
1 **Comparison of water-soluble and insoluble organic compositions attributing to different light**
2 **absorption efficiency between residential coal and biomass burning emissions**

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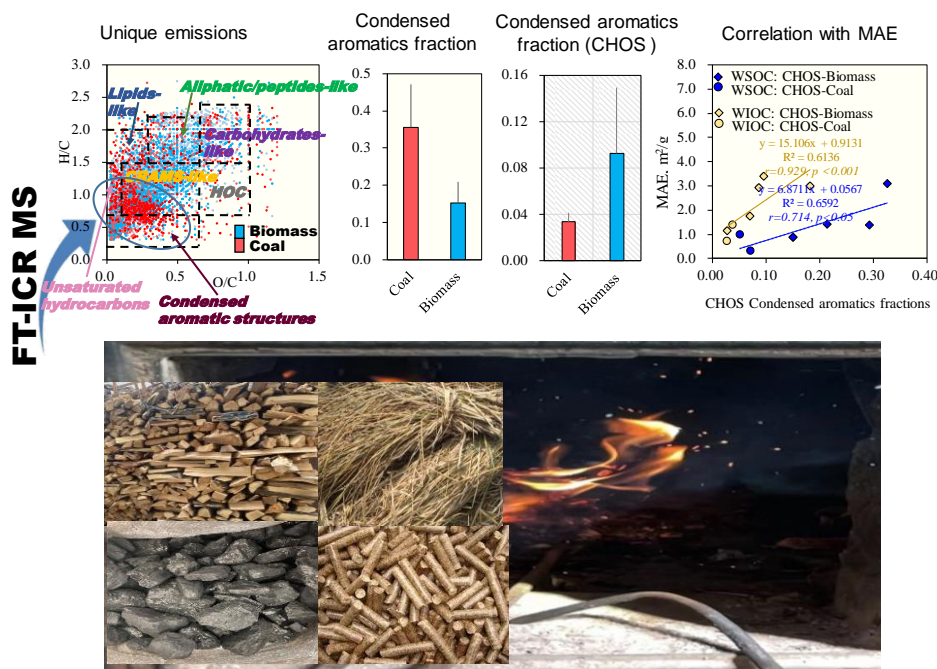
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12 **Abstract**

13 There are growing concerns about the climate impacts of absorbing organic carbon (also known as
14 Brown Carbon, BrC) in the environment, yet its chemical composition and association with the light
15 absorption capabilities remain poorly understood. This study characterized water-soluble and water-
16 insoluble organic carbon (WSOC, WIOC) from residential solid fuel combustion at the molecular
17 level and evaluated their quantitative relationship with mass absorption efficiency (MAE). The
18 MAE values at $\lambda=365$ nm from biomass burning were significantly higher than those from coal
19 combustion ($p<0.05$). Thousands of peaks were identified in the m/z range of 150-800, with the
20 most intense ion peaks occurring between 200-500 m/z for WSOC and 600-800 m/z for WIOC,
21 respectively. The CHO group predominated in WSOC extracts from biomass burning emissions
22 while sulfur-containing compounds (SOCs) including CHOS and CHONS were more intense in the
23 WIOC extracts, particularly from coal emissions. Emissions of the CHON group were positively
24 correlated with the fuel nitrogen content ($r=0.936$, $p<0.05$), explaining their higher abundance in
25 coal emissions compared to biomass. The SOCs emissions were more predominant during flaming
26 phases, as indicated by a positive correlation with modified combustion efficiency (MCE) ($r=0.750$,
27 $p<0.05$). The unique formulas of coal combustion aerosols were in the lower H/C and O/C regions
28 with higher unsaturated compounds in the van Krevelen (VK) diagram. In WIOC extracts, coal
29 combustion emissions contained significantly higher fractions of condensed aromatics (32-59%),
30 compared to only 4.3-9.7% in biomass burning emissions. However, the CHOS group in biomass
31 burning emissions was characterized by larger condensed aromatic compound fractions than those
32 in coal combustion. The CHOS aromatic compound fractions were positively correlated with MAE
33 values in both WSOC ($r=0.714$, $p<0.05$) and WIOC extracts ($r=0.929$, $p<0.001$), suggesting these
34 compounds significantly contributed to MAE variabilities across different fuels.

35 *Keywords:* light absorption properties, atmospheric aerosols, N-containing compounds, S-
36 containing compounds, water-soluble compounds, water-insoluble compounds.

37 **TOC**



38 **1. Introduction**

39 Light-absorbing organic carbon (OC), known as Brown Carbon (BrC), attracts growing
40 concerns due to its direct radiative impact on climate change (Laskin et al., 2015; Wang et al., 2022).
41 The global simulations suggested that BrC may contribute nearly 20% of the surface organic aerosol
42 (OA) burden (Jo et al., 2016), and accounted for 19% of the light absorption by anthropogenic
43 aerosols (Feng et al., 2013). BrC originates from diverse sources, including primary emissions from
44 coal combustion, biomass burning, and vehicular emissions (Du et al., 2014; Olson et al., 2015;
45 Bond, 2001; Sun et al., 2017; Chen and Bond, 2010), and secondary processes like the oxidation of
46 volatile organic compounds (Guan et al., 2020; Laskin et al., 2015). Among these sources,
47 residential solid fuel burning produced large amounts of BrC, accounting for 74% of the total
48 anthropogenic primary emissions (Xiong et al., 2022).

49 Some efforts have been made to explore the BrC physiochemical properties in residential
50 emissions. For example, water-soluble organic carbon (WSOC) and methanol-soluble organic
51 carbon (MSOC) were analyzed in primary emissions from combustions of crop straw, wood fuel,
52 and some coals (Park and Yu, 2016). Available studies suggested that the optical properties of
53 primary BrC varied significantly, influenced by numerous factors including inherent fuel properties
54 and combustion conditions. Mass absorption efficiency (MAE) is a key parameter in assessing the
55 direct radiative forcing of light-absorbing carbonaceous aerosols (Huo et al., 2018). It was found
56 that water-soluble BrC derived from bituminous coals had higher MAE values than anthracites
57 (Tang et al., 2020). However, the specific chemical components responsible for the differences in
58 light absorption among various fuel types are not yet fully understood, especially at the molecular
59 level. It was also reported that the mass absorption of water-insoluble organic carbon (WIOC) could
60 be even greater than that of the water-soluble ones (Chen and Bond, 2010; Huang et al., 2018).
61 However, there is limited information available regarding their chemical components. Given the
62 significant variation in BrC light absorption across different fractions emitted from various fuels
63 (Xie et al., 2017), detailed information of a comprehensive characterization, including the chemical
64 and optical characteristics of the BrC fractions (both water-soluble and water-insoluble) from the
65 combustion source, is needed.

66 This study investigated the chemical compositions and light absorption abilities of WSOC and
67 WIOC in smoke particles generated from the burning of coals with different maturity, raw biomass
68 fuels, and biomass pellets in traditional and improved stoves. The use of biomass pellets has been
69 heavily promoted over the last several years to mitigate air pollutant emissions from traditional solid
70 fuels, and emission characteristics from improved stoves could be different from traditional ones.
71 Optical property variations were quantitatively assessed and analyzed to explore their association
72 with chemical components. Furthermore, unique molecules and fingerprints for coal and biomass

73 sources were discussed, which is critical in pollution source appointment.

74 **2. Materials and methods**

75 2.1 Laboratory combustion emission experiment

76 In the present study, a total of fourteen types of coals with varying maturity degrees, five types
77 of biomass pellets, and twelve types of raw biomass were examined using a laboratory combustion
78 system. Two types of stoves, including a traditional stove (TS) and an improved stove (IS), were
79 utilized for the experiments. The thirty-four fuel-stove combinations are listed in Table S1. The
80 combustion tests were performed in a specially designed system with real-time online monitors
81 (Thermo Scientific Inc., Bremen, Germany), which are capable of continuously measuring gaseous
82 pollutants, including CO, CO₂, hydrocarbons (HC), and nitrogen oxide (NO_x including NO and
83 NO₂). PM_{2.5} (particles with aerodynamic diameters $\leq 2.5 \mu\text{m}$) was collected at a flow rate of 16.7
84 L/min with the quartz filters. Fuel properties including moisture (M_{ad} , %), volatile matter content
85 (V_{daf} , %), ash content (A_{ad} , %), lower heating value (LHV, MJ/kg), and contents of carbon (C, %),
86 hydrogen (H, %), nitrogen (N, %), and sulfur (S, %) are tested, and listed in Table S1. Details of
87 stove construction and combustion processes were available in Zhang et al., (2022).

88 Modified combustion efficiency (MCE) was calculated by integrating the incremental
89 concentrations of CO₂ and CO as:

$$90 \quad \text{MCE} = \frac{\text{CO}_2}{\text{CO}_2 + \text{CO}}$$

91 where CO₂ and CO represent the excess molar mixing ratios of CO₂ and CO, respectively
92 (Pokhrel et al., 2016). The MCE values indicate different phases of combustion: approximately 1
93 during flaming, and between 0.7-0.9 during smoldering (Yokelson et al., 1997)

94 2.2 Bulk carbon and UV-vis absorption spectra

95 For each sample, a 4.9 cm² filter was extracted ultrasonically with 10 mL Milli-Q water (18.2
96 MΩ) for 30 min, and then the supernatant was separated. The extraction process was repeated twice,
97 and the extracts were combined. The water extract was then filtered via a 0.22 μm
98 polytetrafluoroethylene (PTFE) filter to obtain WSOC. Due to the inefficiency of water in extracting
99 BrC, the insoluble PM components remaining on the sample filter were freeze-dried and subjected
100 to methanol extraction through sonication. The resulting extract was then filtered using a PTFE filter,
101 yielding a water-insoluble fraction referred to as WIOC in the following text. The carbon content
102 of WSOC was analyzed with a total organic carbon (TOC) analyzer (TOC-Lcph/cpn, SHIMADZU,
103 Japan), and the WIOC was determined by subtracting WSOC from the total OC loaded on the same
104 4.9 cm² area, given that the methanol extraction efficiency for all combustion samples was up to
105 90% (Zhang et al., 2022). The element carbon (EC) and OC were measured using a thermal-optical
106 analyzer (Sunset OC/EC analyzer) with an interagency monitoring of protected visual environments
107 (IMPROVE) program.

108 The light absorption spectra of the water and methanol extracts were measured between 200
109 nm and 600 nm by UV-vis spectrophotometer (UV-2600, Shimadzu, Japan) at a step size of 1 nm.
110 MAE values of WSOC and WIOC at the wavelength of λ (MAE_{λ, WSOC}; MAE_{λ, WIOC}) were calculated
111 as following equation (Li et al., 2019):

$$112 \quad \text{MAE}_{\lambda, \text{WSOC}} = A_{\lambda, \text{WSOC}} \times \ln(10) / (C_{\text{WSOC}} \times L); \text{MAE}_{\lambda, \text{WIOC}} = A_{\lambda, \text{WIOC}} \times \ln(10) / (C_{\text{WIOC}} \times L)$$

113 where $A_{\lambda, \text{WSOC}}$ and $A_{\lambda, \text{WIOC}}$ is the light absorption value of WSOC extract and WIOC extract
114 at a wavelength of λ, respectively; C is the concentration of WSOC (or WIOC), and L is the optical
115 path length, which is 0.01 m in this study. It is important to note that the reported light absorption
116 of WIOC in this study was underestimated, while such underestimation is insignificant due to the
117 high extraction efficiency. In this study, the MAE values of extractable OC at λ of 365 nm (MAE₃₆₅,

118 wsOC and MAE_{365, WIOC}) were discussed.

119 Absorption Ångström exponent (AAE) values were determined based on the following
120 equation (Li et al., 2019; Li et al., 2020):

$$121 \quad A_{\lambda} = K_{\lambda}^{-AAE}$$

122 where K is a constant and AAE is obtained through the linear regression of $\lg(A_{\lambda})$ against $\lg \lambda$
123 (Fig S1). A wavelength of 300-400 nm is chosen according to the published literature (Yue et al.,
124 2022), and the goodness of fit for all the samples in this study is greater than an r^2 of 0.99.

125 2.3 Molecular composition analysis

126 For further molecular composition analysis, the WSOC and WIOC extracts from seven selected
127 source samples were subjected to Fourier-transform ion cyclotron resonance mass spectrometry (FT-
128 ICR MS) coupled with electrospray ionization (ESI). These samples included two types of coals
129 (high volatile bituminous coal, HVB; medium volatile bituminous coal, MVB) combusted in TS,
130 two types of raw biomass (rice straw and pine wood) combusted in TS, pine wood combusted in IS,
131 and two types of biomass pellets (crop straw pellet and pine wood pellet) combusted in IS as noted
132 in SI. FT-ICR MS has been successfully applied for the molecular-level characterization of
133 compounds (Bianco et al., 2018) due to its ultrahigh resolution and mass accuracy. ESI is well-
134 adopted to characterize soluble aerosols, especially for the detection of polar, hydrophilic molecules
135 like humic-like substances (HULIS)-type compounds (Wozniak et al., 2008), because it is a “soft”
136 ionization technique generating minimal analyte fragments, and thus can detect intact molecules of
137 compounds. Therefore, the negative ESI-FT-ICR was applied here to determine the molecular
138 compositions of WSOC and WIOC from combustion samples from different solid fuels. The
139 methanol extracts were evaporated to dryness under a gentle stream of nitrogen and then dissolved
140 with Milli-Q water. The WSOC and water-reconstituted WIOC were submitted to solid-phase

141 extraction (SPE) using Bond Elut PPL (500 mg, 6 mL, Agilent, U.S.A.). Before the extraction, the
142 PPL cartridges were sequentially conditioned with 12 mL methanol and 12 mL Milli-Q water
143 containing 0.05% hydrochloric acid (HCl). The extract was adjusted to pH=2 using HCl to remove
144 inorganic ions and was then loaded onto the PPL cartridges at a rate of 5 mL/min. Cartridges were
145 washed with 18 mL Milli-Q water containing 0.05% HCL to remove salt and then dried under pure
146 nitrogen. Analytes were eluted with 12 mL methanol, and the combined eluates were concentrated
147 to 1 mL. Then the molecular characterization was conducted using a 15T Solarix XR FT-ICR MS
148 (Bruker Daltonik GmbH, Bremen, Germany) in the negative ESI mode. The capillary inlet voltage
149 was set at -4.0 kV and ion accumulation time was set to 0.06 S. There were 300 continuous 4 M
150 data FT-ICR transients added to improve the signal-to-noise ratio. The FT-ICR MS was calibrated
151 with 10 mmol/L sodium formate in advance, and internal standard calibration with soluble organic
152 matter (known molecular formula) was performed after the test. Finally, <1 ppm absolute mass error
153 was achieved. Data processing details are described in the Supporting Information (SI).

154 **3. Result and discussion**

155 3.1 Optical characteristics of WSOC and WIOC

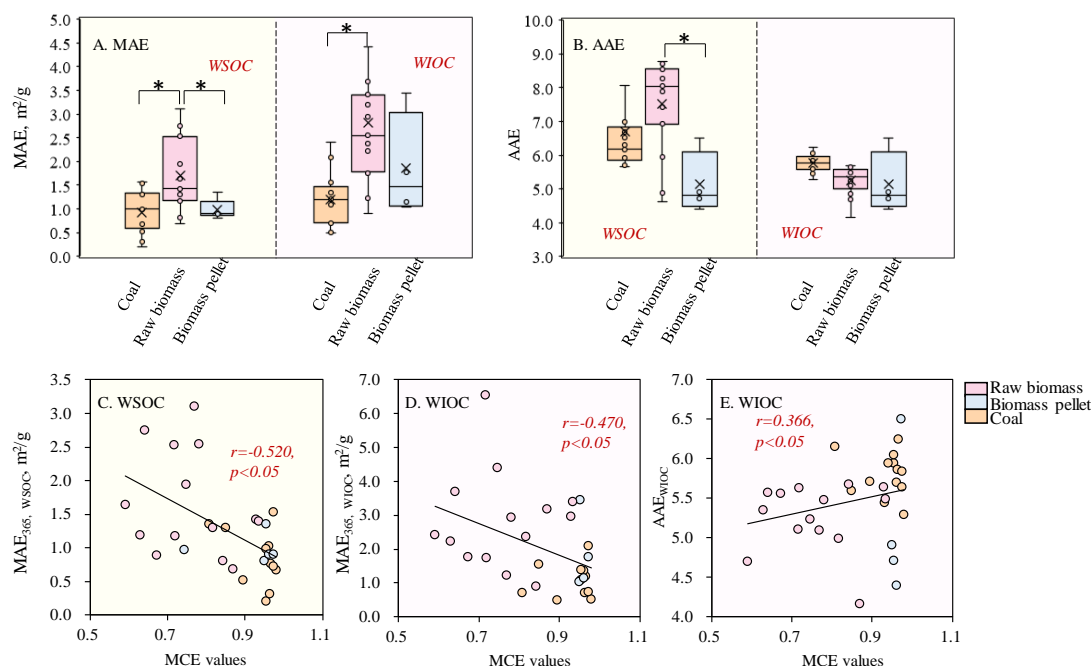
156 MAE is an important parameter reflecting the light absorption capability of the carbonaceous
157 aerosols. The MAE_{365, WSOC} of aerosols from residential sources in this study ranged from 0.21 to
158 3.1 m²/g with an average of 1.3±0.7 m²/g. MAE values of extractable OC in this study were lower
159 than that of 11.3 m²/g for pure BC aerosols (Bond and Bergstrom, 2006) and also lower than that
160 from the filter-based MAE values of OC of 0.16-13 m²/g from residential sources (Zhang et al.,
161 2021b). A significant difference ($p<0.05$) of MAE_{365, WSOC} was observed among the different fuels
162 (Fig.1). The MAE_{365, WSOC} of raw biomass combustion derived aerosols averaged at 1.7±0.8 m²/g,
163 which was significantly higher ($p<0.05$) than those from coal smoke (0.93±0.44 m²/g) and biomass

164 pellet smoke ($1.2 \pm 0.6 \text{ m}^2/\text{g}$). The $\text{MAE}_{365, \text{WIOC}}$ ranged from 0.49 to $6.6 \text{ m}^2/\text{g}$ with an average of
165 $2.0 \pm 1.3 \text{ m}^2/\text{g}$. The absorption capability was higher for the WIOC extract than the WSOC extract.
166 This is thought to be associated with distinct chemical compositions of light-absorbing organics. It
167 was noted that the WIOC had higher MAE values compared to the WSOC, which may be explained
168 by the more hydrophobic PAHs and quinones (Chen and Bond, 2010). The difference in $\text{MAE}_{365, \text{WIOC}}$
169 among the different fuels was also statistically significant ($p < 0.05$), as the value obtained from
170 raw biomass burning ($2.8 \pm 1.4 \text{ m}^2/\text{g}$) was significantly higher than that from coal combustion
171 ($1.2 \pm 0.6 \text{ m}^2/\text{g}$). It was suggested that the MAE values of soluble OC including WSOC and WIOC
172 were dependent on the chemical composition of OC, that is, the chemical structure of the light-
173 absorbing chromophores and the ratio of non-light-absorbing organics to the chromophore
174 components (Cao et al., 2021). The higher MAE values of raw biomass combustion-derived aerosols
175 might be caused by the stronger light absorption capability of chromophores or a higher ratio of
176 chromophores. Our result was comparable to the published data, for example, the $\text{MAE}_{365, \text{WSOC}}$ was
177 reported to average at $1.37 \pm 0.23 \text{ m}^2/\text{g}$ for biomass burning emissions (Park and Yu, 2016). The
178 correlation of MAE with the MCE was investigated. The variability in MAE values, for both WSOC
179 and WIOC, was observed to be negatively correlated with the MCE (or temperature as MCE was
180 found to be positively correlated with the measured temperature in emission exhausts) when pooling
181 all data together as seen in Fig. 1C and 1D. However, within each fuel group, there was no
182 statistically significant correlation between the MAE and MCE values (Fig. S2). Some previous
183 studies found that OC from wood pyrolysis under higher temperature conditions had stronger
184 absorption capability (Chen and Bond, 2010; Saleh et al., 2014), but relatively higher mass
185 absorption coefficient (MAC) values were also reported for the organic aerosol from wood
186 combustion under the $150^\circ\text{C} < T < 250^\circ\text{C}$ compared to emissions at a lower ($T < 150^\circ\text{C}$) or higher
187 ($250^\circ\text{C} < T < 380^\circ\text{C}$) temperature condition (Rathod et al., 2017). Chen et al., (2010) reported that that

188 absorption per mass (α/ρ) of methanol extracts increased with increased wood pyrolysis temperature,
189 but such increase was nonlinear and varied in burning emissions of different fuel types or wood with
190 different sizes. Therefore, the negative correlations between MAE and MCE values when pooling
191 all data together in Fig. 1 were from distinct absorption properties of emissions from different fuel
192 types, rather than conditions like combustion temperature.

193 The AAE which indicates the wavelength dependence of light absorption is also an important
194 parameter in climate models. The calculated AAE in the WSOC extract (AAE_{WSOC}) ranged from
195 3.8 to 11 with an average of 6.9 ± 1.5 , and the difference in AAE_{WSOC} values among the different
196 fuels was statistically significant ($p < 0.05$) (Fig.1). The highest values were observed for the aerosols
197 from raw biomass combustion of 7.5 ± 1.4 , followed by coal smoke of 6.7 ± 1.5 and biomass pellet
198 smoke of 5.6 ± 1.2 . The AAE values in the WIOC extract (AAE_{WIOC}) were slightly lower than
199 AAE_{WSOC} , which ranged from 4.2 to 6.5 with an average of 5.5 ± 0.5 . The differences in AAE_{WIOC}
200 values among the different fuels were statistically insignificant ($p > 0.05$), with averaged values of
201 5.8 ± 0.3 for coal smoke, 5.1 ± 0.9 for biomass pellet smoke, and 5.2 ± 0.4 for raw biomass smoke.
202 There was a weak positive correlation of AAE_{WIOC} between the MCE values ($p < 0.05$). The filter-
203 based analysis also showed that AAE values were positively correlated with MCE values, indicating
204 that more BrC were produced under the smoldering phase compared with BC (Zhang et al., 2020).
205 This study suggested that BrC in WIOC extract was apt to be generated during the smoldering phase
206 in comparison with the non-light-absorbing OC. The AAE values in soluble OC in this study were
207 higher than 4 for all samples, confirming the contribution of BrC to aerosol absorptivity from source
208 emission. The result of this study was comparable to the published literature. For example, it was
209 reported that the AAE_{WSOC} was in the range of 8.6-15 from coal combustion-derived aerosols (Song
210 et al., 2019), and 6.2-9.3 from biomass smoke (Park and Yu, 2016). The AAE_{WIOC} from wintertime

211 urban aerosols were 5.4 ± 0.2 in Xi'an and 5.7 ± 0.2 in Beijing (Huang et al., 2020).



212 **Fig. 1** MAE values at $\lambda=365$ nm (A) and AAE values (B) from the source samples (* represents
 213 $p < 0.05$); Correlation between the MCE values with MAE_{365, WSOC} values (C), MAE_{365, WIOC} values
 214 (D), and AAE_{WIOC} values (E).

215 3.2 Molecular characteristics of WSOC and WIOC

216 The ESI-FT-ICR mass spectra of WSOC and WIOC samples are presented in Fig. 2. Thousands
 217 of peaks were identified in the m/z range of 150-800, indicating a complex chemical composition
 218 of aerosols from residential sources. Formulas detected in the raw biomass burning aerosols were
 219 significantly higher than those in biomass pellet and coal smoke (Fig. S3), which indicated a higher
 220 chemical complexity of raw biomass emissions. In addition, the combustion of pine wood in the
 221 improved stove generated fewer compounds, which was 93% of the peaks identified in the
 222 traditional stove (Table S2). The likely less complexity might be due to higher combustion
 223 efficiencies and temperature in the improved stove. Generally, higher levels of organic aerosol mass
 224 would be emitted during less efficient fuel burning, resulting from prolonged smoldering or

225 incomplete burning (Holder et al., 2016). Our results suggested that corresponding higher chemical
226 complexity was also produced during the incomplete combustion in the traditional stove. The most
227 intense ion peaks were distributed in the 200-500 m/z for WSOC, accounting for 58-86% of the
228 total intensity. Similar results were also found in residential coal combustion (Song et al., 2019),
229 biomass burning (Song et al., 2018), and ambient aerosols (Wozniak et al., 2008).

230 The mass spectra of WIOC are different from the WSOC, especially in aerosol emitted from
231 coal combustion (Fig. 2). WIOC contained more molecules with larger m/z values of 600-800 in
232 range, which indicated that WIOC extract had more compounds with higher molecular weight (MW)
233 than the WSOC extract. According to the molecular formulas and the intensity of each negative ion,
234 the average molecular formulas for the WSOC were obtained with C atom from 20 to 24, H (21-
235 29), N (0.32-0.75), O (5.6-7.0), and S (0.28-0.51) in the WSOC extract. All aerosols from biomass
236 burning, either raw or pelletized ones, had higher relative O atom contents than coal smoke,
237 indicating a higher oxidation degree of biomass emissions. For the WIOC, the average molecular
238 formulas were assigned with 27-33 C, 26-35 H, 0.67-1.2 N, 6.6-11 O, and 0.34-0.92 S. The coal
239 combustion-derived aerosols had more C, N, and S atoms, but less H atom, compared with raw
240 biomass. The combustion of biomass pellets is also assigned with relatively higher S elements. In
241 addition, the WIOC fraction had a higher relative atom content than the corresponding formulas of
242 WSOC from the same source aerosol samples. These results indicated that in addition to the fuel
243 type, extraction solvent also had an important impact on the elemental composition of extractable
244 BrC.



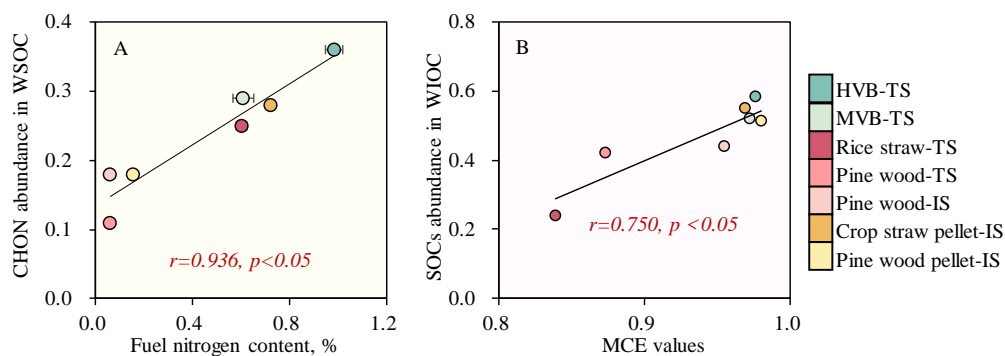
245 **Fig. 2** Negative ESI FT-ICR mass spectra of WSOC (A-G) and WIOC (A'-G') from the seven
 246 aerosol samples, TS represents combustion in the traditional stove and IS represents combustion in
 247 the improved stove. Different formula groups were color-coded. The pie charts showed the relative
 248 intensities of different formula groups.

249 Molecular formulas identified by the FT-ICR-MS can be classified into 4 groups according to
 250 the elemental composition, including CHO (containing only C, H, O), CHON (hereafter similarly),
 251 CHOS, and CHONS. CHO was the most abundant group in the WSOC. The CHO group contributed
 252 51-67% to the total intensity of aerosols from raw biomass burning, which were significantly higher
 253 than those from biomass pellets (29-39%) and coal smoke (30-40%). The CHO compounds with
 254 oxygen-containing functional groups (e.g., hydroxyl, carbonyl, carboxyl, or esters) have been
 255 widely identified in both ambient aerosols (Jiang et al., 2021; Mo et al., 2022) and some source

256 samples (Tang et al., 2020). These compounds contributed a broad range of proportions, from 43%
257 to 69% in residential biomass burning smoke (Tang et al., 2020), 9.7-48% in coal smoke (Song et
258 al., 2019), and 20-39% in ambient aerosols (He et al., 2023) for the WSOC extract, which were
259 comparable to ours. CHON compounds were also an important component in the aerosols,
260 accounting for 29-36% of coal smoke, which was significantly higher than biomass burning smoke
261 of 11-28%. One previous study reported that CHON species were more abundant in biomass burning
262 smoke rather than in coal combustion ones (Song et al., 2018). This high fraction of CHON
263 compounds in coal smoke might be caused by the higher nitrogen content of coals as seen in Table
264 S1. A strongly significantly positive correlation was found between the fuel nitrogen content and
265 CHON species percentage ($r=0.936$, $p<0.05$) (Fig. 3). Such dependence was also found in the
266 emission factors (EFs) of NO_x on fuel nitrogen content in our result (Fig. S4), as well as NO_x , HCN,
267 and NH_3 emissions reported by published literature (Hansson et al., 2004). Sulfur-containing
268 compounds (SOCs; including CHOS and CHONS) abundance was lower than CHO and CHON
269 group, accounting for only 22-42% (13–30% for CHOS and 7.9–28% for CHONS, respectively).
270 The fractions of SOC in the aerosols from coal combustion (25-42%) and biomass pellet burning
271 (~42%) were comparable but significantly higher than those from raw biomass (22-31%). The
272 abundance of SOC was not statistically correlated with fuel S content in the present study for
273 pooled data. However, when fuel subgroups were considered, fuels (coal or raw biomass) with
274 higher sulfur content had higher SOC levels. Also, higher SOC emission was found for pine wood
275 combusted in the improved stove which have higher combustion efficiency than that in the
276 traditional stove. These results suggested that except for fuel sulfur content, the fuel type and
277 combustion efficiency would also influence the SOC emissions.

278 In the WIOC, the intensities of these four groups among different source samples were different
279 from those in the WSOC. The CHO group accounted for 27-45% in raw biomass burning aerosols,

280 22-24% in biomass pellet smokes, and 13-22% of coal smokes. On the contrary, the SOCs
 281 abundance was significantly higher in WIOC, especially for the coals, accounting for 58% for HVB
 282 and 52% for MVB. A significantly positive correlation between the MCE values and SOCs
 283 abundance in the WIOC extract was found ($r=0.750, p<0.05$) (Fig. 3), while no clear association
 284 between fuel sulfur content and SOCs emissions was observed in the WIOC. Combining the result
 285 from the WSOC extract, the strong dependence of SOCs emissions on the combustion conditions
 286 was expected, while the fuel sulfur content had a slight influence. Notably, the SOCs abundance in
 287 aerosols from coal combustion was greatly higher than that in ambient aerosols (Lin et al., 2012),
 288 indicating that residential non-desulfurized coal combustion might be an important emitter of SOCs
 289 in atmospheric samples. The CHON group emissions were determined mainly by the fuel nitrogen
 290 content, while SOCs emissions were strongly related to the combustion conditions (e.g., flaming
 291 phase). It should be noted that the results of our analysis based on the current data without isotopic
 292 internal standard used were semi-quantitative, some uncertainties must be addressed.



293 **Fig. 3** Correlation between fuel nitrogen contents and abundances of CHON in WSOC (A), and the
 294 correlation between MCE values and the abundances of SOCs in WIOC (B). TS represents
 295 combustion in the traditional stove and IS represents combustion in the improved stove.

296 3.3 Detailed CHO/CHON/CHOS/CHONS group differences across fuel types

297 Van Krevelen (VK) diagrams which can provide a visual interpretation of complex mass
 298 spectra can qualitatively identify chemical composition profiles in mixtures (Lv et al., 2016). The

299 classification criteria of the VK diagram are provided in Table S3 (Patriarca et al., 2018; Tang et al.,
300 2020). Fig. S5 and S6 show the VK diagrams of WSOC and WIOC. In both WSOC and WSOC
301 extract, the carboxylic-rich alicyclic molecules (CRAMs-like) were the most abundant component,
302 contributing to 53-69% in the WSOC extract and 37-56% in the WIOC extract, respectively. The
303 condensed aromatics were also an important component in source samples, accounting for 7.3-13%
304 in the WSOC, and for a higher proportion of 8.6-44% in the WIOC. This observation could be
305 attributed to the hydrophobic property of condensed aromatic hydrocarbons, leading to a lower
306 proportion in the highly polar WSOC. Among the identified formulas, 7.0-13% of total intensity in
307 the WSOC and 3.6-17% in the WIOC were the aliphatic/peptide-like compounds. Such fractions
308 were comparable to unsaturated hydrocarbons with percentages of 3.9-15% in the WSOC and 2.0-
309 11% in the WIOC. The compounds including lipids-like species and highly oxygenated compounds
310 (HOC) were in a relatively lower abundance, with less than 10% each in the source samples. The
311 different fuels showed varied chemical composition characteristics. Coal combustion aerosols had
312 lower H/C ratios than those in biomass burning aerosols, indicating higher unsaturated degrees
313 (Table S2). As indicated by the VK diagrams, the coal combustion produced a notable number of
314 condensed aromatics, contributing to 27%-44% of the WIOC, which was significantly higher than
315 8.6-21% in biomass burning emissions. The modified aromaticity index ($AI_{mod,w}$) which is a
316 measure of aromatic and condensed aromatic structure fractions, and the double bonds equivalent
317 (DBE) values which are used as a measure of the unsaturated level in a molecule were all higher for
318 coal emissions compared to the biomass emissions.

319 Distinct compound profiles being identified by the VK diagram classification criteria are
320 consistent with discrepancies in the four groups (CHO/CHON/CHOS/CHONS). For the CHO group,
321 the most intense compounds were CARMs compounds with fractions of 69-84% in the WSOC
322 extract and higher fractions of 51-80% in the WIOC extract. It was observed that raw biomass would

323 emit slightly more CRAMs (76-84% in the WSOC and 61-80% in the WIOC) than coal (69-73% in
324 the WSOC and 62-69% in the WIOC) and biomass pellet (70-75% in the WSOC and 51-56% in the
325 WIOC) combustion. The condensed aromatic compound was also a crucial component, accounting
326 for 13-20% of the WSOC and 19-27% of the WIOC from coal emissions. These fractions were
327 significantly higher than those found in raw biomass (4.5-10% in the WSOC, 7.9-10% in the WIOC)
328 and biomass pellet (3.8-5.4% in the WSOC, and 11% in the WIOC) (Fig. 4). The other components
329 accounted for a small (less than 10% each) of the total CHO intensity.

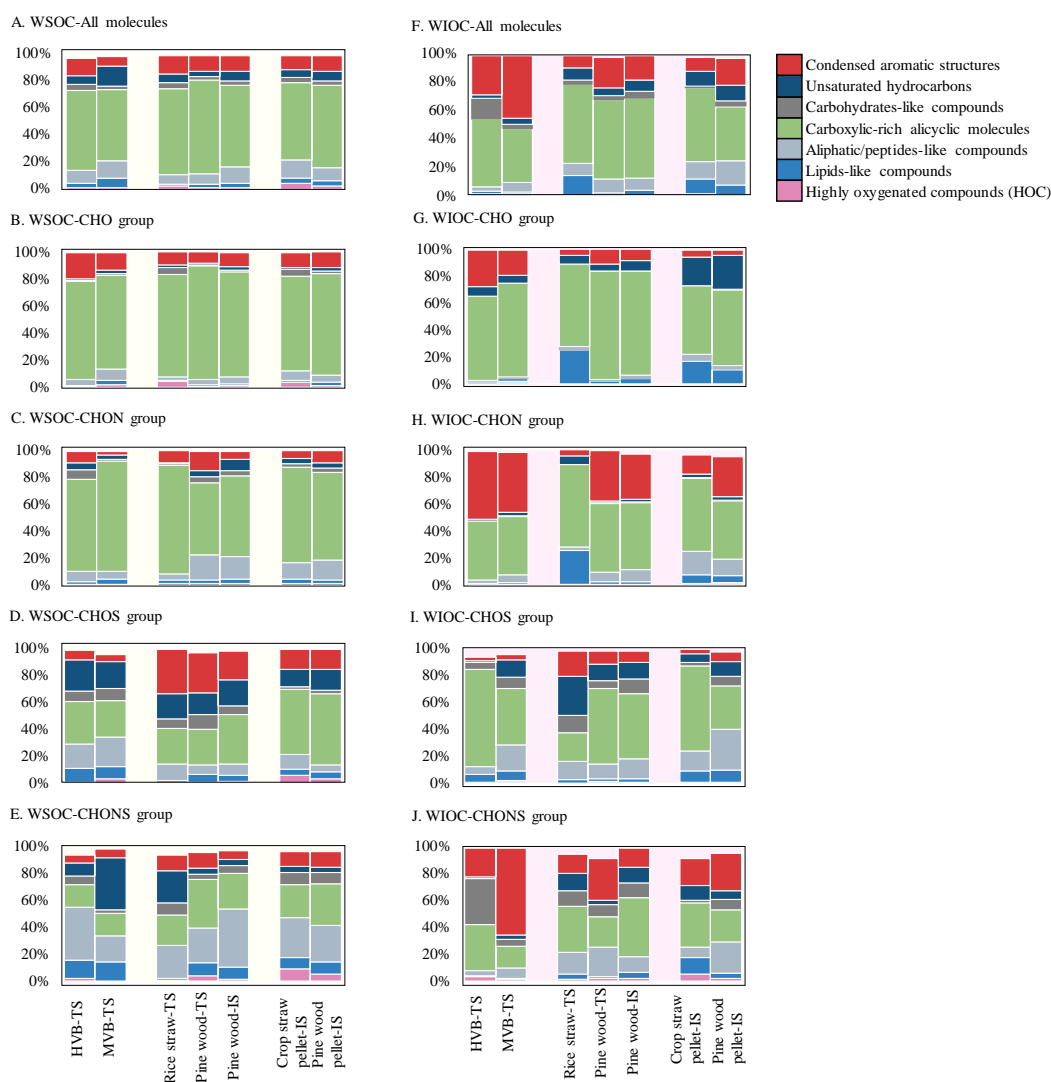
330 It was reported that the CHON compounds with $O/N \geq 3$ might be the organonitrates candidate
331 and nitro-substituted compounds, attributing to the allocation of one nitro (NO_2) or nitrooxy (ONO_2)
332 group (Bianco et al., 2018). In this study, the relative content of $\text{CHON}_{O/N \geq 3}$ compounds (concerning
333 the overall CHON) in the biomass (raw and pellet ones) ranged between 71-82% and 85-91% for
334 the WSOC and WIOC, which is distinctly lower than those found in coal smokes (86-90% for the
335 WSOC, and 86-95% for the WIOC) (Table S4). Moreover, the $\text{AI}_{\text{mod,w}}$ values for CHON compounds
336 from coal combustion were higher than the biomass smokes for both WSOC and WIOC, as indicated
337 by Table S2. It can thus be concluded that more CHON compounds with low aromaticity and a large
338 amount of oxidized nitrogen functional groups were formed during the combustion of biomass fuels.
339 Coal combustion emissions had more intense condensed aromatic compounds with a percentage of
340 45-50% for the CHON group, which were significantly higher than those from raw biomass (4.5-
341 37%) and biomass pellet burning (14-30%). This result confirmed the conclusion that the CHON
342 compounds produced from the combustion of coals were characterized by a relatively high
343 aromaticity and a low degree of oxidation.

344 The O-rich CHOS fraction ($O/S \geq 4$) accounted for 42-71% (concerning the overall CHON
345 group) in the WSOC and 45-95% in the WIOC. O-rich CHOS fraction ($O/S \geq 4$) content from coal

346 (66-70% in the WSOC, and 78-98% in the WIOC) and biomass pellet (67-71% in the WSOC, 76-
347 87% in the WIOC) were relatively higher than the raw biomass smoke (42-56% in the WSOC, and
348 32-57% in the WIOC), suggesting that most of the CHOS compounds in coal and pellet smoke can
349 be potentially assigned with more bearing sulfate (-OSO₃H) or sulfonate (-SO₃) groups. Different
350 from CHO and CHON groups, it was found that the CHOS group in the biomass burning aerosols
351 had a much more intense condensed aromatic structure with a percentage of 14-33% in the WSOC
352 and 2.5-19% in the WIOC, especially for raw biomass (WSOC: 22-33%, WIOC:8.5-19%) than coal
353 combustion (WSOC: 4.9-7.0%, WIOC: 2.5-7.0%). For the CHOS group in the WIOC extract, the
354 unsaturated hydrocarbons were also an important component, accounting for 12-29% of raw
355 biomass burning emissions, which were significantly higher than those in biomass pellet (6.9-11%)
356 and coal (1.5-13%). These results indicated biomass burning emissions were characterized by higher
357 unsaturation levels and aromaticity for the CHOS group than coal combustion.

358 Nearly 34-85% of CHONS formulas have many O atoms (≥ 7), indicating the existence of the
359 -NO₃ group (Table S4). These CHONS compounds are probably nitrooxy-organosulfates (Song et
360 al., 2019). The remaining compounds (15-66%) of the CHONS group had less than 7 atoms,
361 implying that large amounts of CHONS compounds were assigned with reduced N (e.g., amide and
362 nitrile, and heterocyclic aromatics). The condensed aromatic compounds identified by the VK
363 diagram in the WIOC from coal combustion accounted for 22-64%, which was relatively higher
364 than those from biomass burning (14-31%), indicating a higher degree of aromaticity. This was
365 consistent with the difference observed in the AI_{mod,w} values as seen in Table S2. The AI_{mod,w} of
366 WIOC in coal emissions were 0.38 for HVB and 0.60 for MVB, and the values were higher than
367 the raw biomass (0.32-0.38) and biomass pellet (0.35-0.36), confirming a higher aromatic
368 compounds content of coal smoke sample than biomass for CHONS group.

369 Therefore, the CHO, CHON, and CHONS groups generated from coal combustion were
 370 characterized by high unsaturated levels with more aromatic species, while CHOS groups had
 371 higher aromaticity degrees in biomass smoke aerosols. Aromatic compounds might be the strong
 372 BrC chromophores contributing to light absorption (Song et al., 2019). The difference in MAE
 373 between coal and biomass emissions and its association with the chemical components will be
 374 discussed in detail in section 3.5.



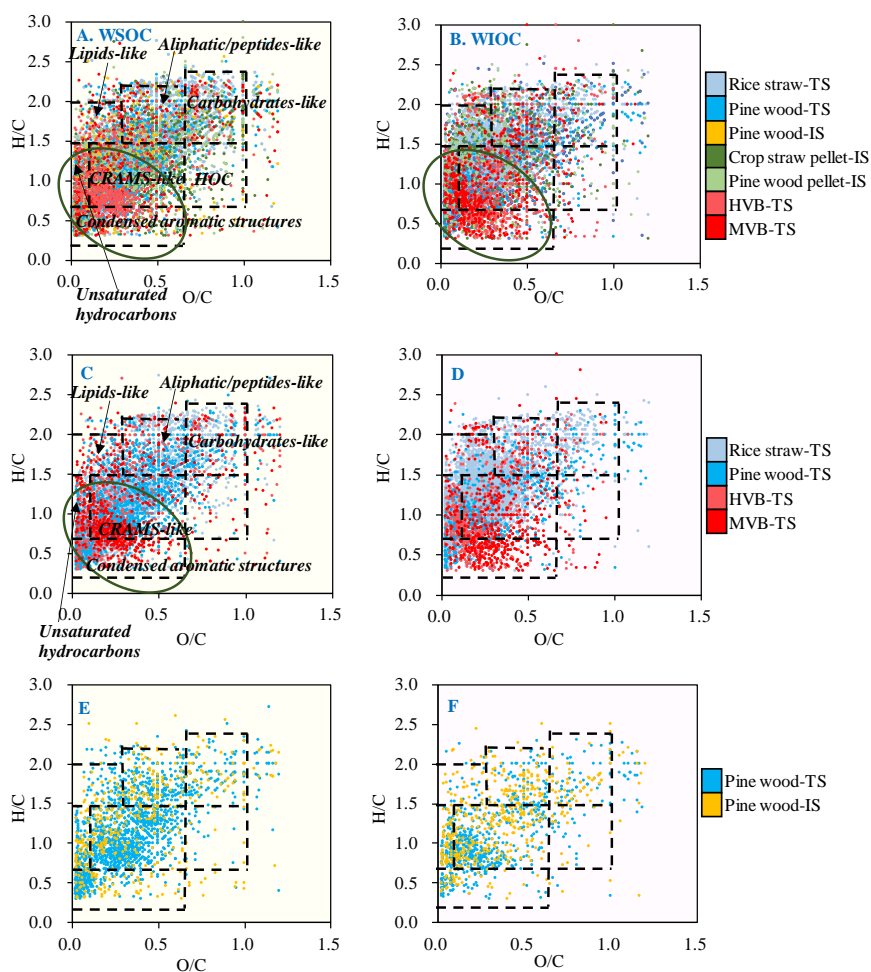
375 **Fig. 4** Each component abundance identified by VK diagram of WSOC (A: all molecules, B: CHO
 376 group, C: CHON group, D: CHOS group, E: CHONS group) and WIOC (F: all molecules, G:
 377 CHO group, H: CHON group, I: CHOS group, J: CHONS group), the classification criteria was
 378 provided in the Table S3.

379 3.4 Likely unique molecules of biomass and coal combustion

380 The unique molecules may be inferred from the Venn diagram of formulas as shown in Fig. S7.
381 Among the observed compounds, 3039 and 1624 unique molecular formulas were detected in the
382 combustion of rice straw and pine wood in the traditional stove in the WSOC fractions, which was
383 significantly higher than those found in biomass pellet-improve stove (638-696) and coal-traditional
384 stove (570-734). This suggested a notable difference between the emissions from raw biomass and
385 coal, even when using the same stove. Interestingly, fewer unique peaks were observed for pine
386 wood-improved stoves (533), indicating that using an improved stove for raw biomass could narrow
387 the difference between coal and biomass emissions. A similar trend was also found in the WIOC
388 fractions (Fig. S7). The CHONS group accounted for a significant portion of unique formulas in
389 source samples, particularly in coal smokes, representing 51-52% in WSOC and 51-69% in WIOC
390 extract. The important role of the CHONS group in unique emissions from coal combustion was
391 also noted by Tang et al., (2020), who reported that CHO and CHON were the main components of
392 unique molecular formulas of raw biomass burning emissions among the raw biomass burning, coal
393 combustion and vehicle emissions, representing 88%-93%. This fraction was higher than our result
394 of 33-77% in raw biomass and only 26-27% in biomass pellet. The distribution of the unique
395 molecules further indicated substantial discrepancies among the fuels. Unique molecules in coal
396 combustion-derived aerosols were in the region with lower H/C and O/C values compared with all
397 other samples (Fig. 5) in both WSOC and WIOC extract, indicating a higher degree of unsaturation
398 and lower level of oxidation. For example, it was observed that specific emissions from coal
399 combustion were mainly composed of condensed aromatics (32-59%), followed by CRAMs
400 compounds (23-39%) in WIOC extract. However, only 4.3-9.7% condensed aromatics of the total
401 unique emissions were observed for biomass burning, with CRAMs being the main component (39-
402 65%). As for different groups of CHO/CHON/ CHONS, the condensed aromatic compound contents

403 for all three groups in coal smoke aerosol (unique detected ones) were relatively higher than biomass.
404 While CHOS group showed a different trend in that condensed aromatic compound contents of
405 biomass smoke (unique detected ones) were relatively higher than coal smoke in WSOC extract and
406 comparable to coal smoke in WIOC extract. This finding highlighted the CHOS group's importance
407 in distinguishing the aerosols from the combustion of coals or biomass, helping conduct the source
408 apportionment of aerosols. Additionally, compared with the pine wood-improved stove emissions,
409 most unique molecules in aerosol from pine wood-traditional stove combustion were in the region
410 with lower H/C and O/C ratios (Fig. 5), which were identified mainly as CRAMs compounds. The
411 emissions from the improved stove were distributed in a wider range with fewer CRAMs
412 compounds and more aliphatic/peptides-like compounds observed. Unique molecules from biomass
413 pellets combusted in the improved stove distributed in the upper region of the VK diagram with
414 higher H/C values in WIOC extract, which indicated a lower unsaturated degree (Fig. S8).

415



416 **Fig. 5** Van Krevelen diagrams of WSOC (left) and WIOC (right) from the source samples, TS
 417 represents combustion in the traditional stove and IS represents combustion in the improved stove.
 418 Different color indicates unique formulas detected in each sample of solid fuel combustion.

419 A total of 484 molecular formulas were detected simultaneously in the seven aerosol samples
 420 in the WSOC extract and 306 in the WIOC extract. Among these commonly detected molecules,
 421 most of which were CHO compounds with the molecular numbers accounting for 60% and 73% in
 422 the WSOC and WIOC, respectively. CHON accounted for 31% of WSOC and 19% of WIOC, while
 423 SOCs only occupied about 10%. As seen in Fig. S9, these CHO compounds were mainly composed
 424 of CRAMs-like compounds and several lipids-like and aliphatic/peptides-like compounds.
 425 Moreover, these compounds were relatively small molecular compounds assigned with 8-28 C
 426 atoms and 2-12 O atoms with DBE values of 2-17 for the WSOC extract. The relatively more C

427 atom assigned with larger DBE values was observed for the CHO compounds in WIOC, which
428 could be partially explained by that the overall CHO compounds in the WIOC extract had larger
429 MW values with a high degree of unsaturation. In total, the CHON compounds were also CRAMs-
430 like compounds, and almost none compounds were expected to be aromatics. CHOS and CHONS
431 species had much fewer formulas, especially for CHONS, only accounting for 1.0% of the
432 commonly detected molecules in the WIOC. The unsaturated levels of commonly detected
433 molecules in all seven source samples were relatively low. For example, the condensed aromatic
434 compounds accounted for 6.5-9.3% and 3.3-4.1% of the total intensity for coal smoke and biomass
435 smoke in the WSOC, respectively, as well as 0.38-8.3% and 18-21% in the WIOC extract. Different
436 from the CHO, CHON, and CHONS groups, high percentages of condensed aromatic compounds
437 were found in the CHOS group (commonly detected ones) from raw biomass burning aerosols with
438 a range of 6.6-51% in the WSOC and 12-46% in the WIOC extract. These fractions were
439 significantly higher than those from coal smokes of 4.1-4.9% in the WSOC and 8.8-25% in the
440 WIOC extract. Combining the finding that the CHOS group in biomass smokes had a higher
441 aromaticity degree in both WSOC and WIOC extract. However, the unique molecules in WIOC
442 extract did not follow this trend. It was thus speculated that the higher aromaticity degree of the
443 CHOS group in biomass smokes was attributed to the intensity variation of these simultaneously
444 detected compounds, rather than the unique emission from a special source for the WIOC extract.

445 To explore the potential influence of fuel properties and combustion conditions on chemical
446 composition, the major factors such as fuel moisture, V_{daf} , and parameters reflecting combustion
447 conditions including MCE, and EC/OC ratios were assessed. The linear correlation analysis was
448 applied to estimate the effect of these factors on each molecular intensity (commonly detected ones)
449 (Fig.S10 and S11). The aliphatic compounds (including lipids-like, aliphatic/peptides-like, and
450 carbohydrates-like compounds) were negatively correlated with the fuel moisture and positively

451 correlated with V_{daf} and EC/OC ratios. Fuel moisture has been recognized as an important factor
452 influencing pollutant formation, but the influence is usually very complicated. The observed
453 relationship between fuel water content and pollutant EFs varies largely among studies, which may
454 be attributed to factors such as different water contents, pollutant types, and interactions among
455 other influencing factors. Fuels with high moisture levels may have high emissions, as extra energy
456 is needed to vaporize water during the burning process; however, a decrease in combustion
457 temperatures under very high-water content conditions may slow the pollutant formation rate and
458 consequently lower emissions. Higher EC/OC ratios and larger MCE values tend to be associated
459 with stronger flaming conditions. The results suggested that the aliphatic compounds were apt to be
460 produced during the period of the flaming phase with higher combustion temperature. This result
461 could partially explain that aliphatic/peptides-like compounds would be apt to be produced in the
462 improved stove rather than the traditional stove. In comparison, the emissions of CRAMs-like
463 compounds which are the most abundant species were decreased with the increased MCE and
464 EC/OC values, indicating that CARMs-like compounds were generated under the smoking phase.
465 No significant correlations of aromatics, unsaturated hydrocarbons, and HOCs were observed with
466 these parameters, resulting from their small proportion in commonly detected molecules. It is worth
467 noting that there may be significant differences, even for the same fuel, in chemical composition
468 depending on other factors such as stove type and combustion conditions. The interactions among
469 these factors make it difficult to assess their influence. It was found that only around 50% of
470 identified molecules were overlapped emissions for pine wood combusted in the traditional stove
471 and in the improved stove, suggesting the importance of the stove used. A smaller fraction of 25-
472 30% molecules were observed compared to the pine wood emissions with coals combusted in the
473 same stove, which suggested that the influence of varied fuels on chemical composition could
474 surpass the differences caused by different stoves. Our previous studies have revealed that fuel type

475 was the most important factor influencing the MAE values (Zhang et al., 2021b) and BrC EFs
476 (Zhang et al., 2021a). The present study highlighted the dominant effect of fuel type on the chemical
477 composition of soluble OC, providing a theoretical basis for source appointment based on molecular
478 composition. Moreover, the combustion conditions would have a significant effect on the molecular
479 intensity, resulting in differences in MAE eventually as indicated in section 3.1.

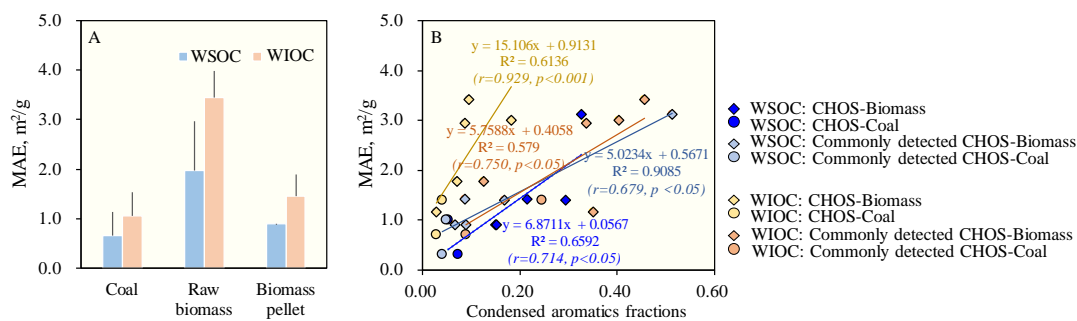
480 3.5 The correlation of light absorption properties with molecular compositions

481 Here we specifically investigate the variability in optical characteristics among the different
482 fuels attributed to their chemical compositions. No significant correlations were observed between
483 the AAE values of soluble OC and the molecular composition, indicating that AAE could be
484 influenced by many factors. The MAE may be influenced by the degree of oxidation and
485 unsaturation degree (Mo et al., 2017) (Tang et al., 2020), however, there was no significant
486 correlation found between the MAE values and O/C values in the present source samples, implying
487 that the BrC light absorption ability might not be directly affected by its oxidation degree.

488 The MAE_{365, WSOC} values were significantly positively correlated with the DBE values ($r=0.786$,
489 $p<0.05$) and the MW values ($r=0.750$, $p<0.05$) (Fig. S12), indicating that unsaturation and MW
490 played a crucial role in the light absorption capability of the source samples. In the above discussion,
491 we have noticed that the CHOS group in biomass was characterized by a higher degree of
492 aromaticity than coal smoke aerosol, while the CHO, CHON, and CHONS groups have a higher
493 aromatic degree in the coal emissions. A significantly positive correlation ($r=0.714$, $p<0.05$) was
494 observed between MAE_{365, WSOC} values, and the condensed aromatics percentages from the CHOS
495 group (Fig. 6), indicating that aromatics in the CHOS group contributed to the high light absorption
496 ability of biomass smokes.

497 For the WIOC extract, no significant correlation was found between the MAE_{365, WIOC} values

498 with MW or DBE values, which might be explained by much different chemical composition for
 499 insoluble compounds compared with soluble parts Although the coal combustion emitted much
 500 more aromatic compounds with higher DBE values for the WIOC, the MAE values were not
 501 significantly higher. The significantly positive correlation ($r=0.929, p<0.05$) between $MAE_{365, WIOC}$
 502 values, and the CHOS condensed aromatics percentages confirmed the importance of CHOS
 503 aromatics in determining the light absorption capability from source samples. As mentioned in
 504 Section 3.4, the higher aromaticity degree of the CHOS group in biomass smoke was largely due to
 505 the intensity variation of these commonly detected compounds in all source samples. Further
 506 analysis revealed that $MAE_{365, WIOC}$ values were positively correlated ($r=0.750, p<0.05$) with the
 507 CHOS condensed aromatics fractions which were the fraction simultaneously detected in all
 508 samples. These results indicated that the light absorption capability of source aerosols may be due
 509 to the higher abundance of some CHOS aromatic compounds commonly emitted from both coal
 510 and biomass, rather than the unique tracers.



511 **Fig. 6** MAE values at $\lambda=365$ nm (A) and correlations between MAE values and condensed aromatics
 512 fractions in CHOS and in commonly detected CHOS (B) from the source samples

513 4. Conclusion and implication

514 The $MAE_{365, WSOC}$ ranged from 0.21 to 3.1 m²/g with an average of 1.3 ± 0.7 m²/g. The $MAE_{365, WIOC}$
 515 was found to be higher with an average of 2.0 ± 1.3 m²/g. There were significant differences
 516 ($p<0.05$) observed among the different fuels for both $MAE_{365, WSOC}$ and $MAE_{365, WIOC}$, as raw

517 biomass burning combustion had significantly higher values than the coal combustion. The
518 AAE_{WSOC} ranged from 3.8 to 11 with an average of 6.9 ± 1.5 . The AAE_{WIOC} was slightly lower than
519 AAE_{WSOC} , which ranged from 4.2 to 6.5 with an average of 5.5 ± 0.5 . Thousands of peaks were
520 identified in the m/z range of 150-800, indicating a complex chemical composition of aerosols from
521 residential sources. CHO group was the most abundant component in the WSOC extracts and the
522 contribution of CHO compounds to the total intensity in aerosols from raw biomass burning was
523 significantly higher than those from biomass pellets and coal smoke. On the other hand, WIOC
524 extract contained more SOCs, especially in the coal combustion aerosol. Notably, CHON
525 compounds were more abundant in the coal combustion emissions, which was due to the higher fuel
526 N content of coals ($r=0.936$, $p<0.05$). The SOCs emissions were more predominant during flaming
527 phases, with a positive correlation between SOCs abundance with the MCE values ($r=0.750$,
528 $p<0.05$). The CHO, CHON, and CHONS groups generated from coal combustion were
529 characterized by high unsaturated levels with more aromatic species, while CHOS groups had
530 higher aromaticity degrees in biomass smoke aerosols. It was found that MAE values were
531 positively correlated with the CHOS condensed aromatics proportion for both WSOC ($r=0.714$,
532 $p<0.05$) and WIOC extract ($r=0.929$, $p<0.001$). These results indicated that higher CHOS condense
533 aromatics abundance in biomass burning aerosols could partly explain the higher MAE values of
534 raw biomass smoke. Further analysis showed a positive correlation of MAE with the CHOS
535 condensed aromatics fractions which were the fractions simultaneously detected in all samples.
536 These results indicated that the light absorption capability of source aerosols may be due to the
537 higher abundance of some CHOS aromatic compounds commonly emitted from both coal and
538 biomass burning, rather than the unique tracers. The unique formulas of coal combustion aerosols
539 were in the lower H/C and O/C regions with higher unsaturated compounds in the VK diagram. This
540 work is potentially applicable to the source appointment based on the molecular characteristics and

541 to future studies developing more scientific control measures by focusing on one major component
542 (e.g., CHOS condensed aromatics) of light absorption aerosols.

543 There are still some questions that need to be investigated in the future study. First, the study on
544 BrC composition in this research used dissolved OC as a substitute. However, this substitute cannot
545 fully represent BrC emissions. Both WSOC and WIOC contain some non-light-absorbing
546 components, and the proportion of these components is unknown, making it difficult to measure the
547 representativeness of extractable OC for BrC. Additionally, the lack of study about the emissions
548 characteristics under controlled combustion conditions limits the obtained results. Controlled
549 experiments including flaming or smoldering burns, airflow, and combustion temperature are
550 needed in future work. Third, fresh burning-derived OC released into the atmosphere can undergo
551 various aging reactions such as photochemical degradation. These reactions can significantly alter
552 the light absorptivity and chemical properties of BrC components. It is essential to consider the
553 optical properties and lifetimes of organic compounds emitted from solid fuel combustion in climate
554 models.

555 **Data availability.**

556 Data are available by contacting the corresponding authors

557 **Supplement**

558 The following information is in the appendix and available via the Internet:

559 Data processing in the ESI FT-ICR MS; fuel properties of coal and biomass fuels tested;
560 number of formulas in each group, values of elemental ratios, MW, and DBE values in the WSOC
561 and WISCO for each source type; Stoichiometric ranges of VK classes; Correlation between fuel N
562 and emission factors of NO_x; the VK diagrams of WSOC and WIOC for different source samples;

563 and correlations between the VK plots of WSOC/WIOC and fuel properties or combustion
564 conditions.

565 **Author contributions**

566 LZ, GS, and ST designed the experiment. LZ and JL prepared the filters used. LZ, YL, XL and
567 ZL conducted the sample collection. LZ and JL performed the data analysis. LZ wrote the paper.
568 GS, and ST reviewed and commented on the paper.

569 **Competing interests**

570 The authors declare that they have no conflict of interest.

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