

1 **Distribution, chemical and molecular composition of high and**
2 **low-molecular-weight humic-like substances in ambient aerosols**

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17 **Abstract**

18 Humic-like Substances (HULIS) encompass a continuum of molecular weight
19 (MW) ranges, yet our understanding of how HULIS characteristics vary with MW is
20 still limited and not well-established. In this study, a combination of ultrafiltration and
21 solid-phase extraction protocols was employed to fractionate the high MW (HMW, >1
22 kDa) and low MW (LMW, < 1kDa) HULIS fractions from ambient aerosols collected
23 during summer and winter at a rural site. Subsequently, comprehensive
24 characterization by using total organic carbon, high-performance size exclusion
25 chromatography (HPSEC), UV-vis and fluorescence spectroscopy, Fourier-transform
26 infrared spectroscopy (FTIR), negative electrospray ionization high resolution mass
27 spectrometry (ESI- HRMS) were conducted. The results revealed that HMW HULIS
28 were dominated by larger-sized chromophores, substantially constituting a higher
29 fraction of total organic carbon and UV absorption at 254 nm than LMW HULIS.
30 While both HMW and LMW HULIS shared similar fluorophore types and functional
31 groups, the former exhibited higher levels of humification and a greater presence of
32 polar functional groups (e.g., -COOH, >C=O). HRMS analysis further unveiled that
33 molecular formulas within HMW HULIS generally featured smaller sizes but higher
34 degrees of unsaturation and aromaticity compared to those within LMW HULIS
35 fractions. This observation suggests the possibility of small molecules assembling to
36 form the HMW HULIS through intermolecular weak forces. Moreover, HMW HULIS
37 contained a higher proportion of CHON but fewer CHO compounds than LMW
38 HULIS. In both HMW and LMW HULIS, the unique molecular formulas were

39 primarily characterized by lignin-like species, yet the former displayed a prevalence
40 of N-enriched and highly aromatic species. Additionally, HMW HULIS contained
41 more unique lipids-like compounds, while LMW HULIS exhibited a distinct presence
42 of tannin-like compounds. These findings provide valuable insights into the
43 distribution, optical properties, and molecular-level characteristics of HULIS in
44 atmospheric aerosols, thereby advancing our understanding of their sources,
45 composition, and environmental implications.

46

47 **Keywords:** Humic-Like Substances, molecular weight fractionation, optical
48 properties, high-performance size exclusion chromatography, negative electrospray
49 ionization-high resolution mass spectrometry

50

51 **1. Introduction**

52 Humic-Like Substances (HULIS) are complex and heterogeneous mixtures of
53 water-soluble organic matters (WSOM) that are of great importance in the
54 atmospheric environment. They usually share similar physicochemical properties (e.g.,
55 acidity, absorption, fluorescence, functional groups) with naturally occurring humic
56 substances (Graber and Rudich, 2006; Zheng et al., 2013) and are prevalent in fog,
57 clouds, rainwater and ambient aerosols (Birdwell and Valsaraj, 2010; Fan et al., 2016a;
58 Santos et al., 2012). With substantial hygroscopic and surface-active properties,
59 HULIS enhance the hygroscopic growth of particles, thereby contributing to the
60 formation of the cloud condensation nuclei and ice nuclei (Chen et al., 2021a; Dinar et
61 al., 2007). Moreover, acting as an important component of brown carbon (BrC),
62 HULIS effectively absorb near-ultraviolet and visible light, thus influencing the
63 global radiative balance and atmospheric chemistry processes (Bao et al., 2022;
64 Zhang et al., 2020). Furthermore, HULIS have the potential to catalyze the formation
65 of reactive oxygen species, leading to potential adverse health effects (Ma et al., 2019;
66 Zhang et al., 2022b).

67 The chemical composition of atmospheric HULIS exhibit significant
68 heterogeneity and typically comprises macromolecular compounds containing
69 aromatic rings with highly conjugated structures, as well as long-chain hydrocarbon
70 with polar groups (e.g., -OH, -COOH, -NO₂) (Fan et al., 2013; Huo et al., 2021). To
71 unravel the structural characteristics and properties of HULIS, a range of analytical
72 techniques, including absorption and fluorescence spectroscopy, Fourier transform

73 infrared spectroscopy (FTIR), nuclear magnetic resonance spectroscopy (^1H NMR),
74 have been utilized (Huo et al., 2021; Qin et al., 2022; Zou et al., 2020). These studies
75 have provided insights into the overall structural characteristics of complex HULIS,
76 including their abundances, chemical and optical characteristics (Huo et al., 2021;
77 Mukherjee et al., 2020; Win et al., 2018; Zhang et al., 2022b; Zheng et al., 2013). In
78 recent years, high-resolution mass spectrometry (HRMS) techniques, such as Fourier
79 transform ion cyclotron resonance mass spectrometry (FT-ICR-MS) and orbitrap
80 HRMS, in combination with electrospray ionization (ESI), have emerged as powerful
81 tools for elucidating the molecular-level characteristics of HULIS (Lin et al., 2012;
82 Sun et al., 2021; Wang et al., 2019; Zou et al., 2023). By utilizing HRMS, researchers
83 have gained deeper insights into the complexity and chemical heterogeneity of
84 HULIS at the molecular level.

85 Operationally, HULIS are defined as the hydrophobic fraction of water-soluble
86 organic matter (WSOM) typically extracted through solid-phase extraction (SPE)
87 protocol (Fan et al., 2012; Zou et al., 2020). Thus, the abundance and characteristics
88 of HULIS are contingent upon the chemical composition of WSOM. Previous studies
89 have shown that aerosol WSOM, often seen as water-soluble brown carbon (BrC), are
90 comprised of a continuum of molecular weight (MW) species, as revealed by
91 high-performance exclusion chromatography (HPSEC) analysis (Di Lorenzo et al.,
92 2017; Fan et al., 2023; Wong et al., 2019). These studies have highlighted that BrC
93 typically consist of both high-MW (HMW) and low-MW (LMW) chromophores in
94 various aerosols. For example, BrC emitted from fresh biomass burning (BB) are

95 dominated by low MW chromophores (Di Lorenzo et al., 2017; Wong et al., 2019).
96 However, BrC derived from aged BB aerosols and ambient aerosols tend to possess
97 more HMW chromophores that are highly chemically resistant (Di Lorenzo et al.,
98 2017; Fan et al., 2023; Wong et al., 2019). Further characterizations of different MW
99 BrC can be conducted using an ultrafiltration (UF) protocol (Fan et al., 2021). This
100 approach enabled researchers to obtain the distributions of content, chromophores and
101 fluorophores within various MW BrC fractions. Despite these advancements, the
102 chemical structures and molecular composition of different MW HULIS fractions
103 remain poorly understood. Consequently, a combination of UF and SPE protocols for
104 the fractionation and characterization of MW-separated HULIS is crucial, as it not
105 only provides insights into MW distributions but also illuminates the chemical
106 heterogeneities of aerosols HULIS.

107 In this study, a combination of UF-SPE isolation protocol was developed to
108 fractionate and characterize the MW HULIS fractions. Two distinct sets of ambient
109 PM_{2.5} samples collected during summer and winter periods were utilized to facilitate a
110 comparative analysis of MW HULIS. Initially, the WSOM were fractionated into
111 high-MW (HMW, >1 kDa) and low-MW (LMW, <1 kDa) species using the UF
112 protocol. Subsequently, the resulting MW WSOM fractions underwent SPE to obtain
113 different MW HULIS fractions. The obtained HMW and LMW HULIS fractions were
114 comprehensively characterized using advanced analytical techniques, including total
115 organic carbon analysis, UV-vis and fluorescence spectroscopy, HPSEC, and HRMS
116 to unveil their abundances, absorption and fluorescence properties, and molecular

117 characteristics. The findings of this study hold great significance in advancing our
118 understanding of the definition and molecular profiles of HULIS, as well as
119 facilitating further investigations into their potential impacts on the atmospheric
120 environment.

121

122 **2. Materials and methods**

123 2.1. Atmospheric fine particles sampling

124 Atmospheric PM_{2.5} were sampled on the rooftop of a building within the campus
125 of Anhui Science and Technology University (32.21°N, 118.72°E), around 20 m above
126 ground level. Detailed information regarding the sampling site can be found in our
127 previous studies (Cao et al., 2022; Fan et al., 2021). The PM_{2.5} samples were collected
128 using a high-volume PM_{2.5} sampler (JCH-1000, Juchuang Ltd., Qingdao) onto
129 prebaked quartz fiber filters (8 × 10 inches, Whatman). Sampling took place from
130 July 25 to August 12, 2021, during summer, and from December 19, 2021 to January
131 6, 2022, during winter. Blank filters were also collected as control samples. All
132 aerosol PM_{2.5} filter samples were stored at -20 °C in a freezer prior to analysis. The
133 atmospheric pollutant data (NO₂, SO₂ and O₃) near sampling site during sampling
134 period were obtained from the website (<https://www.aqistudy.cn>) and are summarized
135 in Table S1. Additionally, the fire spots were investigated using data from the website
136 (<https://firms.modaps.eosdis.nasa.gov/map/>) and are visualized in Fig. S1.

137

138 2.2. Application of UF-SPE for isolating MW HULIS fractions

139 Punctures of summer and winter aerosol PM_{2.5} filter samples were taken and
140 combined for the extraction of water-soluble organic matter (WSOM), respectively.
141 The filters were immersed in 300 mL of ultrapure water and subjected to
142 ultrasonication for 30 min. The resulting suspensions were then filtered through a 0.22
143 μm membrane (Φ 47 mm, Jinteng, China) to obtain bulk WSOM samples. These bulk
144 filtrates were further subjected to UF and SPE in tandem to obtain different MW
145 HULIS fractions. Please refer to Fig. S2 for a schematic representation of the
146 fractionation steps.

147 Before UF, the bulk WSOM were diluted to DOC concentration below 30 mg/L
148 to minimize the concentration effects and prevent the accumulation of organic matters
149 at the membrane surface during UF. The detailed UF procedure followed the profile
150 described in our previous study (Fan et al., 2021). Briefly, each bulk WSOM solution
151 was passed through a pre-cleaned 1 kDa cut-off membrane in a stirred UF cell
152 (Amicon 8200, Millipore, USA), with a pressure at 0.2 MPa applied by ultrapure N₂.
153 The concentration factor was ~10. The resulting retentate was considered as HMW
154 (>1 kDa) WSOM, while the permeate solutions represented the LMW (<1 kDa)
155 WSOM. Finally, each MW fraction was diluted to the initial volume for further
156 treatment and analysis. Mass balances of WSOM during one-step UF process
157 generally ranged from 92% to 99%, as determined by total organic carbon (TOC) and
158 UV absorption at 254 nm (UV₂₅₄), indicating good performance of UF without
159 substantial loss or organic contamination.

160 Subsequently, SPE was applied to isolate the so-called HULIS fractions from the

161 bulk and each MW fraction of WSOM, following the protocol proposed in our
162 previous studies (Fan et al., 2013; Zou et al., 2020). Briefly, the acidified aqueous
163 samples were passed through pre-activated HLB columns (Waters Oasis, 500 mg/6
164 mL, USA). The fractions retained on the resins (referred to as HULIS) were eluted
165 with pure methanol and dried using a gentle stream of pure N₂. Finally, the bulk,
166 HMW and LMW HULIS fractions were obtained. A blank filter control was
167 performed using the same procedure described above, and the analysis signals of
168 samples were corrected by blank control.

169

170 2.3. HPSEC analysis

171 The apparent MW distributions of MW HULIS fractions were analyzed using a
172 high-performance liquid chromatography (HPLC) system (LC-20AT, Shimadzu,
173 Japan) equipped with a refractive index detector (RID-10A, Shimadzu) and a diode
174 array detector (SPD-M20A, Shimadzu). The wavelength of the diode array detector
175 was set at 254 nm. Separation was performed using an aqueous gel filtration column
176 (Polysep-GFC-P 3000, Phenomenex) preceded by a guard column (Polysep-GFC-P,
177 Phenomenex). The mobile phase consisted of a mixture of water and methanol (90:10
178 v/v) containing 25 mM ammonium acetate (Di Lorenzo et al., 2017; Wong et al.,
179 2019). The sample injection volume was 100 μ L, and the flow rate was maintained at
180 1 mL min⁻¹. The HPSEC calibration was performed using a series of polyethylene
181 glycol (PEG) standards (Kawasaki et al., 2011; Zhang et al., 2022c). The
182 chromatographic peak areas were integrated to represent the abundances of

183 corresponding MW species. It should be noted that the MW values estimated here are
184 nominal rather than absolute due to the lack of appropriate standards for column
185 calibration (Fan et al., 2023; Wong et al., 2017).

186 The weight-average MW (M_w), number-average MW (M_n) and polydispersity
187 (ρ), were determined using the following equations (Song et al., 2010):

$$M_w = \frac{\sum_{i=1}^n (h_i MW_i)}{\sum_{i=1}^n h_i} \quad (1)$$

$$M_n = \frac{\sum_{i=1}^n h_i}{\sum_{i=1}^n (h_i/MW_i)} \quad (2)$$

$$\rho = \frac{M_w}{M_n} \quad (3)$$

188 where h_i and MW_i are the absorption intensity of the chromatogram and the MW of
189 molecules corresponding to the i th retention time, respectively.

190

191 2.4. Measurements of WSOC content and optical properties

192 The concentration of water-soluble organic carbon (WSOC) in HMW and LMW
193 HULIS was measured using a Shimadzu TOC analyzer (TOC-VCPN, Japan)
194 following the non-purgeable organic carbon protocol.

195 The UV-vis spectra were recorded using a UV-vis spectrophotometer (UV 2600,
196 Shimadzu, Japan) over a wavelength range of 200-700 nm with 1 nm increments.
197 Excitation-emission matrix (EEM) spectra were determined using a fluorescence
198 spectrophotometer (F4600, Hitachi, Japan). The scanning ranges for excitation (Ex)
199 and emission (Em) wavelengths were 200-400 and 290-520 nm, respectively, with a
200 scanning speed was 12,000 nm/min.

201 To characterize the chemical and optical properties of MW HULIS fractions,

202 several commonly used spectra parameters were calculated, including the specific UV
203 absorbance at 254 nm (SUVA₂₅₄), the UV absorbance ratio between 250 and 365 nm
204 (E_2/E_3), spectra slope ratios (S_R), the absorption Angstrom exponent (AAE), and mass
205 absorption efficiency (MAE₃₆₅), fluorescence indices (FI), biological index (BIX), and
206 humification degree (HIX) (Fan et al., 2021; Li et al., 2022; Wu et al., 2021). Further
207 details can be found in Text S1 of the Supporting Information (SI).

208

209 2.5. HRMS analysis and data processing

210 The MW HULIS fractions were analyzed using a Q-Exactive mass spectrometer
211 (Thermo Scientific, Germany) equipped with a heated electrospray ionization (ESI)
212 source. The system operated in negative ESI mode with a resolution of 140,000 at m/z
213 = 200. The detection mass range was set from 60 to 900 m/z . To ensure accurate mass
214 measurements, mass calibration was carried out using a commercial standard mixture
215 of ESI-L Low Concentration Tuning Mix (G1969-85000, Agilent, USA). It's
216 important to note that the negative ESI source is more sensitive to detecting polar acid
217 compounds, and the reported specific chemical composition here represents a fraction
218 that is biasedly ionized in the negative ESI source, not the entire HULIS composition
219 (He et al., 2023; Song et al., 2022).

220 The acquired mass spectra were processed using Xcalibur software (V2.2,
221 Thermo Scientific). The mathematically possible formulas for all ions were calculated
222 with a signal-to-noise ratio (s/n) ≥ 5 using a mass tolerance of 5 ppm. The assigned
223 molecular formulas followed specific constraints, with limitations on the following

224 elements: $C \leq 50$, $H \leq 100$, $O \leq 20$, $N \leq 3$, $S \leq 2$. Additionally, the elemental ratios of
225 H/C, O/C, N/C, and S/C were constrained to the ranges of 0.3–3.0, 0–3.0, 0–0.5, and
226 0–2.0, respectively. The double-bond equivalents (DBE) and modified aromaticity
227 index (AI_{mod}) values of the assigned neutral assigned formula ($C_cH_hO_oN_nS_s$) were
228 calculated using the equations (4-5) (He et al., 2023; Song et al., 2022):

$$DBE = 1 + \frac{1}{2}(2c - h + n) \quad (4)$$

$$AI_{mod} = \frac{1 + c - 0.5o - 0.5n - 0.5h}{c - 0.5o - n} \quad (5)$$

229 The intensity-weighted molecular parameters (X_w) of MW, H/C, O/C, DBE, and
230 AI values were calculated according to the equation (6) (He et al., 2023; Zhang et al.,
231 2021; Zou et al., 2023):

$$X_w = \frac{\sum(I_i \cdot X_i)}{\sum I_i} \quad (6)$$

232 where X represents the aforementioned parameters, and I_i denote the intensity for each
233 assigned formula i .

234

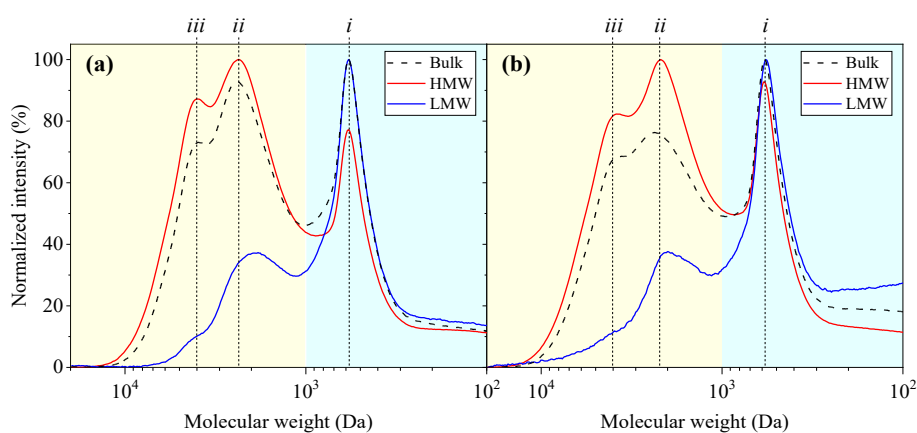
235 3. Results and discussion

236 3.1. Size and distribution of MW HULIS fractions

237 3.1.1. Molecular size of HMW and LMW HULIS

238 Due to the potential differences in molecular densities between the calibration
239 standards (PEG) and BB WSOC, the reported molecular weight of HULIS calculated
240 by HPSEC herein is only estimate. Fig. 1 shows the HPSEC chromatograms of MW
241 HULIS. Both HMW and LMW HULIS exhibit MW continuum distributions ranging
242 from 100 to 20,000 Da, which is consistent with the reported distributions of BrC in

243 BB-derived and various ambient aerosol in previous studies (Di Lorenzo et al., 2017;
 244 Fan et al., 2023; Wong et al., 2017). However, the chromatographic patterns for HMW
 245 HULIS clearly differ from those observed for LMW HULIS in both aerosol samples.
 246 As seen in Fig. 1, HMW HULIS display an additional and stronger absorption peak at
 247 around 4000 Da (peak *iii*), along with a more pronounced peak at 2200 Da (peak *ii*)
 248 and a similar magnitude peak at 570 Da (peak *i*) compared to LMW HULIS. This
 249 suggests that HMW HULIS contain the majority of larger molecular size
 250 chromophores within the bulk WSOM. In addition, the magnitude peak at 570 Da
 251 (peak *i*) in HMW HULIS may indicate the incorporation of small molecule through
 252 weak interactions based on π - π and/or van der Waals forces between the HULIS
 253 components (Fan et al., 2021; Piccolo, 2002).



254
 255 **Fig. 1.** Average HPSEC chromatograms of bulk, HMW and LMW HULIS fractions in
 256 (a) summer and (b) winter aerosol, respectively. The yellow and cyan shadows
 257 represent MW size regions of >1 kDa and <1 kDa, respectively.

258

259 Moreover, the molecular size of MW HULIS can be further reflected by the
 260 differences in Mw and Mn. As listed in Table 1, the average Mw and Mn of HMW

261 HULIS are 2233-2315 and 654-707 Da, respectively, greatly larger than that of LMW
 262 HULIS (989-1071 and 293-394 Da, respectively). Note that these values are estimates
 263 rather than absolute, given the absence of the appropriate aerosol HULIS standards
 264 (Fan et al., 2023; Wong et al., 2017). Moreover, these differences indicate that the
 265 sources and formation processes of HMW HULIS may differ from LMW HULIS.
 266 Many previous studies have demonstrated that fresh BB HULIS generally consist of
 267 small molecular-sized chromophores (Di Lorenzo et al., 2018; Di Lorenzo et al., 2017;
 268 Wong et al., 2017; Wong et al., 2019). However, a notable enhancement in the
 269 formation of large molecular-sized chromophores has been found when they undergo
 270 intricate atmospheric processes (Di Lorenzo et al., 2018; Di Lorenzo et al., 2017;
 271 Wong et al., 2017; Wong et al., 2019). Based on these limited studies, it is suggested
 272 that HMW HULIS, characterized by higher proportions of large-size chromophores
 273 and the resulting larger apparent molecular size, might be associated with the products
 274 from atmospheric aging process rather than being emitted directly by primary sources.
 275
 276 **Table 1.** The summary of typical quantity and quality parameters of each MW HULIS
 277 fraction from BB and ambient aerosols.

		Summer			Winter		
		Bulk	HMW	LMW	Bulk	HMW	LMW
HPSEC-derived parameters	Mw	1975±13	2315±38	1071±24	1918±56	2233±42	989±67
	Mn	591±53	707±48	394±13	525±57	654±17	293±32
	ρ	3.4±0.3	3.3±0.2	2.7±0.2	3.7±0.3	3.4±0.2	3.4±0.2
HULIS/WSOM (%) ^a	TOC	65±1	68±1	51±2	63±2	67±2	41±1
	UV ₂₅₄	66±5	65±2	55±4	67±1	65±1	61±2
Optical parameters	E ₂ /E ₃	12.02±0.54	11.72±0.31	14.98±0.98	6.30±0.24	6.54±0.16	7.24±0.43
	MAE ₃₆₅	0.21±0.02	0.23±0.01	0.20±0.01	1.04±0.02	1.06±0.01	0.88±0.00

AAE	7.11±0.32	7.59±0.00	8.25±0.23	6.66±0.06	6.25±0.06	7.28±0.03
FI	2.00±0.04	1.99±0.03	2.04±0.05	2.06±0.01	1.97±0.03	2.25±0.02
BIX	0.95±0.01	0.86±0.07	1.02±0.01	0.96±0.01	0.81±0.01	1.07±0.02
HIX	2.42±0.06	2.43±0.04	2.40±0.05	3.13±0.25	5.64±0.34	1.94±0.16

278 ^a The ratios of contents of SPE-isolated HULIS fractions to that of corresponding
 279 WSOM fractions determined by TOC and/or absorbance at 254 nm (UV₂₅₄).

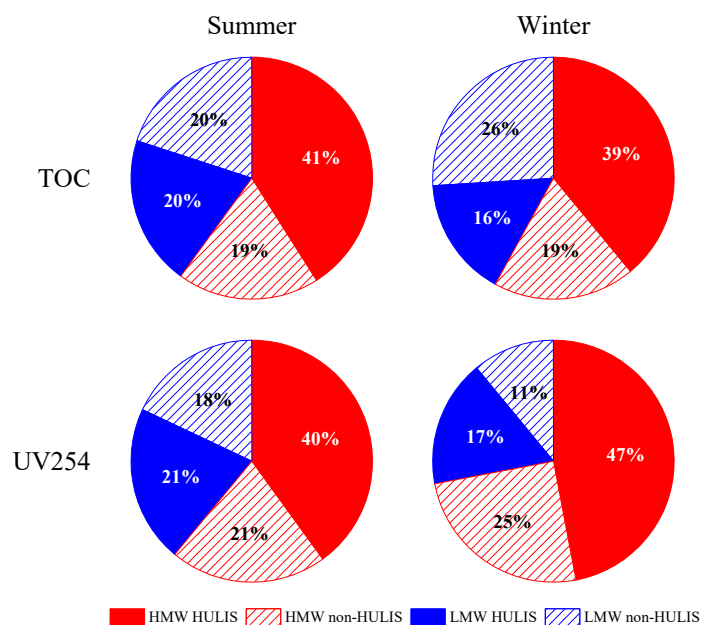
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281 3.1.2. Relative abundances of HMW and LMW HULIS

282 The contribution of MW-HULIS fractions to their corresponding MW-WSOM
 283 fractions, quantified in terms of TOC and UV absorption at 254 nm for both summer
 284 and winter aerosols are summarized in Table 1. In general, the ratios of
 285 HULIS/WSOM of HMW fractions (in terms of TOC and UV₂₅₄) (65-68%) were
 286 higher than the ratios (41-61%) observed for LMW fractions. This finding suggests
 287 that the higher presence of hydrophobic and conjugated aromatic structures in HMW
 288 WSOM, but more hydrophilic OC and non-aromatic species (e.g., aliphatic
 289 dicarboxylic acid) in the LMW WSOM (Fan et al., 2012; Zou et al., 2020).

290 Fig. 2 illustrates the distribution of distinct MW fractions within reconstructed
 291 WSOM, wherein “non-HULIS” refers to the content differences between the MW
 292 WSOM and its HULIS fractions. The HMW HULIS fraction contributed 39-41% of
 293 TOC and 40-47% of UV₂₅₄ to the bulk WSOM. In contrast, the LWM HULIS
 294 fraction only make up a smaller proportion, accounting for 16-20% of TOC and 17-21%
 295 of UV₂₅₄ within the bulk WSOM. Specifically, the ratios between HMW HULIS and
 296 LMW HULIS (H/L) ranged from 1.88 to 2.75 for both summer and winter aerosols in
 297 terms of either TOC or UV₂₅₄. These findings emphasize that HMW HULIS

298 significantly dominate the bulk aerosol HULIS fractions. Notably, the H/L ratio for
 299 winter aerosols was higher than that for summer aerosols, suggesting that larger-sized
 300 HULIS contributed more to the bulk HULIS fractions in winter aerosols.



301

302 **Fig. 2.** Relative proportions of different MW fractions in summer and winter aerosols
 303 determined by TOC and UV254.

304

305 The non-HULIS fractions are also important constituents within aerosol WSOM,
 306 but exhibit some differences between HMW and LMW fractions. The contributions of
 307 HMW non-HULIS to bulk WSOM were ~19% as determined by TOC and 21-25%
 308 measured by UV254. In case of LMW non-HULIS, the contributions were higher in
 309 terms of TOC (20-26%) but lower in terms of UV254 (11-18%). These results
 310 indicate that the LMW WSOM contain a larger proportion of hydrophilic organic
 311 species with weaker light absorption.

312

313 3.2. Optical characteristics of MW HULIS fractions

314 3.2.1. Light absorption characteristics

315 The absorption spectra of MW HULIS fractions in ambient aerosols are shown in
316 Fig. S3. These spectra exhibit a featureless shape with a general decrease in
317 absorbance as the wavelength increases, which is a typical characteristic of HULIS
318 found in rainwater, biomass burning (BB), and ambient aerosols (Huo et al., 2021;
319 Santos et al., 2009; Zhang et al., 2022b). The E_2/E_3 ratio, commonly used as an
320 indicator of the chemical characteristics of organic species, is inversely correlated
321 with higher aromaticity and larger molecular weight (Fan et al., 2021; Li et al., 2022;
322 Sun et al., 2021). As listed in Table 1, the E_2/E_3 of HMW HULIS fractions generally
323 were lower than that of LMW HULIS in both ambient aerosols. This is consistent
324 with the expectation that larger-sized HULIS generally possess more polyconjugated
325 and polymeric structures (Fan et al., 2021; Zhang et al., 2022c), leading to greater
326 aromaticity and larger molecular size.

327 MAE_{365} and AAE are commonly used to characterize the light absorption
328 capacity and the spectral dependence of light absorption by aerosol chromophores,
329 respectively (Bao et al., 2022; Fan et al., 2016b; Kumar et al., 2017; Yuan et al., 2021;
330 Zou et al., 2020). As listed in Table 1, the average MAE_{365} values of HMW HULIS
331 are 0.23 and 1.06 $m^2 g^{-1}$ in summer and winter aerosol, respectively. These values are
332 higher than the corresponding values of 0.20 and 0.88 $m^2 g^{-1}$, respectively, for LMW
333 HULIS. In addition, HMW HULIS presented lower AAE values, being 7.59 and 6.25
334 in summer and winter aerosol, respectively, than the corresponding values of 8.25 and
335 7.28, respectively, for LMW HULIS (Table 1). These findings are consistent with the

336 observation in our previous study that larger WSOM generally own higher MAE₃₆₅
337 but smaller AAE values than smaller WSOM (Fan et al., 2021). Generally, the results
338 suggest that HMW HULIS exhibit stronger light-absorbing ability but with light
339 absorption showing a weaker wavelength dependence. It is worth noting that
340 combustion sources, such as BB and coal combustion, usually emit primary HULIS
341 with high MAE₃₆₅ values due to the enrichment of poly-aromatic and unsaturated
342 species (Cao et al., 2021; Fan et al., 2018; Huo et al., 2021; Zhang et al., 2021), but
343 with small molecular weight distributions (Di Lorenzo et al., 2018; Di Lorenzo et al.,
344 2017; Wong et al., 2017). Furthermore, subsequent pronounced photooxidation and
345 photobleaching processes can lead to a enrichment and/or formation of large sized
346 chromophores (Di Lorenzo et al., 2018; Di Lorenzo et al., 2017; Wong et al., 2017),
347 concurrently resulting in a reduction in their absorption capacity and an enhancement
348 of their spectra dependence on wavelength (Chen et al., 2021b; Fan et al., 2019; Wu et
349 al., 2018; Wu et al., 2020; Zhang et al., 2022a). From this perspective, the SOA
350 formation might induce the generation of HMW HULIS with lower light absorption
351 capacity and weaker light absorption wavelength dependence. In contrast, LMW
352 HULIS is more likely to represent small-sized primary HULIS and/or by-products
353 resulting from the degradation and oxidation of primary large-sized HULIS.
354 Considering the complex sources of ambient HULIS, future studies should explore the
355 MW-dependent light absorption characteristics of HULIS from different sources.

356

357 3.2.2. Fluorescence characteristics

358 The EEM contours of MW HULIS fractions from both summer and winter
359 aerosols are presented in Fig. S4. These HULIS fractions from both seasons exhibit
360 similar EEM spectra features, with a predominance of humic-like fluorophores
361 (Ex/Em = 210-235/395-410 nm). This observation suggest that humic-like
362 fluorophores are fundamental constituents of both HMW and LMW HULIS, which
363 are consistent with previous findings for aerosols MW WSOM (Fan et al., 2021) and
364 bulk HULIS in BB-derived and ambient aerosols (Fan et al., 2020; Qin et al., 2018).
365 The findings imply an assembly of small and heterogeneous molecules to form bulk
366 HULIS through weak intramolecular forces (i.e., π - π , van der Waals, hydrophobic, or
367 hydrogen bonds) (Fan et al., 2021; Piccolo, 2002) and/or charge-transfer interactions
368 (Phillips et al., 2017; Qin et al., 2022). In this study, the fluorescence regional
369 integration (FRI) method was applied to characterize the fluorescent composition of
370 MW HULIS. Using FRI, EEM spectra were divided into five fluorescence regions
371 (labeled as I to V) (Fig. S4), which were successively assigned to simple aromatic
372 proteins (I and II), fulvic acid-like (III), soluble microbial byproduct-like (IV), and
373 humic acid-like (V) substances, respectively, as established in previous studies (Chen
374 et al., 2003; Qin et al., 2018; Wang et al., 2021b). As shown in Fig. S5, the large-size
375 aromatic proteins (II) and fulvic acid-like substances (III) dominated the fluorophores
376 within MW HULIS in both summer and winter aerosols, comprising approximately
377 62-64% of the total fluorescence intensity. This finding is consistent with previous
378 reports on bulk HULIS in summer and winter aerosols from industrial and urban cities
379 (Qin et al., 2018; Wang et al., 2021b). In comparison, the HMW HULIS in both

380 summer and winter aerosols generally exhibited a higher proportion of humic
381 acid-like substances (V), while having a lower abundance of small-size aromatic
382 proteins I compared to LMW HULIS. These differences are particularly pronounced
383 in winter aerosols, with the humic acid-like substances accounting for 23% in HMW
384 HULIS compared to 13% in LMW HULIS, and small-size aromatic proteins I
385 comprising 9% in HMW HULIS compared to 17% in LMW HULIS (Fig. S5).
386 Furthermore, the higher HIX values of HMW HULIS (5.64) in comparison to LMW
387 HULIS (1.94) further support these differences (Table 1). The pronounced BB
388 emissions and potential NO₂-related oxidation of OA, as evidenced by the presence of
389 more hotspots (Fig. S1) and higher concentration of NO₂ (Table S1), are likely driving
390 these marked distinctions between HMW and LMW HULIS in winter aerosols. In
391 general, these findings imply that the HMW HULIS have a stronger level of
392 humification and oxidation, while the LMW HULIS appear to be of a simpler nature
393 and are more likely associated with fresh emissions (e.g., BB).

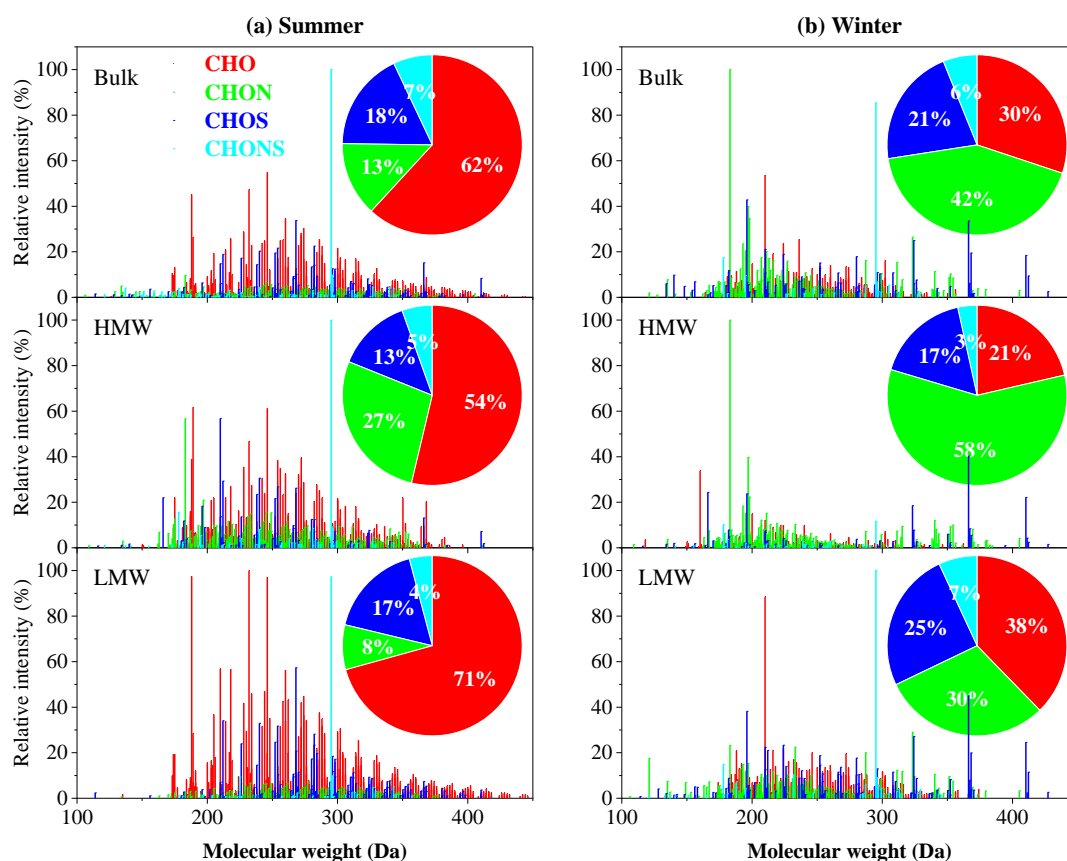
394

395 3.3 Molecular-level insights into MW HULIS

396 3.3.1. Seasonal variations in the molecular composition of MW HULIS

397 The molecular-level characteristics of MW HULIS were examined using
398 negative ESI- HRMS analysis. It is worth noting that the ESI- HRMS could reveal
399 molecular composition of a subset of organic molecules that are biased ionized in the
400 negative ESI source, particularly acid compounds such as carboxylic and sulfonic
401 acids, and may not represent the entire HULIS composition (He et al., 2023; Lin et al.,

402 2012). Additionally, HRMS techniques are often known to be biased towards low
403 masses below 600 Da (Lin et al., 2012; Wang et al., 2019), suggesting that the
404 molecular sizes calculated by HRMS are likely underestimated. Fig. 3 displays the
405 reconstructed mass spectra of all HULIS fraction in both summer and winter aerosols.
406 Hundreds of peaks can be observed in the spectra ranging from m/z 100 to 450 for all
407 samples, with most ions being abundant within the m/z 150-350 range. These
408 spectrum characteristics are similar to the low m/z range previously reported for
409 HULIS in ambient aerosols and BB emissions (He et al., 2023; Song et al., 2022; Sun
410 et al., 2021; Wang et al., 2019; Zhang et al., 2021; Zou et al., 2023). The similarity in
411 low m/z range among HMW and LMW HULIS and other bulk HULIS may indicate
412 their shared fundamental molecules, further suggesting the potential disassembly of
413 larger molecules.



414

415 **Fig. 3.** Mass spectra of bulk and MW HULIS in (a) summer and (b) winter aerosols.

416 The pie charts represent the intensity distributions of four compound categories (CHO,

417 CHON, CHOS, and CHONS).

418

419 As listed in Table 2, the number of assigned formulas within MW HULIS in

420 summer aerosols were 655-672, which was higher than the range of 470-506 observed

421 in winter aerosols. This suggests that the MW HULIS in summer aerosols exhibited

422 greater diversity than those in winter aerosols, mainly due to the stronger SOA

423 formation that enhanced the heterogeneity of HULIS fractions in the summer.

424 Moreover, SOA-derived HULIS generally contain more polar molecules (e.g., -OH,

425 -COOH) (Di Lorenzo et al., 2017; Fan et al., 2020; Huo et al., 2021), making them

426 more susceptible to deprotonation and leading to the generation of a greater number

427 formulas under negative ESI conditions. The identified formulas were then classified
428 into four groups (i.e., CHO, CHON, CHOS and CHONS) according to their elemental
429 composition. As depicted in pie charts in Fig. 3, summer HULISs are predominantly
430 composed of CHO (54-71%), while winter HULISs feature a high concentration of
431 both CHON (30-58%) and CHO (21-38%). The notably higher content of CHO in
432 summer HULISs are likely due to a wide distribution of biogenic VOC-derived SOAs
433 during the summer season (Li et al., 2022; Sun et al., 2023). CHON content in winter
434 HULISs is generally higher than in summer ones, potentially due to more significant
435 contributions from direct BB, as well as secondary nitrogen-related chemical
436 processes during the winter season (He et al., 2023; Song et al., 2022; Zhang et al.,
437 2021; Zou et al., 2023). This finding is supported by the greater number of fire spots
438 (Fig. S1) and higher concentrations of NO₂ (Table S1) during winter. The higher
439 proportions of CHON compounds in aerosol HULIS typically lead to enhanced light
440 absorption capabilities (He et al., 2023; Song et al., 2022; Zeng et al., 2021). This
441 provides a potential explanation for the higher MAE₃₆₅ values observed in winter
442 HULIS compared to summer HULIS. Additionally, CHOS is more abundant in winter
443 HULISs (17-25%) than in summer aerosols (13-18%). Previous studies have
444 demonstrated that both coal combustion and the oxidation initiated by SO₂ can lead to
445 the generation of larger amounts of S-containing compounds (Song et al., 2018; Song
446 et al., 2022; Zou et al., 2023). This finding suggested that the increased levels of coal
447 combustion and SO₂-related SOAs, as evidenced by higher concentration of SO₂
448 (Table S1), are significant contributors to the presence of BrC in winter compared to

449 in summer.

450 **Table 2.** The average values of intensity-weighted molecular weights (MW), elemental ratios, double bond equivalents (DBE), modified
 451 aromaticity index (AI_{mod}) and carbon oxidation state (OS_C) for tentatively identified compounds of the bulk and MW HULIS samples.

		Elemental compositions	Number of formulas	MW _w	H/C _w	O/C _w	N/C _w	S/C _w	O/N _w	O/S _w	OM/OC _w	DBE _w	DBE/C _w	DBE-O _w	AI _{mod,w}	OS _{C,w}
Summer	BULK	CHO	376	276	1.39	0.59					1.91	4.69	0.39	-2.20	0.15	-0.21
		CHON	270	260	1.35	0.60	0.15		5.52		2.10	4.99	0.50	-1.17	0.27	-0.14
		CHOS	133	278	1.74	0.71		0.11		6.61	2.38	2.21	0.24	-4.41	0.02	-0.33
		CHONS	81	265	1.69	0.59	0.18	0.14	4.98	5.20	2.49	3.38	0.36	-1.95	0.09	-0.52
		Total	860	273	1.47	0.61	0.03	0.03	1.09	1.54	2.06	4.20	0.38	-2.44	0.14	-0.24
	HMW	CHO	270	264	1.41	0.55					1.85	4.51	0.38	-1.70	0.16	-0.32
		CHON	264	248	1.42	0.61	0.14		5.23		2.09	4.44	0.47	-1.33	0.24	-0.20
		CHOS	72	247	1.82	0.75		0.14		5.83	2.53	1.81	0.22	-4.07	0.01	-0.32
		CHONS	39	269	1.55	0.64	0.19	0.15	4.87	5.16	2.60	3.82	0.44	-1.65	0.14	-0.28
		Total	645	258	1.48	0.60	0.05	0.03	1.70	1.07	2.05	4.09	0.39	-1.91	0.16	-0.29
	LMW	CHO	365	275	1.34	0.64					1.97	4.82	0.42	-2.47	0.16	-0.05
		CHON	155	272	1.39	0.66	0.11		6.60		2.12	4.80	0.45	-2.09	0.18	-0.08
		CHOS	120	284	1.71	0.74		0.11		7.05	2.43	2.51	0.25	-4.58	0.02	-0.22
		CHONS	32	322	1.71	0.64	0.11	0.11	6.59	6.63	2.42	3.32	0.30	-3.41	0.06	-0.42
		Total	672	278	1.42	0.66	0.01	0.02	0.80	1.48	2.08	4.36	0.39	-2.84	0.13	-0.10
Winter	BULK	CHO	142	247	1.23	0.48					1.75	5.35	0.47	-0.05	0.32	-0.26
		CHON	194	231	1.24	0.53	0.16		3.96		1.99	5.17	0.57	0.52	0.48	-0.18
		CHOS	79	271	1.93	0.51		0.11		4.97	2.14	1.21	0.14	-3.80	0.04	-0.91
		CHONS	20	271	1.61	0.69	0.16	0.13	5.60	5.90	2.60	3.32	0.40	-2.65	0.09	-0.24
		Total	435	247	1.41	0.52	0.08	0.03	2.02	1.42	1.99	4.27	0.44	-0.77	0.31	-0.37
	HMW	CHO	138	232	1.36	0.47					1.74	4.79	0.41	0.02	0.28	-0.42

	CHON	244	232	1.29	0.53	0.16		3.89		2.01	4.87	0.55	0.23	0.44	-0.23
	CHOS	59	292	2.08	0.40		0.12		4.40	2.04	0.46	0.06	-4.06	0.02	-1.27
	CHONS	29	236	1.74	0.48	0.22	0.20	2.82	3.02	2.57	2.88	0.38	-0.57	0.27	-0.77
	Total	470	242	1.46	0.50	0.10	0.03	2.37	0.85	1.98	4.04	0.43	-0.57	0.33	-0.46
LMW	CHO	176	249	1.23	0.54					1.82	5.25	0.48	-0.65	0.30	-0.15
	CHON	195	239	1.34	0.49	0.16		3.99		1.95	4.88	0.52	0.29	0.45	-0.36
	CHOS	107	280	1.94	0.54		0.10		5.42	2.17	1.20	0.13	-4.25	0.02	-0.85
	CHONS	28	272	1.67	0.69	0.16	0.14	5.89	6.03	2.63	2.99	0.37	-3.15	0.13	-0.29
	Total	506	256	1.47	0.54	0.06	0.04	1.61	1.78	2.01	3.96	0.39	-1.45	0.26	-0.40

452 Table 2 summarizes the intensity-weighted molecular parameters for MW
453 HULIS in both summer and winter aerosols. Evidently, the MW_w of summer HULISs
454 are 258-278, which are higher than the corresponding values of 242-256 for winter
455 HULISs. It is noted that these values are considerably smaller than those revealed by
456 HPSEC analysis. This difference could be explained by several factors: (1) ESI-
457 HRMS is biased towards relatively small molecules that easily deprotonated in
458 negative ESI mode (He et al., 2023; Lin et al., 2012); (2) SEC can provide an apparent
459 rather than accurate molecular size of molecules due to the lack of appropriate
460 standards for column calibration (Fan et al., 2023; Wong et al., 2017), and (3) the
461 potential disassembly of larger molecules stabilized by weak forces during
462 electrospray ionization of HRMS (Fan et al., 2021; Phillips et al., 2017). Nevertheless,
463 both HRMS and HPSEC indicate that summer HULISs exhibit larger sizes than
464 winter HULISs. Moreover, summer HULISs exhibit higher O/C_w ranging from 0.60
465 to 0.66, as well as OS_{C,w} ranging from -0.29 to -0.10, which exceed the respective
466 values of 0.50 to 0.54 and -0.46 to -0.37 observed in winter HULISs. Conversely,
467 winter HULISs display higher AI_{mod,w} values (0.26-0.33) than those (0.13-0.16) for
468 summer ones. These findings suggest that summer HULISs are characterized by a
469 high degree of oxidation, while winter HULISs exhibit stronger aromaticity.

470

471 3.3.2. Comparison on molecular composition of HMW and LMW HULIS

472 **CHO compounds.** The CHO compounds are prominent constituents within
473 HULIS fractions, accounting for 54% and 21% in summer and winter HMW HULIS,

474 respectively, whereas these proportions increase to 71% and 38% in LMW HULIS
475 (Fig. 3). It is worth noting that CHO compounds that undergo deprotonation in ESI-
476 mode are likely associated with the presence of carboxyl, carbonyl, alcohol and ester
477 (Lin et al., 2012; Wang et al., 2018). Moreover, CHO compounds in LMW HULIS
478 exhibit a higher oxygenation level compared to HMW HULIS, as evidenced by the
479 higher O/C_w and $OS_{C,w}$ values. As shown in Table 2, the O/C_w for CHO in LMW
480 HULIS are 0.55-0.64, which are higher than 0.47-0.54 observed in HMW HULIS. In
481 contrast, the H/C_w for CHO in HMW HULIS were consistently higher than those in
482 LMW HULIS, with values of 1.41 vs. 1.34 in summer and 1.36 vs. 1.23 in winter
483 (Table 2). This disparity strongly suggests a higher saturation level of CHO
484 compounds within HMW HULIS. This conclusion is further corroborated by the
485 lower DBE_w and $AI_{mod,w}$ observed for CHO in HMW HULIS compared to LMW
486 HULIS (Table 2). It is known that these values serve as estimations of C=C density
487 and aromatic and condensed aromatic structures (Song et al., 2022; Zhang et al.,
488 2021). Taken together, the CHO compounds within HMW HULIS exhibit a more
489 aliphatic nature but lower aromaticity and oxidation levels when compared to those
490 within LMW HULIS.

491 **CHON compounds.** HMW HULIS fractions consist of a higher proportion of
492 CHON compounds compared to LMW HULIS, with proportions of 27% vs. 8% in
493 summer and 58% vs. 30% in winter (Fig. 3). This observation suggests that HMW
494 HULIS contain a higher content of N-containing components. It is noted that the
495 LMW HULIS are generally characterized by higher O/N_w values of 6.60 in summer

496 and 3.99 in winter compared to 5.23 in summer and 3.89 in winter for HMW HULIS.
497 This indicates that the CHON compounds within LMW HULIS are more highly
498 oxidized than those within HMW HULIS. In general, compounds with $O/N \geq 3$ are
499 indicative of oxidized N groups such as nitro ($-NO_2$) or nitrooxy ($-ONO_2$), while
500 compounds with $O/N < 3$ may denote the reduced N-containing functional groups (i.e.,
501 amines) (He et al., 2023; Song et al., 2022; Zeng et al., 2021). In this study, a majority
502 of the CHON compounds, comprising 73-85% in summer and 59-64% in winter,
503 exhibited $O/N \geq 3$ in both MW HULIS fractions. This suggests that high
504 concentrations of nitro compounds or organonitrates dominate the CHON compounds
505 (Sun et al., 2023; Wang et al., 2018; Zeng et al., 2021), especially in summer samples,
506 primarily due to the hydroxyl radical oxidation of biogenic or anthropogenic VOC
507 precursors, as well as BB emissions (Song et al., 2022; Sun et al., 2021; Zhang et al.,
508 2021; Zou et al., 2023). Furthermore, the CHON compounds exhibiting $O/N \geq 3$ were
509 more abundant in LMW HULIS compared to HMW HULIS, accounting for 85% vs.
510 73% in summer and 64% vs. 59% in winter. In contrast, HMW HULIS contained
511 more CHON compounds with $O/N < 3$ compared to LMW HULIS. These findings
512 collectively indicate that the CHON within HMW HULIS possess lower content of
513 nitro compounds or organonitrates than LMW HULIS. Based on FTIR analysis (Text
514 S2 and Fig. S6 in SI), it is known that HMW HULIS contain more carboxylic groups
515 than LMW HULIS, which indicate a higher likelihood of HMW HULIS containing
516 more amino acids.

517 **CHOS and CHONS compounds.** In this study, we observed that CHOS

518 accounted for proportions of 13% to 25% in all MW HULIS fractions, while CHONS
519 had a lower proportion of 3% to 7% (Fig. 3). Notably, the distribution of CHOS
520 differed between HMW and LMW HULIS in both season samples. As depicted in Fig.
521 3, HMW HULIS contained fewer CHOS compounds compared to LMW HULIS, with
522 proportions of 13% vs. 17% in summer and 17% vs. 25% in winter. This finding
523 suggests that a greater number of CHOS compounds are incorporated into the LMW
524 HULIS fractions, which potentially leading to a reduction in the light absorption of
525 LMW HULIS (Zeng et al., 2021; Zhang et al., 2021). Furthermore, as indicated in
526 Table 2, both the CHOS and CHONS within LMW HULIS exhibited higher O/Sw
527 values than HMW HULIS in both seasonal samples. Consequently, the S-containing
528 compounds within LMW HULIS were characterized by a higher degree of oxidation,
529 primarily attributed to SO₂-related chemical oxidation process, in comparison to those
530 in HMW HULIS. Moreover, it was observed that 61% to 92% of CHOS compounds
531 exhibited O/S > 4, and 3% to 43% of CHONS compounds with O/S > 7 for all MW
532 HULIS fractions. Among them, HMW HULIS own lower proportions of CHOS with
533 O/S > 4 and CHONS with O/S > 7 than LMW HULIS, suggesting a reduced presence
534 of potential organosulfates and nitrooxyorganosulfates within HMW HULIS (Sun et
535 al., 2023; Wang et al., 2018; Zeng et al., 2021; Zou et al., 2023).

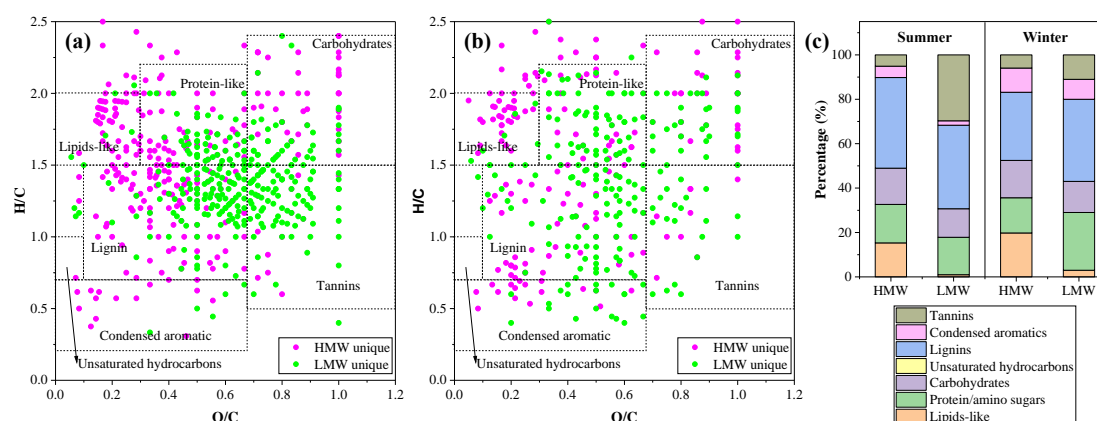
536

537 3.3.3. Comparative analysis of unique molecular formulas in HMW and LMW 538 HULIS

539 In this study, particular emphasis was placed on the unique molecular formulas

540 within the HMW or LMW HULIS fractions. Fig. 4a, b illustrates the Van Krevelen
 541 (VK) diagram depicting the distribution of unique molecular formulas within HMW
 542 and LMW HULIS in summer and winter samples. It is evident that a majority of
 543 unique formulas within LMW HULIS are concentrated around the origin with $O/C >$
 544 0.5 , accounting for 83% in summer and 64% in winter. In contrast, most formulas
 545 within HMW HULIS exhibited $O/C < 0.5$, representing about 58% for both seasonal
 546 samples. These findings indicate that the unique molecules within LMW HULIS
 547 consist of more polar O-containing organic compounds than those within HMW
 548 HULIS.

549



550

551 **Fig. 4.** Van Krevelen diagrams for the unique molecular formulas within HMW and
 552 LMW HULIS from (a) summer and (b) winter aerosols. (c) The contributions of
 553 major substances classes in unique formulas.

554

555 The molecular formulas are further categorized into seven groups based on
 556 previous studies, including lignin-like species, protein/amino sugars, condensed
 557 aromatics, tannin-like species, carbohydrate-like species, unsaturated hydrocarbons,
 558 and lipid-like species (He et al., 2023; Sun et al., 2021; Sun et al., 2023). The

559 classification rules for these formulas can be found in Table S2. Fig. 4c provides an
560 overview of the relative contributions of the number of unique formulas from each of
561 the seven groups for HMW and LMW HULIS. The results indicate that the dominant
562 substance class in the unique formulas within both MW HULIS are lignin-like species,
563 accounting for proportions of 31-40%. This finding indicates that lignin derivatives
564 are fundamental components in both HMW and LMW HULIS either in summer or
565 winter aerosols. Additionally, there are notable differences in the molecular
566 characteristics of lignin-like species within HMW and LMW HULIS. As listed in
567 Table S3, lignin-like species within HMW HULIS exhibit lower MW_w and O/C_w, but
568 higher N/C_w and A_{Imod,w} values than those within LMW HULIS in both seasonal
569 samples. These observations suggest that the unique lignin-like substances in HMW
570 HULIS likely contain more N-enriched and highly aromatic species, while those in
571 LMW HULIS tend to concentrate more aliphatic O-containing compounds. These
572 distinctions in composition and characteristics between HMW and LMW HULIS
573 fractions provide valuable insights into their origins and transformations in the
574 atmosphere.

575 Moreover, there are notable variations in the contributions of lipids-like,
576 protein/amino sugars, carbohydrates, condensed aromatics, and tannins species
577 between HMW and LMW HULIS. In general, HMW HULIS have a higher proportion
578 of lipids-like species, carbohydrates and condensed aromatics than LMW HULIS in
579 both summer and winter aerosols. Among these, the most remarkable difference in
580 composition between HMW HULIS and LMW HULIS is seen in lipids-like species,

581 accounting for 15% versus 1% in summer and 20% versus 3% in winter (Fig. 4). As
582 reported in previous studies, lipids-like species primarily originate from biogenic
583 emissions (He et al., 2023; Li et al., 2022; Sun et al., 2021). This suggests that there is
584 a stronger contribution from biogenic emissions to HMW HULIS. Additionally, these
585 species in HMW HULIS were usually characterized by lower DBE_w and slightly
586 lower $OS_{C,w}$ when compared to LMW HULIS (Table S3), indicating they present
587 stronger saturation and fewer oxidized substituents. On the other hand, tannins species
588 contribute a higher proportion to LMW HULIS, constituting 30% in summer and 11%
589 in winter, while comprising only 5%-6% in HMW HULIS in both season aerosols.
590 Tannin-like species are known to consist of various polyphenolic groups containing
591 hydroxyl and carboxylic functional groups (He et al., 2023; Li et al., 2022; Ning et al.,
592 2019; Sun et al., 2021). The slightly lower DBE_w but much higher $DBE-O_w$ for
593 unique tannin-like species within HMW HULIS were observed compared to LMW
594 HULIS (Table S3), suggesting that the former ones are enriched in more unsaturated
595 O-containing functional groups, particularly carboxylic functional groups.

596

597 3.4. Atmospheric implications

598 This study provides comprehensive comparison between HMW and LMW
599 HULIS regarding their distributions, chemical structures, molecular sizes and
600 compositions. HMW HULIS appear to be larger than LMW HULIS, as evidenced by
601 both ultrafiltration natures and the MW distributions of chromophores analyzed by
602 HPSEC. However, HRMS analysis revealed that the average MW_w of identified

603 formulas within HMW HULIS were lower than those of LMW HULIS (Table 2). This
604 discrepancy can likely be attributed to the “assembled structures” that construct the
605 aerosol HULIS, as suggested in many previous studies focusing on HULIS and BrC
606 characterization (Fan et al., 2021; Fan et al., 2023; Phillips et al., 2017; Qin et al.,
607 2022). In fact, the results from EEM-FRI and FTIR analysis support the notion that
608 HMW and LMW HULIS likely consist of potential structures assembled by similar
609 basic fluorophores and functional groups. Based on this theory, HMW HULIS may
610 consist of macromolecular species primarily assembled from small molecules through
611 weak forces (i.e., π - π , van der Waals, hydrophobic, or hydrogen bonds) and/or
612 charge-transfer interactions (Fan et al., 2021; Phillips et al., 2017), which can
613 potentially disassemble during ESI ionization and form low MW molecules.

614 Based on the molecular-level characterization, significant distinctions in
615 properties between HMW HULIS and LMW HULIS become evident. HMW HULIS
616 generally exhibit stronger aromaticity but lower oxidation degree when compared to
617 LMW HULIS. In terms of molecular composition, HMW HULIS contain higher
618 quantities of CHON species but lower quantities of CHO compounds than LMW
619 HULIS. Furthermore, more lipids-like species were identified as unique molecules in
620 HMW HULIS, while more tannin-like species with abundant carboxylic groups were
621 observed as unique molecules in LMW HULIS. Given these pronounced differences
622 between HMW and LMW HULIS, it can be speculated that the higher levels of
623 aromatic structures, greater presence of CHON molecules and the presence of
624 lipids-like species may serve as driving factors in the formation of potential

625 assembled structures in HMW HULIS. Additionally, it is well-established that CHON
626 can enhance the light absorption of organic aerosols (OA), while CHO species may
627 have the opposite effect, weakening light absorption (He et al., 2023; Song et al.,
628 2022; Wang et al., 2019; Zeng et al., 2021). Therefore, it is reasonable to conclude
629 that HMW HULIS possess stronger light absorbing capability, which is consistent
630 with their larger MAE₃₆₅ values.

631 Importantly, HMW HULIS contain higher amounts of carboxylic functional
632 groups, reduced nitrogen species (e.g., amines) and aromatic species than LMW
633 HULIS. These functional groups have strong complexation abilities with transition
634 metals (Wang et al., 2021a; Wang et al., 2021b), thus influencing the transformation
635 and chemical behavior of metals. Moreover, the OA-metals complex can potentially
636 enhance the catalytic generation of reactive oxygen species (ROS) in organic aerosols
637 (Win et al., 2018; Zhang et al., 2022a), thereby playing significant roles in adverse
638 health effects of OA. These results reinforce the significance of HMW HULIS in light
639 absorption, metal complexation, and the potential ROS generation ability of aerosol
640 BrC.

641

642 **4. Conclusions**

643 This study successfully isolated and characterized HMW and LMW HULIS in
644 atmospheric aerosols using the UF-SPE technique, yielding insights into their
645 distribution, optical properties and molecular-level characteristics. Both HMW and
646 LMW HULIS exhibited a continuum of MW distributions ranging from 100 to 20,000

647 Da. However, HMW HULIS displayed more extensive and intricate MW distributions,
648 suggesting differences in their sources and formation processes compared to LMW
649 HULIS. In general, HMW HULIS constituted a higher percentage of TOC and UV254
650 in aerosols compared to LMW HULIS, indicating the prevalence of hydrophobic and
651 conjugated aromatic structures in the former. Moreover, HMW HULIS exhibited
652 higher aromaticity, stronger light absorption abilities, weaker spectra dependence, and
653 stronger humification and conjugation, compared to LMW HULIS. Interestingly,
654 HRMS analysis revealed slightly lower MW_w values for HMW HULIS than LMW
655 HULIS, which contradicted the HPSEC results and the nature of UF fractionation.
656 This finding strongly suggests the possibility of small molecules assembling to form
657 macromolecules in HMW HULIS. Regarding molecular composition, HMW HULIS
658 contained a higher proportion of CHON compounds, constituting 27% vs. 8% in
659 summer and 58% vs. 30% in winter, while LMW HULIS were primarily composed of
660 CHO compounds, accounting for 71% vs. 54%% in summer and 38% vs. 21% in
661 winter. Both HMW and LMW HULIS featured lignin-like substances as major unique
662 molecular formulas, but HMW HULIS exhibited more N-enriched and highly
663 aromatic species, whereas LMW HULIS contained a higher proportion of polar
664 O-containing functional groups. Additionally, HMW HULIS included a greater
665 number of unique lipids-like compounds, while LMW HULIS tend to concentrate
666 more tannin-like compounds. These observations shed light on the complex nature of
667 MW HULIS, and their diverse sources and transformations. Future research should
668 expand the geographical and seasonal coverage to gain a more comprehensive

669 understanding of the molecular-level characteristics of MW HULIS in various
670 atmospheric environments. Furthermore, exploring additional physicochemical
671 properties of MW HULIS will provide valuable insights into their potential health and
672 environmental implications. Overall, this study offers valuable insights into the
673 molecular-level characteristics of aerosol HULIS, enhancing our understanding of
674 their evolution, sources and potential environmental effects.

675

676 **Author contribution**

677 **Xingjun Fan:** Methodology, Supervision, Funding acquisition, Writing-review &
678 editing. **Ao Cheng:** Sampling, Data curation. **Xufang Yu:** Writing-review & editing.
679 **Tao Cao:** Sampling, Investigation. **Dan Chen:** Investigation, Data curation.
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682 editing. **Pingan Peng:** Writing-review & editing.

683 **Declaration of Competing Interest**

684 The authors declare that they have no known competing financial interests or personal
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693 **References**

694 Bao, M., Zhang, Y.-L., Cao, F., Lin, Y.-C., Hong, Y., Fan, M., Zhang, Y., Yang, X., Xie, F., 2022. Light
695 absorption and source apportionment of water soluble humic-like substances (HULIS) in
696 PM_{2.5} at Nanjing, China. *Environmental Research* 206, 112554.

697 Birdwell, J.E., Valsaraj, K.T., 2010. Characterization of dissolved organic matter in fogwater by
698 excitation–emission matrix fluorescence spectroscopy. *Atmos. Environ.* 44, 3246-3253.

699 Cao, T., Li, M., Xu, C., Song, J., Fan, X., Li, J., Jia, W., Peng, P., 2022. Technical note: Identification of
700 chemical composition and source of fluorescent components in atmospheric
701 water-soluble brown carbon by excitation-emission matrix with parallel factor analysis:
702 Potential limitation and application. *Atmos. Chem. Phys. Discuss.* 2022, 1-41.

703 Cao, T., Li, M., Zou, C., Fan, X., Song, J., Jia, W., Yu, C., Yu, Z., Peng, P., 2021. Chemical composition,
704 optical properties, and oxidative potential of water- and methanol-soluble organic
705 compounds emitted from the combustion of biomass materials and coal. *Atmos. Chem.*
706 *Phys.* 21, 13187-13205.

707 Chen, J., Wu, Z.J., Zhao, X., Wang, Y.J., Chen, J.C., Qiu, Y.T., Zong, T.M., Chen, H.X., Wang, B.B., Lin,
708 P., Liu, W., Guo, S., Yao, M.S., Zeng, L.M., Wex, H., Liu, X., Hu, M., Li, S.M., 2021a.
709 Atmospheric Humic-Like Substances (HULIS) Act as Ice Active Entities. *Geophysical*
710 *Research Letters* 48, e2021GL092443.

711 Chen, Q., Hua, X., Dyussenova, A., 2021b. Evolution of the chromophore aerosols and its driving
712 factors in summertime Xi'an, Northwest China. *Chemosphere* 281, 130838.

713 Chen, W., Westerhoff, P., Leenheer, J.A., Booksh, K., 2003. Fluorescence Excitation-Emission
714 Matrix Regional Integration to Quantify Spectra for Dissolved Organic Matter. *Environ.*
715 *Sci. Technol.* 37, 5701-5710.

716 Di Lorenzo, R.A., Place, B.K., VandenBoer, T.C., Young, C.J., 2018. Composition of Size-Resolved
717 Aged Boreal Fire Aerosols: Brown Carbon, Biomass Burning Tracers, and Reduced
718 Nitrogen. *ACS Earth and Space Chemistry* 2, 278-285.

719 Di Lorenzo, R.A., Washenfelder, R.A., Attwood, A.R., Guo, H., Xu, L., Ng, N.L., Weber, R.J., Baumann,
720 K., Edgerton, E., Young, C.J., 2017. Molecular-Size-Separated Brown Carbon Absorption
721 for Biomass-Burning Aerosol at Multiple Field Sites. *Environ. Sci. Technol.* 51, 3128-3137.

722 Dinar, E., Taraniuk, I., Graber, E.R., Anttila, T., Mentel, T.F., Rudich, Y., 2007. Hygroscopic growth of
723 atmospheric and model humic-like substances. *J. Geophys. Res.* 112.

724 Fan, X., Cai, F., Xu, C., Yu, X., Wang, Y., Xiao, X., Ji, W., Cao, T., Song, J., Peng, P., 2021. Molecular
725 weight-dependent abundance, absorption, and fluorescence characteristics of
726 water-soluble organic matter in atmospheric aerosols. *Atmos. Environ.* 247.

727 Fan, X., Cao, T., Yu, X., Wang, Y., Xiao, X., Li, F., Xie, Y., Ji, W., Song, J., Peng, P., 2020. The
728 evolutionary behavior of chromophoric brown carbon during ozone aging of fine
729 particles from biomass burning. *Atmos. Chem. Phys.* 20, 4593-4605.

730 Fan, X., Cheng, A., Chen, D., Cao, T., Ji, W., Song, J., Peng, P., 2023. Investigating the molecular
731 weight distribution of atmospheric water-soluble brown carbon using high-performance

732 size exclusion chromatography coupled with diode array and fluorescence detectors.
733 Chemosphere 338, 139517.

734 Fan, X., Li, M., Cao, T., Cheng, C., Li, F., Xie, Y., Wei, S., Song, J., Peng, P.a., 2018. Optical properties
735 and oxidative potential of water- and alkaline-soluble brown carbon in smoke particles
736 emitted from laboratory simulated biomass burning. *Atmos. Environ.* 194, 48-57.

737 Fan, X., Song, J., Peng, P., 2013. Comparative study for separation of atmospheric humic-like
738 substance (HULIS) by ENVI-18, HLB, XAD-8 and DEAE sorbents: elemental composition,
739 FT-IR, ¹H NMR and off-line thermochemolysis with tetramethylammonium hydroxide
740 (TMAH). *Chemosphere* 93, 1710-1719.

741 Fan, X., Song, J., Peng, P.a., 2016a. Temporal variations of the abundance and optical properties
742 of water soluble Humic-Like Substances (HULIS) in PM_{2.5} at Guangzhou, China. *Atmos.*
743 *Res.* 172-173, 8-15.

744 Fan, X., Wei, S., Zhu, M., Song, J., Peng, P., 2016b. Comprehensive characterization of humic-like
745 substances in smoke PM_{2.5} emitted from the combustion of biomass materials and fossil
746 fuels. *Atmos. Chem. Phys.* 16, 13321-13340.

747 Fan, X., Yu, X., Wang, Y., Xiao, X., Li, F., Xie, Y., Wei, S., Song, J., Peng, P.a., 2019. The aging
748 behaviors of chromophoric biomass burning brown carbon during dark aqueous
749 hydroxyl radical oxidation processes in laboratory studies. *Atmos. Environ.* 205, 9-18.

750 Fan, X.J., Song, J.Z., Peng, P.A., 2012. Comparison of isolation and quantification methods to
751 measure humic-like substances (HULIS) in atmospheric particles. *Atmos. Environ.* 60,
752 366-374.

753 Graber, E.R., Rudich, Y., 2006. Atmospheric HULIS: How humic-like are they? A comprehensive
754 and critical review. *Atmos. Chem. Phys.* 6, 729-753.

755 He, T., Wu, Y., Wang, D., Cai, J., Song, J., Yu, Z., Zeng, X., Peng, P.a., 2023. Molecular compositions
756 and optical properties of water-soluble brown carbon during the autumn and winter in
757 Guangzhou, China. *Atmos. Environ.* 296, 119573.

758 Huo, Y., Wang, Y., Qi, W., Jiang, M., Li, M., 2021. Comprehensive characterizations of HULIS in
759 fresh and secondary emissions of crop straw burning. *Atmos. Environ.* 248, 118220.

760 Kawasaki, N., Matsushige, K., Komatsu, K., Kohzu, A., Nara, F.W., Ogishi, F., Yahata, M., Mikami, H.,
761 Goto, T., Imai, A., 2011. Fast and precise method for HPLC-size exclusion
762 chromatography with UV and TOC (NDIR) detection: Importance of multiple detectors to
763 evaluate the characteristics of dissolved organic matter. *Water Res.* 45, 6240-6248.

764 Kumar, V., Goel, A., Rajput, P., 2017. Compositional and surface characterization of HULIS by
765 UV-Vis, FTIR, NMR and XPS: Wintertime study in Northern India. *Atmos. Environ.* 164,
766 468-475.

767 Li, X., Yu, F., Cao, J., Fu, P., Hua, X., Chen, Q., Li, J., Guan, D., Tripathee, L., Chen, Q., Wang, Y., 2022.
768 Chromophoric dissolved organic carbon cycle and its molecular compositions and
769 optical properties in precipitation in the Guanzhong basin, China. *Sci. Total Environ.* 814,
770 152775.

771 Lin, P., Rincon, A.G., Kalberer, M., Yu, J.Z., 2012. Elemental composition of HULIS in the Pearl River
772 Delta Region, China: results inferred from positive and negative electrospray high
773 resolution mass spectrometric data. *Environ. Sci. Technol.* 46, 7454-7462.

774 Ma, Y., Cheng, Y., Qiu, X., Cao, G., Kuang, B., Yu, J.Z., Hu, D., 2019. Optical properties, source
775 apportionment and redox activity of humic-like substances (HULIS) in airborne fine

776 particulates in Hong Kong. *Environmental Pollution* 255, 113087.

777 Mukherjee, A., Dey, S., Rana, A., Jia, S., Banerjee, S., Sarkar, S., 2020. Sources and atmospheric
778 processing of brown carbon and HULIS in the Indo-Gangetic Plain: Insights from
779 compositional analysis. *Environmental Pollution* 267, 115440.

780 Ning, C., Gao, Y., Zhang, H., Yu, H., Wang, L., Geng, N., Cao, R., Chen, J., 2019. Molecular
781 characterization of dissolved organic matters in winter atmospheric fine particulate
782 matters (PM_{2.5}) from a coastal city of northeast China. *Sci. Total Environ.* 689, 312-321.

783 Phillips, S.M., Bellcross, A.D., Smith, G.D., 2017. Light Absorption by Brown Carbon in the
784 Southeastern United States is pH-dependent. *Environ. Sci. Technol.* 51, 6782-6790.

785 Piccolo, A., 2002. The supramolecular structure of humic substances: A novel understanding of
786 humus chemistry and implications in soil science, *Advances in Agronomy*. Academic
787 Press, pp. 57-134.

788 Qin, J., Zhang, L., Qin, Y., Shi, S., Li, J., Gao, Y., Tan, J., Wang, X., 2022. pH-Dependent Chemical
789 Transformations of Humic-Like Substances and Further Cognitions Revealed by Optical
790 Methods. *Environ. Sci. Technol.* 56, 7578-7587.

791 Qin, J., Zhang, L., Zhou, X., Duan, J., Mu, S., Xiao, K., Hu, J., Tan, J., 2018. Fluorescence
792 fingerprinting properties for exploring water-soluble organic compounds in PM_{2.5} in an
793 industrial city of northwest China. *Atmos. Environ.* 184, 203-211.

794 Santos, P.S.M., Otero, M., Duarte, R.M.B.O., Duarte, A.C., 2009. Spectroscopic characterization of
795 dissolved organic matter isolated from rainwater. *Chemosphere* 74, 1053-1061.

796 Santos, P.S.M., Santos, E.B.H., Duarte, A.C., 2012. First spectroscopic study on the structural
797 features of dissolved organic matter isolated from rainwater in different seasons. *Sci.*
798 *Total Environ.* 426, 172-179.

799 Song, J., Li, M., Jiang, B., Wei, S., Fan, X., Peng, P., 2018. Molecular Characterization of
800 Water-Soluble Humic like Substances in Smoke Particles Emitted from Combustion of
801 Biomass Materials and Coal Using Ultrahigh-Resolution Electrospray Ionization Fourier
802 Transform Ion Cyclotron Resonance Mass Spectrometry. *Environ. Sci. Technol.* 52,
803 2575-2585.

804 Song, J., Li, M., Zou, C., Cao, T., Fan, X., Jiang, B., Yu, Z., Jia, W., Peng, P.a., 2022. Molecular
805 Characterization of Nitrogen-Containing Compounds in Humic-like Substances Emitted
806 from Biomass Burning and Coal Combustion. *Environ. Sci. Technol.* 56, 119-130.

807 Song, J.Z., Huang, W.L., Peng, P.A., Xiao, B.H., Ma, Y.J., 2010. Humic Acid Molecular Weight
808 Estimation by High-Performance Size-Exclusion Chromatography with Ultraviolet
809 Absorbance Detection and Refractive Index Detection. *Soil. Sci. Soc. Am. J.* 74,
810 2013-2020.

811 Sun, H., Li, X., Zhu, C., Huo, Y., Zhu, Z., Wei, Y., Yao, L., Xiao, H., Chen, J., 2021. Molecular
812 composition and optical property of humic-like substances (HULIS) in winter-time PM_{2.5}
813 in the rural area of North China Plain. *Atmos. Environ.* 252, 118316.

814 Sun, H., Wu, Z., Kang, X., Zhu, C., Yu, L., Li, R., Lin, Z., Chen, J., 2023. Molecular characterization of
815 humic-like substances (HULIS) in atmospheric particles (PM_{2.5}) in offshore Eastern China
816 Sea (OECS) using solid-phase extraction coupled with ESI FT-ICR MS. *Atmos. Environ.*
817 294, 119523.

818 Wang, K., Zhang, Y., Huang, R.-J., Cao, J., Hoffmann, T., 2018. UHPLC-Orbitrap mass
819 spectrometric characterization of organic aerosol from a central European city (Mainz,

820 Germany) and a Chinese megacity (Beijing). *Atmos. Environ.* 189, 22-29.

821 Wang, X., Qin, Y., Qin, J., Long, X., Qi, T., Chen, R., Xiao, K., Tan, J., 2021a. Spectroscopic insight
822 into the pH-dependent interactions between atmospheric heavy metals (Cu and Zn) and
823 water-soluble organic compounds in PM_{2.5}. *Sci. Total Environ.* 767, 145261.

824 Wang, X.B., Qin, Y.Y., Qin, J.J., Yang, Y.R., Qi, T., Chen, R.Z., Tan, J.H., Xiao, K., 2021b. The
825 interaction laws of atmospheric heavy metal ions and water-soluble organic compounds
826 in PM_{2.5} based on the excitation-emission matrix fluorescence spectroscopy. *Journal of
827 Hazardous Materials* 402, 8.

828 Wang, Y., Hu, M., Lin, P., Tan, T., Li, M., Xu, N., Zheng, J., Du, Z., Qin, Y., Wu, Y., Lu, S., Song, Y., Wu,
829 Z., Guo, S., Zeng, L., Huang, X., He, L., 2019. Enhancement in Particulate Organic Nitrogen
830 and Light Absorption of Humic-Like Substances over Tibetan Plateau Due to
831 Long-Range Transported Biomass Burning Emissions. *Environ. Sci. Technol.* 53,
832 14222-14232.

833 Win, M.S., Tian, Z., Zhao, H., Xiao, K., Peng, J., Shang, Y., Wu, M., Xiu, G., Lu, S., Yonemochi, S.,
834 Wang, Q., 2018. Atmospheric HULIS and its ability to mediate the reactive oxygen
835 species (ROS): A review. *J Environ Sci (China)* 71, 13-31.

836 Wong, J.P.S., Nenes, A., Weber, R.J., 2017. Changes in Light Absorptivity of Molecular Weight
837 Separated Brown Carbon Due to Photolytic Aging. *Environ. Sci. Technol.* 51, 8414-8421.

838 Wong, J.P.S., Tsagkaraki, M., Tsiodra, I., Mihalopoulos, N., Violaki, K., Kanakidou, M., Sciare, J.,
839 Nenes, A., Weber, R.J., 2019. Atmospheric evolution of molecular-weight-separated
840 brown carbon from biomass burning. *Atmos. Chem. Phys.* 19, 7319-7334.

841 Wu, G., Fu, P., Ram, K., Song, J., Chen, Q., Kawamura, K., Wan, X., Kang, S., Wang, X., Laskin, A.,
842 Cong, Z., 2021. Fluorescence characteristics of water-soluble organic carbon in
843 atmospheric aerosol ☆. *Environmental Pollution* 268, 115906.

844 Wu, G., Wan, X., Gao, S., Fu, P., Yin, Y., Li, G., Zhang, G., Kang, S., Ram, K., Cong, Z., 2018.
845 Humic-Like Substances (HULIS) in Aerosols of Central Tibetan Plateau (Nam Co, 4730 m
846 asl): Abundance, Light Absorption Properties, and Sources. *Environ. Sci. Technol.* 52,
847 7203-7211.

848 Wu, G., Wan, X., Ram, K., Li, P., Liu, B., Yin, Y., Fu, P., Loewen, M., Gao, S., Kang, S., Kawamura, K.,
849 Wang, Y., Cong, Z., 2020. Light absorption, fluorescence properties and sources of brown
850 carbon aerosols in the Southeast Tibetan Plateau. *Environmental Pollution* 257, 113616.

851 Yuan, W., Huang, R.-J., Yang, L., Ni, H., Wang, T., Cao, W., Duan, J., Guo, J., Huang, H., Hoffmann,
852 T., 2021. Concentrations, optical properties and sources of humic-like substances (HULIS)
853 in fine particulate matter in Xi'an, Northwest China. *Sci. Total Environ.* 789, 147902.

854 Zeng, Y., Ning, Y., Shen, Z., Zhang, L., Zhang, T., Lei, Y., Zhang, Q., Li, G., Xu, H., Ho, S.S.H., Cao, J.,
855 2021. The Roles of N, S, and O in Molecular Absorption Features of Brown Carbon in
856 PM_{2.5} in a Typical Semi-Arid Megacity in Northwestern China. *Journal of Geophysical
857 Research: Atmospheres* 126, e2021JD034791.

858 Zhang, T., Huang, S., Wang, D., Sun, J., Zhang, Q., Xu, H., Hang Ho, S.S., Cao, J., Shen, Z., 2022a.
859 Seasonal and diurnal variation of PM_{2.5} HULIS over Xi'an in Northwest China: Optical
860 properties, chemical functional group, and relationship with reactive oxygen species
861 (ROS). *Atmos. Environ.* 268, 118782.

862 Zhang, T., Shen, Z., Huang, S., Lei, Y., Zeng, Y., Sun, J., Zhang, Q., Ho, S.S.H., Xu, H., Cao, J., 2022b.
863 Optical properties, molecular characterizations, and oxidative potentials of different

864 polarity levels of water-soluble organic matters in winter PM_{2.5} in six China's megacities.
865 *Sci. Total Environ.* 853, 158600.

866 Zhang, T., Shen, Z., Zeng, Y., Cheng, C., Wang, D., Zhang, Q., Lei, Y., Zhang, Y., Sun, J., Xu, H., Ho,
867 S.S.H., Cao, J., 2021. Light absorption properties and molecular profiles of HULIS in PM_{2.5}
868 emitted from biomass burning in traditional "Heated Kang" in Northwest China. *Sci. Total*
869 *Environ.* 776, 146014.

870 Zhang, T., Shen, Z., Zhang, L., Tang, Z., Zhang, Q., Chen, Q., Lei, Y., Zeng, Y., Xu, H., Cao, J., 2020.
871 PM_{2.5} Humic-like substances over Xi'an, China: Optical properties, chemical functional
872 group, and source identification. *Atmos. Res.* 234, 104784.

873 Zhang, W., Li, L., Wang, D., Wang, R., Yu, S., Gao, N., 2022c. Characterizing dissolved organic
874 matter in aquatic environments by size exclusion chromatography coupled with multiple
875 detectors. *Anal. Chim. Acta* 1191, 339358.

876 Zheng, G.J., He, K.B., Duan, F.K., Cheng, Y., Ma, Y.L., 2013. Measurement of humic-like substances
877 in aerosols: A review. *Environmental Pollution* 181, 301-314.

878 Zou, C., Cao, T., Li, M., Song, J., Jiang, B., Jia, W., Li, J., Ding, X., Yu, Z., Zhang, G., Peng, P.a., 2023.
879 Measurement report: Changes in light absorption and molecular composition of
880 water-soluble humic-like substances during a winter haze bloom-decay process in
881 Guangzhou, China. *Atmos. Chem. Phys.* 23, 963-979.

882 Zou, C., Li, M., Cao, T., Zhu, M., Fan, X., Peng, S., Song, J., Jiang, B., Jia, W., Yu, C., Song, H., Yu, Z.,
883 Li, J., Zhang, G., Peng, P.a., 2020. Comparison of solid phase extraction methods for the
884 measurement of humic-like substances (HULIS) in atmospheric particles. *Atmos. Environ.*
885 225, 117370.

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887