We are thankful to the reviewer for his/her thoughtful and constructive comments that help improve the manuscript substantially. We have revised the manuscript accordingly. Listed below is our pointto-point response in blue to each comment that was offered by the reviewer.

#### **Response to Reviewer #1**

This study used different types of instruments with mass spectra, combining with a DMA and AAC for size selection and investigated the effective density of aerosols, during the Beijing 2022 Olympic Winter Games for the impacts of emission controls on particle mixing state and density. The results provide information on changes in aerosol compositions and mixing state due to emission control. A few points need to be addressed before it can be accepted.

We thank the reviewer's comments and have revised the manuscript accordingly.

#### Major:

1) The method in deriving the effective density for different compositions should be given more details. The used equations have not been clearly explained. It is not clear how you have linked the effective density with certain composition.

Thank the reviewer's comments. SPAMS can provide information on the mixing state and particle size, i.e., the vacuum aerodynamic diameter  $(D_{va})$ , of individual particles, while DMA and AAC can select particles with specific mobility diameters  $(D_m)$  or aerodynamic diameters  $(D_a)$ , respectively. The relationship between the three diameters has been given by Decarlo et al. (2004) and the effective density  $(\rho_{eff})$  can be calculated if either two of them are known. For example, when the  $D_{va}$  and  $D_m$  of the particle are known, the  $\rho_{eff}$  can be calculated as:

$$\rho_{eff} = \frac{D_{va}}{D_m} \,\rho_0 \tag{1}$$

It can be used to calculate the  $\rho_{eff}$  of particles captured by the DMA-SPAMS tandem system. Where  $\rho_{\theta}$  is the standard density (1.0 g cm<sup>-3</sup>). Another approach to define the  $\rho_{eff}$  that can be adopted in the AAC-SPAMS tandem system is based on the ratio of particle density ( $\rho_p$ ) and the particle dynamic shape factor ( $\chi_{\gamma}$ ) as follows:

$$\rho_{eff} = \frac{\rho_p}{\chi_{\gamma}} = \frac{D_{va}}{D_{ve}\rho_0} \tag{2}$$

where  $D_{ve}$  represents the volume equivalent diameter. The method of deriving the  $\rho_{eff}$  of particles with the support of  $D_{ve}$  and  $D_{va}$  has been verified in detail in previous studies (Peng et al., 2021; Su et al., 2021). The relationship between the  $D_a$ ,  $D_{va}$  and  $D_{ve}$  can be stated by the following equation:

$$D_a = D_{\nu e} \sqrt{\frac{\rho_p C_c(D_{\nu e})}{\chi_t \rho_0 C_c(D_a)}}$$
(3)

where  $\chi_t$  represents the aerosol dynamic shape factor in the transition regime. Considering the

approximation between  $\chi_t$  and  $\chi_{\gamma}$ , the  $D_{ve}$  can be calculated by combining Eqs. (2) and (3) as follows:

$$C_c(D_a)\frac{D_a^2}{D_{va}} = D_{ve}C_c(D_{ve})$$
(4)

That is, when the aerosol instruments in tandem are same (DMA-SPAMS or AAC-SPAMS), the derivation of  $\rho_{eff}$  of particles with different compositions is uniform and its confidence has been confirmed in previous studies (Su et al., 2021; Spencer et al., 2007; Peng et al., 2021). Based on the diameter values set by DMA or AAC, combined with the  $D_{va}$  and chemical compositions of the particles provided by SPAMS, it is possible to associate the  $\rho_{eff}$  of individual particles with their chemical compositions. Moreover, we have made additional explanations in lines 123–124 of the manuscript and given detailed calculation methods of  $\rho_{eff}$  in the supplementary according to the recommendations.

2) The instrument setup should be given in front in the main texts, with more explanation why running AAC and DMA in parallel. Why the density has been derived using two methods.

Thank the reviewer's comments. In fact, we initially planned to connect DMA and AAC in series with SPAMS at different periods and select particles with  $D_m$  and  $D_a$  in the range of 150–300 nm and 200–700 nm, respectively, to finally obtain two complete datasets. However, only the SPAMS data with  $D_a = 300$  nm were eventually credible in the AAC-SPAMS period, accounting for 13.3% of the total particles captured by SPAMS (322415 of 2416964). This is due to the unstable sheath flow of AAC when selecting particles in the size range of 400–700 nm, and only 1756 particles were captured at  $D_a = 200$  nm due to the SPAMS detection limit. We therefore decided to use the DMA-SPAMS in combination with the AAC-SPAMS dataset, which covers the Olympic Winter Games completely and makes it possible to analyze the changes in the mixing state and  $\rho_{eff}$  of particles under emission control. Considering that DMA and AAC screen particles based on different diameters, it is necessary to calculate the  $\rho_{eff}$  by two methods separately. We have added the experimental system schematic (Fig. 1) to the manuscript as suggested and made additional explanations about the calculation of  $\rho_{eff}$  in lines 114-116.

# 3) The effective density from 1.26 to 1.20 is not significantly different, as emphasized in the abstract.

Thank the reviewer for pointing this out. The 1.20 and 1.26 g cm<sup>-3</sup> in the abstract correspond to the average  $\rho_{eff}$  of particles for the entire period and for the OWG, respectively, which are indeed similar. In order to emphasize the change in particle aging due to reduced emissions, it may be more convincing to compare the  $\rho_{eff}$  of particles during the OWG with the nOWG (1.26 vs. 1.15 g cm<sup>-3</sup>). Actually, the range of average  $\rho_{eff}$  for all particles in this study is wide (0.76–1.68 g cm<sup>-3</sup>) and is influenced by the chemical composition and atmospheric processes. Meanwhile, the  $\rho_{eff}$  ranges for different classes of particles (from 0.36 g cm<sup>-3</sup> for pure-EC to 1.62 g cm<sup>-3</sup> for KAECOC-NS) are also comparable to the results of previous studies in Guangzhou (0.87–1.51 g cm<sup>-3</sup>) and California (0.27–1.48 g cm<sup>-3</sup>) (Spencer et al., 2007; Zhang et al., 2016). The relatively small difference in  $\rho_{eff}$  between the OWG and nOWG may be attributed to the fact that primary emitted KECOC-NS (1.31 vs. 1.30 g cm<sup>-3</sup>) and Biomass-K (1.10 vs. 1.08 g cm<sup>-3</sup>) particles, which contributed up to 36.06% of the total particles, had very little change in  $\rho_{eff}$  between the two periods (Fig. S6). In addition, the

large number of KAECOC-NS particles with high  $\rho_{eff}$  captured during snowfalls on 22 January, 24 January, and 30 January also led to small difference in  $\rho_{eff}$  between the OWG and nOWG periods (for KAECOC-NS, OWG vs nOWG: 1.36 vs 1.71 g cm<sup>-3</sup>). We have added Fig. S6 to the supplementary in order to emphasize the difference in  $\rho_{eff}$  between the OWG and nOWG periods. Moreover, additional explanation on  $\rho_{eff}$  has been added in lines 234-241, while lines 23-24 of the abstract have been revised.



Figure S6: Average effective density of different classes of particles during the OWG and nOWG periods.

4) Why high molecular weight OA has a lower effective density.

As shown in Fig. R1a, the average mass spectra of high-molecular-weight organic matter (HOM) exhibits distinct fragments of polycyclic aromatic hydrocarbons (PAH) such as 152  $[C_{12}H_8]^+$ , 165  $[C_{13}H_9]^+$ , 178  $[C_{14}H_{10}]^+$ , and 189  $[C_{15}H_9]^+$ . Previous studies based on single particles have demonstrated that particles with significant fragments of PAHs in urban areas are mostly associated with vehicle emissions and coal combustion (Sodeman et al., 2005; Su et al., 2021; Giorio et al., 2015; Hu et al., 2021; Wang et al., 2020; Zhang et al., 2022). The lower PA<sub>sulfate</sub>/PA<sub>nitrate</sub> ratio (0.24) indicates that HOM belongs to relatively fresh particles that have not been sufficiently aged. Meanwhile, the peak of the daily cycle of HOM at 8:00 and the high values at night (Fig. R1b) coincide with the traffic emission during the morning rush hours and coal combustion at night, supporting the conclusion that HOM particles are comparatively fresh.

Effective density ( $\rho_{eff}$ ), as a parameter often derived from a combination of two aerosol measurements, can be defined as the ratio of the particle density ( $\rho_p$ ) to the bulk material density ( $\rho_m$ ) (Hand et al., 2002; Mcmurry et al., 2002).  $\rho_p$  is less than  $\rho_m$  when the particles contain voids inside, with the  $\rho_{eff}$  of particles less than 1 (Decarlo et al., 2004). It is recognized that fresh particles emitted by combustion usually show loose structure with irregular shape (Liu et al., 2019; Spencer et al., 2007; China et al., 2014). That is, HOM particles emitted from fossil fuel combustion will have lower  $\rho_{eff}$  due to their loose structure, which is consistent with our results (0.87 g cm<sup>-3</sup> on average). In addition to particle morphological characteristics, the  $\rho_{eff}$  of HOM is also affected by factors including chemical composition and aging process (Katrib et al., 2005; Pagels et al., 2009). As HOM emitted into the atmosphere undergoes the aging process and mixes with more sulfate and nitrate, they become more compact and the  $\rho_{eff}$  peaks at 14:00 (0.96 g cm<sup>-3</sup>). Thanks to the reviewers

for pointing this out, and we have provided additional explanation in lines 231-232.



Figure R1: Average mass spectra of single particles (a) and diurnal cycle of normalized counts (b) for HOM.

5) What can the findings tell from environmental policy point of view?

Thank the reviewer's comments. This study investigated the effects of emission control measures on atmospheric particles during major events by comparing the mass concentration, chemical composition, and effective density of particles during the Olympic Winter Games with other periods. The results indicate that significant improvements in air quality can be achieved in the short term by closing high-emission plants and limiting high-emission vehicles. Control measures during the OWG resulted in a 48.7% and 37.5% decrease in mass concentrations of NR-PM1 and eBC particles, respectively. It is worth noting that the short-term approaches to improving air quality also provide new insights into the development of environmental policy. Due to the implementation of desulphurization, accompanied by increasing NOx emissions from vehicles and industry, nitrates became the most dominant pollutant in urban areas, accounting for 36.1% of NR-PM<sub>1</sub> during this campaign. Therefore, environmental policies need to be formulated not only in terms of its longterm viability, but also with a focus on the emission and control of NO<sub>x</sub>. Utilizing clean energy instead of traditional fossil fuels as much as possible in the production of factories, and further raising the emission standards for factory waste and encouraging the purchase and use of waste treatment equipment. In addition, highly polluting and emitting fuel vehicles can be eliminated gradually while new energy vehicles are actively promoted.

# **Others:**

Please indicate the full name where the abbreviation first appears in the abstract, such as EC and OC on line 17, EC-NC and KEC-N in line 19, and ECOC-NC in line 20.

Thank the reviewer's comments. We have revised the manuscript in lines 17-21 as suggested.

Please give the full name of NR-PM1 when it appears for the first time.

Changed in lines 101-102 as suggested.

Line105-114, Why to use DMA connecting with SPAMS and AAC to connect with SPAMS separately to obtain effective particle size? Why are two different instruments required?

Thank the reviewer's comments. As mentioned in main question 2, we had originally planned to connect DMA and AAC in series with SPAMS at different periods and select particles with  $D_m$  and  $D_a$  in the range of 150–300 nm and 200–700 nm, respectively, to finally obtain two complete datasets. Unfortunately, only the SPAMS data with Da = 300 nm were eventually credible in the AAC-SPAMS period, accounting for 13.3% of the total particles captured by SPAMS (322415 of 2416964). Therefore, in order to make the conclusions of this study more convincing, we decided to use the DMA-SPAMS in combination with the AAC-SPAMS dataset, which fully covers the Olympic Winter Games and makes it possible to analyze the changes in the mixing state and effective density of particles under emission control. Considering that the  $\rho_{eff}$  is usually defined by the combination of two aerosol size measurements (Hand et al., 2002; Mcmurry et al., 2002), two formulas, i.e., Eqs. (1) and (2) above, need to be used in the calculation of  $\rho_{eff}$ .

## Could you provide a detailed explanation how Equation 2 is derived?

Thanks to the reviewer's suggestion, we have provided more detailed derivations both below and in Section 1.2 of the supplementary.

According to the definition given by Hand and Kreidenweis (2002), the effective density of particles is equal to the ratio of the particle density ( $\rho_p$ ) to the dynamic shape factor ( $\chi_\gamma$ ).

$$\rho_{eff} = \frac{\rho_p}{\chi_{\gamma}} \tag{5}$$

The calculation of  $D_{va}$  obtained by Jimenez (2003) is shown in Eq. (6).

$$D_{va} = \frac{\rho_p D_{ve}}{\rho_0 \chi_{\gamma}} \tag{6}$$

Combining Eqs. (5) and (6) to obtain the following formula for  $\rho_{eff}$ , which is mentioned by the reviewer:

$$\rho_{eff} = \frac{D_{va}}{D_{ve}\rho_0} \tag{7}$$

Besides, the relationship between the  $D_a$ ,  $D_{va}$  and  $D_{ve}$  can be stated by Eq. (8):

$$D_a = D_{ve} \sqrt{\frac{\rho_p C_c(D_{ve})}{\chi_t \rho_0 C_c(D_a)}}$$
(8)

where  $\chi_t$  represents the aerosol dynamic shape factor in the transition regime. Considering the approximation between  $\chi_t$  and  $\chi_y$ , Eqs. (6) and (8) can be combined and calculated as follows:

$$C_c(D_a)\frac{D_a^2}{D_{va}} = D_{ve}C_c(D_{ve})$$
<sup>(9)</sup>

 $C_c(D)$  is the Cunningham slip correction factor, which can be calculated by the following equation:

$$Cc(D) = 1 + \frac{\lambda}{D} (A + B \cdot exp(\frac{C \cdot D}{\lambda}))$$
(10)

where  $\lambda$  represents the mean free path of the gas molecules. A, B and C are empirically determined constants specific to the analyzed system, where A is 2.33, B is 0.966 and C is -0.498. Substituting Eq. (10) into Eq. (9) obtains Eq. (11):

$$\frac{D_a^2}{D_{va}} + \frac{D_a \cdot \lambda}{D_{va}} \left( A + B \cdot exp\left(\frac{C \cdot D_a}{\lambda}\right) \right) = D_{ve} + \lambda \left( A + B \cdot exp\left(\frac{C \cdot D_{ve}}{\lambda}\right) \right)$$
(11)

The  $D_a$  and  $D_{va}$  are known in the AAC-SPAMS tandem system, which can be brought into Eq. (11) to obtain  $D_{ve}$ . Finally, the  $\rho_{eff}$  of particles can be derived from Eq. (7).

Line 212-214, Please indicate the corresponding figure number for the conclusive numerical results provided by the authors.

Added as suggested.

Please specify what the color bar in Figure S5 stands for.

Thanks to the reviewer for pointing this out, an explanation has been added in the figure legend of Fig. S5.

Section 3.3 is quite confused. The title indicates that the study focused on the effective density of aerosols during the Olympic Winter Games. However, the effective density is only briefly mentioned but not related to the event.

Thanks to the reviewer for pointing this out. We further compare the average  $\rho_{eff}$  of different classes of particles during the OWG and nOWG periods as shown in Fig. S6. The results show that most of the particles have higher  $\rho_{eff}$  during the OWG, with the most pronounced changes in EC-NS (1.10 vs. 0.99 g cm<sup>-3</sup>) and ECOC-NS (1.22 vs. 1.15 g cm<sup>-3</sup>). In contrast, the  $\rho_{eff}$  of fresh pure-EC (0.36 vs. 0.36 g cm<sup>-3</sup>), KECOC-NS (1.31 vs. 1.30 g cm<sup>-3</sup>) and KOC-N (1.03 vs. 1.00 g cm<sup>-3</sup>) particles from primary emission did not change significantly in both periods. The KAECOC-NS particles with significantly high  $\rho_{eff}$  during the nOWG period were affected by snowfall, and their  $\rho_{eff}$  increased from 1.32 to 1.73 g cm<sup>-3</sup> as RH increased from 10% to 80% (Fig. 9j). We have added Fig. S6 as suggested, with additional description in lines 234-241 of the manuscript.

As this study mainly focused on the BC-containing particles, a few quite related studies also measured the shape of BC-containing particles in Beijing, which showed more spherical particles when polluted (Hu et al., EST Letters, 2022, 10.1021/ acs.estlett.2c00060), and also the

aerodynamic size-selected compositions and density by AAC (Yu et al., ACP, 2022, 10.5194/acp-22-4375-2022). These studies could be referenced to support some of your conclusions.

Thanks to the reviewer's suggestion. The literatures were cited in the revised manuscript.

## 2 is not used in the texts.

Thank the reviewer's comments. The original Fig. 2 is mentioned in lines 127-131 of the manuscript in order to depict the meteorological elements, pollutant concentrations, and the particle counts captured by SPAMS throughout the observation period. The original Fig. S2 is referenced in lines 101-103 of the manuscript to illustrate the representativeness of the SPAMS measurements by comparing them to the AMS and AE33 measurements. The results of the comparison of pollutant concentrations between Winter Olympic and non-Winter Olympic periods given in Table 2 are also mentioned in lines 128-130. The characteristic peak information provided in Table S2, which is essential for particle classification, is mentioned in line 112.

### Please explain the many significantly high values of PAsulfate/PAnitrate in Figure 6.

Thank the reviewer's comments. We selected  $80[SO_3]^-$  and  $97[HSO_4]^-$  as characteristic peaks for sulfate and  $46[NO_2]^-$  and  $62[NO_3]^-$  for nitrate in the calculation of PA<sub>sulfate</sub>/PA<sub>nitrate</sub>. Previous data processing results showed that the average PA<sub>sulfate</sub>/PA<sub>nitrate</sub> values for particles including K-Amine-NS, K-N, and rich-Fe were significantly higher than the 75th percentile. Checks revealed that some of the abnormally high values of PA<sub>sulfate</sub>/PA<sub>nitrate</sub> were not removed during previous data processing, so we further processed the data and redrew the graph (Fig. 6). However, the average PA<sub>sulfate</sub>/PA<sub>nitrate</sub> values of some particles are still high, with K-Amine-NS and rich-Fe being the most obvious. This is due to the fact that local pollutants are removed while particles mixed with sulfate are transported during high wind speed periods, resulting in high PA<sub>sulfate</sub>/PA<sub>nitrate</sub> values (Fig. R3). For example, the average PA<sub>sulfate</sub>/PA<sub>nitrate</sub> for K-Amine-NS and rich-Fe for the entire period were 2.04 and 0.13, respectively, but when the wind speed was higher than 6 m s<sup>-1</sup> (11.4% of the entire period), they were 5.37 and 0.32, respectively (Figs. R2a and R2b). Such high PA<sub>sulfate</sub>/PA<sub>nitrate</sub> values at high wind speeds and low pollutant concentrations lead to the average PA<sub>sulfate</sub>/PA<sub>nitrate</sub> values in the original Fig. 6 being significantly higher than the median values.

While the differences in  $PA_{sulfate}/PA_{nitrate}$  values between different classes of particles are mainly related to their mixing state. Overall, the captured particles were predominantly mixed with nitrate, with an average  $PA_{sulfate}/PA_{nitrate}$  of only 0.25 during the observation period, with average  $PA_{sulfate}/PA_{nitrate}$  values of 0.13, 0.10, 0.12, 0.10 and 0.15 for KEC-N, KNaEC-N, KOC-N, Biomass-K and Total-SIA, respectively.



Figure 6: Peak area ratios of (a-d) sulfate (m/z -80 and -97) to nitrate (m/z -46 and -62) for each type of particles and (e-j) elemental carbon (m/z  $C_n^{\pm}$ , n = 1-5) to organic carbon (m/z 27, 29, 37 and 43) in ECOC-containing particles during OWG and nOWG. Also shown are median (horizontal lines), mean (circles), 25th and 75th percentiles (lower and upper boxes), and 10th and 90th percentiles (lower and upper whiskers).



Figure R2: Time series of PA<sub>sulfate</sub>/PA<sub>nitrate</sub> for (a) K-Amine-NS, (b) rich-Fe, (c) K-N, (d) Total-EC, (e) Total-ECOC, (f) HOM, (g) Biomass-K and (h) Total-OC, and (i) wind speed (WS) colored by wind direction (WD) as well as mass concentration of NR-PM<sub>1</sub>.



Figure R3: Scatter plot of PA<sub>sulfate</sub>/PA<sub>nitrate</sub> values for all captured particles versus the mass

concentration of NR-PM1 and eBC measured by AMS, with scatters colored by wind speed.

### References

- China, S., Salvadori, N., and Mazzoleni, C.: Effect of traffic and driving characteristics on morphology of atmospheric soot particles at freeway on-ramps, Environ. Sci. Technol., 48, 3128-3135, doi:10.1021/es405178n, 2014.
- DeCarlo, P. F., Slowik, J. G., Worsnop, D. R., Davidovits, P., and Jimenez, J. L.: Particle Morphology and Density Characterization by Combined Mobility and Aerodynamic Diameter Measurements. Part 1: Theory, Aerosol Sci. Tech., 38, 1185-1205, doi:10.1080/027868290903907, 2004.
- Giorio, C., Tapparo, A., Dall'Osto, M., Beddows, D. C., Esser-Gietl, J. K., Healy, R. M., and Harrison, R. M.: Local and regional components of aerosol in a heavily trafficked street canyon in central London derived from PMF and cluster analysis of single-particle ATOFMS spectra, Environ. Sci. Technol., 49, 3330-3340, doi:10.1021/es506249z, 2015.
- Hand, J. L. and Kreidenweis, S. M.: A New Method for Retrieving Particle Refractive Index and Effective Density from Aerosol Size Distribution Data, Aerosol Sci. Tech., 36, 1012-1026, doi:10.1080/02786820290092276, 2002.
- Hand, J. L., Kreidenweis, S. M., Kreisberg, N., Hering, S., Stolzenburg, M., Dick, W., and McMurry, P. H.: Comparisons of Aerosol Properties Measured by Impactors and Light Scattering from Individual Particles: Refractive Index, Number and V olume Concentrations, and Size Distributions, Atmos. Environ., 36, 1853–1861, doi:10.1016/S1352-2310(02)00103-6, 2002.
- Hu, J., Xie, C., Xu, L., Qi, X., Zhu, S., Zhu, H., Dong, J., Cheng, P., and Zhou, Z.: Direct Analysis of Soil Composition for Source Apportionment by Laser Ablation Single-Particle Aerosol Mass Spectrometry, Environ. Sci. Technol., 55, 9721-9729, doi:10.1021/acs.est.0c07983, 2021.
- Jayne, J. T., Leard, D. C., Zhang, X., Davidovits, P., Smith, K. A., Kolb, C. E., and Worsnop, D. R.: Development of an aerosol mass spectrometer for size and composition analysis of submicron particles, Aerosol Sci. Technol., 33, 49–70, 2000.
- Jimenez, J. L., Jayne, J. T., Shi, Q., Kolb, C. E., Worsnop, D. R., Yourshaw, I., Seinfeld, J. H., Flagan, R. C., Zhang, X., Smith, K. A., Morris, J. W., and Davidovits, P.: Ambient aerosol sampling using the Aerodyne Aerosol Mass Spectrometer, J. Geophys. Res., 108, 8425, doi:10.1029/2001jd001213, 2003.
- Katrib, Y., Martin, S. T., Rudich, Y., Davidovits, P., Jayne, J. T., and Worsnop, D. R.: Density changes of aerosol particles as a result of chemical reaction, Atmos. Chem. Phys., 5, 275–291, doi:10.5194/acp-5-275-2005, 2005.
- Liu, H., Pan, X., Wu, Y., Wang, D., Tian, Y., Liu, X., Lei, L., Sun, Y., Fu, P., and Wang, Z.: Effective densities of soot particles and their relationships with the mixing state at an urban site in the Beijing megacity in the winter of 2018, Atmos. Chem. Phys., 19, 14791-14804, doi:10.5194/acp-19-14791-2019, 2019.
- McMurry, P. H., Wang, X., Park, K., and Ehara, K.: The Relationship between Mass and Mobility for Atmospheric Particles: A New Technique for Measuring Particle Density, Aerosol Sci. Technol., 36, 227-238, doi:10.1080/027868202753504083, 2002.
- Pagels, J., Khalizov, A. F., McMurry, P. H., and Zhang, R. Y.: Processing of Soot by Controlled Sulphuric Acid and Water Condensation—Mass and Mobility Relationship, Aerosol Sci. Technol., 43, 629-640, doi:10.1080/02786820902810685, 2009.
- Peng, L., Li, L., Zhang, G., Du, X., Wang, X., Peng, P. a., Sheng, G., and Bi, X.: Technical note: Measurement of chemically resolved volume equivalent diameter and effective density of particles by AAC-SPAMS, Atmos. Chem. Phys., 21, 5605-5613, doi:10.5194/acp-21-5605-2021, 2021.
- Sodeman, D. A., Toner, S. M., and Prather, K. A.: Determination of Single Particle Mass Spectral Signatures from Light-Duty Vehicle Emissions, Environ. Sci. Technol. , 39, 4569-4580, doi:10.1021/es0489947 2005.
- Spencer, M. T., Shields, L. G., and Prather, K. A.: Simultaneous Measurement of the Effective Density and Chemical Composition of Ambient Aerosol Particles, Environ. Sci. Technol., 41, 1303-1309, doi:10.1021/es061425+ 2007.
- Su, B., Zhang, G., Zhuo, Z., Xie, Q., Du, X., Fu, Y., Wu, S., Huang, F., Bi, X., Li, X., Li, L., and Zhou, Z.: Different characteristics of individual particles from light-duty diesel vehicle at the launching and idling state by AAC-SPAMS, J. Hazard. Mater., 418, 126304, doi:10.1016/j.jhazmat.2021.126304, 2021.

- Wang, X., Ye, X., Chen, J., Wang, X., Yang, X., Fu, T.-M., Zhu, L., and Liu, C.: Direct links between hygroscopicity and mixing state of ambient aerosols: estimating particle hygroscopicity from their single-particle mass spectra, Atmos. Chem. Phys., 20, 6273-6290, doi:10.5194/acp-20-6273-2020, 2020.
- Zhang, G., Bi, X., Qiu, N., Han, B., Lin, Q., Peng, L., Chen, D., Wang, X., Peng, P. a., Sheng, G., and Zhou, Z.: The real part of the refractive indices and effective densities for chemically segregated ambient aerosols in Guangzhou measured by a single-particle aerosol mass spectrometer, Atmospheric Chemistry and Physics, 16, 2631-2640, doi:10.5194/acp-16-2631-2016, 2016.
- Zhang, Y., Pei, C., Zhang, J., Cheng, C., Lian, X., Chen, M., Huang, B., Fu, Z., Zhou, Z., and Li, M.: Detection of polycyclic aromatic hydrocarbons using a high performance-single particle aerosol mass spectrometer, J. Environ. Sci., 124, 806-822, doi:10.1016/j.jes.2022.02.003, 2022.