Multifactorial effects of warming, low irradiance, and low salinity on Arctic kelps

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Abstract

The Arctic is projected to warm by 2 to 5°C by the end of the century. Warming causes melting of glaciers, shrinking of the areas covered by sea ice, and increased terrestrial runoff from snowfields and permafrost thawing. Warming, decreasing coastal underwater irradiance, and lower salinity are potentially threatening

15   polar marine organisms, including kelps, that are key species of hard-bottom shallow communities. The present study investigates the physiological responses of four kelp species (*Alaria esculenta*, *Laminaria digitata, Saccharina latissima,* and *Hedophyllum nigripes*) to these environmental changes through a perturbation experiment in *ex situ* mesocosms. Kelps were exposed for six weeks to four experimental treatments: an unmanipulated control, a warming condition under the CO_2 emission scenario SSP5-8.5, and

20   two multifactorial conditions combining warming, low salinity, and low irradiance reproducing the future coastal Arctic exposed to terrestrial runoff under two CO_2 emission scenarios (SSP2-4.5 and SSP5-8.5). The physiological effects on *A. esculenta*, *L. digitata* and *S. latissima* were investigated and gene expression patterns of *S. latissima and H. nigripes* were analyzed. Across all species and experimental treatments, growth rates were similar, underlying the acclimation potential of these species to future Arctic conditions.

25   Specimens of *A. esculenta* increased their chlorophyll *a* content when exposed to low irradiance conditions,

suggesting that they may be resilient to an increase in glacier and river runoff with the potential to become more dominant at greater depths. *S. latissima* showed a lower carbon:nitrogen (C:N) ratio under the SSP5-8.5 multifactorial conditions treatment, suggesting tolerance to coastal erosion and permafrost thawing. In contrast, *L. digitata* showed no response to the conditions tested on any of the investigated physiological

30 parameters. The down-regulation of genes coding for heat-shock proteins in *H. nigripes* and *S. latissima* underscores their ability to acclimate to heat stress, which portrays temperature as a key influencing factor. Based on these results, it is expected that kelp communities will undergo changes in species composition that will vary at local scale as a function of the changes in environmental drivers.

35 Keywords: climate change - Arctic - kelps - mesocosm

Summary (500 characters)

We tested the effects of warming, low salinity, and low irradiance on Arctic kelps. We show that growth rates were similar across species and treatments. *Alaria esculenta* is adapted to low light conditions.
40 *Saccharina latissima* exhibited nitrogen limitation suggesting coastal erosion and permafrost thawing could be beneficial. *Laminaria digitata* did not respond to the treatments. Gene expression of *Hedophyllum nigripes* and *S. latissima* indicated acclimation to the experimental treatments.

1. **Introduction**

45 The Arctic region is warming at more than twice the global average rate (Richter-Menge et al., 2017). Over the next 80 years, sea surface temperature is projected to increase by 2°C according to the Shared Socio-economic Pathways (SSP) 1-2.6, which foresees an increasing shift towards sustainable practices, and up to 5°C according to the SSP5-8.5, which assumes an energy-intensive and fossil fuel-based economy (Kwiatkowski et al., 2020). Warming induces glacier and sea ice to melt at a faster rate causing an increase
50 in terrestrial runoff from thawing snowfields and permafrost (Shiklomanov and Shiklomanov 2003; Stroeve

et al., 2014). Total freshwater inflow into the Arctic Ocean rose by around 7% between 1936 and 1999 and 14% between 1980 and 2009 (Peterson et al., 2002; Ahmed et al., 2020). Combined with vertical mixing by waves and wind action, cryosphere melting results in local turbid and low-salinity waters down to 20 m (Karsten 2007). Coastal areas are therefore exposed to warming, changing light and salinity conditions (Lebrun et al., 2022).

In the coastal Arctic, kelps are key ecosystem engineers that form underwater forests. Kelp are large brown macroalgae of the class Phaeophyceae and the order Laminariales. Kelp forests provide a food source, habitat, and nursery ground for numerous fish and invertebrates as well as protection of coastlines from erosion (Filbee-Dexter et al., 2019). They support complex food webs and have a substantial role in storing and sequestering carbon (Krause-Jensen and Duarte 2016). *Saccharina latissima*, *Alaria esculenta, Laminaria digitata,* and *Hedophyllum nigripes* are four abundant kelp species that inhabit the northern hemisphere and extend to subarctic and Arctic waters (Bischof et al., 1999; Müller et al. 2009). As a result of warming, which induces more sea ice-free areas, the surface area suitable for kelps has increased by about 45% from 1940-1950 to 2000-2017 (Krause-Jensen et al., 2020). Temperature requirements and seasonal variability tolerance in irradiance and salinity for reproduction and growth determine the geographical distribution of kelp species (Wiencke et al. 1994, Muth et al., 2021). The temperature tolerance of these kelp species found in the Arctic appears broad (0 - 20°C), however, there is significant variability in temperature optima across ecotypes (Bolton and Lüning, 1982, Andersen et al., 2013, Diehl et al., 2021). Similarly, low salinity is well tolerated (Karsten et al. 2007), yet it is unclear how low salinity stress may interact with other stressors. Irradiance has a major impact on their depth distribution (e.g. Roleda et al. 2005; Krause-Jensen et al., 2012). Turbid waters alter kelp fitness by limiting photosynthesis. This has already induced a shift in the vertical distribution of kelps such as *Laminaria* and *Saccharina* genera to shallower waters in Arctic fjords (Bartsch et al., 2016; Filbee-Dexter et al., 2019). Because optimal temperature, irradiance, and salinity ranges vary between kelp species, their response to environmental changes will likely be species-specific (Eggert 2012; Karsten 2012).

We hypothesized that (1) warming will enhance the growth rate of Arctic kelps in Kongsfjorden during summer, and (2) that the combined effects of high temperature, low salinity and low irradiance will negatively impact their physiology, although responses will be species-specific. To test these hypotheses and

fill knowledge gaps on the multifactorial effects of climate change across species (Renaud et al., 2019; Scherrer et al., 2019), we carried out a land-based mesocosm experiment exposing four kelp species (*S. latissima*, *A. esculenta*, *L. digitata,* and *H. nigripes*) found in common biomass between 5 to 10 m depth to four treatments for six weeks. The treatments consisted of a control, a warming condition mimicking the future offshore (T1), and two multifactorial conditions combining warming, low salinity, and low irradiance mimicking the future coastal Arctic (T2 and T3). In order to best represent *in situ* conditions, the different kelp species were incubated together in each mesocosm at densities mimicking natural communities. The physiological effects on *A. esculenta*, *L. digitata* and *S. latissima* were investigated and gene expression patterns of *S. latissima* and *H. nigripes* were analyzed.

2. Material and methods

2.1 Specimen collection

In June 2021, 188 sporophytes of *A. esculenta*, *L. digitata, S. latissima,* and *H. nigripes* shorter than 1 m were collected by research divers in Kongsfjorden (Svalbard, Norway). They were collected between 2 and 7 m depth at Hansneset and the Old Pier (Fig. 1). All samples were placed into holding tanks ($> 1 \text{ m}^3$) with flow-through ambient seawater until their placement into final mesocosms on 2021-07-03.

2.2 Mesocosm experiment

The experiment was carried out from 2021-07-03 (t_0) to 2021-08-28 (t_{final}), in twelve 1 m^3 mesocosms set up in Ny-Ålesund on the outdoor platform of the Kings Bay Marine Laboratory in order to expose communities to natural light cycles. Each mesocosm received 3 to 6 individuals of *A. esculenta* and *S. latissima,* 2 to 4 individuals of *L. digitata* and 0 to 2 individuals of *H. nigripes* for a total kelp biomass (wet weight) per mesocosm of about 1500 g for *S. latissima* and *L. digitata* (mingled with *H. nigripes*) and 1000 g for *A. esculenta.* These biomasses are representative of those found at Hansneset down to 7 m depth (Hop et al., 2012). Since *H. nigripes* can be mistaken for *L. digitata*, each stipe of these two species was cut at t_{final} to detect individuals with mucilage, corresponding to *H. nigripes* (n=16, Dankworth et al., 2020).

The experimental set-up is briefly described below. More information can be found in Miller et al. (2024a). Seawater flowing through the mesocosms was pumped from 10 m depth in front of the Kings Bay Marine Laboratory (78.929°N, 11.930°E) using a submersible pump (Albatros©). The regulated flow-through system (7 - 8 L min^{-1} in each mesocosm) allowed for the automated control of temperature and salinity. Temperature was adjusted by mixing ambient seawater with warmed seawater (15°C) and salinity was regulated by addition of freshwater. Irradiance was modified by placing spectra and attenuating light filters on top of each mesocosm to simulate current irradiance and future *in situ* irradiance (see below for more details). Each mesocosm was equipped with one 12 W wave pump (Sunsun© JVP-132) to ensure proper mixing.

Four experimental treatments in triplicate (4x3 mesocosms) were used to study conditions representative of present and future Arctic coastal communities at proximity or not to glaciers following two different SSP scenarios (Ctrl, T1, T2, T3; Table 1). Treatments 1 and 2 (T1 and T2) mimicked the conditions expected close to glaciers and, therefore, combined warming, low irradiance, and low salinity. T1 followed the SSP 2-4.5, which describes a middle-of-the-road projection that does not shift markedly from historical patterns, while T2 followed the SSP5-8.5 that assumes an energy-intensive and fossil fuel-based economy. T3 focused on the projected change outside glacial fjords following the SSP 5-8.5, where warming acts as a single driver. Temperature was increased by 3.3°C in T1 and 5.3°C in T2 and T3 as an offset increase from the control condition (Ctrl) which mimicked the *in situ* temperature recorded in real-time during the whole experiment. Based on *in situ* measurements of temperature and salinity in Kongsfjorden taken from week 22 to 35 in 2020, salinity offsets were determined from the *in situ* relationship between temperature-salinity and extrapolated to apply to future warming. This resulted in a salinity decrease by 2.5 in T1 and 5 in T2 (Miller et al., 2024a). Based on *in situ* photosynthetically active radiation (PAR) data collected in May 2021 with a LI-COR (model 192), irradiance was reduced from the control by a mean of 25% for T1, corresponding to the difference between the glacier-proximal inner region and the middle of the fjord, and 40% for T2, corresponding to the difference between the inner and outer parts of the fjord. To simulate the *in situ* light spectrum (Kai Bischof, pers. com.) and reach the irradiance matching the targeted treatments, green (RL244) and neutral Lee filters© (RL211; RL298) were placed on top of each mesocosm accordingly (Table 1). During the first week, all the mesocosms were maintained under *in situ* conditions of temperature, salinity,

and irradiance. The light filters were then added to the mesocosms of T1 and T2 treatments on 2021-07-10, and all treatments gradually reached their targeted temperature and salinity conditions in six days. The experiment then lasted for six weeks.

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2.3 Tissue sampling

Tissue samples were collected in the meristem (located directly above the stipe-frond junction on the frond) of ten individuals of *A. esculenta*, *L. digitata, and S. latissima* at the beginning of the experiment (t_0, 2021-07-03) and on the healthy organisms, namely complete organisms (frond, stipe, and holdfast) that exhibited

140 a firm brown frond without signs of disease at the end (t_{final}), pending determination of chlorophyll a (chl *a*, see section 2.4) and the carbon:nitrogen (C:N) ratio (see section 2.5). Samples were stored in aluminum foil at -20°C. Additional tissue samples were collected in the meristem of *S. latissima* and *H. nigripes* at t_{final} for gene expression analysis (n=8 for each species, see section 2.7). These tissue samples were immediately flash-frozen in liquid nitrogen before being stored at -80°C.

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2.4 Chl *a* content

Samples were blotted dry, weighed (wet weight), and ground with a glass pestle. Chl *a* was extracted in 90% aqueous acetone for 24 h in the dark at 4°C. After cold-centrifugation (0°C, 15 min, 3000 rpm), the supernatants were transferred one at a time into a glass vial and the initial fluorescence (F_0) of chl *a* and

150 pheophytin pigment were measured using a fluorometer (Turner Design 10-AU Fluorometer; 667 nm). The F_a fluorescence (fluorescence after acidification) was measured one minute after the addition of 10 μl of 0.3 N HCl to transform chl *a* into pheophytin pigment and subtract F_a from F_0. The chl *a* content was calculated using the formula of Lorenzen (1967) but modified for determining chl *a* from mass of tissue rather than volume of seawater as (Eq. 1):

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$$Chl\ a\ (\mu g\ gFW^{-1}) \ = \ \frac{F_m}{(F_m-1)}\ \frac{km(F_0-F_a)}{m_{kelp}} \qquad (1)$$

where k is the calibration factor of pigment to fluorescence intensity [(µg chl a mg solvent^{-1}) (instrument fluorescence unit)$^{-1}$], m is mass of acetone used for extraction (mg), and m_{kelp} is the fresh weight of kelp (mg). Chl a content is expressed in µg per g of fresh weight (µg gFW^{-1}). No dilution factor was used as this was dry tissue mass.

2.5 C:N mass ratio

Samples were dried at 60°C for 48 h, weighed (dry weight), and their sizes adjusted to ensure that they did not weigh more than 10 mg, the detection limits specific to the CHN analyzer (PerkinElmer, Inc 2400). C and N contents are expressed in mg per g of dry weight (mg gDW^{-1}).

2.6 Growth rate

Growth rate was determined using the hole puncture method of Parke (1948). Sporophytes were punctured at t_0 in the meristem section of each organism, 2 cm above the base of the frond. The distance from the base of the frond to the hole was measured at t_{final}. The growth rate was calculated as follows (Eq. 2):

$$Growth\ rate\ (cm.d^{-1}) = \frac{dist_{final} - dist_0}{t_{final} - t_0} \qquad (2)$$

with $dist$ as the distance (in cm) from the base of the frond to the hole at time t (in days).

Weekly growth rates for selected individuals were determined at different time points during the experiment for *S. latissima* (weeks 1 and 4) and *A. esculenta* (weeks 2 and 5). Results can be found in the supplementary material (Fig. S2).

2.7 Gene expression analysis

We chose to target two species for gene expression, *S. Latissima* and *H. nigripes*. *S. latissima* was selected because it was the most abundant by biomass in the sampling area and appeared in robust physical health upon visual inspection at t_{final}. *H. nigripes* was selected because it is an endemic Arctic species and, thus, a model comparison specimen.

Total RNA extraction was conducted using the method described by Heinrich et al. (2012) which uses a CTAB extraction followed with a commercial Qiagen kit. The quantity and purity of the extracted RNA were evaluated using a Nanodrop ND-1000 Spectrophotometer (ThermoFisher), which measures RNA concentration at 260 nm and assesses purity by detecting the presence of other compounds such as DNA at 230 nm and proteins at 280 nm. The integrity of total RNA was determined by automated capillary electrophoresis using an Agilent 2100 Bioanalyzer (Agilent Technologies). The cDNA libraries were constructed by poly(A) enrichment and sequenced on a NovaSeq 6000 instrument by the Genome Quebec platform. The 100 bp paired reads were clipped using default values of the Illumina software. The quality of raw sequences was checked using FastQC v.0.11.7 (https://www.bioinformatics.babraham.ac.uk/projects/fastqc/). Sequences of low quality were trimmed using Trimmomatic v.0.39 (Bolger et al., 2014). For each species, a *de novo* transcriptome was constructed using the Trinity v.2.14.0 tool (Grabherr et al., 2011). The most homologous sequences were clustered using the CD-HIT-EST algorithm, part of the CD-HIT v.4.8.1 tool (Li and Godzik, 2006). To ensure the quality of the *de novo* transcriptomes, another transcriptome per species was generated using the rnaSPAdes v.3.14.1 (Bushmanova et al., 2019). Transcriptomes generated using rnaSPAdes and Trinity were compared using BUSCO v.5.4.3, transcriptomes generated with Trinity were retained due to lower duplicated sequences (Simão et al., 2015). Transcript quantification was performed by pseudo alignment using Kallisto v0.46.0, mapping RNA sequences to an index created from *de novo* transcriptomes (Bray et al., 2016). Exploration of differentially expressed genes (DEGs) was performed with the "DESeq2 v1.34.0" R package (Love et al., 2014). For each species, DEGs were obtained from the following comparisons: T1 vs. C, T2 vs. C, T3 vs. C, T2 vs. T1, T3 vs. T1, and T3 vs. T2. Transcripts with an adjusted $p < 0.05$ and $\log_2$ fold change (FC) > 2 or

< -2 were considered significantly differentially expressed genes. Functional annotation of the genes was performed with eggNOG-mapper v2.1.10 against the eggNOG database v.5.0.2 (Huerta-Cepas et al., 2017 & 2019). To ensure they were properly annotated, annotation was also performed with TransDecoder v5.5.0 to predict coding sequences (Haas and Papanicoualo, 2015), which were aligned against a Pfam profile database v35.0 (Mistry et al., 2021) using the HMMER v3.3 alignment tool (Finn et al., 2011). Gene Ontology (Gene Ontology Consortium, 2015) terms were then retrieved from the pfam2go database (https://pypi.org/project/pfam2go/) and functional enrichment was performed with Ontologizer v2.1 to obtain statistically significant GOs from the DEGs of each comparison performed previously (Bauer et al., 2008). Functional enrichment results were summarized as tree plots and scatter plots using REVIGO v1.8.1 (Supek et al., 2011). Investigation of the specific functions of DEGs was carried out by manually checking the involvement of Pfam domains and EggNOG annotations on the SMART database v9.0 (Letunic et al., 2021). Some DEGs whose annotation was questionable (i.e. not referring to plant genomes such as gene collagen) were removed, as well as those whose annotation was not precise enough to be classified. DEGs were then classified into different categories: cytoskeleton, genetic transcription/translation, metabolism, signaling, transport, stress (heat stress and oxydo-reduction processes), and energy production (respiration and photorespiration). A part of DEGs (73.2% in *S. latissima* and 82.3% in *H. nigripes*) were trimmed as they lacked functional annotation. Tools and parameters are summarized in Table S1.

2.8 Statistics

Rosner's generalized Extreme Studentized Deviation (ESD) test was used to detect the outliers using the function rosnerTest of the R package "EnvStats" (Millard, 2013). Out of a total of 165 individual chl *a* measurements, when combining all species and conditions, eleven were identified as outliers and removed. After the removal of the outliers, the normal distribution of the data was verified with a Shapiro-Wilk test using the function shapiro.test from the "stats" R package (R Core Team, 2013 $p > 0.105$). No outliers were identified in the C:N and growth rate data and normality was verified ($p > 0.089$).

Chl *a* content and C:N were analyzed using a linear mixed model with a hierarchical structure (HLM) to evaluate treatment effects by species. The model was fitted using the function lmer in the R package "lme4" (Bates et al., 2015). The fixed factors for the model were treatment and species, while mesocosm was a random factor. For growth rate measurements, a generalized linear mixed model (GLMM) with a Gaussian distribution was preferred - based on an Akaike information criterion - to test for the effects of the species, treatment, and mesocosm replica.

3. Results

3.1 Experimental conditions

The median temperature value in the control treatment was 5.3°C during the experimental period (2021-07-16 to 2021-08-28) calculated based on the mean value across replicates (Fig. 2, Table 1). The median salinity was 33.8 and the median daily PAR was 47.8 µmol photons m^{-2} s^{-1}. In treatment T1, the median temperature, salinity, and PAR were 8.9°C, 31, and 36.1 µmol photons m^{-2} s^{-1}, respectively. For treatments T2 and T3, the median temperature was elevated to 10.8°C. In T2, median salinity and PAR were decreased to 28.5, and 31.4 µmol photons m^{-2} s^{-1}.

3.2 Chl *a* content

The factors species and treatment were found as significant predictors, including their interaction when assessing differences in measured chl *a* content at t_0 and t_{final} ($p < 0.001$; Table S2). For *A. esculenta,* the concentration of chl *a* decreased significantly between t_0 and the control at t_{final} ($p < 0.01$, Fig. 3, Table S3). Values in the T2 treatment were also significantly higher than the control, T1, and T3 treatments (all *p* were < 0.01). Values in the control, T1, and T3 treatments were not statistically different from each other ($p > 0.92$). Similarly to *A. esculenta*, chl *a* content of *S. latissima* significantly decreased between t_0 and t_{final} ($p = 0.02$) for the control, but were not significantly impacted by the treatments ($p > 0.99$). The chl *a* content of *L. digitata* was not significantly impacted by time and treatments ($p > 0.99$).

3.3 C:N ratio

The statistical significance of predictor variables species and treatment for C:N, carbon content, and N content were significantly different for species and treatment ($p < 0.001$; Table S4), with the exception of treatment as a non-significant predictor for carbon content. There were no significant interactions between species and treatment. The pairwise comparisons determined that for *S. latissima,* C:N ratios at t_0 ranged from 24.5 up to 37.1 (Fig. 4). No statistical difference was found between t_0, the control, T1, and T3 treatment at t_{final} ($p > 0.93$, Tables S4, S5). In contrast, C:N ratios of individuals in the T2 treatment were significantly lower than at t_0, ranging from 15.2 to 29.5 (Fig. 4A, p = 0.045). Although carbon content showed no significant difference across treatments and time (Fig. 4B, p = 1), there was a notable increase in nitrogen content in the T2 treatment compared to t_0, but it was not statistically significant (Fig. 4C, p = 0.06). The C:N ratios, carbon, and nitrogen contents of *A. esculenta* and *L. digitata* were not significantly impacted by the treatments (p > 0.32).

3.4 Growth rate

The growth rates of *A. esculenta*, *L. digitata,* and *S. latissima* were not significantly impacted by the treatments (Fig. 5, p = 1, Tables S6, S7). They ranged from 0 to 0.037 cm d^{-1} for *A. esculenta*, 0.007 to 0.046 cm d^{-1} for *L. digitata,* and 0.040 up to 0.509 cm d^{-1} for *S. latissima.* The growth rate of *S. latissima* was significantly higher than for the two other species for each treatment ($p < 0.01$).

The growth rate of *A. esculenta* significantly decreased between week 2 and week 6 ($p < 0.01$, Fig. S1A) over time in the control. For *S. latissima*, significant differences in growth over time were only found in the T3 treatment (p=0.02, Fig. S1B). No intermediate measurements of *L. digitata* growth rate were taken.

3.5 Gene expression analysis

Principal component analysis of global gene expression revealed a clear contrast between the control and the different treatments for both *S. latissima* and *H. nigripes* (Fig. S2). The number of total differentially expressed genes (DEGs, i.e. genes that are either up- or down-regulated when comparing the different

treatments to the control) were close between *S. latissima* (831 including 225 classified; i.e. functionally annotated) and *H. nigripes* (815 including 144 classified, Fig. 6A) and mostly down-regulated for both species (84 and 65% respectively). For *H. nigripes*, the majority of overlapping DEGs were found between treatments T1 and T2 (Fig. 6A). Conversely, for *S. latissima*, the highest number of overlapping DEGs was observed between treatments T1 and T3. In both species, no overlapping genes were identified when comparing the DEGs between treatment pairs T1 vs. T2 and T2 vs. T3 (Fig. 6B).

The highest number of DEGs were exhibited in the transcription/translation and metabolism classes in *H. nigripes* (Fig. 7A) and in the transcription/translation and cytoskeleton classes for *S. latissima* (Fig. 7B). For this last species, the T3 treatments caused the highest number of down-regulated genes (607 including 152 classified) with 60% belonging to the three classes mentioned above, followed by T1 (314 including 47 classified) and T2 (247 including 56 classified; Fig. 6 and 7). For *H. nigripes*, 600 genes were observed to be regulated in T2 including 458 genes down-regulated. A substantial portion of the classified down-regulated genes belongs to the transcription/translation and metabolism class (64%), followed by an approximately equal proportion of genes associated with photorespiration (13%), stress (11%), and transport (8%) and lesser proportions of genes associated with other functions. Genes belonging to the photorespiration/energy production class, involved either in the photosynthesis or respiration process, were found to be down-regulated in *H. nigripes* in T2 and in *S. latissima* in T2 and T3. Stress genes were down-regulated in all treatments for both species. The list of DEGs is available in the Supplementary material.

4. Discussion

The analysis of gene expression combined with the investigated physiological parameters show the ability of Arctic kelps to acclimate to a range of environmental conditions. Indeed, no negative impacts of the treatments were recorded, even according to the highest emission scenarios (SSP5-8.5). This observation confirms that these species, originating from lower latitudes, could thrive in a warmer Arctic. This corroborates the findings of Miller et al. (2024b) which shows a tolerant community level response to the

same experimental conditions. Further, this also refutes our hypothesis that the combined effects of high temperature, low salinity, and low irradiance will necessarily have a negative impact on their physiology.

4.1 Chl *a* content

We hypothesized that different species might have different responses to a changing environment. The chl *a* content of both *A. esculenta* and *S. latissima* in the meristem part of the frond showed a significant decrease from t_0 to t_{final} in the control (-45% and -70% respectively). The same trend was observed in *L. digitata* although this is not significant due to the low number of measurements (-57%, n=3 at t_{final}).The high level of chl *a* measured in early summer matches the anticipation of ice melting and the following increase in turbidity (Aguilera et al., 2002). Decreasing chl *a* content between June and August has already been reported *in situ* in Kongsfjorden for *S. latissima* (Aguilera et al., 2002) with the end of the growth period (Berge et al., 2020).

In contrast to what was observed in the control as well as in the T1 and T3 treatments, for *A. esculenta,* the chl *a* content in the warm, less saline, and lower irradiance treatment (T2) remained as high as it was at t_0. The decrease in irradiance in this treatment may explain the persistence of elevated chl *a* levels. PAR is often negatively correlated with chl *a* content as higher chl *a* can help maintain elevated photosynthetic rates under reduced PAR (e.g. McWilliam and Naylor, 1967; Zhang et al., 2014). Bartsch et al. (2016) showed that the genus *Alaria* was more abundant than *Laminaria* and *Saccharina* between 10 and 15 m depth. Despite a decrease in irradiance caused by glacial and terrestrial runoff, *A. esculenta* is the only species that extended its maximum depth (from 15 to 18 m between 1994/96 to 2014; Bartsch et al., 2016). This shift could be explained by an existing adaptation to low PAR (Niedzwiedz and Bischof, 2023), giving this species a competitive advantage at greater depth. Our findings shed light on the adaptive responses of *A. esculenta* to low light, and seemingly tolerance to low salinity and warming, suggesting that this species will most likely be able to withstand future coastal environmental conditions in the Arctic.

The chl *a* content of *L. digitata* and *S. latissima* was also not affected by the treatments. This is in agreement with the study of Diehl and Bischof (2021) where temperature (up to 10°C), combined with low salinity (down to 25) did not affect the chl *a* content of *S. latissima*. However, their growth rate in low light

conditions remained similar to the other treatments. Other physiological processes such as photosynthetic efficiency, or resource allocation, might have been altered to maintain growth rates similar to the control.

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4.2 C:N ratio

The C:N ratio of *S. latissima* was significantly lower in the T2 treatment compared to t_0. The decrease in C:N ratio seems driven by an increase in nitrogen uptake. Benthic marine macroalgae and seagrasses from temperate and tropical regions have a mean C:N ratio of 22 (Atkinson and Smith, 1983). In northern Norway,
340   Liesner et al. (2020) reported a C:N ratio of 21 for *L. digitata* which is consistent with our measurements for this species as well as for *A. esculenta*, all treatments and sampling times combined. However, *S. latissima* exhibited higher ratios with a mean of 29.7 ± 5.5 (t_0 and t_{final} of the control, T1 and T3 combined), which would suggest nitrogen limitation. While algae in the T2 treatment showed a higher nitrogen content, which is an essential nutrient playing a central role in photosynthesis and protein biosynthesis, the growth rate
345   remained similar to the other treatments. Gordillo et al. (2002) showed higher nitrogen uptake at lower salinity (50% vs. 100% seawater) in *Fucus serratus* that was explained by increased N metabolism. Thus, the higher nitrogen content found here in the low saline T2 treatment (salinity down to 28) could have resulted from increased N metabolism. Indeed, the increase in nitrogen concentration in the macroalgae can induce an increase in the activity of nitrate reductase (Korb and Gerard, 2000). This enzyme catalyzes the first step
350   in the reduction of nitrate to organic forms and protein synthesis. In fact, nitrate concentration in water was higher in the T2 treatment (1.68 ± 0.8 μM/L) than in the control (0.87 ± 0.9 μM/L, data not shown) during the duration of the experiment. Arctic coastal waters are known to be nitrate-limited (Santos-Garcia et al., 2022). The influx of fresh and potentially more nitrate-rich waters may have induced an increase in the N metabolism of *S. latissima* which was nitrogen limited. Higher nutrient input from land through coastal
355   erosion and permafrost thawing may benefit this species in various processes such as photosynthesis, biosynthesis, immunity and/or molecule transport (Campbell, 1988; Meyer et al., 2005).

4.3 Growth rate

We also hypothesized that warming may enhance the growth rate of kelp. None of the growth rates of the three study species were affected by the different treatments over the total duration of the experiment. In contrast, previous studies observed an increase of the growth rate of *S. latissima* when exposed to warmer conditions (8–10°C vs. 0–4°C under replete irradiance; Iñiguez et al., 2016; Olischläger et al., 2017; Li et al., 2020; Diehl and Bischof, 2021). This discrepancy with our results can be explained by the duration of the experiment (7 to 18 days in previous studies vs 6 weeks here), the study period, and the irradiance. Our study was performed at the end of the peak growth (mid-May to July) and after, while other studies were performed in early July or used sporophytes raised from gametophyte cultures. The growth rate of *A. esculenta* significantly decreased over time in the control, indicating the gradual end of the growth peak, with many of the kelp starting to senesce (Fig. S1A). For *S. latissima,* no significant differences were found over time in the control indicating that the experiment started after the growth peak (Fig. S1B; Berge et al., 2020). In the T3 treatment only, growth was stimulated only during the first four weeks of the experiment, suggesting that warming may have prolonged the growth rate of *S. latissima* after the end of the peak growth period. Further studies may focus on this aspect. The T2 treatment did not induce a growth stimulation suggesting a negative effect of salinity and/or low irradiance.

4.4 Gene expression

Both *H. nigripes* and *S. latissima* exhibited different gene expressions in the control compared to the treatments. The fact that treatments are not clustered separately from each other but are grouped together against the control suggests that the common factor among them, which is the increase in temperature, might be the key influencing factor.

Interestingly, and as we hypothesized, the response to these treatments differed between the two species. The analysis of DEGs shows that the low salinity and irradiance treatment (T2) had a higher impact on the number of genes regulated in *H. nigripes* while warming alone (T3) had a higher impact on genes regulation on *S. latissima*. Since no phenotypic response was observed for *S. latissima* in T3, this

suggests that the observed down-regulation might be an acclimation mechanism enabling the organism to maintain its main processes. Other parameters could be measured to validate this hypothesis (lipid content, photosynthesis rates, accessory pigment concentrations, etc). Li et al., (2020) found a regulation of genes involved in reducing the osmotic pressure under low-salinity stress in *S. latissima* (salinity of 20 vs. 30). We did not observe such results with this species nor with *H. nigripes*, most likely because the reduction in salinity was much smaller in our experiment (up to -5 here vs -10 in Li et al., 2020). However, for both species, the T2 treatment induced a down-regulation of photorespiratory genes. This is consistent with previous observations in *S. latissima* (Monteiro et al., 2019). Under stressful conditions like hyposalinity, kelp may prioritize acclimatization and survival strategies over photosynthesis. Photosynthesis was however not measured during the experiment to validate this hypothesis.

Finally, we noticed a down-regulation, rather than the expected up-regulation, of heat-shock proteins (HSP), despite their typical induction under abiotic stress (Sørensen et al., 2003). The regulation of HSP in response to salinity variations occurs to a lesser degree compared to its response to temperature changes (Monteiro et al., 2019). Considering that these species originate from lower latitudes, their current exposure to the low temperatures in the Arctic might induce stress, while future warmer waters may reduce it. For example, the increased growth rate observed for *S. latissima* could likely be a response to the increased activity of RuBisCO inducing faster energy production, and resulting in the observed N-limitation. Further, the down-regulation of stress responses may indeed be a function of reduced energy input to maintain homeostasis in warmer waters. It is important to note, however, that the high temperature exposure treatment approaches the thermal optimum for *S. latissima* (Andersen et al., 2013), and may likely be close to the peak of enzymatic activity. Thus, increased warming beyond what was tested in this experiment may begin to induce a stress response and the subsequent down-regulation of enzymatic activity.

4.5 Future prospects of *Alaria esculenta, Saccharina latissima, Laminaria digitata,* and *Hedophyllum nigripes* in the Arctic

410 Our findings support the hypothesis that *A. esculenta* is more likely to be resilient to future changes in irradiance than other kelp species. In particular, our results reveal its competitive advantage at depth, through its increased chl *a* content. However, no discernible positive impact was observed on its growth rate in low light conditions. This impact may be more evident earlier in the season, during the peak growth. *A. esculenta* seems resilient to increasing glacier and river runoff, becoming more dominant in low-light environments

415 such as greater depths (Bartsch et al., 2016). The dominance of a single kelp species in specific regions may carry ecological consequences, as reduced diversity threatens ecosystem resilience (Loreau et al., 2001).

For *L. digitata*, our results demonstrate neither negative nor positive effects of warming, low salinity, and low irradiance. Franke et al. (2021) also found no effect of a 5°C warming on the growth rate of this species (control: 5°C, warming: 10°C). However, in our study confusion with *H. nigripes* at t_0 has split the

420 data, making the analysis less robust. Indeed, the individuals could only be identified at the end of the experiment, after cutting the stipe. This led to the removal of 16 individuals from the analysis. The slight decrease in the content of chl *a* over time, as observed for the other two species in the study, could not be confirmed statistically. Bartsch et al. (2016) found that *L. digitata* was the only species that experienced a significant increase in biomass between 1994/1996 and 2014 on the entire transect they studied (from 0 to

425 15 m depth). Current and future conditions in the short term seem optimal for this species. Germination of *L. digitata* is enhanced at 9°C compared to 5°C and 15°C (Zacher et al., 2016, 2019) and its growth rate is higher at 15°C compared to 5°C and 10°C (Franke et al., 2021). Although warming alone may be beneficial to this species, its combined effects with other environmental factors as well as biotic interactions (e.g. Zacher et al., 2019) might be detrimental once a certain threshold is reached. Muller et al. (2008) found no difference

430 in the germination rate between 7°C and 12°C, but showed that germination under UV of type A and B decreased down to less than 30% at 12°C compared to almost 80% at 7°C. We acknowledge that the low sample size for *L. digitata* could obscure the observation of potential negative impacts, and, thus, future studies should aim to increase study sample sizes and the temporal variability of *L. digitata* resilience to future climate scenarios. Unfortunately, this would have been logistically challenging given our experimental

435 setup.

S. latissima is widely studied throughout the northern hemisphere. In the Arctic specifically, several studies indicate that future conditions may favor the expansion of this species. This is supported by findings of enhanced germination with temperatures up to 12°C (Muller et al., 2008) and mitigation of the negative effects of UV radiation at high temperatures (12°C; Heinrich et al., 2015). Our results reveal that *S. latissima*

440 may benefit from increasing N input from coastal erosion and permafrost thawing that could enhance immunity, photosynthesis, biosynthesis and/or molecule transport, although this was not measured in this study. *S. latissima* exhibits a high degree of polymorphism, acclimation, and genetic diversity across populations (Bartsch et al., 2008; Guzinski et al., 2016). For example, its growth shows a high phenotypic plasticity that appears to be constrained within specific seasonal growth patterns in accordance with their

445 environment of origin (Spurkland and Iken, 2011). In the Canadian Arctic, Goldsmit et al. (2021) found that suitable habitat of this species may gain 64,000 km^2 by 2050, most of this new area being in the northernmost reaches, where temperature is rising and sea ice is receding. Bartsch et al. (2016) found a 30-time increase in its biomass between 1994/1996 in 2014 at 2.5 m depth at Hansneset (Kongsfjorden, Svalbard, Norway). *S. latissima* will most likely benefit from future conditions although the capacity and time of dispersal, as well

450 as competition with other species, predation, and extreme events must be considered for population projections.

So far, *A. esculenta*, *L. digitata, and S. latissima* have adapted successfully to the shifting Arctic environment and our results suggest that they might thrive in the conditions expected for 2100. In the short term, these species may well continue to spread in this region. Regarding *H. nigripes*, Franke et al. (2021)

455 suggested a true Arctic affinity with a sporophyte growth optimum of 10°C. By 2100, this species might continue to thrive in the Arctic, as evidenced by our gene expression analysis, which suggests efficient acclimation with less stress under future scenarios.

Kelp species will, however, face more competition, grazing, and extreme events such as high sedimentation rate, ice-scouring, and marine heatwaves (Hu et al., 2020). Around Tromsø (Norway), the massive spread of

460 sea urchins may have caused the ecosystem to collapse into a barren state (Sivertsen et al., 1997). Moreover, with warming, the frequency and intensity of marine heatwaves will increase which could have important consequences on marine species of Arctic flora and fauna. These potential effects of climate change should be taken into account to better assess the future of Arctic kelp communities. It therefore appears essential to

continue to study these communities in order to predict and anticipate future changes and impacts on
fisheries, local and indigenous people, and on a global scale.

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Author contributions

AL, CM, SC, PU, SA, RS, JPG, and FG were involved in the fieldwork. AL, CM, SC, JPG, and FG designed the study. SC, PU, and FG designed the system. The experiment was conducted by AL, CM, SC, SA, RS, JPG, and FG. AL and CM performed measurements of the chl *a* content. AL performed the C:N ratio measurement and the RNA extractions. MM processed transcriptomic data. AL analyzed the data and wrote the first draft of the manuscript, which was then finalized by all co-authors.

Code and data availability

The code used to carry out the majority of bio-info processes can be found at: https://github.com/MarcMeynadier/SaccharinaHedophyllumTranscriptomic. All data (except for gene expression) are available at https://doi.pangaea.de/10.1594/PANGAEA.971349.

References

Aguilera, J., Bischof, K., Karsten, U., Hanelt, D., Wiencke, C., 2002. Seasonal variation in ecophysiological patterns in macroalgae from an Arctic fjord. II. Pigment accumulation and biochemical defence systems against high light stress. Marine Biology 140, 1087–1095. https://doi.org/10.1007/s00227-002-0792-y

Ahmed, R., Prowse, T., Dibike, Y., Bonsal, B., O'Neil, H., 2020. Recent trends in freshwater influx to the Arctic Ocean from four major Arctic-draining rivers. Water 12, 1189.

Andersen, G. S., Pedersen, M. F., and Nielsen, S. L., 2013. Temperature acclimation and heat tolerance of photosynthesis in Norwegian *Saccharina latissima* (Laminariales, Phaeophyceae), Journal of Phycology, 49, 689–700. https://doi.org/10.1111/jpy.12077

Atkinson, M.J., Smith, S.V., 1983. C:N:P ratios of benthic marine plants. Limnology and Oceanography 28, 568–574. https://doi.org/10.4319/lo.1983.28.3.0568

Bartsch, I., Paar, M., Fredriksen, S., Schwanitz, M., Daniel, C., Hop, H., Wiencke, C., 2016. Changes in kelp forest biomass and depth distribution in Kongsfjorden, Svalbard, between 1996–1998 and 2012–2014 reflect Arctic warming. Polar Biology 39, 2021–2036.

Bartsch, I., Wiencke, C., Bischof, K., Buchholz, C.M., Buck, B.H., Eggert, A., Feuerpfeil, P., Hanelt, D., Jacobsen, S., Karez, R., Karsten, U., Molis, M., Roleda, M.Y., Schubert, H., Schumann, R., Valentin, K., Weinberger, F., Wiese, J., 2008. The genus *Laminaria* sensu lato: recent insights and developments. European Journal of Phycology 43, 1–86. https://doi.org/10.1080/09670260701711376

Bates, D., Mächler, M., Bolker, B., Walker, S., 2015. Fitting linear mixed-effects models using lme4. Journal of Statistical Software 67. https://doi.org/10.18637/jss.v067.i01

Bauer, S., Grossmann, S., Vingron, M., Robinson, P.N., 2008. Ontologizer 2.0—a multifunctional tool for GO term enrichment analysis and data exploration. Bioinformatics 24, 1650–1651.

Berge, J., Johnsen, G., Cohen, J., 2020. Polar night marine ecology. Advances in Polar Ecology 4.

Bischof, K., Hanelt, D., Wiencke, C., 1999. Acclimation of maximal quantum yield of photosynthesis in the brown alga *Alaria esculenta* under high light and UV radiation. Plant Biology 1, 435–444. https://doi.org/10.1111/j.1438-8677.1999.tb00726.x

Bolger, A.M., Lohse, M., Usadel, B., 2014. Trimmomatic: a flexible trimmer for Illumina sequence data. Bioinformatics 30, 2114–2120.

Bolton, J. J. and Lüning, K., 1982. Optimal growth and maximal survival temperatures of Atlantic Laminaria species (Phaeophyta) in culture, Mar. Biol., 66, 89–94. https://doi.org/10.1007/BF00397259

Bray, N.L., Pimentel, H., Melsted, P., Pachter, L., 2016. Near-optimal probabilistic RNA-seq quantification. Nature biotechnology 34, 525–527.

Bushmanova, E., Antipov, D., Lapidus, A., Prjibelski, A.D., 2019. rnaSPAdes: a de novo transcriptome assembler and its application to RNA-Seq data. GigaScience 8, giz100.

Campbell, W.H., 1988. Nitrate reductase and its role in nitrate assimilation in plants. Physiologia Plantarum 74, 214–219. https://doi.org/10.1111/j.1399-3054.1988.tb04965.x

Dankworth, M., Heinrich, S., Fredriksen, S., Bartsch, I., 2020. DNA barcoding and mucilage ducts in the stipe reveal the presence of *Hedophyllum nigripes* (Laminariales, Phaeophyceae) in Kongsfjorden (Spitsbergen). Journal of Phycology 56, 1245–1254. https://doi.org/10.1111/jpy.13012

Diehl, N., Bischof, K., 2021. Coping with a changing Arctic: mechanisms of acclimation in the brown seaweed *Saccharina latissima* from Spitsbergen. Marine Ecology Progress Series 657, 43–57. https://doi.org/10.3354/meps13532

Diehl, N., Roleda, M. Y., Bartsch, I., Karsten, U., and Bischof, K., 2021. Summer Heatwave Impacts on the European Kelp *Saccharina latissima* Across Its Latitudinal Distribution Gradient, Front. Mar. Sci., 8, https://doi.org/10.3389/fmars.2021.695821

Eggert, A., 2012. Seaweed Responses to Temperature, in: Wiencke, C., Bischof, K. (Eds.), Seaweed biology: novel insights into ecophysiology, Ecology and Utilization, Ecological Studies. Springer, Berlin, Heidelberg, pp. 47–66. https://doi.org/10.1007/978-3-642-28451-9_3

Filbee-Dexter, K., Wernberg, T., Fredriksen, S., Norderhaug, K.M., Pedersen, M.F., 2019. Arctic kelp forests: diversity, resilience and future. Global and Planetary Change 172, 1–14. https://doi.org/10.1016/j.gloplacha.2018.09.005

Finn, R.D., Clements, J., Eddy, S.R., 2011. HMMER web server: interactive sequence similarity searching. Nucleic Acids Research 39, W29–W37. https://doi.org/10.1093/nar/gkr367

Franke, K., Liesner, D., Heesch, S., Bartsch, I., 2021. Looks can be deceiving: contrasting temperature characteristics of two morphologically similar kelp species co-occurring in the Arctic. Botanica Marina 64, 163–175. https://doi.org/10.1515/bot-2021-0014

Gene Ontology Consortium, 2015. Gene ontology consortium: going forward. Nucleic acids research 43, D1049–D1056.

Goldsmit, J., Schlegel, R.W., Filbee-Dexter, K., MacGregor, K.A., Johnson, L.E., Mundy, C.J., Savoie, A.M., McKindsey, C.W., Howland, K.L., Archambault, P., 2021. Kelp in the eastern Canadian Arctic: current and future predictions of habitat suitability and cover. Frontiers in Marine Science 18, 742209. https://doi.org/10.3389/fmars.2021.742209

Gordillo, F., Dring, M., Savidge, G., 2002. Nitrate and phosphate uptake characteristics of three species of brown algae cultured at low salinity. Marine Ecology Progress Series 234, 111–118. https://doi.org/10.3354/meps234111

Grabherr, M.G., Haas, B.J., Yassour, M., Levin, J.Z., Thompson, D.A., Amit, I., Adiconis, X., Fan, L., Raychowdhury, R., Zeng, Q., 2011. Full-length transcriptome assembly from RNA-Seq data without a reference genome. Nature biotechnology 29, 644–652.

Guzinski, J., Mauger, S., Cock, J.M., Valero, M., 2016. Characterization of newly developed expressed sequence tag-derived microsatellite markers revealed low genetic diversity within and low connectivity between European *Saccharina latissima* populations. Journal of Applied Phycology 28, 3057–3070. https://doi.org/10.1007/s10811-016-0806-7

560   Haas, B., Papanicolaou, A., 2015. TransDecoder 5.5. 0.

Heinrich, S., Frickenhaus, S., Glöckner, G., Valentin, K., 2012. A comprehensive cDNA library of light- and temperature-stressed *Saccharina latissima* (Phaeophyceae). European Journal of Phycology 47, 83–94. https://doi.org/10.1080/09670262.2012.660639

Heinrich, S., Valentin, K., Frickenhaus, S., Wiencke, C., 2015. Temperature and light interactively modulate
565   gene expression in *Saccharina latissima* (Phaeophyceae). Journal of Phycology 51, 93–108. https://doi.org/10.1111/jpy.12255

Hop, H., Wiencke, C., Vögele, B., Kovaltchouk, N.A., 2012. Species composition, zonation, and biomass of marine benthic macroalgae in Kongsfjorden, Svalbard. Botanica Marina 55, 399–414. https://doi.org/10.1515/bot-2012-0097

570   Hu, S., Zhang, L., Qian, S., 2020. Marine heatwaves in the Arctic region: variation in different ice covers. Geophysical Research Letters 47. https://doi.org/10.1029/2020GL089329

Huerta-Cepas, J., Forslund, K., Coelho, L.P., Szklarczyk, D., Jensen, L.J., Von Mering, C., Bork, P., 2017. Fast genome-wide functional annotation through orthology assignment by eggNOG-Mapper. Molecular Biology and Evolution 34, 2115–2122. https://doi.org/10.1093/molbev/msx148

575   Huerta-Cepas, J., Szklarczyk, D., Heller, D., Hernández-Plaza, A., Forslund, S.K., Cook, H., Mende, D.R., Letunic, I., Rattei, T., Jensen, L.J., von Mering, C., Bork, P., 2019. eggNOG 5.0: a hierarchical, functionally and phylogenetically annotated orthology resource based on 5090 organisms and 2502 viruses. Nucleic Acids Research 47, D309–D314. https://doi.org/10.1093/nar/gky1085

Iñiguez, C., Carmona, R., Lorenzo, M.R., Niell, F.X., Wiencke, C., Gordillo, F.J.L., 2016. Increased
580   temperature, rather than elevated CO_2, modulates the carbon assimilation of the Arctic kelps *Saccharina latissima* and *Laminaria solidungula*. Marine Biology 163, 248. https://doi.org/10.1007/s00227-016-3024-6

Karsten, U., 2012. Seaweed Acclimation to Salinity and Desiccation Stress, in: Wiencke, C., Bischof, K. (Eds.), Seaweed biology: novel insights into ecophysiology, ecology and utilization, ecological studies. Springer, Berlin, Heidelberg, pp. 87–107. https://doi.org/10.1007/978-3-642-28451-9_5

585 Karsten, U., 2007. Research note: salinity tolerance of Arctic kelps from Spitsbergen. Phycological Research 55, 257–262. https://doi.org/10.1111/j.1440-1835.2007.00468.x

Korb, R.E., Gerard, V.A., 2000. Nitrogen assimilation characteristics of polar seaweeds from differing nutrient environments. Marine Ecology Progress Series 198, 83–92.

Krause-Jensen, D., Archambault, P., Assis, J., Bartsch, I., Bischof, K., Filbee-Dexter, K., Dunton, K.H.,
590 Maximova, O., Ragnarsdóttir, S.B., Sejr, M.K., Simakova, U., Spiridonov, V., Wegeberg, S., Winding, M.H.S., Duarte, C.M., 2020. Imprint of climate change on pan-Arctic marine vegetation. Frontiers in Marine Science 7, 617324. https://doi.org/10.3389/fmars.2020.617324

Krause-Jensen, D., Duarte, C.M., 2016. Substantial role of macroalgae in marine carbon sequestration. Nature Geoscience 9, 737–742. https://doi.org/10.1038/ngeo2790

595 Krause-Jensen, D., Marbà, N., Olesen, B., Sejr, M.K., Christensen, P.B., Rodrigues, J., Renaud, P.E., Balsby, T.J.S., Rysgaard, S., 2012. Seasonal sea ice cover as principal driver of spatial and temporal variation in depth extension and annual production of kelp in Greenland. Global Change Biology 18, 2981–2994. https://doi.org/10.1111/j.1365-2486.2012.02765.x

Kwiatkowski, L., Torres, O., Bopp, L., Aumont, O., Chamberlain, M., Christian, J.R., Dunne, J.P., Gehlen,
600 M., Ilyina, T., John, J.G., Lenton, A., Li, H., Lovenduski, N.S., Orr, J.C., Palmieri, J., Santana-Falcón, Y., Schwinger, J., Séférian, R., Stock, C.A., Tagliabue, A., Takano, Y., Tjiputra, J., Toyama, K., Tsujino, H., Watanabe, M., Yamamoto, A., Yool, A., Ziehn, T., 2020. Twenty-first century ocean warming, acidification, deoxygenation, and upper-ocean nutrient and primary production decline from CMIP6 model projections. Biogeosciences 17, 3439–3470. https://doi.org/10.5194/bg-17-3439-2020

605 Lebrun, A., Comeau, S., Gazeau, F., Gattuso, J.-P., 2022. Impact of climate change on Arctic macroalgal communities. Global and Planetary Change 103980.

Letunic, I., Khedkar, S., Bork, P., 2021. SMART: recent updates, new developments and status in 2020. Nucleic Acids Research 49, D458–D460. https://doi.org/10.1093/nar/gkaa937

Li, H., Monteiro, C., Heinrich, S., Bartsch, I., Valentin, K., Harms, L., Glöckner, G., Corre, E., Bischof, K., 2020. Responses of the kelp *Saccharina latissima* (Phaeophyceae) to the warming Arctic: from physiology to transcriptomics. Physiologia Plantarum 168, 5–26. https://doi.org/10.1111/ppl.13009

Li, W., Godzik, A., 2006. Cd-hit: a fast program for clustering and comparing large sets of protein or nucleotide sequences. Bioinformatics 22, 1658–1659. https://doi.org/10.1093/bioinformatics/btl158

Liesner, D., Fouqueau, L., Valero, M., Roleda, M.Y., Pearson, G.A., Bischof, K., Valentin, K., Bartsch, I., 2020. Heat stress responses and population genetics of the kelp *Laminaria digitata* (Phaeophyceae) across latitudes reveal differentiation among North Atlantic populations. Ecology and Evolution 10, 9144–9177. https://doi.org/10.1002/ece3.6569

Loreau, M., Naeem, S., Inchausti, P., Bengtsson, J., Grime, J.P., Hector, A., Hooper, D.U., Huston, M.A., Raffaelli, D., Schmid, B., Tilman, D., Wardle, D.A., 2001. Biodiversity and Ecosystem Functioning: Current Knowledge and Future Challenges, Science, 294, 804–808, https://doi.org/10.1126/science.1064088.

Lorenzen, C.J., 1967. Determination of chlorophyll and pheopigments: spectrophotometric equations. Limnology and Oceanography 12, 343–346. https://doi.org/10.4319/lo.1967.12.2.0343

Love, M.I., Huber, W., Anders, S., 2014. Moderated estimation of fold change and dispersion for RNA-seq data with DESeq2. Genome Biology 15, 550. https://doi.org/10.1186/s13059-014-0550-8

McWilliam, J.R., Naylor, A.W., 1967. Temperature and plant adaptation. I. Interaction of temperature and light in the synthesis of chlorophyll in corn. Plant Physiology 42, 1711–1715. https://doi.org/10.1104/pp.42.12.1711

Meyer, C., Lea, U.S., Provan, F., Kaiser, W.M., Lillo, C., 2005. Is nitrate reductase a major player in the plant NO (nitric oxide) game? Photosynthesis Research 83, 181–189. https://doi.org/10.1007/s11120-004-3548-3

Millard, S.P., 2013. EnvStats: An R Package for Environmental Statistics. Springer, New York. ISBN 978-1-4614-8455-4, https://www.springer.com.

Miller, C.A., Urrutti, P., Gattuso, J.-P., Comeau, S., Lebrun, A., Alliouane, S., Schlegel, R.W., Gazeau, F., 2024a. Technical note: An autonomous flow-through salinity and temperature perturbation mesocosm system for multi-stressor experiments. *Biogeosciences*, *21*(1), 315–333. https://doi.org/10.5194/bg-21-315-2024

Miller, C. A., Gazeau, F., Lebrun, A., Gattuso, J.-P., Alliouane, S., Urrutti, P., Schlegel, R. W., and Comeau, S., 2024. Productivity of mixed kelp communities in an Arctic fjord exhibit tolerance to a future climate, Science of The Total Environment, 930, 172571, https://doi.org/10.1016/j.scitotenv.2024.172571

Mistry, J., Chuguransky, S., Williams, L., Qureshi, M., Salazar, G.A., Sonnhammer, E.L.L., Tosatto, S.C.E., Paladin, L., Raj, S., Richardson, L.J., Finn, R.D., Bateman, A., 2021. Pfam: The protein families database in 2021. Nucleic Acids Research 49, D412–D419. https://doi.org/10.1093/nar/gkaa913

Monteiro, C.M.M., Li, H., Bischof, K., Bartsch, I., Valentin, K.U., Corre, E., Collén, J., Harms, L., Glöckner, G., Heinrich, S., 2019. Is geographical variation driving the transcriptomic responses to multiple stressors in the kelp *Saccharina latissima*? BMC Plant Biol 19, 513. https://doi.org/10.1186/s12870-019-2124-0

Müller, R., Laepple, T., Bartsch, I., Wiencke, C., 2009. Impact of oceanic warming on the distribution of seaweeds in polar and cold-temperate waters. Botanica Marina 52, 617–638. https://doi.org/10.1515/BOT.2009.080

Müller, R., Wiencke, C., Bischof, K., 2008. Interactive effects of UV radiation and temperature on microstages of Laminariales (Phaeophyceae) from the Arctic and North Sea. Climate Research 37, 203–213. https://doi.org/10.3354/cr00762

Muth, A.F., Bonsell, C., Dunton, K.H., 2021. Inherent tolerance of extreme seasonal variability in light and salinity in an Arctic endemic kelp (*Laminaria solidungula*). Journal of Phycology 57, 1554–1562. https://doi.org/10.1111/jpy.13187

Niedzwiedz, S., Bischof, K., 2023. Glacial retreat and rising temperatures are limiting the expansion of temperate kelp species in the future Arctic. Limnology and Oceanography 68, 816-830. https://doi.org/10.1002/lno.12312

Olischläger, M., Iñiguez, C., Koch, K., Wiencke, C., Gordillo, F.J.L., 2017. Increased pCO_2 and temperature reveal ecotypic differences in growth and photosynthetic performance of temperate and Arctic populations of *Saccharina latissima*. Planta 245, 119–136. https://doi.org/10.1007/s00425-016-2594-3

Parke, M., 1948. Studies on British Laminariaceae. I. Growth in *Laminaria Saccharina* (L.) Lamour. Journal of the Marine Biological Association of the United Kingdom 27, 651–709. https://doi.org/10.1017/S0025315400056071

Peterson, B.J., Holmes, R.M., McClelland, J.W., Vörösmarty, C.J., Lammers, R.B., Shiklomanov, A.I., Shiklomanov, I.A., Rahmstorf, S., 2002. Increasing River Discharge to the Arctic Ocean. Science 298, 2171–2173. https://doi.org/10.1126/science.1077445

R Core Team, 2023. R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. URL https://www.R-project.org/

Renaud, P.E., Wallhead, P., Kotta, J., Włodarska-Kowalczuk, M., Bellerby, R.G., Rätsep, M., Slagstad, D., Kukliński, P., 2019. Arctic sensitivity? Suitable habitat for benthic taxa is surprisingly robust to climate change. Frontiers in Marine Science 6, 538.

Richter-Menge, J., Overland, J.E., Mathis, J.T., Osborne, E., 2017. Arctic Report Card: Arctic shows no sign of returning to reliably frozen region of recent past decades.

Roleda, M., Wiencke, C., Hanelt, D., 2005. Response of Arctic kelp zoospores to ultraviolet and photosynthetically active radiation in relation to growth depth. 8th International Phycological Congress, 13-19 August 2005, Durban, South Africa.

Santos-Garcia, M., Ganeshram, R.S., Tuerena, R.E., Debyser, M.C.F., Husum, K., Assmy, P., Hop, H., 2022. Nitrate isotope investigations reveal future impacts of climate change on nitrogen inputs and cycling in Arctic

fjords: Kongsfjorden and Rijpfjorden (Svalbard). Biogeosciences 19, 5973–6002. https://doi.org/10.5194/bg-19-5973-2022

Scherrer, K.J.N., Kortsch, S., Varpe, Ø., Weyhenmeyer, G.A., Gulliksen, B., Primicerio, R., 2019. Mechanistic model identifies increasing light availability due to sea ice reductions as cause for increasing macroalgae cover in the Arctic: light causes arctic macroalgal increase. Limnology and Oceanography 64, 330–341. https://doi.org/10.1002/lno.11043

Shiklomanov, I.A., Shiklomanov, A.I., 2003. Climatic change and the dynamics of river runoff into the Arctic Ocean. Water Resources 30, 593–601.

Simão, F.A., Waterhouse, R.M., Ioannidis, P., Kriventseva, E.V., Zdobnov, E.M., 2015. BUSCO: assessing genome assembly and annotation completeness with single-copy orthologs. Bioinformatics 31, 3210–3212. https://doi.org/10.1093/bioinformatics/btv351

Sivertsen, K., 1997. Geographic and environmental factors affecting the distribution of kelp beds and barren grounds and changes in biota associated with kelp reduction at sites along the Norwegian coast. Canadian Journal of Fisheries and Aquatic Sciences 54, 2872–2887. https://doi.org/10.1139/f97-186

Sørensen, J.G., Kristensen, T.N., Loeschcke, V., 2003. The evolutionary and ecological role of heat shock proteins: Heat shock proteins. Ecology Letters 6, 1025–1037. https://doi.org/10.1046/j.1461-0248.2003.00528.x

Springer, K., Lütz, C., Lütz-Meindl, U., Wendt, A., Bischof, K., 2017. Hyposaline conditions affect UV susceptibility in the Arctic kelp *Alaria esculenta* (Phaeophyceae). Phycologia 56, 675–685. https://doi.org/10.2216/16-122.1

Spurkland, T., Iken, K., 2011. Salinity and irradiance effects on growth and maximum photosynthetic quantum yield in subarctic *Saccharina latissima* (Laminariales, Laminariaceae). Botanica Marina 54. https://doi.org/10.1515/bot.2011.042

Stroeve, J.C., Markus, T., Boisvert, L., Miller, J., Barrett, A., 2014. Changes in Arctic melt season and implications for sea ice loss. Geophysical Research Letters 41, 1216–1225. https://doi.org/10.1002/2013GL058951

705 Supek, F., Bošnjak, M., Škunca, N., Šmuc, T., 2011. REVIGO Summarizes and Visualizes Long Lists of Gene Ontology Terms. PLoS ONE 6, e21800. https://doi.org/10.1371/journal.pone.0021800

Vihtakari M (2023). ggOceanMaps: plot data on oceanographic maps using 'ggplot2'. R package version 2.0.4, https://mikkovihtakari.github.io/ggOceanMaps/

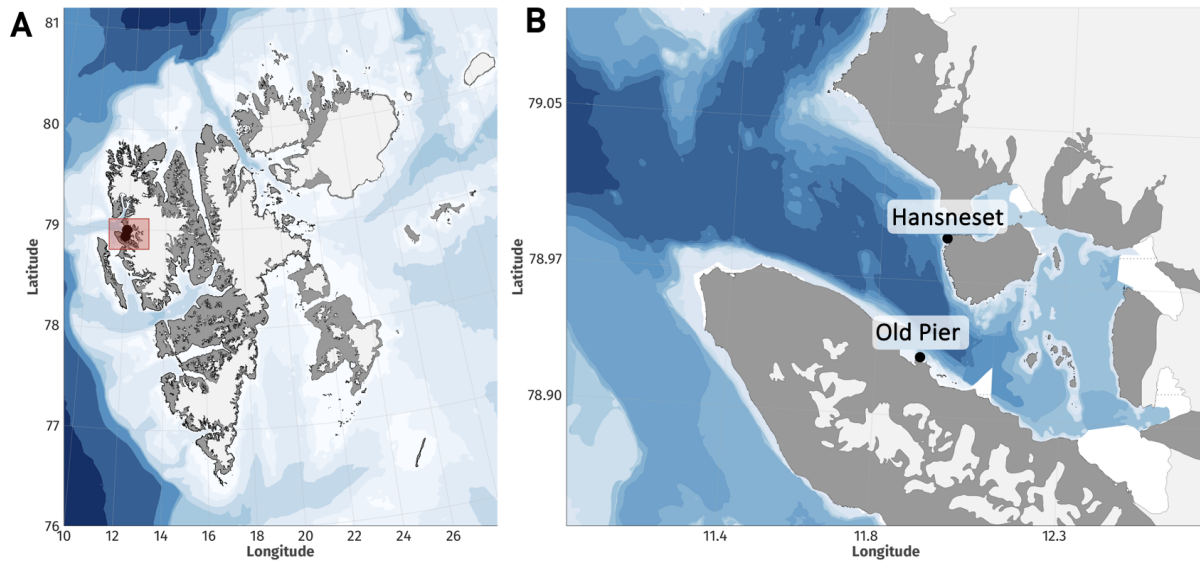
Wiencke, C., Bartsch, I., Bischoff, B., Peters, A.F., Breeman, A.M., 1994. Temperature requirements and 710 biogeography of Antarctic, Arctic and Amphiequatorial Seaweeds. Botanica Marina 37. https://doi.org/10.1515/botm.1994.37.3.247

Zacher, K., Bernard, M., Bartsch, I., Wiencke, C., 2016. Survival of early life history stages of Arctic kelps (Kongsfjorden, Svalbard) under multifactorial global change scenarios. Polar Biology 39, 2009–2020. https://doi.org/10.1007/s00300-016-1906-1

715 Zacher, K., Bernard, M., Daniel Moreno, A., Bartsch, I., 2019. Temperature mediates the outcome of species interactions in early life-history stages of two sympatric kelp species. Marine Biology 166, 161. https://doi.org/10.1007/s00227-019-3600-7

Zhang, D.-W., Yuan, S., Xu, F., Zhu, F., Yuan, M., Ye, H.-X., Guo, H.-Q., Lv, X., Yin, Y., Lin, H.-H., 2016. Light intensity affects chlorophyll synthesis during greening process by metabolite signal from mitochondrial 720 alternative oxidase in A rabidopsis: Signalling by AOX regulates greening process. Plant, Cell & Environment 39, 12–25. https://doi.org/10.1111/pce.12438

Table 1: Temperature, salinity, and photosynthetically active radiation (PAR) during the experiment. The T1 and T2 treatments represent future coastline exposed to runoff conditions, whereas the T3 treatment represents future conditions on shores not exposed to runoff. The quartiles and medians were calculated based on data acquired from 2021-07-10 for PAR and 2021-07-16 for temperature and salinity (once the targeted treatments were reached) until the end of the experiment.

Treatment	Scenario	Temperature (°C)				Salinity				PAR (μmol photons m^{-2} s^{-1})			
		Δ	1st quartile	Median	3rd quartile	Δ	1st quartile	Median	3rd quartile	Δ	1st quartile	Median	3rd quartile
Ctrl	control	*in situ*	4.8	5.3	5.8	*in situ*	33.4	33.8	34.3	*in situ*	35.1	47.8	59.5
T1	SSP2-4.5 - coastline	+ 3.3°C	8.4	8.9	9.2	- 2.5	30.8	31.0	31.8	- 20 %	27.8	36.1	43.9
T2	SSP5-8.5 - coastline	+ 5.3°C	10.3	10.8	11.2	- 5	28.2	28.5	29.5	- 30 %	23.8	31.4	40.7
T3	SSP5-8.5 - offshore	+ 5.3°C	10.3	10.8	11.2	*in situ*	33.4	33.9	34.5	*in situ*	40.3	54.8	69.9

730 **Figure 1:** The study was carried out in Svalbard (A) on kelp sampled in Kongsfjorden (B) in Hansneset and the Old Pier. Maps were created using the R package "ggOceanMaps" (Vihtakari, 2023).

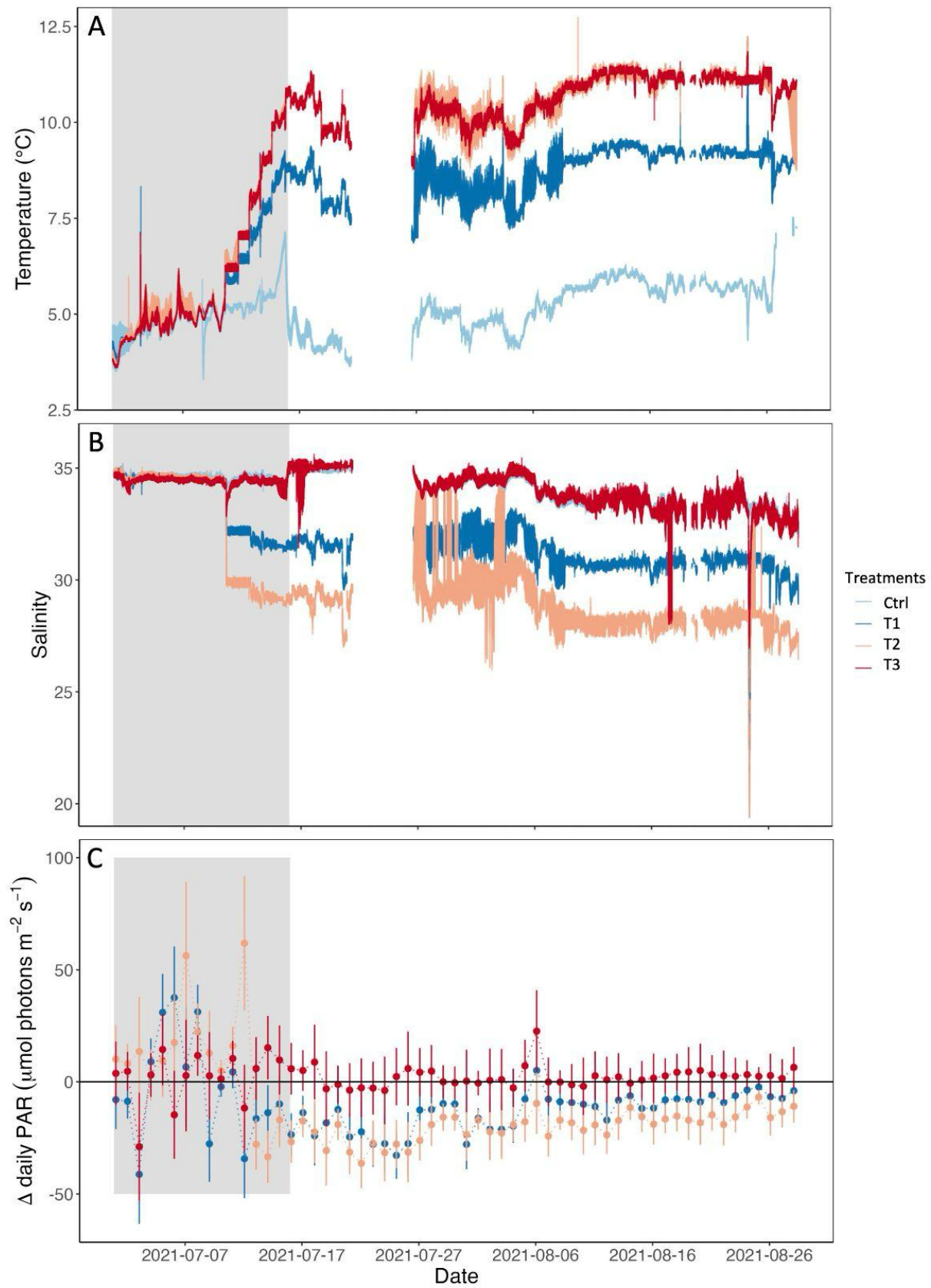



Figure 2: **A**) Temperature, **B**) salinity, and **C**) Δ Daily Photosynthetically Active Radiation (PAR) between the control and the treatments. Temperature and salinity were measured every minute. PAR values were integrated over 10-minute intervals and averaged over the day. The gray-shaded region corresponds to the beginning of the experiment, before the treatment conditions of temperature, salinity and irradiance were reached. A few days of temperature and salinity data were lost (from 2021-07-21 to 2021-07-26).

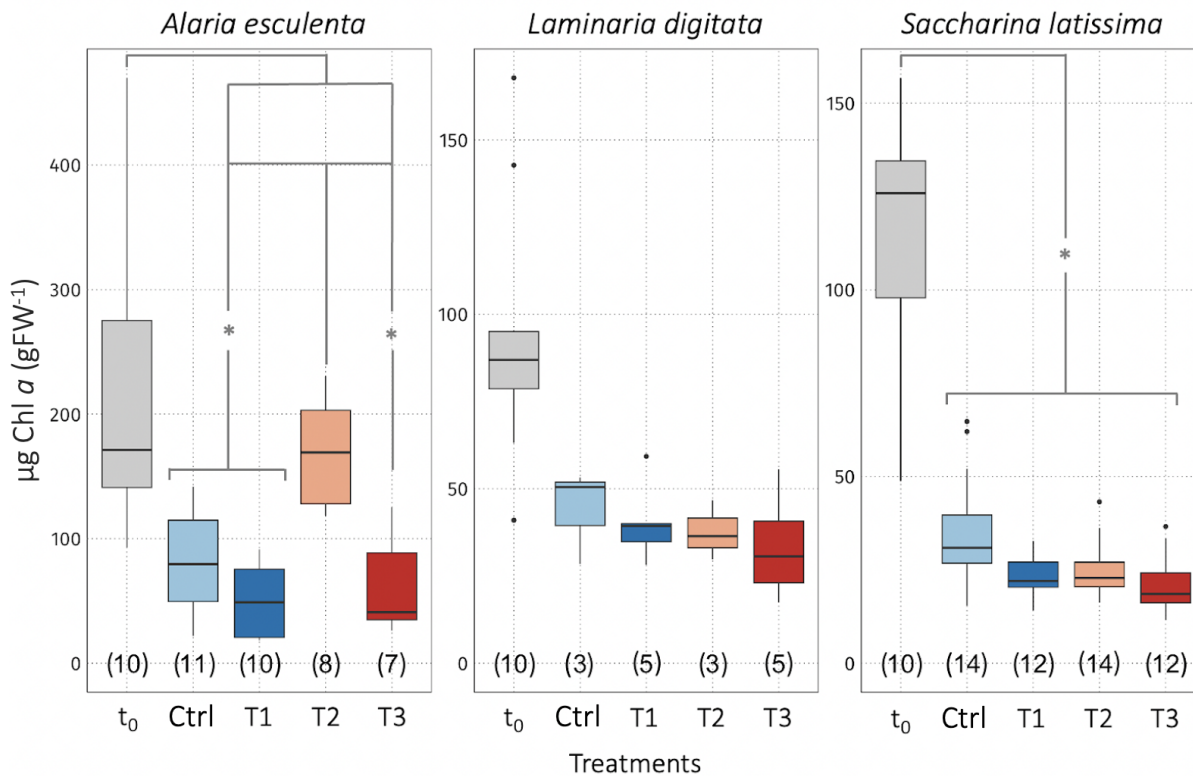



Figure 3: Chlorophyll *a* (chl *a*) content of *Alaria esculenta*, *Laminaria digitata,* and *Saccharina latissima* exposed to the four treatments, expressed per unit of fresh weight (gFW). t_0 values correspond to the chl *a* content at the start of the experiment, while Ctrl, T1, T2, and T3 correspond to the final chl *a* content of organisms maintained in the respective treatments for six weeks. The horizontal lines in each boxplot represent the median. The whiskers extend to the furthest data points within 1.5 times the interquartile range (the top and bottom of the box). Statistically significant differences are shown with an asterisk ($p < 0.05$). The number in parentheses below each boxplot corresponds to the sample size.

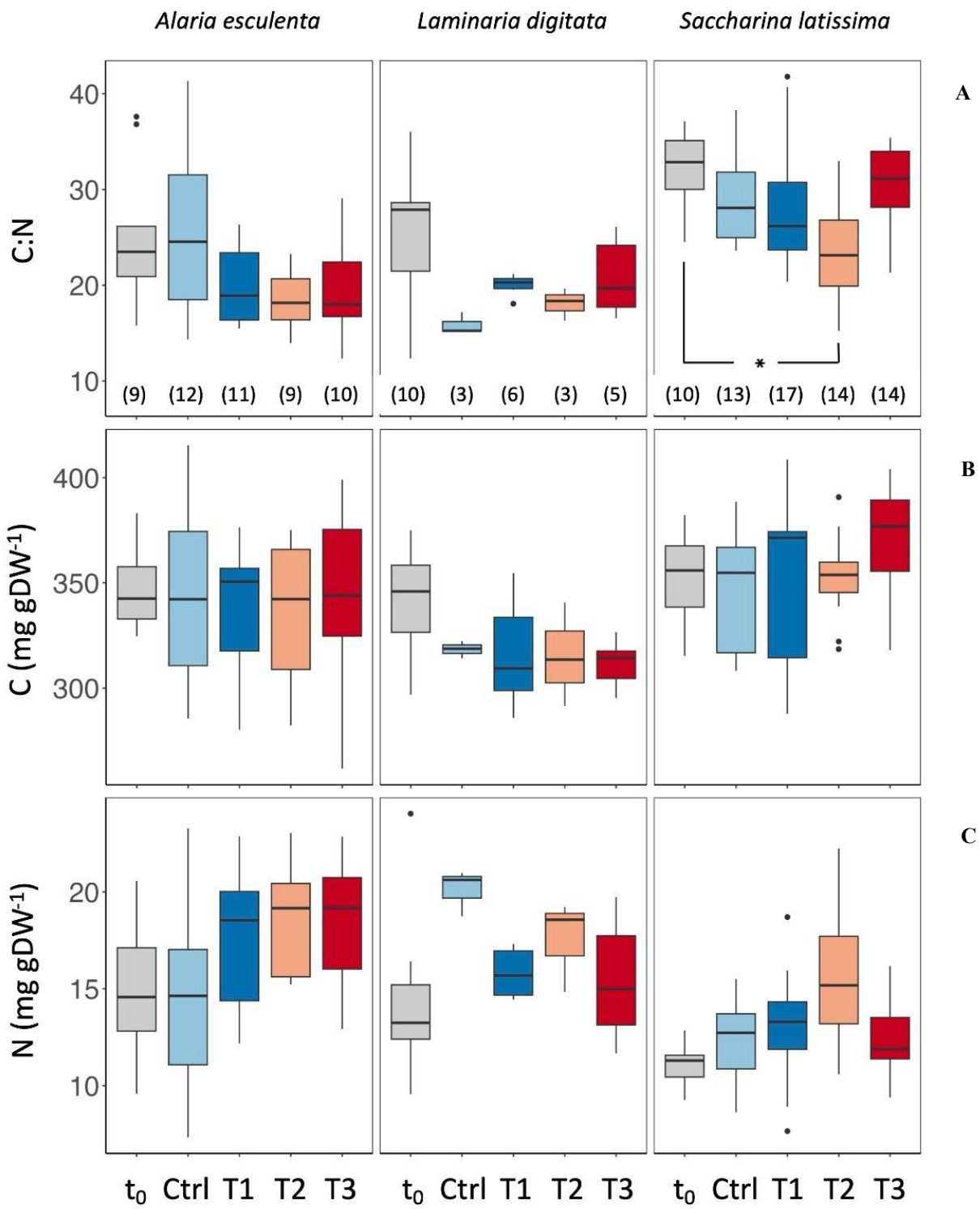



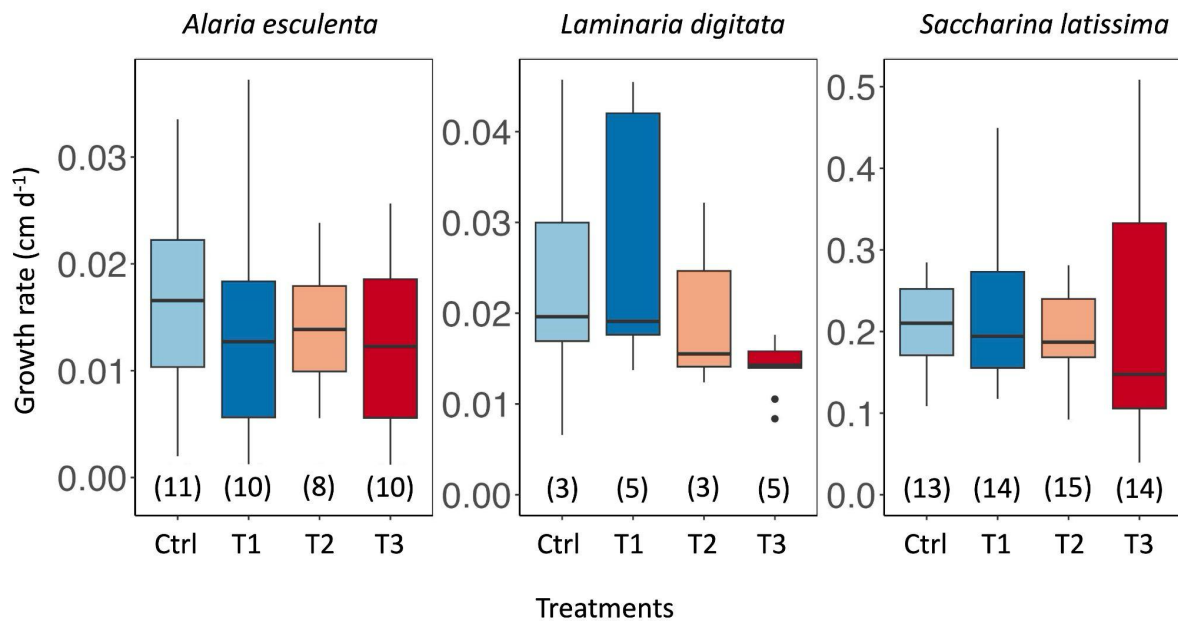
Figure 4: **A**) Carbon:nitrogen (C:N), **B**) carbon contents, and **C**) nitrogen contents of *Alaria esculenta, Laminaria digitata,* and *Saccharina latissima* exposed to the four treatments, expressed per unit of dry weight (gDW). t_0 values correspond to samples taken at the start of the experiment, while Ctrl, T1, T2, and T3 correspond to the final values from organisms maintained in the respective treatments for six weeks. The horizontal lines in each boxplot represent the median. The whiskers extend to the furthest data points within 1.5 times the interquartile range (the top and bottom of the box). Statistically significant differences are shown with an asterisk ($p < 0.05$). The number in parentheses below each boxplot in (**A**) corresponds to the sample size, respectively the same in (**B**) and (**C**).

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Figure 5: Growth rate of *Alaria esculenta*, *Laminaria digitata,* and *Saccharina latissima* exposed to the four treatments during six weeks. The horizontal lines in each boxplot represent the median. The horizontal lines in each boxplot represent the median. The whiskers extend to the furthest data points within 1.5 times the interquartile range (the top and bottom of the box). The number in parentheses below each boxplot

760 corresponds to the sample size.
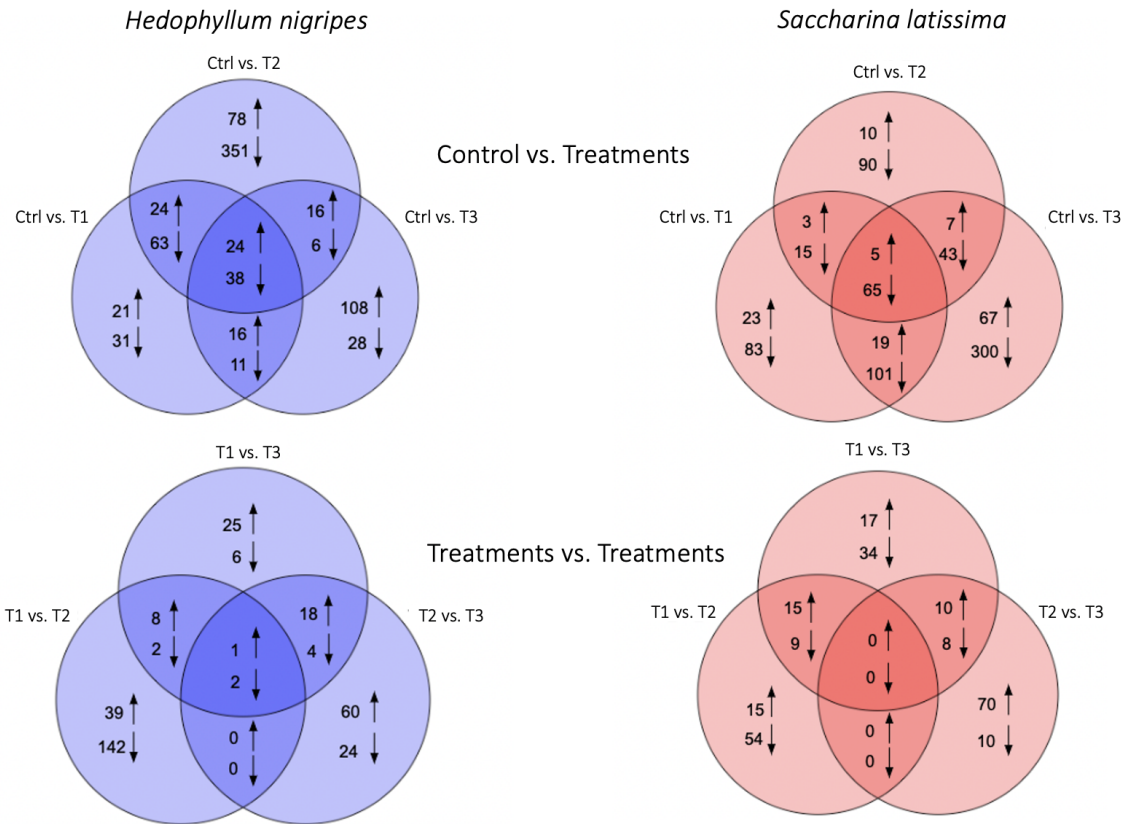
Figure 6: Venn diagrams of differentially up-regulated (↑) and down-regulated (↓) genes of *Saccharina latissima* and *Hedophyllum nigripes* between the control and the treatments (T1, T2, and T3) and between treatments.
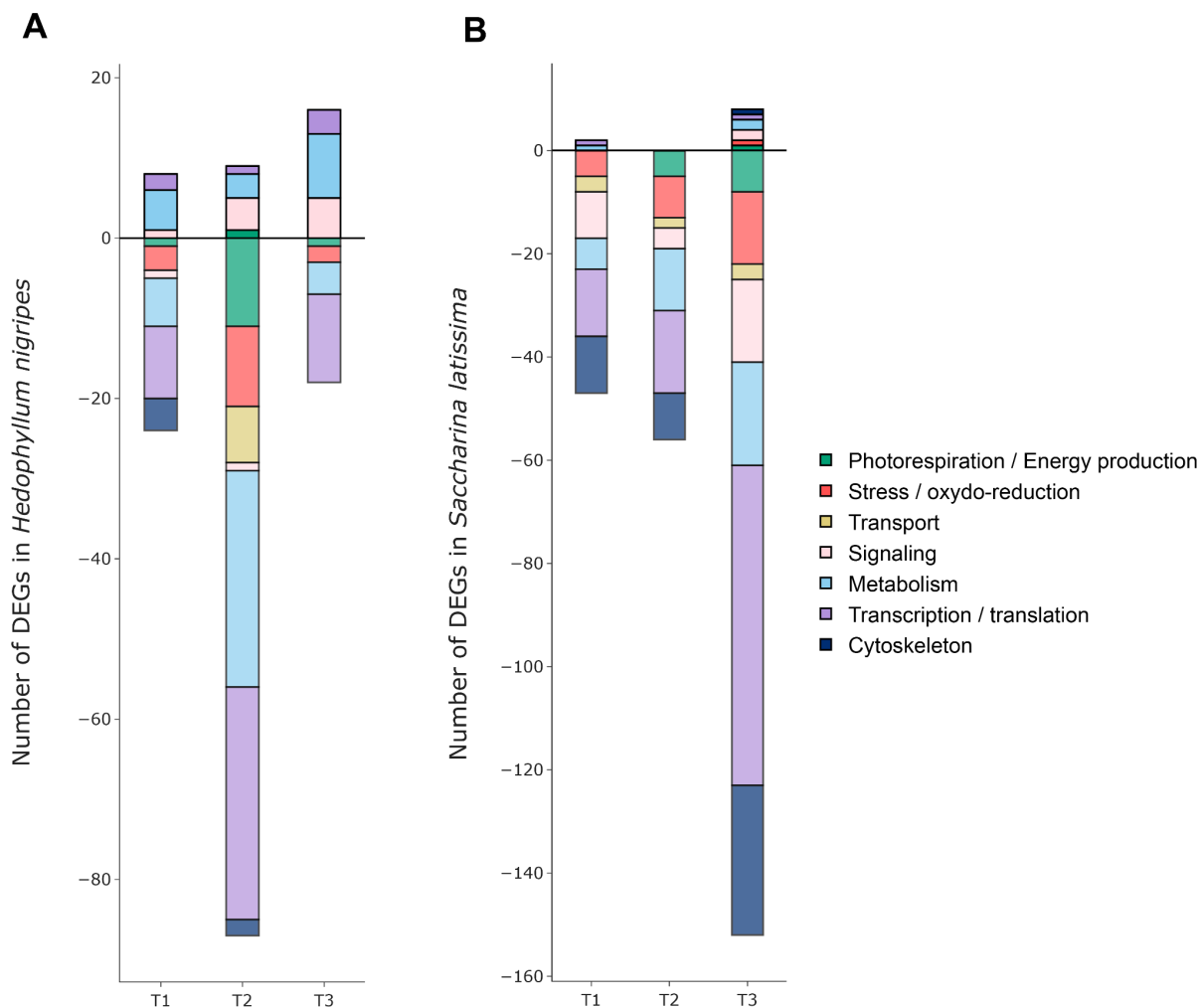
Figure 7: Number of classified differentially expressed genes (DEGs) in **A**) *Hedophyllum nigripe*s and **B**) *Saccharina latissima* in response to T1, T2, and T3. The upper part of the graph displays up-regulated DEGs and the lower part down-regulated DEGs. Genes were classified with their Pfam and EggNOG annotations (see 2.7).