

Why Did Ozone Concentrations Remain HighIncrease During Shanghai's Static Management? A Statistical and Radical Chemistry Perspective

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Abstract

During the period of April and May 2022, Shanghai implemented city-wide static management measures to control the spread of the Omicron variant. Compared to the lockdown in early 2020, the static management in 2022 occurred during the high-ozone season and lasted for a longer duration. It can be considered as a “large-scale field experiment” to study the response of ambient ozone levels to emission reductions. During this period, we conducted comprehensive observations at Fudan University Jiangwan Campus in the northeast corner of Shanghai. Similar experiments were also conducted during the same period in 2020 and 2021. Despite the significant reduction of approximately 30% in VOCs and around 50% in NO₂ due to static management in 2022, the average ozone level increased by nearly 23%, compared to 2020 and 2021. This suggests that the reduction in ozone precursors and other pollutants did not lead to a corresponding decrease in ozone concentrations as expected. Cluster analysis of diurnal patterns of ozone concentration revealed four distinct types of diurnal ozone variations. Cluster 3 and Cluster 4, with high ozone levels, experienced significant increases in their share during static management, ultimately leading to an overall increase in average ozone levels in 2022. According to the Observation-Based Model (OBM) simulation analysis, the average peak concentrations of OH, HO₂, and RO₂ in 2022 were estimated to be 5.13×10^6 , 4.79×10^8 , and 2.6×10^8 molecules cm⁻³, respectively, representing an increase of ever about 30% compared to the levels in 2020 and 2021. Although HONO photolysis was the main contributor to the primary source of ROx radicals, the radical cycling process remained dominant for the overall production of ROx radicals. Due to a significant decrease in NO₂ concentration relative to VOCs, the average VOCs/NO₂ ratio increased from 1.6 in 2020 to 3.0 in 2022, which is also reflected in the radical cycling. The ratio of OH radical propagation (OH+VOCs) to termination (OH+NO₂) was 2.3749, higher than 1.1803 in 2020 and 1.7869 in 2021, indicating that the different reduction proportions of precursors led to a higher VOCs/NO₂ ratio, strengthening the radical cycling. The differential reduction in precursor VOCs and NO₂ levels due to static management is the underlying cause for the increase in ozone concentration in Shanghai.

1 Introduction

To curb the spread of Omicron variant in Shanghai, China, the local government decided to implement city-wide static management in early April 2022. The strict two-month lockdown severely impacted the economic activities and human life of this mega city. According to official statistics (<https://tjj.sh.gov.cn/sjfb/index.html>), in April and May 2022, Shanghai experienced a year-on-year decrease of 42% in its total industrial output value. Moreover, the total volume of transported goods decreased by 30% year-on-year, with road transport witnessing a significant drop of 64%. Additionally, the port cargo

37 throughput decreased by 31% year-on-year. The direct effect of such lockdown policies on air quality is a significant reduction in
38 anthropogenic emissions, which can be considered as an ideal experiment on emissions control in a mega-city to explore the
39 reduction potential and response of air quality . It makes sense to take advantage of this rare yet regrettable window to study the
40 causes and management of air pollution, especially in ~~these~~ countries like China that face complex air pollution ~~complex~~ issues.
41 Prior to this, China had implemented a series of nationwide lockdown measures against the occurrence and spread of the virus in
42 early 2020. This reduction in human activity is expected to significantly reduce air pollutant emissions, as confirmed by lots of
43 studies on lockdown in 2020 (Bao and Zhang, 2020; Huang et al., 2021; Li et al., 2021b; Liu et al., 2020; Tian et al., 2021; Wang
44 et al., 2021; Zhang et al., 2022b). Reports on the impact of the lockdown on air quality most commonly focus on measuring
45 nitrogen dioxide (NO₂) and fine particulate matter (PM_{2.5}) (Agarwal et al., 2020; Hua et al., 2021; Chu et al., 2021). According
46 to satellite data, tropospheric nitrogen oxides (NO_x) emissions have decreased by 30-60% compared to pre-lockdown levels
47 (Feng et al., 2020; Ding et al., 2020; Venter et al., 2020). Similarly, surface PM_{2.5} levels in northern China have also decreased
48 by approximately 35%. Meanwhile, the average O₃ concentration has increased 1.5-2 times (Shi and Brasseur, 2020). In Wuhan,
49 the urban area that implemented stringent measures to limit the spread of the coronavirus, concentrations of PM_{2.5}, NO₂, and
50 ozone also exhibited similar changes (Shi and Brasseur, 2020). Among these pollutants, NO₂ from traffic sources has shown the
51 most significant reduction, with traffic-related NO₂ exhibiting the largest decrease (Rana et al., 2021; Wang et al., 2020). Huang
52 et al. (2021) suggest that increase in O₃ enhances atmospheric oxidation capacity, providing favorable conditions for the
53 formation of secondary particulate matter. Due to the lockdown taking place during winter, which is a season of high particulate
54 matter pollution in China, the reports on the impact of the lockdown on air quality have focused more on the changes in
55 particulate matter. The lockdown in Shanghai in 2022 was implemented in April and May during the high-ozone (O₃) season and
56 lasted for a longer duration, providing an opportunity to study atmospheric pollution primarily caused by O₃.

57 Previous extensive research has demonstrated that the formation of O₃ in response to its precursors is highly nonlinear, rather
58 than linear, which presents a challenge in ozone control (Liu and Shi, 2021; Wang et al., 2017; Sillman, 1999). The COVID-19
59 pandemic provided a costly experiment to validate this. ~~De~~uring the static management period in Shanghai, despite a reduction in
60 precursor emissions, the ozone levels increased compared to the previous year. The cause of this increase is attributed to an
61 imbalance in the reduction ratio of nitrogen oxides (NO_x) and volatile organic compounds (VOCs) rather than meteorological
62 conditions, according to satellite observation results (Tan and Wang, 2022; Xue et al., 2022). In this current study, aim is to
63 elucidate the reasons behind the increase in ozone levels in Shanghai through a comprehensive approach involving in-situ
64 observations, mathematical analysis, and modeling. Benefitting from our conducted routine observational campaigns, we have
65 obtained comprehensive observational data for both the static management period and corresponding historical periods. We
66 initiated our analysis by comparing pollutant levels and diurnal variations during the static management period with those from
67 historical reference periods. Subsequently, by clustering diurnal ozone profile patterns, we examined the reasons behind the
68 elevated ozone levels from a statistical perspective. Furthermore, we discuss changes in radical chemistry compared to historical
69 periods, shedding light on the increase in ozone concentrations from a photochemical process standpoint. It is worth mentioning
70 that, based on previous researchs and our analysis, meteorological conditions are not considered to be the primary cause of the
71 ozone increase during the static management period, even though we acknowledge that meteorological conditions are indeed
72 important factors influencing ozone levels. The in-depth comparison of meteorological conditions is presented in Text S1 of the
73 Supplement.

74 2 Experimental Details and Methods

75 2.1 Location and Experimental Setup

76 During the static management period in April and May of 2022, we conducted comprehensive observational experiments at
77 Fudan University's Jiangwan Campus (31.34°N, 121.51°E), located in an urban area in northeastern Shanghai, China. Similar
78 experiments were also conducted during the corresponding period in 2020 and 2021. The ambient concentrations of O₃, NO₂,
79 SO₂, HONO, and HCHO were measured using the Differential Optical Absorption Spectroscopy (DOAS) system located on the
80 rooftop of the Environmental Science Building (Zhu et al., 2020; Guo et al., 2021; Zhu et al., 2022). Based on the given optical
81 path length and integration time, the detection limits for O₃, NO₂, SO₂, HONO, and HCHO were approximately 1.3 ppbv, 0.5
82 ppbv, 0.1 ppbv, 0.1 ppbv, and 0.5 ppbv, respectively. The measurements were carried out with a time resolution of 5-6 minutes
83 and detailed fitting configurations are available in Table S1. Non-methane volatile organic compounds (NMVOCs) were
84 monitored in real-time using the TH-300B online monitoring instrument, which has been previously described in detail in
85 previous reports (Gu et al., 2022; Zhu et al., 2020). The photolysis rate of NO₂ ($j(\text{NO}_2)$) was measured using a filter radiometer
86 (Meteorologieconsult GmbH, Germany). The meteorological parameters data such as pressure (P), temperature (TEMP), relative
87 humidity (RH), wind speed (WS), wind direction (WD), and boundary layer height (BLH) are derived from the European Center
88 for Medium-Range Weather Forecasts (ECMWF) atmospheric reanalysis product ERA5 and extracted from the nest where the
89 measurement site is located. The PM_{2.5} data was obtained from the Yangpu environment monitoring station (31.53°N,
90 ~~31.21~~21.25°E) near the measurement site.

91 2.2 Observation-Based Model (OBM)

92 The open-source zero-dimensional box model tool AtChem2 was used to simulate atmospheric chemical processes, which is
93 specifically designed for use with the Master Chemical Mechanism (MCM) (Sommariva et al., 2020). The MCM, one of most
94 widely used chemical mechanism for chemistry, is a near-explicit chemical mechanism which describes the degradation of
95 methane and 142 nonmethane VOCs and over 17000 elementary reactions of 6700 primary, secondary and radical species
96 (<http://mcm.york.ac.uk/>, last access: 16 January 2023) (Jenkin et al., 2003; Saunders et al., 2003). AtChem2 software and
97 documentation can be found on <https://github.com/AtChem/> (last accessed on January 16, 2023).

98 In this study, the observed data of O₃, NO₂, SO₂, HONO, HCHO, NMVOCs, $j(\text{NO}_2)$, P, T, RH, and BLH were used as inputs to
99 constrain the model calculations. The photolysis rates of other molecules such as O₃, HCHO, HONO, and OVOCs can be
100 calculated in this model platform with the basic principle driven by the solar zenith angle and scaled by the measured J_{NO_2}
101 (Sommariva et al., 2020). The removal of all unconstrained and simulated species caused by the deposition is determined by a
102 parameterization approach, and is determined by the accumulation of the deposition velocity of 0.01 m s⁻¹ within the boundary
103 layer (Santiago et al., 2017). The sensitivity of simulation results to the deposition velocity has been studied in previous research,
104 and the impact is limited (Zhu et al., 2020). The model outputs include the concentration of the hydroxyl radical (OH) and
105 hydroperoxy radical (HO₂), as well as the reaction rates at each step of the simulation process.

106 2.3 Machine learning

107 Two machine learning methods, k-means clustering and the stacking model, were utilized in this study. In order to investigate the
108 reasons for the overall increase in ozone levels during the static management period from the perspective of its diurnal variation,
109 the k-means clustering method was applied to cluster 24-hour time series of O₃ concentration (Zhang et al., 2004). Time-series
110 clustering is a specific application of curve clustering, which is similar to trajectory clustering in the transport of air masses

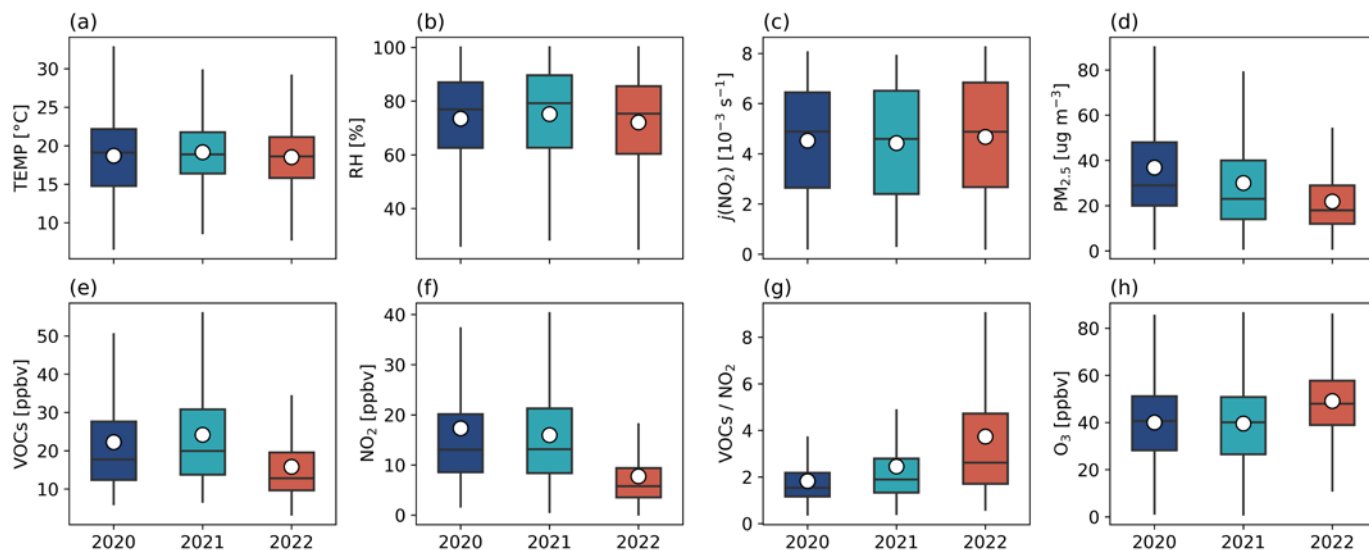
(Darby, 2005; Suris et al., 2022). The procedure for k-means clustering is as follows: (i) randomly initialize k clusters and then calculate the cluster centroid or mean, (ii) assign each data point to the nearest cluster using an appropriate distance measure, (iii) re-calculate the cluster centroids based on the current cluster members, (iv) repeat steps ii and iii until there is no further change. Additionally, the stacking model was applied to address missing values in DOAS observations caused by uncontrollable factors, ensuring the continuity and variation characteristics of the data. This step was deemed necessary for two main reasons. Firstly, the clustering analysis of O₃ diurnal variation demands a continuous time series without any missing values. Secondly, compared to the conventional method of handling missing data in the input of the OBM model through simple linear interpolation, the stacking model preserves the diurnal variation characteristics of the data, ensuring the correct constraints on the OBM model. The stacked model is an ensemble machine learning algorithm that consists of two levels, with two or more base models at level 0 and one meta-model at level 1. The meta-model is trained using predictions made by the base models on out-of-sample data. In other words, data that was not used to train the base models is fed into them to make predictions. These predictions, along with the corresponding expected outputs, form the input and output pairs of the training dataset used to fit the meta-model. The stacking model has been previously described in detail and demonstrated good performance in Zhu et al. (2022), and the architecture of the stacking model can also be found in Figure S1. In this study, the models for O₃, NO₂, SO₂, HONO, and HCHO demonstrated good performance, as shown in Figure S2-S6 of the Supplement.

126 **3 Results and Discussion**

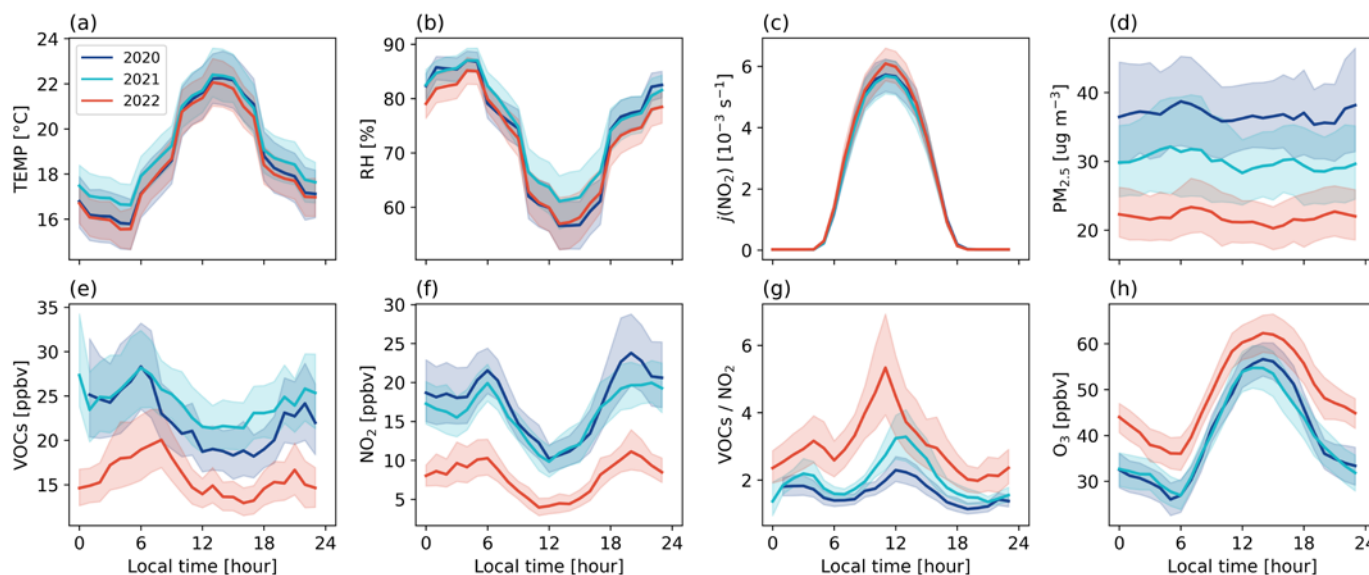
127 **3.1 Year-on-year changes on air quality**

128 Figure 1 compares the average levels of the meteorological parameters and air pollutants during the period from April to May of
129 2020 to 2022, while Figure 2 compares the diurnal variations. In terms of meteorological parameters, the temperature and
130 relative humidity, and $j(\text{NO}_2)$ during the static management period in 2022 were almost unchanged compared to the same period
131 in 2020 and 2021. The average temperature difference in 2022 was 6.5°C, which was similar to that of 2020 and slightly higher
132 than that of 2021, while the average relative humidity at noon in 2022 was also comparable to that of 2020 and was 5% lower
133 than that of 2021. Furthermore, we also ruled out the contribution of transport from the surrounding areas to the increase in
134 ozone concentration in Shanghai during the 2022 static management period (see Figures S7 and S10). The abrupt reduction of
135 emissions across the entire industry led to a significant decrease in primary pollutant concentrations. The average concentrations
136 of PM_{2.5} in April and May from 2020 to 2022 were $36.8 \pm 24.1 \mu\text{g m}^{-3}$, $30.0 \pm 23.1 \mu\text{g m}^{-3}$, and $21.8 \pm 14.0 \mu\text{g m}^{-3}$, respectively,
137 showing a decreasing trend over the years. And the diurnal variation profile in Figure 2d shows that PM_{2.5} levels decreased
138 proportionally throughout the entire 24-hour period, without any particularly prominent periods of decrease. The VOCs and NO₂
139 declined by 29% and 55% respectively compared to 2020, and by 35% and 51% respectively compared to 2021. Due to the
140 significant decrease in NO₂ concentration compared to VOCs, the average ratio of VOCs/NO₂ has increased from 1.6 in 2020 to
141 3.0 in 2022. However, the precursor reduction at different magnitudes has led to an increase of approximately 23% in the
142 average level of ozone. The photochemical production of ozone is controlled by the non-linear chemistry of the precursors VOCs
143 and NO_x (NO₂+NO). The literatures have shown that Shanghai in the spring largely operates under VOCs-limited regime (Li et
144 al., 2021a; Xue et al., 2022). Therefore, the reduction in VOCs during the static management period may not be enough to
145 counteract the titration effect of NO_x, and may even alter the ozone formation regime in Shanghai. From the perspective of
146 diurnal variation (see Figure 2), the period with a significant difference in the magnitude of the decrease between VOCs and
147 NO₂ occurred during the strong photochemical process in the morning until noon. Therefore, the VOCs/NO₂ ratio during the
148 static management period was significantly higher in the morning compared to the same period in 2020 and 2021. The

149 weakening of the titration of nitrogen oxides on ozone during nighttime led to significantly higher nighttime average levels
 150 during the static management period compared to 2020 and 2021. Due to the higher O_3 baseline concentration and higher
 151 VOCs/ NO_2 ratio, there was a significant increase in overall ozone levels.



152
 153 **Figure 1. Comparison of meteorological parameters (TEMP, RH, $j(NO_2)$) and air pollutants ($PM_{2.5}$, VOCs, NO_2 , VOCs/ NO_2 , O_3)**
 154 **during the periods from April to May of 2020, 2021, and 2022. The top and bottom of the vertical line for each box correspond to the**
 155 **95th and 5th percentiles, respectively. The dots represent the averages, and the top, middle, and bottom lines of the box mark the 75th,**
 156 **50th, and 25th percentiles, respectively.**



157
 158 **Figure 2. The mean diurnal profiles of meteorological parameters and air pollutants during the periods from April to May of 2020,**
 159 **2021, and 2022. Colored areas denote 95% confidence intervals.**

160 As VOCs are crucial precursors for ozone formation, we conducted a comparison of each VOCs component during the 2022
 161 static management period with those of the same period in 2020 and 2021, as shown in Figure 3. We classified the 103 VOCs
 162 into six categories based on functional groups, including alkanes, alkenes, alkynes, aromatics, oxygenated VOCs (OVOCs), and
 163 halohydrocarbons. The detailed classification is available in the Table S2. The results revealed that aromatics experienced the
 164 most significant year-over-year reduction in absolute terms. The chemical raw materials and chemical products manufacturing
 165 industry, which is the main source of aromatics (Liu et al., 2019), accounts for 10% of the total industrial output value, and this
 166 industry experienced a 32% year-on-year decrease in total output value during the static management period

(<https://tjj.sh.gov.cn/sjfb/index.html>). In contrast, OVOCs remained relatively stable, primarily because they are sourced from biogenic (Liu et al., 2019), and thus, were less impacted by lockdown measures relative to other VOCs. From the perspective of ozone formation potential (OFP), in the years 2020 and 2021, the primary contributors were aromatics, followed by alkenes, while in 2022, the primary contributors shifted to alkenes, with OVOCs coming in second (see Text S4 and Figure S11).

The reduction in the imbalance of VOCs has altered the average proportion of each component. Specifically, the proportion of aromatics decreased from 8.9% and 11.5% to 4.1%, while the proportion of OVOCs increased from 12.7% and 13.7% to 17.2%. The photolysis of OVOCs is a major source of the important radicals RO_x (OH+HO₂+RO₂) in the photochemical cycle, with a daily average contribution rate of over 30% (Xue et al., 2016). In the radical chemistry section, the photolysis of OVOCs, as well as the reactions of O₃ and NO₃ with VOCs, have been quantified for their contributions to the radicals. Additionally, the role of VOCs in the propagation of radicals has been quantified. ~~Consequently, the rise in OVOC proportion during the static management period has the potential to enhance the photochemical process.~~ The mean diurnal profiles of the VOCs indicated that the daily average concentration range in 2020, 2021, and 2022 was between 18-28 ppbv, 21-28 ppbv, and 13-20 ppbv, respectively.

In 2022, the peak time of VOCs was observed at 08:00, which exhibited a delay compared to the peak times observed in 2020 and 2021 at 06:00, resembling the previously reported “weekend effect” on VOCs that the peak time of VOCs is delayed on weekends in comparison to weekdays (Cai et al., 2010). In Figure S12 of Supplement, we observe that the peak of VOCs diurnal profile in June 2023, after the lifting of restrictions, returned to 06:00. This finding indicates that the reduced human activities during the 2022 period, similar to weekends, led to a decline in anthropogenic VOC emissions in the morning.

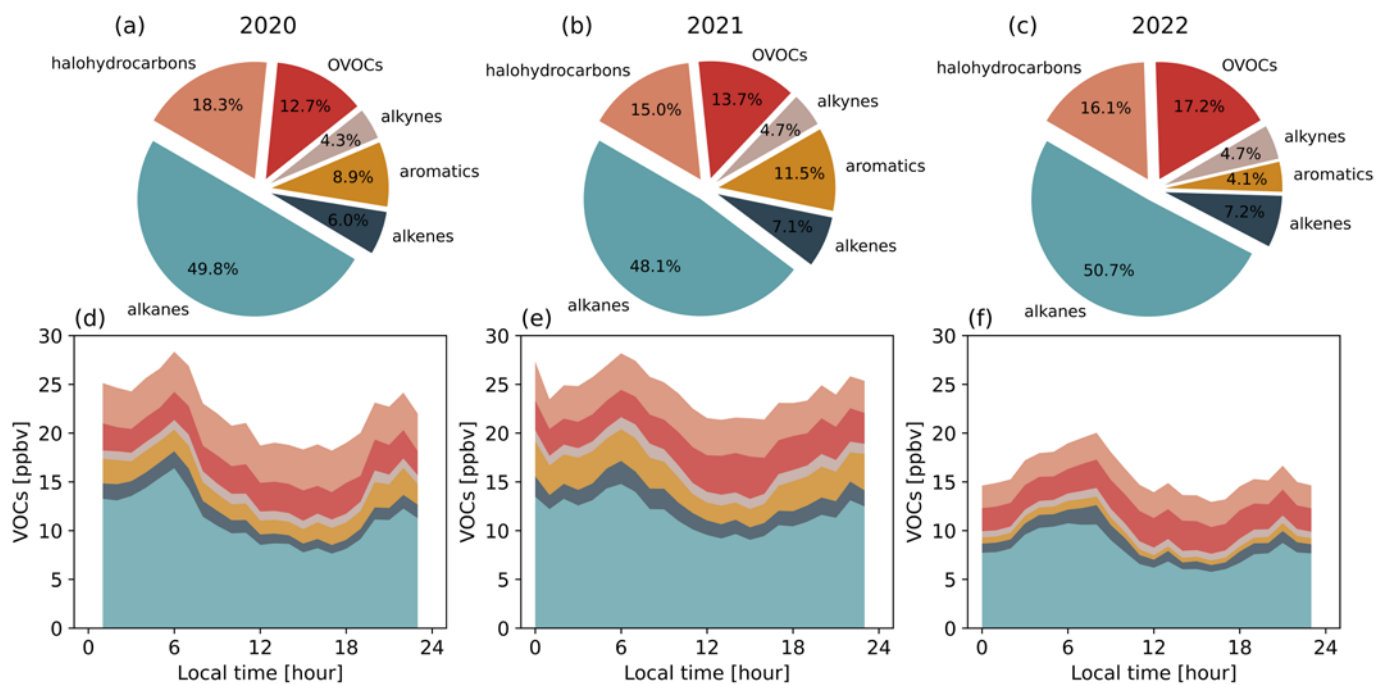


Figure 3. The proportions (a, b, c) and the mean diurnal profiles (d, e, f) of different VOCs components during the periods from April to May of 2020, 2021, and 2022.

3.2 Clustering of O₃ diurnal profiles

The k-means algorithm clustered the ozone diurnal profiles over the three years into four types, as shown in Figure 4a. We define the minimum concentration in the ozone diurnal profile as the background concentration, and the difference between the midday

192 [peak and the morning trough represents the net ozone production](#). These four types of profiles can be described as follows:
193 Cluster 1 with low background concentration [\(22.4 ppbv\)](#) and low net production [\(14.8 ppbv\)](#); Cluster 2 with low background
194 concentration [\(16.9 ppbv\)](#) and high net production [\(45.9 ppbv\)](#); Cluster 3 with high background concentration [\(40.6 ppbv\)](#) and
195 low net production [\(17.5 ppbv\)](#); and Cluster 4 with high background concentration [\(33.3 ppbv\)](#) and high net production [\(50.9](#)
196 [ppbv\)](#). The background concentration of ozone is mainly determined by the nighttime loss of ozone and the titration of nitrogen
197 oxides in the morning, while the net production depends on the intensity of the photochemical reactions. In Figure 4b, the four
198 ozone profiles occurred for 13, 24, 20, and 4 days in 2020, comparable occupation of 19, 21, 17, and 4 days in 2021, respectively.
199 During the static management period of 2022, when the nitrogen oxide titration effect weakened, the number of days on which
200 Cluster 3 and Cluster 4 appeared increased to 34 and 16, respectively. The average ozone levels for the four types were
201 comparable across 2020, 2021, and 2022, ranging from 27-30 ppbv for Cluster 1, 38-40 ppbv for Cluster 2, 46-49 ppbv for
202 Cluster 3, and 54-61 ppbv for Cluster 4 (Figure 4c). As depicted in Figure 4d, assuming that the proportions of the four types in
203 2022 are the same as those in 2020 and 2021, the average ozone levels in 2022 would decrease by 14.5% and 17.3%,
204 respectively, remaining comparable to the levels in 2020 and 2021. Alternatively, if the proportions of the four types in 2020 and
205 2021 were the same as those in 2022, the average ozone levels in 2020 and 2021 would increase by 15.4% and 20.2%,
206 respectively, which would be very close to the levels in 2022. Purely statistical analysis indicated that the significant [increase](#)
207 [high and stable level of](#) ~~ozone levels~~ in 2022 was due to a higher proportion of Cluster 3 and Cluster 4, which had higher ozone
208 concentrations during the static management period.

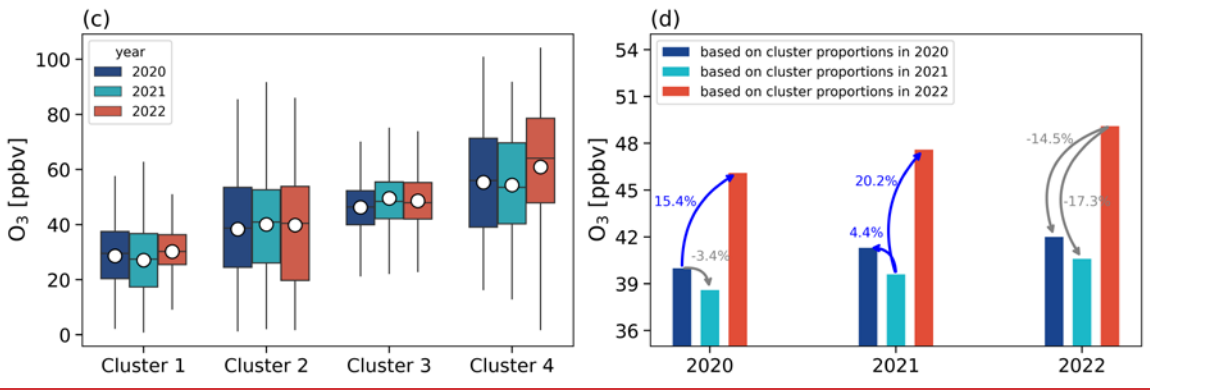
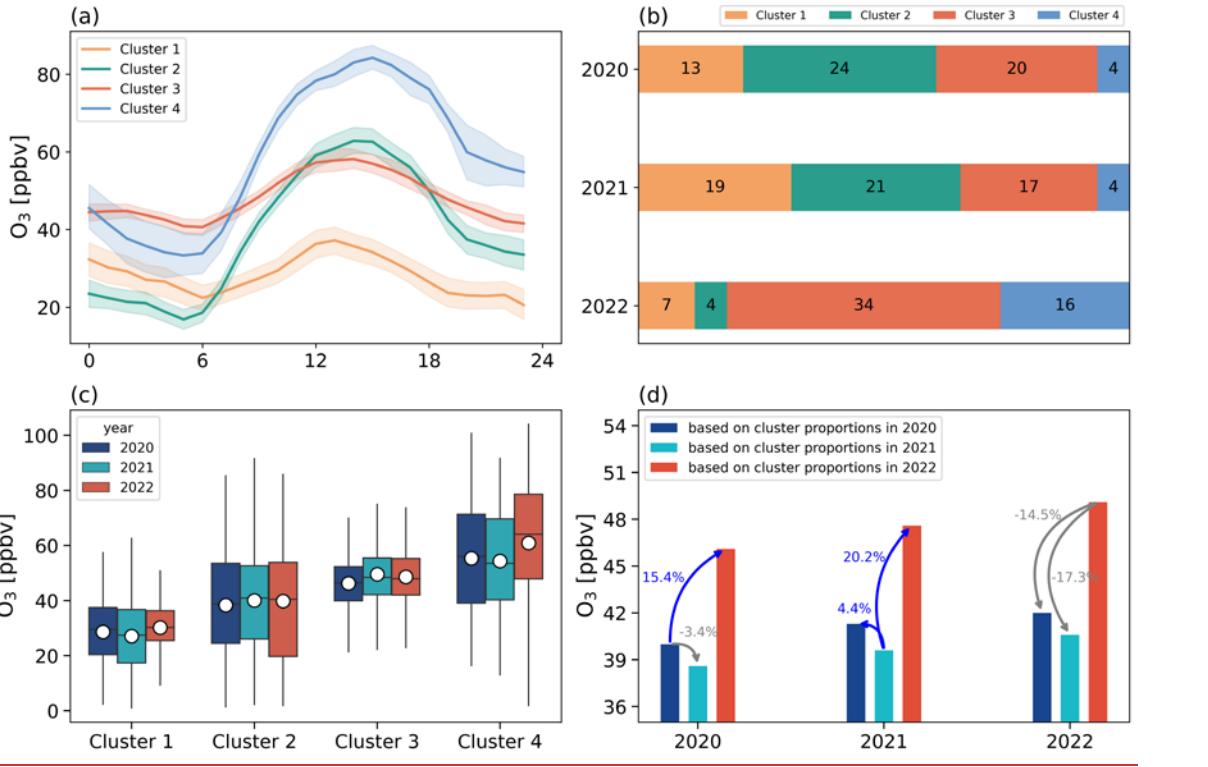
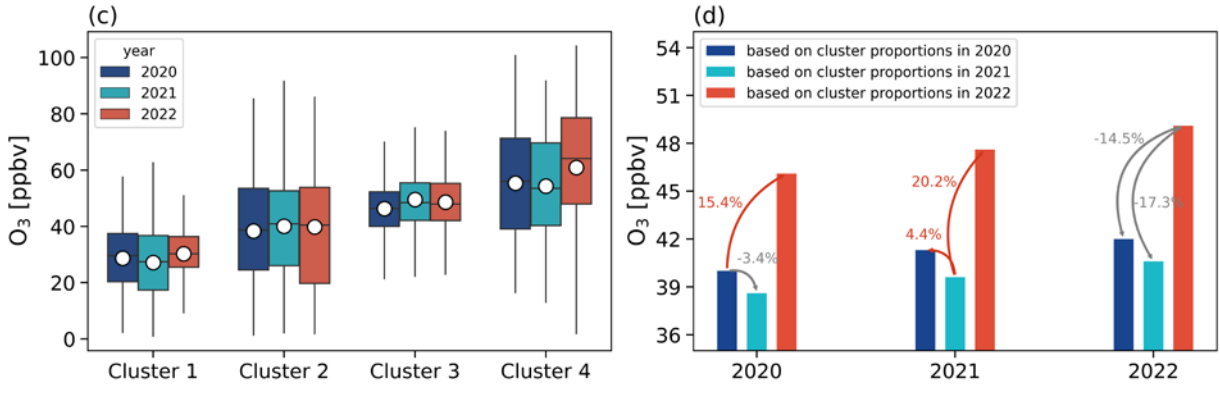
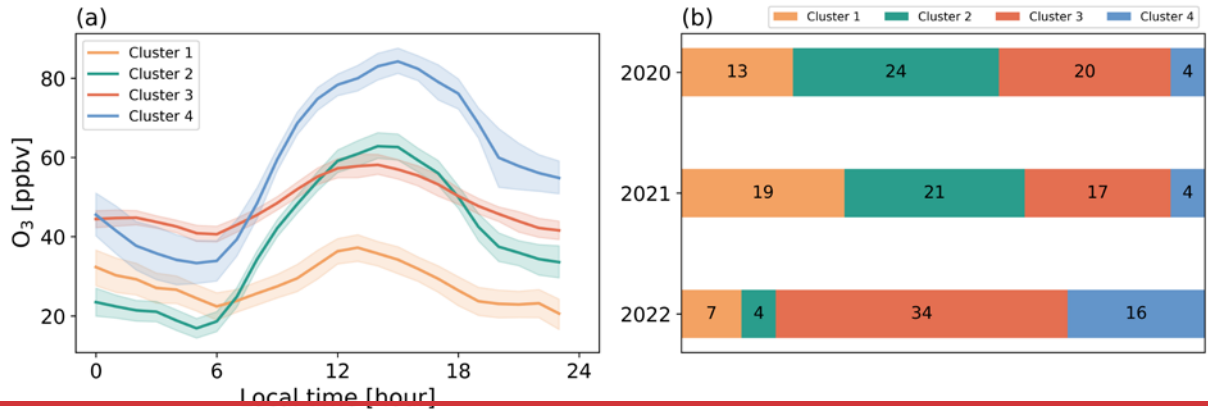
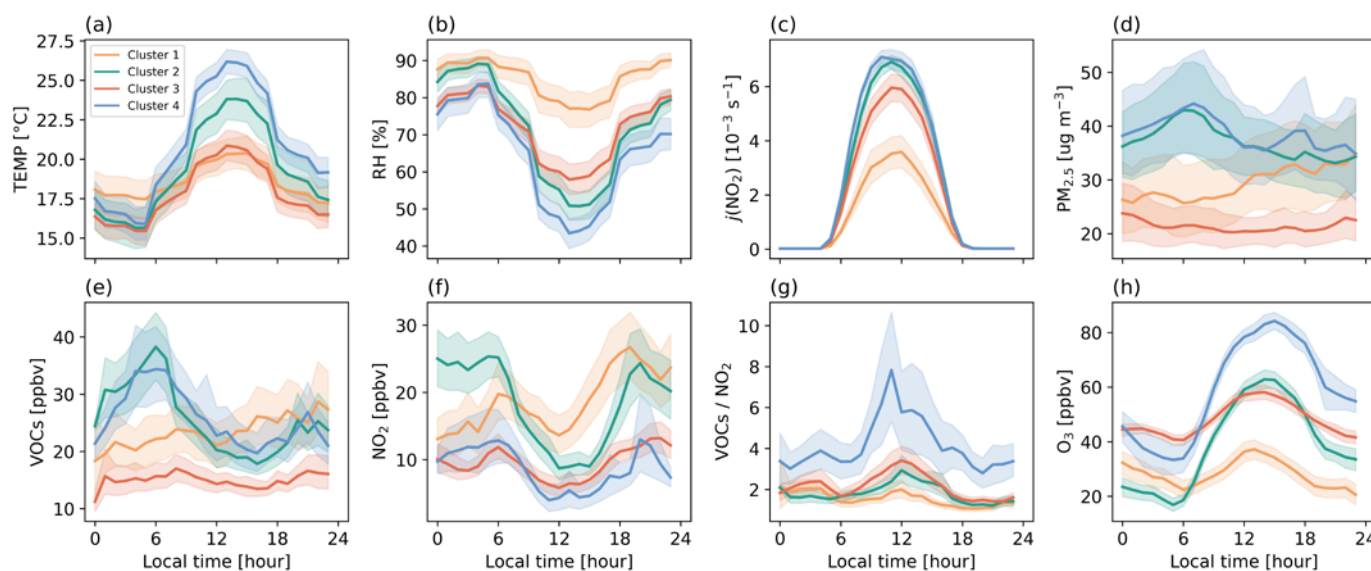


Figure 4. (a) Comparison of the mean diurnal profiles of the four types of O₃ after clustering. Colored areas denote 95% confidence intervals; (b) The proportions of the four clusters in 2020, 2021 and 2022. (c) Comparison of the O₃ levels of the four clusters in 2020, 2021 and 2022. The top and bottom of the vertical line for each box correspond to the 95th and 5th percentiles, respectively. The dots represent the averages, and the top, middle, and bottom lines of the box mark the 75th, 50th, and 25th percentiles, respectively; (d) Comparison of the average ozone concentrations in 2020, 2021 and 2022 for different ratios of the four clusters.

216 Different types of ozone profiles are formed under different meteorological conditions and pollution environments with
 217 distinctive diurnal variations. As shown in Figure 5, the meteorological conditions during the periods of Cluster 2 and Cluster 4,
 218 with high net ozone production, were characterized by high temperature, low relative humidity, and high radiation, compared to
 219 those during Cluster 1 and Cluster 3 periods. This is consistent with the well-known favorable condition promoting ozone
 220 production. The valley values of ozone profiles are closely related to the titration of nitrogen oxides, as shown in Figures 5f and
 221 5h, where the valley values of ozone are inversely related to the peak values of NO₂ during the morning rush hour. Indeed, it is
 222 these meteorological conditions and titration that result in the formation of the corresponding four clusters of ozone profiles.
 223 During periods of Cluster 2 and Cluster 4, a large amount of VOCs accumulated before sunrise and were rapidly consumed after
 224 sunrise. Similarly, the precursor NO₂ was also rapidly consumed after sunrise, with the difference that the NO₂ level during
 225 Cluster 4 was lower than that of Cluster 2. The diurnal profiles of PM_{2.5} under four clusters exhibit similar patterns to those of
 226 VOCs, with Cluster 2 and Cluster 4 exhibiting a distinct morning peak. The VOCs/NO₂ ratio in Cluster 4 was significantly
 227 higher than that in other clusters, which may explain the substantial net ozone production despite the relatively high ozone
 228 background levels. The differences among the clusters are also reflected in the photochemical processes. In the following
 229 sections, we investigated the reasons for the increase in ozone levels during the static management period from the perspective of
 230 atmospheric oxidizing capacity and free radical chemistry.



231
 232 **Figure 5. The mean diurnal profiles of meteorological parameters and air pollutants for four clusters over three years. Colored areas**
 233 **denote 95% confidence intervals.**

234 3.3 Radical chemistry

235 Figures 6a, 6e and 6i show the mean diurnal profiles of simulated OH, HO₂ and RO₂ radical concentrations for the years 2020,
 236 2021, and 2022. These radicals exhibit clear diurnal variations, with peaks occurring at midday. The mean diurnal profiles
 237 display that the average peak concentrations of OH were $4.95.2 \times 10^6$, 4.84×10^6 , and 5.13×10^6 molecules cm⁻³, those of HO₂
 238 were 2.29×10^8 , 2.76×10^8 , and 4.79×10^8 molecules cm⁻³, and those of RO₂ were 0.98×10^8 , 1.64×10^8 and 2.6×10^8
 239 molecules cm⁻³ in 2020, 2021 and 2022, respectively. The differences in radicals concentration levels were mainly reflected in
 240 the HO₂ and RO₂ radicals. The average peak values of the HO₂ and RO₂ radicals in 2022 were about 1 to 3 times higher than
 241 those in the same periods of 2020 and 2021. Reviewing previously observational results, peak concentrations of OH and HO₂
 242 were observed at various locations and times (see Table S3): $(4-17) \times 10^6$ molecules cm⁻³ and $(2-24) \times 10^8$ molecules cm⁻³ at a
 243 suburban site in Yufa from Aug 18-31, 2006 (Lu et al., 2013); $(5-15) \times 10^6$ molecules cm⁻³ and $(3-14) \times 10^8$ molecules cm⁻³ at a

244 rural site in Wangdu from June 8 to July 8, 2014 (Tan et al., 2017); 4.5×10^6 molecules cm^{-3} and 3×10^8 molecules cm^{-3} at a
245 suburban in Heshan from October 22 to November 5, 2014 (Tan et al., 2019); $(2-9) \times 10^6$ molecules cm^{-3} and $(2-14) \times 10^8$
246 molecules cm^{-3} at a urban sites in Shenzhen from Oct 5-28, 2018 (Yang et al., 2022); $(8-24) \times 10^6$ molecules cm^{-3} and $(4-28) \times$
247 10^8 molecules cm^{-3} at a suburban site in Taizhou from May 23 to June 18, 2018 (Ma et al., 2022); and $(10-20) \times 10^6$ molecules
248 cm^{-3} and $(6-18) \times 10^8$ molecules cm^{-3} at a suburban site in Chengdu from Aug 10-25, 2019 (Yang et al., 2021). The simulated
249 concentrations of OH and HO₂ in this study were comparable to the observed levels during autumn in Shenzhen, which is also an
250 urban site, however, were generally lower than those observed at non-urban sites. This difference can be attributed to the site
251 types, but more importantly, to the fact that most observations were conducted during periods of stronger radiation. For RO₂, the
252 average maximum concentration was simulated to be 4.5×10^8 molecules cm^{-3} at urban site of Beijing in August 2007 (Liu et al.,
253 2012). At coastal site of Xiamen, the simulated average daily peak reached 4.7×10^8 molecules cm^{-3} in September 2019, while at
254 the coastal site of Ningde, the simulated value was 0.9×10^8 molecules cm^{-3} in spring 2019 (Liu et al., 2022). Overall, our ROx
255 concentrations fell within the range of observations and simulated results in other regions of China. During the static
256 management period in 2022, the levels of ROx were significantly higher compared to the same period in 2020 and 2021,
257 indicating an enhanced atmospheric oxidation capacity in Shanghai in 2022.

258 Figures 6b-6d, Figures 6f-6h, and Figures 6j-6l illustrate the mean diurnal variation of primary OH, HO₂, and RO₂ sources for
259 the years of 2020, 2021, and 2022. For OH, HONO photolysis peaked at around 07:00 and remained high until around 12:00,
260 with peak values reaching approximately 0.7257 ppbv h⁻¹, 0.5244 ppbv h⁻¹, and 0.6059 ppbv h⁻¹ in 2020, 2021, and 2022,
261 respectively. Meanwhile, ozone photolysis peaked at noon, with peak values reaching around 0.8365 ppbv h⁻¹, 0.8974 ppbv h⁻¹,
262 and 0.887 ppbv h⁻¹, respectively. In addition, the ozonolysis of unsaturated VOCs was another source of OH radical, with an
263 average production rate of less than 0.10 ppbv h⁻¹, while other sources such as the photolysis of H₂O₂, HNO₃, and OVOCs were
264 generally negligible. Overall, HONO photolysis for the daytime accounted for 57%, 432%, and 48% of the total OH primary
265 production rates in 2020, 2021, and 2022, respectively, with O₃ photolysis accounting for 3940%, 48%, and 467% in the
266 corresponding years. For HO₂ radical, the most important source was HCHO photolysis, with average production rates during
267 daytime of 0.280 ppbv h⁻¹, 0.217 ppbv h⁻¹, and 0.254 ppbv h⁻¹ in the corresponding years, respectively. The secondary source was
268 OVOCs photolysis, which produce HO₂ at the average rate of 0.140 ppbv h⁻¹, 0.2015 ppbv h⁻¹, and 0.086 ppbv h⁻¹, respectively.
269 Another source to consider was reactions of O₃ and unsaturated VOCs, which had an average rate of around 0.072 ppbv h⁻¹ for
270 three years. For RO₂ radical, the daytime average peak of primary production rates contributed by OVOCs photolysis from 2020
271 to 2022 were 0.124 ppbv h⁻¹, 0.1622 ppbv h⁻¹, and 0.0740 ppbv h⁻¹, respectively. The reactions of O₃ and NO₃ with VOCs were
272 also important primary sources of RO₂ radical. The average daily contributions of O₃+VOC reactions from 2020 to 2022 were
273 0.092 ppbv h⁻¹, 0.1508 ppbv h⁻¹, and 0.084 ppbv h⁻¹, while the contributions of NO₃+VOC reactions were 0.1398 ppbv h⁻¹,
274 0.2043 ppbv h⁻¹, and 0.064 ppbv h⁻¹, respectively. The primary production rate of RO₂ in 2022 was lower compared to the years
275 2020 and 2021. This can be attributed to the fact that the primary sources of RO₂ are the reactions involved VOCs, which
276 significantly decreased during the static management period and further led to the decreased sources of RO₂.

277 Overall, the total primary production rates of ROx in 2022 was 2.34 ppbv h⁻¹, which is lower than 2.94 ppbv h⁻¹ and 2.85 ppbv h⁻¹
278 in 2021, as shown in Figure S13 of the Supplement. remained relatively stable from 2020 to 2022, with values of 1.20 ppbv h⁻¹,
279 1.35 ppbv h⁻¹, and 1.28 ppbv h⁻¹, respectively. These values are close to higher than the simulated value of 1.56 ppbv h⁻¹ in
280 November 2019 in downtown Shanghai (Zhang et al., 2022a), but lower than close to the value of 2.55 ppbv h⁻¹ during the ozone
281 episode in the suburban of Shanghai in 2018 (Zhang et al., 2021). During these three years, reactions involving VOCs (excluding
282 HCHO) accounted for 2518.3%, 4228.1%, and 4926.5%, respectively, which is correlated with the observed VOCs abundances.
283 In 2020 and 2022, HONO photolysis accounted for over 30.7% and 28.4% of the total primary production rates, respectively.

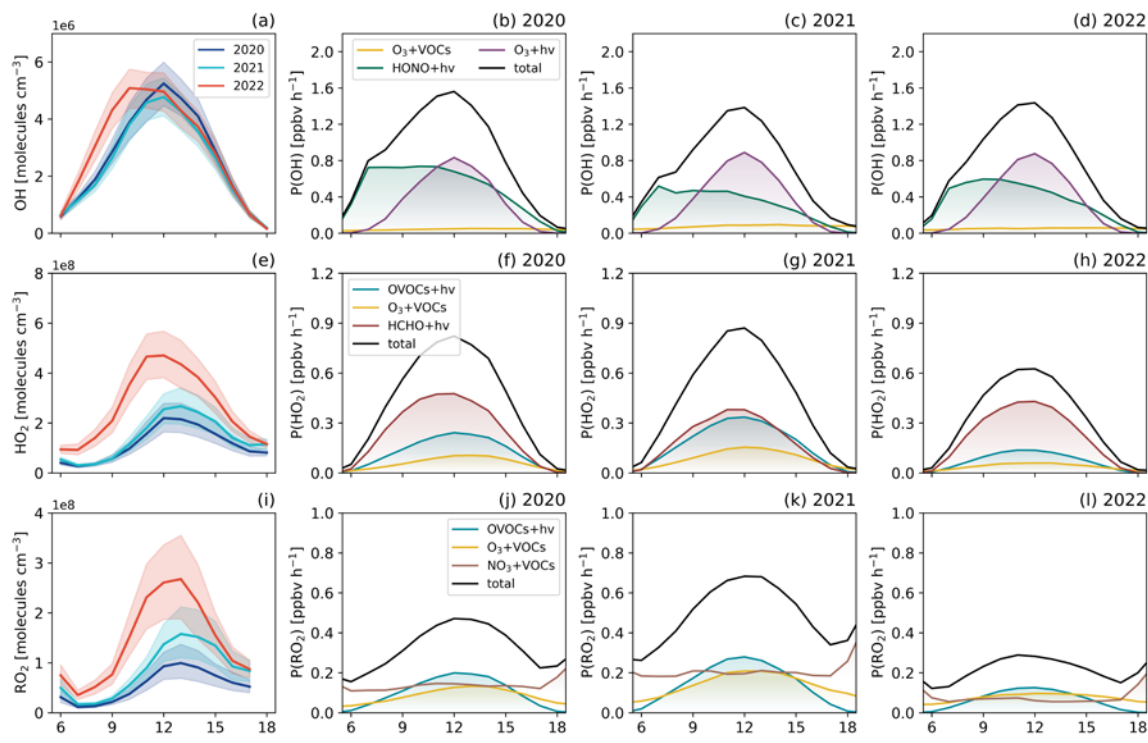
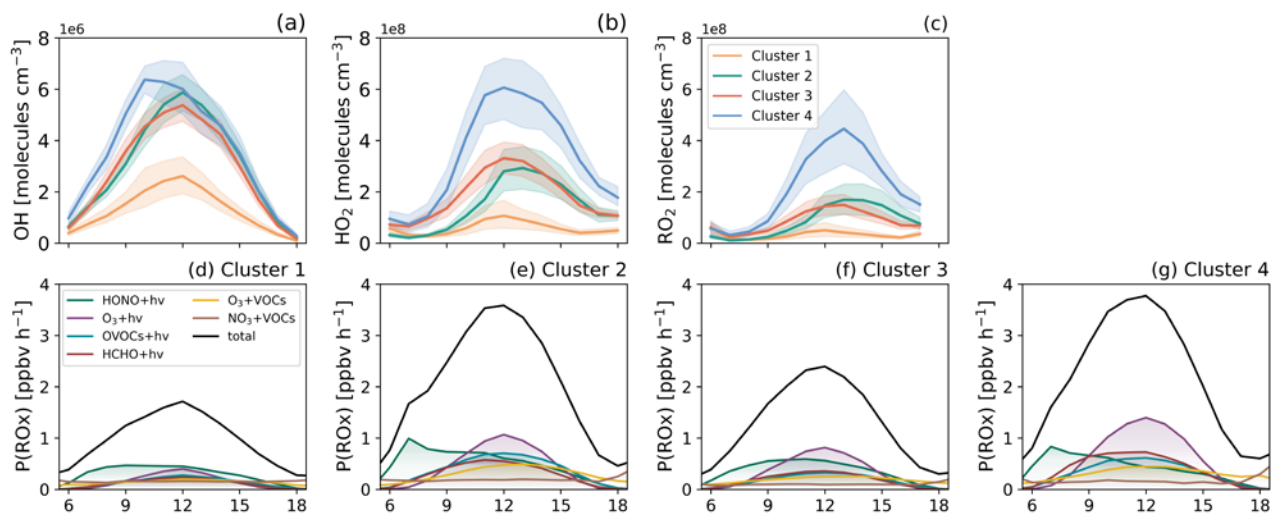


Figure 6. The mean diurnal profiles of simulated OH (a), HO₂ (e), and RO₂ (i) concentrations in 2020, 2021, and 2022. Colored areas denote 95% confidence intervals; The mean diurnal profiles of primary sources of OH radical (b-d), HO₂ radical (f-h), and RO₂ radical (j-l) from model calculations in 2020, 2021 and 2022.

We also investigated the radical chemistry under different ozone profile clusters (Figure 7). The average peak of OH in Cluster 1, Cluster 2, Cluster 3, and Cluster 4 were 2.65×10^6 molecules cm^{-3} , $4.45.9 \times 10^6$ molecules cm^{-3} , $4.85.4 \times 10^6$ molecules cm^{-3} , and 6.44×10^6 molecules cm^{-3} , those of HO₂ were $0.91.1 \times 10^8$ molecules cm^{-3} , 2.95×10^8 molecules cm^{-3} , 3.3×10^8 molecules cm^{-3} , and 6.16×10^8 molecules cm^{-3} , and those of RO₂ were 0.54×10^8 molecules cm^{-3} , 1.73×10^8 molecules cm^{-3} , 1.54×10^8 molecules cm^{-3} , and 4.52×10^8 molecules cm^{-3} , respectively. Clearly, Cluster 4 had the highest levels of OH, HO₂, and RO₂ radicals among all clusters, with HO₂ and RO₂ radicals being particularly prominent. The order of those concentrations among the four clusters is consistent with the order of average ozone concentration. The average peak of P(ROx) for Clusters 1 to 4 were 1.71 ppbv h⁻¹, 3.58 ppbv h⁻¹, 2.39 ppbv h⁻¹, and 3.77 ppbv h⁻¹, respectively. Cluster 2 and Cluster 4, characterized by significant net ozone production, exhibit distinct features in radical chemistry. The daily average of P(ROx) was 1.502.28 ppbv h⁻¹ in Cluster 2 and 1.892.37 ppbv h⁻¹ in Cluster 4, which is higher than the values of 0.781.08 ppbv h⁻¹ in Cluster 1 and 1.441 ppbv h⁻¹ in Cluster 3. In addition, HONO photolysis during the morning rush hour was particularly prominent in Cluster 2 and Cluster 4, with peak values reaching 0.9967 ppbv h⁻¹ and 0.8378 ppbv h⁻¹, respectively. In Cluster 2, HONO photolysis was the dominant source with a daily average of 0.5640 ppbv h⁻¹, accounting for 256% of the total, and ozone photolysis followed with 0.4734 ppbv h⁻¹, accounting for 213%. On the other hand, in Cluster 4, ozone photolysis took the lead with 0.658 ppbv h⁻¹, representing 3027% of the total, and HONO photolysis came next with 0.470 ppbv h⁻¹, accounting for 204%. Additionally,

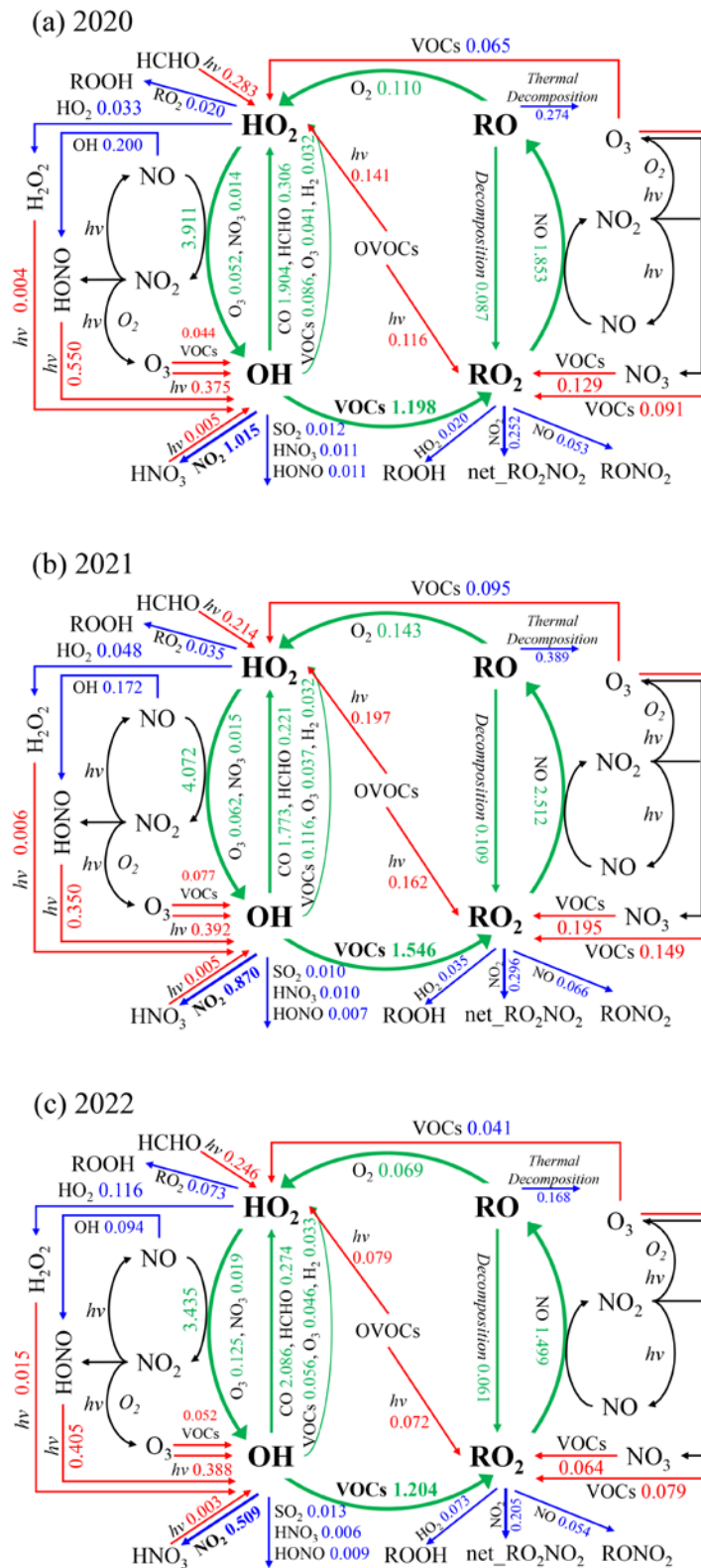
309 OVOCs photolysis (including HCHO) in Cluster 2 and Cluster 4 showed a significant increase compared to Cluster 1 and Cluster
 310 3. In conclusion, The correlation analysis between the each primary source of radicals and the net ozone production reveals a
 311 significant correlation (see Figure S15), which shows that a large amount of net ozone production implies the presence of active
 312 photochemical processes. Cluster 4 is characterized by higher concentrations of HO₂ and RO₂ radicals and a higher primary
 313 source of radicals. Cluster 3 has moderate levels of radical concentrations and primary sources of radicals, but it features a high
 314 ozone background value. The increase in the number of days associated with these two clusters in 2022 has contributed to the
 315 elevated ozone levels.



316
 317 **Figure 7. The mean diurnal profiles of simulated OH (a), HO₂ (b), and RO₂ (c) concentrations for Cluster 1, Cluster 2, Cluster 3, and**
 318 **Cluster 4. Colored areas denote 95% confidence intervals; The mean diurnal profiles of primary sources of ROx radical (d-g) from**
 319 **model calculations for Cluster 1, Cluster 2, Cluster 3, and Cluster 4.**

320 In tropospheric chemistry, radical initiation, which involves the breakdown of closed-shell species to generate new radicals,
 321 plays a crucial role in initiating the formation of secondary pollutants. However, in highly polluted atmospheric environments,
 322 radical cycling becomes the dominant process, with the amplification of new radicals in the ROx cycle playing a crucial role.
 323 Volkamer et al. (2010) quantified the production of new radicals and found that approximately 20% of radical production is
 324 attributed to the breakdown of closed-shell species, while 80% is derived from radical cycling. Therefore, in addition to
 325 understanding the sources of radicals, it is essential to comprehend the impacts of radical cycling and recycling processes on
 326 ozone formation. Figure 8 demonstrated the daytime average of ROx radicals budgets during the periods from April to May of
 327 2020, 2021, and 2022. Taking the simulation of 2022 as an example, OH oxidation of CO and VOCs produces HO₂ and RO₂
 328 with daytime average rates of 2.0944 ppbv h⁻¹ (1.9044 ppbv h⁻¹ in 2020 and 1.7745 ppbv h⁻¹ in 2021) and 1.2007 ppbv h⁻¹
 329 (0.781.20 ppbv h⁻¹ in 2020 and 1.5547 ppbv h⁻¹ in 2021), respectively. The reactions of RO₂+NO and HO₂+NO further lead to
 330 the strong production of RO with a rate of 1.5043 ppbv h⁻¹ (0.981.85 ppbv h⁻¹ in 2020 and 42.51 ppbv h⁻¹ in 2021) and OH with a
 331 rate of 3.4435 ppbv h⁻¹ (2.763.91 ppbv h⁻¹ in 2020 and 3.224.07 ppbv h⁻¹ in 2021), while generating O₃ as a by-product. Clearly,
 332 these recycling processes dominate the overall production of ROx radicals compared to the primary sources. The OH-HO₂ cycle
 333 is an important process in ozone formation. According to the statistics in Table S3, no evidence was found to suggest that the
 334 cycle in 2022 was more efficient than in 2020 and 2021. In terms of termination processes, the loss of ROx radicals was
 335 primarily dominated by their reactions with NOx. Specifically, the reactions of OH+NO₂ and RO₂+NO₂ accounted for
 336 approximately 0.51 ppbv h⁻¹ (0.761.01 ppbv h⁻¹ in 2020 and 0.8774 ppbv h⁻¹ in 2021) and 0.2180 ppbv h⁻¹ (0.2596 ppbv h⁻¹ in
 337 2020 and 4.550.30 ppbv h⁻¹ in 2021) of the ROx radical loss on daytime average, respectively. This is in line with the
 338 understanding that reactions involving NOx typically play a dominant role in the removal of radicals in high NOx environments

339 (Zhang et al., 2021; Xue et al., 2016; Volkamer et al., 2010; Tan et al., 2019; Liu et al., 2012). To sum up, the changes resulting
340 from the approximately 55% reduction in NO₂ and 30% reduction in VOCs due to static management are reflected in both the
341 radical propagation and termination processes. The ratio of OH radical propagation (OH+VOCs) to termination (OH+NO₂)
342 reached 2.3740, which is higher than 1.1803 in 2020 and 1.7860 in 2021. It can be inferred the different proportions of NO₂
343 and VOCs reduction did not weaken the radical cycling. On the contrary, a higher VOCs/NO₂ ratio promotes the radical
344 recycling efficiency in the reaction chain of radicals.



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Figure 8. Averaged budgets (in ppbv h⁻¹) of ROx radicals in daytime (06:00-18:00) during the periods from April to May of 2020, 2021, and 2022. The red, blue, and green lines and words indicate the primary production, termination, and recycling pathways of radicals, respectively.

349 4 Conclusions

350 The two-month city-wide static management was implemented in April and May 2022 in Shanghai aiming to control the spread
351 of the Omicron variant, ~~which provides~~ a valuable opportunity to study the causes of ozone pollution. The comprehensive
352 observations during the static management and same period in 2020 and 2021 shows that, there was a decrease of 29%-35% in
353 VOCs and 51%-55% in NO₂ concentrations, while the average ozone level increased by nearly 23%. By statistics, the ozone
354 profiles were classified into four clusters: Cluster 1, characterized by low background concentration and low net production;
355 Cluster 2, characterized by low background concentration and high net production; Cluster 3, characterized by high background
356 concentration and low net production; and Cluster 4, characterized by high background concentration and high net production.
357 The average O₃ concentration relationship among these clusters is Cluster 4 > Cluster 3 > Cluster 2 > Cluster 1. The significant
358 increase in the proportion of Cluster 4 and Cluster 3 during the period of static management ~~led to an resulted in the~~ overall
359 increase in ~~the~~ average ozone levels. Secondly, from the perspective of radical chemistry, we explored the changes in
360 photochemical processes due to the reduction in precursor species. The OBM model simulated the levels of radicals and their
361 processes of initiation, propagation, and termination. The average peak concentrations of OH, HO₂, and RO₂ in 2022 were 5.13
362 × 10⁶, 4.79 × 10⁸, and 2.6 × 10⁸ molecules cm⁻³, respectively, which were higher than those in the same period in 2020 and 2021.
363 HONO photolysis was the main contributor to the primary source of RO_x, accounting for about 30% of the total. However, in
364 terms of the overall production of RO_x radicals, the radical recycling process remained dominant. The ~~different~~
365 ~~proportions~~ reduction of NO₂ and VOCs ~~reduction in different proportions due to~~ during static management ~~has~~ led to an increased
366 ~~ratio of~~ OH radical propagation (OH+VOCs) to termination (OH+NO₂) ~~ratio~~, reaching 2.37+0, ~~which is~~ higher than 1.1803 in
367 2020 and 1.7860 in 2021. This enhanced ratio indicates a strengthened radical cycling as a result of a higher VOCs/NO₂ ratio.
368 The important lesson ~~we learned~~ from the static management “large-scale field experiment” ~~during the period of static~~
369 ~~management~~ is that Shanghai operates in a VOCs-limited regime. When the reduction in VOCs is not able to catch up with or
370 exceed the reduction in nitrogen oxides, it is not sufficient to curb the formation of secondary pollutants. In terms of ozone
371 control strategies, it is necessary to strengthen the regulation and control of VOCs.

372 Data Availability

373 The observed and predicted hourly time series data in the study are presented in Figure S164-S182, and Code and data used for
374 our analyses are available at <https://data.mendeley.com/datasets/3kmhg7r2df/1> (Zhu, 2023).

375 Competing interests

376 The contact author has declared that none of the authors has any competing interests.

377 Author contributions

378 Jian Zhu: Conceptualization, Methodology, Software, Validation, Investigation, Writing original draft, Visualization. Shanshan
379 Wang: Conceptualization, Methodology, Supervision, Funding acquisition. Chuanqi Gu, Zhiwen Jiang, Sanbao Zhang, Ruibin
380 Xue, and Yuhao Yan: Methodology, Validation, Investigation. Bin Zhou: Conceptualization, Methodology, Supervision,
381 Funding acquisition.

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