. The black line in each panel is the MAC for the parent phenol; the absorbance of the remaining parent phenol was removed from the aqSOA MAC values at each illumination time (SI Section S4S2). The MAC for TYR is zero in this wavelength range.

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Figure 2 shows MAC values for three of the phenols we studied — tyrosol, guaiacyl acetone, and syringyl acetone — and their resulting aqSOA. These phenols are substituted versions of the three most abundant phenols from BB: phenol, guaiacol (2-methoxyphenol), and syringol (2,6-dimethoxyphenol). Results for the other three phenols we studied, which are derivatives of

guaiacol and syringol, are shown in Figure S6-S7. For each solution, we first sampled the illumination solution when approximately 25% of the parent phenol had decayed (i.e., $0.5 t_{1/2}$), so the corresponding spectrum likely does not capture the maximum absorbance exhibited by the solution. For most of the $\bullet\text{OH}$ reactions with the six phenols, the MAC of aqSOA increases from the parent ArOH (which typically has little to no absorbance in the solar wavelengths) to the initial aqSOA. This is because $\bullet\text{OH}$ reactions with ArOH form hydroxylated products, which can be important contributors to light absorption (Sun et al., 2010; Yu et al., 2014). Additionally, the formation of quinones, which can contain extensive conjugated pi-electrons, might explain our observation of increased aqSOA light absorption at longer wavelengths (Dulo et al., 2021). The aqSOA of TYR is the least absorbing, and aqSOA MAC tends to increase with increased methoxy substitution on the aromatic ring (Figures 2 and S6S7). For example, the $\text{MAC}_{\text{aqSOA},11}$ of SyrAcid is greater than that of TYR by a factor of approximately 10 at 300 nm. FA oxidation is somewhat of an exception since the MAC of FA is higher than that of its aqSOA at wavelengths below 350 nm (Figure S6S7). The decrease in absorbance during oxidation for FA is probably because of a loss of unsaturation in the acrylic acid substituent, which decreases conjugation and light absorption.

For $^3\text{C}^*$ reactions, the trends in absorbance are similar to the $\bullet\text{OH}$ results above, including a general increase in MAC from the parent phenol to the initial daughter aqSOA, as shown in Figure 2 for TYR, GA, and SA (and Figure S6S7 for VAL, FA, and SyrAcid). More highly substituted ArOH (FA, SA, and SyrAcid) produce aqSOA that absorbs more strongly compared to the less substituted phenols (TYR, GA, and VAL). Also, $^3\text{C}^*$ -derived aqSOA absorbs more light compared to aqSOA from the corresponding $\bullet\text{OH}$ reactions. The $\text{MAC}_{\text{aqSOA}}$ at 300 nm for SyrAcid at this condition was $9.0 \text{ m}^2 \text{ g}^{-1}$, the highest MAC we observed for both $\bullet\text{OH}$ and $^3\text{C}^*$ reactions. One reason for higher MAC values in $^3\text{C}^*$ -produced aqSOA may be because triplet reactions produce more oligomers, which can strongly absorb light, compared to $\bullet\text{OH}$ reactions (Yu et al., 2014; Jiang et al., 2021).

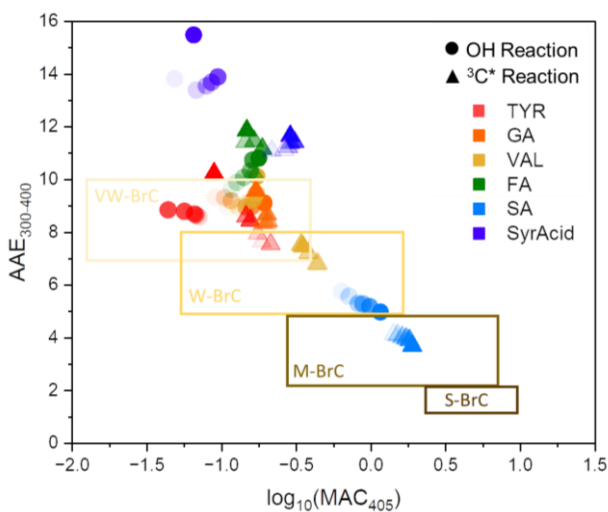


Figure 3. aqSOA light absorption based on the brown carbon (BrC) classification scheme (Saleh, 2020). The boxes from top to bottom represent very weakly absorbing BrC (VW-BrC), weakly absorbing BrC (W-BrC), moderately absorbing BrC (M-BrC), and strongly absorbing BrC (S-BrC). Triangles represent aqSOA formed via reactions of each phenol with the triplet state of DMB, while circles are data from reactions of each phenol with hydroxyl radical. The time series of each reaction follows a gradient: the first time point is represented as the darkest marker followed by a decrease in color with increasing time.

In Figure 3 we plot our MAC data at 405 nm in the BrC classification scheme from Saleh (2020). Most of the phenolic BrC is very weakly absorbing, but the products from SA (our most strongly absorbing aqSOA at 405 nm) are weakly to moderately absorbing. The magnitude of $AAE_{300-400nm}$ (3.7 to 15) and MAC_{405} (0.05 to 1.9) (Table S7) depend on parent ArOH and generally increase with more substituted ArOH. BrC from SA is an exception because the MAC at both 300 and 400 nm are high and do not change drastically across these wavelengths compared to BrC from the other phenols. Oxidant identity also plays an important role in the extent of light absorption by the BrC formed. Most of the BrC formed by $\bullet OH$ oxidation was very weakly absorbing, while BrC formed by triplet oxidation had a broader range of absorption, spanning very weakly to moderately absorbing. Also, in most cases, continued oxidation photobleaches BrC towards the very-weakly absorbing bin (Figure 3). Our observation that most phenolic BrC is very weak to weakly absorbing is consistent with results for water-soluble BrC in ambient particles impacted by residential wood combustion and wildfires (Jiang et al., 2023; Ma et al., 2023b)(Jiang et al., 2023a; Ma et al., 2024).

3.3 Rates of Sunlight Absorption

We next calculated R_{abs} , the rate of sunlight absorption, for the parent phenol and resulting aqSOA at each time point, as shown in Figure 4 for TYR, GA and SA (and Figure S7S8 for VAL, FA, and SyrAcid). Except for FA, the parent phenols generally absorb very weakly at short solar wavelengths (i.e., below 400 nm) or do not absorb any sunlight (i.e., TYR and VAL). As such, the rate of sunlight absorption by aqSOA is generally much higher than that of the parent ArOH, with the exception of FA. For example, going from ArOH to aqSOA, R_{abs} for GA and SA increase by factors of 14 to 17 for $\bullet\text{OH}$ reactions, and 5.5 to 24 for triplet reactions. The increase for the R_{abs} of SyrAcid aqSOA is much higher, with factors of 74 to 250 for $\bullet\text{OH}$ and triplet reactions, respectively, due to the very weak absorption by the parent phenol. In contrast, ferulic acid aqSOA shows very small changes compared to FA, with factors of 1.6 for $\bullet\text{OH}$ reactions and 0.52 for triplet reactions. For the entire set of six phenols, continued oxidation after the first illumination time point can either increase or decrease R_{abs} . For $\bullet\text{OH}$ reactions, the rate of sunlight absorption for aqSOA from VAL, GA, FA, and SA decreases with increasing oxidation, i.e., continued aging produces less-absorbing aqSOA, possibly because of fragmentation by $\bullet\text{OH}$ (Sun et al., 2010). Continued $^3\text{C}^*$ reactions generally decrease R_{abs} for aqSOA, but much more slowly than $\bullet\text{OH}$ reactions. This may be because triplets are less effective at oxidizing aqSOA than is $\bullet\text{OH}$ (Jiang et al., 2021).

We also calculated the fraction of sunlight absorption due to UV wavelengths, i.e., those below 400 nm. As summarized in Table 1 for the initial time point, the fraction of R_{abs} due to UV wavelengths varies significantly across different ArOH but also, to a smaller extent, across the two oxidants. Fractions of R_{abs} due to UV wavelengths for each illumination time point are shown in Table S7. Generally, more substituted ArOH produce aqSOA where shorter wavelengths dominate sunlight absorption, and this remains throughout the course of reaction. In contrast, for simpler ArOH (i.e., TYR, GA, and VAL), the fraction of R_{abs} due to wavelengths below 400 nm generally increases with reaction time (Table S7). Also, except for VAL, the fraction of R_{abs} contributed by short wavelengths is always higher for the triplet reaction compared to the $\bullet\text{OH}$ reaction. Interestingly, the VAL-aqSOA formed by both oxidants (Figure S7S8, top plot) and TYR-aqSOA by $\bullet\text{OH}$ (Figure 4, top plot) contain characteristic peaks at 500 nm, possibly due to quinones. In the case of VAL, this peak behaves differently for the two oxidants. For VAL reaction with $\bullet\text{OH}$, aqSOA light absorption diminishes across the entire wavelength range due to photobleaching, with the fastest loss between 400 to 800 nm, possibly due to faster direct photodegradation of chromophores. In contrast, during reactions with triplets, the VAL-aqSOA becomes more absorbing between 300 and 425 nm but photobleaches at longer wavelengths. We observe similar behaviour for TYR-aqSOA with $\bullet\text{OH}$. This indicates that during the evolution of aqSOA light absorption can simultaneously increase in some wavelength regions and decrease in others.

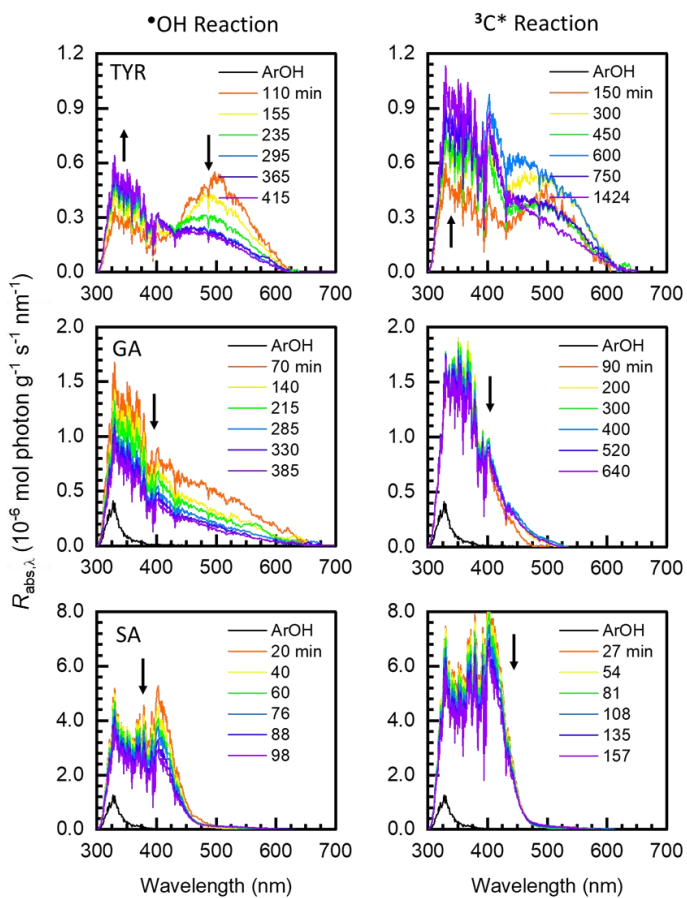


Figure 4. Wavelength-specific rates of sunlight absorption by aqSOA formed via reactions with $\bullet\text{OH}$ (left column) and ${}^3\text{C}^*$ (right column) for tyrosol (top plots), guaiacyl acetone (middle plots) and syringyl acetone (bottom plots). For a given phenol, the black line represents sunlight absorption by the parent ArOH and colored lines represent absorption for aqSOA at different illumination times (see legend). Arrows represent the time trends of aqSOA MAC values after the initial illumination time point. Rates of aqSOA sunlight absorption for vanillyl alcohol, ferulic acid, and syringic acid are shown in SI Figure S7S8.

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3.4 Lifetimes of Light-Absorbing AqSOA

We calculated the lifetime of light absorption by aqSOA (i.e., the lifetime of phenolic BrC) by monitoring the decline in R_{abs} with continued illumination, as shown in Figure 5 for SA. We show the equivalent figures for the other five phenols in Figures S8S9 and S9S10. Plotting the natural log of R_{abs} versus reaction time yields the experimental first-order decay rate constant for loss of light absorption ($k'_{\text{Rabs,exp}}$). In some cases, the natural log of R_{abs} increases at short reaction times (e.g., the triplet reactions of TYR, GA, and FA) before decreasing; we did not include these increasing points in our determination of $k'_{\text{Rabs,exp}}$. We then normalized this rate constant to Davis midday winter solstice sunlight to obtain (k'_{Rabs}) (Table S8); values are summarized in Table 1. k'_{Rabs} for $\bullet\text{OH}$ reactions are in the range $(1.0 \text{ to } 9.4) \times 10^{-3} \text{ min}^{-1}$, while rate constants for triplet reactions are lower and range from $(0.0053 \text{ to } 3.9) \times 10^{-3} \text{ min}^{-1}$, i.e., the loss of light-absorbing aqSOA by triplets is much slower than with $\bullet\text{OH}$ under our experimental conditions. It is also Jiang et al. (2023b) also measured the first-order decay rate constant for loss of light absorption for BrC formed via reactions of GA with $\bullet\text{OH}$ and $^3\text{C}^*$. Compared to their experiments (with no additional oxidant), the rate of decay of $^3\text{C}^*$ -aqSOA in Jiang et al. (2023) is higher than our value because their $[^3\text{C}^*]$ is higher by a factor of 3.8. In contrast, for $\bullet\text{OH}$ -aqSOA, our decay value is higher than that of Jiang et al. (2023) because our $[\bullet\text{OH}]$ is higher by a factor of 3.2. However, despite these differences in oxidant concentrations, the observed trends for aqSOA are similar. That is, initial photooxidation forms light absorbing aqSOA, but during extended aging $\bullet\text{OH}$ -aqSOA photobleaches rapidly while $^3\text{C}^*$ -aqSOA photobleaches more slowly. It is possible that molecular differences in the aqSOA formed by $^3\text{C}^*$ and $\bullet\text{OH}$ play a role in the different rates of photobleaching. ~~Triplets more~~ Triplet-mediated reactions efficiently form oligomers, while ~~$\bullet\text{OH}$ tends-mediated reactions tend~~ $\bullet\text{OH}$ -mediated reactions tend to form hydroxylated products that eventually fragment (Jiang et al., 20232023a; Yee et al., 2013). Also, ~~with~~during continuous oxidation, triplets ~~seem to be~~are less reactive towards aqSOA, especially ~~if possibly because~~ the ~~molecule does~~BrC molecules do not have a-phenolic hydroxyl ~~group~~groups or aliphatic double ~~bond~~bonds, which are both reactive with triplets (Kaur et al., 2018; Ma et al., 2021). In contrast, $\bullet\text{OH}$, which reacts rapidly with most organic molecules, more quickly ~~reaacts with~~degrades phenolic aqSOA.

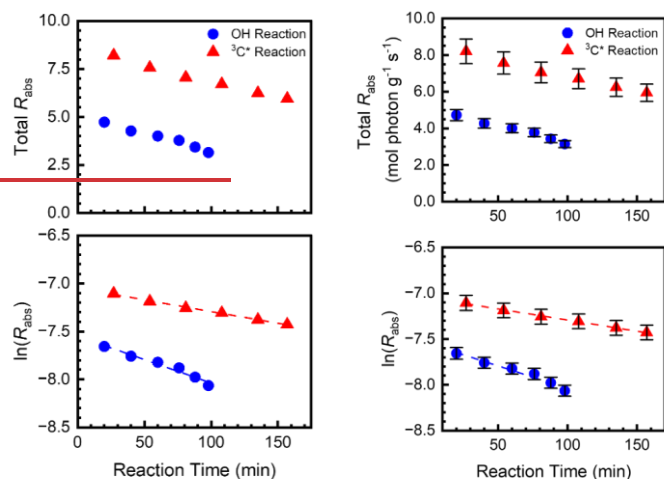


Figure 5. Top panel: Total rate of aqSOA sunlight absorption (R_{abs} , $10^{-4} \text{ mol photon g}^{-1} \text{ s}^{-1}$) as a function of reaction time during the oxidation of SA by $\bullet\text{OH}$ (blue circles) and $^3\text{C}^*$ (red triangles). Bottom panel: Natural log of R_{abs} versus reaction time, which was used to determine rate constants for the loss of light absorbance by aqSOA. **Error bars represent one standard deviation propagated from the standard deviations of the dark controls and aqSOA mass yields.**

The inverse of k'_{Rabs} is the lifetime of sunlight absorption by phenolic aqSOA during continued $\bullet\text{OH}$ and $^3\text{C}^*$ reactions under our experimental conditions, i.e., the timescale for loss of absorbance by brown carbon compounds in the aqueous SOA. Under our experimental conditions, these BrC lifetimes range from 1.8 to 16 hours for $\bullet\text{OH}$ solutions and 4.3 to 3,200 hours for $^3\text{C}^*$ experiments (Figure S40S11). Next, we extrapolate these experimental lifetimes to ambient lifetimes by assuming that the concentration of oxidant ($\bullet\text{OH}$ or $^3\text{C}^*$) is driving the loss of brown carbon. We consider two ambient conditions: cloud/fog drops (with a PM mass/liquid water mass ratio of $3 \times 10^{-5} \mu\text{g-PM} / \mu\text{g-water}$) and ALW ($1 \mu\text{g-PM} / 1 \mu\text{g-water}$). Concentrations of the two oxidants under our experimental conditions, as well as values for cloud/fog drops and ALW, are described in Tables S9 and S10; the cloud/fog and ALW oxidant concentrations are based on measurements and extrapolations in Ma et al. (2023a2024).

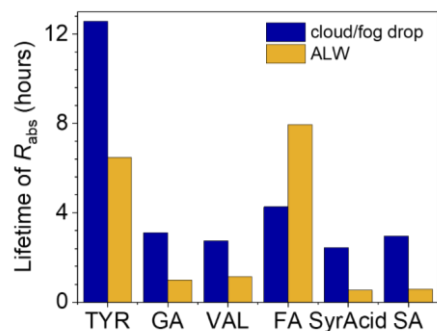
Because the aqueous concentration of $\bullet\text{OH}$ is similar in our experimental solutions, in cloud/fog drops, and in ALW, we find similar lifetimes for sunlight absorption by aqSOA under all three conditions for $\bullet\text{OH}$, as shown in Figure S40S11. Lifetimes of sunlight absorption BrC during reactions in the $\bullet\text{OH}$ solutions range from 3.04.2 to 26.21 hours across both ambient conditions for aqSOA from all six highly substituted phenols (Figure S40S11). Differences in ArOH substitution do not significantly influence the lifetime of light absorption for aqSOA BrC formed from $\bullet\text{OH}$, with lifetimes mostly within a factor

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of two. Even for TYR-aqSOA, which forms longer-lived absorbing aqSOA, the BrC lifetime-of-sunlight absorption only differs by a factor of 4 compared to that for the other phenols. As mentioned above for TYR-aqSOA, this is complicated by the simultaneous occurrence of both photobleaching and photo-enhancement in different wavelength regions.

375 Lifetimes for the light-absorbing aqSOA in the triplet solutions are different/differ from the $\cdot\text{OH}$ -associated values in two main ways. First, extrapolation to ambient conditions results in a broader range of lifetimes (Figure S10S11) due to the 2216-times increase in triplet concentrations as we move from cloud/fog drops to particle water. The higher concentration of triplets in ALW makes for shorter lifetimes of light-absorbing aqSOA, ranging from 0.5883 to 8.612 hours (except for FA, which is 180250 hours). However, under cloud/fog conditions, where triplet concentrations are lower, lifetimes of phenolic BrC with
380 respect to triplet reactions are much longer, ranging from 13 to 190200 hours. The exception again is FA, where the light-absorbing aqSOA had a lifetime of 4,000 hours. Unlike the other reaction systems, This resistance of FA-aqSOA formed by FA is more resistant to $^3\text{C}^*$ oxidation, is possibly because a large fraction of FA-tripletaqSOA interactions with triplets result in physical quenching of the triplet rather than chemical loss/transformation of the aqSOA, likely because the unsaturated side chains from FA undergo isomerization readily (Ma et al., 2021).

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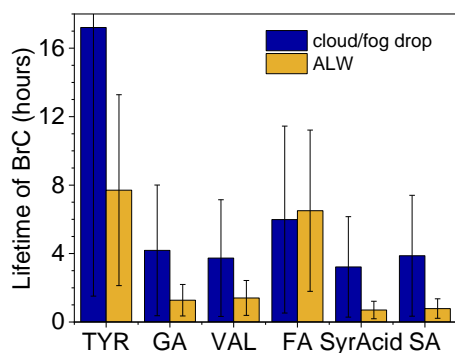


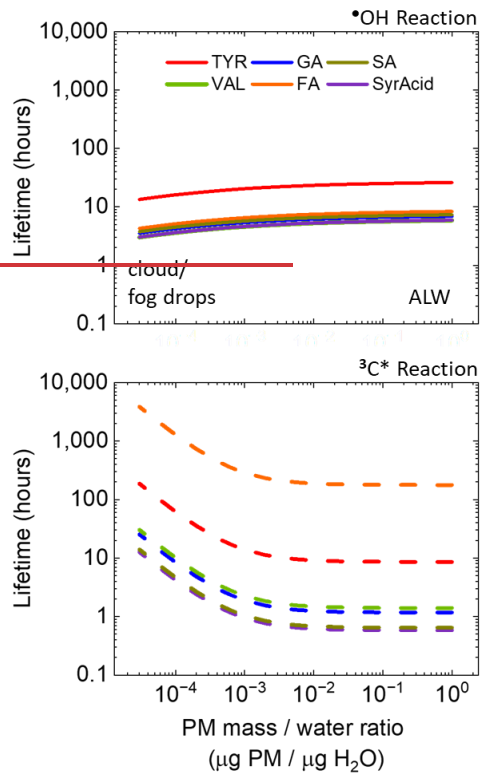
Figure 6. Overall lifetimes of light-absorbing aqSOA brown carbon (BrC) with respect to photooxidation by $\bullet\text{OH}$ and $^3\text{C}^*$ under cloud/fog drop and aerosol liquid water (ALW) conditions—calculated using Equation 5. Lifetimes with respect to photooxidation from the two oxidants individually are shown in Figure S10,S11. Error bars represent one standard deviation, primarily representing the variability in photooxidant concentrations averaged across the four different sample types (Ma et al., 2024), but also include the uncertainty in the rate constant for loss of brown carbon light absorption.

We next calculate the overall lifetime of light absorption by phenolic BrC with respect to both $\bullet\text{OH}$ and $^3\text{C}^*$ for the two atmospheric conditions. As shown in Figure 6, for cloud/fog drops the phenolic BrC lifetimes are typically a few hours but range up to 17 hours for TYR aqSOA; these lifetimes are controlled by $\bullet\text{OH}$ (Figure S10,S11). In particle water, phenolic BrC lifetimes range from 0.5370 to 8.07.7 hours; these are typically shorter than in cloud/fog drops and are generally controlled by reactions with $^3\text{C}^*$. The exception is FA, where the lifetimes are controlled by $\bullet\text{OH}$ and thus the lifetime is shorter in cloud/fog drops. While many of the light-absorbing aqSOA lifetimes are under 4 hours for the two conditions, the longer lifetimes for brown carbon from TYR and FA show that identity of the precursor phenol also plays a role.

Lastly, we compare the lifetimes of light-absorbing aqSOA across a continuum of liquid water content (LWC), from cloud/fog conditions to ALW (Figure 7). Combining winter $\bullet\text{OH}$ and $^3\text{C}^*$ concentrations as a function of liquid water content from Ma et al. (2023a)(Ma et al., 2023, 2024), we calculated the lifetimes of light-absorbing aqSOA with respect to the two oxidants individually. Lifetimes of phenolic aqSOA with respect to oxidation by $\bullet\text{OH}$ change relatively little across the enormous range of LWC, a consequence of the predicted small change in hydroxyl radical concentrations (Ma et al., 2023a,2023). In contrast, lifetimes of aqSOA produced and oxidized by $^3\text{C}^*$ show a much larger variation. In the dilute conditions of cloud/fog drops, aqSOA is less susceptible to aging by $^3\text{C}^*$, resulting in some of the longer lived aqSOA across the liquid water content range.

But as solutions become more concentrated, the triplet-mediated BrC lifetime drops quickly and then plateaus at a PM mass/water mass ratio of approximately 10^{-2} and above (Figure 7). This reflects changes in predicted triplet concentrations, which increase by approximately a factor of 2016 from cloud/fog to ALW. The net result is that lifetimes of light absorbance

by aqSOA are generally controlled by $\bullet\text{OH}$ under the more dilute conditions of cloud/fog drops, while triplets dominate phenolic BrC lifetimes under more concentrated conditions, including in ALW. The clear exception again is FA, where the BrC lifetime is always controlled by $\bullet\text{OH}$, while TYR follows the trend of the other phenols but with a less dramatic shift from $\bullet\text{OH}$ to $^3\text{C}^*$ controlling the lifetime as we move from cloud/fog drops to ALW (Figure 7). Otherwise, the other four phenols behave as a group, with no clear trend between ArOH substitution and aqSOA lifetime. One caveat to our calculations is that we assume the phenolic BrC lifetimes are governed by reactions with $\bullet\text{OH}$ and $^3\text{C}^*$ rather than direct photodegradation. But the latter pathway might also be important, especially for BrC species that absorb at longer wavelengths where the photon flux is higher (Figure S7SS). Although it is difficult for us to constrain this, any contribution of direct photodegradation to BrC loss under wintertime Davis conditions would be similar to that in our experimental conditions since our laboratory photon fluxes are similar to ambient values (Figure S1).



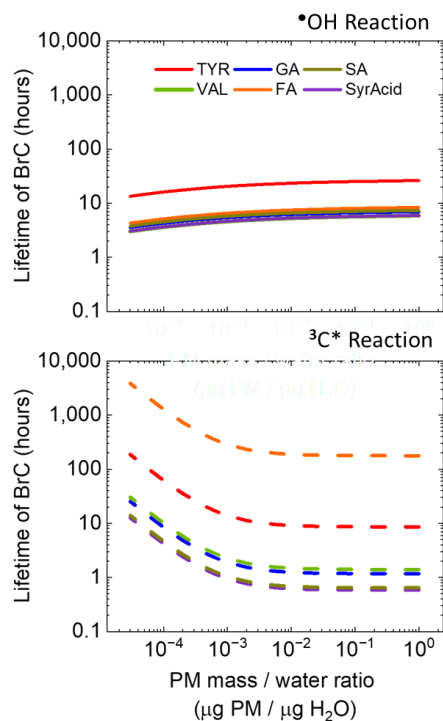


Figure 7. Lifetimes of light-absorbing aqSOA (i.e., BrC) as a function of particle mass-to-water mass ratio for TYR (red), GA (blue) and SA (mustard), VAL (green), FA (orange), and SyrAcid (purple). Solid and dashed lines indicate lifetimes with respect to reaction with $\bullet\text{OH}$ (top panel) and ${}^3\text{C}^*$ (bottom panel), respectively. For a PM concentration of $10 \mu\text{g m}^{-3}$, the range of PM mass/water mass ratios corresponds to liquid water contents of 1 to $1 \times 10^{-5} \text{ g m}^{-3}$, i.e., conditions of cloud/fog drops and ALW, respectively.

5 Conclusions

We examined light absorption by the aqueous secondary organic aerosol formed from reactions of phenols with two aqueous oxidants: hydroxyl radical ($\bullet\text{OH}$) and a triplet excited state (${}^3\text{C}^*$). These two routes are important for ArOH processing and significant contributors to aqSOA formation, but the light absorption trends with continuous reaction have not been evaluated previously. Initial reactions of highly substituted phenols with both oxidants produced aqSOA that absorbs much more sunlight compared to the parent phenol, a result of functionalization and/or oligomerization. Continued photo-aging alters the light

absorption by aqSOA. Overall, continued •OH oxidation more rapidly causes photobleaching, while the aqSOA formed and oxidized by triplet excited states loses its light absorption much more slowly. Our reactions are restricted with respect to each
440 oxidant, i.e., we did not study how susceptible aqSOA formed by •OH is to triplet oxidation or vice versa.

We also extrapolated from our experimental conditions to ambient conditions based on previously measured and estimated ambient oxidant concentrations. Lifetimes of ~~sunlight-absorption-by~~ phenolic aqSOA secondary BrC range from ~~3.2.5~~ to ~~1317~~ hours for cloud/fog drops, where •OH is the major sink, and from ~~0.5370~~ to ~~8.0-hours~~ ~~7.7~~ for particle water, where triplet excited states are the major sink. Overall, our results will help constrain atmospheric lifetimes of phenolic BrC, which should aid predictions of the global impact of BrC from biomass burning.
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Author Contributions

CA and SA developed the research goals and designed the experiments. SA and CM performed the •OH oxidation experiments, while LM and CG performed the triplet oxidation experiments. SA analysed the data and prepared the manuscript. CA
450 reviewed and edited the manuscript and provided supervision and oversight during the experiments and writing.

Competing Interests

The authors disclose that they have no conflict of interest.

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References

460 Aguilera, R., Corringham, T., Gershunov, A. and Benmarhnia, T.: Wildfire smoke impacts respiratory health more than fine particles from other sources: observational evidence from Southern California., Nat. Commun., 12(1), 1493, <https://doi.org/10.1038/s41467-021-21708-0>, 2021.

- 465 Akagi, S. K., Yokelson, R. J., Wiedinmyer, C., Alvarado, M. J., Reid, J. S., Karl, T., Crouse, J. D. and Wennberg, P. O.: Emission factors for open and domestic biomass burning for use in atmospheric models, *Atmospheric Chemistry and Physics*, 11(9), 4039–4072, <https://doi.org/10.5194/acp-11-4039-2011>, 2011.
- Anastasio, C. and McGregor, K. G.: Chemistry of fog waters in California's Central Valley: 1. In situ photoformation of hydroxyl radical and singlet molecular oxygen, *Atmos. Environ.*, 35(6), 1079–1089, [https://doi.org/10.1016/S1352-2310\(00\)00281-8](https://doi.org/10.1016/S1352-2310(00)00281-8), 2001.
- 470 Andreae, M. O.: Emission of trace gases and aerosols from biomass burning – an updated assessment, *Atmospheric Chemistry and Physics*, 19(13), 8523–8546, <https://doi.org/10.5194/acp-19-8523-2019>, 2019.
- Arciva, S., Niedek, C., Mavis, C., Yoon, M., Sanchez, M. E., Zhang, Q. and Anastasio, C.: Aqueous ·OH oxidation of highly substituted phenols as a source of secondary organic aerosol., *Environ. Sci. Technol.*, 56(14), 9959–9967, <https://doi.org/10.1021/acs.est.2c02225>, 2022.
- 475 Berndt, T. and Böge, O.: Gas-phase reaction of OH radicals with phenol, *Phys. Chem. Chem. Phys.*, 5(2), 342–350, <https://doi.org/10.1039/B208187C>, 2003.
- Bond, T. C., Streets, D. G., Yarber, K. F., Nelson, S. M., Woo, J.-H. and Klimont, Z.: A technology-based global inventory of black and organic carbon emissions from combustion, *Journal of Geophysical Research: Atmospheres*, 109(D14), 2004.
- Buxton, G. V., Greenstock, C. L., Helman, W. P. and Ross, A. B.: Critical Review of rate constants for reactions of hydrated electrons, hydrogen atoms and hydroxyl radicals ($\bullet\text{OH}/\bullet\text{O}^-$ in Aqueous Solution, *J. Phys. Chem. Ref. Data*, 17(2), 513–886, <https://doi.org/10.1063/1.555805>, 1988.
- 485 Chang, J. L. and Thompson, J. E.: Characterization of colored products formed during irradiation of aqueous solutions containing H₂O₂ and phenolic compounds, *Atmos. Environ.*, 44(4), 541–551, <https://doi.org/10.1016/j.atmosenv.2009.10.042>, 2010.
- Coeur-Tourneur, C., Cassez, A. and Wenger, J. C.: Rate coefficients for the gas-phase reaction of hydroxyl radicals with 2-methoxyphenol (guaiacol) and related compounds., *J. Phys. Chem. A*, 114(43), 11645–11650, <https://doi.org/10.1021/jp1071023>, 2010.
- Di Lorenzo, R. A., Washenfelder, R. A., Attwood, A. R., Guo, H., Xu, L., Ng, N. L., Weber, R. J., Baumann, K., Edgerton, E. and Young, C. J.: Molecular-Size-Separated Brown Carbon Absorption for Biomass-Burning Aerosol at Multiple Field Sites., *Environ. Sci. Technol.*, 51(6), 3128–3137, <https://doi.org/10.1021/acs.est.6b06160>, 2017.
- 490 Dulo, B., Phan, K., Githaiga, J., Raes, K. and De Meester, S.: Natural quinone dyes: A review on structure, extraction techniques, analysis and application potential, *Waste Biomass Valor.*, 12(12), 6339–6374, <https://doi.org/10.1007/s12649-021-01443-9>, 2021.
- 495 Fleming, L. T., Lin, P., Roberts, J. M., Selimovic, V., Yokelson, R., Laskin, J., Laskin, A. and Nizkorodov, S. A.: Molecular composition and photochemical lifetimes of brown carbon chromophores in biomass burning organic aerosol, *Atmospheric Chemistry and Physics*, 20(2), 1105–1129, <https://doi.org/10.5194/acp-20-1105-2020>, 2020.
- Garofalo, L. A., Pothier, M. A., Levin, E. J. T., Campos, T., Kreidenweis, S. M. and Farmer, D. K.: Emission and Evolution of Submicron Organic Aerosol in Smoke from Wildfires in the Western United States, *ACS Earth Space Chem.*, 3(7), 1237–1247, <https://doi.org/10.1021/acsearthspacechem.9b00125>, 2019.
- 500 Graber, E. R. and Rudich, Y.: Atmospheric HULIS: How humic-like are they? A comprehensive and critical review, *Atmospheric Chemistry and Physics*, 6(3), 729–753, <https://doi.org/10.5194/acp-6-729-2006>, 2006.

- Grieshop, A. P., Logue, J. M., Donahue, N. M. and Robinson, A. L.: Laboratory investigation of photochemical oxidation of organic aerosol from wood fires 1: measurement and simulation of organic aerosol evolution, *Atmospheric Chemistry and Physics*, 9(4), 1263–1277, <https://doi.org/10.5194/acp-9-1263-2009>, 2009.
- 505 Gurol, M. D. and Nekouinaini, S.: Kinetic behavior of ozone in aqueous solutions of substituted phenols, *Ind. Eng. Chem. Fund.*, 23(1), 54–60, <https://doi.org/10.1021/i100013a011>, 1984.
- Hallquist, M., Wenger, J. C., Baltensperger, U., Rudich, Y., Simpson, D., Claeys, M., Dommen, J., Donahue, N. M., George, C., Goldstein, A. H., Hamilton, J. F., Herrmann, H., Hoffmann, T., Iinuma, Y., Jang, M., Jenkin, M. E., Jimenez, J. L., Kiendler-Scharr, A., Maenhaut, W., McFiggans, G., Mentel, Th. F., Monod, A., Prévôt, A. S. H., Seinfeld, J. H., Surratt, J. D., Szmigielski, R. and Wildt, J.: The formation, properties and impact of secondary organic aerosol: current and emerging issues, *Atmospheric Chemistry and Physics*, 9(14), 5155–5236, <https://doi.org/10.5194/acp-9-5155-2009>, 2009.
- 510 Heald, C. L., Henze, D. K., Horowitz, L. W., Feddes, J., Lamarque, J. F., Guenther, A., Hess, P. G., Vitt, F., Seinfeld, J. H., Goldstein, A. H. and Fung, I.: Predicted change in global secondary organic aerosol concentrations in response to future climate, emissions, and land use change, *J. Geophys. Res.*, 113(D5), n/a-n/a, <https://doi.org/10.1029/2007JD009092>, 2008.
- 515 Hecobian, A., Zhang, X., Zheng, M., Frank, N., Edgerton, E. S. and Weber, R. J.: Water-Soluble Organic Aerosol material and the light-absorption characteristics of aqueous extracts measured over the Southeastern United States, *Atmospheric Chemistry and Physics*, 10(13), 5965–5977, <https://doi.org/10.5194/acp-10-5965-2010>, 2010.
- Hems, R. F., Schnitzler, E. G., Bastawrous, M., Soong, R., Simpson, A. J. and Abbatt, J. P. D.: Aqueous photoreactions of wood smoke brown carbon, *ACS Earth Space Chem.*, <https://doi.org/10.1021/acsearthspacechem.0c00117>, 2020.
- 520 Hems, R. F., Schnitzler, E. G., Liu-Kang, C., Cappa, C. D. and Abbatt, J. P. D.: Aging of atmospheric brown carbon aerosol, *ACS Earth Space Chem.*, <https://doi.org/10.1021/acsearthspacechem.0c00346>, 2021.
- Hennigan, C. J., Miracolo, M. A., Engelhart, G. J., May, A. A., Presto, A. A., Lee, T., Sullivan, A. P., McMeeking, G. R., Coe, H., Wold, C. E., Hao, W. M., Gilman, J. B., Kuster, W. C., de Gouw, J., Schichtel, B. A., Collett, J. L., Kreidenweis, S. M. and Robinson, A. L.: Chemical and physical transformations of organic aerosol from the photo-oxidation of open biomass burning emissions in an environmental chamber, *Atmospheric Chemistry and Physics*, 11(15), 7669–7686, <https://doi.org/10.5194/acp-11-7669-2011>, 2011.
- 525 Huang, D. D., Zhang, Q., Cheung, H. H. Y., Yu, L., Zhou, S., Anastasio, C., Smith, J. D. and Chan, C. K.: Formation and Evolution of aqSOA from Aqueous-Phase Reactions of Phenolic Carbonyls: Comparison between Ammonium Sulfate and Ammonium Nitrate Solutions., *Environ. Sci. Technol.*, 52(16), 9215–9224, <https://doi.org/10.1021/acs.est.8b03441>, 2018.
- 530 Jiang, W., Misovich, M. V., Hettiyadura, A. P. S., Laskin, A., McFall, A. S., Anastasio, C. and Zhang, Q.: Photosensitized Reactions of a Phenolic Carbonyl from Wood Combustion in the Aqueous Phase-Chemical Evolution and Light Absorption Properties of AqSOA., *Environ. Sci. Technol.*, 55(8), 5199–5211, <https://doi.org/10.1021/acs.est.0c07581>, 2021.
- Jiang, W., Ma, L., Niedeck, C., Anastasio, C. and Zhang, Q.: Chemical and Light-Absorption Properties of Water-Soluble Organic Aerosols in Northern California and Photooxidant Production by Brown Carbon Components., *ACS Earth Space Chem.*, 7(5), 1107–1119, <https://doi.org/10.1021/acsearthspacechem.3c00022>, [20232023a](https://doi.org/10.1021/acsearthspacechem.3c00022).
- 535 [20232023a](https://doi.org/10.1021/acsearthspacechem.3c00022), [20232023b](https://doi.org/10.1021/acsearthspacechem.3c00022).
- <https://doi.org/10.5194/acp-23-7103-2023>, [2023b](https://doi.org/10.5194/acp-23-7103-2023).
- 540 Jimenez, J. L., Canagaratna, M. R., Donahue, N. M., Prevot, A. S. H., Zhang, Q., Kroll, J. H., DeCarlo, P. F., Allan, J. D., Coe, H., Ng, N. L., Aiken, A. C., Docherty, K. S., Ulbrich, I. M., Grieshop, A. P., Robinson, A. L., Duplissy, J., Smith, J. D., Wilson, K. R., Lanz, V. A., Hueglin, C., Sun, Y. L., Tian, J., Laaksonen, A., Raatikainen, T., Rautiainen, J., Vaattovaara, P., Ehn, M., Kulmala, M., Tomlinson, J. M., Collins, D. R., Cubison, M. J., Dunlea, E. J., Huffman, J. A., Onasch, T. B.,

- Alfarra, M. R., Williams, P. I., Bower, K., Kondo, Y., Schneider, J., Drewnick, F., Borrmann, S., Weimer, S., Demerjian, K., Salcedo, D., Cottrell, L., Griffin, R., Takami, A., Miyoshi, T., Hatakeyama, S., Shimono, A., Sun, J. Y., Zhang, Y. M., Dzepina, K., Kimmel, J. R., Sueper, D., Jayne, J. T., Herndon, S. C., Trimborn, A. M., Williams, L. R., Wood, E. C., Middlebrook, A. M., Kolb, C. E., Baltensperger, U. and Worsnop, D. R.: Evolution of organic aerosols in the atmosphere., *Science*, 326(5959), 1525–1529, <https://doi.org/10.1126/science.1180353>, 2009.
- 545 Jo, D. S., Park, R. J., Lee, S., Kim, S.-W. and Zhang, X.: A global simulation of brown carbon: implications for photochemistry and direct radiative effect, *Atmospheric Chemistry and Physics*, 16(5), 3413–3432, <https://doi.org/10.5194/acp-16-3413-2016>, 2016.
- 550 Kaeswurm, J. A. H., Scharinger, A., Teipel, J. and Buchweitz, M.: Absorption coefficients of phenolic structures in different solvents routinely used for experiments., *Molecules*, 26(15), <https://doi.org/10.3390/molecules26154656>, 2021.
- Kanakidou, M., Seinfeld, J. H., Pandis, S. N., Barnes, I., Dentener, F. J., Facchini, M. C., Van Dingenen, R., Ervens, B., Nenes, A., Nielsen, C. J., Swietlicki, E., Putaud, J. P., Balkanski, Y., Fuzzi, S., Horth, J., Moortgat, G. K., Winterhalter, R., Myhre, C. E. L., Tsigaridis, K., Vignati, E., Stephanou, E. G. and Wilson, J.: Organic aerosol and global climate modelling: a review, *Atmospheric Chemistry and Physics*, 5(4), 1053–1123, <https://doi.org/10.5194/acp-5-1053-2005>, 2005.
- 555 [Kaur, R., Hudson, B. M., Draper, J., Tantillo, D. J. and Anastasio, C.: Aqueous Reactions of Organic Triplet Excited States with Atmospheric Alkenes, *Atmos. Chem. Phys. Discuss.*, 1–29, <https://doi.org/10.5194/acp-2018-1259>, 2018.](https://doi.org/10.5194/acp-2018-1259)
- 560 Kaur, R., Labins, J. R., Helbock, S. S., Jiang, W., Bein, K. J., Zhang, Q. and Anastasio, C.: Photooxidants from brown carbon and other chromophores in illuminated particle extracts, *Atmospheric Chemistry and Physics*, 19(9), 6579–6594, <https://doi.org/10.5194/acp-19-6579-2019>, 2019.
- Kim, K.-H., Kabir, E. and Kabir, S.: A review on the human health impact of airborne particulate matter., *Environ. Int.*, 74, 136–143, <https://doi.org/10.1016/j.envint.2014.10.005>, 2015.
- 565 Kirchstetter, T. W. and Thatcher, T. L.: Contribution of organic carbon to wood smoke particulate matter absorption of solar radiation, *Atmospheric Chemistry and Physics*, 12(14), 6067–6072, <https://doi.org/10.5194/acp-12-6067-2012>, 2012.
- Kleinman, L. I., Sedlacek III, A. J., Adachi, K., Buseck, P. R., Collier, S., Dubey, M. K., Hodshire, A. L., Lewis, E., Onasch, T. B., Pierce, J. R., Shilling, J., Springston, S. R., Wang, J., Zhang, Q., Zhou, S. and Yokelson, R. J.: Rapid evolution of aerosol particles and their optical properties downwind of wildfires in the western US, *Atmospheric Chemistry and Physics*, 20(21), 13319–13341, <https://doi.org/10.5194/acp-20-13319-2020>, 2020.
- 570 Laskin, A., Laskin, J. and Nizkorodov, S. A.: Chemistry of atmospheric brown carbon., *Chem. Rev.*, 115(10), 4335–4382, <https://doi.org/10.1021/cr5006167>, 2015.
- Liu, D., Li, S., Hu, D., Kong, S., Cheng, Y., Wu, Y., Ding, S., Hu, K., Zheng, S., Yan, Q., Zheng, H., Zhao, D., Tian, P., Ye, J., Huang, M. and Ding, D.: Evolution of Aerosol Optical Properties from Wood Smoke in Real Atmosphere Influenced by Burning Phase and Solar Radiation., *Environ. Sci. Technol.*, 55(9), 5677–5688, <https://doi.org/10.1021/acs.est.0c07569>, 2021.
- 575 [Li, X., Tao, Y., Zhu, L., Ma, S., Luo, S., Zhao, Z., Sun, N., Ge, X. and Ye, Z.: Optical and chemical properties and oxidative potential of aqueous-phase products from OH and ³C-initiated photooxidation of eugenol, *Atmospheric Chemistry and Physics*, 22\(11\), 7793–7814, <https://doi.org/10.5194/acp-22-7793-2022>, 2022.](https://doi.org/10.5194/acp-22-7793-2022)
- 580 Li, Y. J., Huang, D. D., Cheung, H. Y., Lee, A. K. Y. and Chan, C. K.: Aqueous-phase photochemical oxidation and direct photolysis of vanillin – a model compound of methoxy phenols from biomass burning, *Atmospheric Chemistry and Physics*, 14(6), 2871–2885, <https://doi.org/10.5194/acp-14-2871-2014>, 2014.
- Ma, L., Guzman, C., Niedeck, C., Tran, T., Zhang, Q. and Anastasio, C.: Kinetics and mass yields of aqueous secondary organic aerosol from highly substituted phenols reacting with a triplet excited state, *Environ. Sci. Technol.*, 55(9), 5772–5781, <https://doi.org/10.1021/acs.est.1c00575>, 2021.

- 585 Ma, L., Worland, R., Jiang, W., Niedeck, C., Guzman, C., Bein, K. J., Zhang, Q. and Anastasio, C.: Predicting photooxidant concentrations in aerosol liquid water based on laboratory extracts of ambient particles, *Atmospheric Chemistry and Physics*, 23(15), 8805–8821, <https://doi.org/10.5194/acp-23-8805-2023>, 2023a2023.
- 590 [Ma, L., Worland, R., Heinlein, L., Guzman, C., Jiang, W., Niedeck, C., Bein, K. J., Zhang, Q. and Anastasio, C.: Seasonal variations in photooxidant formation and light absorption in aqueous extracts of ambient particles, *egusphere*, 2023b, <https://doi.org/10.5194/egusphere-2023-861>, 2023b.](https://doi.org/10.5194/egusphere-2023-861)
- [Ma, L., Worland, R., Heinlein, L., Guzman, C., Jiang, W., Niedeck, C., Bein, K. J., Zhang, Q. and Anastasio, C.: Seasonal variations in photooxidant formation and light absorption in aqueous extracts of ambient particles, *Atmospheric Chemistry and Physics*, 24\(1\), 1–21, <https://doi.org/10.5194/acp-24-1-2024>, 2024.](https://doi.org/10.5194/acp-24-1-2024)
- 595 McFall, A. S., Johnson, A. W. and Anastasio, C.: Air-Water Partitioning of Biomass-Burning Phenols and the Effects of Temperature and Salinity., *Environ. Sci. Technol.*, 54(7), 3823–3830, <https://doi.org/10.1021/acs.est.9b06443>, 2020.
- [Misovich, M. V., Hettiyadura, A. P. S., Jiang, W., Zhang, Q. and Laskin, A.: Molecular-Level Study of the Photo-Oxidation of Aqueous-Phase Guaiacyl Acetone in the Presence of \$^{\bullet}C^*\$: Formation of Brown Carbon Products, *ACS Earth Space Chem.*, <https://doi.org/10.1021/acsearthspacechem.1c00103>, 2021.](https://doi.org/10.1021/acsearthspacechem.1c00103)
- 600 O'Dell, K., Bilsback, K., Ford, B., Martenies, S. E., Magzamen, S., Fischer, E. V. and Pierce, J. R.: Estimated mortality and morbidity attributable to smoke plumes in the united states: not just a western US problem., *GeoHealth*, 5(9), e2021GH000457, <https://doi.org/10.1029/2021GH000457>, 2021.
- Ortega, A. M., Day, D. A., Cubison, M. J., Brune, W. H., Bon, D., de Gouw, J. A. and Jimenez, J. L.: Secondary organic aerosol formation and primary organic aerosol oxidation from biomass-burning smoke in a flow reactor during FLAME-3, *Atmospheric Chemistry and Physics*, 13(22), 11551–11571, <https://doi.org/10.5194/acp-13-11551-2013>, 2013.
- 605 Palm, B. B., Peng, Q., Fredrickson, C. D., Lee, B. H., Garofalo, L. A., Pothier, M. A., Kreidenweis, S. M., Farmer, D. K., Pöhrl, R. P., Shen, Y., Murphy, S. M., Permar, W., Hu, L., Campos, T. L., Hall, S. R., Ullmann, K., Zhang, X., Flocke, F., Fischer, E. V. and Thornton, J. A.: Quantification of organic aerosol and brown carbon evolution in fresh wildfire plumes., *Proc Natl Acad Sci USA*, 117(47), 29469–29477, <https://doi.org/10.1073/pnas.2012218117>, 2020.
- 610 Powelson, M. H., Espelien, B. M., Hawkins, L. N., Galloway, M. M. and De Haan, D. O.: Brown carbon formation by aqueous-phase carbonyl compound reactions with amines and ammonium sulfate., *Environ. Sci. Technol.*, 48(2), 985–993, <https://doi.org/10.1021/es4038325>, 2014.
- Reid, J. S., Koppmann, R., Eck, T. F. and Eleuterio, D. P.: A review of biomass burning emissions part II: intensive physical properties of biomass burning particles, *Atmos. Chem. Phys.*, 27, 2005.
- 615 Robinson, A. L., Donahue, N. M., Shrivastava, M. K., Weitkamp, E. A., Sage, A. M., Grieshop, A. P., Lane, T. E., Pierce, J. R. and Pandis, S. N.: Rethinking organic aerosols: semivolatile emissions and photochemical aging., *Science*, 315(5816), 1259–1262, <https://doi.org/10.1126/science.1133061>, 2007.
- Saleh, R., Hennigan, C. J., McMeeking, G. R., Chuang, W. K., Robinson, E. S., Coe, H., Donahue, N. M. and Robinson, A. L.: Absorptivity of brown carbon in fresh and photo-chemically aged biomass-burning emissions, *Atmospheric Chemistry and Physics*, 13(15), 7683–7693, <https://doi.org/10.5194/acp-13-7683-2013>, 2013.
- 620 [Schauer, J. J., Kleeman, M. J., Cass, G. R. and Simoneit, B. R.: Measurement of emissions from air pollution sources. 3. C1-C29 organic compounds from fireplace combustion of wood., *Environ. Sci. Technol.*, 35\(9\), 1716–1728, <https://doi.org/10.1021/es001331e>, 2001.](https://doi.org/10.1021/es001331e)
- 625 Schnitzler, E. G., Gerrebos, N. G. A., Carter, T. S., Huang, Y., Heald, C. L., Bertram, A. K. and Abbatt, J. P. D.: Rate of atmospheric brown carbon whitening governed by environmental conditions., *Proc Natl Acad Sci USA*, 119(38), e2205610119, <https://doi.org/10.1073/pnas.2205610119>, 2022.

- Shrivastava, M., Cappa, C. D., Fan, J., Goldstein, A. H., Guenther, A. B., Jimenez, J. L., Kuang, C., Laskin, A., Martin, S. T., Ng, N. L., Petaja, T., Pierce, J. R., Rasch, P. J., Roldin, P., Seinfeld, J. H., Shilling, J., Smith, J. N., Thornton, J. A., Volkamer, R., Wang, J., Worsnop, D. R., Zaveri, R. A., Zelenyuk, A. and Zhang, Q.: Recent advances in understanding secondary organic aerosol: Implications for global climate forcing, *Rev. Geophys.*, 55(2), 509–559, <https://doi.org/10.1002/2016RG000540>, 2017.
- 630 Smith, J. D., Kinney, H. and Anastasio, C.: Aqueous benzene-diols react with an organic triplet excited state and hydroxyl radical to form secondary organic aerosol, *Phys. Chem. Chem. Phys.*, 17(15), 10227–10237, <https://doi.org/10.1039/c4cp06095d>, 2015.
- 635 Smith, J. D., Kinney, H. and Anastasio, C.: Phenolic carbonyls undergo rapid aqueous photodegradation to form low-volatility, light-absorbing products, *Atmos. Environ.*, 126, 36–44, <https://doi.org/10.1016/j.atmosenv.2015.11.035>, 2016.
- Sun, Y. L., Zhang, Q., Anastasio, C. and Sun, J.: Insights into secondary organic aerosol formed via aqueous-phase reactions of phenolic compounds based on high resolution mass spectrometry, *Atmospheric Chemistry and Physics*, 10(10), 4809–4822, <https://doi.org/10.5194/acp-10-4809-2010>, 2010.
- 640 Volkamer, R., Jimenez, J. L., San Martini, F., Dzepina, K., Zhang, Q., Salcedo, D., Molina, L. T., Worsnop, D. R. and Molina, M. J.: Secondary organic aerosol formation from anthropogenic air pollution: Rapid and higher than expected, *Geophys. Res. Lett.*, 33(17), <https://doi.org/10.1029/2006GL026899>, 2006.
- 645 Wang, Y., Huang, W., Tian, L., Wang, Y., Li, F., Huang, D. D., Zhang, R., Go Mabato, B. R., Huang, R.-J., Chen, Q., Ge, X., Du, L., Ma, Y. G., Gen, M., Hoi, K. I., Mok, K. M., Yu, J. Z., Chan, C. K., Li, X. and Li, Y. J.: Decay Kinetics and Absorption Changes of Methoxyphenols and Nitrophenols during Nitrate-Mediated Aqueous Photochemical Oxidation at 254 and 313 nm, *ACS Earth Space Chem.*, <https://doi.org/10.1021/acsearthspacechem.2c00021>, 2022.
- Yee, L. D., Kautzman, K. E., Loza, C. L., Schilling, K. A., Coggon, M. M., Chhabra, P. S., Chan, M. N., Chan, A. W. H., Hersey, S. P., Crounse, J. D., Wennberg, P. O., Flagan, R. C. and Seinfeld, J. H.: Secondary organic aerosol formation from biomass burning intermediates: phenol and methoxyphenols, *Atmospheric Chemistry and Physics*, 13(16), 8019–8043, <https://doi.org/10.5194/acp-13-8019-2013>, 2013.
- 650 Yu, L., Smith, J., Laskin, A., Anastasio, C., Laskin, J. and Zhang, Q.: Chemical characterization of SOA formed from aqueous-phase reactions of phenols with the triplet excited state of carbonyl and hydroxyl radical, *Atmospheric Chemistry and Physics*, 14(24), 13801–13816, <https://doi.org/10.5194/acp-14-13801-2014>, 2014.
- 655 Yu, L., Smith, J., Laskin, A., George, K. M., Anastasio, C., Laskin, J., Dillner, A. M. and Zhang, Q.: Molecular transformations of phenolic SOA during photochemical aging in the aqueous phase: competition among oligomerization, functionalization, and fragmentation, *Atmospheric Chemistry and Physics*, 16(7), 4511–4527, <https://doi.org/10.5194/acp-16-4511-2016>, 2016.
- Zhang, A., Wang, Y., Zhang, Y., Weber, R. J., Song, Y., Ke, Z. and Zou, Y.: Modeling the global radiative effect of brown carbon: a potentially larger heating source in the tropical free troposphere than black carbon, *Atmospheric Chemistry and Physics*, 20(4), 1901–1920, <https://doi.org/10.5194/acp-20-1901-2020>, 2020.
- 660 Zhao, R., Lee, A. K. Y., Huang, L., Li, X., Yang, F. and Abbatt, J. P. D.: Photochemical processing of aqueous atmospheric brown carbon, *Atmospheric Chemistry and Physics*, 15(11), 6087–6100, <https://doi.org/10.5194/acp-15-6087-2015>, 2015.
- Zhong, M. and Jang, M.: Dynamic light absorption of biomass-burning organic carbon photochemically aged under natural sunlight, *Atmospheric Chemistry and Physics*, 14(3), 1517–1525, <https://doi.org/10.5194/acp-14-1517-2014>, 2014.
- 665 Zhou, X., Josey, K., Kamareddine, L., Caine, M. C., Liu, T., Mickley, L. J., Cooper, M. and Dominici, F.: Excess of COVID-19 cases and deaths due to fine particulate matter exposure during the 2020 wildfires in the United States., *Sci. Adv.*, 7(33), <https://doi.org/10.1126/sciadv.abi8789>, 2021.