- 1 Using Eddy Covariance Observations to Determine the Carbon
- 2 Sequestration Characteristics of Subalpine Forests in the

3 Qinghai-Tibet Plateau

- 4 Niu Zhu^{1,2,4}, Jinniu Wang^{1,2}, Dongliang Luo³, Xufeng Wang³, Cheng Shen^{1,2}, Ning
 5 Wu¹
- 6 1 Chengdu Institute of Biology, Chinese Academy of Science, Chengdu 610041,
 7 China
- 8 2 Mangkang Ecological Monitoring Station, Tibet Ecological Security Barrier
- 9 Ecological Monitoring Network, Qamdo 854500, China
- 10 3 Northwest Institute of Eco-environmental Resources, Chinese Academy of
- 11 Sciences, Lanzhou 730000, China
- 12 4 College of Resources and Environmental Sciences, Gansu Agricultural University,
- 13 Lanzhou 730070, China
- 14 Correspondence: Jinniu Wang (wangjn@cib.ac.cn)

15 Abstract: The subalpine forests are one of the crucial components in the carbon cycling system in 16 the Qinghai-Tibet Plateau (QTP) in the context of climate change and ecosystem dynamics. The 17 subalpine forests in the Qinghai-Tibet Plateau (QTP) act as carbon sinks in the context of climate 18 change and ecosystem dynamics. In this study, we investigated the carbon exchange dynamics for 19 a subalpine forest on the QTP using the eddy covariance method from November 2020 to October 20 2022. We first revealed the seasonal characteristics of carbon dynamics in the subalpine forest, 21 revealing the pattern of higher rates in summer and autumn and lower rates in winter and spring, 22 and found that autumn is the peak period for carbon sequestration in the subalpine forest. Subsequently, we explored the climatic factors influencing the carbon sequestration function. The 23 24 PCA analysis show that photosynthetically active radiation (PAR) was major climatic factor 25 driving the net ecosystem exchange (NEE), significantly influencing forest and carbon absorption. 26 The spatial distribution of NEE was significantly positively correlated with temperature, while the 27 average annual precipitation shows a minor effect on NEE at the regional scale. At the annual 28 scale, the subalpine forest was a strong carbon sink, with an average NEE of -342 g C m⁻² (from 29 November 2020 to October 2022). Despite the challenges caused by climate change, forests 30 remain a robust carbon sink, currently, they are the ecosystems with the highest carbon 31 sequestration capacity in the QTP, with an average annual CO_2 absorption rate of 368 gC m⁻². this 32 study provides essential insights for understanding the carbon cycling mechanism in plateau 33 ecosystems and the global carbon balance. We propose that, to positively influence global carbon 34 cycling and promote "carbon neutrality and peak carbon," strengthening the protection and 35 management of subalpine forests is crucial. Although our research has shown that these forest is 36 currently playing a role in continuous carbon absorption, there are significant data gaps on the 37 Qinghai-Tibetan Plateau. Therefore, it is essential to enhance continuous monitoring of forest 38 carbon absorption processes in the future. In this study, we investigated the carbon sequestration 39 function using the *in-situ* observations from an eddy covariance system for the subalpine forests. 40 With two-year contiguous observations, the factors driving the seasonal variations in carbon 41 sequestration potential were quantified. We first revealed the seasonal characteristics of carbon-42 dynamics in the subalpine forests during the growing and dormant seasons, respectively. The 43 diurnal carbon exchange exhibited significant fluctuations, as high as 10.78 µmol CO₂-s⁺-m⁻² 44 (12:30, autumn). The period from summer to autumn was identified as the peak in carbon sequestration rate in the subalpine forests. Subsequently, we explored the climatic factors-45 46 influencing the carbon sequestration function. Photosynthetically active radiation (PAR) wasfound to be a major climatic factor driving the net ecosystem exchange (NEE) within the same-47 season, significantly influencing forest growth and carbon absorption. Increasing altitude 48 49 negatively impacts carbon absorption at the regional scale and the rising annual temperature significantly enhances carbon uptake, while the average annual precipitation shows a minor effect 50 51 on NEE. At the annual scale, the observations at the subalpine forests demonstrated a strong 52 carbon sequestration capability, with an average NEE of 389.03 g C m². Furthermore, we roughly assessed the carbon sequestration status of subalpine forests in the OTP. Despite challenges caused 53 by climate change, these forests possess enormous carbon sequestration potential. Currently, they 54 55 represent the most robust carbon sequestration ecosystem in the QTP. We conclude that enhancing 56 the protection and management of subalpine forests under future climate change scenarios will-57 positively impact global carbon cycling and contribute to climate change mitigation. Moreover, 58 this study provides essential insights for understanding the carbon cycling mechanism in plateau-59 ecosystems and global carbon balance.

Keywords: Subalpine forest; Qinghai-Tibet Plateau; The eddy covariance system; Three Parallel
Rivers Region; Carbon sinks

62 **1 Introduction**

63 Carbon dioxide (CO_2) is a prominent greenhouse gas, and its atmospheric concentration has 64 reached an unprecedented high level in recent years, in May 2021, a recorded peak of 419 parts 65 per million (ppm) was observed at the Mauna Loa Observatory in Hawaii (Stein, 2021).with a 66 recorded peak of 419 parts per million (ppm). The global atmospheric CO_2 concentration is 67 rapidly increasing at a rate of 2 to 3 ppm per year, compared to pre-industrial levels, the average global temperature has already risen by 1.1°C by 2019 (World Meteorological Organization, 68 69 2019). Human activities have been the primary catalyst behind the significant surge in 70 atmospheric CO₂ concentrations (Schweizer et al., 2020). Extensive research conducted by 71 numerous scholars has consistently demonstrated that human activities have been the primary 72 catalyst behind the significant surge in atmospheric CO₂ concentrations since the 18th century

73 (Stein, 2021). CO_2 and CH_4 collectively contribute approximately 70% to the global warming 74 potential among the six greenhouse gases specified in the Kyoto Protocol (Zhang et al., 2022). As 75 atmospheric CO₂ concentrations continue to rise, global climate warming is gradually intensifying. 76 Therefore, The Paris Agreement urges national governments to restrict the increase in global average temperature to well below 2.0 °C above pre-industrial levels and to strive to limit it to 77 78 1.5 °C. The increasing atmospheric CO₂ levels will lead to irreversible ecological disasters. For 79 instance, the concentration of CO₂ in the atmosphere is projected to double within approximately 80 50 years if global consumption of fossil fuels continues to rise at the current rate, the 81 concentration of CO_2 in the atmosphere is projected to double within approximately 50 years. The 82 rise in temperatures at 80°S latitude could result in the melting of glaciers, leading to a sea-level rise of 5 m (Mercer, 1978). By the year 2040, most countries are projected to experience at least 83 84 one annual disaster with a 50% or higher probability (Fortunato et al., 2022). Addressing the 85 greenhouse effect caused by carbon dioxide and reducing its impact is a crucial challenge facing 86 human society today. Reducing regional carbon emissions or per capita carbon emissions is 87 widely regarded as an effective approach to carbon reduction (Wang et al., 2023a). Nevertheless, 88 countries around the world have already begun to commit to carbon reduction and carbon 89 neutrality efforts. On September 22, 2020, during the 75th session of the United Nations General 90 Assembly, the Chinese government announced "double carbon" goals, which aim to achieve 91 carbon emission peaking by 2030 and carbon neutrality by 2060, in alignment with ecological 92 conservation and sustainable development objectives (Yu, 2022). It is predicted that China's 93 average forest carbon sequestration rate will reach 0.358 Pg C year-1 (petagrams of carbon per 94 year) by 2060 (Cai et al., 2022). This significant rate of carbon sequestration is expected to have a 95 substantial impact on the environment and economy, providing negative feedback to global 96 warming (Pan et al., 2011).

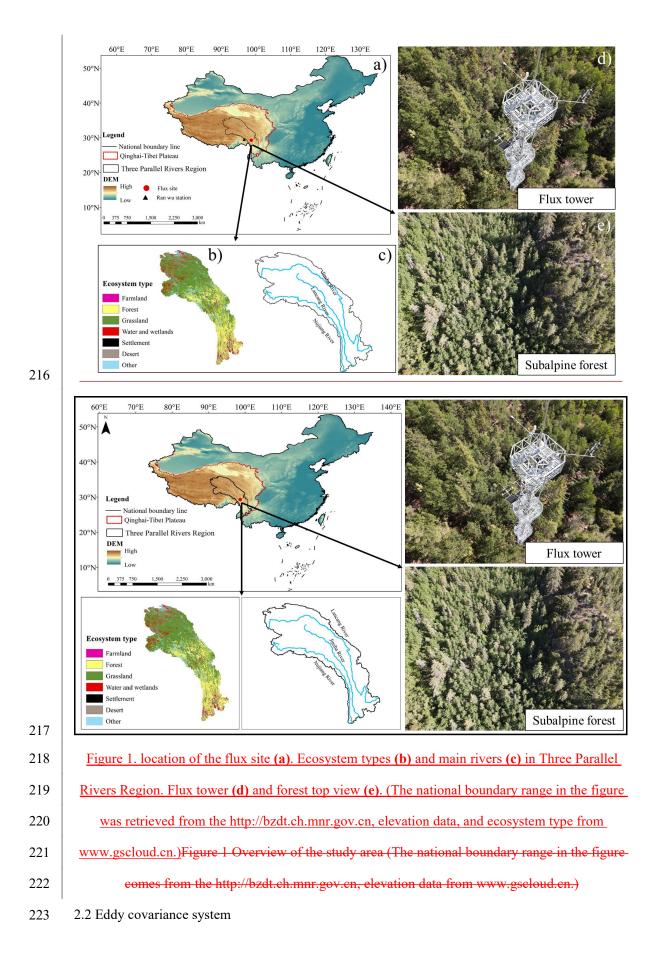
97 <u>Currently, there are various methods available to accurately quantify the carbon sequestration</u> 98 potential of forests, each with its advantages and disadvantages. Forests cover approximately 30%-99 of the earth's land surface and store around 90% of the terrestrial vegetation carbon (Le Quéré et 100 al., 2018). However, currently, there is no method available to accurately quantify the carbon 101 sequestration potential of forests. Quantitative estimation of carbon sequestration potential still 102 requires scientists to establish more *in-situ* sites and generate comprehensive datasets to assess a 103 wide range of areas. Initially, individuals' biomass measurements were used to estimate forest 104 carbon sequestration capacity (Ebermayer, 1876). However, this method was time-consuming, 105 labor-intensive, and prone to inaccuracies due to the omission of various variables during the 106 calculation process. The development of modeling techniques allowed for the use of simulation 107 methods - forest management models and land ecosystem-climate interaction models, such as the 108 Ecological Assimilation of Land and Climate Observation (EALCO), have been widely applied in 109 this regard (Landsberg and Waring, 1997; Wang et al., 2001). Currently, remote sensing 110 monitoring and the eddy covariance method are widely used. Remote sensing techniques can be 111 used to extract vegetation parameters (such as NDVI) from multispectral bands and estimate the 112 carbon sequestration of entire forests through regression analysis (Laurin et al., 2014). The eddy 113 covariance (EC) method, allowing continuous, long-term carbon flux calculation, provides fundamental data for model establishment and calibration. It is widely applied across ecosystems, 114 115 including urban areas, farmlands, grasslands, forests, and water bodies (Konopka et al., 2021; 116 Vote et al., 2015; Du et al., 2022a; Kondo et al., 2017; Li et al., 2022). The theoretical foundation-117 of the eddy covariance method was initially proposed by Swibank et al., (Swinbank, 1951). It 118 started to be applied in carbon flux studies of forest ecosystems in the 1980s (Anderson et al., 119 1984). Nowadays, this method not only accurately measures the carbon exchange between forests-120 and the atmosphere but also integrates other instruments to measure meteorological variables such 121 as light intensity and temperature. It allows for long term and continuous calculation of carbon flux between forests and the atmosphere. Additionally, it provides fundamental data for 122 123 establishing and calibrating other models. The eddy covariance method has been widely applied in-124 various ecosystems, including urban areas (Konopka et al., 2021), farmlands (Vote et al., 2015), 125 grasslands (Du et al., 2022a), forests (Kondo et al., 2017), and water bodies (Li et al., 2022). 126 The forest ecosystem's Net ecosystem exchange (NEE) of carbon dioxide is influenced by 127 multiple environmental factorsNet Ecosystem Exchange (NEE) of carbon dioxide is a 128 fundamental parameter in the biogeochemical feedback of the climate system (Graf et al., 2013). 129 The carbon flux in forest ecosystems is influenced by multiple environmental factors. Previous 130 studies have shown that NEE is significantly influenced by air temperature (AT),

photosynthetically active radiation (PAR), vapor pressure deficit (VPD), relative humidity (RH), 131 132 and soil temperature (ST) (Liu et al., 2022). For instance, temperature variables, especially annual 133 or seasonal average temperature variations, serve as the optimal single predictor for carbon flux, 134 explaining variations in carbon flux between 19% and 71% (Banbury Morgan et al., 2021). 135 Photosynthetically active radiation not only influences the absorption of carbon dioxide by the 136 forest canopy but also affects the utilization of carbohydrates by roots due to its association with canopy processes and soil respiration (Baumgartner et al., 2020). Furthermore, research suggests 137 138 that the NEE is influenced by biotic factors such as NDVI (Normalized Difference Vegetation 139 Index) and LAI (Leaf Area Index) (Tang et al., 2022). Given the projected future global warming 140 trends, the role of forests as a vast carbon reservoir becomes highly significant and worthy of 141 attention. The Qinghai-Tibet Plateau (QTP) is the highest and largest plateau in the world, with an 142 extensive area of alpine forests covering approximately 2.3×10^5 km². These forests hold tremendous economic and ecological benefits. The southeastern region of the QTP boasts one of 143 144 the world's highest-altitude alpine forest ecosystems. Research indicates that the alpine forest 145 ecosystem in this area has a remarkable capacity to consume methane, reaching up to 5.06 kg ha⁻¹ 146 yr⁻¹, and playing a significant role in mitigating the impact of greenhouse gases (Qu et al., 2023). 147 Since the 1960s, the QTP has experienced a faster warming rate than lowland areas. It is projected that this phenomenon will be intensified by the end of the 21st century (Li et al., 2019). Currently, 148 149 the QTP is considered a weak carbon sink at the overall level, but the carbon source-sink 150 dynamics vary among different ecosystems (Chen et al., 2022). For instance, most lakes in the 151 QTP are currently characterized by supersaturated CO₂ levels (Cole et al., 1994). Mu et al. (2023) 152 found that the thermokarst lakes serve as significant carbon sources through carbon flux 153 measurements in 163 thermokarst lakes during the summer and autumn seasons. Wang et al. (2021) 154 discovered that these ecosystems act as sinks for carbon dioxide by comparing carbon fluxes in 155 ten high-mountain ecosystems with different grassland types. Wang et al. (2021), discovered that 156 these ecosystems act as sinks for carbon dioxide in their study comparing carbon fluxes in ten 157 high-mountain ecosystems with different grassland types. The alpine meadows in the eastern QTP 158 were identified as strong carbon sinks, with the highest annual average NEE recorded at -284 g C 159 m⁻². Forest ecosystems play a crucial role in the south-eastern edge of the QTP, providing

160 important support for climate regulation and forestry-based economic activities. Moreover, recent 161 predictive studies suggest that under both current and future climate scenarios, the forested area in 162 this region is expected to expand further, with coniferous forests continuing to grow into higher altitudes (Liu et al., 2021). Due to the extensive presence of permafrost in the QTP, forest net 163 164 primary productivity exhibits a most pronounced response to surface temperatures in the continuous permafrost zone over multiple years. Therefore, the changes in permafrost in the QTP 165 should not be overlooked, as they also have a significant impact on carbon absorption by forests 166 167 (Mao et al., 2015). However, the QTP is a vast region with a widespread distribution of high-altitude and subalpine forests. Researchers need to conduct long-term monitoring to 168 169 understand how these forests will respond to climate change. Furthermore, there is a significant data gap concerning the monitoring of carbon exchange capacity in the forests of the QTP, 170 171 indicating the need for further data collection efforts. Based on this, we have established a carbon 172 flux monitoring site in the subalpine ecosystem of the Three Parallel Rivers Region, which is 173 located on the south-eastern edge of the QTP and lies in the transitional zone between the QTP 174 and the Yunnan-Kweichow Plateau and is renowned as a global hotspot for biodiversity (Wang et 175 al., 2022). Our research objectives are as follows: Determine whether the subalpine forests in the Three Parallel Rivers Region act as a carbon 176 1) sink or source, and quantify the annual uptake or release of carbon dioxide; 177 178 2) Investigate the main environmental factors influencing the carbon exchange process in the 179 subalpine forests and identify the factors with the greatest impact; 180 3) Since the carbon sink potential of forest ecosystems in the QTP is currently unknown, we 181 evaluated the carbon exchange capacity of subalpine forests by comparing existing data with 182 other ecosystems in the QTP.Assess the carbon exchange capacity of the subalpine forests in 183 comparison to other ecosystems of the QTP.

- 184 This study will provide a data foundation and background support for accurately estimating
- 185 the carbon balance of forests in high-altitude areas and for model simulations in the future.
- 186 2 Materials and Methods
- 187 2.1 Overview of the study site

188 The study site is located in the Hongla Mountain Yunnan Snub-nosed Monkey National 189 Nature Reserve in Mangkang County, Tibet, China (29.28633°N, 98.69096°E), the core area of 190 the Three Parallel Rivers (Nujiang River, Lancang River, and Jinsha River) Region. The elevation 191 of the study site is 3755 m. The observation period was from November 2020 to October 2022. 192 The study area experiences large diurnal temperature variations and dry conditions in winter, while the summers are warm and humid. The climate of the region is characterized as a typical 193 194 mountainous climate. The average daily sunshine duration exceeds 10 h, with an annual average 195 temperature of 5 °C and an average annual precipitation of around 600 mm (Niu et al., 2023). The 196 main tree species in the area include Picea likiangensis var. rubescens, Abies squamata, Sabina 197 tibetica Kom, and Abies ernestii. They are accompanied by the growth of some Quercus aquifolioides, Rhododendron lapponicum, and Potentilla fruticosa shrubs. The average height of 198 199 the trees is below 30 meters, and the forest is in a relatively active growth phase, reaching the state 200 of a mature forest. The vegetation coverage ranges from 70% to 80%. The vegetation coverage 201 ranges from 70% to 80%, indicating rich vegetation resources. The dominant soil type is 202 yellow-brown soil. The mountainous terrain contributes to distinct vertical climate characteristics 203 and significant variations in water and heat conditions. Characterized by numerous dry and hot 204 river valleys and widespread distribution of canyons, the climate in the study area exhibits a clear 205 impact from the southwest and southeast monsoons. The varying elevations give rise to diverse 206 ecosystems, transitioning from alpine forests to mountain shrubs, and above 4000 meters, high 207 alpine grasslands and meadows, forming a noticeable vegetation transition zone. The mountainous 208 topography results in evident vertical climate features and significant fluctuations in water and 209 heat conditions, with precipitation showing a pronouncedly uneven distribution throughout the 210 region (Zemin et al., 2023). The study site is located in the core area of the Three Parallel Rivers-211 (Nujiang River, Lancang River, and Jinsha River) Region. The area exhibits a complex and 212 diverse climatic environment influenced by the southwest and southeast monsoon. The 213 mountainous terrain contributes to distinct vertical climate characteristics and significant 214 variations in water and heat conditions. The region is characterized by numerous dry and hot river-215 valleys and widespread distribution of canyons.



The flux data in this study were collected from a 35 m-high tower located at the study 224 225 site. The flux data in this study were collected from a 35 m-high tower located at the study site. At 226 the top of the tower, a 3-D wind velocity (Wind Master, Gill, UK) and an open-path infrared CO₂/H₂O analyzer (LI-7500DS, Li-Cor, USA) were installed to measure CO₂ flux. The 227 228 instruments had a measurement frequency of 10 Hz. The instruments had a response frequency of 229 10 Hz. Additionally, micro-meteorological sensors were placed at different heights on the tower, 230 including sensors at 15 m for observing air temperature and humidity (HMP155A, Vaisala, 231 Finland), sensors at -5 cm for soil temperature (TEROS11, LI-Cor, USA), and sensors at 35 m for photosynthetically active radiation (LI-190R, LI-Cor, USA), including sensors for observing air 232 temperature and humidity (HMP155A, Vaisala, Finland), soil temperature (TEROS11, LI-Cor, 233 USA), and photosynthetically active radiation (LI-190R, LI-Cor, USA), among other 234 235 environmental variables. All data were recorded at 30-m intervals and stored in a SmartFlux 3 data 236 logger (Li-Cor, USA) for future download. 237 2.3 Data processing and quality control

238 When considering only the turbulent transport of matter and energy in the vertical direction, 239 the carbon dioxide flux (Fc) can be represented by the following equation (Yu and Sun, 2006; 240 Monteith et al., 1994):Turbulent transport is the primary form of gas exchange between the 241 near surface and the atmosphere. In the case of a homogeneous and flat underlying surface, 242 considering only the turbulent transport of substances in the vertical direction, the CO₂ flux *Fc*-243 (µmol m⁻² s⁻¹ or mg m⁻² s⁻¹) within the region can be calculated using the following.

244 $F_{c} = \overline{W' \text{CO}_{2}'} \quad (1)$

245 Where W' is the vertical component of 3-D wind speed fluctuations (m s⁻¹), and CO₂' represents 246 the fluctuations in measured CO₂ mole concentration. A positive Fc indicates carbon emissions, 247 while a negative value represents carbon uptake. Where W' represents the vertical component of 248 3-D wind speed fluctuations (m/s), CO₂' represents the fluctuations in measured CO₂ mole 249 concentration (µmol m⁻³), and the overline denotes the average value over a half-hour time period. 250 A positive value of Fc indicates carbon emissions from the underlying surface during the given 251 time interval, while a negative value represents carbon uptake. 252 The acquired 10 Hz raw data was processed and corrected using the EddyPro software 253 (EddyPro 7.06, Li-Cor, USA). The calibration process correction process involved outlier detection 254 for flux data, lag elimination, coordinate rotation (Jia et al., 2020), ultrasonic temperature correction (Schotanus et al., 1983), frequency correction (Moncrieff et al., 1997), and 255 256 Webb-Pearman-Leuning (WPL) correction (Leuning and King, 1992), After these controls, the 257 integrity of the effective FC raw valid data we obtained reached 92.95 %. We removed outliers 258 caused by environmental disturbances such as power outages, rain, snow, and dust particles that 259 interfered with the instrument. Due to the slope of the underlying surface being around 5 degrees, 260 we also corrected from non-uniform and non-flat surfaces using EddyPro for double coordinate 261 rotation (Cao et al., 2019). We also corrected errors resulting from non-uniform and non-flatunderlying surfaces (Cao et al., 2019). As a result, we obtained half-hourly flux data with 262 263 associated data quality indicators. To evaluate the turbulence steadiness, we employed the "0-1-2" 264 quality assessment method, which classified flux results into three quality levels: 0 for excellent 265 data quality, 1 for moderate data quality, and 2 for low data quality (Mauder and Foken, 2011; 266 Foken et al., 2005)(Mauder and Foken, 2011). We removed data points labeled with a quality level 267 of "2". We further eliminated flux data with negative values during nighttime since plants do not 268 perform photosynthesis at night. Additionally, we conducted spectral analysis to identify and 269 remove data points with values significantly deviating from normal. Finally, friction velocities (u*) 270 for each of the two years were determined separately using the method of moving point, and 271 deleted data recorded during nighttime when u* was less than 0.28 and 0.39 m s⁻¹ (Reichstein et al., 272 2005). After excluding outliers from the data, the data integrity is 72.67%. Tovi software (Tovi, 273 Li-Cor, USA) was used in the process. Finally, we utilized the friction velocity (U*) as a criterion 274 and deleted data recorded during nighttime when U* was less than 0.28 and 0.39 m s⁺ (Papale et 275 al., 2006). 276 When turbulence is weak, a portion of CO_2 is stored in the vegetation canopy and the 277 atmosphere below the measurement height. At this time, the NEE is calculated as (Zhang et al., 278 <u>2018):NEE of CO₂ can be represented by the following:</u>

279 $NEE = F_C + F_S (2)$

- 280 Where NEE represents the net ecosystem exchange of CO₂, F_C stands for the observed flux during
- 281 a specific period, F_S represents the CO₂ storage in the forest canopy, F_S is calculated as $(\Delta c/\Delta t)$ ·h,
- 282 where Δc is the difference in CO₂ concentration between two consecutive measurements, Δt is the
- 283 time interval between two consecutive measurements, and h is 35m.
- 284 We adopted the following formula as a gap-filling strategy for daytime NEE (NEE_{day})
- 285 concerning PAR, aiming to address missing values during the daytime (Falge et al., 2001):
- 286 Where NEE represents the net ecosystem exchange of CO₂, F_C stands for the observed flux during-
- 287 a specific time period, F_8 -represents the CO₂-storage in the forest canopy, which is assumed to be-288 zero in this case.
- We used the Michaelis-Menten model to fit the daytime NEE (NEE_{day}) with respect to PAR
 to fill in the missing values during the daytime (Falge et al., 2001):

291
$$NEE_{day} = \frac{\alpha \cdot PAR \cdot P_{max}}{\alpha \cdot PAR + P_{max}} - R_{day} \quad (3)$$

292 where: a (µmol CO2/µmol PAR) represents the apparent photosynthetic quantum efficiency, which 293 characterizes the maximum efficiency of converting light energy during photosynthesis. PAR 294 (μ mol m⁻² s⁻¹) is the photosynthetically active radiation, a measure of the amount of light energy available for photosynthesis. Pmax (µmol CO2 m-2 s-1) is the apparent maximum photosynthetic rate, 295 representing the maximum CO₂ uptake rate under optimal conditions. R_{dav} (µmol CO₂ m⁻² s⁻¹) is 296 297 the daytime dark respiration rate, which denotes the rate of CO_2 release during daylight hours. The 298 parameters $\underline{\alpha}_{,a}$, P_{max} , and R_{day} are obtained through the non-linear fitting of the Michaelis-Menten 299 model to the observed data.

300 During the nighttime, the NEE is modeled using an exponential function of <u>ecosystem</u> 301 <u>respiration</u> and soil temperature to fill in the missing values of NEE during the night 302 (NEE_{night}) (Lloyd and Taylor, 1994; Kato et al., 2006):

303 $NEE_{night} = a \cdot exp^{(bt)}$ (4)

The parameters a and b are estimated values for the exponential function used in modeling NEE_{night}. The variable t represents the soil temperature measured at the depth of 5 cm. <u>Origin 2023</u> (Originlab Corporation, USA) is the data processing software used for this analysis. For the missing data, interpolation was performed using Tovi software allows for data interpolation to fill

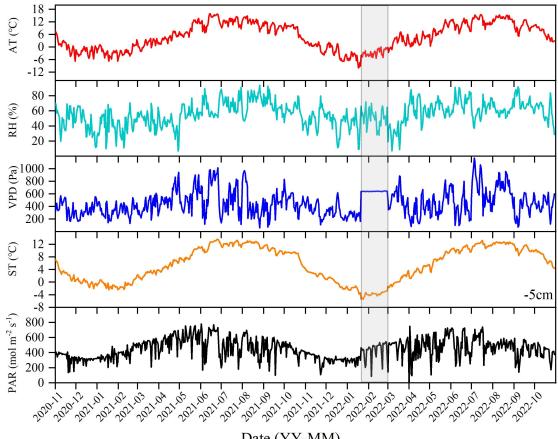
308	in the gapsThe data processing software used for this analysis is Origin 2023 (Originlab-
309	Corporation, USA). For the missing data, interpolation was performed using Tovi software (Tovi,
310	Li-Cor, USA) that allows for data interpolation to fill in the gaps and ensure a continuous dataset
311	for further analysis (Reichstein et al., 2005). 27.33% of missing data were interpolated, The final
312	flux data achieved a data integrity of 100%.
313	In flux analysis, the significance of source area contributions cannot be overlooked. In this
314	study, the peak distances of the 90% flux contribution areas averaged over two years are 364.2 and
315	357.1m, respectively. Looking at seasons, the average peak distances of the 90% flux contribution
316	areas for winter, spring, summer, and autumn over the two years are 353.9, 358.2, 350.05, and
317	344.34m, respectively.
318	2.4 <u>Flux partitioning</u> Flux splitting
319	Ecosystem respiration (RE) is the sum of plant and heterotrophic respiration in an ecosystem
320	and is obtained by adding the measured nighttime data to the extrapolated daytime data. Gross
321	primary productivity (GPP) is the total amount of organic carbon fixed by green plants through
322	photosynthesis per unit of time and per unit of area:
323	$RE=R_{day}+R_{night}(5)$
324	GPP=-NEE+RE (6)
325	Carbon use efficiency (CUE) is a crucial parameter that reflects the ability of an ecosystem to
326	sequester carbon. It is defined as the ratio of net ecosystem productivity (NEP) to gross primary
327	productivity. It is defined as the ratio of net primary productivity to gross primary productivity.
328	CUE can be expressed using the following equation:
329	$CUE = \frac{NEP}{GPP} = \frac{-NEE}{GPP} (7)$
330	To study the variation of ecosystem respiration rates with environmental
331	factorsenvironmental conditions, we considered the dependence of nocturnal ecosystem
332	respiration on soil temperature (Pavelka et al., 2007; Mamkin et al., 2023):
333	$Q_{10}=\exp(10\cdot\alpha)$ (8)
334	$\ln (\text{NEE}_{\text{night}}) = \alpha \cdot T + \gamma (9)$
335	Where T is the soil temperature (°C) and γ is an empirical parameter of the equation.
	13

To clarify the carbon sink potential of forests in the QTP and to compare it with other ecosystems, a search was conducted in two authoritative databases, Web of Science and China National Knowledge Internet, for research articles on the current utilization of EC systems in the QTP. A total of 82 research results were collected from 48 studies, and their annual average

- 340 <u>environmental factors, such as air temperature, precipitation, and altitude, were obtained.</u>
- **341 3 Results**

342 3.1 Daily average changes in main environmental factors

343 During the observational period, the environmental conditions exhibited significant fluctuations. The winter and spring seasons were characterized by cold and dry conditions, while 344 345 the summer and autumn seasons were warm and humid. The daily maximum air temperature (AT) recorded was 15.87 °C (on June 15, 2021), and the minimum temperature was -9.88 °C (on 346 347 January 17, 2022), with an average of 5.5 °C over the two years. The relative humidity (RH) with an annual average of 55.89%. The vapor pressure deficit (VPD) with an annual average of 4.46 348 349 hPa. Soil temperature (ST) exhibited a similar trend to air temperature. The highest observed soil 350 temperature was 13.53 °C (on June 27, 2021), while the minimum was -3.78 °C (on January 18, 351 2022), with an annual average of 6.11 °C. Photosynthetically active radiation (PAR) with an annual average of 447.24 mol m⁻² s⁻¹. The relative humidity (RH) ranged from a maximum of 352 93.98% (on August 26, 2021) to a minimum of 6.74% (on April 29, 2021), with an annual average-353 354 of 55.89%. The vapor pressure deficit (VPD), which represents the difference between the 355 saturated vapor pressure and the actual vapor pressure in the air, influences plant stomatal closure and regulates physiological processes such as transpiration and photosynthesis. The highest 356 357 recorded VPD was 1169.8 hPa (on July 5, 2022), and the lowest one was 60.8 hPa (on August 26, 358 2021), with an annual average of 446.4 hPa. Soil temperature (ST) exhibited a similar trend to air 359 temperature and remained relatively stable over short periods. The highest observed soil temperature was 13.53 °C (on June 27, 2021), while the minimum was -3.78 °C (on January 18, 360 361 2022), with an annual average of 6.11 °C. Photosynthetically active radiation (PAR) reached a maximum value of 779.06 mol m² s⁺ (on June 2, 2021), with an annual average of 447.24 mol 362 363 m²-s⁴. From March to October, the radiation conditions were favorable for photosynthesis, but 364 reduction in radiation intensity was observed during rainy, snowy, and cloudy weather conditions.



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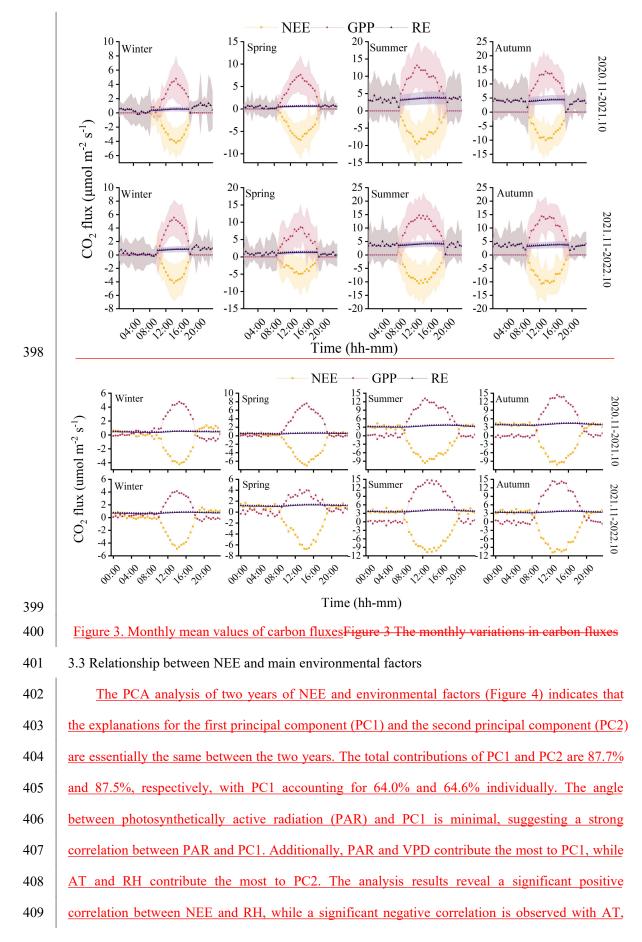
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366 Figure 2. Daily values of main environmental factors, air temperature (AT), relative humidity (RH), vapor pressure deficit (VPD), soil temperature (ST), and Photosynthetically active radiation 367 368 (PAR). (The data of the shadow part in the figure comes from the Ranwu forest site (Figure 1). 369 Since there was no interpolated data source for VPD, the annual average was used instead.)(Theshaded part of the figure represents the data interpolated by the nearby station) 370

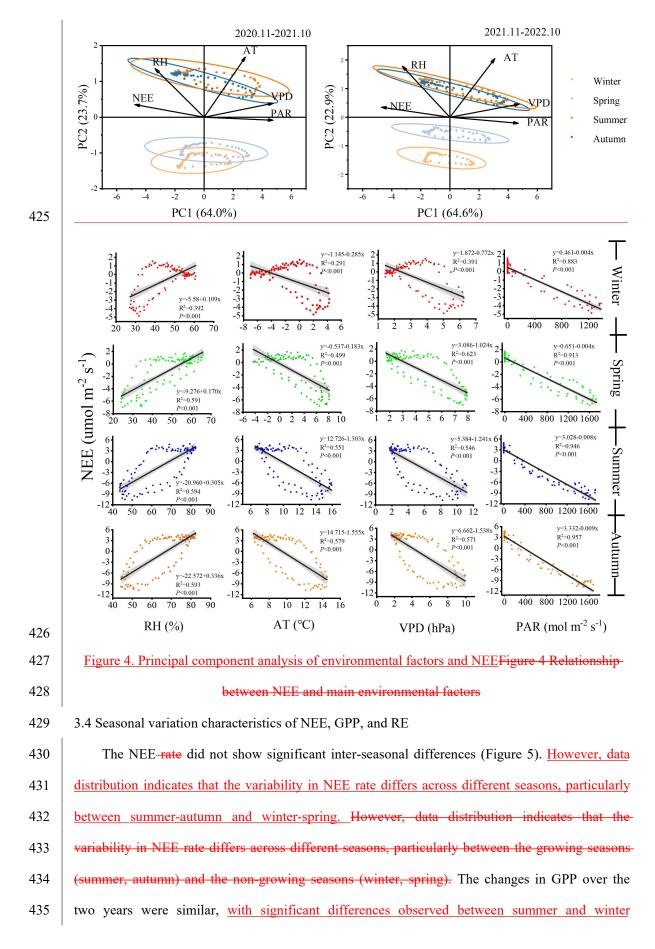
371 3.2 The seasonal variations in NEE, RE, and GPP

372 The observations from the forest ecosystem indicate distinct diurnal and seasonal variations in NEE and GPP. The NEE and GPP exhibit a pronounced U-shaped curve, with significant 373 374 seasonal differences. The summer and autumn are characterized by peak carbon uptake, with the 375 maximum NEE reaching 10.78 umol CO₂ m⁻² s⁺⁺ (12:30, autumn). During the nighttime, the 376 ecosystem generally releases carbon, while during favorable daytime meteorological conditions, it 377 demonstrates a carbon uptake capacity. The peak carbon absorption of the forest ecosystem occurs 378 from 12:00 to 15:00 (Beijing time, UTC+8:00)(Beijing time, UTC+8:00). daily carbon 379 sequestration The carbon sequestration period in summer and autumn is 1.5-3 hrs longer than in 380 winter. The timing of maximum carbon sequestration capacity changes with each season. In

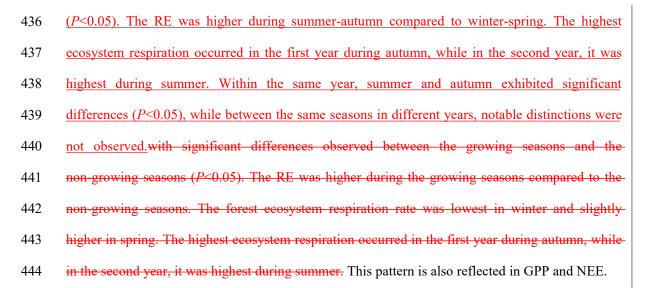
winter, the transition from nighttime carbon release to daytime carbon uptake occurs around 08:30, 381 382 which is approximately 1 hour later than in summer. GPP characterizes the forest's carbon 383 sequestration capacity, and since photosynthesis does not occur at night, GPP is zero during nighttime. The maximum daily total productivity is recorded at $14.76 \pm 7.34 \ \mu mol \ CO_2 \ m^{-2} \ s^{-1}$ 384 during the summer of the second year, with a standard deviation indicating greater variability in 385 386 GPP and NEE during the summer and autumn compared to the winter and spring. Although 387 diurnal variations in RE are relatively small, there are significant seasonal differences. During the 388 night, when only respiration occurs, RE equals NEE. However, as photosynthesis becomes active 389 during the day, RE gradually increases and stabilizes. The respiratory rate of the coniferous forest 390 is highest in autumn, being eight times greater than in winter. In winter, the transition from 391 nighttime carbon release to daytime carbon uptake occurs around 08:30, while in summer, it shifts-392 to around 07:30 (Beijing time, UTC+8:00). GPP reflects the carbon sequestration capacity of the forest, with the recorded daily total productivity highest at 14.76 umol CO₂ m⁻² s⁺-during summer 393 season of second year, RE exhibits minor diurnal variations but shows significant seasonal 394 differences, with maximum and minimum diurnal RE values of 0.73 umol CO2 m⁻² s⁻¹ and 0.17 395 umol CO₂ m⁻² s⁻¹, respectively. The respiration rate of the coniferous forest during the summer and 396 397 autumn is 5-8 times higher than that in the winter and spring.



VPD, and PAR. This implies that an increase in RH is unfavorable for the forest's absorption of 410 411 carbon dioxide. Among these environmental factors, PAR plays a dominant role. Furthermore, the 412 figure illustrates the relationships between environmental factors, showing a positive correlation between RH and TA, and a negative correlation with VPD and APR. The indicators exhibit some 413 414 seasonality, with notable differences between the winter-spring and summer-autumn seasons, indicating limited similarity between seasons. The fitting results between NEE and environmental 415 factors indicate that the selected environmental factors have a significant impact on NEE (P<0.001) 416 417 (Figure 4). However, the influence of individual environmental factors on NEE varies across 418 different seasons. RH has the smallest impact on NEE during the summer, while AT, VPD, and 419 PAR exhibit the strongest influence on NEE during the autumn. These factors consistently have the least impact on NEE during autumn. In the same season, PAR primarily controls NEE, with an 420 421 R² value reaching up to 0.957. Positive values of NEE indicate carbon emissions, while negative 422 values indicate carbon uptake. Therefore, air temperature, vapor pressure deficit, and PAR all have a significant positive effect on carbon uptake, while an increase in humidity leads to a 423 424 noticeable reduction in carbon uptake.







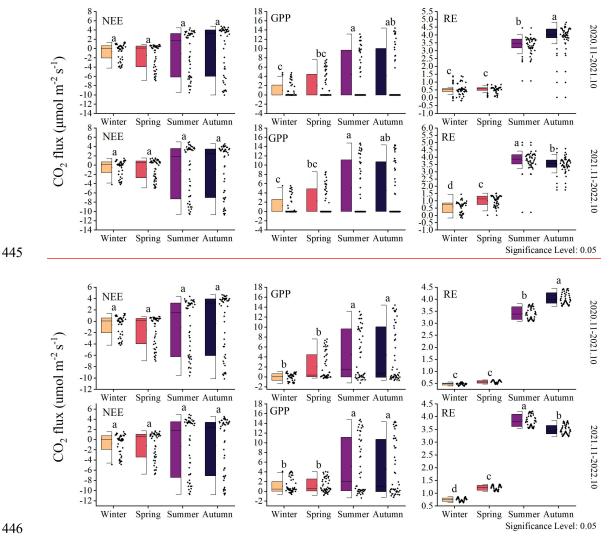
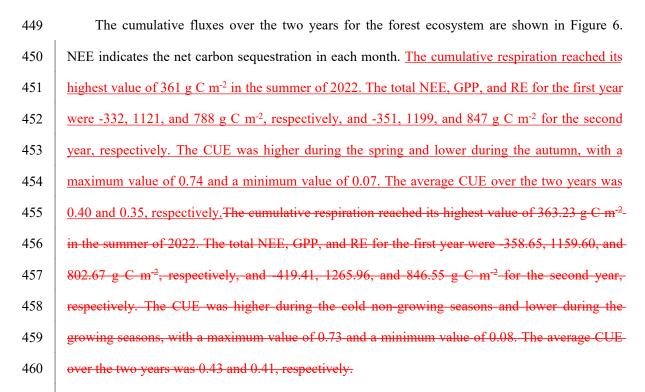
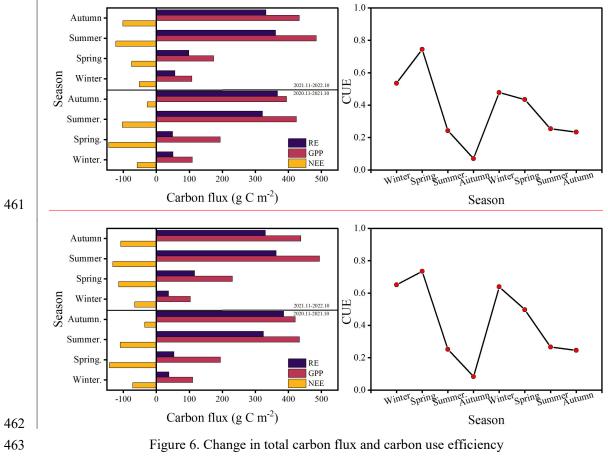




Figure 5. Seasonal variation of carbon fluxes

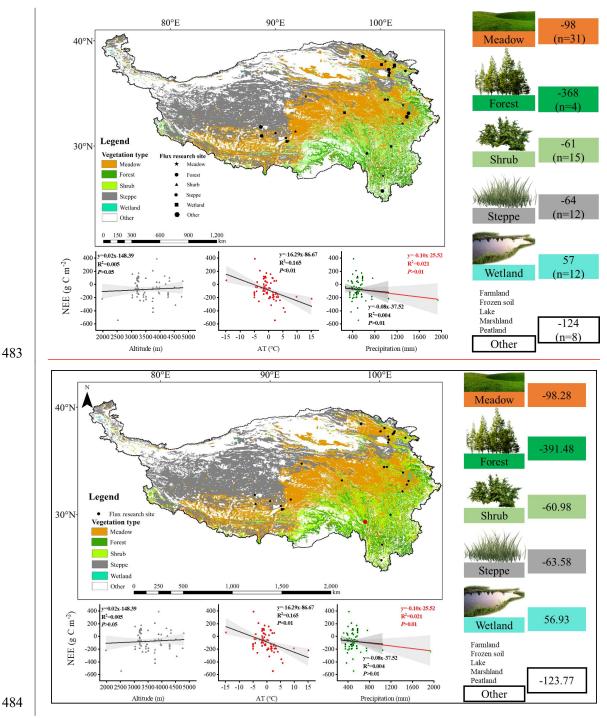




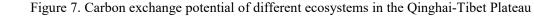


464 3.6 The carbon sequestration potential of subalpine forests of QTP

465	To clarify the carbon sequestration contribution of the subalpine forests found in the QTP, we
466	compared these research results (Figure 7). Found that ecosystems with high vegetation cover
467	exhibited higher annual cumulative carbon sequestration. Among these ecosystems, the subalpine
468	forests in the QTP showed the highest carbon sequestration potential, reaching an average of
469	<u>368</u> 391.48 g C m ⁻² per year. The carbon sequestration potential of different ecosystems ranked as
470	follows: forest > meadow > steppe > shrub. The average value for wetlands indicated that they are
471	a significant source of CO ₂ , releasing 5756.93 g C m ⁻² into the atmosphere annually. We also
472	analyzed the influence of altitude, mean annual air temperature, and precipitation on NEE at these
473	sites in the QTP. It has been observed that these sites cover a wide range of altitudes, ranging from
474	1977 to 4800 m. According to existing results, an increase in elevation may lead to a reduction in
475	carbon uptake, while the range of mean annual temperature varies between -14.8 to 15.1 °C, and
476	higher mean annual temperatures significantly increase carbon uptake. Forests exhibit the highest
477	mean annual precipitation, averaging 827 mm, with mean annual precipitation having a relatively
478	weak impact on the NEE. It was found that increasing elevation had a negative impact on carbon
479	uptake, while higher mean annual temperatures significantly increased carbon uptake. Mean
480	annual precipitation had a weak influence on NEE. These findings highlight the important role of
481	subalpine forests in carbon sequestration in the QTP and provide insights into the factors that
482	affect carbon exchange in the QTP, such as altitude, temperature, and precipitation.







486 4 Discussion

487 4.1 Main factors affecting the carbon sequestration function of subalpine forests

488 Climate change significantly affects the vegetation's carbon sequestration capacity,
489 particularly at the seasonal scale due to phenological changes (Acosta-Hernández et al., 2020). In
490 the short term, these factors (PAR, AT, RH, and VPD) play important roles in regulating
491 vegetation photosynthesis and, consequently, carbon uptake. For instance, PAR representing the

portion of solar energy that can be utilized by plants and is an essential component in chloroplast 492 493 reactions. PAR drives a nonlinear response of GPP to Solar-induced fluorescence (SIF) across 494 different seasons, resulting in a strong positive correlation between GPP and SIF (Wang et al., 495 2023b). VPD affects photosynthesis and transpiration of leaves, with stomata serving as tiny pores 496 mediating carbon dioxide uptake. Research has demonstrated that excessive increases in VPD are 497 detrimental to photosynthesis. For instance, a moderate increase in VPD significantly reduces 498 photosynthetic efficiency under light fluctuations, due to changes in RH and/or AT often 499 accompany fluctuations in light, studies also indicate that the impact of VPD on sunlight 500 utilization efficiency is primarily determined by relative RH rather than AT (Liu et al., 2024). In 501 different seasons, the same influencing factors exhibit varying degrees of contribution to NEE. For 502 example, during winter, when the climatic conditions are relatively harsh with low air temperature 503 and humidity, the forest maintains a low level of carbon uptake. Climate change is the significant 504 factor affecting the vegetation's carbon sequestration capacity, particularly at the seasonal scale 505 due to phenological changes (Acosta-Hernández et al., 2020). Our study has demonstrated that, in-506 the short term, NEE is primarily influenced by factors such as PAR, AT, RH, and VPD. These 507 factors play a role in regulating vegetation photosynthesis and, consequently, carbon uptake. For 508 instance, PAR represents the portion of solar energy that can be utilized by plants and is an 509 essential component in chloroplast reactions. AT regulates the activity of enzymes involved in 510 light and dark reactions, which may contribute to seasonal variations in NEE. RH and VPD impact 511 the entire process of photosynthesis by influencing the concentration of CO₂-in the air and the 512 stomatal conductance (the pathway for CO2 exchange). In different seasons, the same influencing 513 factors exhibit varying degrees of contribution to NEE. For example, during winter, when the 514 elimatic conditions are relatively harsh with low air temperature and humidity, the forest 515 maintains a low level of carbon uptake. While the forest continues to absorb carbon dioxide, the 516 uptake remains limited at a low level under such unfavorable conditions. On longer time scales, 517 such as annual and decadal variations, the inherent changes in forest NEE may be attributed to 518 disturbances and recovery (Hayek et al., 2018). In this study, significant differences in ecosystem 519 respiration were observed during the summer and autumn in different years. Past research 520 suggested that due to leaf aging or water stress, the photosynthetic light use efficiency of the

521 ecosystem peaks after spring leaf expansion and gradually declines (Wehr et al., 2016). This 522 implies a peak in carbon exchange during the summer, followed by higher productivity and 523 ecosystem respiration in the following seasons. The variation in different years may be attributed to rainfall regulating the availability of natural resources such as water, biomass, litter, and soil 524 525 nutrients (Schwinning and Sala, 2004). For instance, in temperate forests, when microbial biomass 526 undergoes seasonal changes, microbial activity exhibits a seasonal lag in response to temperature 527 variation, resulting in a seasonally delayed effect between litter heterotrophic respiration and 528 temperature (Ataka et al., 2020). Whether such differences persist between different years on 529 longer time scales remains to be demonstrated through more sustained and detailed research in the 530 future. Ecosystem respirationResearch by Amiro (2001) has demonstrated that disturbances-531 caused by fire and logging have been found to regulate the carbon balance of northern forests in 532 Canada over several decades. Additionally, there are close relationships between subtle climate-533 changes, stand dynamics, tree age, post-disturbance time, and forest carbon storage and cycling 534 (Bradford et al., 2008). Compared to naturally regenerating forests, actively restored forests-535 exhibit higher rates of carbon accumulation. Restoration efforts have been shown to increase 536 aboveground carbon density recovery rates by more than 50% over a decade, from 2.9 to 4.4 537 megagrams per hectare per year (Philipson et al., 2020). The carbon dioxide generated by soil microbial activity is an essential component of forest ecosystem respiration. Soils contain the 538 539 largest organic carbon reservoir on Earth, three times more than the carbon content in the 540 atmosphere (Tifafi et al., 2018). With climate warming, soil microorganisms, and root systems-541 will decompose soil organic carbon at a faster rate, releasing carbon dioxide into the atmosphere-542 more rapidly. Temperature plays a more sensitive role in soil carbon turnover in cold climate 543 regions compared to warmer conditions (Koven et al., 2017). Ecological respiration sensitivity to 544 temperature is represented by the Q_{10} coefficient. In this study, seasonal variations influenced the 545 magnitude of Q_{10} (as shown in Figure 8). The calculated Q_{10} for each season are as follows: 9.025, 546 2.22, 2.71, and 4.48. The winter season exhibited the highest sensitivity of forest ecosystem 547 respiration to temperature, indicating that respiration rates in the winter are more responsive to 548 changes in temperature compared to other seasons. The main reason for such differences is that 549 ecosystem respiration consists of heterotrophic respiration and autotrophic respiration, which are

550 typically governed by different factors (Edwards, 1975). For instance, the high activity of soil 551 microbes contributes to heterotrophic respiration, a process dominated by soil temperature and 552 moisture conditions, which are severely restricted during the cold and dry conditions of winter 553 (Falge et al., 2002). Simultaneously, due to the changing relative roles of growth and maintenance 554 respiration, the allocation of autotrophic respiration varies seasonally. In winter, soil CO₂ 555 emissions constitute a significant portion of ecosystem CO₂ emissions, and in some boreal forests, 556 the ratio between the two can reach 0.6 or even higher (Davidson et al., 2006), In winter, under the 557 frequent coverage of snow, cold-adapted microorganisms thriving in a relatively narrow sub-zero 558 temperature range engage in respiration and exhibit relatively high sensitivity to warming or 559 cooling beyond this range (Monson et al., 2006). The seasonal patterns of the Q_{10} value are jointly 560 determined by the variation in the ratio of soil respiration to ecosystem respiration, reflecting these 561 seasonal changes. The winter season exhibited the highest sensitivity of forest ecosystem 562 respiration to temperature, indicating that respiration rates in the winter are more responsive to-563 changes in temperature compared to other seasons.

564 Our integrated analysis (as shown in Figure 7) reveals that despite the high elevation of the 565 "Third Pole", the topographic factor of elevation does not have a significant impact on carbon 566 uptake. Instead, NEE gradually increases with a steep rise in elevation. Research conducted by Wang et al. (2023c) WANG et al. (2023b), indicates that mean annual average temperature and 567 568 precipitation are the main driving factors of interannual variations in NEE in alpine meadows and 569 alpine steppes. Decreased precipitation resulting in a transition into carbon sources at some 570 regions with high precipitation-dependent alpine grasslands. It is worth noting that, among all data 571 collection sites, alpine wetlands show an average carbon source trend. Due to prolonged flooding 572 and low temperatures, microbial activity in alpine wetlands is hindered, and the accumulation of 573 organic carbon from plant litter decomposition is substantial. As a result, approximately 5756.93 g C m⁻² is emitted into the atmosphere annually. Previous studies have indicated that NEE in alpine 574 575 wetlands is increasing with global warming (Yasin et al., 2022).

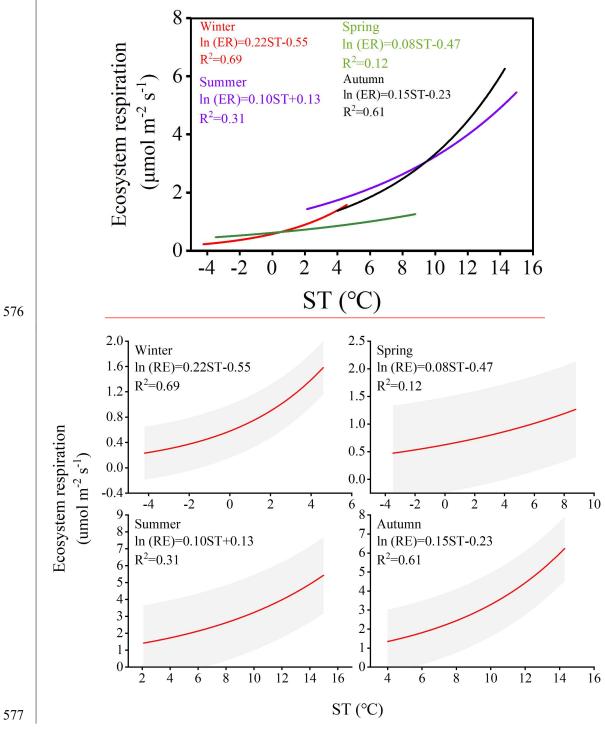




Figure 8. Relationship between NEEnight and soil temperature in different seasons

579 4.2 Sustained carbon sequestration of subalpine forests

580 Subalpine forests are integral components of global alpine ecosystems and play crucial roles 581 in the global carbon cycle. Our study on subalpine forests demonstrates a continuous absorbing of 582 carbon dioxide even during winter, which aligns well with measurements taken in the vicinity of 583 Mount Fuji in Japan (Mizoguchi et al., 2012). The age of subalpine forests is a crucial factor 584 influencing sustained carbon sequestration. Based on NPP simulations of natural subalpine forests 585 in the Northern Rockies, Carey. (2001) found that aboveground net primary productivity reaches 586 its maximum after approximately 250 years, followed by a decline, this challenges the previous 587 view that forests older than 100 years are generally considered to be unimportant carbon sinks. 588 Compared to the forest (mature forest) of Mount Gongga in the QTP (e.g., Zhang et al., 2018), the 589 subalpine forest in this study exhibits a stronger carbon sequestration capacity. However, its 590 carbon sequestration ability is slightly weaker than that of the Qilian Mountains high-mountain 591 forests (approximately 60-70 years old) in the QTP (Zhang et al., 2018; Du et al., 2022b). 592 Although existing flux monitoring results of high-altitude forests in the QTP indicate that these 593 forest ecosystems act as carbon sinks, it is important to consider that globally there are still many 594 cold regions with coniferous forests serve as carbon sources. For example, continuous CO₂ flux 595 monitoring from native boreal forests in Sweden for over 10 years indicates that they are a net 596 carbon source, which is attributed to the contribution of woody debris to RE due to disturbances 597 such as extreme weather events, fires, insect infestations, and pathogen attacks (Hadden and 598 Grelle, 2017). In the summer of 2018, Europe experienced a heatwave that affected the carbon 599 cycling in forests. The mixed coniferous-deciduous forest in southern Estonian, under the 600 influence of the heatwave, transitioned from a net carbon sink to a net carbon source in 2018 601 (Krasnova et al., 2022). Particular attention should be paid to the long-term monitoring in 602 high-altitude environments of the impact of disturbances on forest carbon sequestration capacity. 603 Our study has shown that forests in the QTP have the strongest carbon sink capacity, indicating 604 that alpine forests will have an important sustained effect on carbon reduction in the QTP in the 605 context of future climate change, but whether this sustained effect will be longer than other 606 ecosystems is still unknown. However, a modeling experiment in a large semi-arid area of 607 California predicted that grasslands are more resilient carbon sinks than forests in responding to 608 climate change in the 21st century (Dass et al., 2018). In terms of carbon sequestration rate, forests 609 in the QTP were significantly stronger than other ecosystems, followed by grasslands, while 610 alpine deserts and alpine grasslands in the north-western and southern regions were the main 611 carbon sources (Wu et al., 2022). Forests are mostly distributed in the south-eastern margin of the 612 QTP and the mid-altitude area near 3000 m in the Sichuan-Tibet alpine gorge area, with an area of

- 613 19.3×10^4 km² (Yu et al., 2022)(Y et al., 2022). Based on the average value of a few current 614 carbon flux monitoring, the forest in the QTP will absorb about 71×10^6 Mg C year⁻¹.
- 615 5 Conclusion

This study explores the carbon sequestration function, seasonal variations, and climate 616 617 drivers of subalpine forests in the QTP. Over the observational period, We synchronously monitored ecosystem carbon exchange and primary environmental factors using an eddy 618 619 covariance system. The research reveals that the subalpine forest is a carbon sink, with a total 620 NEE, GPP, and RE of -332, 1121, and 788 g C m⁻², respectively, and -351, 1199, and 847 g C m⁻² for two years, respectively.with a total NEE, GPP, and RE of 358.65, 1159.60, and 802.67 g C 621 m², respectively, and 419.41, 1265.96, and 846.55 g C m² for two years, respectively. 622 Photosynthetically active radiation was identified as the primary control of NEE. The NEE did not 623 624 exhibit significant differences across seasons. Combining results from other eddy covariance sites 625 on the QTP, this study highlights those forests have the highest carbon sequestration potential, reaching 368 g C m⁻² annually, followed by meadows, steppes, and shrubs. Wetlands, however, 626 were identified as a substantial carbon dioxide source. Despite the challenges posed by climate 627 628 change, the subalpine forests in the QTP retain substantial carbon sequestration potential. 629 Strengthening conservation and management efforts for subalpine forests is crucial to ensure their continued and significant carbon sequestration function in the future. Overall, this research 630 631 underscores the vital role of subalpine forests in the QTP as essential carbon sink regions, playing 632 a critical role in the context of global climate change.

633 **Data availability.** The data is available from the authors on request.

Authorship contributions. Niu Zhu: Conceptualization, study design, data analyses,
visualization, writing-original draft. JinNiu Wang: study design, writing—review & editing,
supervision, project administration, funding acquisition. Dongliang Luo and Xufeng Wang:
writing-reviewing & editing. Cheng Shen and Ning Wu: resources, data curation, supervision. all
authors approved the final manuscript.

639 **Declaration of competing interest.** The authors declare that they have no conflict of interest.

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