

1 Alkalinity sources in the Dutch Wadden Sea

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9 Abstract

10 The oceanic buffering capacity total alkalinity (TA), as the major global CO₂ sink, is of growing scientific interest. TA is
11 mainly generated by weathering, and further by various anaerobic metabolic processes. The Wadden Sea, located in the
12 southern North Sea is hypothesized to be a source of TA for the North Sea, but quantifications are scarce. This study observed
13 TA, dissolved inorganic carbon (DIC), and nutrients in the Dutch Wadden Sea in May 2019. Along several transects, surface
14 samples were taken to investigate spatial distribution patterns and to compare them with data from the late 1980s. A tidal cycle
15 was sampled to further shed light on TA generation and potential TA sources. We identified the Wadden Sea as a source of
16 TA with an average TA generation of 7.6 μmol kg⁻¹ h⁻¹ TA during ebb tide in the Ameland Inlet. TA was generated in the
17 sediments with deep pore water flow during low tide enriching the surface water. A combination of anaerobic processes and
18 CaCO₃ dissolution were potential sources of TA in the sediments. We assume that seasonality and the associated nitrate
19 availability in particular influence TA generation by denitrification, which we assume is low in spring and summer.

20 1 Introduction

21 As the regulator of the ocean carbon dioxide (CO₂) sink, total alkalinity (TA) is of increasing scientific interest and is
22 investigated worldwide in the Anthropocene (Abril and Frankignoulle, 2001;Bozec et al., 2005;Chen and Wang,
23 1999;Dickson, 1981;Middelburg et al., 2020;Norbistrath et al., 2022;Renforth and Henderson, 2017;Thomas et al.,
24 2004;2009;Sabine et al., 2004). The Anthropocene describes the current era of our planet, when environmental changes, driven
25 by humans, have become identifiable in geological records (Zalasiewicz et al., 2010;Crutzen, 2002). The climate and the
26 increasing atmospheric CO₂ content is mainly regulated by the open ocean. Around 30 % of the global anthropogenic CO₂
27 emissions are removed by the ocean (Gruber et al., 2019). The carbon storage capacity of the North Sea is an important
28 atmospheric CO₂ sink as it exports the absorbed CO₂ in the deep layers of the Atlantic Ocean where it is stored on longer time
29 scales (Borges et al., 2005;Bozec et al., 2005;Burt et al., 2016;Brenner et al., 2016;Hu and Cai, 2011;Schwichtenberg et al.,

30 2020;Thomas et al., 2004;2009). Two important aspects of the oceanic climate regulation are the oceanic circulation and TA.
31 Both interact well in highly active and shallow ocean areas such coastal zones and continental and marginal shelves. In these
32 shallow areas, TA is very susceptible due to various metabolic processes and the influence of adjacent zones like rivers,
33 estuaries, marshes, and tidal flats (e.g., Norbistrath et al., 2022;2023;Wang et al., 2016;Voynova et al., 2019). A previous study
34 by Norbistrath et al. (2022) showed that an enhanced riverine, metabolic alkalinity would lead to increasing CO₂ absorption in
35 the coastal zones of the North Sea, highlighting the need to further investigate TA regulation in adjacent zones of coastal
36 oceans.

37 Coastal zones, which are the direct interface between most, if not all, compartments of the Earth system (i.e., terrestrial, aquatic,
38 oceanic) and human societies, appear particularly vulnerable to environmental and climate change (Glavovic et al., 2015). This
39 holds true for the Wadden Sea, the shallow, coastal sea along an approximately 500 km coastline of the Netherlands, Germany,
40 and Denmark, in the southern North Sea, which is declared as an UNESCO world natural heritage site since 2009. Most of the
41 Wadden Sea is located between the protecting barrier Islands and the Mainland, which makes it the world's largest
42 uninterrupted stretch of tidal flats with multiple tidal inlets (Fig. 1). Due to the topography, the Wadden Sea is a highly dynamic
43 ecosystem with influences from the mainland and the North Sea (Hoppema, 1993;Postma, 1954;van Raaphorst and van der
44 Veer, 1990). Driving forces of the biogeochemical dynamics in the Wadden Sea are nutrient and organic matter (OM) imports
45 by rivers and high suspended particulate matter (SPM) imports from the North Sea (Van Beusekom et al., 2012;Postma, 1954).
46 Physical sources of variability in the Wadden Sea are oceanic driven wind, waves, and tidal currents, as well as the
47 counterclockwise circulation of the North Sea (Elias et al., 2012). Large tidal amplitude and currents in conjunction with
48 shallow water depths allow for vertical water column mixing and an exchange between the pelagic and benthic realms
49 including deep pore water exchange (Røy et al., 2008). The high tidal currents also impact the biogeochemistry of the North
50 Sea (Postma, 1954), as they cause an exchange of water between the North Sea and the Wadden Sea and play an important
51 role in the import of particulate matter from the North Sea (Burchard et al., 2008).

52 TA, primarily consisting of bicarbonate and carbonate, is generated by chemical rock weathering (Suchet and Probst,
53 1993;Meybeck, 1987;Berner et al., 1983), and in various stoichiometries by calcium carbonate (CaCO₃) dissolution and
54 anaerobic metabolic processes, such as denitrification, which is the reduction process of nitrate to dinitrogen gas in the nitrogen
55 cycle (Hu and Cai, 2011;Wolf-Gladrow et al., 2007;Chen and Wang, 1999;Brewer and Goldman, 1976). Understanding of TA
56 sources have recently become increasingly important due to increasing anthropogenic CO₂ emissions, and the resulting demand
57 for ocean based net-negative CO₂ emissions (e.g., Keith et al., 2006;Matthews and Caldeira, 2008;Zhang et al., 2022). In
58 previous studies, the Wadden Sea was identified as a TA source for the North Sea with a loading between 39 Gmol yr⁻¹
59 (Schwichtenberg et al., 2020) and 73 Gmol yr⁻¹ (Thomas et al., 2009). Both studies suggested the entire Wadden Sea as one of
60 the most important TA sources of the carbon storage capacity for the North Sea. Burt et al. (2016) highlighted the importance
61 of coastal TA production for regulating the buffer system in the North Sea, and suggested denitrification as the major TA
62 source. Due to the strong connection between the North Sea and the Wadden Sea, a better understanding of TA generation in
63 the latter is required. Here, we focus on the Dutch Wadden Sea that has been well-studied during the past decades (Hoppema,

64 1990, 1991, 1993;De Jonge et al., 1993;Elias et al., 2012;Ridderinkhof et al., 1990;Postma, 1954;Van Beusekom et al.,
65 2019;Schwichtenberg et al., 2017). In particular Hoppema (1990);(1993) observed the spatial and temporal variability of TA
66 in May, which we compare with our observed transect data to detect potential differences over the last 30 years. In addition,
67 we further shed light on potential TA sources in the Dutch Wadden Sea.

68 **2 Methods**

69 **2.1 Study site and sampling**

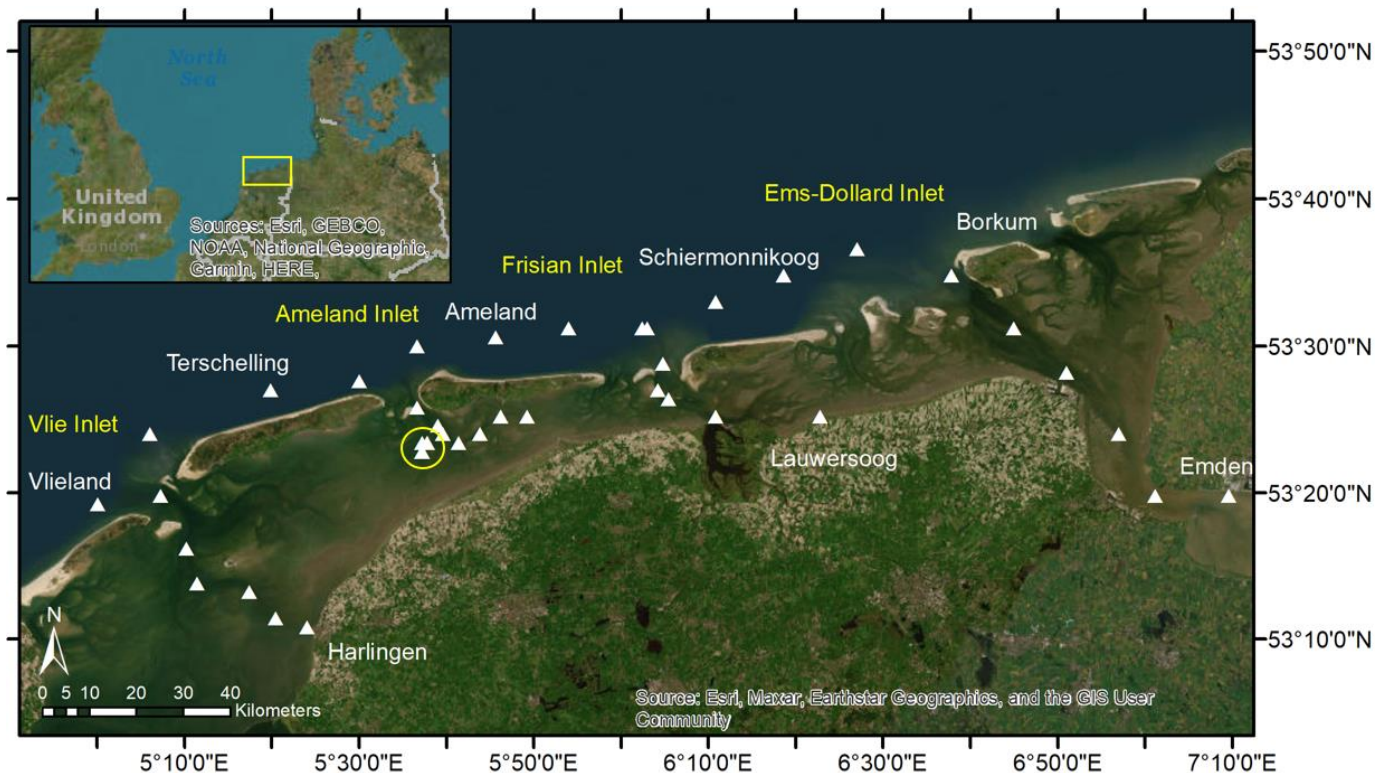
70 This study is based on samples collected on a research cruise (LP20190515) in the Dutch Wadden Sea (Frisian Islands) on RV
71 *Ludwig Prandtl* in May 2019 (Fig. 1). We collected water samples in the Wadden Sea starting at Harlingen, through the Vlie
72 Inlet around the islands Vlieland and Terschelling, through the Ameland Inlet to Ameland Island, from there on via the Frisian
73 Inlet to Lauwersoog, and around Schiermonnikoog Island via the Ems-Dollard Inlet to Emden. In addition, we sampled a tidal
74 cycle from high tide to low tide (21.05.2019) and from low tide to high tide (23.05.2019) on each day as an anchor station in
75 the waterway at the western side of Ameland in the Ameland Inlet.

76 Nearly half-hourly, we continuously collected discrete surface (1.2 m depth) water samples with a bypass from the onboard
77 flow-through FerryBox system (Petersen et al., 2011), which also provided essential physical parameters such as salinity and
78 temperature.

79 For TA and DIC measurements we sampled water with overflow into 300 mL BOD (biological oxygen demand) bottles and
80 preserved them with 300 μ L saturated mercury chloride (HgCl_2) to stop biological activity. Each BOD bottle was filled without
81 air bubbles and closed by using a ground-glass stopper coated in Apiezon® type M grease and a plastic cap. The samples were
82 stored in a cool dark environment until measurements in the lab.

83 Water for nutrient samples was filtered through pre-combusted (4 h, 450 °C) GF/F filters and the filtrate was stored frozen in
84 three 15 mL Falcon tubes for triplicate measurements in the lab.

85 To determine the total carbon (C), organic carbon (C_{org}) and nitrogen (N) concentrations in SPM and associated $C_{\text{org}}:\text{N}$ ratios,
86 we used pre-combusted (4 h, 450 °C) GF/F filters, which were dried after sampling at 50 °C to remove all humidity and were
87 stored frozen afterwards until measurement.



88
 89 **Figure 1** Sampling site in the Dutch Wadden Sea. The sampling stations around the Frisian Islands in May 2019 are visualized
 90 with white triangles. The yellow circle highlights the anchor stations for the tidal cycle sampling in the Ameland Inlet on two
 91 days. During the sampling day from low tide to high tide, we had two samples that we took slightly more western due to
 92 drifting. The island and city names are shown in white, the inlets in yellow. The tidal flats and sedimentary structures are well
 93 visible between the barrier islands and the mainland.

94 2.2 Carbon species analyses

95 The parallel analyses of TA and DIC were carried out by using the VINDTA 3C (Versatile Instrument for the Determination
 96 of Total dissolved inorganic carbon and Alkalinity, MARIANDA - marine analytics and data), which measures TA by
 97 potentiometric titration and DIC by coulometric titration with a measurement precision $< 2 \mu\text{mol kg}^{-1}$ (Shadwick et al., 2011).
 98 To ensure a consistent calibration of both measurements, certified reference material (CRM batch # 187) provided by Andrew
 99 G. Dickson (Scripps Institution of Oceanography) was used.

100 The calcite and aragonite saturation states (Ω) and the pH were computed with the CO₂SYST program (Lewis and Wallace,
 101 1998), using the measured parameters TA, DIC, salinity, temperature, silicate and phosphate as input variables, together with
 102 the dissociation constants from Mehrbach et al. (1973), as refit by Dickson and Millero (1987).

103 **2.3 Nutrient analyses**

104 The nutrients were measured with a continuous flow automated nutrient analyzer (AA3, SEAL Analytical) and a standard
105 colorimetric technique (Hansen and Koroleff, 2007) for nitrate (NO_3^-), nitrite (NO_2^-), phosphate (PO_4^{3-}), and silicate (Si), and
106 a fluorometric method (K  rouel and Aminot, 1997) for ammonium (NH_4^+) (Grasshoff et al., 2009).

107 For the C_{org} determination, filters were acidified with 1N HCl and dried overnight to remove all inorganic carbon content.
108 Filters were measured with a CHN-elemental analyzer (Eurovector EA 3000, HEKAtech GmbH) in the Institute of Geology,
109 University Hamburg, and calibrated against a certified acetanilide standard (IVA Analysentechnik, Germany). The standard
110 deviations were 0.05 % for carbon and 0.005 % for nitrogen.

111 **2.4 Data analyses**

112 The data analyses were performed by using RStudio Version 1.3.1073    2009-2020 RStudio, PBC. The linear regression
113 Model II was performed by using the “lmodel2” R package, and the plots were created with the “ggplot2” R package.

114 **3 Results**

115 **3.1 Spatial parameter distribution**

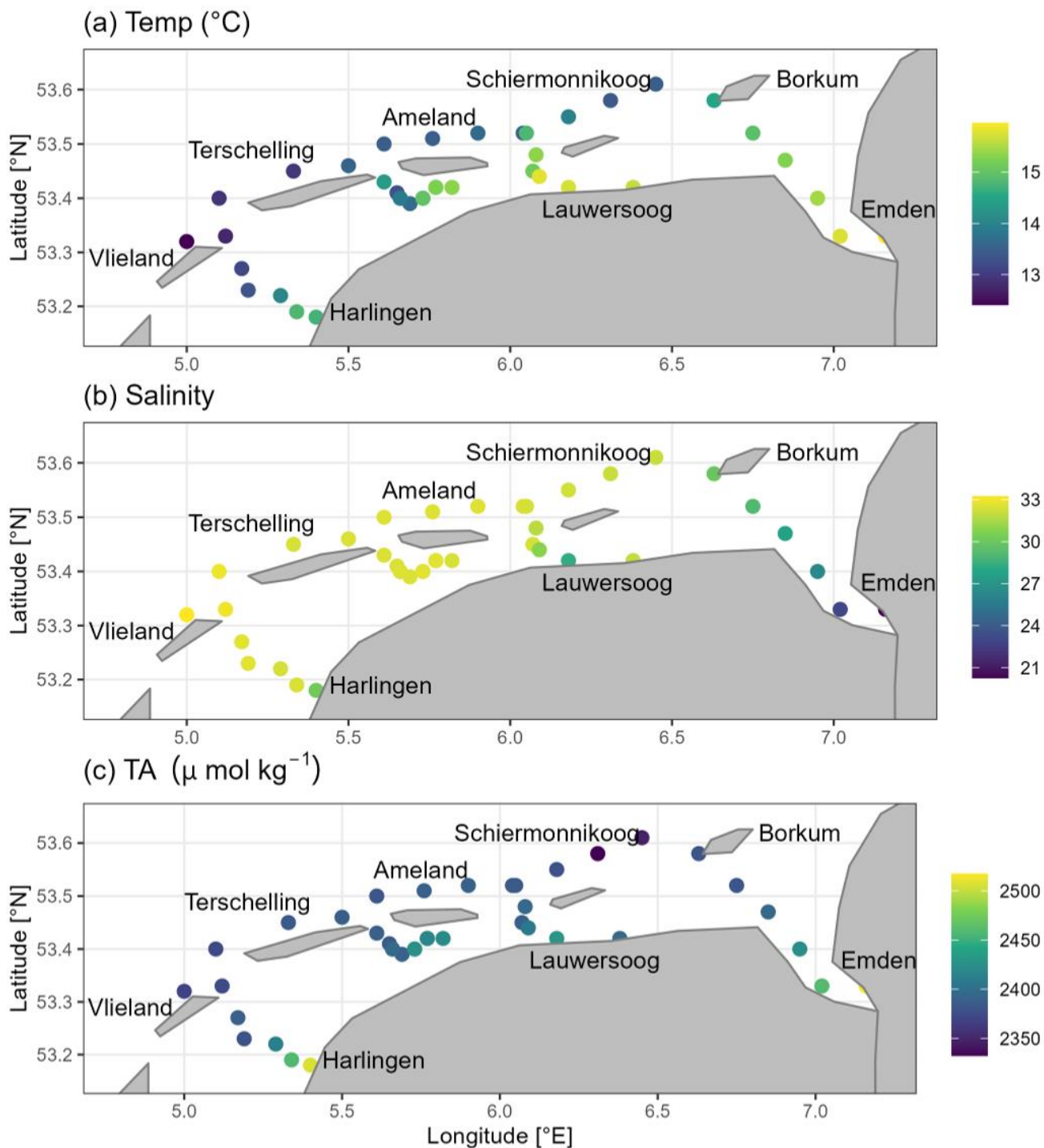
116 To investigate the spatial distribution of total alkalinity (TA) in the Dutch Wadden Sea and compare its general status with
117 earlier studies (e.g., Hoppema, 1990), we observed the spatial distribution of TA and related parameters from the coastal
118 mainland towards the open North Sea as a surface water transect.

119 The temperatures varied between 12 and 16   C with higher temperatures towards the coastal mainland (Fig. 2a). Salinity was
120 relatively stable with only minor differences varying from 28 to 33 (Fig. 2b). Lower salinities were only observed in the four
121 sampling stations in the Ems Estuary with the minimum value of 20.25 at the most upstream station.

122 Spatial transect TA contents ranged from 2332 $\mu\text{mol kg}^{-1}$ to 2517 $\mu\text{mol kg}^{-1}$. We observed lower contents on the oceanic, i.e.,
123 North Sea side of the Frisian Islands with somewhat higher contents around Ameland (Fig. 2c). In contrast to the oceanic side,
124 the values were higher ($> 2380 \mu\text{mol kg}^{-1}$ TA) in the Wadden Sea. Only in the Ems Estuary, the contents were even higher,
125 with values up to 2517 $\mu\text{mol kg}^{-1}$ TA at the most upstream station. Higher TA values in the Wadden Sea than in the open
126 ocean, supporting the assumption of TA being generated in this tidal flat area.

127 Silicate (Si), showed a similar pattern with higher concentrations in the Wadden Sea and lower ones towards the ocean (Fig.
128 A1). Highest concentrations were observed at the coastal mainland and in the Ems Estuary. Silicate concentrations were
129 between 0.26 and 56.32 $\mu\text{mol L}^{-1}$. Both, the calcite and aragonite saturation states (Ω) were supersaturated in the entire study
130 site. Values ranged from 2.3 to 4.6 for calcite (Fig. A2), and from 1.4 to 3.0 for aragonite (Table B1). Highest values were
131 observed at the oceanic side of the barrier islands in the North Sea. The lowest values were observed near Harlingen and in the
132 Ems Estuary. Like the calcite and aragonite saturation states, the pH values were higher in the North Sea, and lower in the

133 Wadden Sea and near the coastal mainland (Fig. A3). The pH values ranged from 7.86 to 8.19. The lowest values were observed
134 near Harlingen and in the Ems Estuary. The nitrate (NO_3^-) concentrations were in a similarly low range ($< 3 \mu\text{mol L}^{-1} \text{NO}_3^-$)
135 throughout the transect. Higher concentrations ($< 6 \mu\text{mol L}^{-1} \text{NO}_3^-$) were observed only at a few stations close to land, and
136 maximum concentrations ($< 38 \mu\text{mol L}^{-1} \text{NO}_3^-$) were observed in the Ems Estuary (Fig. A4). DIC contents ranged from 2097
137 $\mu\text{mol kg}^{-1}$ to 2430 $\mu\text{mol kg}^{-1}$ (Fig. A5). DIC values showed a similar pattern as TA values, with higher concentrations near the
138 coastal mainland and in the Ems Estuary, and decreasing concentrations toward the North Sea, where DIC reached minimum
139 values.



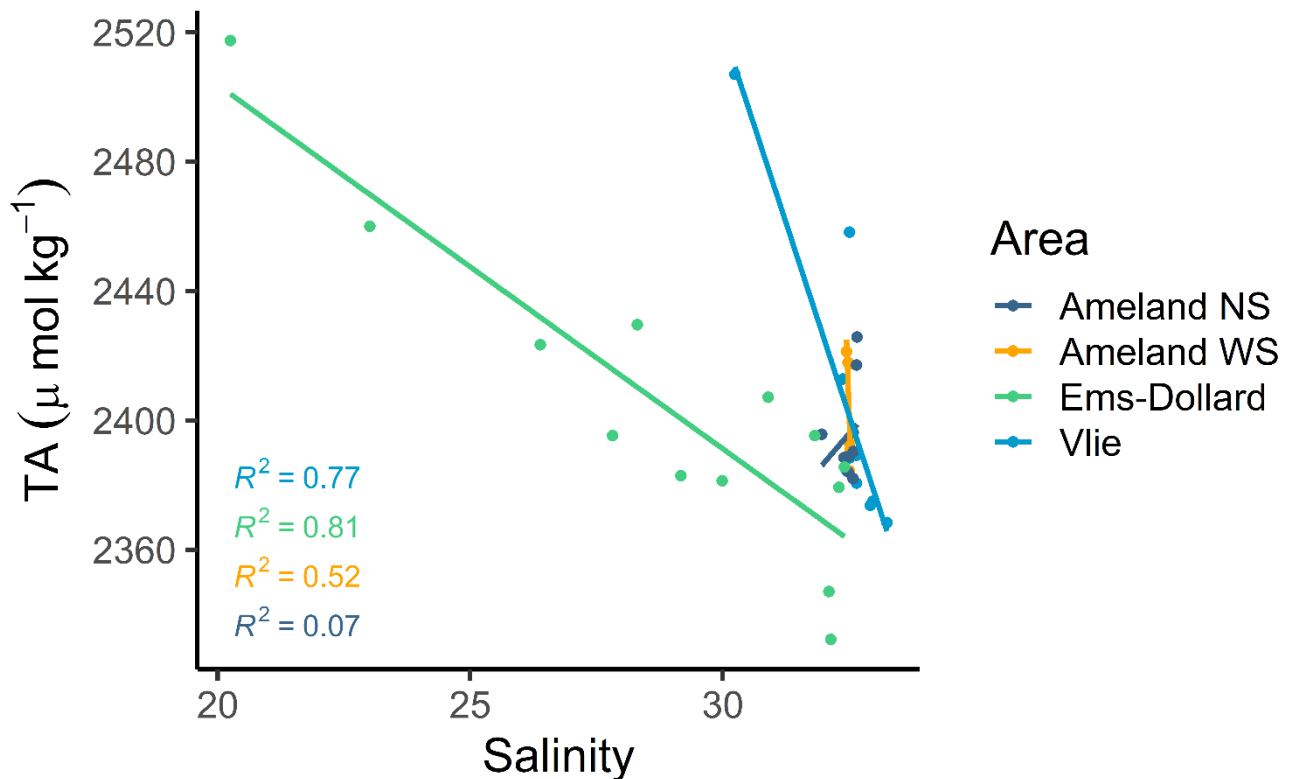
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141 **Figure 2** Latitudinal and longitudinal distribution of a) temperature ($^{\circ}\text{C}$), b) salinity, and c) total alkalinity (TA; $\mu\text{mol kg}^{-1}$)

142 from surface water samples in May 2019.

143

144 The strong impact from the inner Ems Estuary is visible in all parameters with higher values in the outer estuary and its adjacent
145 zones, or with lower values in case of pH and the calcite saturation state. Furthermore, we observed higher values around
146 Ameland Island than in the western part of the transect starting from Harlingen towards the Vlie Inlet. At the oceanic side of
147 the Vlie Inlet, the impact of the North Sea is visible through lower temperatures and higher salinities. The North Sea impact is
148 also visible in the mixing between TA and salinity (Fig. 3). A relatively linear mixing behavior was only observed in the
149 transect through the Ems-Dollard Inlet and Vlie Inlet, where TA contents decreased with increasing salinities from the mainland
150 towards the ocean (Fig. 3), identifying the Dutch Wadden Sea as a source of TA. In contrast to the TA content computed for
151 the salinity end-member in the Ems-Dollard Inlet, we detected higher TA contents around Ameland, both at the North Sea side
152 (Ameland NS) and the Wadden Sea side (Ameland WS), as well as in the Vlie Inlet, which further support additional TA
153 sources in the Wadden Sea (Fig. 3). The Ameland NS and Ameland WS data clearly indicated non-conservative behavior with
154 increasing TA contents at constant salinities.



155

156 **Figure 3** Mixing between total alkalinity (TA) and salinity in the oceanic side of Ameland and the Frisian Inlet (Ameland NS),
157 in the Wadden Sea site of Ameland (Ameland WS), around Schiermonnikoog and in the Ems-Dollard Inlet, and in the Vlie
158 Inlet.

159 **3.2 Tidal cycle**

160 To estimate TA generation in the Dutch Wadden Sea, to shed light on potential TA sources, and to estimate the potential TA
161 amount that is exported to the North Sea, we observed a tidal cycle at an anchor station in the Ameland Inlet on two days
162 during flood tide and ebb tide, respectively. Here, the focus is on ebb tide data that we used to identify patterns in the several
163 biogeochemical parameters in off running water (Table B1). The salinity observations allow us to exclude addition by mixing
164 with freshwater sources, since the salinity was constant between 32.50 and 32.52 (Table B1). Salinity values were in the range
165 of saline waters like water of the North Sea.

166 During ebb tide, TA ranged from high tide with 2387 $\mu\text{mol kg}^{-1}$ TA to low tide with 2438 $\mu\text{mol kg}^{-1}$ TA (Fig. 4a). We identified
167 an increase of 51.6 $\mu\text{mol kg}^{-1}$ TA (ΔTA) over a duration of 6.8 h during ebb tide, resulting in a TA increase of 7.6 $\mu\text{mol kg}^{-1}$
168 h^{-1} TA at the sampling location.

169 DIC contents were similar to TA with minimum values at high tide (2172 $\mu\text{mol kg}^{-1}$ DIC) and highest values (2273 $\mu\text{mol kg}^{-1}$
170 DIC) at low tide. During ebb tide, we observed an increase of 101.3 $\mu\text{mol kg}^{-1}$ DIC (ΔDIC) resulting in a DIC increase of 14.9
171 $\mu\text{mol kg}^{-1} \text{h}^{-1}$ DIC (Fig. 4b). DIC increased almost twice as much as TA.

172 Nitrate concentrations approached seawater concentrations with an observed minimum of 1.26 $\mu\text{mol L}^{-1} \text{NO}_3^-$ and a maximum
173 of 2.17 $\mu\text{mol L}^{-1} \text{NO}_3^-$ (Fig. 4c). During ebb tide, nitrate slightly increased by 0.92 $\mu\text{mol L}^{-1} \text{NO}_3^-$ (ΔNO_3^-), resulting in a nitrate
174 increase of 0.13 $\mu\text{mol L}^{-1} \text{h}^{-1} \text{NO}_3^-$.

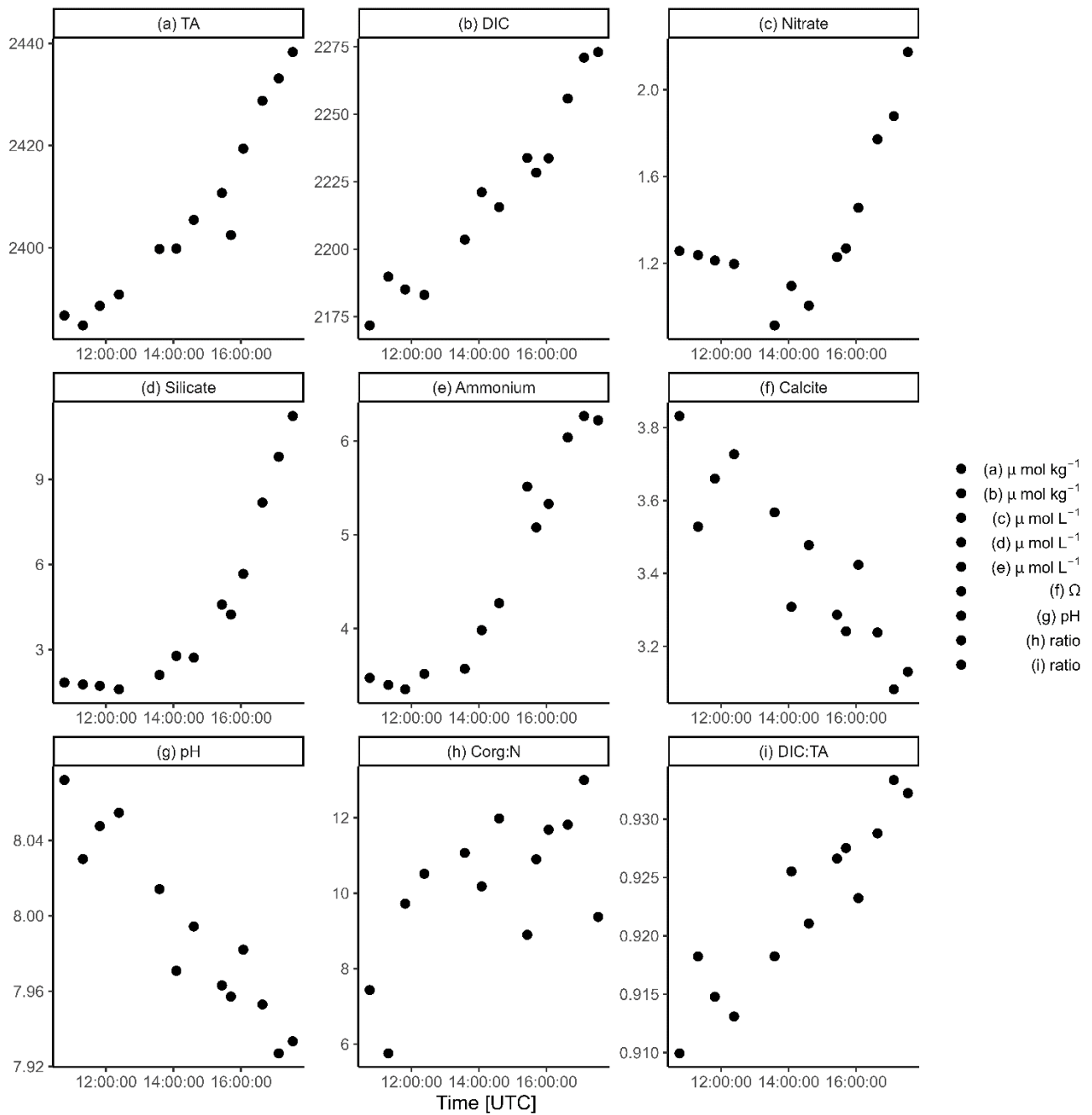
175 In silicate, a similar pattern with low values (1.8 $\mu\text{mol L}^{-1} \text{Si}$) at high tide and increasing concentrations during ebb tide to a
176 maximum of 11.2 $\mu\text{mol L}^{-1} \text{Si}$ was detected. The silicate increase (ΔSi) of 9.4 $\mu\text{mol L}^{-1} \text{Si}$ resulted in a silicate increase of 1.4
177 $\mu\text{mol L}^{-1} \text{h}^{-1} \text{Si}$ during ebb tide (Fig. 4d).

178 Ammonium increased from 3.47 $\mu\text{mol L}^{-1} \text{NH}_4^+$ to 6.22 $\mu\text{mol L}^{-1} \text{NH}_4^+$ during ebb tide (Fig. 4e). We observed an ammonium
179 increase (ΔNH_4^+) of 2.74 $\mu\text{mol L}^{-1} \text{NH}_4^+$ resulting in an increase of 0.4 $\mu\text{mol L}^{-1} \text{h}^{-1} \text{NH}_4^+$.

180 The calcite and aragonite saturation states had maximum values ($\Omega_{\text{Ca}} = 3.8$, $\Omega_{\text{Ar}} = 2.4$) at high tide and decreased to their
181 minimum ($\Omega_{\text{Ca}} = 3.1$, $\Omega_{\text{Ar}} = 2.0$) during ebb tide (Fig. 4f, Table B1). The maximum at high tide indicated the influence of the
182 North Sea that decreased during the ebb.

183 Like omega, the pH had maximum values (8.07) at high tide and decreased to a minimum (7.93) during ebb tide (Fig. 4g).

184 $C_{\text{org}}:\text{N}$ ratios of SPM increased during ebb tide (Fig. 4h). A minimum $C_{\text{org}}:\text{N}$ ratio of 5.6 was observed around high tide and
185 increased to a maximum of 13.0 during ebb tide. Simultaneously, the SPM concentration increased during ebb tide, from 12.8
186 mg L^{-1} SPM to a maximum of 82.4 mg L^{-1} SPM at the second last station (Table B1).



187

188 **Figure 4** Tidal cycle from high tide to low tide. Temporal distribution of a) total alkalinity (TA), b) dissolved inorganic
 189 carbon (DIC), c) nitrate, d) silicate, e) ammonium, f) calcite saturation state (Ω), g) pH, h) $C_{\text{org}}:\text{N}$ ratio of SPM, and i)
 190 DIC:TA ratio. Note the different y-axes.

191 3.3 TA generation

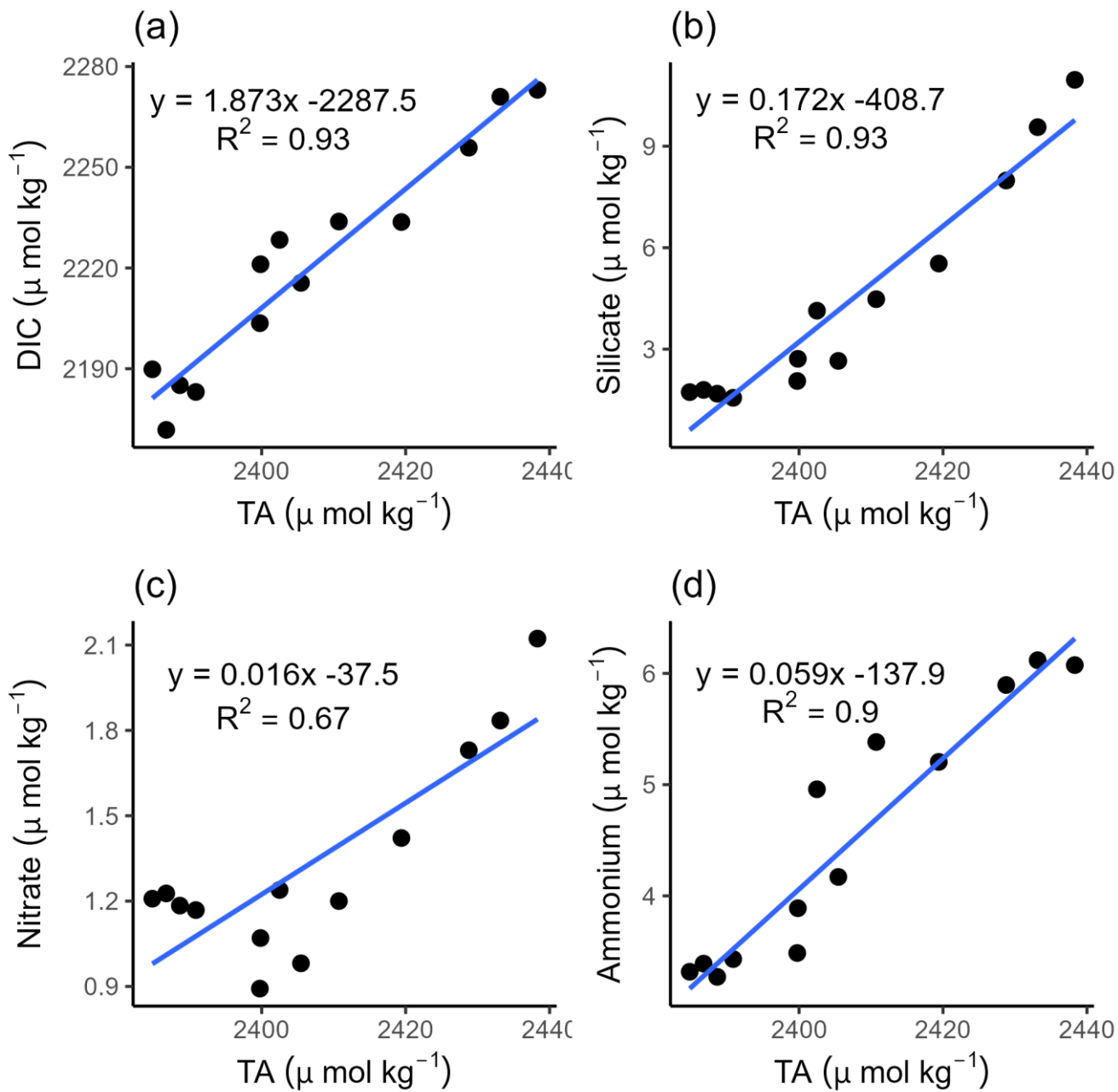
192 The Dutch Wadden Sea is exposed to strong tidal forcing by the North Sea leading to a bi-diurnal exchange of water. The
193 strong tidal forcing induces a strong benthic-pelagic coupling. Many studies support that the outflowing water exports material
194 from the sediment (e.g., Billerbeck et al., 2006; Røy et al., 2008). To support our assumption of a sediment source of TA, we
195 further investigated potential TA sources.

196 For a first rough estimate of a maximum TA export during ebb tide, we used the observed TA increase (Δ TA) of 51.6 μ mol
197 kg^{-1} TA during ebb tide (in the Ameland Inlet), and the tidal prism of $478 \cdot 10^6 \text{ m}^3$ of the Ameland Inlet (Louters and Gerritsen,
198 1994). With this estimate, we obtained a rough TA export on the order of 23.9 mol TA during ebb tide from the Ameland Inlet
199 into the North Sea. With a tidal duration of 6.8 h, the estimated TA export of 23.9 mol would result in a TA export of 3.5 mol
200 h^{-1} TA.

201 Based on the correlation of TA and silicate ($R^2 = 0.93$), and on the nonlinear relation between both, TA and salinity ($R^2 =$
202 0.32), as well as silicate and salinity ($R^2 = 0.21$), we were able to determine whether TA originates in this part of the Dutch
203 Wadden Sea or is carried by river runoff. Both TA and silicate increased almost proportionally during ebb tide, pointing to the
204 same origin (Fig. 5b). The non-conservative behavior of TA over salinity and silicate over salinity rules out any sources due
205 to freshwater dilution and river runoff (Table B1).

206 To further elucidate potential TA sources in the Dutch Wadden Sea, we correlated TA with DIC, silicate, nitrate, and
207 ammonium from high tide to low tide (Fig. 5). The first four samples on the left side of the plots in Fig. 5 show values probably
208 at the tipping point of high tide, whereby we recommend neglecting these samples in the interpretation of ebb tide samples.

209 First, the correlation between TA and DIC reveals the ratio between anaerobic and aerobic processes, which identifies a strong
210 positive correlation between DIC and TA ($R^2 = 0.93$) with TA contents higher than DIC contents (Fig. 5a). However, even
211 with contents of TA higher than DIC, the slope of 1.87 indicated DIC release excess with an increase in DIC (Δ DIC = 101.3
212 μ mol kg^{-1}) almost twice as high as TA (Δ TA = 51.6 μ mol kg^{-1}) (Fig. 5a). This may be caused by strong CO_2 production due
213 to high aerobic OM degradation, or uptake from the atmosphere due to water movement by tidal forcing. The TA increase can
214 be fueled by various processes which we will discuss below. We detected an almost linear positive correlation of increasing
215 TA and silicate ($R^2 = 0.93$) during ebb tide, supporting the pore water outflow (Fig. 5b). A stronger influence of the pore water
216 with ongoing ebb tide is indicated by increasing values. The positive correlation between nitrate and TA ($R^2 = 0.67$) (Fig. 5c)
217 was less strong than the correlations between TA and DIC, and TA and Si, which could be traced back on an effect of the first
218 four samplings as mentioned above. In the remaining samples, the increasing nitrate and TA contents suggest a stronger effect
219 of TA generation than nitrate production.



220

221 **Figure 5** Correlations of TA with a) dissolved inorganic carbon (DIC), b) silicate, c) nitrate, and d) ammonium during ebb

222 tide in the Ameland Inlet.

223 4 Discussion

224 4.1 Spatial TA variability

225 Hoppema (1990) reported TA distributions in the westernmost part of the Dutch Wadden Sea around the barrier islands Texel,
226 Vlieland, and Terschelling. He focused on the tidal basins drained by the tidal inlets Marsdiep and Vlie located more to the
227 west than our sampling stations (not visible on the map). Hoppema (1990) did not observe an increase of salinity in the Wadden
228 Sea from the freshwater source towards the ocean and associated this to the influence of tidal differences and an arbitrary
229 sampling scheme. The presence of seawater in the Dutch Wadden Sea and on the tidal flats is supported by our transect data,
230 which show relatively high salinities at coastal North Sea level. Brackish salinities were only detected in the large Ems Estuary,
231 which receives high discharges of freshwater from the river Ems, and also close to Harlingen and Lauwersoog, which have
232 direct freshwater inflows by smaller rivers and streams. The absence of clear salinity gradients in this part of the Dutch Wadden
233 Sea suggest that most of the IJsselmeer discharge was exchanged with the North Sea through the Marsdiep.

234

235 The spatial TA data by Hoppema (1990), show lower TA contents at stations with more freshwater influence and higher TA
236 contents in the tidal inlets. The data of this study also show high TA contents in the tidal inlets, suggesting TA generation in
237 sediments, which is fueled by high imports of nutrients and OM (Van Beusekom and De Jonge, 2002). The even higher TA
238 contents at stations with lower salinities close to the mainland observed in this study also show the influence from the
239 catchment area on the coast, and TA generation in the shallow sediments near the coast. In May (1986), Hoppema's (1990)
240 data showed TA contents ranging between 2319 and 2444 $\mu\text{mol kg}^{-1}$ TA at salinities between 18.62 and 29.17, while our
241 lowest observed TA content was 2332 $\mu\text{mol kg}^{-1}$ TA at a salinity of 32.14, and our highest TA content was 2517 $\mu\text{mol kg}^{-1}$
242 TA at a salinity of 20.25 close to the coastal mainland. Comparing both studies, one can say that the general level of TA was
243 in a similar range.

244 A conservative mixing between TA and salinity is only visible in the Ems-Dollard Inlet and the Vlie Inlet (Fig. 3). The
245 conservative mixing in the Vlie Inlet can be explained by the fact that more North Sea water pass through the deeper inlets
246 and transport more seawater towards the coast. The Vlie Inlet has the highest average tidal prism and is the second largest inlet
247 after the Marsdiep Inlet in the Dutch Wadden Sea (Elias et al., 2012). Similar to our findings, Hoppema (1990) noted a linear
248 mixing of TA and salinity in the Vlie Inlet, and suspected a lower freshwater contribution there as well.

249 In the Ems-Dollard Inlet, conservative mixing was observed indicating minor contributions from other sources. In a previous
250 study, Norbistrath et al. (2023) observed very high TA contents and TA generation in the upper tidal river of the highly turbid
251 Ems Estuary, which may explain the high levels of TA in the estuary (at low salinities) observed in this study.

252 Hoppema (1990) also identified varying TA contents within the Dutch Wadden Sea and related these to different sinks and
253 sources. TA sinks can be calcium carbonate (CaCO_3) precipitation, or extraction of seawater carbonate by mollusks (e.g., Chen
254 and Wang, 1999; Hoppema, 1990). Variable freshwater inflows can either serve as a sink or a source (e.g., Chen and Wang,

255 1999;Hoppema, 1990). Other TA sources can be CaCO₃ dissolution, anaerobic metabolic processes in the sediment, or erosion
256 of TA enhancing sediments (e.g., Hoppema, 1990;Chen and Wang, 1999).
257 Since we observed constant marine salinities (> 30), and higher TA values in the Dutch Wadden Sea than in the North Sea, we
258 exclude TA sinks and focus only on TA sources. According to Hoppema (1990), the main causes for TA variations in the
259 Dutch Wadden Sea were freshwater inflows and sources in the sediment. In our study, freshwater inflows with high TA
260 contents were only observed in the Ems Estuary and Ems-Dollard Inlet, but not around the islands and the tidal flats. For a
261 further TA source identification in the Dutch Wadden Sea, we investigated the TA variability during a tidal cycle close to
262 Ameland.

263 **4.2 Determination of TA generation**

264 Burt et al. (2016) and Schwichtenberg et al. (2020) assumed TA generation in the Wadden Sea as an important source for the
265 North Sea's carbon storage capacity. Here, we want to further identify TA generation and potential TA sources.

266 In a study from the early 1990s, Hoppema (1993) observed a tidal cycle in the Marsdiep in May and September. Focusing on
267 TA, DIC, and oxygen, he also observed increasing TA values during ebb tide and assumed the tidal flats and discharging rivers
268 and canals as TA sources. Our present TA data and the historical TA data show no large differences in the range of values
269 observed during a tidal cycle. However, an in-depth interpretation and comparison of both data sets would exceed the capacity
270 of these data, leading us to focus on TA generation during our cruise.

271 Our observation of a TA generation of 7.6 μmol kg⁻¹ h⁻¹ TA during ebb tide supports the assumption of the Wadden Sea being
272 a TA source for the North Sea. With the tidal prism of the Ameland Inlet estimated by Louters and Gerritsen (1994), we
273 estimated an upper bound TA export from the Dutch Wadden Sea into the North Sea on the order of 23.9 mol TA during ebb
274 tide (3.5 mol h⁻¹ TA) between spring and summer. This TA export is a rough maximum estimate, because some waters in the
275 main tide channels have no direct contact to the areas of the tidal flats. Schwichtenberg et al. (2020) assumed an annual export
276 of 10 to 14 Gmol yr⁻¹ TA for the entire Dutch Wadden Sea. However, an inclusion of our TA export into the model used by
277 Schwichtenberg et al. (2020) would be unreliable, since our TA export based only on one tidal observation (personal
278 communication J. Pätsch, 2022). To test whether the observed TA generation matches their suggested TA export, more
279 observational data are required. We suggest at least seasonal observational data in order to run the model and gain a
280 representative result (personal communication J. Pätsch, 2022) as future work.

281 **4.3 TA source attribution**

282 **4.3.1 Local sediment outwash**

283 In order to gain further insight into potential sources of TA, we included nutrients in our investigation. The main focus was on
284 silicate that we used as a natural tracer since it is not directly provided anthropogenically and allowed us to determine the
285 silicate source (Van Der Zee and Chou, 2005). We identified a silicate increase of 1.4 μmol L⁻¹ h⁻¹ Si during ebb tide. Silicate

286 originates from dissolution of diatoms and pore water exchange in the Wadden Sea (Van Bennekom et al., 1974; Van Der Zee
287 and Chou, 2005). Here, we relate the silicate increase to pore water exchange, because it markedly increased during ebb tide.
288 This assumption is supported by Van Bennekom et al. (1974), who suggested silicate diffusion from interstitial water in the
289 sediment as potential source, since very high silicate concentrations were found in deeper sediment layers (Rutgers van der
290 Loeff, 1974). Due to the absence of large estuaries nearby and salinity consistently being above 32, we exclude freshwater
291 runoff as a major silicate source. This can be supported by the relation between silicate and salinity in which we observed a
292 non-conservative behavior (Table B1). Since TA behaves also non-conservative relative to salinity (Table B1), the silicate
293 observations support the occurrence of TA sources in the tidal flats of the Wadden Sea.

294 Submarine groundwater discharge (SGD) was identified as a source for nutrient fluxes in tidal flat ecosystems in previous
295 studies (e.g., Billerbeck et al., 2006; Røy et al., 2008; Santos et al., 2021; Waska and Kim, 2011; Wu et al., 2013). Waska and
296 Kim (2011) identified strong SGD contributing up to 50 to 70 % of the nutrient fluxes that fuel primary production in a tidal
297 embayment (Hampyeong Bay) in southwest Korea. In May, they observed low salinities indicating freshwater. However, in
298 September they observed constant marine salinities referring them to be exclusively composed of recirculating seawater. Since
299 we constantly observed marine salinities, we suspect that recirculating marine groundwater enriched with nutrients act as a
300 source for our observed increasing TA and nutrients parameters.

301 TA generation in tidal flats was also observed by Faber et al. (2014), who focused on a large macro tidal embayment in southern
302 Australia. They also found increasing TA values during ebb tide, associated the TA increase with a higher fraction of pore
303 water, and determined the tidal cycle as the controlling force for pore water exchange. Their findings and the observed silicate
304 outwash support our assumption that TA is generated in the sediments of the tidal flats and is washed out during ebb tide. In
305 addition, we exclude lateral advected signals from more western regions as the Vlie Inlet, since the TA contents in the surface
306 transect samples in the Vlie Inlet (except of the two samples close to the coastal mainland near Harlingen) were in the same
307 range as the other observed TA contents and were below the increasing TA contents during ebb tide. Both increases in TA and
308 silicate are clearly tidal signals, and we clearly identify TA generation in the sediments of the tidal flats here as local TA
309 sources.

310 **4.3.2 TA generating processes**

311 The observed TA generation of $7.6 \mu\text{mol kg}^{-1} \text{h}^{-1}$ TA and the silicate increase of $1.4 \mu\text{mol L}^{-1} \text{h}^{-1}$ Si indicated an excess of TA
312 given a TA:silicate ratio of 2:1 (Marx et al. 2017), and under the condition of silicate being bound in minerals, which would
313 then account for a TA generation of $2.8 \mu\text{mol kg}^{-1} \text{h}^{-1}$ TA. However, when silicate occurred dissolved in water it does not
314 contribute to TA generation (Meister et al., 2022). A TA excess related to silicate was also observed in the correlation between
315 TA and silicate (Fig. 5b). Since we observed more TA generated than silicate being washed out, other biogeochemical
316 processes must be responsible for the TA generation in the sediments of the tidal flats in the Dutch Wadden Sea.

317

318 With the observed omega values, we exclude CaCO_3 dissolution as TA source in the overlying water, since the omega values
319 were clearly supersaturated with $\Omega > 1$ (Fig. 4f, Table B1). The continuous calcite supersaturation nicely indicated the inflow
320 and dominance of North Sea water during the flood, with omega values similar to previously observed North Sea values ($\Omega \sim$
321 3.5 to 4) (Charalampopoulou et al., 2011; Carter et al., 2014). However, because of the Ω supersaturation of the overlying water
322 and a lack of pore water data, we were unable to determine if TA generation by CaCO_3 dissolution occurs in deeper sediment
323 layers. There, CaCO_3 dissolution can only be driven metabolically, when CO_2 is produced during OM remineralization, or
324 when reduced compounds that were previously produced during anaerobic processes are oxidized and lead to undersaturation
325 with respect to carbonates (Brenner et al., 2016; Jahnke et al., 1994).

326 A more detailed interpretation of ΔDIC , ΔTA , and various nutrient ratios, further narrows down the potential sources of TA
327 generation in the sediments and used an upper bound estimate for CaCO_3 dissolution. The correlation of DIC and TA reveals
328 an excess of released DIC compared to TA (Fig. 5a), as indicated by the slope of 1.87, while we observed an increase in DIC
329 (ΔDIC) almost twice as high as in TA (ΔTA). The high ΔDIC points to high aerobic OM degradation and remineralization,
330 resulting in high CO_2 export. High aerobic OM degradation was also previously observed in the heterotrophic Wadden Sea
331 (e.g., De Beer et al., 2005; Van Beusekom et al., 1999), assuming an OM degradation and remineralization occurring in the
332 water and sediment in about equal parts (Van Beusekom et al., 1999). High OM degradation is also indicated by the increasing
333 $\text{C}_{\text{org}}:\text{N}$ ratios of SPM during ebb tide (Fig. 4h, Table B1). Because we observed constant coastal North Sea salinities, we rule
334 out freshwater runoff and terrestrial signals as source for the increasing $\text{C}_{\text{org}}:\text{N}$ ratios of SPM. We assume that fresh OM is
335 rapidly degraded in the water column, and the older OM settles on and in the sediment where the degradation continuous and
336 where it is resuspended by the water exchange with outflowing water. Therefore, we assume that the increase of SPM
337 concentrations and their $\text{C}_{\text{org}}:\text{N}$ ratios is an indicator for older and more refractory OM. The increase in TA contents point to
338 anaerobic processes, CaCO_3 dissolution, or a combination thereof as TA sources occurring in the sediments.

339
340 For the upper bound estimate, we assumed CaCO_3 dissolution in the sediments with the DIC:TA ratio of 1:2 as source of TA.
341 Considering this ratio and the observed ΔTA of $51.6 \mu\text{mol kg}^{-1}$ TA, a potential ΔDIC of $25.8 \mu\text{mol kg}^{-1}$ DIC of the observed
342 ΔTA would be produced by CaCO_3 dissolution. The remaining potential $75.5 \mu\text{mol kg}^{-1}$ DIC ($101.3 - 25.8 \mu\text{mol kg}^{-1}$ DIC) of
343 the observed (ΔDIC) could then be produced by OM degradation and remineralization, and would, using the expected Redfield
344 ratio of C:N (6.6), correspond to an estimated potential dissolved inorganic nitrogen (DIN) production of $11.4 \mu\text{mol kg}^{-1}$ DIN.
345 However, this estimated potential DIN production ($11.4 \mu\text{mol kg}^{-1}$ DIN) of OM degradation and remineralization exceeded
346 the observed increase of ΔDIN ($3.97 \mu\text{mol L}^{-1}$ DIN; Table B1, sum of NO_3^- , NO_2^- and NH_4^+) during ebb tide. With this
347 estimation and the assumption that all DIN produced is released and thus lost, TA is probably produced by CaCO_3 dissolution
348 and anaerobic metabolic processes other than denitrification in the sediment. In addition to that, and with a N-focused
349 perspective, the DIN loss also hints to the occurrence of other processes that consume nitrogen species but have no net effect
350 on TA, such as anammox and coupled nitrification-denitrification (Hu and Cai, 2011; Middelburg et al., 2020). The suggested
351 DIN loss can be supported by considering the marine DIN:Si ratio, which is supposed to be 1:1 (Brzezinski, 1985). We

352 observed DIN:Si ratios decreasing from 2.7 to 0.8 from high tide to low tide. The decreasing ratios show that both parameter
353 concentrations increased during ebb tide, whereby DIN concentrations increased lower than silicate concentrations. We
354 observed a silicate excess with respect to DIN at the end of ebb tide, supporting the DIN loss.

355 Denitrification, the anaerobic irreversible reduction of NO_3^- to N_2 that generates 0.9 mole TA by using 1 mole NO_3^- as electron
356 acceptor (Chen and Wang, 1999) is a net TA source. Denitrification depends on the supply of nitrate, which seasonally varies
357 (Van Der Zee and Chou, 2005 and references therein). Generally, nitrate is depleted in summer due to high turnover rates and
358 occurs in higher concentrations in winter (Kieskamp et al., 1991; Jensen et al., 1996; Van Der Zee and Chou, 2005). This
359 seasonality lead to denitrification rates also being lower in summer and higher in winter (Kieskamp et al., 1991; Jensen et al.,
360 1996). In previous studies, Faber et al. (2014) identified denitrification as a minor source of TA due to low denitrification rates,
361 and also Kieskamp et al. (1991) observed low denitrification rates in the Wadden Sea, with low nitrate concentrations (< 2.5
362 $\mu\text{mol L}^{-1}$) in the overlying water. We observed nitrate concentrations ($< 2.17 \mu\text{mol L}^{-1}$) lower than the concentration sufficient
363 for denitrification assumed by Kieskamp et al. (1991). Therefore, we do not exclude denitrification, but suspect it as a minor
364 source of TA in the Dutch Wadden Sea at least in spring and summer due to the seasonal lack of nitrate. Thomas et al. (2009)
365 detected TA seasonality in the southern bight of the North Sea, which is also influenced by the TA generation in the Wadden
366 Sea. We support their findings of lowered TA generation by denitrification in late spring and early summer. In addition, the
367 calculated potential DIN excess compared to the observed DIN not only hints to other N consuming processes that have no
368 effect on TA, but also suggests that allochthonous nitrate would be needed to fuel the TA increase by denitrification. In
369 addition, the albeit low availability of nitrate indicates to predominantly aerobic metabolic activity during the time of our
370 observations, which is in line with earlier studies reporting an enhanced relevance of anaerobic activity later in summer (Luff
371 and Moll, 2004; Thomas et al., 2009).

372

373 Another source of TA in sediments is aerobic OM respiration with the associated formation of ammonium while consuming
374 H^+ (Blackburn and Henriksen, 1983; Berner et al., 1970; Brenner et al., 2016). The observed increasing ammonium
375 concentrations (Fig. 4e) could be associated with aerobic OM degradation leading to ammonium formation. The resulting
376 ammonium formation in the upper oxygenated sediment layers would increase DIC by the production of CO_2 , and increase
377 TA by the consumption of H^+ (Fig.5) (Brenner et al., 2016). In sediments, the production of one mole ammonium (from
378 ammonia) would then generate one mole TA (Berner et al., 1970; Meister et al., 2022). In contrast, in the water column, the
379 aerobic respiration of OM produce CO_2 and increase DIC, also visible in decreasing pH values (Fig. 4g), but consume TA and
380 would not produce ammonium (Chen and Wang, 1999). Therefore, aerobic OM respiration in the water column could only
381 explain the higher increase in DIC than in TA, but not the simultaneous increase in TA and ammonium (Fig. 5). Based on this,
382 we assume that OM respiration associated with TA generation occurs in the sediments, leading to TA and DIC generation and
383 also to ammonium production, being washed-out during ebb tide. The produced ammonium is then also accessible for
384 nitrification that produces nitrate. A slightly increased nitrate concentration in the most upper sediment layers was observed
385 by Beck et al. (2008a) in the German Wadden Sea. This observation, a potential nitrate reservoir, nitrate production due to OM

386 degradation and nitrification occurring in the upper oxygenated sediment layers (Martin and Sayles, 1996), or a combination
387 thereof could explain the observed low increasing nitrate concentrations during ebb tide. However, as we rule out terrestrial
388 nitrate imports as nitrate source here, the simultaneous increase of TA and nitrate is noticeable for us, because nitrification
389 consumes TA (Brenner et al., 2016). We assume that potential nitrification has a minor effect on TA, since we observed only
390 low nitrate concentrations and a really low increase of nitrate compared to the increase of ammonium and TA during ebb tide.
391 Low nitrate concentrations resulting in a reduced availability of bound oxygen, i.e., electron acceptors. This promotes the
392 occurrence of other anaerobic processes of the redox system to generate TA in the deeper, anoxic sediment layers in the Dutch
393 Wadden Sea, such as sulfate and iron reduction.

394 Sulfate reduction followed by iron reduction and the formation and burial of pyrite are net sources of TA, since TA
395 consumption by reoxidation is excluded when buried in sediments (Berner et al., 1970;Faber et al., 2014). Whether these
396 processes contribute to TA generation in the deeper sediments of the Dutch Wadden Sea cannot be further identified without
397 the necessary data. However, sulfate reduction was also mentioned as source of TA by Thomas et al. (2009). The temporary
398 slight appearance of noticeable sulfuric odor could be another indirect indicator for the occurrence of sulfate reduction. In
399 previous studies of tidal flats in the German Wadden Sea, Beck et al. (2008a);(2008b) observed increasing TA contents with
400 depth and identified sulfate reduction as the most important process for anaerobic OM remineralization in pore water cores.
401 Sulfate reduction releases 1.14 mole TA with the oxidation of one mole carbon of POC , and iron reduction releases 8.14 mole
402 TA with the oxidation of one mole carbon of POC, indicating that both processes are large sources of TA generation (Brenner
403 et al., 2016), but further studies are needed to support this.

404

405 A strict comparison of the northern and the western parts of the Wadden Sea is difficult because the areas vary in terms of OM
406 import and eutrophication effects (Van Beusekom et al., 2019), sediment composition, and extent between the barrier islands
407 and the mainland, all of which influence the occurrence and interaction of biogeochemical processes (Schwichtenberg et al.,
408 2020). The area characteristics of the northern and western Wadden Sea differ especially in terms of OM turnover being lower
409 in the norther Wadden Sea. A previous study by Brasse et al. (1999) identified high TA and DIC contents in the sediment of
410 the North Frisian Wadden Sea and identified CaCO_3 dissolution and sulfate reduction as major TA sources, which is consistent
411 with our findings.

412 **5 Conclusion**

413 The Dutch Wadden Sea is a unique and highly dynamic ecosystem. While observing the spatial TA distribution and TA
414 generation in the Dutch Wadden Sea, we detected higher TA values than in the North Sea and identified the Dutch Wadden
415 Sea clearly as a TA source for the North Sea's carbonate system. Compared to previous studies (Hoppema, 1990, 1993), the
416 TA values we observed were in a similar range, with high TA values in the tidal basins. Beside the need for seasonal

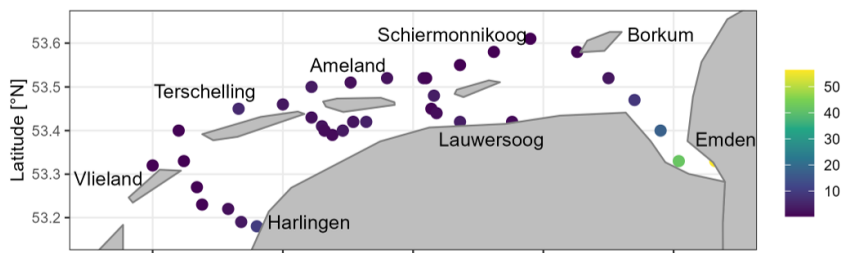
417 observations, future work should also focus on the tidal end-members to better understand the general and seasonal influence
418 of freshwater inflows on the TA status in the Dutch Wadden Sea.

419 By observing salinity and using silicate as a tracer, we excluded freshwater dilution and river runoff as TA sources on the tidal
420 flats, and instead, identified local outwash from the sediments as sources of TA. Aerobic, metabolic processes such as CaCO_3
421 dissolution and ammonium formation seem to dominate TA generation in the upper oxic sediment layers and the overlying
422 water, while anaerobic, metabolic processes such as denitrification, sulfate and iron reduction are potential TA sources in the
423 deeper anoxic sediment layers. However, in spring and early summer, denitrification seems to play a minor role in generating
424 TA in the sediments of the Dutch Wadden Sea due to seasonality and associated limited nitrate availability.

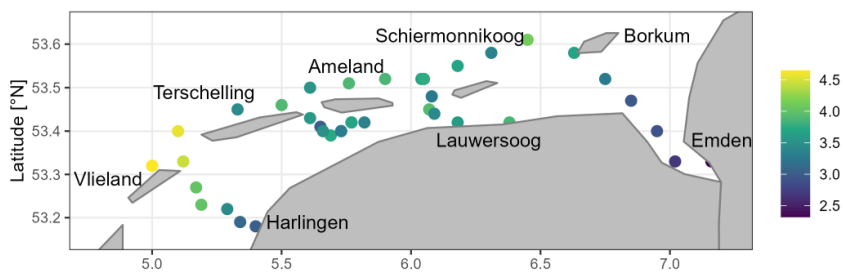
425 **6 Appendices**

426 **Appendix A**

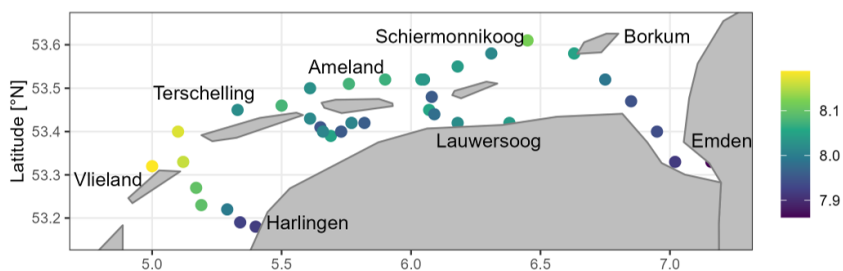
(A1) Silicate ($\mu\text{ mol L}^{-1}$)



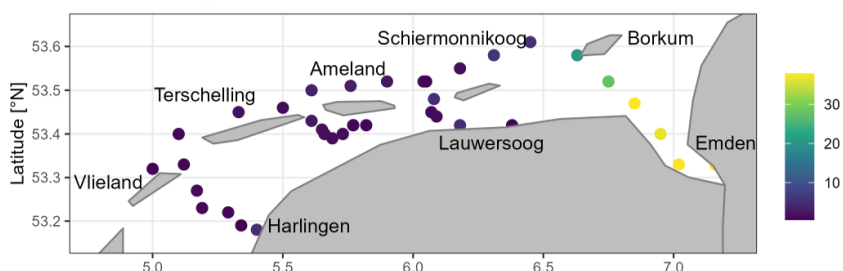
(A2) Calcite saturation state (Ω)



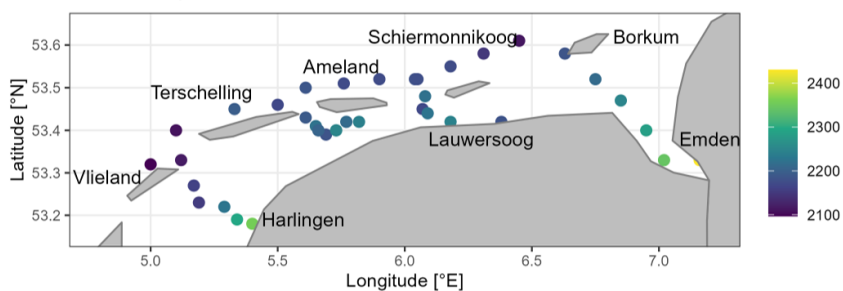
(A3) pH



(A4) Nitrate ($\mu\text{ mol L}^{-1}$)



(A5) DIC ($\mu\text{ mol kg}^{-1}$)



428 **Figure A** Latitudinal and longitudinal distribution of A1) silicate (Si; $\mu\text{mol L}^{-1}$), A2) calcite saturation state (Ω), A3) pH, A4)
 429 nitrate (NO_3^- ; $\mu\text{mol L}^{-1}$), and A5) dissolved inorganic carbon (DIC; $\mu\text{mol kg}^{-1}$) from surface water samples in May 2019.

430 **Appendix B**

431 **Table B1** Tidal cycle sample parameter during ebb tide. Sample no. 545 is the first sample at high tide and sample no. 557 is
 432 the last sample at ebb tide on May 21st 2019. Shown are rounded up values of temperature (Temp), salinity (Sal), total alkalinity
 433 (TA), dissolved inorganic carbon (DIC), silicate (Si), nitrate (NO_3^-), nitrite (NO_2^-), ammonium (NH_4^+), dissolved inorganic
 434 nitrogen (DIN), the amount of carbon (C) and organic carbon (C_{org}) of SPM, the amount of nitrogen (N) of SPM, the calcite
 435 (Ca) and aragonite (Ar) saturation states, the pH, and phosphate (PO_4^{3-}) per sample.

Sample No.	Temp [°C]	Sal	TA [$\mu\text{mol kg}^{-1}$]	DIC [$\mu\text{mol kg}^{-1}$]	Si [$\mu\text{mol L}^{-1}$]	NO_3^- [$\mu\text{mol L}^{-1}$]	NO_2^- [$\mu\text{mol L}^{-1}$]	NH_4^+ [$\mu\text{mol L}^{-1}$]
545	13.26	32.52	2387	2172	1.84	1.26	0.19	3.47
546	13.25	32.52	2385	2190	1.77	1.24	0.19	3.40
547	13.28	32.52	2389	2185	1.72	1.21	0.19	3.35
548	13.38	32.52	2391	2183	1.6	1.19	0.19	3.52
549	14.32	32.50	2400	2204	2.11	0.91	0.25	3.57
550	14.61	32.50	2400	2221	2.78	1.09	0.29	3.98
551	14.64	32.51	2405	2216	2.72	1.01	0.29	4.27
552	14.73	32.51	2411	2234	4.59	1.23	0.34	5.51
553	14.77	32.51	2402	2228	4.24	1.26	0.33	5.08
554	14.72	32.51	2419	2234	5.66	1.46	0.36	5.33
555	14.66	32.51	2428	2256	8.18	1.77	0.43	6.04
556	14.68	32.51	2433	2271	9.79	1.87	0.47	6.27
557	14.70	32.50	2438	2273	11.22	2.17	0.50	6.22
Sample No.	DIN [$\mu\text{mol L}^{-1}$]	C / C_{org} (SPM) [$\mu\text{mol L}^{-1}$]	N (SPM) [$\mu\text{mol L}^{-1}$]	$\text{C}_{\text{org}}:\text{N}$ (SPM)	SPM [mg L^{-1}]	Ca / Ar [Ω]	pH	PO_4^{3-} [$\mu\text{mol L}^{-1}$]
545	4.93	86.8 / 65.1	8.8	7.4	12.8	3.8 / 2.4	8.07	0.12
546	4.83	72.7 / 42.4	7.4	5.8	8.7	3.5 / 2.3	8.03	0.11
547	4.76	112.4 / 93.4	9.6	9.7	15.4	3.7 / 2.3	8.05	0.11
548	4.91	108.5 / 104.6	9.9	10.5	16.8	3.7 / 2.4	8.05	0.12
549	4.73	111.1 / 97.8	8.8	11.1	13.9	3.6 / 2.3	8.01	0.32
550	5.37	233.0 / 180.3	17.7	10.2	32.2	3.3 / 2.1	7.97	0.42
551	5.56	193.2 / 174.3	14.5	12.0	29.6	3.5 / 2.2	7.99	0.47
552	7.08	248.6 / 163.5	18.4	8.9	34.3	3.3 / 2.1	7.96	0.57
553	6.67	257.6 / 199.3	18.3	10.9	41.6	3.2 / 2.1	7.95	0.54
554	7.15	324.4 / 271.1	23.2	11.7	55.0	3.4 / 2.2	7.98	0.54

555	8.24	440.4 / 345.2	29.2	11.8	75.7	3.2 / 2.1	7.95	0.58
556	8.61	430.5 / 363.3	27.9	13.0	82.4	3.1 / 2.0	7.93	0.62
557	8.90	308.9 / 199.1	21.2	9.4	48.8	3.1 / 2.0	7.93	0.63

436 **Data availability**

437 The data of this study are either presented in the article or are available upon request from the corresponding author.

438 **Author Contributions**

439 MN wrote the manuscript, did the carbon sampling and sample measurement, analyzed and evaluated the data, and led the
440 study. JvB led the research cruise. JvB and HT contributed with editorial and scientific recommendations. MN prepared the
441 manuscript with contribution from all co-authors.

442 **Competing interests**

443 The contact author has declared that none of the authors has any competing interests.

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451 **References**

452 Abril, G., and Frankignoulle, M.: Nitrogen–alkalinity interactions in the highly polluted Scheldt basin (Belgium), *Water Research*, 35, 844-
453 850, [https://doi.org/10.1016/S0043-1354\(00\)00310-9](https://doi.org/10.1016/S0043-1354(00)00310-9), 2001.
454 Beck, M., Dellwig, O., Holstein, J. M., Grunwald, M., Liebezeit, G., Schnetger, B., and Brumsack, H.-J.: Sulphate, dissolved organic carbon,
455 nutrients and terminal metabolic products in deep pore waters of an intertidal flat, *Biogeochemistry*, 89, 221-238,
456 <https://doi.org/10.1007/s10533-008-9215-6>, 2008a.

457 Beck, M., Dellwig, O., Liebezeit, G., Schnetger, B., and Brumsack, H.-J.: Spatial and seasonal variations of sulphate, dissolved organic
458 carbon, and nutrients in deep pore waters of intertidal flat sediments, *Estuarine, Coastal and Shelf Science*, 79, 307-316,
459 <https://doi.org/10.1016/j.ecss.2008.04.007>, 2008b.

460 Berner, R. A., Scott, M. R., and Thomlinson, C.: Carbonate alkalinity in the pore waters of anoxic marine sediments I, *Limnology and*
461 *Oceanography*, 15, 544-549, <https://doi.org/10.4319/lo.1970.15.4.0544>, 1970.

462 Berner, R. A., Lasaga, A. C., and Garrels, R. M.: Carbonate-silicate geochemical cycle and its effect on atmospheric carbon dioxide over
463 the past 100 million years, *Am. J. Sci.:(United States)*, 283, doi:10.2475/ajs.283.7.641., 1983.

464 Billerbeck, M., Werner, U., Polerecky, L., Walpersdorf, E., DeBeer, D., and Huettel, M.: Surficial and deep pore water circulation governs
465 spatial and temporal scales of nutrient recycling in intertidal sand flat sediment, *Marine Ecology Progress Series*, 326, 61-76, 2006.

466 Blackburn, T., and Henriksen, K.: Nitrogen cycling in different types of sediments from Danish waters 1, *Limnology and Oceanography*,
467 28, 477-493, <https://doi.org/10.4319/lo.1983.28.3.0477>, 1983.

468 Borges, A. V., Delille, B., and Frankignoulle, M.: Budgeting sinks and sources of CO₂ in the coastal ocean: Diversity of ecosystems counts,
469 *Geophysical research letters*, 32, doi.org/10.1029/2005GL023053, 2005.

470 Bozec, Y., Thomas, H., Elkalay, K., and de Baar, H. J.: The continental shelf pump for CO₂ in the North Sea—evidence from summer
471 observation, *Marine Chemistry*, 93, 131-147, <https://doi.org/10.1016/j.marchem.2004.07.006>, 2005.

472 Brasse, S., Reimer, A., Seifert, R., and Michaelis, W.: The influence of intertidal mudflats on the dissolved inorganic carbon and total
473 alkalinity distribution in the German Bight, southeastern North Sea, *Journal of Sea Research*, 42, 93-103, 1999.

474 Brenner, H., Braeckman, U., Le Guitton, M., and Meysman, F. J.: The impact of sedimentary alkalinity release on the water column CO₂
475 system in the North Sea, *Biogeosciences*, 13, 841-863, <https://doi.org/10.5194/bg-13-841-2016>, 2016.

476 Brewer, P. G., and Goldman, J. C.: Alkalinity changes generated by phytoplankton growth, *Limnology and Oceanography*, 21, 108-117,
477 <https://doi.org/10.4319/lo.1976.21.1.0108>, 1976.

478 Brzezinski, M. A.: The Si: C: N ratio of marine diatoms: interspecific variability and the effect of some environmental variables, *Journal of*
479 *Phycology*, 21, 347-357, <https://doi.org/10.1111/j.0022-3646.1985.00347.x>, 1985.

480 Burchard, H., Flöser, G., Staneva, J. V., Badewien, T. H., and Riethmüller, R.: Impact of density gradients on net sediment transport into
481 the Wadden Sea, *Journal of Physical Oceanography*, 38, 566-587, 2008.

482 Burt, W., Thomas, H., Hagens, M., Pätsch, J., Clargo, N., Salt, L., Winde, V., and Böttcher, M.: Carbon sources in the North Sea evaluated
483 by means of radium and stable carbon isotope tracers, *Limnology and Oceanography*, 61, 666-683, <https://doi.org/10.1002/lno.10243>, 2016.

484 Carter, B. R., Toggweiler, J., Key, R. M., and Sarmiento, J. L.: Processes determining the marine alkalinity and calcium carbonate saturation
485 state distributions, *Biogeosciences*, 11, 7349-7362, <https://doi.org/10.5194/bg-11-7349-2014>, 2014.

486 Charalampopoulou, A., Poulton, A. J., Tyrrell, T., and Lucas, M. I.: Irradiance and pH affect coccolithophore community composition on a
487 transect between the North Sea and the Arctic Ocean, *Marine Ecology Progress Series*, 431, 25-43, <https://doi.org/10.3354/meps09140>,
488 2011.

489 Chen, C. T. A., and Wang, S. L.: Carbon, alkalinity and nutrient budgets on the East China Sea continental shelf, *Journal of Geophysical*
490 *Research: Oceans*, 104, 20675-20686, <https://doi.org/10.1029/1999JC900055>, 1999.

491 Crutzen, P.: Geology of mankind, *Nature*, 415, <https://doi.org/10.1038/415023a>, 2002.

492 De Beer, D., Wenzhöfer, F., Ferdelman, T. G., Boehme, S. E., Huettel, M., van Beusekom, J. E., Böttcher, M. E., Musat, N., and Dubilier,
493 N.: Transport and mineralization rates in North Sea sandy intertidal sediments, Sylt-Rømø basin, Wadden Sea, *Limnology and*
494 *Oceanography*, 50, 113-127, 2005.

495 De Jonge, V., Essink, K., and Boddeke, R.: The Dutch Wadden Sea: a changed ecosystem, *Hydrobiologia*, 265, 45-71,
496 <https://doi.org/10.1007/BF00007262>, 1993.

497 Dickson, A., and Millero, F. J.: A comparison of the equilibrium constants for the dissociation of carbonic acid in seawater media, *Deep Sea*
498 *Research Part A. Oceanographic Research Papers*, 34, 1733-1743, [https://doi.org/10.1016/0198-0149\(87\)90021-5](https://doi.org/10.1016/0198-0149(87)90021-5), 1987.

499 Dickson, A. G.: An exact definition of total alkalinity and a procedure for the estimation of alkalinity and total inorganic carbon from titration
500 data, *Deep Sea Research Part A. Oceanographic Research Papers*, 28, 609-623, [https://doi.org/10.1016/0198-0149\(81\)90121-7](https://doi.org/10.1016/0198-0149(81)90121-7), 1981.

501 Elias, E. P., Van der Spek, A. J., Wang, Z. B., and De Ronde, J.: Morphodynamic development and sediment budget of the Dutch Wadden
502 Sea over the last century, *Netherlands Journal of Geosciences*, 91, 293-310, <https://doi.org/10.1017/S0016774600000457>, 2012.

503 Faber, P. A., Evrard, V., Woodland, R. J., Cartwright, I. C., and Cook, P. L.: Pore-water exchange driven by tidal pumping causes alkalinity
504 export in two intertidal inlets, *Limnology and Oceanography*, 59, 1749-1763, <https://doi.org/10.4319/lo.2014.59.5.1749>, 2014.

505 Glavovic, B., Limburg, K., Liu, K., Emeis, K., Thomas, H., Kremer, H., Avril, B., Zhang, J., Mulholland, M., and Glaser, M.: Living on the
506 Margin in the Anthropocene: engagement arenas for sustainability research and action at the ocean-land interface, *Current Opinion in*
507 *Environmental Sustainability*, 14, 232-238, <https://doi.org/10.1016/j.cosust.2015.06.003>, 2015.

508 Grasshoff, K., Kremling, K., and Ehrhardt, M.: *Methods of seawater analysis*, John Wiley & Sons, 2009.

509 Gruber, N., Clement, D., Carter, B. R., Feely, R. A., Van Heuven, S., Hoppema, M., Ishii, M., Key, R. M., Kozyr, A., and Lauvset, S. K.:
510 The oceanic sink for anthropogenic CO₂ from 1994 to 2007, *Science*, 363, 1193-1199, DOI: 10.1126/science.aau5153, 2019.

511 Hansen, H., and Koroleff, F.: *Determination of nutrients. Methods of Seawater Analysis: Third, Completely Revised and Extended Edition.*
512 Grasshoff, K., Kremling, K., and Ehrhardt, M. (Eds.), Weinheim, Germany: Wiley-VCH Verlag GmbH, 2007.

513 Hoppema, J.: The distribution and seasonal variation of alkalinity in the southern bight of the North Sea and in the western Wadden Sea,
514 Netherlands journal of sea research, 26, 11-23, [https://doi.org/10.1016/0077-7579\(90\)90053-J](https://doi.org/10.1016/0077-7579(90)90053-J), 1990.

515 Hoppema, J.: The oxygen budget of the western Wadden Sea, The Netherlands, Estuarine, Coastal and Shelf Science, 32, 483-502,
516 [https://doi.org/10.1016/0272-7714\(91\)90036-B](https://doi.org/10.1016/0272-7714(91)90036-B), 1991.

517 Hoppema, J.: Carbon dioxide and oxygen disequilibrium in a tidal basin (Dutch Wadden Sea), Netherlands journal of sea research, 31, 221-
518 229, [https://doi.org/10.1016/0077-7579\(93\)90023-L](https://doi.org/10.1016/0077-7579(93)90023-L), 1993.

519 Hu, X., and Cai, W. J.: An assessment of ocean margin anaerobic processes on oceanic alkalinity budget, Global Biogeochemical Cycles,
520 25, doi.org/10.1029/2010GB003859, 2011.

521 Jahnke, R. A., Craven, D. B., and Gaillard, J.-F.: The influence of organic matter diagenesis on CaCO₃ dissolution at the deep-sea floor,
522 Geochimica et Cosmochimica Acta, 58, 2799-2809, 1994.

523 Jensen, K., Jensen, M., and Kristensen, E.: Nitrification and denitrification in Wadden Sea sediments (Königshafen, Island of Sylt, Germany)
524 as measured by nitrogen isotope pairing and isotope dilution, Aquatic Microbial Ecology, 11, 181-191, doi:10.3354/ame011181, 1996.

525 Keith, D. W., Ha-Duong, M., and Stolaroff, J. K.: Climate strategy with CO₂ capture from the air, Climatic Change, 74, 17-45,
526 <https://doi.org/10.1007/s10584-005-9026-x>, 2006.

527 Kérouel, R., and Aminot, A.: Fluorometric determination of ammonia in sea and estuarine waters by direct segmented flow analysis, Marine
528 Chemistry, 57, 265-275, [https://doi.org/10.1016/S0304-4203\(97\)00040-6](https://doi.org/10.1016/S0304-4203(97)00040-6), 1997.

529 Kieskamp, W. M., Lohse, L., Epping, E., and Helder, W.: Seasonal variation in denitrification rates and nitrous oxide fluxes in intertidal
530 sediments of the western Wadden Sea, Marine ecology progress series. Oldendorf, 72, 145-151, 1991.

531 Lewis, E., and Wallace, D.: Program developed for CO₂ system calculations, Environmental System Science Data Infrastructure for a Virtual
532 Ecosystem, 1998.

533 Louters, T., and Gerritsen, F.: The Riddle of the Sands: A Tidal System Vs Answer to a Rising Sea Level, report RIKZ-94.040 (isbn 90-369-
534 0084-0), 1994.

535 Luff, R., and Moll, A.: Seasonal dynamics of the North Sea sediments using a three-dimensional coupled sediment–water model system,
536 Continental Shelf Research, 24, 1099-1127, <https://doi.org/10.1016/j.csr.2004.03.010>, 2004.

537 Martin, W., and Sayles, F.: CaCO₃ dissolution in sediments of the Ceara Rise, western equatorial Atlantic, Geochimica et Cosmochimica
538 Acta, 60, 243-263, [https://doi.org/10.1016/0016-7037\(95\)00383-5](https://doi.org/10.1016/0016-7037(95)00383-5), 1996.

539 Matthews, H. D., and Caldeira, K.: Stabilizing climate requires near-zero emissions, Geophysical research letters, 35,
540 <https://doi.org/10.1029/2007GL032388>, 2008.

541 Mehrbach, C., Culberson, C., Hawley, J., and Pytkowicz, R.: Measurement of the apparent dissociation constants of carbonic acid in seawater
542 at atmospheric pressure, Limnology and oceanography, 18, 897-907, <https://doi.org/10.4319/lo.1973.18.6.0897>, 1973.

543 Meister, P., Herda, G., Petrishcheva, E., Gier, S., Dickens, G. R., Bauer, C., and Liu, B.: Microbial Alkalinity Production and Silicate
544 Alteration in Methane Charged Marine Sediments: Implications for Porewater Chemistry and Diagenetic Carbonate Formation, Geochemical
545 Signals in Dynamic Sedimentary Systems Along Continental Margins, 2022.

546 Meybeck, M.: Global chemical weathering of surficial rocks estimated from river dissolved loads, American journal of science, 287, 401-
547 428, 10.2475/ajs.287.5.401, 1987.

548 Middelburg, J. J., Soetaert, K., and Hagens, M.: Ocean alkalinity, buffering and biogeochemical processes, Reviews of Geophysics, 58,
549 e2019RG000681, <https://doi.org/10.1029/2019RG000681>, 2020.

550 Norbistrath, M., Pätsch, J., Dähnke, K., Sanders, T., Schulz, G., van Beusekom, J. E., and Thomas, H.: Metabolic alkalinity release from
551 large port facilities (Hamburg, Germany) and impact on coastal carbon storage, Biogeosciences, 19, 5151-5165, <https://doi.org/10.5194/bg-19-5151-2022>, 2022.

552 Norbistrath, M., Neumann, A., Dähnke, K., Sanders, T., Schöl, A., van Beusekom, J. E., and Thomas, H.: Alkalinity and nitrate dynamics
553 reveal dominance of anammox in a hyper-turbid estuary, Biogeosciences, 20, 4307–4321, <https://doi.org/10.5194/bg-20-4307-2023>, 2023.

554 Petersen, W., Schroeder, F., and Bockelmann, F.-D.: FerryBox-Application of continuous water quality observations along transects in the
555 North Sea, Ocean Dynamics, 61, 1541-1554, <https://doi.org/10.1007/s10236-011-0445-0>, 2011.

556 Postma, H.: Hydrography of the Dutch Wadden sea, Arch. Neerl. Zool, 10, 405-511, 1954.

557 Renforth, P., and Henderson, G.: Assessing ocean alkalinity for carbon sequestration, Reviews of Geophysics, 55, 636-674,
558 <https://doi.org/10.1002/2016RG000533>, 2017.

559 Ridderinkhof, H., Zimmerman, J., and Philippart, M.: Tidal exchange between the North Sea and Dutch Wadden Sea and mixing time scales
560 of the tidal basins, Netherlands Journal of Sea Research, 25, 331-350, [https://doi.org/10.1016/0077-7579\(90\)90042-F](https://doi.org/10.1016/0077-7579(90)90042-F), 1990.

561 Røy, H., Lee, J. S., Jansen, S., and de Beer, D.: Tide-driven deep pore-water flow in intertidal sand flats, Limnology and oceanography, 53,
562 1521-1530, <https://doi.org/10.4319/lo.2008.53.4.1521>, 2008.

563 Rutgers van der Loeff, M.: Transport van reactief silikaat uit Waddenzee sediment naar het bovenstaande water, NIOZ-rapport, 1974.

564 Sabine, C. L., Feely, R. A., Gruber, N., Key, R. M., Lee, K., Bullister, J. L., Wanninkhof, R., Wong, C., Wallace, D. W., and Tilbrook, B.:
565 The oceanic sink for anthropogenic CO₂, science, 305, 367-371, DOI: 10.1126/science.1097403, 2004.

567 Santos, I. R., Chen, X., Lecher, A. L., Sawyer, A. H., Moosdorf, N., Rodellas, V., Tamborski, J., Cho, H.-M., Dimova, N., and Sugimoto,
568 R.: Submarine groundwater discharge impacts on coastal nutrient biogeochemistry, *Nature Reviews Earth & Environment*, 2, 307-323,
569 <https://doi.org/10.1038/s43017-021-00152-0>, 2021.

570 Schwichtenberg, F., Callies, U., and van Beusekom, J. E.: Residence times in shallow waters help explain regional differences in Wadden
571 Sea eutrophication, *Geo-Marine Letters*, 37, 171-177, <https://doi.org/10.1007/s00367-016-0482-2>, 2017.

572 Schwichtenberg, F., Pätzsch, J., Böttcher, M. E., Thomas, H., Winde, V., and Emeis, K.-C.: The impact of intertidal areas on the carbonate
573 system of the southern North Sea, *Biogeosciences*, 17, 4223-4245, <https://doi.org/10.5194/bg-17-4223-2020>, 2020.

574 Shadwick, E., Thomas, H., Gratton, Y., Leong, D., Moore, S., Papakyriakou, T., and Prowe, A.: Export of Pacific carbon through the Arctic
575 Archipelago to the North Atlantic, *Continental Shelf Research*, 31, 806-816, <https://doi.org/10.1016/j.csr.2011.01.014>, 2011.

576 Suchet, P. A., and Probst, J.-L.: Modelling of atmospheric CO₂ consumption by chemical weathering of rocks: application to the Garonne,
577 Congo and Amazon basins, *Chemical Geology*, 107, 205-210, DOI:10.1016/0009-2541(93)90174-H, 1993.

578 Thomas, H., Bozec, Y., Elkalay, K., and De Baar, H. J.: Enhanced open ocean storage of CO₂ from shelf sea pumping, *Science*, 304, 1005-
579 1008, DOI: 10.1126/science.1095491, 2004.

580 Thomas, H., Schiettecatte, L.-S., Suykens, K., Koné, Y., Shadwick, E., Prowe, A. F., Bozec, Y., de Baar, H. J., and Borges, A.: Enhanced
581 ocean carbon storage from anaerobic alkalinity generation in coastal sediments, *Biogeosciences*, 6, 267-274, [https://doi.org/10.5194/bg-6-](https://doi.org/10.5194/bg-6-267-2009)
582 [267-2009](https://doi.org/10.5194/bg-6-267-2009), 2009.

583 Van Bennekom, A., Krijgsman-van Hartingsveld, E., Van der Veer, G., and Van Voorst, H.: The seasonal cycles of reactive silicate and
584 suspended diatoms in the Dutch Wadden Sea, *Netherlands Journal of Sea Research*, 8, 174-207, [https://doi.org/10.1016/0077-](https://doi.org/10.1016/0077-7579(74)90016-7)
585 [7579\(74\)90016-7](https://doi.org/10.1016/0077-7579(74)90016-7), 1974.

586 Van Beusekom, J., Brockmann, U., Hesse, K.-J., Hickel, W., Poremba, K., and Tillmann, U.: The importance of sediments in the
587 transformation and turnover of nutrients and organic matter in the Wadden Sea and German Bight, *Deutsche Hydrografische Zeitschrift*, 51,
588 245-266, 10.1007/BF02764176, 1999.

589 Van Beusekom, J., and De Jonge, V.: Long-term changes in Wadden Sea nutrient cycles: importance of organic matter import from the
590 North Sea, in: *Nutrients and Eutrophication in Estuaries and Coastal Waters*, Springer, 185-194, 2002.

591 Van Beusekom, J. E., Buschbaum, C., and Reise, K.: Wadden Sea tidal basins and the mediating role of the North Sea in ecological processes:
592 scaling up of management?, *Ocean & coastal management*, 68, 69-78, <https://doi.org/10.1016/j.ocecoaman.2012.05.002>, 2012.

593 Van Beusekom, J. E., Carstensen, J., Dolch, T., Grage, A., Hofmeister, R., Lenhart, H., Kerimoglu, O., Kolbe, K., Pätzsch, J., and Rick, J.:
594 Wadden Sea Eutrophication: long-term trends and regional differences, *Frontiers in Marine Science*, 6, 370,
595 doi.org/10.3389/fmars.2019.00370, 2019.

596 Van Der Zee, C., and Chou, L.: Seasonal cycling of phosphorus in the Southern Bight of the North Sea, *Biogeosciences*, 2, 27-42,
597 <https://doi.org/10.5194/bg-2-27-2005>, 2005.

598 van Raaphorst, W., and van der Veer, H. W.: The phosphorus budget of the Marsdiep tidal basin (Dutch Wadden Sea) in the period 1950–
599 1985: importance of the exchange with the North Sea, in: *North Sea—Estuaries Interactions*, Springer, 21-38, 1990.

600 Voynova, Y. G., Petersen, W., Gehrung, M., Aßmann, S., and King, A. L.: Intertidal regions changing coastal alkalinity: The Wadden Sea–
601 North Sea tidally coupled bioreactor, *Limnology and Oceanography*, 64, 1135-1149, 2019.

602 Wang, Z. A., Kroeger, K. D., Ganju, N. K., Gonnee, M. E., and Chu, S. N.: Intertidal salt marshes as an important source of inorganic
603 carbon to the coastal ocean, *Limnology and Oceanography*, 61, 1916-1931, 2016.

604 Waska, H., and Kim, G.: Submarine groundwater discharge (SGD) as a main nutrient source for benthic and water-column primary
605 production in a large intertidal environment of the Yellow Sea, *Journal of Sea Research*, 65, 103-113,
606 <https://doi.org/10.1016/j.seares.2010.08.001>, 2011.

607 Wolf-Gladrow, D. A., Zeebe, R. E., Klaas, C., Körtzinger, A., and Dickson, A. G.: Total alkalinity: The explicit conservative expression and
608 its application to biogeochemical processes, *Marine Chemistry*, 106, 287-300, <https://doi.org/10.1016/j.marchem.2007.01.006>, 2007.

609 Wu, Z., Zhou, H., Zhang, S., and Liu, Y.: Using ²²²Rn to estimate submarine groundwater discharge (SGD) and the associated nutrient fluxes
610 into Xiangshan Bay, East China Sea, *Marine pollution bulletin*, 73, 183-191, <https://doi.org/10.1016/j.marpolbul.2013.05.024>, 2013.

611 Zalasiewicz, J., Williams, M., Steffen, W., and Crutzen, P.: *The new world of the Anthropocene*. ACS Publications, 2010.

612 Zhang, C., Shi, T., Liu, J., He, Z., Thomas, H., Dong, H., Rinkevich, B., Wang, Y., Hyun, J.-H., and Weinbauer, M.: Eco-engineering
613 approaches for ocean negative carbon emission, *Science Bulletin*, 67, 2564-2573, <https://doi.org/10.1016/j.scib.2022.11.016>, 2022.