Modelling CO₂ and N₂O emissions from soils in silvopastoral systems of the West-African Sahelian band

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Abstract. Silvopastoral systems (SPSs) have been shown to improve ecosystem resilience and provide sustainable land management solutions in the Sahel. However, accurately estimating the contribution of Sahelian ecosystems to the overall greenhouse gas (GHG) balance is a challenge, in particular the magnitude of carbon dioxide (CO_2) and nitrous oxide (N_2O) emissions from soils. In this work, we spatialized and applied the process-based model STEP-GENDEC-N_2O to investigate the magnitude, spatial, and temporal patterns of herbaceous mass, as well as CO_2 and N_2O emissions from soil in Sahelian SPSs. Our results show that over the last decade (2012-2022), there was a heterogeneous spatial distribution of herbaceous mass production, as well as of soil CO_2 and N_2O emissions in Sahelian SPSs. Spatial variations in soil CO_2 emissions are primarily controlled by soil carbon content, temperature, herbaceous mass, and animal load, while soil nitrogen content, soil water content, and animal load are the main factors driving the spatial variations in N_2O emissions from soil. The estimated CO_2 and N_2O emissions from soil in Sahelian SPSs over the 2012-2022 period were equal to 58.79 ± 4.83 Tg CO_2-C yr^{-1} (1 Tg = 10^{12} g) and 21.59 ± 3.91 Gg N_2O-N yr^{-1} (1 Gg = 10^9 g), respectively. These values are generally lower than estimates reported in the literature for tropical areas and croplands. Furthermore, our simulations indicated a significant annual rising trend of soil CO_2 and N_2O emissions between 2012-2020 as herbaceous mass increases, making more C and N available for nitrification, denitrification and decomposition processes. By mapping soil CO_2 and N_2O emissions, we provide crucial insights into the localization of emission hotspots in Sahelian SPSs, thereby offering valuable information that can be used to devise and implement effective strategies aimed at fostering carbon sequestration in the Sahel.

Keywords: Sahelian Silvopastoral system, soil carbon dioxide emission, soil nitrous oxide emission, process-based model, greenhouse gas emission

1 Introduction

35 Carbon dioxide (CO_2) and nitrous oxide (N_2O) are two important greenhouse gases (GHG) that contribute significantly (>90%) to anthropogenic climate warming (Hansen et al., 2000). With 298 times the warming potential of CO_2 over 100 years (Myhre et al., 2013), N_2O is also a stratospheric ozone-depleting substance (Ravishankara et al., 2009). The concentrations of atmospheric CO_2 and N_2O have increased substantially since the late 1700s (Bloch-Johnson et al., 2021; Prinn et al., 2018), mainly due to emissions from terrestrial soils (Butterbach-Bahl et al., 2013; Chevallier et al., 2015; Tian et al., 2020). CO_2

40 emissions from soil are due to organic matter decomposition (Robertson and Paul, 2000), while N_2O is produced in soils through nitrification (*i.e.,* oxidation of ammonium to nitrate) and denitrification (*i.e.,* reduction of nitrate to molecular N) (Davidson and Verchot, 2000). These processes are regulated by a range of environmental factors (Aulakh et al., 1991; Bajracharya et al., 2000; Reth et al., 2005), making it difficult to up-scale soil CO_2 and N_2O emissions from local sites to the regional and global scale.

45 Nevertheless, in the last decade, several works provided estimates of CO_2 and N_2O emissions from terrestrial soils at the large scale (Dangal et al., 2020; Leahy, 2004; Tian et al., 2020, 2019, 2018, 2016, 2015a). However, regions such as Africa, especially West-African Sahelian livestock production systems have not received much attention. Our knowledge of the magnitude, and the spatio-temporal distribution of soil CO_2 and N_2O emissions in these systems is limited and subject to large uncertainties (Assouma et al., 2017). This is mainly due to a lack of experimental and modelling studies focused on the region.

50 Silvopastoral systems (SPSs) are one of the most common livestock production systems in the West-African Sahel. They are composed of a mix of trees and herbaceous cover, grazed by livestock. As an attractive nature-based climate solution, SPSs offer long-term climate benefits thanks to the presence of trees that have the potential to sequester carbon and offset GHG emissions (Agbohessou et al., 2023; Torres et al., 2017). On the other hand, it has been reported that the livestock component of SPSs has an impact on the nitrogen (N) and carbon (C) cycles and therefore on GHG emissions (Butterbach-Bahl et al.,

55 2020). Indeed, livestock affects substrate availability in soil through N input from their excreta, then impacting CO_2 and N_2O emissions (Butterbach-Bahl et al., 2020; Dangal et al., 2020). It has been also reported that direct agricultural N_2O emissions from Africa mainly arise from livestock manure deposited in pastures and rangelands (Xu et al., 2019). Livestock movements result in heterogeneous spatial and temporal distributions of excreta, which increases spatial heterogeneity in soil properties and available nutrients which promote microbiological processes driving soil CO_2 and N_2O emissions (Assouma et al., 2017;

60 Smith et al., 2003). Actually, rangeland soils, combined with livestock productions, were reported to be responsible for a large share of GHG emissions (Assouma et al., 2017; Soussana et al., 2010; Valentini et al., 2014). The importance of rangelands in the global CO_2 and N_2O cycles and their potentialities to increase atmospheric CO_2 and N_2O levels, have been highlighted in a number of studies (Chang et al., 2015; Dangal et al., 2020; Leahy, 2004). Accordingly, to better understand the magnitude of GHG emissions in these systems and develop effective and spatially targeted climate solutions it is important to identify

65 CO_2 and N_2O emission hotspots and accurately estimate emissions from Sahelian SPSs.

The different bottom-up approaches used to estimate large-scale soil CO_2 and N_2O emissions include the use of "emission factors" (EFs) as proposed by the Intergovernmental Panel on Climate Change (IPCC) (Hergoualc'h et al., 2019; IPCC, 2006), statistical extrapolation of field measurements, and process-based models (Bigaignon et al., 2020; Delon et al., 2019; Li, 2000; Parton et al., 2001). Besides, the top-down approaches integrate atmospheric measurements and atmospheric inversion models

70 (Saikawa et al., 2014). Each method has its uncertainties and limitations, resulting in significant divergences in results across studies (Tian et al., 2019), especially in underrepresented regions like West-Africa (Tian et al., 2020). The IPCC defined N_2O emission as 1% of the applied N in the Tier 1 level (IPCC, 2006). This assumption of constant EF can neither depict spatial variations in N_2O emissions nor reflect the impacts of changing environments over time (Davidson and Kanter, 2014). Statistical extrapolation can also fail to depict the spatial heterogeneity in emissions, especially when the spatial variability in

75 the parameters exceeds the conditions prevailing during the calibration step (Tian et al., 2019). On the other hand, the process-based model simulation approach has the advantage of describing the overall C and N cycle within the terrestrial systems and can integrate various driving factors controlling soil CO_2 and N_2O production and emissions (Tian et al., 2019). This approach involves the use of extensive data, such as meteorological, soil and ecosystem management data. However, estimating the model parameters can be challenging as there is a scarcity of experimental studies that encompass comprehensive details on

80 local and regional pedoclimatic conditions and agricultural practices in West-Africa.

In this study, we selected the STEP-GENDEC-N_2O process-based model (Agbohessou et al., 2023), which couples water budget, herbaceous aboveground and belowground vegetation growth and decay, herbaceous and tree foliage litterfall (Jarlan et al., 2005; Mougin et al., 1995; Tracol et al., 2006), soil biogeochemistry and gaseous emissions (Bigaignon et al., 2020; Delon et al., 2019; Moorhead and Reynolds, 1991) to investigate the spatial and temporal patterns of herbaceous vegetation

85 mass, CO_2 and N_2O emissions from soil, and estimate their annual budget in the Sahelian SPSs. The STEP-GENDEC-N_2O model was specifically designed for Sahelian semi-arid ecosystems and has been validated locally for soil CO_2 and N_2O emissions in several sites representative of the Sahelian SPSs (Agbohessou et al., 2023; Bigaignon et al., 2020; Delon et al., 2019, 2015). In this study, this model was upscaled and used at the regional scale, *i.e.* at the west Sahelian region scale.

The specific objectives of our study are to: (1) investigate the spatio-temporal patterns of herbaceous vegetation mass, CO_2

90 and N_2O emissions from soils in the Sahelian SPSs over the last decade (2012-2022); (2) identify the environmental factors responsible for the changes in the spatial patterns of soil CO_2 and N_2O emissions; and (3) estimate the soil CO_2 and N_2O budget of the Sahelian SPSs during the 2012-2022 period.

2 Materials and methods

2.1 Characteristics of the study area

95   The Sahel region is a semi-arid strip stretching across the African continent from Senegal to the red sea (Le Houérou, 1989). The region is characterized by high temperatures, low soil fertility and a long dry season alternating with a short rainy season, with precipitation occurring mostly between June and September, making it challenging to grow crops. As a result, a large portion of the region is used for pastoral activities, which serve as the primary means of subsistence (Touré et al., 2012). The focus of this study is put on Sahelian SPSs of West-Africa from longitude 18°W to 20°E and latitude 13°N to 18°N (Fig. 1 and A1), which cover approximately 40% ($\approx$ 892,353 km^2) of the Sahelian band. The dynamics of rainfall in the Sahel are strongly linked to the dynamics of the West African monsoon (Biasutti, 2019). The Sahel experienced a dry period from the late 1960s to the mid-90ies, marked by years of extreme droughts such as in 1973-1974 and 1984-1985. Several studies have reported a recovery period (Galle et al., 2018; Nicholson, 2017) for the Sahel since 1984, which is defined by an increasing trend in total seasonal rainfall (Biasutti, 2019; Dai et al., 2004). However, rainy season characteristics have changed; rainfall is more intense and intermittent (especially in areas with the lowest rainfall) and wetting is concentrated in the late rainy season (Biasutti, 2019; Chagnaud et al., 2022).
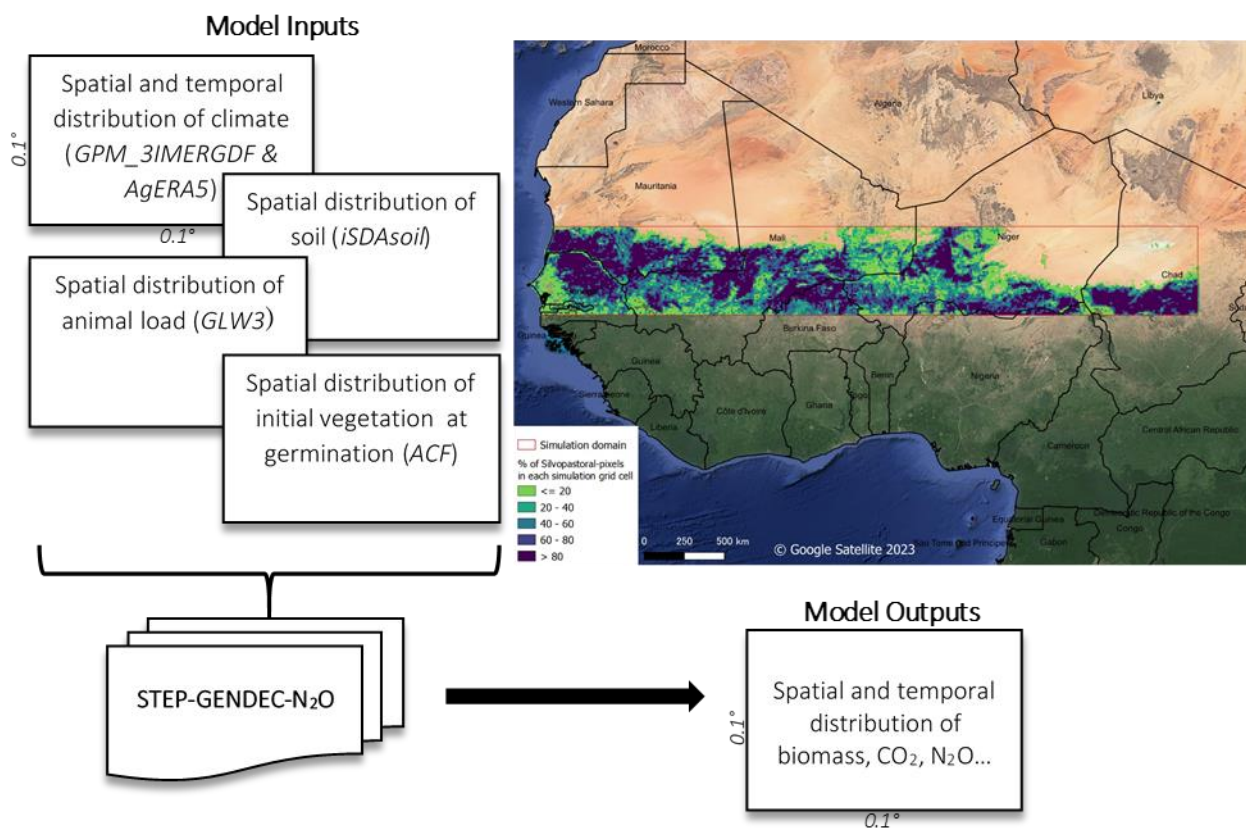
Figure 1: Illustration of the up-scaling approach used: model inputs and outputs and the simulation domain (Sahelian SPSs) are shown on the map. Silvopastoral areas were filtered from cultivated areas in the simulation area.

2.2 Model used from 1D processes to 2D up scaling: STEP-GENDEC-N$_2$O

2.2.1 Model description

STEP-GENDEC-N$_2$O is a process-based model developed for the Sahelian herbaceous savanna, coupling water budget, aboveground and belowground herbaceous vegetation growth and decay, litter fall (Mougin et al., 1995), soil biogeochemistry (Moorhead and Reynolds, 1991), and soil gaseous emissions (Agbohessou et al., 2023; Bigaignon et al., 2020; Delon et al., 2019). The model simulates the main processes describing the water, C, and N cycling between the atmosphere, vegetation, and soil at daily time steps and finally CO$_2$ and N$_2$O emissions. STEP-GENDEC-N$_2$O is forced daily by rain, global radiation, air temperature, wind speed, and relative air humidity. The model has been applied and evaluated to estimate herbaceous vegetation mass in Senegal, Mali (Mougin et al., 1995; Tracol et al., 2006), and Niger (Hiernaux et al., 2009), CO$_2$, NO, and N$_2$O emissions in Mali (Delon et al., 2015) and Senegal (Agbohessou et al., 2023; Bigaignon et al., 2020; Delon et al., 2019). In the litter decomposition GENDEC sub-model, the soil C content is calculated from the total litter input provided by STEP while soil N is derived from the quantity of C using C/N ratios (Moorhead and Reynolds, 1991). Soil moisture, soil temperature, and biomass (*i.e.,* herbal aerial mass, herbaceous root mass, ligneous leaves mass and faecal matter by livestock) are used as inputs to simulate microbial respiration. This is done by examining the interaction between buried litter, decomposer microorganisms and six C and N pools (*i.e.,* labile compounds, holo-cellulose, resistant compounds, dead microbial biomass, living microbial biomass and soil N). N$_2$O production and emissions from nitrification and denitrification are simulated using DNDC's (DeNitrification-DeComposition) equations (Li, 2000; Liu, 1996), adapted to the semi-arid region as described in Bigaignon et al. (2020) and Agbohessou et al. (2023). STEP alone has already been run to simulate aboveground biomass production, at the local scale (Jarlan et al., 2008, 2005, 2003; Mougin et al., 1995), meso scale (Grippa et al., 2017) and West-African Sahel scale (Pierre et al., 2016).

2.2.2 Model up-scaling

We used STEP-GENDEC-N$_2$O to simulate daily herbaceous vegetation mass, CO$_2$ and N$_2$O emissions from soil in West Sahelian SPSs. We developed a framework to run the model at a regional scale, using the parameterizations developed in the above-cited studies. Simulations were performed at the West Sahelian band scale (Fig. 1) divided into 18271 grid cells of 0.1° x 0.1°, from 2012 to 2022. Input variables were extracted from different datasets available at the global or regional scale as described below (Table 1). For the soil dataset which is provided at finer resolution (<0.1° x 0.1°), pixel values for each centroid of the 18271 simulation grid cells were extracted. Simulations were performed using a 6-year spin-up to reach carbon and nitrogen pools and stability as in Agbohessou et al. (2023).

Table 1:

Summary of the datasets used for input variables and land cover/use

Dataset	Input variable (unit)	Spatial resolution	Temporal resolution	Ref	URL
iSDAsoil	Soil pH (-) and soil texture (clay, silt, and sand content (%))	30 m	01-01-2012 (taken as constant)	(Hengl et al., 2021)	https://developers.google.com/earth-engine/datasets/tags/isda?hl=en
ERA5-Land	Initial soil water content (mm) and initial soil temperature (°C)	0.1° x 0.1°	01-01-2012 (taken as constant)	(Muñoz Sabater, 2019)	https://cds.climate.copernicus.eu/cds app#!/dataset/reanalysis-era5-land?tab=form
GPM_3IMERGDF	Precipitation (mm)	0.1° x 0.1°	01-01-2012 to 31-12-2021 (daily)	(Huffman et al., 2019)	https://disc.gsfc.nasa.gov/datasets/GPM_3IMERGDF_06/summary?keywords=%22IMERG%20final%22
AgERA5	Temperature (°C), Solar radiation (MJ m^{-2}), Vapour pressure (hPa), Wind speed (m s^{-1}), Soil albedo (-)	0.1° x 0.1°	01-01-2012 to 31-12-2021 (daily)	(Boogaard et al., 2020)	https://cds.climate.copernicus.eu/cds app#!/dataset/sis-agrometeorological-indicators?tab=overview
Gridded Livestock of the World version 3 (GLW3)	Animal load (-)	0.083333 decimal degrees	2012 (taken as constant)	(Gilbert et al., 2018)	https://dataverse.harvard.edu/dataverse/gld
Action Contre la Faim Surveillance West Africa	Proxy of herbaceous mass at germination (Kg ha^{-1})	1 km	2019-2021 (taken as constant)	(Lambert et al., 2019; Bernard and Fillol, 2020, 2021)	https://data.humdata.org/dataset/acf_biomass_west-africa_raster
Tree area density	Proxy of trees' foliar biomass	100 m	2023 (taken as constant)	(Tucker et al., 2023)	https://daac.ornl.gov/cgi-bin/dsviewer.pl?ds_id=2117
Land Cove/Use product					
Global Land Cover – SHARE (GLC – SHARE)	Land cover and land use (-)	1-km	2013 (taken as constant)	(Latham et al., 2014)	https://data.apps.fao.org/map/catalog/srv/eng/catalog.search#/metadata/ba4526fd-cdbf-4028-a1bd-5a559c4bff38

140 ## 2.3 Model input data

2.3.1 Climate data

The climate data required for the simulation were derived from two different datasets (GPM_3IMERGDF and AgERA5). Precipitation (mm) data were taken from the IMERG (Integrated Multi-satellitE Retrievals for GPM) dataset, GPM_3IMERGDF (Huffman et al., 2019). GPM_3IMERGDF or GPM IMERG Final Precipitation L3 1 day 0.1 degree x 0.1 degree V06, is derived from the half-hourly GPM_3IMERGHH dataset (GPM IMERG Final Precipitation L3 Half Hourly 0.1

145 degree x 0.1 degree V06, Greenbelt, MD, Goddard Earth Sciences Data and Information Services Center (GES DISC), 2022)

and represents the final estimate of the daily accumulated precipitation. The selected product is "*precipitationCal**: multi-satellite precipitation estimates with gauge calibration". Dezfuli et al. (2017) validated the IMERG product in Africa using gauge data from West and East Africa. They showed that the precipitation diurnal cycle is relatively better captured by IMERG than by the TMPA (TRMM Multi-Satellite Precipitation Analysis) product. Maranan et al., (2020) did a process-based validation of GPM IMERG in Africa using gauge data from a West African forested zone.

The spatial distribution of the GPM_3IMERGDF average precipitation over the last decade (2012-2022) exhibits significant gradients, with precipitation reaching as low as 0 mm at the northern border, exceeding 500 mm at the south-eastern border, and exceeding 1000 mm at the south-western borders (Fig. A2). Additionally, there is a significant increasing trend in annual mean precipitation amounts from 2010 to 2021, along with interannual variability (Fig. 3c).

Temperature (°C), solar radiation (MJ m^{-2}), vapour pressure (hPa) and wind speed (m s^{-1}), were extracted from the AgERA5 dataset (Boogaard et al., 2020) using the R package "ag5Tools" (ag5Tools: Toolbox for Downloading and Extracting Copernicus AgERA5 Data, 2022). AgERA5 dataset provides daily surface meteorological data matching the input needs of STEP-GENDEC-N$_2$O. The dataset is actually based on the ECMWF (European Centre for Medium-Range Weather Forecasts) re-analysis ERA5-Land dataset (Muñoz Sabater, 2019). ERA5-Land is an enhanced global dataset for the land component of the fifth-generation reanalysis produced by the ECMWF. It combines extensive historical observations from satellites, aircraft, land and marine weather sensors into global estimates using advanced modelling and data assimilation systems to generate consistent time series of multiple climate variables. More information about ERA5-Land product can be found in Muñoz Sabater et al. (2021) and Gleixner et al. (2020). In the data used, no significant trend (p>0.01) was observed in average air temperature (range: 25-35°C), minimum air temperature (range: 16-27°C), maximum air temperature (range: 25-39°C), global radiation (range: 19-25 MJ m^{-2} d^{-1}), wind speed (range: 2-7 m s^{-1}) and vapour pressure (range: 5-25 hPa) (extracted from ERA5-Land) in the Sahel between 2012 and 2022.

2.3.2 Soil data

Soil pH and soil texture (i.e. clay, silt, and sand content) were obtained from the iSDA (Innovative Solutions for Decision Agriculture Ltd.) soil dataset (Hengl et al., 2021). The iSDAsoil dataset contains soil property predictions at 30 m pixel size using machine learning coupled with remote sensing data and a training set of over 100,000 analysed soil samples all over Africa (Hengl et al., 2021; Miller et al., 2021). Dry soil albedo, soil moisture (mm) and soil temperature (°C) at the beginning of the simulation were extracted from the ECMWF re-Analysis ERA5_Land (Muñoz Sabater, 2019).

Exploration of the extracted soil datasets showed that the soils in the Sahel region are typically sandy, with high levels of sand and low levels of clay (Fig. A3a and A3b). This results in well-drained soils but low in nutrients. The soil pH in the south-western part of the Sahel ranges from 5 to 7, while in the north and east it is higher than 7 (Fig. A3c). The pH levels of the

soils in the Sahel vary also depending on their texture. Sandier soils typically have a high pH (7-8.5), while clay soils have a lower pH (5-7).

2.3.3 Animal load data

180 Information about livestock population and animal load distribution were obtained from the total livestock number for the reference year 2010 provided by the Gridded Livestock of the World version 3 (GLW3) (Gilbert et al., 2018) dataset. GLW3 provides global population densities of cattle, buffaloes, horses, sheep and goats in each land pixel at a spatial resolution of 0.083333 decimal degrees (approximately 10 km at the equator). The relative spatial distribution of livestock over the simulation period was assumed to be the same as the one indicated by the GLW3 database for the year 2010. To our knowledge

185 no measurement data are available on the temporal variation of livestock across the Sahel. We used the annual values of the GLW3 database to distribute the animal load on a monthly basis, taking into account the temporal variation of the livestock population from one month to the next throughout the year. We assumed an increase of the livestock up to 100% (in reference to the GLW3 database) in the pixels during the rainy season, and a gradual decrease down to 20% as we approach the middle of the dry season.

190 Analysis of the GLW3 dataset revealed that livestock is heterogeneously distributed across the Sahel and the animal load is dominated by bovines, ovines, caprines and some equines (Gilbert et al., 2018). High livestock densities were observed in north-western Senegal, southern Mauritania, central Mali, northern Burkina-Faso, southern Niger, northern Nigeria and south-western Chad (Fig. A3f).

2.3.4 Initial biomass data

195 The model calibration input parameters related to herbaceous vegetation such as Initial mass (B_g0) and Initial Specific Leaf Area (SLA_g0) at germination date were computed using data from the biomass dataset provided by "Action Contre la Faim (ACF) Surveillance West Africa" (Bernard and Fillol, 2021, 2020; Lambert et al., 2019). ACF biomass data were produced from 10-day images of Dry Mass Production (DMP) from SPOT-VEGETATION 4&5, PROBA-V and SENTINEL-3 satellites (Lambert et al., 2019). The retrieval algorithm of the DMP product is described as follows (Monteith, 1972; Swinnen et al.,

200 2022) Eq. (1):

$$DMP = R \bullet fAPAR \bullet \mathcal{E}_{LUEc} \bullet \mathcal{E}_c \bullet \mathcal{E}_T \bullet \mathcal{E}_{CO2} \bullet CUE \tag{1}$$

DMP is the 10-day Dry Matter Production (kgDM/ha/day), R is the 10-day total shortwave incoming radiation (GJ_T/ha/day), fAPAR is the PAR-fraction absorbed by green vegetation (J_{AP}/J_P), $\mathcal{E}_{LUEc}$ is the light use efficiency at optimum (kgDM/GJ_{AP}), $\mathcal{E}_c$ is the fraction of PAR in total shortwave (J_P/J_T), $\mathcal{E}_T$ is the normalized temperature effect, $\mathcal{E}_{CO2}$ is the normalized CO_2

205 fertilization effect and CUE is the carbon use efficiency.

The 1 km^2 resolution biomass raster product showing biomass production in the Sahel in kg ha^{-1} yr^{-1} was downloaded for the study period. We extracted the biomass value for each centroid of the simulation grid cells and performed a normalization by linearly scaling the dataset to a range between 0 and 2.5 g m^{-2} (the min and max values of B_g0 in the STEP model) to get the spatial distribution of the initial biomass (B_g0) at germination date. To obtain the spatial distribution of the initial Specific Leaf Area (SLA_g0) at germination date, we normalized ACF biomass dataset to a range between 0 and 280 cm^2 g^{-1} (min and max values of SLA_g0 given in Jarlan et al. (2008)). Here is the normalization formula used to linearly scale biomass values to B_g0 and SLA_g0 ranges Eq. (2):

$$X_{norm} = a + \frac{(x - min(x)) \cdot (b - a)}{max(x) - min(x)} \tag{2}$$

With X_{norm} representing the value of B_g0 or SLA_g0, a and *b* being the smallest and the largest value that B_g0 or SLA_g0 can take, respectively, and *x* being the biomass values in the ACF dataset.

In the model B_g0 and SLA_g0 are calibration parameters. B_g0 mainly affects the date of peak biomass (Tracol et al., 2006), whereas SLA_g0 is used to estimate LAI (and fAPAR). The maximum conversion efficiency (ε_c) of absorbed radiation into biomass (i.e., g of dry matter per MJ of absorbed photosynthetically active radiation) was set to 5 g MJ^{-1} which corresponds to the central value of the ε_c range possible values (Mougin et al., 1995; Pierre et al., 2011; Tracol et al., 2006) for all simulation grid cells.

2.3.5 Foliar mass of trees

Using the allometric equation developed by Hiernaux et al. (2023), we transformed the tree area density product provided by Tucker et al. (2023) into an estimate of tree foliar biomass in each simulation grid cell (Fig A3e). The conversion formula employed was Eq. (3):

$$DM_{foliar} = 0.2693 \cdot A^{0.9441} \tag{3}$$

Here, DM_{foliar} represents the mass of trees' leaves in kilograms, and *A* denotes the tree crown area in square meters.

2.4 Accounting for SPSs distribution in model outputs

Global Land Cover – SHARE (GLC – SHARE) dataset (Latham et al., 2014) provides information about the spatial distribution of a set of eleven major land cover classes (i.e., artificial surfaces, cropland, grassland, tree covered areas, shrubs covered areas, herbaceous vegetation, aquatic or regularly flooded, mangroves, sparse vegetation, bare soil, snow and glaciers, and water bodies) for the year 2013 and at 1-km^2 pixel resolution. First, we assumed that land cover change intensity was negligible in the Sahel during the last decade (the study period). Second, a new land cover class called silvopastoral areas was created and represents the sum of pixels of the classes: shrubs covered areas and grassland (Fig. 1).

The proportions of silvopastoral areas pixels within the 0.1° × 0.1° simulation grid cells (pixel resolution ≈ 123.21 km^2) were calculated using the GLC – SHARE dataset to obtain the spatial distribution of silvopastoral systems in the Sahel (Fig. A1).

In our analysis and interpretation of the spatial distribution of herbaceous mass, CO_2, and N_2O emissions, we consider the model outputs for simulation pixels where silvopastoral areas are > 80%. Additionally bivariate maps were proposed, which display both model outputs and the distribution of SPSs in the simulation domain, to provide a more comprehensive view of the results.

240   To estimate the annual budget of soil CO_2 and N_2O emissions, the model outputs were weighted by the proportion of silvopastoral area within each simulation grid cell (Fig. 1 and Fig. A1), therefore considering all SPSs, even those which %SPS < 80%, across the simulation domain.

2.5 Random Forest algorithm for the analysis of soil CO_2 and N_2O emissions driving parameters

Random Forest (RF) is a machine learning method developed by Breiman (2001), it is a natural non-linear modelling tool that
245   has proven valuable in many fields (Liu et al., 2022; Webb et al., 2021). We used the RF algorithm to identify the most important factors influencing the spatial distribution of soil CO_2 and N_2O emissions. The main advantages of RF algorithms are its low number of tunable factors, good tolerance to outliers and noise, general resistance to overfitting, and ability to identify and rank the most important variables (Liu et al., 2022; Webb et al., 2021). The RF algorithm was implemented into the R software (R Core Team, 2019) and the modelling framework provided by the *randomForest* R package (Liaw and
250   Wiener, 2002) is used in our study. The target variables of the RF are the spatial distribution of the simulated soil CO_2 and soil N_2O emissions, while the predictors are the spatial distribution of a range of environmental and biological factors susceptible to impact the simulated soil CO_2 and N_2O emissions. These factors include: soil sand content, soil clay content, soil water content, soil temperature, soil pH, soil C content, soil N content; air temperature, albedo, annual precipitation; herbaceous mass and animal load. The dataset used was composed of 18271 individual observations (outputs of each simulation grid cell),
255   and we conducted the RF with the default parameters proposed by the *randomForest* package.

The method is composed of three critical steps, each of which plays a crucial role in the overall performance of the model. In the first step, a bootstrap sample of observations (equal to the number of trees) is randomly drawn from the dataset, with replacement. Approximately one third of the total observations are left out and used as "out-of-bag" (OOB) data to evaluate the model's performance and prevent the need for a separate validation dataset (Efron and Tibshirani, 1986; Philibert et al.,
260   2013). This provides a resampling procedure that generates multiple versions of the training dataset, which helps to mitigate overfitting and improves the accuracy of the model. In the second step, a random subset of predictor variables is selected at each node of the decision tree (Ghattas, 2000; Philibert et al., 2013; Prasad et al., 2006). The number of variables selected (mtry) was set to the integer part of the square root of the total number of variables (Breiman, 2001; Liaw and Wiener, 2002; Philibert et al., 2013). This approach involves considering a subset of variables at each node of the decision tree and selecting
265   the best variable that maximizes the information gain. This randomization technique reduces the correlation among the trees and makes the model more robust and accurate. In the final step, multiple decision trees are grown from the bootstrapped

dataset and the random subsets of features. The trees are grown using recursive binary partitioning of the data, with the best split determined by optimizing a quality criterion such as information gain by the Gini impurity index (Breiman et al., 1984). The final prediction is made by aggregating the predictions of all trees in the forest by averaging the outputs. The process is repeated multiple times until a stable estimate of model performance is obtained.

We assessed variable importance using the percentage increase in Mean Squared Error (*%IncMSE*) after a factor was randomly permuted. *%IncMSE* estimates the contribution of each variable to the reduction in the mean squared error of the model (Breiman, 2001; Echeverry-Galvis et al., 2014). Factors with higher *%IncMSE* values are considered as more important in explaining the spatial distribution of soil CO_2 and N_2O emissions. The importance of each factor was displayed with the variable importance plot developed from the RF.

2.6 Statistical analysis and mapping

We conducted linear regression analysis to examine trends over time in herbaceous vegetation mass, soil CO_2 and N_2O emissions and relevant emission driving variables. The Pearson correlation was used to assess the relationship between the different variables. All statistical analysis and mapping were performed using R (R Core Team, 2019).

3 Results

3.1 Spatio-temporal patterns of aboveground herbaceous mass in the Sahelian SPSs (2012-2022)

The annual production of aboveground herbaceous mass in the Sahelian SPSs, simulated from 2012 to 2022, displays a latitudinal gradient characterized by higher herbaceous mass in the southern regions, which diminishes as we progress towards the northern latitudes (Fig. 2). The same spatial pattern is observed in Figure 2b, which highlights results for Sahelian SPSs (pixel %SPS>80%). The maximum annual mean production (2012-2022) reaches 3 t DM ha^{-1} yr^{-1} and the annual minimum production is 0 t DM ha^{-1} yr^{-1}.
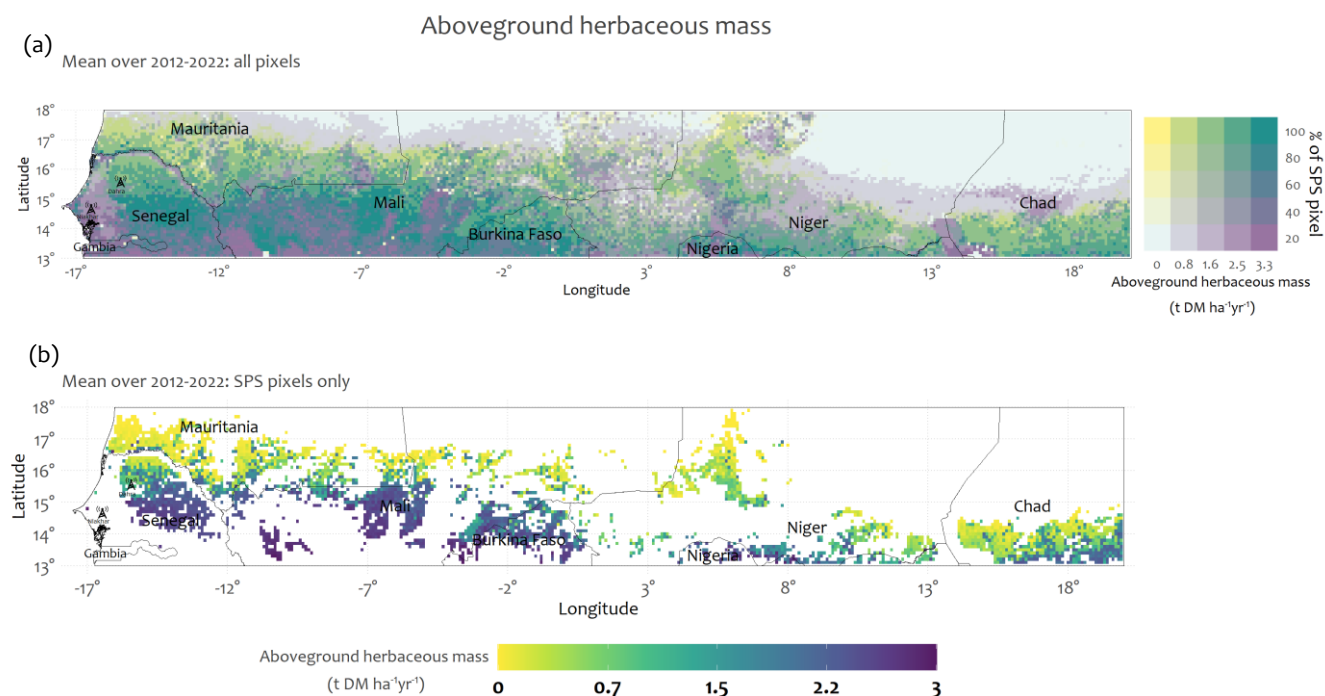
Figure 2: Regional distribution of simulated herbaceous mass in the Sahelian SPSs (annual mean over 2012-2022), in tDM ha-1 yr-1. (a) Bivariate map, which displays both simulated herbaceous mass and the distribution of SPSs in the simulation domain, (b) map filtering the simulated herbaceous mass for areas with Sahelian SPSs >80% only.

Herbaceous mass in Sahelian SPSs exhibited inter-annual variations with standard deviations reaching up to 1.3 t DM ha^{-1} yr^{-1} at some locations (Fig. A4a). We observed a significant increasing trend (p<0.001) in the annual herbaceous mass anomaly (deviation from the 2012-2022 average) from 2012 to 2020 (Fig. 3a). This rising trend is evident in the Hovmöller representation, which depicts a gradual increase in herbaceous mass, particularly in the southern Sahel region around the latitudes 13°N and 15°N (Fig. 3b), with the highest production simulated in the wettest years (2019, 2020 and 2021, Fig. 3c). In the southern Sahel (13°N to 15°N), herbaceous mass in SPSs can reach 2.5 t DM ha^{-1} yr^{-1}, while in the northern Sahel (16°N to 18°N), it does not exceed 0.5 t DM ha^{-1} yr^{-1} (Fig. 3b). Overall, herbaceous mass in the Sahelian SPSs is highly correlated to the wet season total precipitation which shows large inter-annual variation (Fig. 3c: p < 0.001 and r = 0.6).
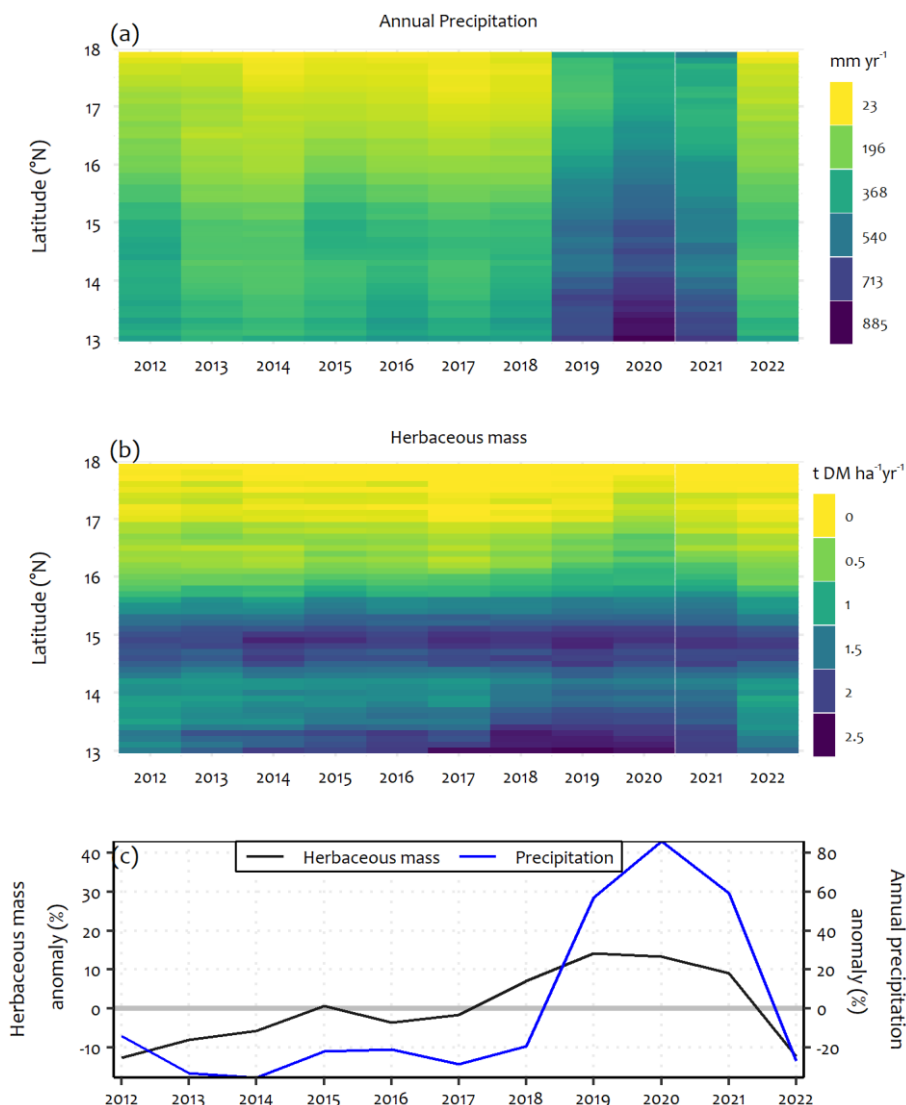
Figure 3: (a) Hovmöller (latitude-year) plot of the annual precipitation. (b) Hovmöller (latitude-year) plot of herbaceous mass in the domain indicated in Fig. 2b. (c) Interannual variations of anomalies (relative to the mean value for the period 2012-2022).

3.2 Soil CO_2 and N_2O emissions in Sahelian SPSs

3.2.1 Spatial distribution across the Sahel

The simulation results reveal a heterogeneous spatial distribution of soil CO_2 and N_2O emissions, with the lowest emissions in the north and the highest emissions in the south (Fig. 4). SPSs in the pastoral zones of central Senegal, in southern and central Mali, in northern Burkina Faso, and in southern Niger (between longitudes 7°E and 8°E) exhibit high levels of soil CO_2 emissions (Fig. 4a and 4b). The average soil CO_2 emissions for the period 2012-2022, reached up to 1.7 t CO_2-C ha^{-1} yr^{-1}, as

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shown in figure 4b. SPSs located in the northern regions of Niger, as well as in Mauritania, were generally not significant sources of CO_2 (Fig. 4b). Only SPSs of central Senegal, northern Burkina Faso and Mali remain constant CO_2 emission

310 hotspots throughout the study period, with emissions as high as 2.6 t CO_2-C ha^{-1} yr^{-1} in some years, as shown in the all-years detailed maps in figure A6. Inter-annual variabilities of up to 0.7 t CO_2-C yr^{-1} ha^{-1} have been observed in some SPSs (Fig. A4b).

Figure 4c depicts heterogeneous soil N_2O emissions ranging from 0 to 3 kg N_2O-N ha^{-1} yr^{-1} and high emissions in some areas where the percentage of SPSs pixel is lower than 80%. The figure 4d exclusively show cases areas that are representative of

315 the Sahelian SPSs (%SPS>80), showing that soil N_2O emissions were as high as 2.3 kg N_2O-N ha^{-1} yr^{-1} (mean 2012-2022 period) in SPSs located within the sandy pastoral zones of central Senegal, and in southern Mali between latitudes 13°N and 15N°E. In contrast, smaller N_2O emissions were observed in the other SPSs of the region, especially in Niger and Chad. High inter-annual variabilities have been observed in the southern part of the Sahel (Fig. A4c).
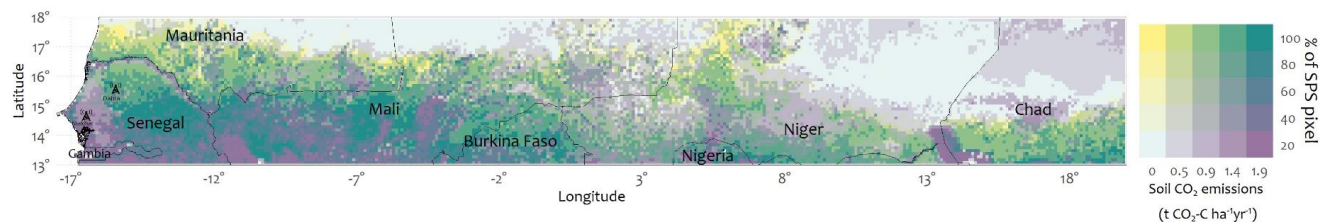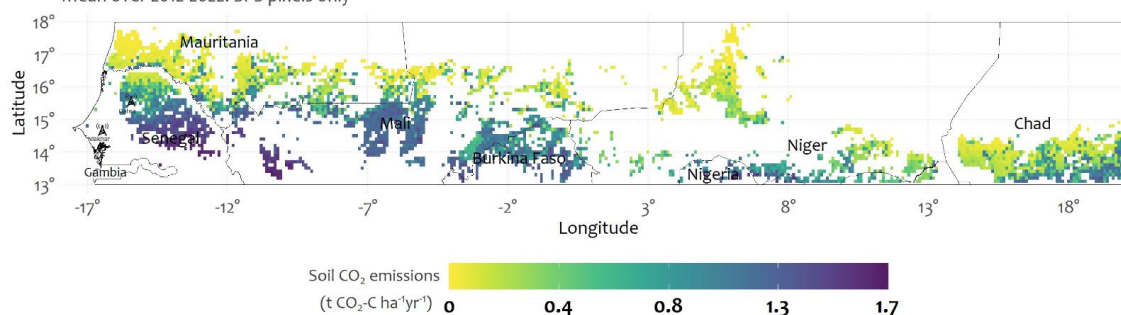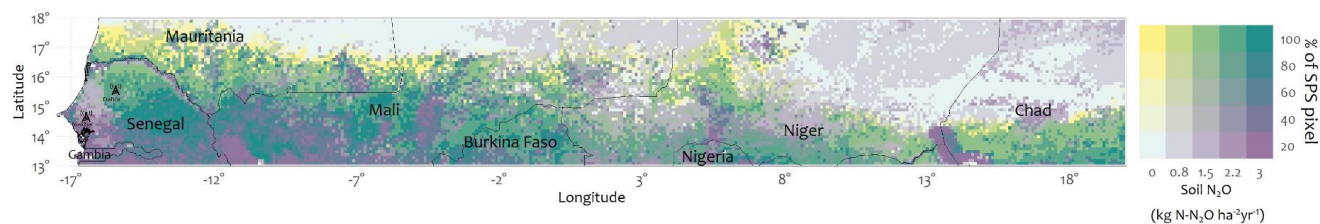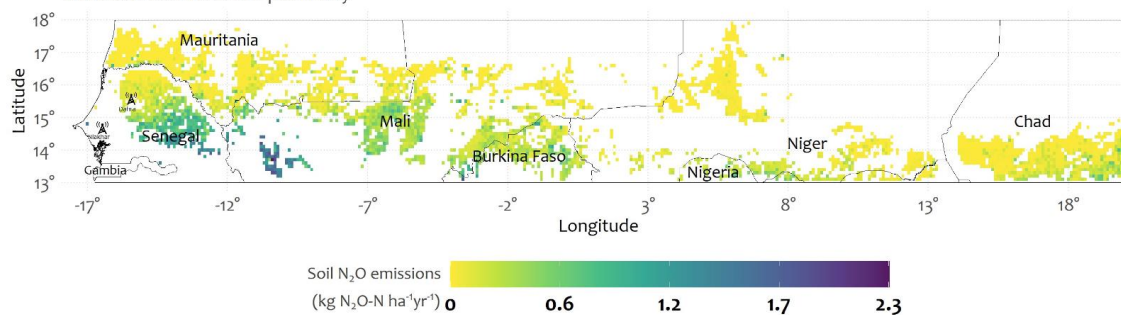
Figure 4: **Regional distribution of simulated soil CO₂ andN₂O emissions in the Sahelian SPSs (annual means over 2012-2022). (a) and (c): Bivariate maps display both model outputs and the distribution of SPSs in the simulation domain. (b) and (d): Maps displaying model outputs only on areas representative of the Sahelian SPS (>80%).**

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3.2.2 Exploring the temporal dynamics of model outputs

Figure 5 shows the temporal dynamics of wet season precipitation, soil CO_2 emissions, soil N_2O emissions, soil water content

325   and soil total C at two contrasted sites showing different emission levels (low and high), located in Niger (longitude 10.7, latitude 14.2) and Senegal (longitude -15.4, latitude 15.4) respectively. These sites were on predominant sandy soils. The observed dynamics of the different variables (precipitation, soil CO_2 emissions, soil N_2O emissions, soil water content and soil C content) at these sites show the model's ability to simulate realistically seasonal variations at fine time-scales in soil CO_2 and soil N_2O emissions in the Sahel.

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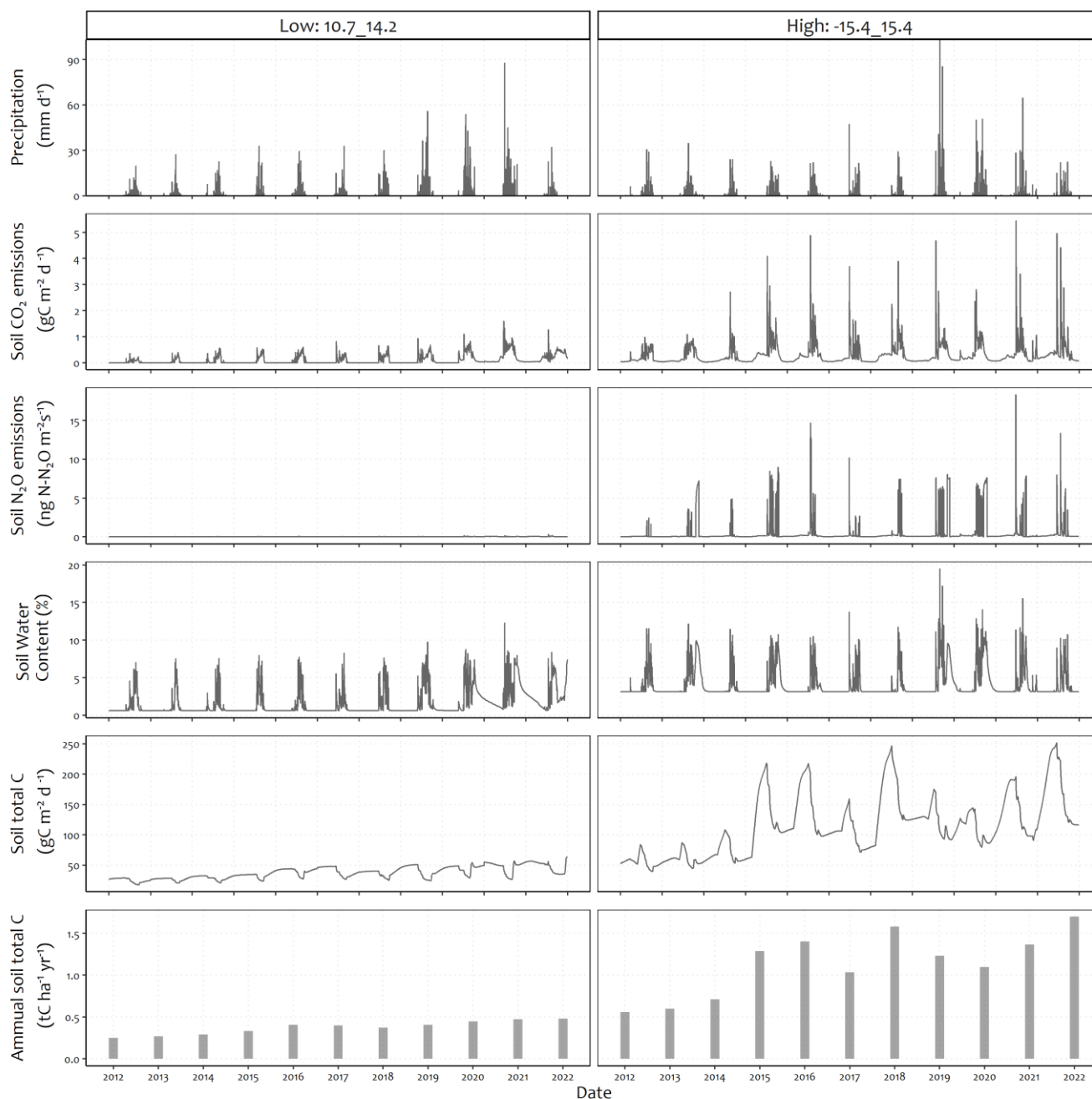
Figure 5: Temporal dynamics of model outputs across two sites with different levels of soil CO_2 &N_2O emissions. From top to bottom: precipitation, soil CO_2 emissions, soil N_2O emissions, soil water content and soil total C. At the left, a site exhibiting low emissions (latitude -10.7, longitude 14.2), at the right, a site with high emissions (longitude -15.4, latitude 15.4).

3.2.3 Factors controlling the spatial distribution of soil CO_2 and N_2O emissions

335   The observed variations in the spatial patterns of soil CO_2 and N_2O emissions were attributed to a complex interaction between meteorological, edaphic, bio-physical factors. According to a statistical analysis assessed by Random Forest over the model output in grid cells containing more than 80% of SPSs, the soil carbon and nitrogen contents were found to be the primary factors controlling the spatial distribution of soil CO_2 and N_2O emissions, respectively, as shown in figure 6. Soil C content, air temperature and soil temperature were identified as the three most significant factors controlling the spatial patterns of soil

340   CO_2 emissions. For soil N_2O, the two most significant factors after soil N content were soil water content and animal load. The results further showed that for soil CO_2, the other driving factors were herbaceous mass, animal load, annual precipitation (or soil water content), soil clay content, and soil water content (Fig. 6a). For soil N_2O, herbaceous mass, soil temperature, soil clay content, annual precipitation (or soil water content), and air temperature (in that order) also appeared as key driving factors (Fig. 6b). Soil pH was found to have the least influence on the spatial pattern of soil N_2O emissions (Fig 6).
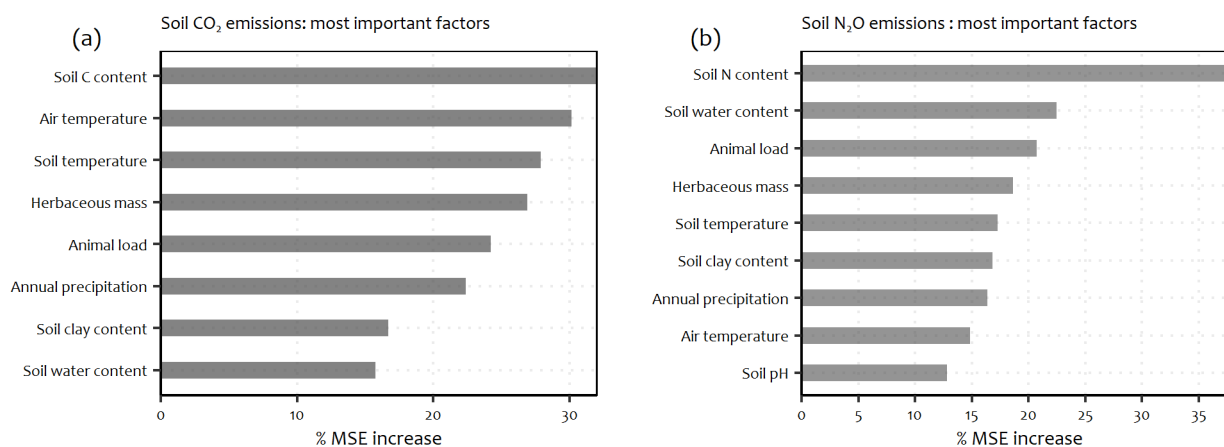
Figure 6: Factors controlling the spatial changes in (a) soil CO_2 and (b) N_2O emissions, from Random Forest Analysis. MSE = Mean Squared Error

### 3.2.4 Annual budgets across the Sahel (2012-2022)

The simulated soil CO_2 emissions include both microbial respiration and root respiration of herbaceous. Between 2012 and

350   2022, the estimated average soil CO_2 emissions in the Sahelian SPSs was 58.79 ±4.83 Tg CO_2-C yr^{-1} (1 Tg = 10^{12} g). The highest soil CO_2 annual emission (65.80 Tg CO_2-C yr^{-1}) was found in 2020, while the lowest (50.77 Tg CO_2-C yr^{-1}) was in 2012 (Fig. 7a). During this same period, the mean soil N_2O emission was 21.59 ± 3.91 Gg N_2O-N yr^{-1} (1 Gg = 10^9 g), ranging from 17.31 Gg N_2O-N yr^{-1} in 2012 to 27.43 Gg N_2O-N yr^{-1} in 2020 (Fig. 7b). From 2012 to 2020, annual soil CO_2 and N_2O emissions showed significant ($p < 0.01$) rising trends of 4.30e-3 ± 6.05e-4 Tg CO_2-C yr^{-1} and 3.75e-3 ± 4.47e-4 Gg N_2O-N yr^{-}

355   ^{1}, respectively. However, emissions dropped after 2021, with a 17.5% decrease in soil CO_2 emissions and 25.5% decrease in soil N_2O emissions (Fig. 7c). Figure 7c reveals that the inter-annual variations in soil CO_2 and soil N_2O emissions are quite

homothetic, as indicated by a Pearson's correlation coefficient of 0.86. Annual precipitation over the 2012-2022 period and averaged over the study domain was significantly correlated to both soil CO_2 ($p<0.05$, $r=0.48$) and N_2O ($p<0.05$, $r = 0.79$) emissions.
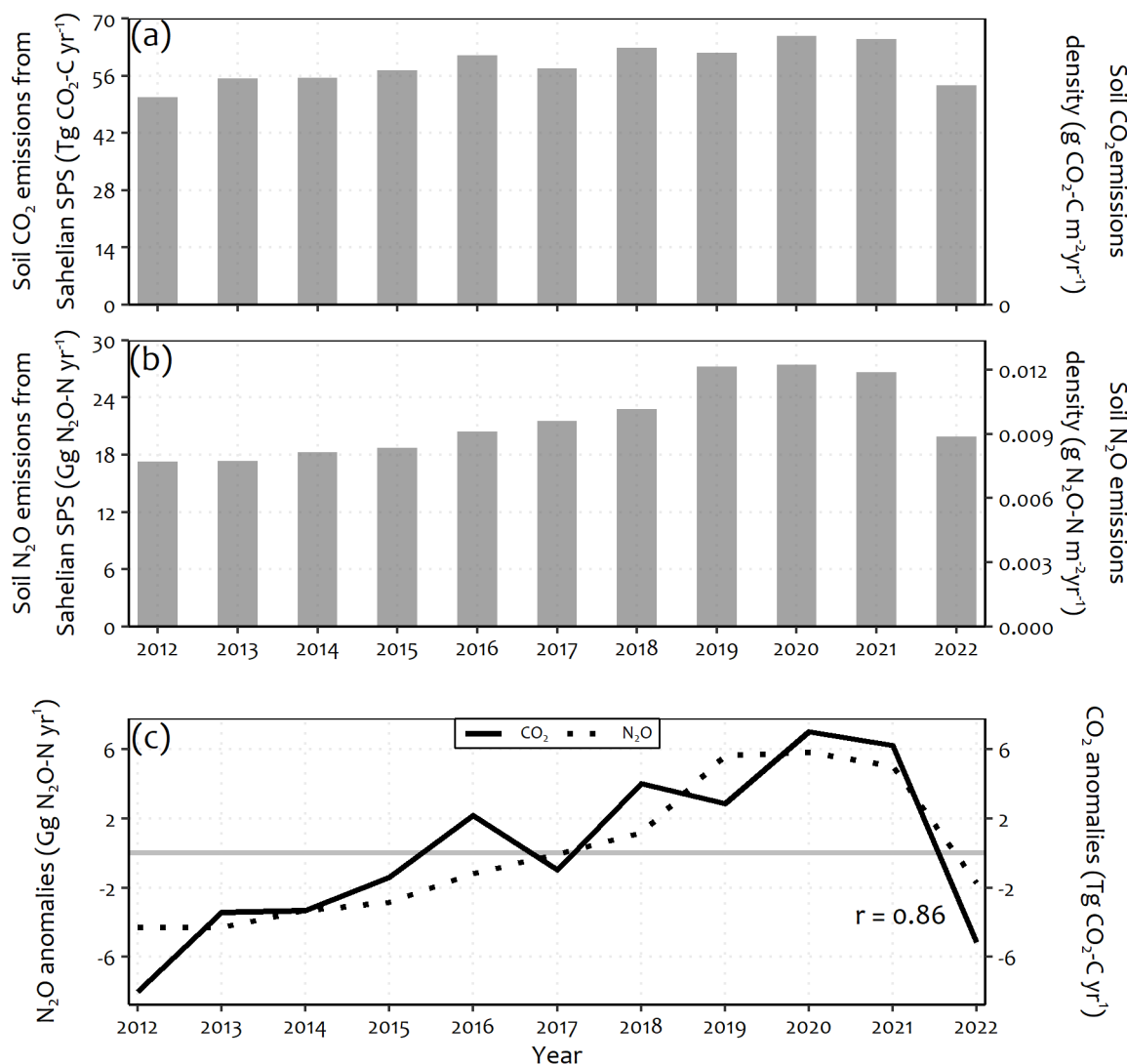
Figure 7: Interannual variation in soil CO_2 and N_2O emissions in the Sahelian SPSs (which covers *approx.* 892000 km²) during 2012-2022. (a) Soil CO_2 emissions in Tg C yr^{-1} (1 Tg = 10^{12} g) and (b) soil N_2O emissions in Gg C yr^{-1} (1 Gg = 10^9 g). (c) Interannual variations of soil CO_2 and soil N_2O anomalies (relative to the mean value for the period 2012-2022). The Pearson correlation coefficient between CO_2 and N_2O anomalies was 0.86. We calculated the proportion of SPSs area pixels within each 0.1° x 0.1° simulation grid cell, and used these values to weight the model outputs for each grid cell.

4 Discussion

Previous studies at global and regional scales have estimated greenhouse gas (GHG) emissions from various ecosystems, especially agricultural systems (Tian et al., 2020, 2015b), forests (Tian et al., 2020; Verchot et al., 1999), and rangelands (Dangal et al., 2020). These studies have frequently highlighted significant uncertainties when estimating emissions from underrepresented regions, like in Africa. In addition, different modeling techniques often give divergent results when estimating emissions from these regions. In this study, we have up-scaled the 1D STEP-GENDEC-N$_2$O model, building upon previous local applications and validations of the model in different representative sites of the Sahelian SPS, to provide the first large-scale estimate of soil CO$_2$ and N$_2$O emissions from W-Sahelian SPSs. In this section, we discuss the magnitude of soil CO$_2$ and N$_2$O emissions reported in this study, the role of environmental and biological factors that drive the spatial heterogeneity observed in soil CO$_2$ and N$_2$O emissions in Sahelian SPSs, and the uncertainties and limitations associated with these estimations.

4.1 Spatial and temporal patterns of herbaceous vegetation, soil CO2 emissions, and their relationship

In a previous study, Pierre et al. (2016) demonstrated the ability of the STEP model to simulate the dynamics of herbaceous vegetation at regional scale in the W-Sahel. They found a good agreement between the regional spatial patterns of STEP-simulated vegetation masses and the Moderate Resolution Imaging Spectroradiometer (MODIS) vegetation indices. They observed a latitudinal gradient in herbaceous vegetation mass, caused by the rainfall gradient, as also shown in our results. The magnitudes of herbaceous mass in their study and ours are comparable and the spatial patterns are similar although the study periods and the input data employed are not exactly the same. Previous estimates of mass production in the Sahel, using the LandscapeDNDC model (Rahimi et al., 2021) and remotely sensed data obtained from the ACF product, Proba-V, and SPOT-vegetation satellites (Bernard and Fillol, 2021, 2020; Lambert et al., 2019), exhibited relatively stable temporal dynamics in mass production from 2010 to 2019. These estimates encompassed all land use types in the Sahel region, which could explain the divergence with our results showing a gradual increase in mass production in Sahelian SPSs between 2012 and 2022. Moreover, the trend observed in this study is mainly driven by the most recent years, with the highest values occurring in 2019, 2020 and 2021.

Plants' litter is the main source of carbon entering the soil, which explains the similar spatial patterns observed in both annual herbaceous mass (Fig. 2b) and annual soil CO$_2$ emissions (Fig. 4b). This illustrates the effect of the C substrate on CO$_2$ emissions, as confirmed by the Random Forest analysis (Fig. 6). The size and composition (nature of substrate, molecules, C/N ratio, etc.) of the available carbon pool actually control the magnitude of the CO$_2$ emissions from soil (Barnard et al., 2020). Soil CO$_2$ emissions include the respiration of soil microorganisms (microbial or heterotrophic respiration) and plant roots (autotrophic respiration), including all respiratory processes occurring in the rhizosphere (Raich and Potter, 1996; Xu and Shang, 2016). Root cells perform cellular respiration, metabolizing carbohydrates that are sent down from the leaves. Depending on the vegetation density, root respiration can contribute significantly to the total soil respiration (Macfadyen,

1970). In some SPSs in the north-western Sahel (*e.g.,* Mauritania, Mali and Niger), we simulated significant soil CO_2 emissions despite the low herbaceous mass. These areas also exhibit high interannual variabilities in soil CO_2 emissions (Fig. A4b: up to 0.7 t CO_2-C ha^{-1} yr^{-1}). The northern Sahel is generally characterized by a long dry season and very low rainfall. In such semi-arid areas, at the onset of the wet season, the first rainfall events rewet the dry soil resulting in a mineralization peak leading to a large soil CO_2 efflux pulse, also known as the "Birch effect" (Birch, 1958). The STEP-GENDEC-N_2O model accounts for this "Birch effect" (Delon et al., 2019), which could explain the soil CO_2 emissions hotspots simulated in some SPSs of the north-western Sahel. The site (simulation pixel) located at longitude -15.4°W and latitude 15.4°N (0.1 * 0.1 degrees), as depicted in Figure 5, actually illustrates the Birch effect in soil respiration dynamics, with notably high emissions simulated at the onset of the rainy seasons. This simulation pixel encompasses the Dahra site in northern Senegal (longitude -15.43222°W and latitude 15.40277°N), where the 1D STEP-GENDEC-N_2O model results were in good agreement with observations (Agbohessou et al., 2023; Delon et al., 2019). According to Fan et al. (2015), up to 20% of the annual soil CO_2 emissions to the atmosphere occurs in African savanna ecosystems following intense rainfall. The CO_2 pulses associated with rewetting can represent a large part of the annual C budget in semi-arid and arid ecosystems (Barnard et al., 2020; Jarvis et al., 2007; Ma et al., 2012; Rey et al., 2017).

In a SPSs located in northern Senegal, Delon et al. (2017) measured soil respiration ranging from 2.4 ± 0.62 gC m^{-2} d^{-1} at the onset of the wet season to 0.7 ± 0.01 gC m^{-2} d^{-1} at the end of the wet season in 2013. Our estimated mean soil CO_2 emissions density for Sahelian SPSs between 2012 and 2022 (0.06 gC m^{-2} d^{-1}) is lower than estimates at the global scale for grasslands (2.2 gC m^{-2} d^{-1}) and partially vegetated deserts (1.0 gC m^{-2} d^{-1}) by Xu and Shang (2016). On a global scale, for these grasslands, the substrate (soil C content) is probably much more important than in SPSs, which explains the higher values of CO_2 emissions.

4.2 Soil N_2O and CO_2 emissions in Sahelian SPSs and importance of livestock

Between 2012 and 2022, the simulated soil N_2O emissions from Sahelian SPSs were 0.022 ± 0.004 Tg N_2O-N yr^{-1}. The regional natural soil N_2O emissions in Africa was estimated at 1.6 Tg N_2O-N yr^{-1} for the period 2007-2016 (this estimate includes natural emission from oceans, inland waters, estuaries, coastal zones) (Tian et al., 2020). The simulated average soil N_2O emissions from Sahelian SPSs is lower than the median total N_2O emissions of 0.05 Tg N_2O-N yr^{-1} from bomas (livestock enclosure where livestock excreta accumulates) in sub-Saharan Africa's semi-arid and arid climates (Butterbach-Bahl et al., 2020). The average soil N_2O emission density (per unit area) in Sahelian SPSs (2012-2022) was found to be 0.01 g N_2O-N m^{-2} yr^{-1} (range: 0-0.23 g N_2O-N m^{-2} yr^{-1}), which is comparatively lower than the average estimate in tropical regions (0.11 ± 0.02 g N_2O-N m^{-2} yr^{-1}) and the global average ($\approx$ 0.05 g N_2O-N m^{-2} yr^{-1}) reported for the period 2007-2016 (Tian et al., 2019). The soil N_2O emission density in Sahelian SPSs (2012-2022) was also lower than global emission densities estimated in croplands (0.21 ± 0.08 N_2O-N m^{-2} yr^{-1}) and other ecosystems (0.06 ± 0.01 g N_2O-N m^{-2} yr^{-1}), respectively, during the period 2007-2016 (Tian et al., 2019). The most significant soil N input in Sahelian SPSs actually originates from livestock excreta, which is

lower than the N input in most fertilized agricultural fields (Dangal et al., 2020), explaining the lower emission density in SPSs compared to the global average emissions density in croplands. In fact, studies have shown that nitrogen fertilizer application in croplands is the leading factor responsible for the increases in emission from agriculture (Cao et al., 2018; Davidson, 2009; Maavara et al., 2019; Shcherbak et al., 2014; Yao et al., 2020), followed by a minor yet significant rise in emissions from

435   livestock manure (Tian et al., 2020). But on the other hand, in regions where very little nitrogen fertilizer is used in cropland such as in Africa, soil N_2O emissions mainly arise from livestock manure deposited in pastures and rangelands (Butterbach-Bahl et al., 2020; Dangal et al., 2020; Xu et al., 2019). This confirms the N_2O emission hotspots simulated in locations where the density of livestock is high (Fig. 4c & 4d, Fig A3f), as also highlighted by the Random Forest analysis. Indeed, the animal load distribution also affects the spatial distribution of soil N_2O and CO_2 emissions, as shown in figure 6. Several authors have

440   already mentioned this impact (Assouma et al., 2017; Butterbach-Bahl et al., 2020; Dangal et al., 2020; Smith et al., 2003). Livestock influences the spatial distribution of soil C and N, which, in turn, significantly affects soil N_2O and CO_2 emissions.

4.3 Common features of soil CO_2 and N_2O emissions in Sahelian SPSs

Figure 7c shows that the inter-annual variations in soil CO_2 and soil N_2O emissions are quite homothetic, as indicated by a Pearson's correlation coefficient of 0.86. This suggests that they are both responding in a similar manner to the different

445   ecological drivers. Some authors stated that the main processes responsible for CO_2 (decomposition) and N_2O (nitrification and denitrification) emissions from soils are influenced by the same environmental factors, namely soil moisture, soil temperature, soil texture, soil C and N content (Davidson and Swank, 1986; Oertel et al., 2016; Rastogi et al., 2002; Signor and Cerri, 2013). Several studies have shown how soil CO_2 and N_2O emissions evolve over time in response to changes in environmental driving factors (Cuhel et al., 2010; Davidson and Swank, 1986; Khalil, 2015; Ray et al., 2020) but the

450   complexity of the interactions between these different factors make it difficult to assess the importance of each driver responsible for the spatial distribution of the emissions. From our results, the main factor responsible for the spatial distribution of soil CO_2 and N_2O emissions in SPSs (Fig. 6) is substrate availability (soil C and N content), which outweighs other factors such as soil water content, temperature, and soil texture. Moreover, substrate availability is directly linked to herbaceous mass productivity (as mentioned in section 4.1) and to animal load (see section 4.2). This is consistent with the findings of Ray et

455   al. (2020), who showed that soil CO_2 emissions are affected more by substrate availability than by rainfall, although their experiment was performed in a cropping system. In addition to influencing the spatial pattern of soil CO_2 and N_2O emissions, soil C and N also impact the temporal variation of these emissions, as shown in figure 5, where the largest emissions are found where the C content was the highest. Furthermore, our simulations revealed a rise in emissions between 2012 and 2020 (Fig. 7c), that is correlated to the increase in herbaceous mass during the same period (Fig. 3a).

460   In the literature, soil water content is often highlighted as the major driver of the temporal variation of soil N_2O emissions as it regulates the oxygen availability to soil microbes (Butterbach-Bahl et al., 2013; Davidson and Verchot, 2000). The effect of soil moisture is actually predominant on denitrification processes, which lead to large amounts of N_2O emissions when soil

water-filled pore space (WFPS) reaches 70 to 80% (Davidson and Verchot, 2000). This is consistent with the result of our RF analysis, which ranks soil water content as the second most important factor responsible for spatial changes in soil N_2O emissions (Fig. 6b). The impact of air temperature and soil temperature on the spatial distribution of soil CO_2 emissions suggests a positive feedback loop between climate warming and these emissions. The impact of global change drivers such as temperature on ecosystem processes and greenhouse gas emissions has been well studied and proven (Aulakh et al., 1992; Bajracharya et al., 2000; Lloyd and Taylor, 1994; Ray et al., 2020). The annual budgets of CO_2 and N_2O emissions (Fig. 7a and 7b) throughout the period of simulation show a low interannual variability. This can be attributed to the low interannual variability of influencing factors such as substrate availability (C: 33.60 ± 2.38 gC m^{-2} d^{-1} N: 5.89 ± 0.46 gN m^{-2} d^{-1}), and soil water content ($4.87 \pm 0.19\%$ yr^{-1}). Our simulation results do not allow to explore possible interactions between climate warming and annual soil CO_2 and N_2O emissions, as the average annual air temperature (averaged over the study domain) did not vary much over the simulation period (28.37 ± 0.25 °C). Regional-scale observations show a temperature increase ranging from 1 to 2°C between 1950 and 2010 (Guichard et al., 2020). Therefore, over a 10-year period, this corresponds to a maximum increase of approximately 0.33°C, which is less than 0.5°C. This order of magnitude is comparable to the one computed for air temperature from the climate dataset used, and it is too small to be detected by temperature-versus-time regression.

4.4 Uncertainties and limitations

The lack of comprehensive dataset on the annual spatial distribution and growth dynamics of the livestock population in the Sahel between 2012 and 2022 remains a significant source of uncertainty in the CO_2 and N_2O emissions reported in this study. Actually, information on the spatial distribution and population of livestock was only available for the 2010 year (Gilbert et al., 2018). Only the spatial and the seasonal variability of the grazing pressure was taken into account in our simulation. We assumed that the annual distribution and growth dynamics of livestock in Sahelian SPSs did not change significantly between 2010 and 2022, although it might have been affected by the interannual variability of herbaceous mass. Given the significant impact of livestock on CO_2 and N_2O emissions in these ecosystems (Agbohessou et al., 2023; Assouma et al., 2017; Soussana et al., 2010; Valentini et al., 2014), an increase in livestock population during the study period could result in the misestimation of soil CO_2 and N_2O emissions. Significant changes in the spatial distribution of animal load from one year to another could also lead to some uncertainties in the simulated spatial distribution of the emissions. Furthermore, it's worth noting that our estimate does not account for tree root respiration, which can lead to an underestimate of the total soil CO_2 emissions in regions with high tree density.

The soil C and N contents are significant factors influencing the spatial distribution of soil CO2 and N2O emissions in Sahelian SPSs, as indicated by our RF analysis. However, despite the availability of some local measurement data (Elberling et al., 2003a, b) and databases related to soil C and N content (Hengl et al., 2021) in the Sahel region, accurately assessing the temporal variability of these elements in Sahelian SPSs soils remains challenging.

Finally, we assumed that there were no natural or anthropogenic disturbances such as wildfires in Sahelian SPSs during our
495 simulation period. Uncertainties related to disturbances like wildfire are actually difficult to estimate as there are varying
perspectives and conflicting findings in the literature regarding the impact of burning on N_2O emissions (Karhu et al., 2015;
Takakai et al., 2006).

5 Conclusions and perspectives

Our study advances the understanding of the spatial distribution and annual budget of CO_2 and N_2O emissions from soil in the
500 Sahel. Information on the magnitude of CO_2 and N_2O emissions from soils in underrepresented areas are actually important to
shed light on the contribution of these areas to the overall GHGs budget and thereby inform the development of effective
mitigation strategies that can help reduce GHG emissions. SPSs represent a significant portion of the West African drylands,
where they have expanded due to global warming and are expected to continue expanding in the near future (Thornton and
Herrero, 2015). Previous studies at the local scale in the Sahel have shown that soils in semi-arid ecosystems are notable
505 contributors to GHGs emissions (Assouma et al., 2017; Brümmer et al., 2009; Delon et al., 2017). Our results extended these
local estimates to a broader spatiotemporal scale, showing that, overall, Sahelian SPSs soil emits less CO_2 and N_2O than
tropical areas and croplands, on a global scale. Furthermore, by mapping emissions we provided crucial insights into the
localization of soil CO_2 and N_2O emission hotspots, thereby offering indirect assessments of soil health in the Sahel region.
This information can be a valuable asset for land managers who can leverage it to devise and implement effective strategies
510 aimed at minimizing emissions and fostering carbon sequestration.

To further refine estimates of soil CO_2 and N_2O emissions in Sahelian SPSs, efforts to collect comprehensive datasets on
livestock spatial distribution and temporal dynamics, tree densities and fire are needed. Additionally, more experimental
studies should investigate the roles of nitrification and denitrification processes for soil N_2O emissions and the role of the
decomposition process for CO_2 emissions in semi-arid ecosystems to better parameterize the model.

6 Appendices

Appendix A:

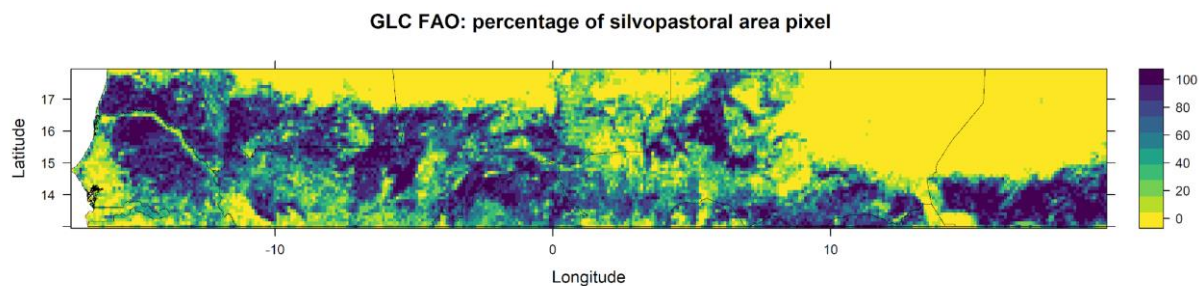**GLC FAO: percentage of silvopastoral area pixel**

24

Figure A1: Spatial distribution of silvopastoral areas in the Sahel. (Details on how the percentage of silvopastoral area pixels within the simulation grid cells were computed are provided in Methodology, Session 2.4)
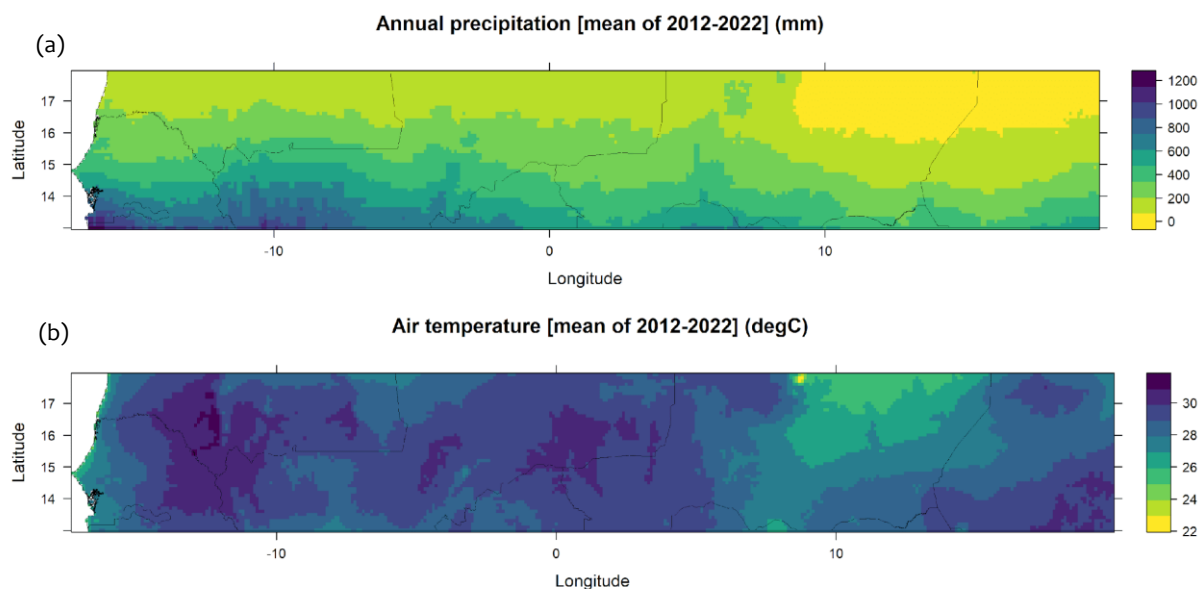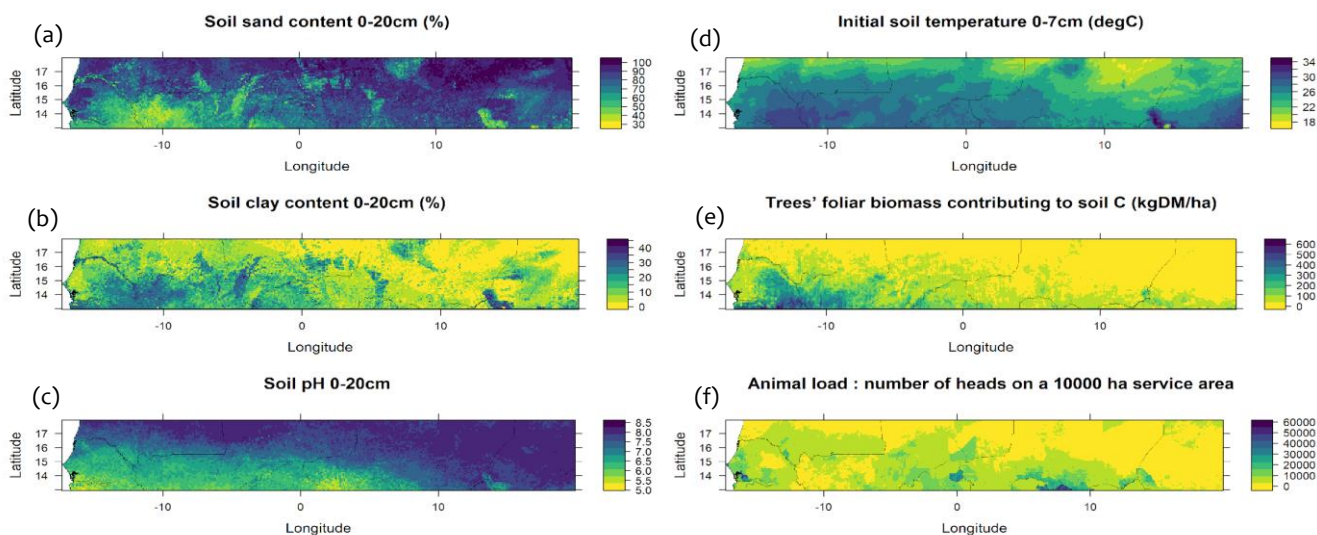
520

Figure A2: Spatial distribution of precipitation and air temperature (mean over 2012-2022).

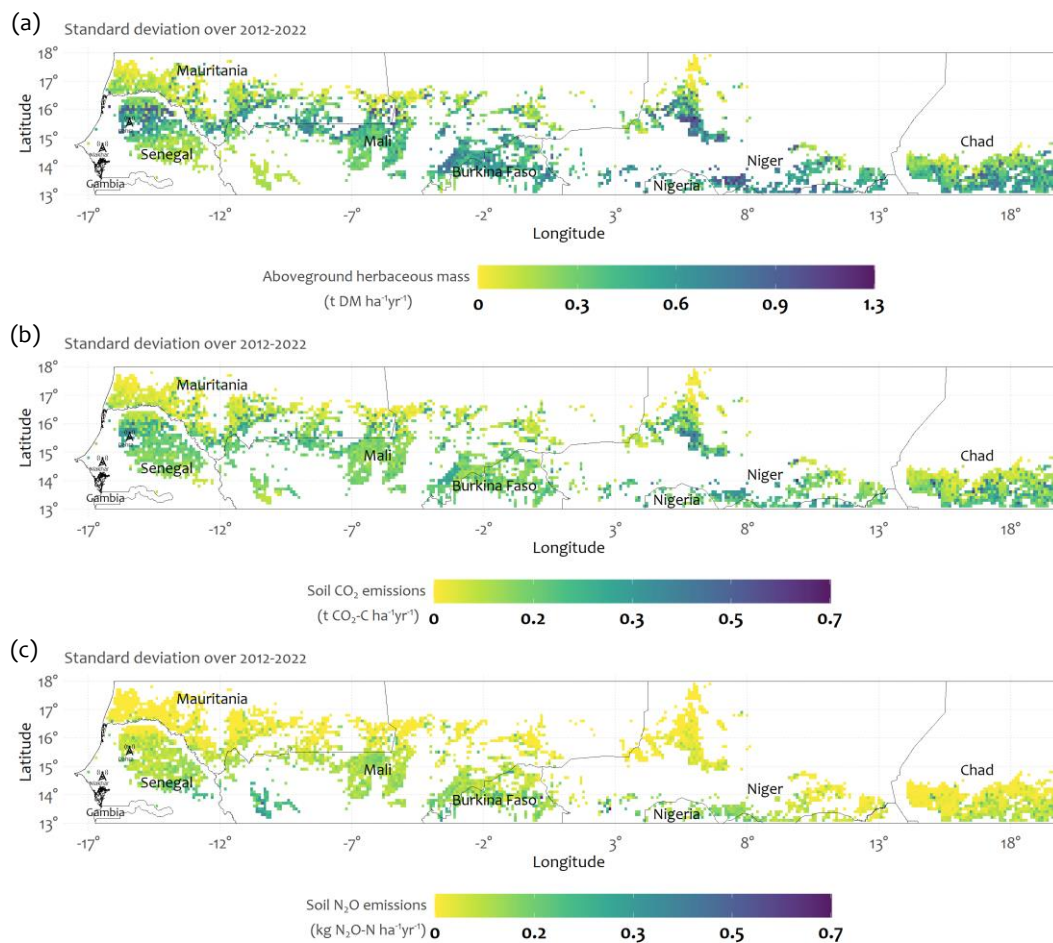525   **Figure A3: Spatial distribution of soil properties, trees' foliar biomass and livestock**

(a)

(b)

(c)

Figure A4: Standard deviations of the spatial distribution of (a) herbaceous biomass, (b) soil CO₂ and (c) soil N₂O emissions in Sahelian SPSs (over 2012-2022). Only pixels dominated by SPSs (>80%) are displayed

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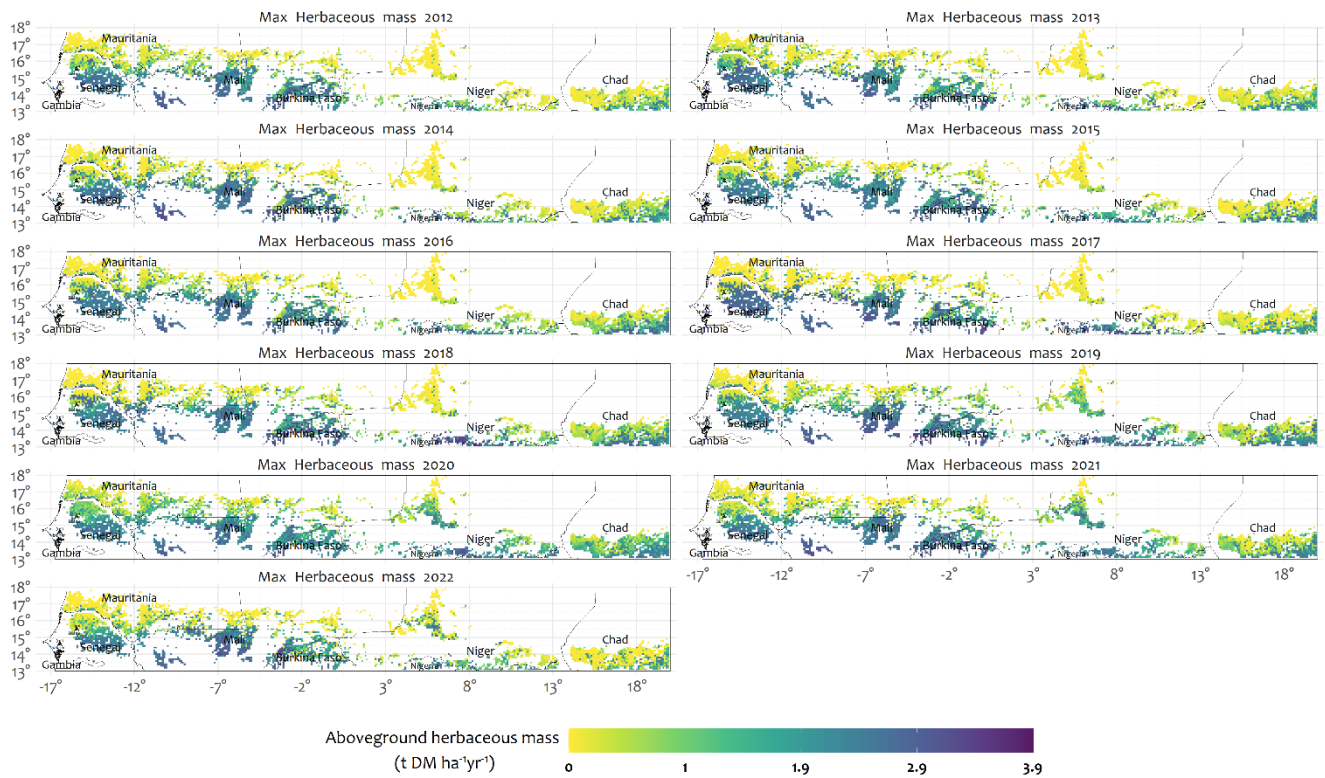
Figure A5: Annual spatial distribution of herbaceous biomass in Sahelian SPSs (2012-2022). Only pixels dominated by SPSs (>80%) are displayed
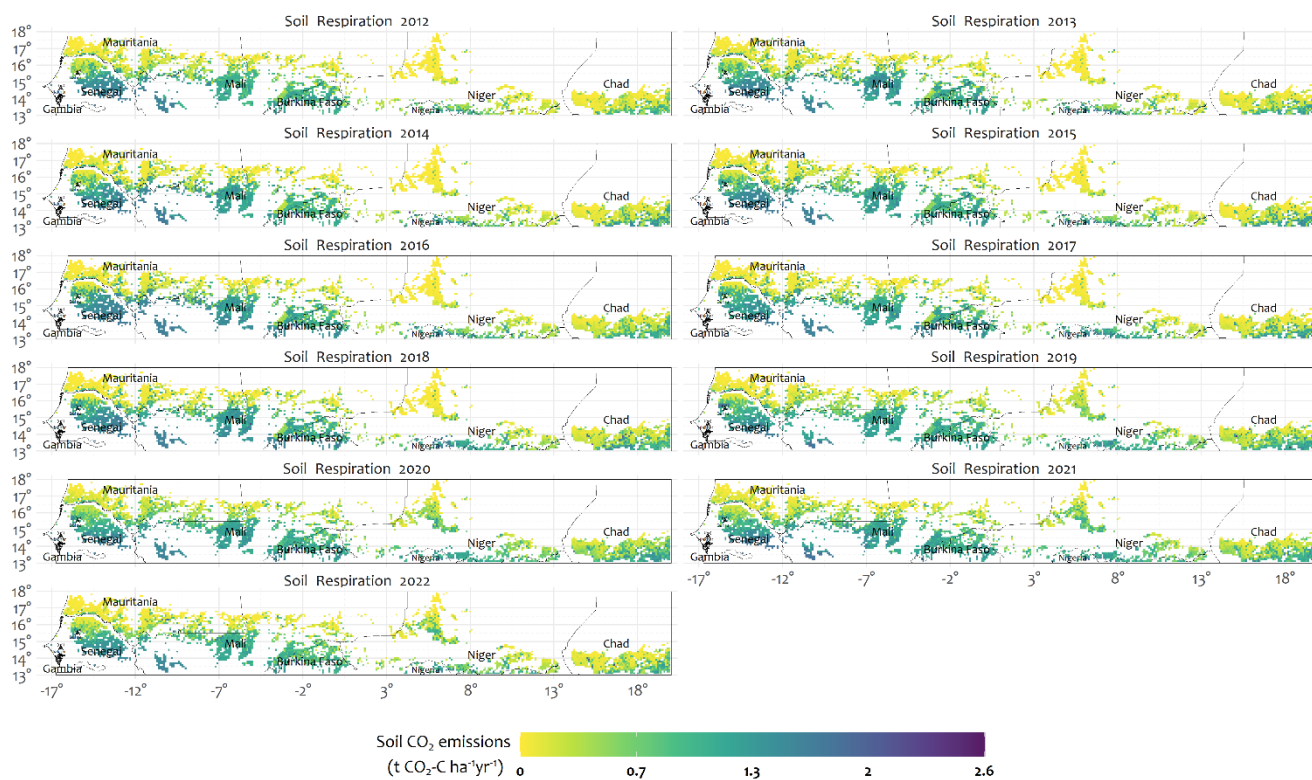
Figure A6: Annual spatial distribution of soil CO₂ emissions in Sahelian SPSs (2012-2022). Only pixels dominated by SPSs (>80%) are displayed
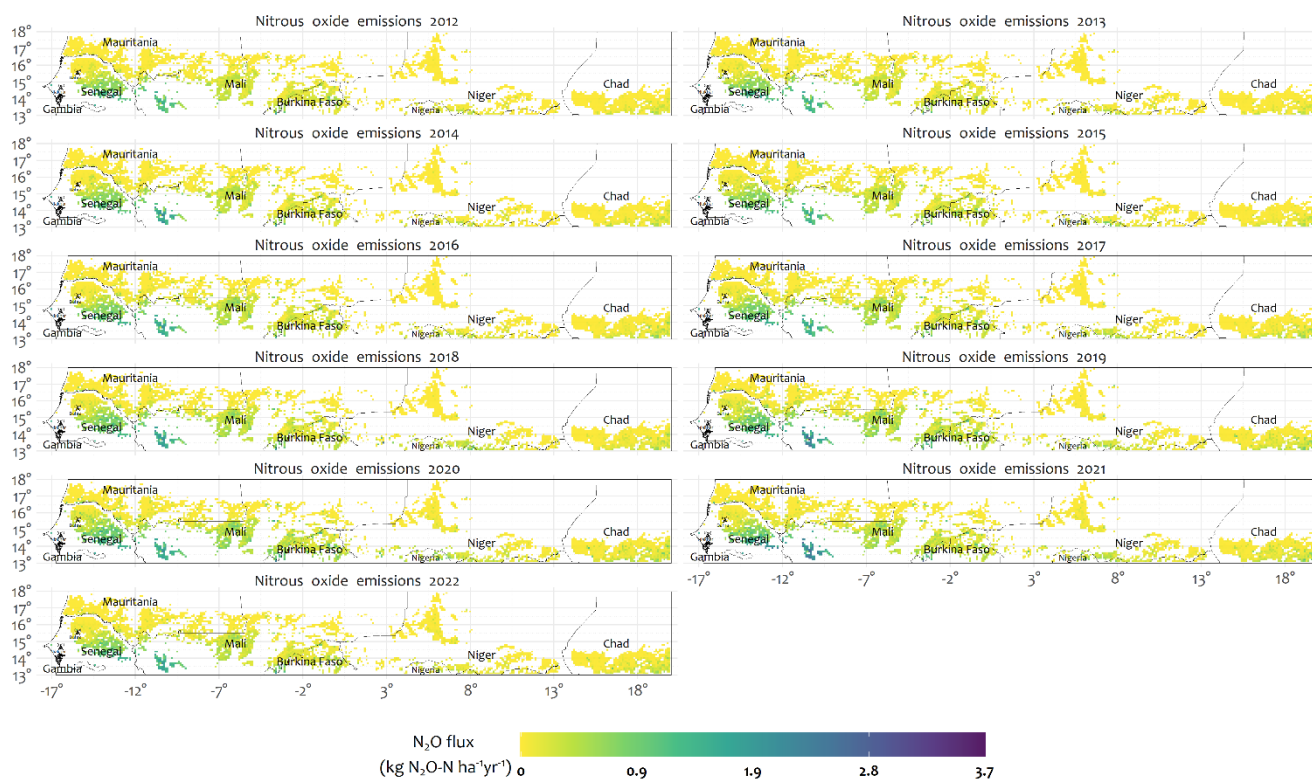
Figure A7: Annual spatial distribution of soil N₂O emissions in Sahelian SPSs (2012-2022). Only pixels dominated by SPSs (>80%) are displayed
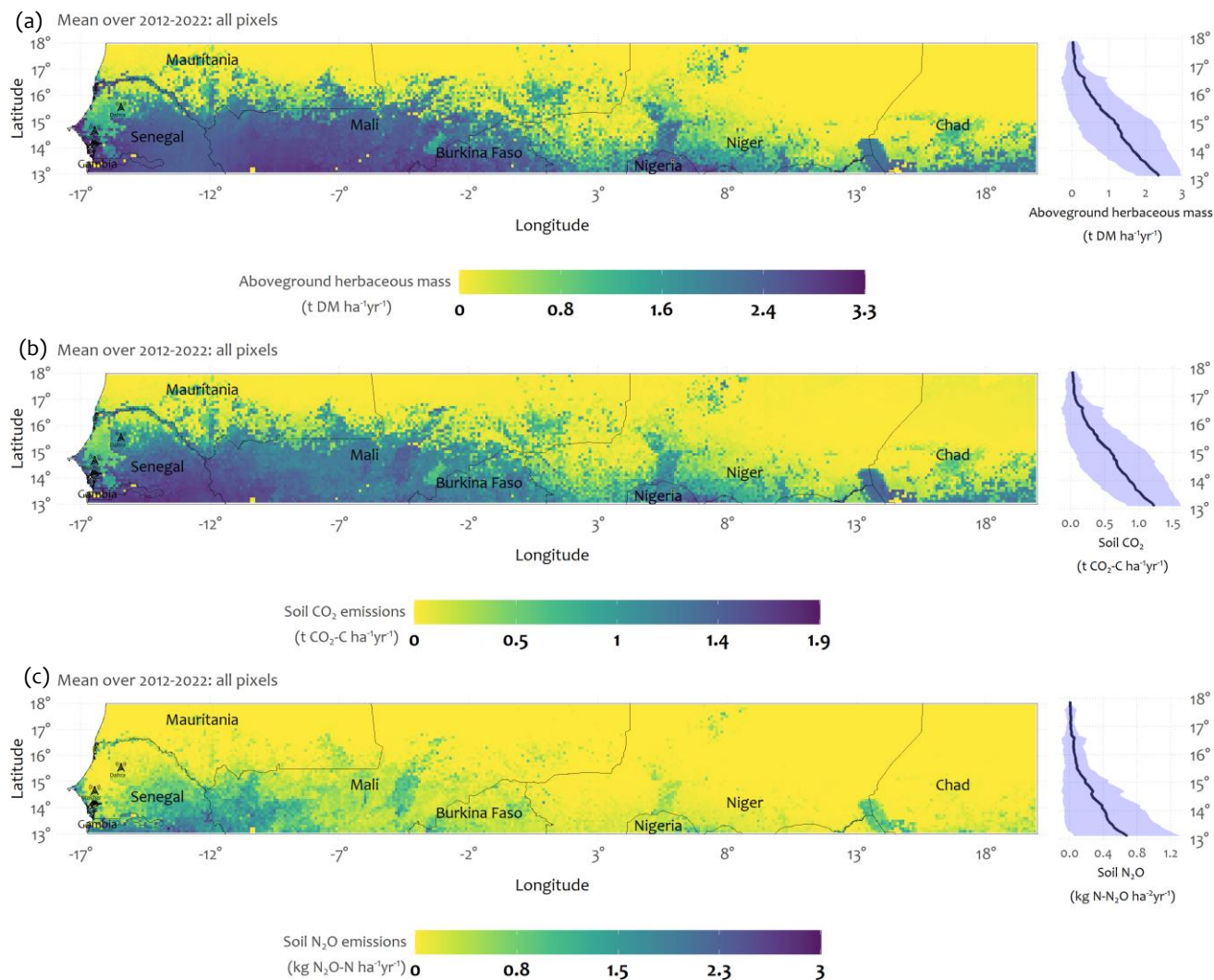
Figure A8: Regional distribution of simulated (a) herbaceous biomass, (b) soil CO₂ and (c) soil N₂O emissions in Sahelian SPSs (annual mean over 2012-2022). All pixels are displayed. The right panel shows (a) herbaceous biomass, (b) soil CO₂ and (c) soil N₂O emissions along a latitudinal gradient of 0.1°, while the shaded area indicates the standard deviation.

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Code availability. The 2D STEP-GENDEC-N₂O model is available on Zenodo at https://doi.org/10.5281/zenodo.7866671

Author contribution. YA, CD, MG, EM, and OR conceived and designed the study. EM, CD, MG, and YA developed the 2D STEP-GENDEC-N₂O model code and YA performed the simulations. YA prepared the manuscript with contributions from all co-authors.

550   **Competing interests.** The authors declare that they have no conflict of interest.

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