



# 1 Non-mycorrhizal root-associated fungi increase soil C stocks and 2 stability via diverse mechanisms

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8 **Abstract.** While various root-associated fungi could facilitate soil carbon (C) storage and therefore aid climate change  
9 mitigation, so far research in this area has largely focused on mycorrhizal fungi, and potential impacts and mechanisms for  
10 other fungi are largely unknown. Here, we assessed the soil C storage potential of 12 root-associated, non-mycorrhizal fungal  
11 isolates (spanning nine genera and selected from a wide pool based on traits potentially linked to soil C accrual) and  
12 investigated fungal, plant and microbial mediators. We grew wheat plants inoculated with individual isolates in chambers  
13 allowing continuous <sup>13</sup>C labelling. After harvest, we quantified C persistence, and pools of different origin (plant vs soil) and  
14 of different stability with long-term soil incubations and size/density fractionation. We assessed plant and microbial  
15 community responses, as well as fungal physiological and morphological traits in a parallel *in vitro* study. While inoculation  
16 with three of the 12 isolates resulted in significant total soil C increases, soil C stability improved under inoculation with most  
17 isolates – as a result of increases in resistant C pools and decreases in labile pools and respired C. Further, these increases in  
18 soil C stability were positively associated with various fungal traits and plant growth responses, including greater fungal hyphal  
19 density and plant biomass, indicating multiple direct and indirect mechanisms for fungal impacts on soil C storage. We found  
20 more evidence for metabolic inhibition of microbial decomposition than for physical limitation under the fungal treatments.  
21 Our study provides the first direct experimental evidence in plant-soil systems that inoculation with specific non-mycorrhizal  
22 fungal strains can improve soil C storage, primarily by stabilising existing C. By identifying specific fungi and traits that hold  
23 promise for enhancing soil C storage, our study highlights the potential of non-mycorrhizal fungi in C sequestration and the  
24 need to study the mechanisms underpinning it.

## 25 1 Introduction

26 Despite soils having the capacity to sequester large amounts of atmospheric CO<sub>2</sub> and mitigate catastrophic climate change, the  
27 full potential of soil carbon (C) sequestration is yet to be realised (Field and Raupach, 2004; Scharlemann et al., 2014;  
28 Schlesinger, 1990). Moreover, rather than being protected, soils are becoming increasingly degraded globally due to intensive  
29 agricultural practices - a situation that may worsen as C loss potentially accelerates with future climate scenarios (Hannula and  
30 Morriën, 2022; Lal, 2018). While soil C sequestration is becoming more broadly recognised as an important climate mitigation



31 strategy, and as an approach to recover the multiple ecosystem services provided by soil C (Kopittke et al., 2022), its successful  
32 implementation first requires understanding of processes underpinning soil C storage (Dynarski et al., 2020; Smith and Wan,  
33 2019; Von Unger and Emmer, 2018). Knowledge of soil C storage has improved substantially in recent years, with it now  
34 understood to result from the balance of multiple, dynamic processes (that are further complicated by pedoclimatic context)  
35 determining C inputs to soil and their stabilisation - which ultimately determine the persistence of soil C at the ecosystem scale  
36 (Cotrufo and Lavallee, 2022; Derrien et al., 2023; Dignac et al., 2005; Dynarski et al., 2020; Jackson et al., 2017; Schmidt et  
37 al., 2011). Soil microbes act as key participants of these processes: they regulate the persistence of soil C primarily via their  
38 abilities to mineralise soil organic matter, which determine how long C of plant or microbial origin persists in soil, and can  
39 also influence how much C is available for stabilisation from their necromass and from plant inputs. However, the soil  
40 microbial community is complex, and largely unknown; hence, referred to as a “black box” (Mishra et al., 2023; Tiedje et al.,  
41 1999). Within this black box, fungi, both free-living and plant-associated, are considered particularly important for soil C  
42 storage; however, their impacts on soil C storage are both multifaceted and diverse.

43 The complexity in fungal impacts on soil C storage firstly arises from their abilities to influence both soil C inputs and  
44 stabilisation via multiple direct and indirect mechanisms occurring simultaneously (Hannula and Morriën, 2022; Kallenbach  
45 et al., 2016; Liang et al., 2019; Starke et al., 2021). In general, fungi that are present in soil (1) all produce hyphae and with  
46 them hyphal C inputs, (2) can alter plant health, growth, and C chemistry and allocation to soil, and (3) can influence the rest  
47 of soil microbial community structure and composition, thus impacting fungal-, plant-, and microbial-derived C, respectively  
48 (Clocchiatti et al., 2020; Hannula and Morriën, 2022; Rai and Agarkar, 2016; Stuart et al., 2022). All of these inputs, but  
49 particularly fungal and plant C, are potentially available for soil C storage but they require stabilisation in order to persist in  
50 soil long term. The broad and efficient enzymatic capabilities and extensive mycelial structure of fungi, as compared to the  
51 rest of the microbial community, allow them to competitively obtain soil C and transform it so that it can be readily sorbed  
52 and stabilised onto mineral surfaces (Boer et al., 2005; Hannula and Morriën, 2022). In addition, fungal necromass is  
53 considered to have a particularly strong affinity for mineral surfaces and is therefore an important source of stabilisable C  
54 (Sokol et al., 2019). The impact of fungi on soil structure and spatial heterogeneity, including promoting aggregate formation  
55 by enmeshing soil particles with their hyphae and producing various extracellular biopolymers, further protects C by physically  
56 constraining microbial decomposition, leading to greater persistence (Berg and Mcclaugherty, 2014; Dynarski et al., 2020;  
57 Kleber et al., 2011; Lehmann et al., 2017; Lützow et al., 2006; Schmidt et al., 2011).

58 These various impacts of fungi on soil C storage are further complicated by fungal diversity, which occurs at the inter-genus,  
59 inter-species, and even down to the sub-species level (Andrade et al., 2016; Hiscox et al., 2015; Johnson et al., 2012; Juan-  
60 Ovejero et al., 2020; Plett et al., 2021). In plant-soil ecosystems, fungi exist either as free-living saprotrophs or as plant-  
61 associated fungi, including mycorrhizal, endophytic, and parasitic fungi (Rai and Agarkar, 2016). Saprotrophic fungi are often  
62 assumed to promote soil C output, as they decompose soil organic matter due to being outcompeted by mycorrhizal fungi for



63 plant C exudates, but as decomposition can increase the availability of C to be sorbed onto mineral surfaces, thereby fostering  
64 soil C stability, their net impacts on soil C storage may need further exploration (Fr ac et al., 2018; Hannula and Morri en, 2022;  
65 Lehmann and Rillig, 2015). Meanwhile, much of the research on the impacts of plant-associated fungi on soil C has focused  
66 on mycorrhizal fungi, particularly arbuscular mycorrhizal fungi and ectomycorrhizal fungi due to their dominance in their  
67 respective habitats (Jackson et al., 2017; Smith and Read, 2008). These fungi have additional impacts, to the general fungal  
68 impacts outlined above, on the inputs, stabilisation, and persistence of C. As they transform and funnel plant C belowground,  
69 mycorrhizal fungi can increase and modify the quality of C inputs, for example by synthesising melanin for cell walls, which  
70 is considered to be highly stable and has been associated with decreased hyphal decomposability and increased soil C content  
71 (Fernandez and Kennedy, 2018; Fernandez and Koide, 2013; Zak et al., 2019; Zhu and Michael Miller, 2003). Due to their  
72 nutrient requirements and abilities to mine soil resources, they are thought to be strong competitors against saprotrophs for not  
73 only plant C but also soil nutrients, thereby suppressing microbial respiration, and resulting in greater C persistence (Gadgil  
74 and Gadgil, 1971; Averill and Hawkes, 2016). Some mycorrhizal fungi have limited abilities to directly and partially decay  
75 organic matter, and they can also prime saprotrophic microbes to decompose pre-existing soil C, thus having the potential to  
76 decrease C persistence – though their net impact on soil C storage is not well understood (Frey, 2019). Despite the large  
77 diversity amongst fungi in plant-soil ecosystems, influences of non-mycorrhizal fungi, particularly other plant-associated  
78 fungi, on soil C storage have not been studied in as great detail compared to mycorrhizal fungi but do hold promise. For  
79 example, endophytic fungi could potentially be important for soil C storage due to their abilities to produce melanin and  
80 promote plant growth (Berthelot et al., 2017; He et al., 2019; Mandyam and Jumpponen, 2005; Rai and Agarkar, 2016).  
81 However, there are conflicting reports regarding their lifestyles, benefits or harms imposed on host plants, enzymatic and  
82 nutrient acquisition ability, or even whether they produce extraradical mycelium, suggesting there may be wide functional  
83 variation or plasticity within this fungal group (Addy et al., 2005; Mukasa Mugerwa and Mcgee, 2017; Rai and Agarkar, 2016).  
84 To better understand the diversity of fungal impacts on soil C storage, particularly soil C stability, more focus is needed on  
85 fungal types other than mycorrhizal fungi.

86 There is growing interest in searching and screening for organisms that, in addition to supporting plant productivity, may  
87 improve soil C storage in agricultural systems (Kaminsky et al., 2019; Islam et al., 2021; Salomon et al., 2022). Thus far,  
88 mycorrhizal fungi have received much attention in this area due to their well-established impacts on plant health and soil C.  
89 However, as discussed above, other fungal types may also offer advantages to soil C storage and plant productivity but have  
90 been largely unexplored. With this objective in mind, in the current study we aimed to determine the potential of diverse non-  
91 mycorrhizal fungi to impact soil C stocks, formation (by impacting the origin of soil C), and persistence (by impacting C pools,  
92 dynamics, and fractions), and to investigate the mechanisms underpinning these impacts, both direct and indirect. We assessed  
93 12 separate fungal species (spanning nine genera in the orders Chaetosphaeriales, Helotiales, and Pleosporales), isolated from  
94 roots collected from multiple soil environments across Australia and screened for traits that may support plant growth and soil  
95 C storage, such as capabilities to capture and solubilise nutrients from the soil. In a pot study, we inoculated spring wheat



96 (*Triticum aestivum*), an important cereal crop, with one of the 12 fungi and grew the plants for a full life cycle in  $^{13}\text{C}$ -depleted  
97  $\text{CO}_2$  growth chambers to homogeneously label the plants during the full growth cycle, in order to distinguish soil C from plant-  
98 derived soil C. Following harvest, we assessed total C and its isotopic composition, and assessed C distribution among pools  
99 of different stability and persistence (labile, intermediate, and resistant) via four-month soil incubations, and evaluated the  
100 contribution of soil and plant C to these pools using isotopic analysis. These incubation-based assessments were accompanied  
101 by size and density fractionation analyses to quantify mineral-associated organic matter (MAOM), aggregate carbon (AggC),  
102 and particulate organic matter (POM). We then measured traits of the fungi and of the plants and microbial community to  
103 explore the potential direct and indirect mechanisms behind these impacts, respectively. We hypothesised that if a fungal  
104 species increased total soil C storage, this would be due primarily to increasing plant C inputs by supporting plant growth and  
105 also to stabilising existing soil C - so that fungi-driven increases in total soil C would be associated with more persistent and  
106 stable pools and fractions of C. We expected that these changes to soil C would be associated with fungal traits, alluding to  
107 direct mechanisms, as well as to increases in plant growth and shifts in microbial community composition, alluding to indirect  
108 mechanisms.  
109



## 110 2 Materials and methods

111 The overall study design consisted of a wheat growth pot experiment, in which changes to soil, plant, and soil microbial  
112 communities in response to fungal inoculation were assessed, and a separate *in vitro* fungal growth assay, to measure fungal  
113 traits that could potentially be linked to observations made in the main experiment (Fig. A1).

### 114 2.1 Experiment set up and maintenance

115 Twelve fungal isolates were originally obtained from plant roots and screened for traits that may support plant growth and soil  
116 C storage by Loam Bio Pty Ltd (Orange, New South Wales, Australia). The fungal isolates, including endophytic fungi and  
117 potentially saprotrophic or other fungi, comprised: a *Thozetella* species, a *Paraconiothyrium* species, three *Darksidea* species,  
118 a *Leptodontidium* species, a *Clohesyomyces* species, two *Phialocephala* species, an *Acrocalymma* species, a *Periconia* species,  
119 and an *Ophiosphaerella* species. Pure cultures of these isolates were maintained on 1/10 strength potato dextrose agar (PDA).  
120 Surface-sterilised (2% NaOCl) and moistened seeds of Australian wheat cultivar Condo (*Triticum aestivum*) were incubated  
121 at room temperature for 48 h. Soil was obtained from an agricultural field, sieved through 2 mm, and was a clay loam (4.3%  
122 C, 0.39% N, pH 5.85; Table B1).

123 The experimental setup consisted of seven planted replicates inoculated with one of the 12 fungal isolates, and six replicates  
124 of uninoculated planted pots, distributed among six CO<sub>2</sub>-controlled growth chambers (Climatron-1260; Thermoline, Wetherill  
125 Park, New South Wales, Australia) that had been modified to achieve continuous <sup>13</sup>C-labeling of plant tissues. For “planted”  
126 replicates, three 7 mm agar squares from actively growing 1/10 PDA fungal culture plates were placed near three sterile seeds  
127 in 2 L plastic pots (at a depth of 2-3 cm) containing 1800 g of non-sterile soil. Uninoculated planted pots (“absent/control”)  
128 received three agar squares from uninoculated plates. Each agar square contained approximately 1.3 mg C. Smaller pots  
129 (containing 500 g of soil) for “unplanted” control pots (see below) were set up three days later using two agar squares, as  
130 controls for impacts of fungi in the absence of plants, adding to 142 pots in total. After 10 days of growth, seedlings were  
131 thinned to one per pot.

132 Pots were regularly and uniformly watered with tap water. Pots within each tub and tubs within each chamber were randomly  
133 relocated four times throughout the experiment. The chamber atmosphere was sampled weekly to confirm that the atmospheric  
134 CO<sub>2</sub> was sufficiently depleted in <sup>13</sup>C via a pump system into a Tedlar® SCV Gas Sampling Bag and δ<sup>13</sup>C analysis in a  
135 PICARRO G2201i isotopic CO<sub>2</sub>/CH<sub>4</sub> analyser (Picarro Inc., Santa Clara, CA, USA).

### 136 2.2 Harvest and plant biomass measurement

137 Once the plants had senesced and the grain had ripened, at 18<sup>th</sup> weeks of growth, wheat spikes and shoots were cut off, dried  
138 at 70°C and weighed. The intact root-containing soil was preserved in the pots by freezing at -20°C immediately after shoots



139 were cut to stop all decomposer activity to retain the C status generated by the treatment until ready for subsampling and  
140 processing. After two days of thawing at 4°C, soil was removed from the pots and a subsample for fractionation analysis was  
141 collected from near the root crown and oven-dried at 40°C. The main root system was gently shaken of soil and 1/3 of the  
142 roots were cut, washed, patted dry, frozen at -20°C prior to root morphology measurement. The rest of the soil was  
143 homogenised before subsamples collection. A subsample for phospholipid fatty acid (PLFA) analysis was frozen at -20°C. A  
144 subsample for soil moisture content was weighed and dried at 105°C. A subsample for soil incubations was oven-dried at  
145 40°C and sieved at 2 mm, and of this, a further subsample for isotope analysis was dried at 105°C. To obtain total root mass,  
146 first the root/soil ratio outside the main root system was estimated by collecting the root mass of the remaining soil (after all  
147 subsampling) via wet sieving (500 µm) and oven-drying at 40°C. The root mass of the soil subsamples was calculated using  
148 this ratio and the amount of soil in all subsamples.

### 149 **2.3 Root morphology**

150 To evaluate root morphology, a potential indirect mechanism for fungal impacts on soil C storage, washed, dried, and frozen  
151 root subsamples were arranged with minimal overlap for digital scanning (Epson Expression 11000XL scanner, Epson,  
152 Macquarie Park, Australia). Images were analysed with WinRhizo Pro software 2015 (Regent Instruments Inc., Quebec City,  
153 Canada) to obtain root average diameter (mm), specific length as the ratio of length to dry mass (cm mg<sup>-1</sup>), tissue density as  
154 mass per unit volume (g cm<sup>-3</sup>), specific surface area as the ratio of area to dry mass (cm<sup>2</sup> g<sup>-1</sup>), and branching as the number of  
155 forks per unit of mass (number mg<sup>-1</sup>). Following root morphology assessment, the root subsample was oven-dried at 40°C  
156 for determination of total root mass.

### 157 **2.4 Plant and soil isotope and chemical analysis**

158 To determine the contribution of soil- versus plant-derived C to total C in soils under wheat, isotopic compositions and C/N  
159 content of ground shoots and soil were assessed using an elemental analyser interfaced to a continuous flow isotope ratio mass  
160 spectrometer (UC Davis Stable Isotope Facility, Davis, California, USA). The proportion of original soil C present in the soil  
161 of each pot after plant growth was calculated via isotopic partitioning following Eq. (1):

$$162 \text{ Soil proportion. Soil} = \frac{(\delta^{13}\text{C}_{\text{Soil}} - \delta^{13}\text{C}_{\text{UP-Soil}})}{\delta^{13}\text{C}_{\text{P}} - \delta^{13}\text{C}_{\text{UP-Soil}}},$$

163 where  $\delta^{13}\text{C}_{\text{Soil}}$  is the <sup>13</sup>C isotopic composition of soil measured in each planted pot,  $\delta^{13}\text{C}_{\text{UP-Soil}}$  is the mean <sup>13</sup>C isotopic  
164 composition of soil in unplanted controls, and  $\delta^{13}\text{C}_{\text{P}}$  is the <sup>13</sup>C isotopic composition of the plant shoots in each planted pot.  
165 The plant C proportion (including C from other biological sources) was defined as 1 minus the soil C proportion. These  
166 proportions were then applied to the measured C concentrations in each pot to calculate plant- and soil-derived C amounts.



## 167 **2.5 Soil incubations**

168 To evaluate the impact of fungal isolates on overall C persistence and C distribution across pools of different stability (labile,  
169 intermediate, and resistant), we assessed microbial CO<sub>2</sub> production during 135-day laboratory incubations of soil harvested at  
170 the time of wheat harvest under standard temperature and moisture conditions, and fitted a decay model to estimate decay  
171 kinetic parameters. Kinetic parameters derived from mid- to long-term soil incubation are sensitive functional measures of  
172 changes in the distribution and stability of C pools resulting from previous exposure to experimental treatments (Carney et al.,  
173 2007; Carrillo et al., 2011; Jian et al., 2020; Langley et al., 2009; Taneva and Gonzalez-Meler, 2008). Measured CO<sub>2</sub>  
174 production rates over time were fitted to a two-pool exponential decay model to estimate the size of the labile and intermediate  
175 C pools and their mean residence time (MRT; Cheng and Dijkstra, 2007; Wedin and Pastor, 1993). The size of the resistant  
176 pool was calculated as the difference between the total measured organic C and the sum of the estimated labile and intermediate  
177 pools. This same procedure was also applied to the portion of CO<sub>2</sub> that was released from the originally present soil C (soil-  
178 derived C, i.e. not plant-derived C), which was determined via isotopic partitioning of plant vs. soil-derived CO<sub>2</sub>. Based on  
179 these, we calculated total CO<sub>2</sub> released from plant- and soil-derived C during the full length of the incubation. See  
180 Supplementary Methods for full details on incubations, isotopic partitioning, and decay modelling.

## 181 **2.6 Soil fractionation analysis**

182 Soil fractionation analysis was performed as an alternative method to soil incubations for understanding fungal impacts on C  
183 stability and potential persistence. Hereafter we refer to the pools measured via fractionation analysis as “fractions”, as  
184 opposed to “pools” measured via soil incubations. The analysis was performed according to a method developed by (Poeplau  
185 et al., 2017; Poeplau et al., 2018) and adapted by Buss et al. (2023, in review) involving high throughput physical  
186 fractionation into conceptually designed soil C fractions - mineral-associated organic matter (MAOM), aggregate carbon  
187 (AggC), and particulate organic matter (POM). See Supplementary Methods for further details.

## 188 **2.7 Soil PLFA analysis**

189 Total microbial community size and composition are also potential indirect drivers of fungal impacts on soil C storage.  
190 Microbial PLFAs in soils were extracted from 2 g of freeze-dried soil harvested from the wheat growth experiment, following  
191 the high throughput method developed and described by Buyer and Sasser (2012; see Supplementary Methods).

## 192 **2.8 *In vitro* fungal assessment**

193 To assess morphological and chemical properties of the fungal isolates (used in the wheat growth experiment) as potential  
194 drivers of fungal impacts on soil C storage, a separate *in vitro* plate assay was performed using 1/2 PDA plates incubated in  
195 the dark at 23-25°C (see Supplementary Methods). Radial growth rate was calculated by measuring colony areas every two-



196 to-three days using ImageJ (National Institutes of Health, Bethesda, Maryland, US; Schneider et al., 2012). Growth rate was  
197 calculated by subtracting the colony area from an earlier sampling point from that of the following sampling point. Hyphal  
198 density was calculated as the final fungal biomass per final colony area. C and N content were measured by Dumas combustion  
199 using a El Vario cube analyser (Elementar, Langenselbold, Germany).

## 200 **2.9 Data and statistical analysis**

201 ANOVA of soil C properties and experimental variables was performed in R (v. 4.1.2; R Core Team, 2021), followed by  
202 Dunnett's post-hoc test to determine which treatment groups were significantly different to the uninoculated control group or  
203 Tukey's post-hoc test to determine significant differences between inoculated groups. Principal component analysis (PCA)  
204 and redundancy analysis (RDA) were performed using the vegan package in R (Oksanen et al., 2020). Missing values in the  
205 PCA and RDA datasets were replaced with the variable mean.

206 Curve fitting of CO<sub>2</sub> rate dynamics was done using the non-linear modelling platform in JMP 16.1.0 and the biexponential  
207 four-parameter decay model using all replicates of a treatment. We used nonlinear least square curve fitting to test if the models  
208 were significantly different between a fungal treatment and uninoculated control, using the nls function in R.

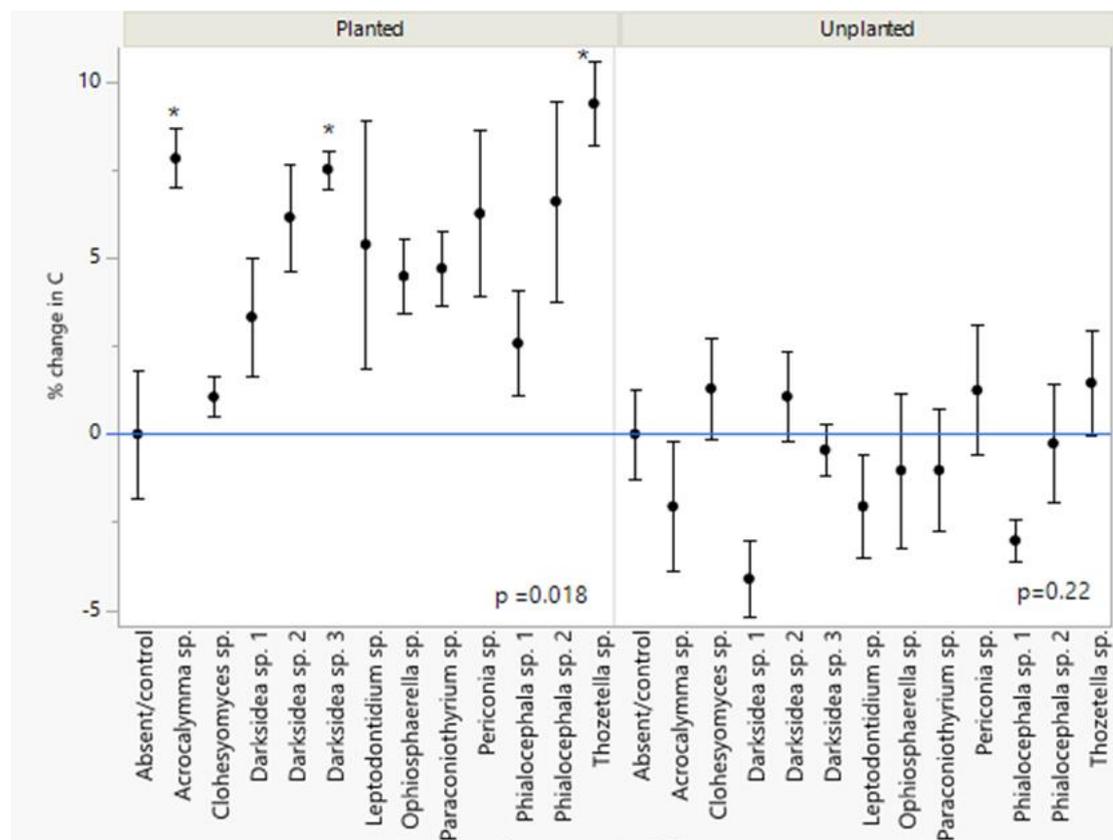
209



210 **3 Results**

211 **3.1 Several non-mycorrhizal fungal species increased soil C under wheat plants**

212 We inoculated wheat plants (*Triticum aestivum*) with one of 12 fungi (non-mycorrhizal) isolated from plant roots. After four  
 213 months of plant growth, there was a positive but varied effect of fungal inoculation on soil C content compared to the  
 214 uninoculated control group ( $p < 0.05$ ; Fig. 1, Table B2). This effect was not observed in soils that received the same fungi but  
 215 were unplanted ( $p = 0.22$ ; Fig. 1). We found significant isolate-specific increases in soil C content of the planted treatments  
 216 under inoculation with *Thozetella* sp., *Darksidea* sp. 3, and *Acrocalymma* sp., relative to the uninoculated control, of 9.4%  
 217 (percentage of change), 7.5, and 7.8, respectively. Nitrogen levels were generally higher in the soils of the inoculated and  
 218 planted treatments compared to the uninoculated control and were generally higher in the treatments where C was also higher  
 219 (Table B2).



220

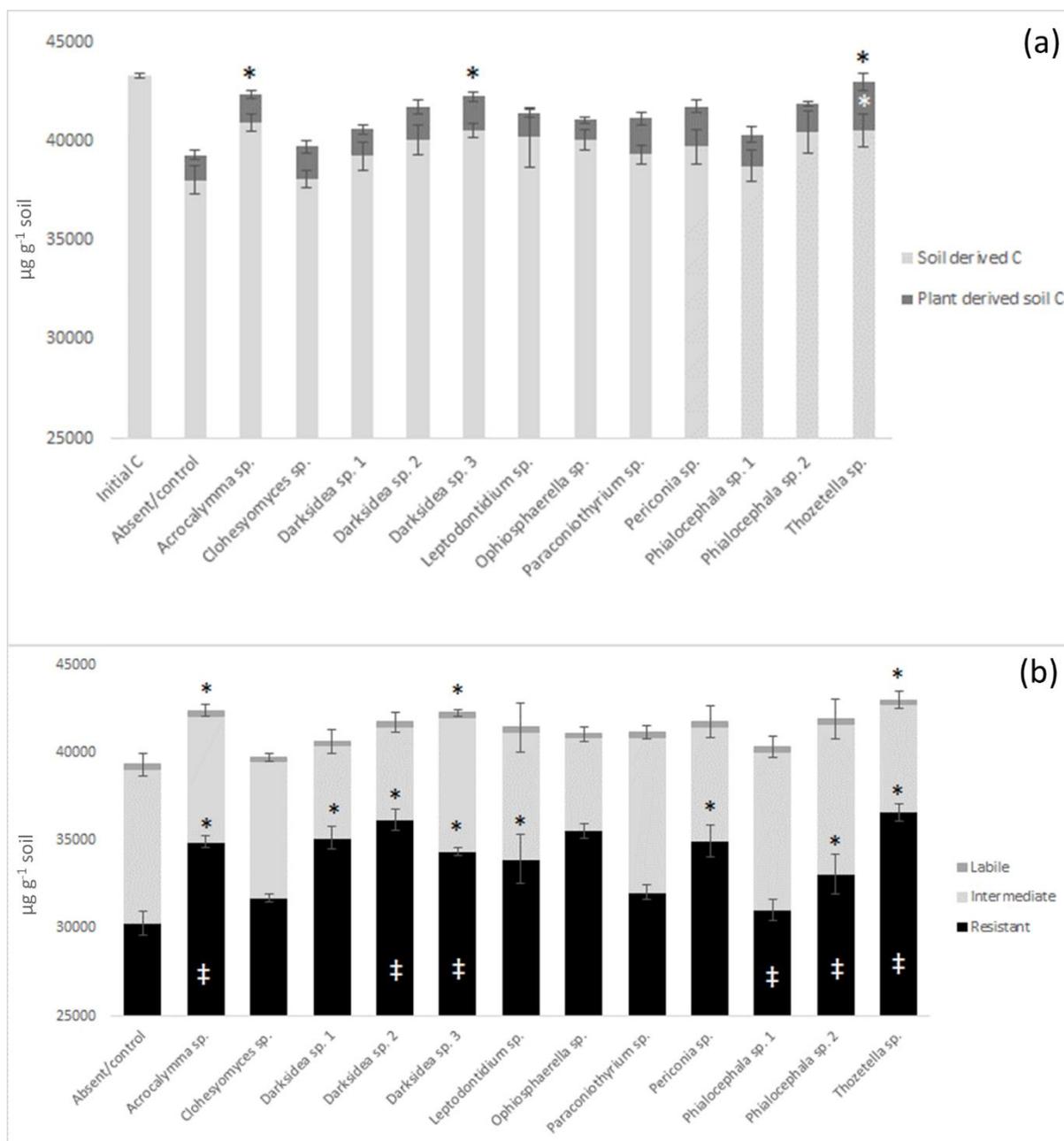
221 **Figure 1. Changes in total soil C under inoculation with different fungal isolates compared to an uninoculated control.**  
 222 **Values indicate percentage of change relative to mean of uninoculated control (blue line). Error bars indicate standard**  
 223 **error, n=7 for inoculated treatments, n=6 for control. ANOVA results for planted and unplanted are presented.**



224 Asterisks indicate significant differences with control (Dunnett test,  $p < 0.05$ ). C concentrations are presented in Table  
225 B2.

### 226 3.2 Fungi-dependent increases in soil C are associated with changes in soil C pools, origin, and persistence

227 To understand the underlying mechanisms of the fungal isolate-dependent increases in soil C content and potential shifts in  
228 sources and stability of the resulting soil C, we performed C isotope analysis, soil incubations, and soil C fractionation analysis.  
229 Isotopic partitioning of C into plant- and soil-derived C revealed how changes in these pools contributed to changes in total  
230 soil C (Fig. 2a, Table B2). Planting reduced total soil C, compared to initial C prior to planting, as expected due to C inputs  
231 stimulating decomposition (rhizosphere priming). This reduction was due to decreases in soil-derived C, which were generally  
232 not counteracted by newly added plant-derived soil C - which on average represented 3.8% ( $\pm 0.2$ ) of total soil C. Some  
233 increases in total soil C compared to the planted uninoculated controls could be explained by plant- and soil-derived C. Namely,  
234 one of the fungal treatments whereby total soil C significantly increased (*Thozetella* sp.) exhibited higher amounts of plant-  
235 derived C - at a level that was marginal in its non-significance. However, overall, the higher total soil C content relative to  
236 controls corresponded more closely with higher soil-derived C ( $R = 0.93$ ,  $p < 0.01$ ), than with plant-derived C ( $R = 0.02$ ,  $p =$   
237 0.83). All three fungal treatments resulting in significant increases in total soil C showed increases in soil-derived C but these  
238 were not statistically significant.



239

240 **Figure 2. Distribution of total soil C in plant- and soil-derived pools (A) and among labile, intermediate, and resistant**  
 241 **pools (B) in soil under inoculation with different fungal isolates or under no inoculation (Absent/control). (A): Plant-**  
 242 **and soil-derived C from C isotope partitioning (see Materials and methods). Black asterisks indicate significant**  
 243 **differences in total C with control and white asterisks differences in plant-derived soil C with control (Dunnnett test, p**  
 244 **< 0.1); (B): Pools estimated from decay models derived from soil incubation (see Materials and methods). Crosses**



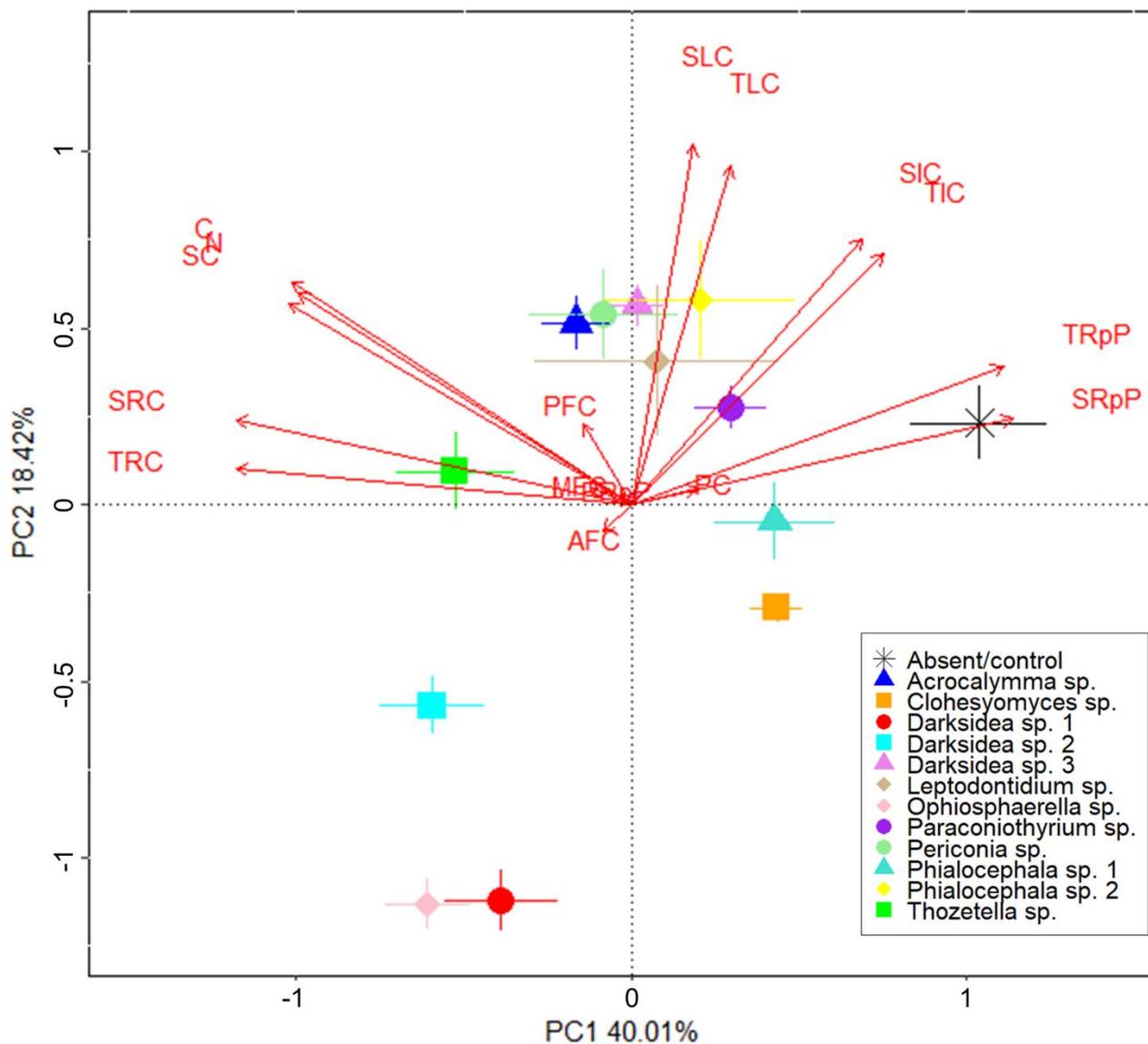
245 **indicate significant differences in the dynamics of total C decomposition (decay curves models, Table B3) compared to**  
246 **the uninoculated control. Asterisks indicate significant differences in total C or resistant C against control (Dunnett**  
247 **test,  $p < 0.05$ ). Error bars indicate standard error of total C,  $n=7$  for inoculated treatments,  $n=6$  for uninoculated**  
248 **control. Note y axis scale.**

249 Incubation of soils after plant harvest demonstrated impacts of several fungal species on the dynamics of C decomposition and  
250 the distribution of C among soil pools of different stability. The dynamics of total C decomposition (decay curves models  
251 derived from incubations) were significantly different to the control under half of the isolates (Table B3, Fig. A2). These  
252 include the three isolates that produced higher total C pools: *Thozetella* sp., *Darksidea* sp. 3, and *Acrocalymma* sp. Soil-derived  
253 C decomposition curves (from isotopic partitioning of respiration) were also significantly different to the controls under the  
254 same fungal treatments as well as *Leptodontidium* sp. Estimated pools from these decay curves showed significantly higher  
255 total resistant C (up to 86% of C), compared to controls (76% of C), under eight of the 12 isolates, including the three treatments  
256 where total C increased the most (Fig. 2b, Table B3). In terms of other pools, MRT of the total labile C was significantly lower  
257 under inoculation with *Darksidea* sp. 1 compared to the control, whereas MRT of the soil-derived labile C was significantly  
258 higher under inoculation with *Periconia* sp. (Table B3). In terms of intermediate pool MRTs, controls and fungal treatments  
259 were not significantly different.

260 Soil incubations and partitioning of respiration revealed fungal effects on the degree of persistence of total C, soil-derived C,  
261 and plant-derived C over time, which we assessed as the proportion of what was present at harvest that was respired over the  
262 full incubation. Significantly lower proportions of total and soil-derived C were respired under all fungal treatments compared  
263 to the controls ( $p < 0.001$ ; Fig. A3), indicating increased persistence. In contrast, plant-derived respired C was significantly  
264 lower (more persistent) than the controls only with *Thozetella* sp. ( $p < 0.05$ ).

265 From fractionation analysis, %C and %N of the AggC fraction, i.e. the fraction of intermediate stability whereby C is protected  
266 in aggregates, were found to have significant fungal effects, with *Thozetella* sp. and *Periconia* sp. exhibiting significantly  
267 higher levels of both C and N, and *Ophiosphaerella* sp. and *Phialocephala* sp. 1 exhibiting significantly higher levels of N  
268 compared to controls (Table B4). Significant fungal effects were not observed in the MAOM and POM fractions.

269 We performed PCA to identify soil C properties associated with fungi-driven increases in soil C (Fig. 3). Most of the variance  
270 was explained by PC1 and 2 (58%). Greater total soil C (C) was closely associated with soil-derived C (SC), but not plant-  
271 derived C (PC), at time of harvest and soil N. Soil C was also related with the resistant C pools (total (TRC) and soil-derived  
272 (SRC)). The treatments with lowest total soil C (mainly the control, followed by *Clohesyomyces* sp., and *Phialocephala* sp. 1;  
273 Fig. 1) were associated with higher proportions of total and soil-derived C respired during incubation indicating that the C  
274 remaining at harvest was inherently less persistent. %C of the AggC and MAOM fractions, considered to be more stable  
275 fractions of C, were not clearly associated with soil C or the resistant C pools, nor with any fungal treatments.



276

277 **Figure 3. Fungi-dependent increases in soil C largely relate to measures for soil C stability. Principal component**  
 278 **analysis showing soil C properties (red text) associated with various fungal isolates (symbols). Soil C properties were**  
 279 **measured via isotope analysis, soil incubations, and fractionation analysis of soil from wheat experiment. Soil C**  
 280 **property abbreviations: AFC, aggregate C fraction %C; C, %C; MFC, MAOM fraction %C; N, %N; PC, plant-**  
 281 **derived C ( $\mu\text{g g}^{-1}$  soil); PFC, POM fraction – %C; PRpP, plant-derived C respired proportion; SC, soil-derived C ( $\mu\text{g}$**   
 282  **$\text{g}^{-1}$  soil); SIC, soil-derived intermediate C ( $\mu\text{g C g}^{-1}$  soil); SLC, soil-derived labile C ( $\mu\text{g C g}^{-1}$  soil); SRC, soil-derived**

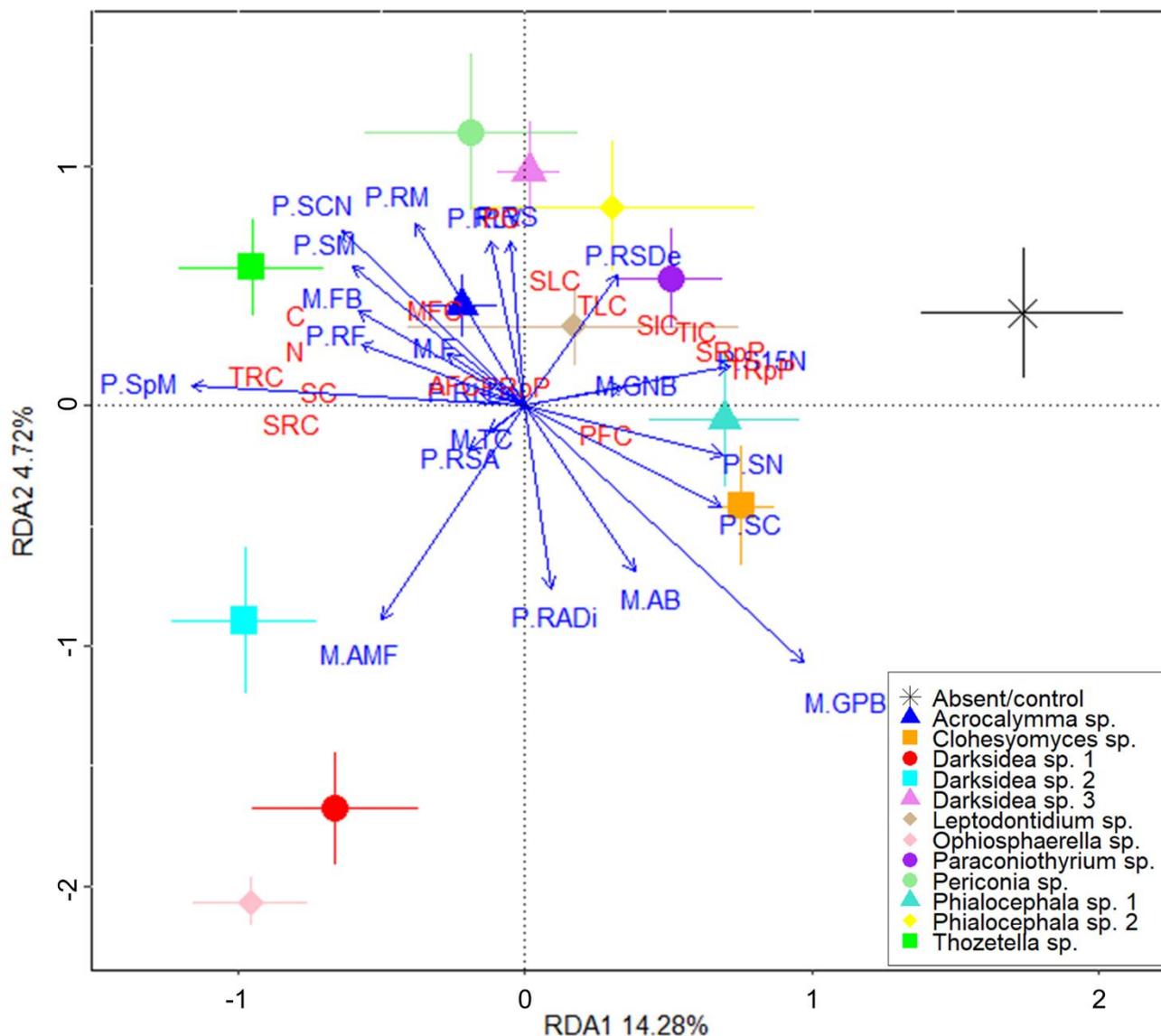


283 resistant C ( $\mu\text{g C g}^{-1}$  soil); SRpP, soil-derived C respired proportion; TIC, total intermediate C ( $\mu\text{g g}^{-1}$  soil); TLC, total  
284 labile C ( $\mu\text{g g}^{-1}$  soil); TRC, total resistant C ( $\mu\text{g g}^{-1}$  soil); TRpP, total C respired proportion.

285 **3.3 Fungi-dependent increases in soil C and its stability and persistence are positively associated with plant growth and**  
286 **microbial community composition**

287 We assessed plant and microbial community variables, including plant biomass, shoot C/N content, root morphology, and total  
288 microbial community size and composition derived from PLFA analysis. Overall, while variation among fungal isolates was  
289 observed, no significant differences were observed between the inoculated and uninoculated plants for any of the plant or  
290 microbial community variables measured, although average spike mass of *Thozetella*-inoculated plants was significantly  
291 higher than that of uninoculated plants (Table B5-6).

292 To identify plant and microbial community variables potentially involved in the fungal isolate-dependent changes in soil C  
293 properties, we performed RDA using plant and microbial community data and the soil C property data used in the PCA (Fig.  
294 4). Variance explained by RDA1 and 2 was 14.28 and 4.72%, respectively. The cluster of soil C properties that were found to  
295 be closely associated with *Thozetella* sp. in the PCA (e.g. soil-derived C, resistant C pools; Fig. 3) also trended positively with  
296 plant biomass and growth (spike and shoot mass, shoot C/N ratio, and root fork number) and with the PLFA-assessed fungal  
297 to bacterial ratio. *Acrocalymma* sp. and *Darksidea* sp. 3 were more associated with root growth traits, and were also associated  
298 with plant-derived C. The low soil C treatments (uninoculated control, *Clohesyomyces* sp., and *Phialocephala* sp. 1) and their  
299 associated soil C properties (i.e. respired C) were related to shoot C and N.



300

301 **Figure 4. Fungal treatments resulting in increased soil C and its stability are associated with plant growth. Redundancy**  
 302 **analysis showing microbial community and plant variables (blue text) driving changes in soil C properties (red text)**  
 303 **associated with various fungal isolates (symbols). Soil C properties were measured via isotope analysis, soil incubations,**  
 304 **and fractionation analysis of soil from wheat experiment. Microbial community and plant variables were measured**  
 305 **using samples harvested from the wheat experiment. Microbial community (M.) and plant (P.) variable abbreviations:**  
 306 **M.AB, actinobacteria (% of total community); M.AMF, arbuscular mycorrhizal fungi (% of total community); M.F,**  
 307 **fungi (% of total community); M.FB, fungal to bacterial biomass ratio; M.GNB, gram negative bacteria (% of total**

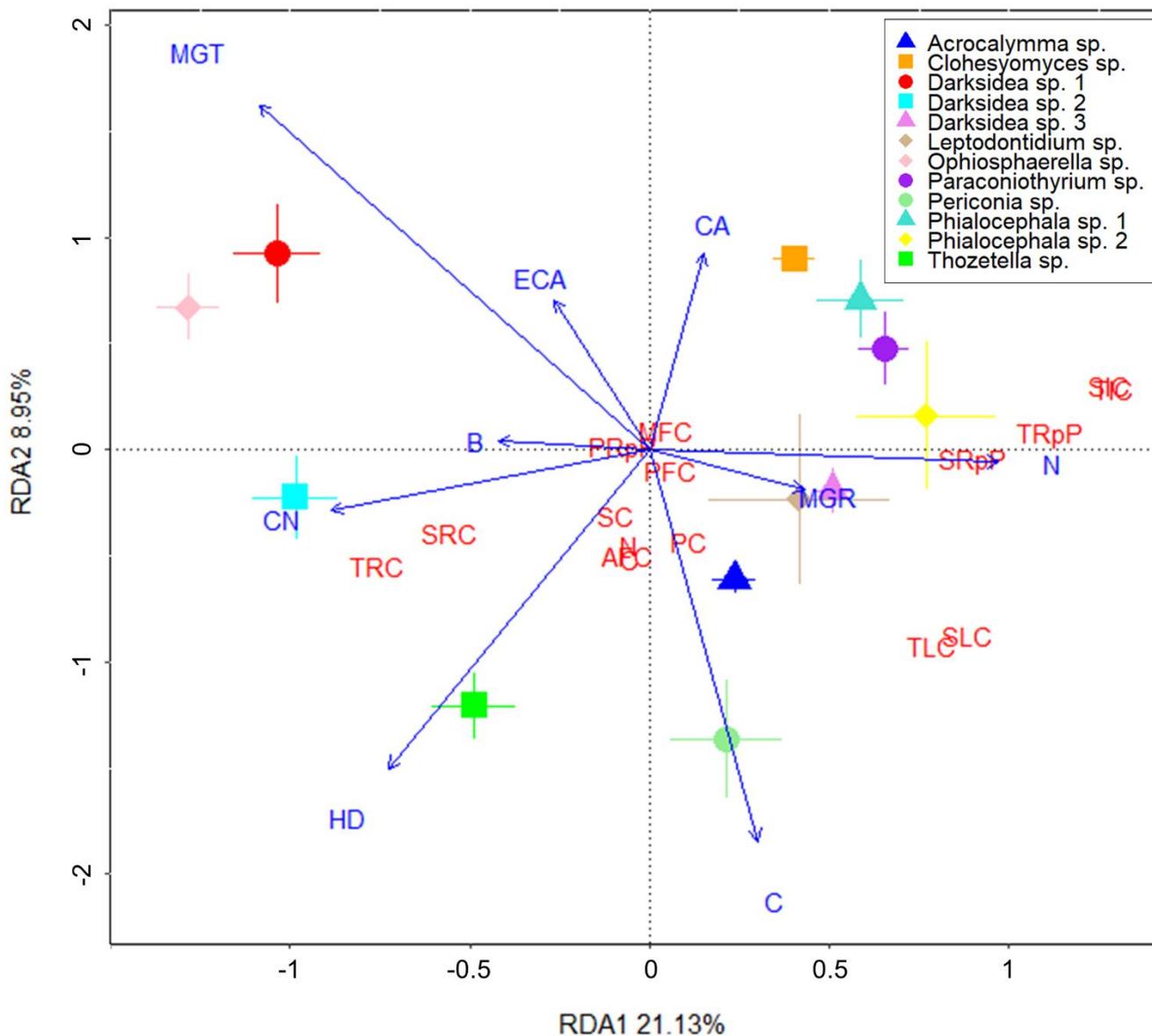


308 community); M.GPB, gram positive bacteria (% of total community); M.TC, total community size ( $\mu\text{g PLFA g}^{-1}$  soil);  
309 P.RADi, root average diameter (mm); P.RF, root fork number ( $\text{g}^{-1}$ ); P.RLDe, root length density ( $\text{cm g}^{-1}$ ); P.RLV, root  
310 length per volume ( $\text{cm m}^{-3}$ ); P.RM, root mass (g); P.RS, root/shoot ratio; P.RSA, root specific surface area ( $\text{cm}^2 \text{g}^{-1}$ );  
311 P.RSDe, root specific density ( $\text{g cm}^{-3}$ ); P.S15N, shoot  $\delta^{15}\text{N}$  (‰); P.SC, shoot %C; P.SCN, shoot C/N ratio; P.SM, shoot  
312 mass (g); P.SN, shoot %N; P.SpM, total spike mass (g). Soil C properties: AFC, aggregate C fraction – %C; C, %C;  
313 MFC, MAOM fraction – %C; N, %N; PC, plant-derived C ( $\mu\text{g g}^{-1}$  soil); PFC, POM fraction – %C; PRpP, plant-  
314 derived C respired proportion; SC, soil-derived C ( $\mu\text{g g}^{-1}$  soil); SIC, soil-derived intermediate C ( $\mu\text{g C g}^{-1}$  soil); SLC,  
315 soil-derived labile C ( $\mu\text{g C g}^{-1}$  soil); SRC, soil-derived resistant C ( $\mu\text{g C g}^{-1}$  soil); SRpP, soil-derived C respired  
316 proportion; TIC, total intermediate C ( $\mu\text{g g}^{-1}$  soil); TLC, total labile C ( $\mu\text{g g}^{-1}$  soil); TRC, total resistant C ( $\mu\text{g g}^{-1}$  soil);  
317 TRpP, total C respired proportion.

#### 318 3.4 Fungi-dependent increases in soil C and its stability and persistence are associated with denser fungal hyphae

319 Fungal isolates showed strong differentiation in all of the *in vitro*-assessed variables relating to growth and C/N content  
320 (statistically significant effects on all variables,  $p < 0.001$ ; Table B7). Biomass, colony area, and growth rate tended to be  
321 positively associated variables, and were higher in several treatments including *Acrocalymma* sp., *Darksidea* sp. 3, and  
322 *Phialocephala* sp. 1. In contrast, *Thozetella* sp. and *Clohesyomyces* sp. tended to have lower values for these variables, but  
323 *Thozetella* sp. had significantly higher hyphal density than all other treatments.

324 We performed a separate RDA to identify fungal variables potentially involved in fungi-dependent soil %C and soil C stability  
325 increases, using *in vitro* fungal assessment data and the soil C property data (Fig. 5). Compared to the RDA using plant and  
326 microbial community data (Fig. 4), greater proportions of variance were explained in this RDA by RDA1 and 2 (21.1 and 9%,  
327 respectively). Fungal colony area and hyphal density were close to opposite in their direction, with the high soil C treatment  
328 *Thozetella* sp. closely associated with hyphal density and the low soil C treatment *Clohesyomyces* sp. more associated with  
329 colony area. Similarly, fungal colony maximum growth time and rate (denoting slower and faster fungal growth, respectively)  
330 were in opposing directions. Along this axis, the high soil C treatment *Darksidea* sp. 3 was closely associated with maximum  
331 fungal growth rate. Respired C proportions were closely associated with fungal N content and were opposite resistant C  
332 fractions, which were associated with fungal C/N ratio and hyphal density.



333

334 **Figure 5. Fungal isolates involved in increased soil C and its stability have denser hyphae. Redundancy analysis (RDA)**  
 335 **showing the fungal variables (blue text) driving changes in soil C properties (red text) associated with the various fungal**  
 336 **isolates (symbols). Soil C properties were measured via isotope analysis, soil incubations, and fractionation analysis of**  
 337 **soil from wheat experiment. Fungal variables were measured in an *in vitro* plate assay and values were averaged for**  
 338 **the RDA. Fungal (F.) variable abbreviations: F.B, biomass (g); F.C, %C; F.CA, final colony area (cm<sup>2</sup>); F.CN, C/N**  
 339 **ratio; F.ECA, estimated final colony area (cm<sup>2</sup>); F.HD, hyphal density (mg cm<sup>-2</sup>); F.MGR, maximum growth rate (cm<sup>2</sup>**  
 340 **day); F.MGT, time to maximum growth (days); F.N, %N. Soil C properties: AFC, aggregate C fraction – %C; C,**



341 %C; MFC, MAOM fraction – %C; N, %N; PC, plant-derived C ( $\mu\text{g g}^{-1}$  soil); PFC, POM fraction – %C; PRpP, plant-  
342 derived C respired proportion; SC, soil-derived C ( $\mu\text{g g}^{-1}$  soil); SIC, soil-derived intermediate C ( $\mu\text{g C g}^{-1}$  soil); SLC,  
343 soil-derived labile C ( $\mu\text{g C g}^{-1}$  soil); SRC, soil-derived resistant C ( $\mu\text{g C g}^{-1}$  soil); SRpP, soil-derived C respired  
344 proportion; TIC, total intermediate C ( $\mu\text{g g}^{-1}$  soil); TLC, total labile C ( $\mu\text{g g}^{-1}$  soil); TRC, total resistant C ( $\mu\text{g g}^{-1}$  soil);  
345 TRpP, total C respired proportion.



#### 346 **4 Discussion**

347 Discussions on soil C sequestration as a climate change strategy have largely focused on one side of the soil C storage system  
348 - increasing C inputs into soil (promoting soil C formation). However, due to the complex and dynamic nature of soil C,  
349 reductions of soil C outputs (or, increases in soil C stability and persistence) must also be attained in order to foster soil C  
350 storage. In this study, we drew our attention to fungi that have potential in improving soil C storage but that are often  
351 overlooked in this area of research, using a high resolution, multifaceted approach combining isotopic labelling, soil  
352 incubations, and soil fractionation analysis, as well as an *in vitro* study in parallel. Our study supports the notion that non-  
353 mycorrhizal root-associated fungi can improve soil C storage via multiple direct and indirect mechanisms determining C inputs  
354 and stabilisation. Mechanisms that increased the stability of existing C were more common across the diverse fungal treatments  
355 than those increasing the input of new C.

356 Despite our finding that bulk soil C increased significantly under only three fungal treatments, in support of our hypothesis  
357 our incubations revealed significant increases in directly and functionally assessed soil C stability (i.e. increases in resistant  
358 pools and decreases in respired C during incubation) under most of the fungal treatments, with the stabilised C being original  
359 soil C, not new inputs of C. Thus, as well as contributing to evidence that fungi can lead to increased soil C content (e.g.  
360 Kallenbach et al., 2016), our study provides direct evidence from plant-fungi soil systems for non-mycorrhizal fungi-driven  
361 improvements to soil C storage primarily via enhanced stability of soil C. This is emphasised by our findings that the treatments  
362 whereby soil C content was the lowest (control, *Clohesyomyces* sp., and *Phialocephala* sp. 1) were associated with higher  
363 proportions of total and soil-derived C respired during incubation - indicating that the C remaining at harvest under these  
364 treatments was inherently more prone to decomposition (i.e. less persistent). Increased stability and persistence of soil C  
365 primarily results from inhibition of microbial decomposition (Cotrufo and Lavalley, 2022), which can occur by a variety of  
366 reasons including reduced saprotrophic activity due to microbes being outcompeted for nutrients (Boer et al., 2005), increased  
367 input of fungal, more readily stabilised C (Sokol et al., 2019), and increased soil aggregation (Lehmann et al., 2020). We  
368 investigated multiple potential mediators for the observed increases in soil C stability and persistence in our study and found  
369 some leads. We found that increased fungal C/N ratio and hyphal density may be important for stability and persistence of soil  
370 C (while fungal N corresponded with decreased stability and persistence). Fungi with denser hyphae can promote soil  
371 aggregation, as soil particles get more entangled and stabilised in dense hyphae (Dignac et al., 2017). Our study substantiates  
372 previous assertions that fungal trait expression is relevant to soil C stability: fungi that exhibited an exploitative growth strategy  
373 (denser hyphae) were found to more closely associated with soil C stability and persistence, while fungi that exhibited a more  
374 exploratory strategy (faster growth) were positively associated with respired C and less stable C pools (Camenzind et al., 2020;  
375 Fernandez et al., 2019; Fernandez and Koide, 2013; Jackson et al., 2017; Lehmann et al., 2020; Schmidt et al., 2011; Zanne et  
376 al., 2020). These findings support the notion that an exploitative growth strategy may be more conducive to competition with  
377 saprotrophs for nutrients, leading to reduced decomposition (Bödeker et al., 2016).



378 Our PLFA-assessed finding regarding fungal to bacterial ratio points towards a second likely mechanism for the increases in  
379 soil C stability – increased input of fungal C, which becomes necromass. Fungal necromass is a significant source of soil C  
380 inputs, and can in some cases make up the majority of SOM (Wang et al., 2021). Substrates with high C/N ratios, such as  
381 fungal biomass or necromass, are generally associated with reduced decomposition rates, although C/N ratio is not the sole  
382 determinant of substrate decomposition and C/N ratios can in fact be altered by, rather than alter the activity of, soil microbial  
383 communities (Marañón-Jiménez et al., 2021; Smith and Wan, 2019; Schneckner et al., 2019). Compared with other substrates,  
384 however, necromass is a particularly stabilisable form of C as it can bind to the surfaces of MAOM or be stabilised on  
385 aggregates, where it is physically protected from decomposition (Sokol et al., 2019). For these reasons, we expected to see  
386 positive associations between soil C stability and aggregate and MAOM soil fractions, which are considered to signify  
387 increased and longer-term stability and persistence (Dynarski et al., 2020; Hemingway et al., 2019; Islam et al., 2022; Poeplau  
388 et al., 2018; Poeplau et al., 2017). However, in our study these fractions were not strongly associated with soil C content, its  
389 distribution in pools (stability) or persistence, nor were they as influential on differences between fungal treatments. While  
390 this lends support to the notion that microbial decomposition of soil C was metabolically inhibited (as discussed above), rather  
391 than physically limited, our findings may be explained to some extent by methodology. A potential explanation for our findings  
392 is that although fungal necromass may have been abundant, the experimental conditions may have been unresponsive of  
393 MAOM formation (e.g. the high C content of the unplanted soil may have meant that MAOM content was already at saturation  
394 level and new MAOM was not able to form). Other potential explanations are that the MAOM fraction could possibly take  
395 longer than the experimental timeframe to change substantially, or that the MAOM estimation method may carry greater error,  
396 thus making detection of responses more difficult. Nonetheless, our study detected increases in total C, and C resistance and  
397 persistence that were not associated with MAOM, suggesting that soil fractionation analyses do not entirely accurately reflect  
398 natural soil C distribution and stability which can be detected functionally via soil incubations. Further studies utilising the  
399 combined approach of soil incubations and soil fractionation analysis, such as studies using soil with lower C content or studies  
400 over a longer time period, may shed light on how findings from the two methods can be compared. However, our findings call  
401 for caution in directly equating operationally defined MAOM pools and their size with C stability and suggest that functionally  
402 assessing C dynamics may be more effective in some cases.

403 In terms of improvements to soil C content, of the three fungal treatments whereby soil C increases were significant, these  
404 were accompanied by increases in plant-derived C only under inoculation with *Thozetella* sp. While we expected that there  
405 would be some variation in the fungal impacts on soil C storage due to the diversity amongst the fungi included in this study,  
406 this finding is in contrast to our expectation that increases in plant-derived C would be the main mechanism involved in C  
407 increase. As plant growth promotion and changes in nutrient uptake is a well-known characteristic of some fungi (Hossain et  
408 al., 2017), the increase in plant-derived C with *Thozetella* sp. may have been related to the increases in quantity or quality of  
409 plant inputs related to the shifts in plant variables of *Thozetella* sp. (spike mass, shoot biomass, and shoot C/N ratio). Our  
410 results from the isotopic partitioning of respiration from soil incubations further indicate that the plant-derived C present in



411 soil and that contributed to total soil C increase under inoculation with *Thozetella* sp. was more persistent compared to the  
412 control or other treatments. Fungal-derived C could also have contributed to size and persistence of plant-derived C, if the  
413 fungi took up plant-derived C. Thus, in addition to increasing plant inputs, *Thozetella* sp. appears to have been more active in  
414 stabilising those inputs via the mechanisms discussed above.

415 Our study addresses key knowledge gaps in the ways fungi affect soil C storage. We have explicitly demonstrated that  
416 inoculation with non-mycorrhizal fungi can improve soil C content and, moreover, soil C stability - supporting the general  
417 agreement in this field that microbial transformations of soil C and microbially driven changes to soil structure are as important,  
418 if not more important, than the characteristics of the inputs themselves for soil C storage (Dynarski et al., 2020; Hannula and  
419 Morriën, 2022). When it comes to evaluating the potential of fungi to support soil C storage, our findings indicate that it is  
420 important to consider not only increases in soil C but also their impact on the stability of C. Among the diverse fungi studied,  
421 these improvements largely resulted from reductions in C outputs by increasing stable C pools and resistance of existing soil  
422 C to decomposition. While potential mechanisms behind these improvements depended on fungal identity, our study points  
423 towards metabolic inhibition (rather than physical limitation) of microbial decomposition for which growth characteristics  
424 such as density of fungal hyphae and fungal C/N ratio may be important indicators – thus, fungal trait expression may be a  
425 proxy for fungal influences on soil C storage. However, more work is needed to test whether or not physical limitation of  
426 microbial decomposition leads to enhanced soil C stability by these fungi. More rarely, the improvements to soil C storage  
427 involved the effects of fungal inoculation on host plant growth and C inputs. While total soil C content increased significantly  
428 only under a minority of fungal treatments, the significant fungi-driven increases in stability we observed could potentially  
429 lead to even greater increases in soil C content over time - however experiments with longer timeframes are needed to test this  
430 idea. This study and continued work will advance knowledge of these mechanisms and support the search and potential  
431 implementation of root-associated fungi to improve soil C storage, which will aid soil C sequestration strategies.

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445 **Appendices**

446

447 **Appendix A**

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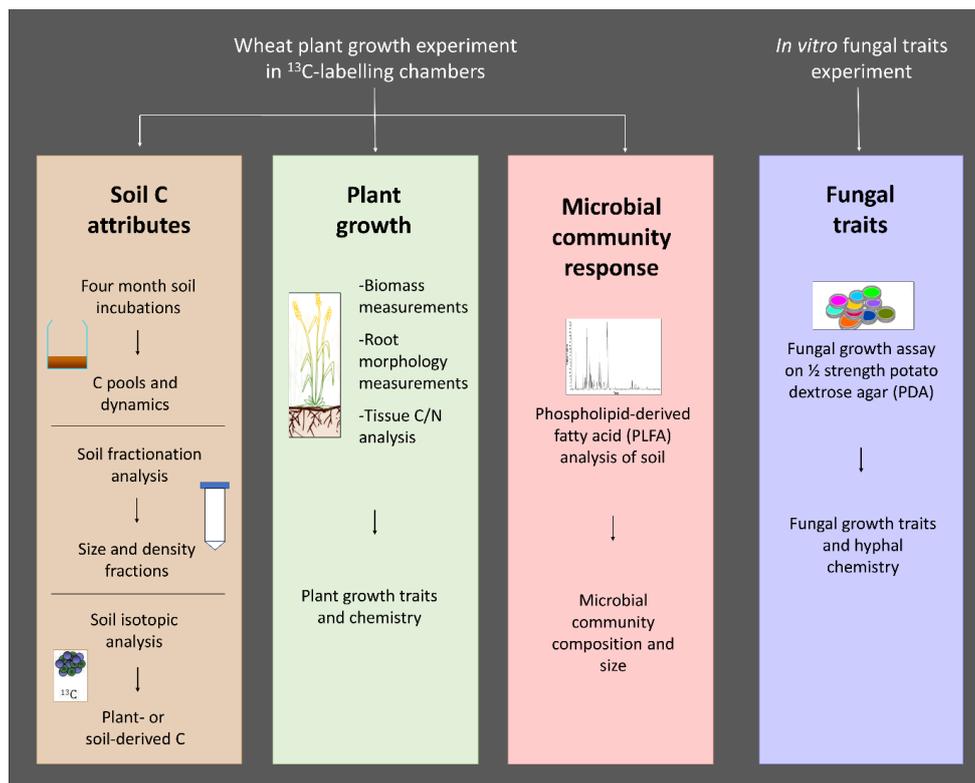


Figure A1. Overview of the study design, measured traits, and methodology used. C, carbon, N, nitrogen.

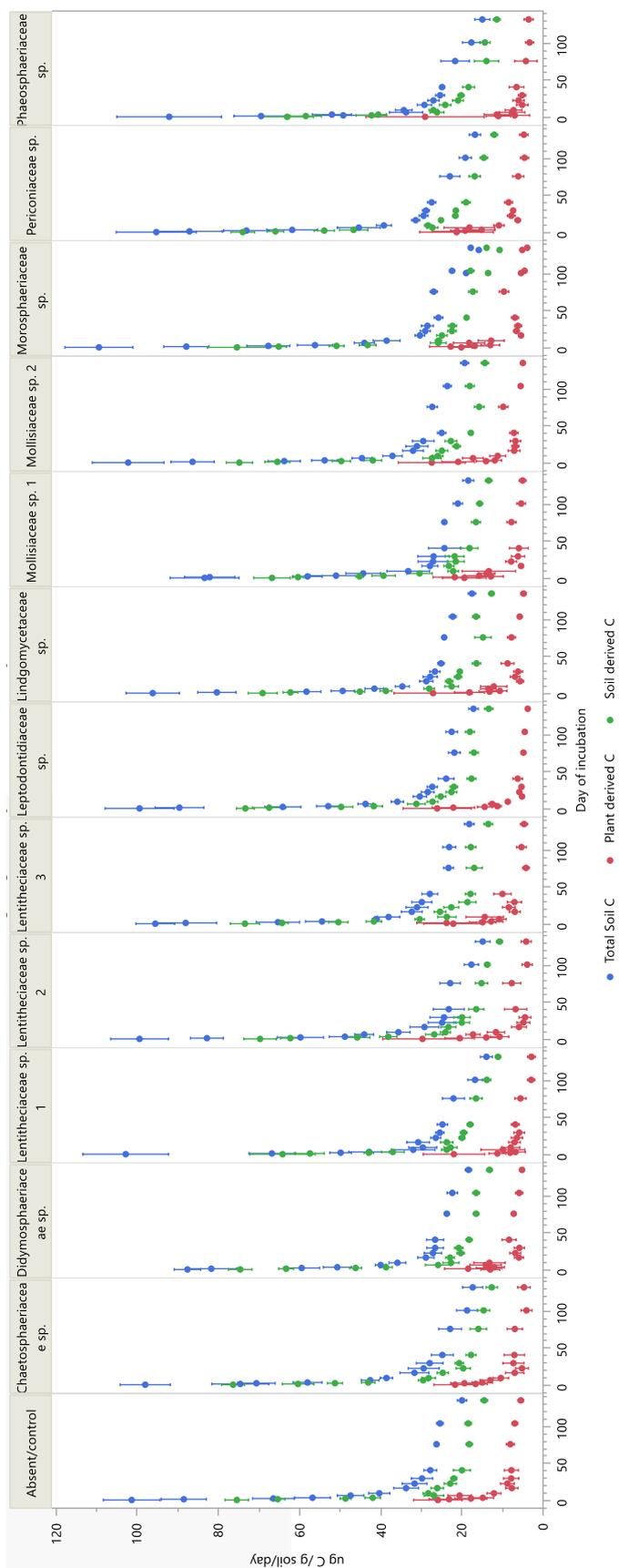


Figure A2. Total soil respiration and its soil- and plant-derived components during laboratory soil incubations of soils collected after plant growth with inoculation of 12 fungal species and a control (Absent/control). Data points are means (n=7 for inoculated pots; n=6 for controls). Soil and plant components calculated from isotopic partitioning based on planted and unplanted soil  $\delta^{13}\text{C}$ . Error bars are standard error.

Family (Genus): Chaetosphaeriaceae sp. (*Thozetella* sp.); Didymosphaeriaceae sp. (*Paraconiothyrium* sp.); Lentitheciaceae sp. 1 (*Darksidea* sp. 1); Lentitheciaceae sp. 2 (*Darksidea* sp. 2); Lentitheciaceae sp. 3 (*Darksidea* sp. 3); Leptodontidiaceae sp. (*Leptodontidium* sp.); Lindgomycetaceae sp. (*Chloesyomyces* sp.); Mollisiaceae sp. 1 (*Phialocephala* sp. 1); Mollisiaceae sp. 2 (*Phialocephala* sp. 2); Morosphaeriaceae sp. (*Acrocalymma* sp.); Periconiaceae sp. (*Periconia* sp.); Phaeosphaeriaceae sp. (*Ophiosphaerella* sp.)

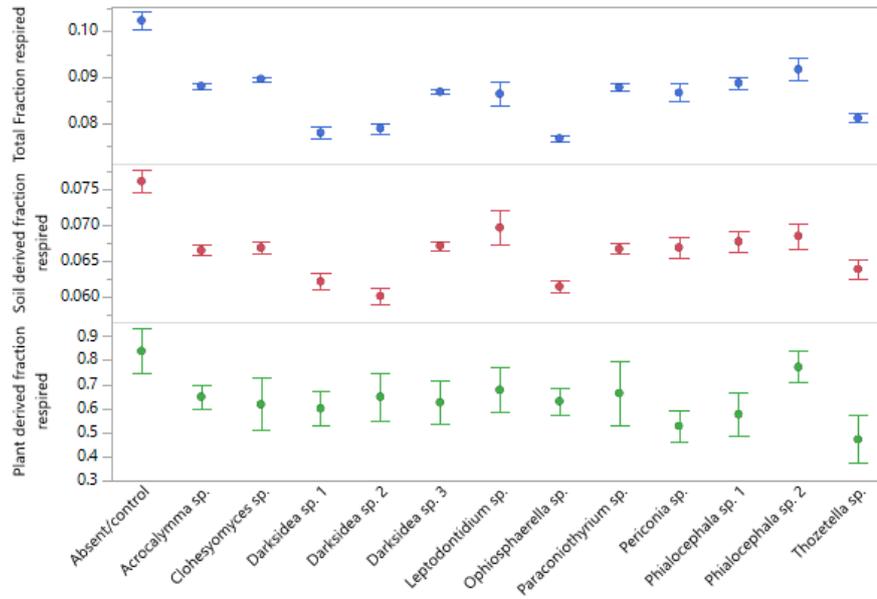


Figure A3. Fraction of soil carbon (C) respired over the course of 135-day incubation of soils under wheat and 12 types of fungal inoculum. Total C is all C respired, and soil- and plant-derived C were obtained from isotopic partitioning of respiration over time (See Materials and methods). Values are means of n=7 for treatments and n=6 for control. Error bars are standard error.

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449 **Appendix B**

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**Table B1.** Chemical and physical analysis of pre-planted soil used in wheat experiment. Analysis was performed by Environmental Analysis Laboratory (East Lismore, Australia).

Parameter	Units	Value
Phosphorus	mg kg <sup>-1</sup>	151
pH		5.85
Electrical conductivity	dS m <sup>-1</sup>	0.232
Estimated organic matter	% OM	7.5
	cmol kg <sup>-1</sup>	8.9
Exchangeable calcium	kg ha <sup>-1</sup>	4010
	mg kg <sup>-1</sup>	1790
	cmol kg <sup>-1</sup>	2.9
Exchangeable magnesium	kg ha <sup>-1</sup>	795
	mg kg <sup>-1</sup>	355
	cmol kg <sup>-1</sup>	3.1
Exchangeable potassium	kg ha <sup>-1</sup>	2719
	mg kg <sup>-1</sup>	1214
	cmol kg <sup>-1</sup>	0.32
Exchangeable sodium	kg ha <sup>-1</sup>	164
	mg kg <sup>-1</sup>	73
	cmol kg <sup>-1</sup>	0.02
Exchangeable aluminium	kg ha <sup>-1</sup>	3.1
	mg kg <sup>-1</sup>	1.4
	cmol kg <sup>-1</sup>	0.06
Exchangeable hydrogen	kg ha <sup>-1</sup>	1.2
	mg kg <sup>-1</sup>	<1
Effective cation exchange capacity	cmol kg <sup>-1</sup>	15
Calcium	%	58
Magnesium	%	19
Potassium	%	20
Exchangeable sodium	%	2.1
Aluminium	%	0.1
Hydrogen	%	0.36
Calcium/magnesium ratio		3.1
Total carbon	%	4.3
Total nitrogen	%	0.39
Carbon/nitrogen ratio		11
Basic texture		Clay loam
Basic colour		Brownish
Chloride estimate	(equiv. mg kg <sup>-1</sup> )	148



**Table B2.** Properties of soil in which inoculated wheat plants were grown for four months. P-values from ANOVA are displayed in the bottom row. Asterisks/dots in other rows (if present) indicate significant differences to uninoculated controls as determined via Dunnett's post-hoc test ( $p < 0.1$ , \*  $p < 0.05$ , \*\*  $p < 0.01$ , \*\*\*  $p < 0.001$ ). C, carbon, N, nitrogen.

Treatment	%C	%N	$\delta^{13}\text{C}$ (‰)	$\delta^{15}\text{N}$ (‰)	Plant-derived C (µg/g soil)	Soil-derived C (µg/g soil)
Absent/control	3.93 ± 0.07	0.36 ± 0.01	-25.31 ± 0.03	9.72 ± 0.04	1279.03 ± 247.66	38060.63 ± 712.28
<i>Acrocalymma</i> sp.	4.24 ± 0.03 *	0.39 ± 0.003 **	-25.33 ± 0.02	9.65 ± 0.01	1448.55 ± 188.76	40966.09 ± 416.19
<i>Clothesyromyces</i> sp.	3.98 ± 0.02	0.36 ± 0.003	-25.33 ± 0.03	9.58 ± 0.03	1611.13 ± 319.08	38142.72 ± 394.1
<i>Darksidea</i> sp. 1	4.07 ± 0.06	0.37 ± 0.004	-25.32 ± 0.03	9.61 ± 0.06	1364.06 ± 220.06	39281.97 ± 668.04
<i>Darksidea</i> sp. 2	4.18 ± 0.06	0.38 ± 0.004	-25.35 ± 0.03	9.62 ± 0.03	1635.09 ± 320.66	40122.22 ± 683.05
<i>Darksidea</i> sp. 3	4.23 ± 0.02 *	0.38 ± 0.003 *	-25.37 ± 0.02	9.69 ± 0.02	1747.74 ± 243.68	40544.37 ± 332.86
<i>Leptodontidium</i> sp.	4.15 ± 0.13	0.38 ± 0.01	-25.34 ± 0.04	9.72 ± 0.03	1208.67 ± 207.32	40246.15 ± 1395.36
<i>Ophiosphaerella</i> sp.	4.11 ± 0.04	0.38 ± 0.003	-25.29 ± 0.04	9.82 ± 0.03	1004.45 ± 142.31	40094.79 ± 501.62
<i>Paraconiothyrium</i> sp.	4.12 ± 0.04	0.38 ± 0.004	-25.39 ± 0.03	9.72 ± 0.03	1830.47 ± 282.22	39356.27 ± 415.96
<i>Periconia</i> sp.	4.18 ± 0.09	0.38 ± 0.01	-25.44 ± 0.04	9.75 ± 0.05	2038.42 ± 288.09	39760.5 ± 820.79
<i>Phialocephala</i> sp. 1	4.04 ± 0.05	0.37 ± 0.01	-25.36 ± 0.05	9.81 ± 0.03	1582.66 ± 368.69	38769.63 ± 739.07
<i>Phialocephala</i> sp. 2	4.19 ± 0.10	0.38 ± 0.01 *	-25.35 ± 0.02	9.71 ± 0.03	1422.66 ± 130.89	40511.25 ± 998.06
<i>Thozzeria</i> sp.	4.30 ± 0.04 **	0.39 ± 0.01 **	-25.47 ± 0.04 *	9.69 ± 0.03	2434.52 ± 418.15	40592.71 ± 756.54
<b>p-value (ANOVA)</b>	0.02 *	0.01 *	0.03 *	<0.001 ***	0.06	0.15



**Table B3.** Model fit, model comparisons, pool sizes (resistant, intermediate, and labile) and pool mean residence times (labile and intermediate) estimated from four parameter exponential decay models fitted to CO<sub>2</sub> released over 135-day incubations of soil under wheat and fungal inocula. Total C is C in all CO<sub>2</sub> released, soil-derived C is C from non-plant origin calculated through isotopic partitioning of CO<sub>2</sub> based on plant and CO<sub>2</sub> δ<sup>13</sup>C. Asterisks indicate significant difference with uninoculated controls (. p < 0.1, \* p < 0.05, \*\* p < 0.01, \*\*\* p < 0.001). Crosses indicate variables for which no statistical test was possible as they were estimated from average curves per treatment. For details of parameter estimation and isotopic partitioning see Materials and methods. C, carbon, MRT, mean residence time.

Treatment	Model R <sup>2</sup>	Decomposition dynamic p-value (comparison with absent/control group)	Resistant C (µg/g soil)	Intermediate C (µg/g soil)†	Intermediate C MRT (days)	Labile C (µg/g soil)†	Labile C MRT (days)
<b>Total C</b>							
Absent/control	0.89	NA	30276 ± 655	8777.69	247 ± 74	285.57	3.07 ± 0.40
<i>Acrocalymma</i> sp.	0.89	< 0.001 ***	34923 ± 304 ***	7195.55	210 ± 67	295.37	2.70 ± 0.33
<i>Clohesyomyces</i> sp.	0.91	ns	31704 ± 206	7797.19	246 ± 67	252.13	2.63 ± 0.28
<i>Darksidea</i> sp. 1	0.84	ns	35164 ± 613 ***	5275.69	164 ± 51	206.06	1.51 ± 0.22 **
<i>Darksidea</i> sp. 2	0.88	< 0.001 ***	36182 ± 556 ***	5322.69	160 ± 44	252.16	2.51 ± 0.37
<i>Darksidea</i> sp. 3	0.87	< 0.01 **	34398 ± 195 **	7620.96	222 ± 65	272.88	3.01 ± 0.42
<i>Leptodontidium</i> sp.	0.89	ns	33941 ± 1285 **	7216.05	227 ± 69	297.45	3.04 ± 0.37
<i>Ophiosphaerella</i> sp.	0.79	ns	35583 ± 380 ***	5317.96	161 ± 60	198.12	2.09 ± 0.45
<i>Paraconiothyrium</i> sp.	0.89	ns	32053 ± 379	8866.63	291 ± 97	266.34	3.25 ± 0.41
<i>Periconia</i> sp.	0.87	ns	34970 ± 859 ***	6485.94	196 ± 77	342.66	4.17 ± 0.81
<i>Phialocephala</i> sp. 1	0.79	< 0.001 ***	31058 ± 540	9011.62	309 ± 193	282.05	3.76 ± 0.77
<i>Phialocephala</i> sp. 2	0.88	< 0.01 **	33098 ± 1041.	8563.14	249 ± 79	271.87	2.73 ± 0.35
<i>Thozetella</i> sp.	0.86	< 0.001 ***	36615 ± 439 ***	6127.71	182 ± 54	284.05	3.41 ± 0.53
<b>Soil-derived C</b>							
Absent/control	0.95	NA	31337 ± 712	6517.67	258 ± 55	205.43	2.70 ± 0.22
<i>Acrocalymma</i> sp.	0.9	< 0.001 ***	35086 ± 416 *	5660.13	234 ± 77	219.30	2.90 ± 0.34
<i>Clohesyomyces</i> sp.	0.94	ns	32351 ± 394	5586.36	252 ± 60	205.31	2.99 ± 0.25
<i>Darksidea</i> sp. 1	0.85	ns	34436 ± 668.	4669.97	206 ± 75	175.08	2.78 ± 0.43
<i>Darksidea</i> sp. 2	0.92	< 0.001 ***	35757 ± 683 **	4165.06	181 ± 45	199.37	2.86 ± 0.33
<i>Darksidea</i> sp. 3	0.93	< 0.001 ***	33927 ± 332	6389.46	277 ± 78	227.75	3.18 ± 0.30
<i>Leptodontidium</i> sp.	0.92	< 0.001 ***	34232 ± 1395	5791.95	235 ± 58	221.83	3.13 ± 0.32
<i>Ophiosphaerella</i> sp.	0.87	ns	35804 ± 501 **	4113.89	169 ± 52	175.91	3.10 ± 0.56
<i>Paraconiothyrium</i> sp.	0.95	ns	32887 ± 415	6258.33	281 ± 64	209.99	2.64 ± 0.19
<i>Periconia</i> sp.	0.96	ns	34874 ± 820 *	4644.09	187 ± 37	242.11	3.58 ± 0.34 *
<i>Phialocephala</i> sp. 1	0.91	< 0.001 ***	32988 ± 739	5584.94	241 ± 74	196.62	3.14 ± 0.38
<i>Phialocephala</i> sp. 2	0.93	< 0.001 ***	33891 ± 998	6399.73	270 ± 72	220.25	2.94 ± 0.27
<i>Thozetella</i> sp.	0.94	< 0.001 ***	35864 ± 756 **	4509.96	184 ± 37	217.77	3.05 ± 0.29



**Table B4.** Properties of carbon fractions of soil in which inoculated wheat plants were grown for four months. Properties were measured using soil fractionation analysis. P-values from ANOVA are displayed in the bottom row. Asterisks/dots in other rows (if present) indicate significant differences to uninoculated controls as determined via Dunnett's post-hoc test ( $p < 0.1$ , \*  $p < 0.05$ , \*\*  $p < 0.01$ , \*\*\*  $p < 0.001$ ). C, carbon, N, nitrogen, AggC, aggregate carbon, MAOM, mineral-associated organic matter, POM, particulate organic matter.

Treatment	AggC fraction – %C	AggC fraction – %N	MAOM fraction – %C	MAOM fraction – %N	POM fraction – %C	POM fraction – %N
Absent/control	1.96 ± 0.05	0.16 ± 0.01	0.57 ± 0.02	0.05 ± 0.002	0.92 ± 0.07	0.06 ± 0.01
<i>Acrocalymma</i> sp.	2.18 ± 0.10	0.18 ± 0.01	0.48 ± 0.02	0.04 ± 0.001	0.98 ± 0.05	0.07 ± 0.004
<i>Clohesomyces</i> sp.	2.14 ± 0.07	0.18 ± 0.01	0.51 ± 0.02	0.05 ± 0.002	0.94 ± 0.05	0.06 ± 0.003
<i>Darksidea</i> sp. 1	2.09 ± 0.06	0.17 ± 0.01	0.58 ± 0.04	0.05 ± 0.003	0.87 ± 0.04	0.06 ± 0.003
<i>Darksidea</i> sp. 2	2.13 ± 0.03	0.17 ± 0.002	0.54 ± 0.05	0.05 ± 0.004	0.89 ± 0.03	0.06 ± 0.002
<i>Darksidea</i> sp. 3	2.13 ± 0.05	0.17 ± 0.004	0.60 ± 0.02	0.05 ± 0.002	1.00 ± 0.06	0.07 ± 0.004
<i>Leptodontidium</i> sp.	2.12 ± 0.07	0.17 ± 0.01	0.53 ± 0.02	0.05 ± 0.002	0.98 ± 0.04	0.06 ± 0.003
<i>Ophiosphaerella</i> sp.	2.18 ± 0.04	0.19 ± 0.004 *	0.55 ± 0.03	0.05 ± 0.003	0.96 ± 0.04	0.07 ± 0.003
<i>Paraconiothyrium</i> sp.	2.15 ± 0.05	0.18 ± 0.004	0.56 ± 0.03	0.05 ± 0.002	1.00 ± 0.06	0.07 ± 0.01
<i>Periconia</i> sp.	2.25 ± 0.06 *	0.19 ± 0.01 *	0.55 ± 0.05	0.05 ± 0.004	0.89 ± 0.03	0.06 ± 0.002
<i>Phialocephala</i> sp. 1	2.22 ± 0.06	0.19 ± 0.01 **	0.53 ± 0.02	0.05 ± 0.002	0.86 ± 0.09	0.06 ± 0.01
<i>Phialocephala</i> sp. 2	2.09 ± 0.07	0.17 ± 0.01	0.56 ± 0.03	0.05 ± 0.003	0.86 ± 0.03	0.06 ± 0.002
<i>Thozetella</i> sp.	2.37 ± 0.07 ***	0.20 ± 0.01 ***	0.52 ± 0.04	0.05 ± 0.003	0.91 ± 0.10	0.06 ± 0.01
<b>P-value (ANOVA)</b>	0.03 *	0.002 **	0.63	0.62	0.65	0.41





**Table B6.** Microbial community variables potentially influencing soil (in which inoculated wheat plants were grown for four months). P-values from ANOVA are displayed in the bottom row. Asterisks/dots in other rows (if present) indicate significant differences to uninoculated controls as determined via Dunnett's post-hoc test (.  $p < 0.1$ , \*  $p < 0.05$ , \*\*  $p < 0.01$ , \*\*\*  $p < 0.001$ ).

Treatment	Total community size ( $\mu\text{g PLFA/g soil}$ )	Fungal to bacterial biomass ratio	Gram positive bacteria (% of total community)	Gram negative bacteria (% of total community)	Actinobacteria (% of total community)	Fungi (% of total community)	Arbuscular mycorrhizal fungi (% of total community)
Absent/control	8.30 $\pm$ 0.33	0.22 $\pm$ 0.01	19.50 $\pm$ 0.01	26.19 $\pm$ 0.55	8.20 $\pm$ 0.14	10.19 $\pm$ 0.47	2.41 $\pm$ 0.09
<i>Acrocalymma</i> sp.	8.59 $\pm$ 0.57	0.23 $\pm$ 0.01	19.88 $\pm$ 0.01	26.10 $\pm$ 0.72	7.68 $\pm$ 0.74	10.44 $\pm$ 0.42	2.45 $\pm$ 0.07
<i>Clothesomyces</i> sp.	8.35 $\pm$ 0.28	0.22 $\pm$ 0.01	20.38 $\pm$ 0.01	26.48 $\pm$ 0.48	8.48 $\pm$ 0.14	10.11 $\pm$ 0.28	2.52 $\pm$ 0.07
<i>Darksidea</i> sp. 1	8.54 $\pm$ 0.30	0.22 $\pm$ 0.01	20.14 $\pm$ 0.01	26.06 $\pm$ 0.61	8.37 $\pm$ 0.11	9.98 $\pm$ 0.26	2.63 $\pm$ 0.10
<i>Darksidea</i> sp. 2	7.72 $\pm$ 0.32	0.21 $\pm$ 0.01	20.10 $\pm$ 0.01	26.59 $\pm$ 0.47	8.23 $\pm$ 0.16	9.79 $\pm$ 0.32	2.71 $\pm$ 0.12
<i>Darksidea</i> sp. 3	7.50 $\pm$ 0.71	0.22 $\pm$ 0.01	19.03 $\pm$ 0.01	25.32 $\pm$ 0.40	7.90 $\pm$ 0.08	9.54 $\pm$ 0.34	2.41 $\pm$ 0.08
<i>Leptodontidium</i> sp.	7.89 $\pm$ 0.51	0.23 $\pm$ 0.01	20.01 $\pm$ 0.01	26.02 $\pm$ 0.57	8.16 $\pm$ 0.20	10.36 $\pm$ 0.41	2.62 $\pm$ 0.07
<i>Ophiosphaerella</i> sp.	8.61 $\pm$ 0.21	0.24 $\pm$ 0.01	19.28 $\pm$ 0.01	26.27 $\pm$ 0.33	8.21 $\pm$ 0.17	10.97 $\pm$ 0.47	2.72 $\pm$ 0.08
<i>Paraconiothyrium</i> sp.	7.98 $\pm$ 0.27	0.21 $\pm$ 0.01	20.65 $\pm$ 0.01	26.64 $\pm$ 0.43	8.69 $\pm$ 0.15	9.88 $\pm$ 0.29	2.65 $\pm$ 0.05
<i>Periconia</i> sp.	8.50 $\pm$ 0.34	0.21 $\pm$ 0.01	20.37 $\pm$ 0.01	27.02 $\pm$ 0.34	8.25 $\pm$ 0.09	9.83 $\pm$ 0.34	2.61 $\pm$ 0.09
<i>Phialocephala</i> sp. 1	8.69 $\pm$ 0.29	0.21 $\pm$ 0.01	20.52 $\pm$ 0.01	26.34 $\pm$ 0.42	8.30 $\pm$ 0.09	9.79 $\pm$ 0.27	2.75 $\pm$ 0.09
<i>Phialocephala</i> sp. 2	8.75 $\pm$ 0.20	0.23 $\pm$ 0.01	19.30 $\pm$ 0.01	25.89 $\pm$ 0.27	8.25 $\pm$ 0.19	10.16 $\pm$ 0.43	2.62 $\pm$ 0.09
<i>Thozetella</i> sp.	8.27 $\pm$ 0.37	0.22 $\pm$ 0.01	19.39 $\pm$ 0.01	26.23 $\pm$ 0.50	8.23 $\pm$ 0.11	9.80 $\pm$ 0.24	2.53 $\pm$ 0.09
<b>p-value (ANOVA)</b>	0.72	0.50	0.45	0.81	0.61	0.50	0.13





451 **Author contribution**

452  
453 YC, JP and LCG designed the study; ES, LCG and WB performed the research; ES wrote the first draft of the manuscript, and  
454 all authors contributed to revisions.

455  
456  
457 **Competing interests**

458  
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463

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