Impacts of atmospheric circulation patterns and cloud inhibition on aerosol radiative effect and boundary layer structure during winter air pollution in Sichuan Basin, China

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16 Abstract. Winter persistent aerosol pollution frequently occurs in the Sichuan Basin (SCB) due to its unfavorable weather conditions, such as low wind, wetness, and cloudiness. Based on long-term 17 18 observational data analyses from 2015-2021, it has been found that the four representative stations in 19 the SCB often simultaneously experience PM_{2.5} pollution accompanied by variations in meteorological 20 conditions above 850 hPa, which indicates a connection between regional winter air pollution in the 21 SCB and large-scale synoptic patterns. The dominant 850 hPa synoptic patterns of winter SCB were 22 classified into six patterns using T-model principal component analysis: (1) strong high pressure in the 23 north, (2) east high west low (EHWL) pressure, (3) weak high pressure in the north, (4) weak ridge of 24 high pressure after the trough, (5) low trough (LT), and (6) strong high pressure. Pattern 2 25 characterized with EHWL pressure system, and Pattern 5 featured with LT, were identified as key synoptic patterns for the beginning and accumulation of pollution processes. Pattern 1, characterized by 26 27 a strong high pressure in the north, was the cleanest pattern associated with reduced PM_{2.5} 28 concentrations. The EHWL and LT patterns were associated with a remarkably high cloud liquid 29 content, attributed to upper southerly winds introducing humid air. Clouds reduce solar radiation 30 through reflection and scattering, resulting in more stable stratification and aerosol accumulation. This 31 cloud radiation interaction (CRI) was more pronounced in the LT pattern due to denser isobaric lines 32 and stronger southerly winds than in the EHWL pattern. Numerical simulation experiments utilizing 33 WRF-Chem indicated that there is a upper-level heating during afternoon and surface cooling in the 34 morning forced by the aerosol radiation interaction (ARI) under the EHWL and LT patterns. 35 Additionally, strong surface cooling in the evening influenced by valley winds could be found. With 36 wet and cloudy synoptic forcing, ARI directly affects the stability of the boundary layer and is 37 modulated through CRI inhibition. For example, Chongqing exhibited lower PM2.5 concentrations and stronger ARI compared to the western and southern SCB due to lower cloud liquid content and weaker 38 39 CRI inhibition on ARI. The CRI inhibition caused a 50 % reduction in solar radiation and boundary

40 layer height during the daytime under the LT pattern, which was larger than that under the EHWL 41 pattern. This study comprehensively analyzed the spatial disparities in cloud inhibition on ARIs, their 42 impacts on the boundary layer structure, and the discrepancies of these interactions under different 43 synoptic patterns during pollution processes. The findings hold important implications for effective 44 management of pollution processes in cloudy and foggy weather.

Key words: Synoptic patterns, Cloud radiation interaction inhibition, Aerosol radiation interaction,
Boundary layer structure, Sichuan Basin.

47 1 Introduction

48 Particulate matter (PM) pollution has become a significant environmental concern in China (Xie et al., 2016a; 2016b; Che et al., 2019). High concentrations of aerosols not only worsen air quality 49 and pose serious health risks to residents, but also have implications for weather and climate 50 51 through their effects on radiation and clouds (Li et al., 2019; Zhao et al., 2020; Alexeeff et al., 52 2021; Yang et al., 2021). The interactions between aerosols and clouds present the largest 53 uncertainty in anthropogenic radiative forcing of the Earth's climate (Liao et al., 2017; Haywood 54 et al., 2021). Studying interactions among cloud, aerosol and radiation from an air quality perspective 55 is crucial for a scientific understanding of relationship between weather and pollution.

56 Excessive emissions are the essential cause of air pollution, with primary aerosol and secondary 57 aerosol formation playing significant roles in comprehending the complete picture of air pollution 58 (Peng et al., 2021). Besides, meteorological conditions not only influence on the formation of 59 secondary aerosols, but also govern the transportation and distribution of both primary and secondary aerosols, and thereby impact regional and long-range air pollution (Zhu et al., 2018; 60 61 Luo et al., 2018; Nichol et al., 2020; Zhang et al., 2020; Jiang et al., 2021). PM and gaseous 62 pollutants, primarily transported by the planetary boundary layer (PBL), are directly or indirectly influenced by various meteorological factors such as wind, relative humidity, PBL height (PBLH), 63 64 and solar radiation. These factors contribute to the multi-temporal and spatial distribution characteristics through vertical and horizontal diffusion, physicochemical reactions, and dry and 65 wet deposition (Park et al., 2017; Shu et al., 2017; Zhan et al., 2019; Huang et al., 2019). Large-66 67 scale synoptic forcing is considered the primary driving condition for meteorological factors, PBL 68 structure, and the resulting distribution of atmospheric pollutants (Miao et al., 2019; Ning et al., 69 2019; Jiang et al., 2020; Li et al., 2021). Specific synoptic patterns can induce advection, which 70 largely determines the local PBL structure and development. PBL, located at the bottom of the 71 atmosphere, is responsible for the main exchange of heat, moisture, and matter between the surface 72 and the free troposphere (Stull, 1988). The fate of pollutants emitted near the surface, a significant 73 source of aerosols in the air, is largely controlled by the PBL (Garratt, 1994). The PBLH is often used as a metric to characterize the capacity and dilution of pollutants (Seidel et al., 2010). 74 75 Synoptic patterns can directly determine the meteorological conditions of emitted pollutants and

⁷⁶ influence their transport by regulating PBL thermal stratification and mechanical turbulence (Stull,

77 1988; Ning et al., 2018; Zhan et al., 2019; Jiang et al., 2021; Zhang et al., 2022).

78 Unfavorable meteorological conditions play a significant role in contributing to aerosol pollution. 79 When pollutants accumulate to a certain degree, aerosols can reduce surface solar radiation through backscattering or absorbing solar radiation, leading to surface cooling. This decrease in 80 81 solar radiation and temperature near the ground weakens turbulent diffusion, suppresses the 82 convective development of the PBL, and lowers PBLH, which in turn exacerbates aerosol pollution (Ding et al., 2016; Wang et al., 2018). Moreover, the increase in humidity caused by the decreased 83 84 surface saturation vapor pressure and inhibited water vapor diffusion enhances aerosol 85 hygroscopic growth accelerates liquid-phase and heterogeneous reactions, and contributes to aerosol pollution (Pilinis et al., 1989). The positive feedback between unfavorable PBL 86 87 meteorology and increasing aerosols was found to be responsible for the majority of the increase 88 in PM_{2.5} during cumulative stages in various regions of eastern China affected by aerosol pollution, 89 including the North China Plain, the Guanzhong Plain, the Yangtze River Delta, the Two Lakes 90 Basin, the Pearl River Delta and the Northeast China Plain. But in the Sichuan Bain (SCB), the 91 feedback is weak due to the suppression of the cloudy mid-upper layer (Zhong et al., 2018; Zhong et al., 2019). As for the aerosol-cloud interactions, arise from increasing aerosols acting as cloud 92 93 condensation nuclei in cloud and translating into larger concentrations of smaller cloud droplets, 94 leading to an increased cloud albedo reflecting more radiation back to space (Twomey, 1977; 95 Lohmann and Feichter, 2005). Even a marginal increase in cloud droplets above pristine conditions 96 in deep convective clouds causes more droplets to reach supercooled levels, which enhances latent 97 heat release and invigorates convection (Rosenfeld et al., 2009; Possner et al., 2015). Further 98 increases in cloud droplets result in direct radiative effects, reducing downward solar radiation, 99 cooling the surface, and inhibiting convection (Scott et al., 2016).

100 The SCB is surrounded by high mountains with cloudy and wet weather conditions. The mean 101 annual relative humidity in the SCB is around 75%, with cloud fraction exceeding 80%, and an 102 average of 1200 hours of sunshine per year. The Chengdu-Chongqing city cluster in the SCB 103 serves as the economic center of the upper reaches of the Yangtze River in China, accounting for 104 approximately 10 % of the country's population. However, rapid industrialization and urbanization 105 in this region have resulted in severe air pollution. The SCB is recognized as one of the most 106 polluted regions in China, with high black carbon concentrations (Li et al., 2016; Cao et al., 2021). 107 The Qinghai-Tibet Plateau on the western edge of the SCB significantly influences the transport 108 and accumulation of pollutants through thermal and dynamic effects (Ning et al., 2017; Shu et al., 109 2021). In addition, the Oinghai–Tibet topography leads to higher cloud water content over the SCB 110 than the other regions (Yu et al., 2004; Yang et al., 2012). Many studies have emphasized the 111 importance of the interactions between cloud, aerosols and radiation in air pollution processes (Wang et 112 al., 2018; Hu et al., 2021). High pollutant emissions, combined with the prevalence of cloudy and 113 foggy weather, make these interactions in the SCB even more complex than those in other regions. The 114 aerosol radiation interactions (ARI) can be inhibited by cloud in cities like Chengdu (Zhong et al.,

115 2019). However, there is a lack of in-depth quantitative discussions regarding this aspects in the SCB. 116 On one hand, the complex terrain in the SCB leads to differences in the meteorological conditions 117 between them (Ning et al., 2017; Lu et al., 2022). For example, Chengdu is a typical basin city while 118 Chongqing is a mountain city located on the basin slope, so they have markedly different climate 119 conditions. It remains to be elucidated whether these conditions will result in spatial disparities in cloud 120 inhibition on the ARI. On the other hand, synoptic forcing, as the primary driver of meteorological 121 variations, undoubtedly play an unneglectable role in shaping cloud cover and boundary layer 122 structures (Miao et al., 2020; Wang et al., 2022; Painemal et al., 2023). The discrepancies in cloud 123 inhibition on ARI under different synoptic patterns also need to be revealed. Addressing these issues is 124 crucial for understanding the persistent pollution processes and the intricate interactions between 125 weather and pollution in the SCB. It holds important implications for the effective management of 126 pollution processes in cloudy and foggy weather.

127 Characterized with high aerosol loadings and semi-permanent cloudy weather, the SCB 128 provides an ideal region for studying the complex interactions between clouds, aerosols, and the 129 PBL. This study objectively classifies the synoptic patterns influencing the SCB based on long-130 term data. By conducting an integrated analysis of pollutants and meteorological factors, the 131 primary pollution sources and clean synoptic patterns are identified. To further investigate the 132 inhibition of cloud radiation interaction (CRI) on ARI under different synoptic patterns in the SCB, 133 WRF-CHEM simulation experiments are conducted. The results contribute to a deeper 134 understanding of CRI, ARI, and the PBL interactions in regions influenced by plateau-basin 135 topography with wet and cloudy weather. The data and methods are presented in Section 2, 136 whereas Section 3 describes the synoptic patterns and their corresponding impacts on clouds, 137 aerosols, radiation, and PBL. Finally, the conclusions are presented in Section 5.

138 2 Data and method

139 2.1 Observation data

Air quality monitoring data used in this study were obtained from air quality monitoring sites
established by the Ministry of Ecology and Environment of China across the SCB. Hourly PM_{2.5}
observations from 18 stations in the SCB were collected during the winter period from 2015 to 2021
for data analysis and model verification purposes (Fig. 1b). The abbreviations CQ, CD, MY, DY,
LS, MS, YA, ZY, ZG, YB, LZ, NJ, GA, NC, SN, GY, DZ, and BZ represent the following cities:
Chongqing, Chengdu, Mianyang, Deyang, Leshan, Meishan, Yaan, Ziyang, Zigong, Yibin,
Luzhou, Neijiang, Guangan, Nanchong, Suining, Guangyuan, Dazhou, and Bazhong, respectively.



Figure 1. (a) Three layers of simulation domains in WRF-Chem with topography map as shading; (b) the locations of 18 air quality monitoring stations (red dots) and 4 sounding stations (black dots) in the Domain 3.

The SCB has four sounding stations: Wenjiang (CD), YB, DZ, and Shapingba (CQ), situated in

the western, southern, northwestern, and eastern regions of the basin, respectively (Fig. 1b), and

represent different pollution and meteorological conditions in different regions within the SCB. In all, the air pollution over the SCB exhibits a gradual decrease from southwest to northeast. Statistical analysis indicates that the western and the southern basin experience the most severe pollution. The western basin shows the highest pollution proportion, while the southern basin exhibits the highest occurrence of heavy pollution. In the northeastern basin, specifically in DZ, heavy pollution is more likely to occur during winter, which verifies it to be the third highest pollution zone outside the western and southern basin. This makes the spatial distribution during winter differs from the overall annual pollution pattern in the SCB (Lu et al., 2022; Qi et al., 2022). Regarding meteorological conditions, research reveals that DZ has the lowest ventilation coefficient during winter, while CQ has the highest. The SCB experiences frequent temperature inversions, with CD having a higher occurrence of inversions compared to the other three cities. CD also exhibits the strongest inversion intensity and is prone to multi-layer inversions. On the other hand, YB and CQ have greater inversion thickness, while CD has the smallest inversion thickness (Feng et al., 2020). The vertical distribution of the meteorological factors used in the study was obtained from an L-band sounding radar, collecting temperature, pressure, humidity, and wind data at 00:00 and 12:00 Coordinated Universal Time (UTC) on vertical levels every second from the surface up to 30 km. Ground observation data from the four cities, including

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167 temperature and dew point temperature, were used for meteorological factor simulation verification.

168 All meteorological data were obtained from the China Weather Website Platform maintained by

169 the China Meteorological Bureau. As for the calculation of PBLH, there are various methods to

determine the PBLH, and differences in methods, data or threshold values may yield quite
different PBLH results (Seibert et al., 2000; Eresmaa et al., 2006; Jiang et al., 2021). The bulk

172 Richardson number (*Ri*) method was adopted to calculate the PBLH with sounding data in
173 the study by assuming that the PBLH is the height at which *Ri* reaches its critical value (*Rc*).
174 *Ri* at a certain height *h* is calculated as follows:

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$$Ri = \frac{(g / \theta_{v0})(\theta_{vh} - \theta_{v0})h}{u_h^2 + v_h^2}$$

Where g is the acceleration of gravity, θ_{v0} and θ_{vh} are the virtual potential temperature at surface and the height *h*, respectively, and u_h and v_h are the meridional and zonal wind components at *h*. We adopted the *Ri* method and *Rc* to be 0.25, because the EAR5 and YSU schemes use the same method and threshold value when calculating PBLH (Hong et al., 2006; ECMWF, 2017).

180 CD, YB, DZ and CQ were selected as representative cities for analysis in the study. These four 181 representative cities are located in the western, southern, northwestern, and eastern regions of the basin, 182 to capture diverse pollution and meteorological conditions within the SCB. These cities were chosen to 183 represent the most polluted regions (Zhao et al., 2018; Lu et al., 2022), as well as typical basin and 184 mountainous cities. Furthermore, there are only four sounding stations in the SCB available, which are 185 located in these four representative cities. They can provide valuable vertical and surface 186 meteorological observations, as well as pollution data, contributing comprehensive dataset used in this 187 study.

188 ERA5 reanalysis data from the European Centre for Medium-Range Weather Forecasts (ECMWF), 189 which assimilates comprehensive observation data, including ground observation, sounding data, 190 aircraft observation data, and satellite observation data, were obtained for synoptic pattern 191 classification and their impact on meteorological factors in four representative cities. The EAR5 192 data at the 850 hPa pressure level were collected for the synoptic pattern study. Additionally, cloud 193 liquid water content and downward solar radiation derived from the EAR5 single-level datasets 194 were obtained to assess the influences of synoptic forcing on CRI studies, while PBLH were 195 adopted to conduct the simulation verification. Previous studies have demonstrated the reliability 196 of ERA5 data in estimating cloud properties, including the cloud liquid content (Yao et al., 2019; 197 Nandan et al., 2022; Ojo et al., 2023).

198 2.2 Synoptic pattern classification

199 The objective classification was conducted on the synoptic patterns of the SCB using ERA5 data, 200 including geopotential height, u, and v components of winds at the 850 hPa pressure level. The 201 analysis covered an area of 97–117° E and 24–37° N with a horizontal resolution of $0.25^{\circ} \times 0.25^{\circ}$. 202 Given that PM pollution in the SCB is primarily prevalent during winter months (Zhao et al., 2018; 203 Lu et al., 2022), the synoptic pattern classification was performed for winter seasons from 2015 to 204 2021 (December, January, and February) using the principal component analysis in the T-model 205 (T-PCA) objective method. Compared with the subjective classification method, the objective method can process large amounts of data without relying on subjective experience (Huth et al., 206 207 2008; Miao et al., 2017). Among various classification methods, the T-PCA method accurately reflects the characteristics of the original synoptic circulations and exhibits spatial and temporal stability (Huth et al., 1996; Huth et al., 2008). Consequently, the T–PCA has been widely used in synoptic pattern classification researches (Ning et al., 2019; Miao et al., 2020; Li et al., 2021).

211 **2.3 Model configuration and simulation experiments**

212 To understand the combined effects of synoptic patterns and CRI inhibition on ARI and PBL, a 213 series of parallel experiments were conducted on the simulation of a typical pollution episode 214 using the Weather Research and Forecasting model with Chemistry (WRF-Chem v3.9.1) (Grell et 215 al., 2005). The Advanced Research WRF (ARW) dynamics solver integrates the compressible, 216 nonhydrostatic Euler equations, for example, the momentum equation, the continuity equation, the 217 thermodynamic equation, the moisture equation and the ideal-gas equation of state (Skamarock et 218 al., 2008). The model domain (Fig. 1a) was centered over the SCB and utilized three layers of nested grids with horizontal resolutions of 27, 9, and 3 km, respectively. A total of 32 vertical 219 220 layers spanning from the surface to 100 hPa were defined. Initial and boundary meteorological fields 221 were obtained from the National Centers for Environmental Prediction Final reanalysis data with a horizontal resolution of $1^{\circ} \times 1^{\circ}$ and 6 h time interval. For chemical process simulations, 222 223 anthropogenic emissions were sourced from the Multiresolution Emission Inventory for China 224 (MEIC) in 2016, featuring a grid resolution of $0.25^{\circ} \times 0.25^{\circ}$. To address the empirically 225 overestimated PM2.5 emissions by the MEIC in the SCB (Zhan et al., 2023), the ensemble square 226 root Kalman filter were implemented on the PM_{2.5} emission during simulation (Wu et al., 2018; 227 Lu et al., 2021). Biogenic emissions were calculated online using the Guenther scheme (Guenther 228 et al., 2006). Table 1 provides a summary of the chosen physical and chemical parameterization 229 schemes. The parameterization schemes employed in this study is the one used by the Chongqing 230 Meteorological Bureau in the daily operational activities. The schemes have been obtained 231 through multiple sets of control experiments and are considered suitable for the simulation in the 232 SCB.

Items	Contents
Domains (x, y)	(155, 110), (184, 160), (320, 250)
Grid spacing (km)	27, 9, 3
Center	(29.1° N, 106.2° E)
Time step (s)	60
Microphysics	WRF Single-Moment 5 class (WSM5) scheme
Longwave radiation	RRTMG scheme (Iacono et al., 2008)
Shortwave radiation	RRTMG scheme (Iacono et al., 2008)
Planetary boundary layer	Younsei University scheme (Hong et al., 2006)
Land surface	United Noah land surface model (Tewari et al., 2004)
Cumulus parameterization	Grell–Freitas ensemble scheme

233 Table 1 The main options of WRF-Chem

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(Grell et al., 2013)	
fifth- and third-order differencing for horizontal and vertical	
advection respectively	
Fast-J photolysis (Fast et al., 2006)	
RADM2 (Stockwell et al., 1990)	
MADE/SORGAM (Schell et al., 2001)	

To assess the impact of CRI inhibition on ARI under typical synoptic pollution patterns, four parallel experiments were conducted using simulation models. The selected simulation period for these experiments was January 1-7, 2017. The period was selected for two reasons: the Chinese government announced clean-air action in the year of 2013, aiming to reduce PM_{2.5} concentrations in the next 5 year. Specifically, the year of 2017 was identified as a key year for assessing PM_{2.5} pollution in China, as significant practical actions were implemented during the period (Wang et al., 2020) and the selected period encompassed both typical pollution and clean weather patterns.

242 The baseline experiment (BASE) included both CRI and ARI in the simulations. In contrast, the 243 three sensitivity experiments focused on excluding either ARI or CRI. Experiment 1 (EXP1) did not 244 consider ARI, Experiment 2 (EXP2) did not include CRI, and Experiment 3 (EXP3) omitted ARI 245 when CRI was not included. The differences between BASE and EXP1 represented the 246 disturbances caused by ARI, while EXP2 and EXP3 represented the influences of ARI without 247 CRI inhibition. Detail differences between the experiments could be found in Table 2. The 248 numerical experiments were initiated at 00:00 UTC on December 30, 2016, and ran until 00:00 249 UTC on January 8, 2017, with the first 48 hours designated as a model spin-up period.

250 Table 2 Four numerical simulation experiments are conducted in the study

Experiments	Description	Results	Meaning
BASE	Baseline simulation	BASE-EXP1 Disturbances by ARI	
EXP1	Only shutting down ARI		
EXP2	Only shutting down CRI Influenc		Influences of ARI
EXP3	Shutting down both ARI and CRI	EAF 2-EAF 3	without CRI

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*ARI: aerosol radiation interaction; CRI: cloud radiation interaction

252 3 Results and discussions

253 3.1 Relationships between synoptic patterns and PM_{2.5} pollution in the SCB

Figure 2 illustrates the daily mean variations in PM_{2.5} concentration and vertical distributions of potential temperature (PT) during winter period from 2015 to 2021, with a focus on the pollution episodes. The four sounding stations located in separate areas of the SCB (CD, YB, CQ, and DZ), consistently experienced pollution processes characterized by simultaneous changes in vertical thermal structures. For example, during the pollution events in January 2017 and December 2020, 259 the PM_{2.5} concentrations in all four cities reached their peak levels at the same time before rapid 260 declining (Fig. 2). Interestingly, these pollution episodes were accompanied by warming in the 261 upper layer atmosphere, while a decrease in PM_{2.5} concentration correlated with cooling. Despite 262 the significant distances between these cities (approximately 200-400 km), the synchronized 263 changes in pollutant concentrations and vertical thermal structures could be attributed to large-264 scale synoptic patterns (Miao et al., 2020; Li et al., 2021). While the four cities with sounding 265 stations were selected as representatives for vertical thermal structure analysis, other cities in the 266 SCB also experienced pollution episodes and relevant physical processes, except for GY (Fig. S1). 267 GY is located in the northern edge of the SCB, bordering Shaanxi and Gansu Provinces. The proportion 268 of heavy PM_{2.5} pollution in GY is the lowest in the basin, but the proportion of PM₁₀ pollution is higher 269 than other cities of SCB (Lu et al., 2022). Due to the lower PM2.5 concentration, the two pollution 270 processes in January 2017 in GY were not as significant as in other cities whithin the basin. However, 271 the warming of upper air coincided with PM_{2.5} increase could still be observed.



Figure 2. Time series of daily mean $PM_{2.5}$ and potential temperature derived from the sounding data during 2015-2021 winter months. The $PM_{2.5}$ pollution episodes are marked with black dotted boxes.

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273 The time series of daily mean PM_{2.5} from air quality monitoring sites and the accompanying 274 vertical distributions of temperature, relative humidity, and wind in CD, CQ, YB, and DZ derived from 275 the sounding stations are shown in Fig. S2, focusing on January 2017 as an example for analysis. During this month, two severe PM2.5 pollution episodes occurred: one from January 1 to 7 and another 276 277 from January 24 to 31 in 2017. These pollution episodes had a significant impact on air quality in all 278 four cities. The highest daily PM2.5 concentrations recorded during these episodes were 291.17 µg/m3 in CD and 276.2 μ g/m³ in YB. Pollution in early January exhibited a gradual increase in PM_{2.5} levels from 279 280 January 1 to 3, with upper air warming and the emergence of an inversion above the PBL. Additionally, 281 lower humidity and higher wind speeds above 1500 m were observed during the pollution 282 accumulation period. Similarly, the late January pollution episode showed a rapid increase in PM_{2.5}, 283 from January 24 to 27, together with warming, dryness, and high wind speed above 1500 m in all four

284 cities. These consistent meteorological conditions during the pollution periods indicated significant 285 synoptic forcing. The previous study has found that winter heavy pollution processes in the SCB are 286 usually associated with abnormal warming above the 850 hPa (Lu et al., 2022). The warming is 287 induced by strong southerly airflow above the basin. The southerly airflow in winter over the SCB 288 originates from the Yunnan-Guizhou Plateau or the Indian Peninsula, characterized with high 289 temperature, dryness, and high wind speed. The strong southerly airflow forms a warm lid over the 290 basin, suppressing the vertical exchange of pollutants within the basin. As a result, pollutants 291 accumulate rapidly, which may explain the phenomenon of rapid $PM_{2.5}$ growth accompanied by 292 warming, dryness, and strong winds above 1500 m. Notably, the key layer for studying the connection 293 between synoptic patterns and PM_{2.5} pollution is approximately 850 hPa, corresponding to a height of 294 approximately 1500 m within the PBL, where changes in specific meteorological conditions primarily 295 affect surface-emitted pollutants.

296 Using ERA5 reanalysis data for winter (December, January, and February) from 2015 to 2021, 297 the 850 hPa synoptic patterns over the SCB were objectively classified into six types (Fig. 3). 298 According to the relative positions of the high-pressure and low-pressure systems in the basin, 299 these synoptic patterns could be described as follows: (1) strong high pressure in the north, (2) east 300 high west low (EHWL) pressure, (3) weak high pressure in the north, (4) weak ridge of high 301 pressure after the trough, (5) low trough (LT), and (6) strong high pressure. Patterns 1 and 3 302 exhibited high pressure in the northern SCB, which differed from the high-pressure intensity. With 303 strong high pressure, the basin was primarily controlled by northerly airflow. Under weak high-304 pressure conditions, the basin was dominated by an easterly backflow. Patterns 2 and 5 had high 305 and low pressures near the basin, forming a relatively dense isopotential altitude gradient and 306 resulting in strong southerly winds over 850 hPa. Pattern 4 was a weak high-pressure ridge after a 307 trough controlled the SCB with sparse isobaric lines and weak winds leading to static and stable 308 weather conditions. During Pattern 6, the SCB was controlled by the cold high-pressure system, 309 accompanying weak northerly airflow on the basin. Pattern 6 usually evolved from either Pattern 1 310 or Pattern 3.



Figure 3. The 850hPa geopotential height field (shading) with wind vector fields (blue vectors), and frequency of occurrence for 6 synoptic patterns during 2015-2021 winter months. The SCB was outlined with an altitude contour of 750 m terrain height (black lines).

Patterns 2, 4, and 5 exhibited higher frequencies of pollution occurrence (PM_{2.5} daily 312 313 concentration $\geq 75 \ \mu g/m^3$) according to statistical results from 18 cities in the SCB during the 314 2015–2021 winters (Fig. 4a). These patterns were associated with high PM_{2.5} concentrations in 50– 315 70 % days, including CD, DY, and MY in the northern SCB, 40-60 % for cities in the southern 316 SCB, such as ZG and YB, and also 40-60 % of days for cities in the northern SCB, such as CQ, DZ, 317 NC, and GA. Furthermore, the average $PM_{2.5}$ concentrations in the respective cities for the six 318 synaptic patterns were calculated (Fig. 4b), aligning with the frequency of pollution occurrence. 319 The days under Patterns 2, 4, and 5 exhibited higher average daily PM_{2.5} concentrations. The average 320 concentrations under these three synoptic patterns were 99.19, 103.43, and 111.97 μ g/m³ for CD, 321 95.44, 87.98, and 94.26 μg/m³ for YB, 79.14, 83.96, and 74.77 μg/m³ for CQ, and 91.02, 104.64, 322 and 91.51 μ g/m³ for DZ, respectively. Regarding the impact of synoptic patterns on the 323 accumulation or dispersion of PM2.5, Fig. 4c illustrates the average daily changes in PM2.5 324 concentration compared with the previous day for CD, CQ, YB, and DZ under the six synoptic 325 patterns. Patterns 2 and 5 exhibited the most significant PM2.5 accumulation under the influence of 326 southerly airflow. The average PM_{2.5} concentration under Pattern 1, 3 and 6 was lower in all cities 327 of SCB than other three pollution patterns (Fig. 4a). Besides, the day to day PM_{2.5} variations under Pattern 1, 3 and 6 exhibited negative growth trend in the four representative cities (Fig. 4c). As a 328 329 result, Pattern 1,3 and 6 were identified as the "clean pattern". In addition, the pollution 330 occurrence frequency of which was found higher for cities located in the eastern part of the SCB 331 than other parts. Under Pattern 6, strongest northerly airflow affects the basin. The eastern part of the basin consists of parallel ridges and valleys, which reduces wind speed. The stronger the wind 332 333 is, the more obvious the reduction of wind by terrain is. In contrast, the western part is relatively 334 flat, which can result in higher surface wind speeds. The difference in wind impacted by terrain

led to a weaker pollution removal effect in the eastern region, thus contributing to a higher proportion of pollution days under Pattern 6. Besides, differences in precipitation rates between eastern cities and other regions were not significant (the proportion of rainfall with a daily accumulated precipitation exceeding 10 mm in CD, CQ, YB and DZ under Pattern 6 were all less than 3%), which might not the main reason why eastern cities in the SCB experience higher pollution frequency.



Figure 4. (a) The pollution occurrence frequency at 18 air pollution stations in SCB, (b) (c) average $PM_{2.5}$ concentrations and $PM_{2.5}$ day to day variations at 4 representative SCB cities, under 6 synoptic patterns.

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342 The time series of daily mean PM_{2.5} and the day-to-day classification of 850 hPa synoptic 343 patterns are shown in Fig. S3, from December 2016 to January 2017. Six pollution episodes 344 occurred during this period (December 03-12 and 16-26, 2016; January 1-7, 16-19, and 20-28, 345 2017; and February 14–23). It is observed that pollution episodes consistently began with Pattern 2 346 and ended with Pattern 1, accompanied by a rapid decline in PM_{2.5}. This finding suggests that 347 Pattern 2 acted as a key synoptic forcing for the initation of pollution episodes. Additionally, 348 statistical results revealed that Pattern 2 accounted for a high proportion of PM_{2.5} increase during the 349 six pollution episodes, reaching 48.48 %, while Pattern 5 had the second highest proportion of 350 21.88 %, with Patterns 2 and 5 combined accounting for more than 70 % of the pollution episodes. 351 For example, during the two heavy pollution events in early and late January 2017, PM_{2.5} rapidly 352 accumulated with the interplay of Patterns 2 and 5. These two patterns represented a substantial 353 proportion of 31.96 % and 12.97 %, respectively, during winters from 2015 to 2020 at 850 hPa 354 level in the SCB (Fig. 3). Based on this analysis, Patterns 2, 4, and 5 were identified as synoptic 355 pollution patterns, whereas Patterns 1, 3, and 6 were as clean patterns. In summary, Patterns 2 and 356 5 played crucial roles in the initiation and accumulation of PM_{2.5} during pollution episodes. 357 The discussion above showed that pollution in the SCB tended to occur when southerly airflow

358 controlled the upper-layer of the basin (Pattern 2 and 5), while the dispersion of pollutants was

359 accompanied by northerly winds, which aligns with the findings of Lu et al. (2022). This study 360 indicated that southerly airflow in the upper-layer could bring warm air, leading to warming above 361 the basin and forming a "warm lid". The surrounding mountains and plateau with the "warm lid" 362 contributed to the formation of a relative enclosed space within the SCB, facilitating local 363 circulations and allowing for the thorough mixing and secondary reactions of local emission and 364 pollutants transported from outside. As a result, persistent and severe pollution often occurred 365 under the influence of southerly airflow. When the northerly airflow began to dominate the SCB, 366 the "warm lid" and local circulation were disrupted, leading to dispersion of pollutants through 367 advection and vertical transport. Northerly winds were often associated with cold air and sometimes accompanied by weak precipitation, resulting to wet deposition and the removal of 368 369 pollutants. Therefore, the arrival of northerly airflow often signified the ending of the pollution 370 episode. The evolution of 850 hPa synoptic forcing and vertical meteorological conditions (Fig. 2 371 and 6) aligns with the study of Lu et al (2022). Therefore, there are also similar pollution change 372 mechanisms.

373 Due to the convergence of air moving eastward across the Tibetan Plateau, the SCB experiences 374 frequent wet and cloudy weather, with cloud cover fraction exceeding 80 % (Yu et al., 2004; 375 Zhang and Lin, 1985). Clouds undoubtedly play an unneglectable role in the interactions of 376 aerosols, radiation, and the PBL under typical synoptic forcing in this region. This study evaluated 377 the average cloud liquid water content, downward solar radiation, and PBL under the influence of 378 the six classified synoptic patterns in CD, CQ, DZ, and YB, using data from ERA5 (Fig. 5). The 379 reanalysis data revealed significant higher cloud liquid water contents with Patterns 2 and 5, likely 380 triggered by robust southerly air prevailing at 850 hPa over the SCB (Fig. 3). This southerly air 381 brought warm and moist air, contributing to cloud formation. Dense clouds reduced solar radiation 382 through reflection and scattering, resulting in surface cooling and inhibiting PBL development. 383 The PBLH under Patterns 2 and 5 was approximately 900-1000 m, lower than that under the 384 influence of clean synoptic Pattern 6 at 1500 m or Pattern 1 and 3 at 1200-1300 m (Fig. 5). In 385 contrast, the clean synoptic Pattern 1 was characterized by a strong northerly flow at 850 hPa, 386 resulting in lower cloud liquid water content over the basin and increased solar radiation, promoting 387 PBL development. The lower PBLH with more stable stratification caused by the CRI in Patterns 2 388 and 5 could partially explain the rapid accumulation of $PM_{2.5}$ during these two pollution patterns. 389



Figure 5. The averaged (a) $PM_{2.5}$ concentrations, (b) cloud liquid water contents, (c) downwards solar radiation and (d) boundary layer height derived from 2015-2021 winter months ERA5 reanalysis data, at 4 representative SCB cities under 6 synoptic patterns, the dot sizes represent $PM_{2.5}$ concentrations.

391 3.2 Integrate impacts of synoptic patterns and the CRI inhibition on ARI

392 Based on the above analysis, Patterns 2 and 5 were identified as the key pollution synoptic 393 patterns accompanying dense clouds and, thus, strong CRI. However, the effects of pollution 394 patterns on ARI and their interaction with CRI in the SCB remain unclear and warrant further 395 investigation. A typical pollution episode from January 1–7, 2017, was selected to understand these 396 complex processes and simulated using WRF-Chem. The BASE simulations were verified with 397 observations to determine the accuracy and reliability of the simulation results. The simulation and 398 observation of PM_{2.5}, T2, TD2 and wind speed values with some statistical metrics in CD from 399 January 1–7 are shown in Fig. 6a-d. Similar information at CQ, YB and DZ can be found in the 400 Fig. S4. The MB of the simulated and observed PM_{2.5} concentrations were -15.59, -13.42, 2.10 401 and -13.11 μ g/m³, with NMB values of -4.12%, -4.22%, 6.01% and -0.68% at four cities, 402 respectively, which are within the acceptable standards (NMB $< \pm 15\%$). The R of PM_{2.5} were 403 78.91%, 57.23%, 61.15% and 62.86% for four representative cities, respectively. The statistical 404 metrics for PM_{2.5} are consistent with previous studies (Wang et al., 2020; Shu et al., 2021; Zhan et 405 al., 2023), indicating that our model results for PM_{2.5} are reasonable and acceptable. Regarding to 406 the surface meteorological factors, low MB and high R for both temperature and dew point 407 temperature suggested good simulation performance for these variables. However, the simulation 408 results for wind speed were poor, which was expected under conditions of low wind and complex 409 terrain. The high observed calm wind frequency, influenced by the starting speed of the 410 anemometer, led to an overestimation in the simulation (Shu et al., 2021; Zhan et al., 2023). 411 Additionally, it could be argued that unresolved topographic features introduce additional drag, 412 beyond that generated by vegetation, which was not considered in the WRF model (Jimenez and 413 Dudhia, 2012).

414 In addition, the temporal averaged and variations of vertical profiles for potential temperature, 415 relative humidity and wind speed in the model were compared with the sounding data in CD (Fig. 416 6e-m). Model evaluation of vertical structures in CQ, YB and DZ can be found in Fig. S5. The 417 SCB is characterized with cloudy and foggy conditions, which result in abundant water vapor and near 418 100% relative humidity above the nocturnal boundary layer. Models often underestimate the humidity 419 above the boundary layer during night in the SCB (Shu et al., 2021). Furthermore, due to complex 420 terrain and measurement bias of the anemometer for weak winds, the evaluation of simulation results 421 for wind speed often exhibit certain deviations (Jimenez and Dudhia, 2021; Shu et al., 2021; Zhan et al., 422 2023). For the verification of PBLH, sounding data are commonly regarded as reliable vertical 423 observation records, and PBLH calculated based on sounding data can be used as the true values to 424 compare with other data for long-term validation (Guo et al., 2016). However, for short-term studies, 425 due to limited availability of sounding data at only 00:00 and 12:00 UTC, the ERA5 data were also 426 incorporated for the model evaluation of PBLH in this study (Fig. 6 and Fig.S5). The simulation PBLH 427 showed a consistent trend with those calculated from ERA5 and sounding data. Overall, the simulation 428 results can capture the meteorological and PM2.5 variation trends. According to the simulation 429 evaluation standards for the SCB in previous studies (Wang et al., 2020; Zhan et al., 2023), the results 430 is acceptable and reasonable; thus, the simulation can be used for subsequent analysis and discussion.



Figure 6. Time series of hourly simulated and observed (a) $PM_{2.5}$ concentration, (b) temperature at 2 m, (c) dew temperature at 2m and (d) wind speed near surface, and comparison of simulated and observed mean vertical profile of (e) potential temperature, (f) relative humidiy and (g) wind speed, the red and grey shaded areas represent deviations to the mean values of observation and simulation, respectively. The simulated and observed time-height sections of (h)(k) potential temperature, (i)(l) relative humidity and (j)(m) wind speed are also given, while the red lines in (h)(k) are time series the boundary layer heights calculated with sounding data. The above figures display information in CD. Additionally, the model verification information regarding CQ, YB and DZ can be found in Supplement Figure 3 and 4.

432 During the pollution episode that occurred from January 1 to 7, 2017, the pollution synoptic 433 patterns controlled the SCB as follows: Pattern 2 from January 1 to 3, Pattern 5 from January 4 to 6, 434 and Pattern 1 on January 7. Consequently, PM_{2.5} pollution in the SCB occurred on January 1–6 and 435 rapidly dissipated on January 7 (Fig. 7). The mean geopotential height at 850 hPa derived from the 436 simulation of January 1-3 under Pattern 2 showed EHWL, with southerly flow prevailing over the 437 SCB (Fig. 7a). The resulting upper air warming suppressed PBL development (Fig. 7d). During 438 January 1–3 under Pattern 2, the average PBL heights were lower (Fig. 7d), acting as a lid above the 439 SCB and hindering the airflow within the basin due to the surrounding mountains. Low wind 440 speeds provided adverse diffusion conditions for pollutants emitted into the basin, resulting in 441 severe pollution in the western and southern SCB (Fig. 7g). From January 4 to 6, the low pressure 442 over the SCB evolved into a LT pattern, termed Pattern 5 in the previous analysis. Compared with 443 Pattern 2, the isobaric lines were denser under the influence of the LT, leading to stronger southerly 444 winds above the SCB (Fig. 7a-b). Lower average PBL heights were observed during January 4-6

445 under Pattern 5 compared with those of January 1-3 under Pattern 2 (Fig 7d-e), primarily due to 446 stronger upper air warming and more stable stratification. Pollutants that accumulated during 447 January 4-6 from the earlier pollution episode (January 1-3) further increased (Fig. 7g-i). On January 7, high pressure in the north dominated the SCB, with a prevailing northerly flow over the 448 449 basin (Fig. 7c). The PBL height quickly increased due to upper-layer cold advection (Fig. 7f), 450 resulting in a rapid decrease in PM2.5 (Fig. 7i). Overall, synoptic patterns play a key role in the 451 accumulation and diffusion of PM2.5 during pollution episodes by modulating PBL development 452 and stratification stability.



Figure 7. The simulated (a)-(c) 850 hPa geopotential height field (shading) with wind vector fields (blue vectors), (d)-(f) boundary layer height and (g)-(i) $PM_{2.5}$ concentrations (shading) and wind vector fields at 900hPa (blue vectors) for 1-3, 4-6 and 7 January. The size and color of scatters in (g)-(i) show corresponding observed $PM_{2.5}$ concentrations at 18 air quality monitoring stations. The SCB was outlined with an altitude contour of 750 m terrain height (black lines).

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Pollutant accumulation can regulate the PBL structure through the ARI, further exacerbating pollution (Wang et al., 2018; Miao et al., 2020). In the SCB, this positive feedback is weaker than in the other regions and may be inhibited by cloud radiation (Zhong et al., 2019). A series of simulation experiments were conducted to investigate the aerosol radiation feedback in the SCB

under the influence of two typical synoptic pollution patterns, as described in Section 2.4. BASE-458 459 EXP1 represents the perturbations caused by ARI, whereas EXP2-EXP3 demonstrates changes 460 through ARI without CRI inhibition. Aerosols led to surface cooling through absorbing and 461 scattering solar radiation, thereby inhibiting the development of the PBL, which in turn facilitated 462 pollutant accumulation (Fig. 8). Compared with Pattern 2, the aerosol concentrations in Pattern 5 463 were higher, resulting in greater reduction of downward solar radiation reduction due to ARI, 464 leading to more pronounced cooling near the ground and a lower PBLH. Overall, the ARI in Pattern 465 5 was more significant than that in Pattern 2, regardless of CRI inhibition (Fig. 8).



Figure 8. Spatial distribution of perturbations induced by (a)-(h) aerosol radiation interactions (ARI), and (i)-(p) ARI without the cloud radiation interaction (CRI) inhibition during 1-3 and 4-6 January representing Pattern 2 and Pattern 5 synoptic forcing, respectively. The SCB was outlined with an altitude contour of 750 m terrain height (black lines).

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Furthermore, parallel simulation experiments revealed that the CRI significantly attenuated the ARI in the SCB under both pollution synoptic patterns. When the CRI was not considered, more solar radiation penetrated the PBL. Dense aerosols accumulated near the surface, intercepting more downward shortwave radiation, and resulting in stronger cooling near the ground. This suppressed the development of the PBL, and contributed to a more remarkable ARI (Fig. 8). Regarding the horizontal spatial distribution, a strong ARI was primarily observed in CQ, as well as in the western and southern SCB, despite CQ experiencing lower pollutant concentrations compared to the other 474 two regions (Figs. S4 and 7). This weaker ARI phenomenon in the western SCB was also reported 475 by Zhong et al. (2019) and was attributed to CRI inhibition on ARI. Considering the statistical 476 results in Fig. 5, the average cloud liquid water content in CD and YB was significantly higher than 477 that in CQ under the influence of Patterns 2 and 5. Consequently, a more remarkable CRI 478 inhibition on the ARI would occur in the western and southern SCB compared to CQ, leading to a 479 relatively weaker ARI distribution in these regions. Without considering the CRI, the ARI in the 480 western and southern SCB would be much more pronounced than that in CQ. As for the 481 northwestern SCB (DZ), the ARI in DZ is lower than in the other three regions. When the CRI is not 482 considered, the ARI in DZ is higher than in CQ but lower than in CD and YB. This is because DZ has 483 lower aerosol concentrations compared to CD and YB (Fig. 7), but exhibits higher cloud cover than CQ 484 under Patterns 2 and 5 (Fig. 5).

485 Using the western SCB, which exhibited the highest pollution concentration, as an example, Fig. 486 9 illustrates the vertical diurnal variations in temperature and solar radiation caused by the ARI. 487 The results in Fig. 9–11 were derived from the simulation experiments in CD, as CD is one of the 488 most polluted cities with typical meteorological and geographical characteristics of the western SCB. 489 The ARI caused surface cooling in the morning and upper-air warming in the afternoon. As local 490 solar radiation increased from 8 am to 12 pm, the reduction in solar radiation caused by the ARI 491 also increased. Surface cooling reached its peak at approximately 10 am to 12 pm, and gradually 492 weakened in the afternoon. This diurnal variation might be attributed to the enhanced turbulence 493 during morning PBL evolution (Wang et al., 2018). Afternoon surface cooling was partly 494 compensated by the turbulent transport of warm air above the PBL. In addition, strong surface 495 cooling between 5 pm and 8 pm in the SCB, was possibly influenced by remarkable valley wind 496 circulations forced by the Qinghai–Tibet Plateau adjacent to the western SCB (Lu et al., 2022). 497 The evening cooling of the plateau induced strong mountain winds, promoting surface cooling, 498 while the upper-layer warming mainly occured around 1-1.5 km in the afternoon. In general, the 499 ARI reduces solar radiation, causing surface cooling and upper air warming, thereby regulating the 500 vertical atmospheric thermal structure, suppressing convection, and consequently decreasing PBL 501 heights (Fig. 10).



Figure 9. Diurnal variations of vertical temperature perturbations and downward solar radiation under influences of Pattern 2 and Pattern 5 induced by (a) (c) ARI and (b) (d) ARI without CRI inhibition.



Figure 10. Diurnal variations of (a) (c) boundary layer height(lines) and downward solar radiation(shading), and (b)(d)the perturbations of boundary layer height induced by ARI and ARI without CRI inhibition, under Pattern 2 and Pattern 5 synoptic forcing respectively.

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Figure 11. Meridional vertical distribution of averaged PM_{2.5} between 104°E and 105°E under (a) Pattern 2 and (b) Pattern 5, and (c) average profiles of PM_{2.5} within 28°N and 31°N.

505 Synoptic patterns play a role in the interaction between the ARI and PBL (Wang et al., 2018; Miao et al., 2020). Zonal average of PM_{2.5} concentration between 104°E and 105°E was conducted, 506 and the meridional vertical distribution of PM2.5 between 27°N and 35°N was illustrated in Fig. 507 508 11a-b. Fig. 11(c) provides an average of PM_{2.5} concentration within 28°N and 31°N, showing the 509 vertical distribution profiles under Pattern 2 and 5. Due to the inhibition of "warm lid" above the 510 SCB, the vertical exchange was not prominent under both Pattern 2 and 5, and PM_{2.5} was more 511 concentrated at the middle and lower levels. The PM_{2.5} concentration under Pattern 5 was higher 512 than Pattern 2 throughout the atmospheric column, indicating stronger aerosol radiative forcing 513 and a more significant impact on the boundary layer structure under Pattern 5. During January 4-6, 514 the surface cooling reached 1 K, with cooling layers higher than those observed on January 1-3. 515 The differences in thermal structure modulations contributed to a lower diurnal PBLH in Pattern 5 516 than in Pattern 2 (Figs. 10a and c), indicating that Pattern 5 was more conducive to ARI. Based on 517 the simulation experiments, this study further discussed the impact of synoptic forcing on the CRI 518 inhibition of ARI. When the CRI was not considered, the solar radiation reduction at noon on 519 January 4-6 by the ARI was nearly twice as high as when the CRI was considered. 520 Correspondingly, surface cooling at noon was remarkably enhanced. In the evening, surface 521 cooling occurred earlier and was stronger without the CRI (Fig. 9). The regulation of CRI on ARI 522 was further reflected in changes in PBLH. Without the CRI, the diurnal PBLH increased 523 significantly, with the PBLH decreased more with ARI without CRI inhibition. The PBLHs were 524 decreased by the ARI during January 13-17 afternoon, reaching 2-3 times the decrease observed 525 with CRI inhibition (Fig. 10). More significant CRI inhibition of ARI was revealed under Pattern 526 5 compared with that under Pattern 2, owing to the stronger ARI itself with higher aerosol

- 527 concentrations in Pattern 5 and the more apparent CRI inhibition with denser cloud liquid water
- 528 contents under the LT pattern (Fig. 5). Therefore, the intensity of CRI inhibition of ARI in the
- 529 SCB was altered by synoptic forcing, with stronger effects under the influence of LT.



Figure 12. The aerosol radiative effect can be significantly inhibitted by cloud under influence of pollution synoptic patterns with dense cloud.

531 4 Conclusion

532 This study utilized synoptic classification and numerical simulation to gain insights in to the 533 combined effects of synoptic patterns and CRI inhibition on ARI and PBL structures in the wet and 534 cloudy SCB. Based on the long-term PM2.5 observations and sounding data in the SCB, it was 535 found that large-scale synoptic circulations at 850 hPa played crucial roles in the variations of 536 PM_{2.5} pollution. Synoptic classification was performed with the T-PCA method, which reveaed 537 that Pattern 2 and 5 characterized with low pressure system and southerly airflow on 850 hPa were 538 key synoptic patterns for onset and accumulation of PM2.5, while Pattern 1 controlled by the 539 northerly airflow represented a clean pattern associated with significant decrease in PM_{2.5}. 540 Moreover, it was indicated that Pattern 2 and 5 exhibited denser cloud liquid water content and 541 thus stronger compared to other patterns. Among these patterns, Pattern 5 exhibited the highest 542 cloud liquid water content and CRI. This could be attributed to the robust southerly airflow 543 induced by the dense isobaric lines, which brought warm and humid air masses into the region. 544 To illustrate the interactions among cloud, aerosol and PBL under pollution synoptic patterns, a pollution episode occurred from January 1 to 7, 2017, was simulated with using WRF-Chem. The 545

546 simulation results showed that ARI remarkably reduced solar radiation was during the two 547 pollution patterns. This reduction led to surface cooling in the morning and upper-air warming in 548 the afternoon. Additionally, the enhanced evening surface cooling was impacted by the mountain-549 valley wind circulations forced by the plateau-basin topography of the SCB. This modulation in 550 the vertical thermal structure by the ARI would then suppress the development of the PBL, 551 favoring pollution outbreaks (Fig. 12). Furthermore, parallel simulation experiments indicated that 552 CRI impacted stratification stability and modulated the vertical thermal structure by inhibiting ARI 553 (Fig. 12). Regarding the spatial distribution, a stronger ARI appeared in Chongqing, despite lower 554 PM_{2.5} concentrations compared to the western and southern SCB. This was due to the lower cloud 555 liquid water content and weaker CRI inhibition of ARI in Chongqing. When CRI inhibition was 556 not considered, the ARI in the western and southern SCB was significantly stronger than that in 557 Chongqing. In addition, under Pattern 5, the reduction in solar radiation and PBLH during the 558 daytime due to ARI could be more than doubled when the CRI influence was neglected. This was 559 primarily due to higher aerosol concentrations and cloud liquid water contents associated with a 560 low trough in Pattern 5. This study provided insights into the interaction among aerosols, clouds, 561 and PBL under different synoptic patterns, considering the complex terrain and foggy/cloudy 562 climate of the SCB. The findings highlighted the significant role of CRI inhibition on ARI during 563 wet and cloudy conditions, shedding light in the multi-scale atmospheric physical processes in the 564 SCB.

565 **Author contributions.** HL and MX had the original idea for the study, designed the experiments, 566 conducted the numerical simulation and prepared the initial draft manuscript. BL, YZ and KZ 567 collected the data. TW and BZ helped perform the analysis with constructive discussions, reviewed and 568 edited the manuscript. HL, MX, TW and BL acquired financial support for the project leading to this 569 publication. SL and ML reviewed the manuscript.

- 570 Competing Interest: The authors declare no conflict of interest.
- 571 Data Available Statement

572 The ERA5 pressure layer and single layer data can be respectively downloaded from
573 <u>https://cds.climate.copernicus.eu/cdsapp#!/dataset/reanalysis-era5-pressure-levels</u> and
574 <u>https://cds.climate.copernicus.eu/cdsapp#!/dataset/reanalysis-era5-single-levels?tab=form</u>.
575 The NCEP FNL data are available at <u>https://rda.ucar.edu/datasets/ds083.2/</u>. The MEIC data can
576 be accessed in Zheng et al (2018) at <u>https://doi.org/10.5194/acp-18-14095-2018</u>. Air quality and
577 meteorological monitoring data can be acquired from <u>https://doi.org/10.7910/DVN/USX59F</u>.
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