

1 **Fractional solubility of iron in mineral dust aerosols over coastal**
2 **Namibia: a link with marine biogenic emissions?**

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29 **Abstract**

30 Mineral dust is the largest contributor to elemental iron in the atmosphere, and, by deposition, to the
31 oceans, where elemental iron is the main growth limiting nutrient. Southern Africa is an important
32 source at the regional scale, and for the Southern Ocean, however limited knowledge is currently
33 available about the fractional solubility of iron from those sources, as well as on the atmospheric
34 processes conditioning its dissolution during deposition.

35 This paper presents the first investigation of the solubility of iron in mineral dust aerosols from 176
36 filter samples collected at the Henties Bay Aerosol Observatory (HBAO), in Namibia, from April to
37 December 2017. During the study period, 10 intense dust events occurred. Elemental iron reached
38 peak concentrations as high as $1.5 \mu\text{g m}^{-3}$, significantly higher than background levels. These events
39 are attributed to wind erosion of natural soils from the surrounding gravel plains of the Namib desert.
40 The composition of the sampled dust is found to be overall similar to that of aerosols from northern
41 Africa, but characterised by persistent and high concentrations of fluorine, which are attributed to
42 fugitive dust from mining activities and soil labouring for construction.

43 The fractional solubility of Fe (%SFe) for both the identified dust episodes and background conditions
44 ranged between 1.3 to 20 % and averaged at 7.9% ($\pm 4.1\%$) and 6.8 ($\pm 3.3\%$), respectively, in the
45 range of values previously observed in the remote Southern Ocean. Even in background conditions,
46 the iron fractional solubility was correlated to that of aluminium and silicon ~~solubility~~. The solubility
47 was lower between June and August, and increased from September onwards, during the austral
48 spring months. The relation with measured concentrations of particulate MSA (methane_sulfonic acid),
49 solar irradiance and wind speed suggests a possible two-way interaction whereby marine biogenic
50 emissions from the coastal Benguela upwelling to the atmosphere would increase the solubility of
51 iron-bearing dust, according to the photo-reduction processes proposed by Johansen and Key (2006).
52 The subsequent deposition of soluble iron could act to further enhance marine biogenic emissions.
53 This first investigation points to the west coast of southern Africa as a complex and dynamic environ-
54 ment with multiple processes and active exchanges between the atmosphere and the Atlantic Ocean,
55 requiring further research.

56

57 **Keywords:** aerosols, mineral dust, water-soluble Fe, atmospheric processing, marine biogenic emis-
58 sions

59 1. Introduction

60 Through the processes of atmospheric transport and deposition, mineral dust is known to provide
61 nutrients and metals to the terrestrial and marine ecosystems (Hooper et al., 2019; Ventura et al.,
62 2021). Amongst those, mineral dust provides iron (Jickells et al., 2005), which plays a major role for
63 the primary productivity of the nutrient-limited oceans, modulating the marine carbon cycle (Hooper
64 et al., 2019) as well as that of key continental ecosystems such as the Amazon rainforest (Reichholf,
65 1986).

66 To date, much attention has been paid to the soluble Fe in mineral dust emitted from arid and semi-
67 arid areas in the northern Hemisphere, in particular the Saharan and Chinese deserts (e.g. Baker et
68 al., 2006; Paris et al., 2010; Takahashi et al., 2011; Rodriguez et al., 2021), where emissions are the
69 most intense (Tegen and Schepanski, 2009).

70 Nonetheless, the southern Hemisphere accounts for approximately 10% of the global atmospheric
71 dust loading (Kok et al., 2017). Large sources are found in southern Africa, mostly in Namibia (Kala-
72 hari and Namib deserts, Etosha Pan), numerous ephemeral riverbeds along the Namibian coastline)
73 and Botswana (Makgadikgadi Pan) (Prospero et al., 2002; Bryant et al., 2007; Mahowald et al., 2003;
74 Ginoux et al., 2012; Vickery and Eckardt, 2013; Von Holdt et al., 2017).

75 Previous research has shown that the long-range transport of dust emitted from southern African
76 sources can reach the south-eastern Atlantic and the Indian Oceans (Swap et al., 1996; Jickells et
77 al., 2005; Bhattachan et al., 2012; 2015; Ito and Kok, 2017). In particular, Gili et al. (2022) demon-
78 strated recently that mineral dust from Namibia can also be transported across the Southern Oceans
79 to eastern Antarctica. Furthermore, the research by Dansie et al. (2022) has suggested that mineral
80 dust from Namibia could dominate the atmospheric deposition to the coastal Benguela Upwelling
81 System (BUS), where biomass burning aerosols, a significant source of soluble Fe to the Southern
82 and Indian Oceans (Hamilton et al., 2021; Ito et al., 2021; Liu, et al., 2022), are limited by atmospheric
83 stratification (Formenti et al., 2019; Redemann et al., 2021). The inputs of Namibian (and Angola)
84 dust in the upwelled waters could also modulate the migration of skipjack tuna between Gulf of Guinea
85 and equatorial Atlantic, by contributing to support phytoplankton growth and hence upper trophic lev-
86 els in this area (Rodriguez et al., 2023).

87 There is, however, very little data available on the concentrations and composition of soluble Fe in
88 dust aerosols from southern Africa, both near the sources and over the oceans. Previous research in
89 Namibia focussed on soils and sediments (Dansie et al., 2017a; 2017b; Kanguuehi, 2021). The At-
90 lantic Meridional Transect (AMT) cruise programme conducted recurrent observations between Oc-
91 tober and March in the South Atlantic Ocean (Baker et al., 2013), while Heimbürger et al. (2013) and
92 Gao et al. (2013) report on sparse measurements of deposited aerosols and in rainwaters over the
93 Southern Indian Ocean.

94 Within this context, this paper investigates the fractional solubility of Fe in samples of atmospheric
95 aerosol particles smaller than 10 μm in diameter collected in 2017 at the Henties Bay Aerosols Ob-
96 servatory (HBAO; 22.09°S, 14.26°E) on the Namibian coast. In section 2 we outline the experimental
97 and analytical methodology for elemental and water-soluble analysis of ions and metals, including
98 iron, ~~obtained by Inductively Coupled Plasma (ICP) analysis~~. We also provide the definition of frac-
99 tional solubility and method for estimating the total dust mass. We introduce the supporting tools used
100 to evaluate the source regions of the collected mineral dust, their pathways during transport, and the
101 presence of fog, a recurrent feature on coastal Namibia favouring multi-phase ageing processes.
102 Section 3 provides the results of the analysis. We present the iron soluble concentrations and solu-
103 bility, and explore their links to the load, emission area and transport of mineral dust, as well as at-
104 mospheric processing. Section 4 discusses the observations, suggesting that the fractional solubility
105 of iron in the Namibian dust is higher when the ~~particulate~~-MSA, a tracer of marine biogenic emissions,
106 is also detected in highest concentrations. This points to the photo-oxidation of DMS as a process for
107 increasing the ~~dust~~-Fe solubility, and suggests a possible positive feedback loop of the iron fertilisation
108 by dust to the ocean. Section 5 summarizes the findings and suggests directions for future research.

109

110 **2. Methodology**

111 **2.1. Study area**

112 The Henties Bay Aerosol Observatory (HBAO, 22.09°S, 14.26°E; <http://www.hbao.cnrs.fr/>, last ac-
113 cess: 10 October 2022) is located at the Sam Nujoma Marine and Coastal Resources Research Cen-
114 tre (SANUMARC) of the University of Namibia in Henties Bay, Namibia (**Fig 1**).



115

116 **Fig 1.** Location of Henties Bay Aerosol Observatory (HBAO, red star) and main dust source regions (©
 117 Google Maps). The position of Walvis Bay (blue dot), the major harbour in the area, and the Wlotzkabaken
 118 meteorological station (blue star) are also indicated.

119

120 Three kilometers to the south of the University campus hosting HBAO is the small town of Henties
 121 Bay, with no industrial activity and very little traffic, and approximately 170 km north from Walvis Bay,
 122 the major harbour in Namibia. Directly east of HBAO are the Namibian gravel plains, which are one
 123 of the dominant features of the Namib desert together with the sand dunes. Approximately 100 m to
 124 the north is the Omaruru riverbed, one of the coastal sources of mineral dust identified by Vickery and
 125 Eckardt (2013).

126 Our previous results show that, at the surface level, the atmosphere at HBAO is a receptor of different
 127 air masses dominated by marine aerosols, but also the seasonal occurrence of light-absorbing aerosols
 128 from biomass burning or pollution in northern wind regimes, and mineral dust detected episodically
 129 from various wind directions (Formenti et al., 2018; Klopper et al., 2020, hereafter KL20).

130 **2.2. Sample collection and analysis**

131 Aerosol particles smaller than 10 μm in aerodynamic diameter (PM_{10}) were collected by an automated
 132 sampler (model Partisol Plus 2025i, Thermo Fisher Scientific, Waltham, MA USA) on 47 mm What-
 133 man Nuclepore polycarbonate filters (1- μm pore size). The air was drawn through a certified sampling
 134 inlet (Rupprecht and Patashnick, Albany, New York, USA) located at approximately 30 m above
 135 ground and operated at a flow rate of 1 $\text{m}^3 \text{h}^{-1}$. Samples were collected for 9 hours during the daytime
 136 (from 9:00 to 18:00 UTC time) and night-time (21:00 to 06:00 UTC time) for 12 non-consecutive weeks

137 from April to December 2017 (7-14 April, 26 April-3 May, 19-26 May, 07-14 July, 2-9 August, 15-22
138 August, 18-25 September, 02-09 October, 31 October-7 November, 13-20 November, 28 November-
139 04 December, 12-19 December). In total, 176 samples (+ including 13 blanks, one per week of sam-
140 pling) were collected.

141 The elemental analysis of 24 elements from Na to Pb and including some major tracers of mineral
142 dust (Fe, Al and Si) was performed at the LISA laboratory by Wavelength-dispersive X-ray fluores-
143 cence (WD-XRF) using a PW-2404 spectrometer (Panalytical, Almelo, Netherlands), as detailed by
144 KL20. The total mass concentration per element x will be referred to as T_{X^*} .

145 The measured elemental concentrations are used to calculate the estimated dust mass (*EDM*) ac-
146 cording to Lide (1992) as

$$147 \quad EDM = 1.12 \times \{1.658 \times [\text{nss-Mg}] + 1.889 \times [\text{Al}] + 2.139 \times [\text{Si}] + 1.399 \times [\text{nss-Ca}] + 1.668 \times [\text{Ti}] + 1.582 \\ 148 \quad \times [\text{Mn}] + (0.5 \times 1.286 + 0.5 \times 1.429 + 0.47 \times 1.204) \times [\text{Fe}]\} \quad (1)$$

149

150 where, as explained by KL20, nss-Mg and nss-Ca represent the non-sea salt fractions of Mg and Ca,
151 respectively.

152 The analysis of the water-soluble fraction was also performed at LISA. Individual filters were placed
153 in 20 mL of ultrapure water (MilliQ® 18.2 MΩ.cm) for 30 minutes. The solution was filtered (Nuclepore
154 polycarbonate filters with 0.2µm pore size) then divided into two sub-samples. One half was analysed
155 by Ion Cehromatography (IC) using a Metrohm IC 850 device equipped with a column MetrosepA
156 supp 7 (250/4.0 mm) for anions and with a Metrosep C4 (250/4.0 mm) for cations. The IC analysis
157 provided the concentrations of the following water-soluble ions: F⁻, formate, acetate, MSA⁻ (me-
158 thanesulfonic acid), Cl⁻, NO₃⁻, SO₄²⁻, formate, acetate, oxalate, MSA⁻ (methanesulfonic acid) oxalate,
159 Na⁺, NH₄⁺, K⁺, Ca²⁺ and Mg²⁺. A calibration with certified standard multi-ions solutions of concentra-
160 tions ranging from 5 to 5000 ppb was performed and the uncertainty of the analysis was estimated to
161 be 5% (KL20).

162 The second half of the solution was acidified to 1% with ultrapure nitric acid (HNO₃) and analysed by
163 Inductively Coupled Plasma-Atomic Emission Spectroscopy (ICP-AES) using a Spectro ARCOS
164 Ametek® ICP-AES and by High-resolution Inductively Coupled Plasma-Mass Spectrometry (HR-ICP-
165 MS) using a Neptune Plus™ instrument by Thermo Scientific™. The calibration curve was performed
166 using standard multi-element solutions ranging from 2 to 1000 ppb for ICP-AES and 1 to 1000 ppt for
167 HR-ICP-MS (Desboeufs et al., 2022). These analyses provided the dissolved mass concentrations
168 (D_xD_X) of 25 water-soluble metals and metalloids, including Fe, Al, and Si. All sample concentrations
169 were corrected using the filter blanks for each sampling period.

170 Based on those analyses, the fractional solubility ($\%S_xSX$) representing the percentage solubility
171 value was calculated as

$$\%S_xSX = 100 \times \frac{D_xDX}{T_xTX} \quad (2)$$

172
173
174
175 with D_xDX and T_xTX , the dissolved and total elemental concentration respectively.

176 Here, a leaching protocol using ultrapure water (UPW) was used to simulate wet deposition of parti-
177 cles, since the wet deposition dominates the total iron supply in the Southern Atlantic Ocean (Chance
178 et al., 2015). Moreover, the UPW leach enables the chemical reaction between iron with organic or
179 inorganic ligands, naturally dissolved from the particulate aerosols into rain droplets. However, it is
180 known that the extraction protocol modulates dissolution process and hence the values of iron frac-
181 tional solubility, in particular the estimates using UPW are higher in comparison to these one using
182 seawater, but lower than the acidic, buffered or reduction agent leach (Perron et al., 2020).

183 2.3.

184 **Ancillary data**

185 Maps of the emission fluxes of mineral dust were calculated using the dust emission model described
186 by Feuerstein and Schepanski (2019), driven with hourly 10m wind fields at a $0.1^\circ \times 0.1^\circ$ grid from
187 the European Centre for Medium-range Weather Forecasts (ECMWF). The dust emission parame-
188 terisation follows Marticorena and Bergametti (1995). Additional information on the soil type was taken
189 from the ISRC soil data set (FAO/IIASA/ISRIC/ISSCAS/JRC, 2012) and information on the aerody-
190 namic roughness length was obtained from POLDER/ADEOS surface products following the works
191 of Marticorena et al. (2004) and Laurent et al. (2005). The MODIS monthly vegetation product
192 (MYD13A3 v6) was used to describe the vegetation cover, while the vegetation type was defined
193 using the BIOME4 database (Kaplan et al., 2003). We additionally differentiated between different
194 dust source types (alluvial fines, dunes and sand sheets) which allowed us to reflect the source di-
195 versity over Namibia and thus the spatial diversity in the soil's susceptibility to wind erosion. This layer
196 was compiled following Feuerstein and Schepanski (2019) using MODIS surface reflectance
197 (MOD09A1 v6). A MODIS retrieved map on surface water cover was used to eliminate flooded areas
198 as active dust sources.

199 Back-trajectories of the air masses during the dust event were calculated from Meso-NH model (ver-
200 sion 5.3). The model set-up is similar to the one used for the AErosols, RadiatiOn and CLouds in
201 southern Africa (AEROCLO-sA) field campaign (Formenti et al. 2019) and related case studies (Fla-
202 mant et al. 2022; Chaboureau et al. 2022). In short, the model was run on a 5 km grid covering the
203 southern tip of Africa and 67 stretched levels spaced by 60 m close to the surface and 600 m at high

204 altitude. Meso-NH was run for 24 h for each dust event using initial and boundary conditions provided
205 by the ECMWF operational analysis. Emission, transport and deposition of dust is described by the
206 scheme of Grini et al. (2006). Back trajectories were computed online using three passive tracers
207 initialized with the 3D-field of their initial conditions. Further details on the dust prognostic scheme,
208 the backward trajectories and the physical parameterizations are given in Chaboureau et al. (2022).

209 The presence of fog and low clouds (FLC) along the Namibian coastline during dust events was an-
210 alysed using an existing satellite-based fog and low-cloud data set (Andersen et al., 2019). The FLC
211 detection algorithm used to create this data set was developed and validated specifically for this re-
212 gion. The algorithm is based on infrared observations from the Spinning Enhanced Visible and Infra-
213 red Imager (SEVIRI) aboard the geostationary Meteosat Second Generation (MSG) satellites, making
214 use of both spectral and textural information. The FLC product is available at the native spatial and
215 temporal resolutions of the SEVIRI sensor (3 km nadir, every 15 minutes), as described in Andersen
216 and Cermak (2018). The FLC product does not specifically distinguish between fog and low clouds
217 but captures the coastal boundary-layer cloud regime typical for the region and at HBAO that could
218 interact with mineral dust. It has been shown to be consistent with synoptic-scale atmospheric dy-
219 namics (Andersen et al. 2020). The FLC data are used to calculate maps of average fog and low
220 cloud coverage for the time periods of all dust events given in Table 1.

221 Observations of the local meteorology, including measurements of air temperature, relative humidity
222 and fog, at the nearby Wlotzkasbaken meteorological station (22.31°S, 14.45°E, 73 m asl, see **Fig.**
223 **1**) part of the Southern African Science Service Centre for Climate Change and Adaptive Land Man-
224 agement (SASSCAL) ObservationNet (<https://www.sasscal.org/>; last accessed 14/04/2023), are
225 used.

226

227 **3. Results**

228 **3.1. Description of the dust episodes**

229 The dataset discussed in this paper is based on 176 aerosol samples collected at HBAO, 42 of which
230 were associated with 10 dust episodes. As detailed by KL20, events of mineral dust were identified
231 as peaks in the time series of the mass concentrations of Al and non-sea-salt Ca^{2+} (nss- Ca^{2+}). The
232 dust episodes investigated in this study are a subset of those presented by KL20, we therefore use
233 their naming convention to facilitate the connections between the two papers (**Table 1**). In the follow-
234 ing, we refer to samples collected during the dust episodes as “dust”. Samples collected outside the
235 dust events will be indicated as “background”.

236

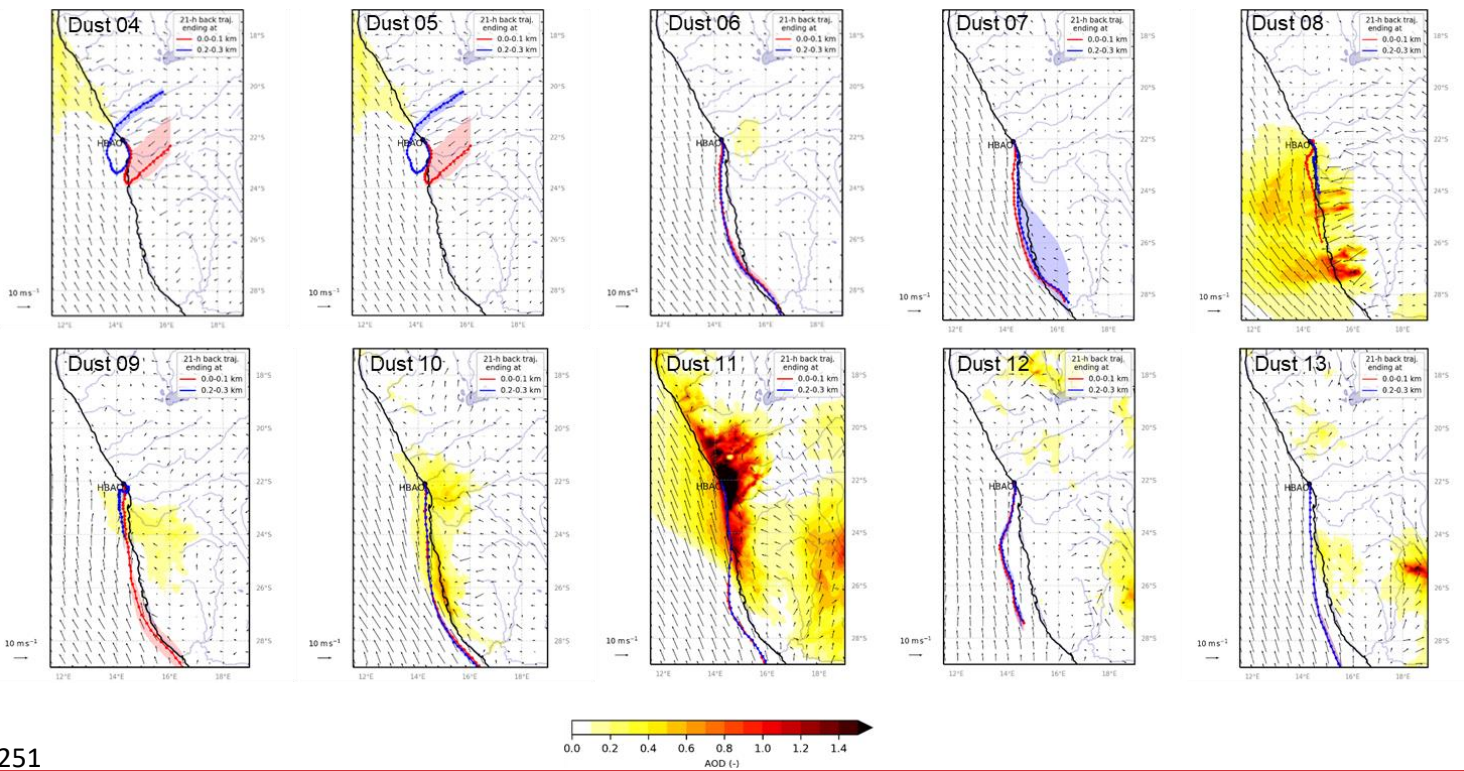
237 **Table 1.** *Dates of dust events identified at HBAO from May to December 2017, following KL20. The number of*
238 *samples collected during each episode is indicated in the column called “N”. The average air temperature,*
239 *relative humidity, wind speed and direction recorded at the nearby meteorological station in Wlotzkasbaken are*

240 reported. The maxima wind speed and corresponding direction are indicated in brackets in the corresponding
 241 columns. The average EDM is reported with in brackets the maximum of EDM during the event.

Episode identifier	Start and end date (UTC)	N	Air temperature (°C)	RH (%)	Wind speed (m s ⁻¹)	Wind direction (degN)	EDM (µg m ⁻³)
Dust 04	19/05 09h – 20/05 18h	3	17.7	73.7	2.7 (6.2)	186 (185)	13 (14)
Dust 05	24/05 21h – 26/05 09h	3	18.1	63.3	2.3 (6.3)	183 (188)	21 (42)
Dust 06	11/07 09h – 13/07 09h	4	13.2	82.9	1.2 (5.4)	235 (193)	27 (45)
Dust 07	04/08 21h – 06/08 09h	4	12.5	87.0	1.2 (5.4)	233 (201)	10 (16)
Dust 08	17/08 21h – 19/08 09 h	4	11.9	80.6	1.3 (4.6)	324(129)	18 (21)
Dust 09	23/09 21h – 24/09 18h	2	15.6	84.3	3.1 (6.2)	309 (330)	11 (17)
Dust 10	05/10 21h – 08/10 09h	8	14.0	74.6	2.1 (5.9)	249 (228)	14 (23)
Dust 11	15/11 09h – 18/11 09h	6	16.7	66.1	3.2 (11.7)	231 (232)	31 (56)
Dust 12	30/11 09h – 01/12 18h	3	16.7	78.1	1.9 (5.7)	244 (195)	2 (3)
Dust 13	15/12 09h – 19/12 09h	7	16.9	76.9	2.9 (6.5)	252 (238)	10 (19)

242

243 The dust episodes were long-lasting (generally a few days). The dynamic of the emissive areas, air
 244 mass transport and fog coverage during the episodes (**Fig. 2 and S4S1**) is driven by the synoptic
 245 circulation, which, in southern Africa, is primarily affected by the high-pressure belt under the de-
 246 scending limb of the Hadley cell (Tyson and Preston-Whyte, 2014). The maps of dust emission fluxes
 247 and the air mass back-trajectories reflect this seasonality. During the first part of the year (episodes
 248 Dust 04 to 05), dust emissions originated from the gravel plains and the Etosha pan north of HBAO.
 249 During this time of the year the transport to HBAO below 300 m asl was north- to south-easterly
 250 originating inland from the coast.



251

252 **Fig 2.** Maps of dust optical depth (shading) and 10-m wind (vector) overlaid by pathway of 21-hour air mass
 253 back trajectories ending in the first 100 m (red line) and between 200 and 300 m (blue line) above HBAO for
 254 dust episodes, as calculated by the Meso-NH model (version 5.3). Dots are plotted every hour and shadings
 255 around these lines are the interquartile ranges for latitudes.

256

257 From July onwards, the active source areas were identified in the southern gravel plains, Namib sand
 258 dunes and Kalahari Desert (this former only for Dust 11 to 13). Air mass transport was southerly and
 259 travelled over the sea and along the coastline. It is worth noticing that all the air masses experienced
 260 maritime air during their last hours of transport, including the episodes Dust 04 and 05 associated
 261 with berg wind conditions, due to the coastal low that develops to the west of HBAO.

262 The formation of fog events at Henties Bay is also highly seasonal. The frequency of occurrence of
 263 fog events is highest during austral winter at the coast, whereas lifted stratus clouds dominate during
 264 austral summer, when overall FLC occurrence peaks. The occurrence of fog over Namibia correspond
 265 to the advection of low-level clouds which is modulated both by local meteorology along the coastline
 266 of Namibia (trade winds) and synoptic-scale radiative processes (Spirig et al., 2019; Andersen et al.,
 267 2019; 2020). Henceforth, as shown in Fig. S1, the presence of fog and low clouds correlates with
 268 wind directions and aerosol source regions. Overall, three episodes (Dust 04, Dust 05 and Dust 11 in
 269 April, May and November, respectively) occurred in fog-free or low-fog conditions. The remaining
 270 episodes were characterised by extensive fog and low cloud coverage throughout the study area. The
 271 meteorological observations at the nearby Wlotzkasbaken station (**Fig. S2**) confirm these findings,
 272 and show in particular that the relative humidity always exceeded 60 %, and 80 % when fog or low
 273 clouds were present (Table 1). As a consequence, the aerosol can be considered deliquescent even

274 in the fog-free conditions. The seasonality is also observed in the average downwelling solar irradiance, with the lowest values during July and September, associated with austral winter. Finally, it is interesting to note that the fog-free conditions, associated with the predominance of continental air masses, corresponded to the highest estimated dust mass (EDM), possibly because of the reduced wet removal during transport and the increase of emission fluxes with the decrease of soil moisture (Kok et al., 2014), but possibly also because of the high wind speed prevailing during these conditions, which in principle, enhancing both dust emissions and transport (Table 1).

281 3.2. Iron solubility

282 The total and dissolved concentrations, and fractional solubility of Fe, Al and Si, during the dust episodes are reported in **Table 2**, where they are compared to background conditions. For iron, the average values over the entire sampling period are also shown.

285 **Table 2.** Average and standard deviations of water-soluble ($D \times DX$), total elemental ($T \times X$) mass concentrations and fractional solubility ($\%S \times X$) for Fe, Al and Si at HBAO measured for the total period and during the dust and background events from April to December 2017. Concentrations values are expressed in ng m^{-3} , while fractional solubility is expressed in percent. The numbers of considered samples is presented between the parentheses.

	Fe			Al		Si	
	All period	Dust	Background	Dust	Background	Dust	Background
$D \times DX$	28 ± 51 (N=175)	80 ± 84 (N=42)	11 ± 10 (N=131)	322 ± 296 (N=42)	56 ± 46 (N=131)	529 ± 616 (N=42)	78 ± 83 (N=124)
$T \times TX$	364 ± 482 (N=176)	955 ± 633 (N=42)	177 ± 155 (N=133)	1204 ± 870 (N=42)	284 ± 222 (N=94)	4158 ± 3037 (N=42)	776 ± 674 (N=133)
$\%S \times X$	7.1 ± 3.6 (N=175)	7.9 ± 4.1 (N=42)	6.8 ± 3.3 (N=130)	27 ± 10 (N=42)	26 ± 11 (N=90)	12 ± 7 (N=42)	11 ± 8 (N=116)

290

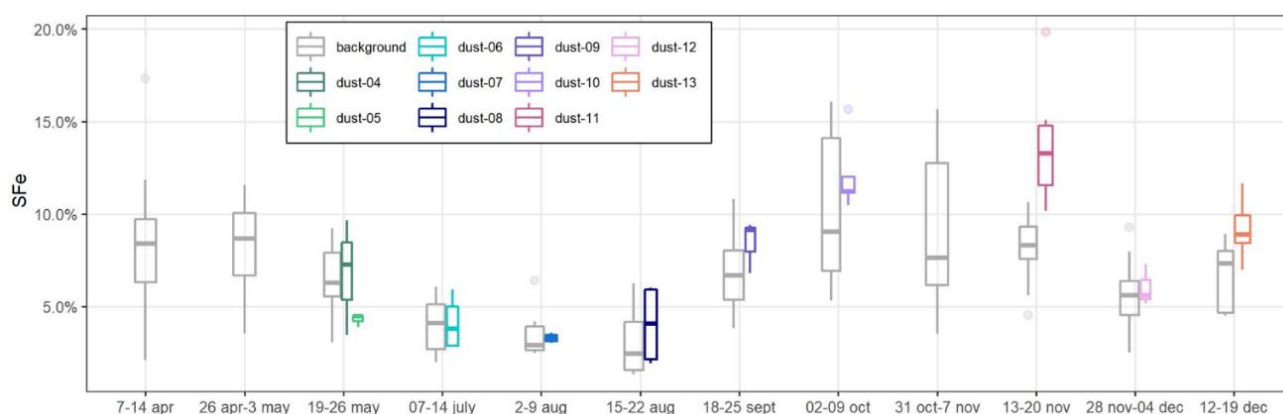
291 The total Fe concentrations varied significantly from one episode to the other, and so did EDM, which was larger than $10 \mu\text{g m}^{-3}$ for all of them (except Dust 12) and as high as $56 \mu\text{g m}^{-3}$ during Dust 11 event (Table 1). By contrast, the total Fe-to-EDM ratio was virtually constant, with an average of 5.8 % (± 0.6 %) for the dust events and 5.6 % (± 1.1 %) for the entire dataset. These values are quite superior to usual Fe content recommended in upper continental crust models (3.5% for Taylor and McLennan or 5.04 +/- 0.53 % Rudnick and gao, 2004) and estimated in Saharan dust (4.45% for Guieu et al., 2002; 4.3 to 6.1% for Lafon et al., 2006 or 4.5% for Formenti et al., 2008). Keeping in mind that Fe abundance is estimated, this suggests that the Namibian aerosol dust could be enriched in iron in comparison to upper crust and dust provided by Saharan sources.

300 The total dissolved concentrations of Fe during the sampling period ranged from 1.5 to 427 ng m^{-3} , with a median and average of 10.5 and 28 ng m^{-3} . During the dust episodes, the average mass concentration of dissolved Fe was $80 \pm 84 \text{ ng m}^{-3}$, almost an order of magnitude higher than for background conditions ($11 \pm 10 \text{ ng m}^{-3}$). The dissolved concentrations in dust periods are higher than

304 those observed in the South Atlantic Ocean for air masses associated with transport from continental
305 southern Africa (Baker et al., 2013; Chance et al., 2015; Baker and Jickells, 2017), which are of the
306 order as those observed at HBAO for background periods.

307 The calculated fractional solubility of Fe ranged from 1.3 to 19.8 %, with a median and average of 6.7
308 and 7.1 %. The average %SFe during dust events ($7.9 \pm 4.1\%$) was higher, but quite similar than in
309 background conditions ($6.8 \pm 3.3\%$). It is interesting to note that Dust 11 event, the most intense
310 recorded event, presents the highest %SFe (between 10.2 and 19.8 % with an average at 13.8 %).
311 Apart from this event, the average fractional solubility seems to be independent of the EDM. Excluding
312 this event, the average solubility of Fe for dust event ($6.9 \% \pm 3.3 \%$) is equivalent to the one for
313 background samples. The uniformity of iron solubility values between background and dust periods
314 contrasted with the observations made in regions where the dust influence is sporadic and the origin
315 of Fe is associated to various sources (e.g. Shelley et al., 2018). That is consistent with a main dust
316 source of iron in our samples, as indicated in KL20. For both conditions, the observed range of vari-
317 ability is high and consistent with previous observations over the Southern Atlantic Ocean (2.4-20 %,
318 Baker et al., 2013; 1.3-22 %, Chance et al., 2015), as well as with measurements over and the South-
319 ern Indian Ocean (0.76-27 %, Gao et al., 2013), using acetate buffer leach at pH 4.7 (0.4 μ m) which
320 can extract 1.4 times more Fe than UPW protocol (Perron et al., 2020). Moreover, the measured iron
321 fractional solubility is significantly higher than obtained from dissolution experiments, with an identical
322 protocol, of mineral dust aerosol samples collected on filters after laboratory generation from the soils
323 collected in Namibian sources (< 1%; Formenti et al., in preparation, 2024).

324
325 The temporal variability of %SFe is presented in Fig. 2, where dust and background episodes are
326 shown separately.

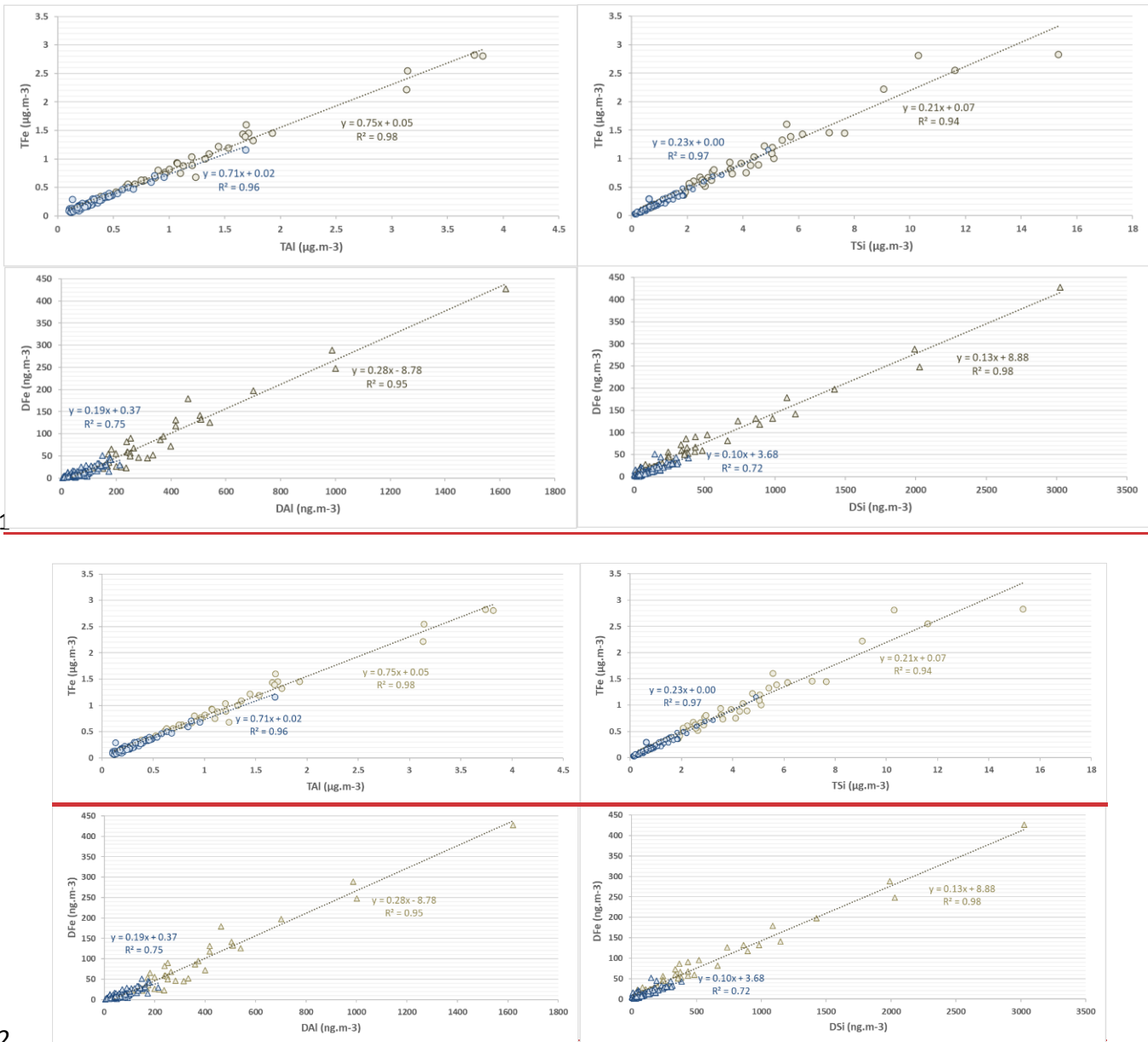


327
328 **Fig.32:** Temporal variability of %SFe average for dust and background samples during the different periods of
329 sampling. In the box plots, the box indicates the interquartile range, i.e. the 25th and the 75th percentile, and
330 the line within the box marks the median. The whiskers indicate the quartiles ± 1.5 times the interquartile range.
331 Points above and below the whiskers indicate outliers outside the 10th and 90th percentile.

332

333 The temporal variability of %SFe is presented in Fig. 3, where dust and background episodes are
 334 shown separately. The temporal variability is similar during dust and background conditions. The
 335 highest %SFe occurred during austral spring (October-November), and in particular during episode
 336 Dust 11 from 13 to 20 November 2017, when the average %SFe reached 13.8 %. The %SFe was
 337 quite similar along the year between dust and background, except between 13-20 November where
 338 the iron solubilities during Dust 11 event was very superior to the one of background samples, and to
 339 a lesser extent, in September (Dust 09) and December (Dust 13).

340 ~~Fig. 3 represents the correlations of Fe with Al and Si, both for the total and the dissolved concentrations.~~



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342

343 **Fig. 34.** Scatterplot of TFe with respect to TAI and TSi (top panels) and DFe with respect to DAI and DSi (bot-
 344 tom panels) for dust (sand dots and triangles) and background events (blue dots and triangles). The Pearson
 345 coefficient are shown for both.

346

347 **Fig. 4 represents the correlations of Fe with Al and Si, both for the total and the dissolved concentra-**
348 **tions.** For both dust and background samples, the total Fe concentration is linearly correlated with
349 total Al ($R^2=0.98$ and 0.96 , slope= 0.75 and 0.71 , for dust and background conditions respectively)
350 and total Si ($R^2=0.94$ and 0.97 , slope= 0.21 and 0.23 , respectively). The slopes are consistent with
351 typical Fe/Al and Fe/Si ratios found in desert dust from northern Africa (Formenti et al., [2014](#)[2011](#);
352 Shelley et al., 2014), confirming the main crustal origin of Fe during all the sampling periods. Likewise,
353 the concentrations of dissolved iron (DFe) show a strong linear correlation with both DAI and DSI, for
354 both for dust and background events ($R^2=0.96$ and 0.75 with respect to DAI and $R^2=0.98$ and 0.73
355 with respect to DSI). The slopes for Al and Si are also comparable (0.19 and 0.28 for DAI and 0.10
356 and 0.13 for DSI, respectively in dust and in background events). A very strong linear correlation was
357 also observed between DFe and DTi ($R^2=0.96$ and 0.84 ; not shown), another unique marker of min-
358 eral dust. Significant correlations of soluble concentrations for several elements associated with min-
359 eral dust (Fe, Al, Si, Ti) have been previously obtained in remote aerosols over ocean area (Baker et
360 al., 2016). Additionally, DFe during dust events correlate very closely with F^- ($R^2=0.94$, not shown),
361 which has been indicated by KL20 as being emitted in the atmosphere by the wind erosion as well as
362 the labouring of the Namibia soil, rich in fluoride mineral deposits.

364 4. Discussion

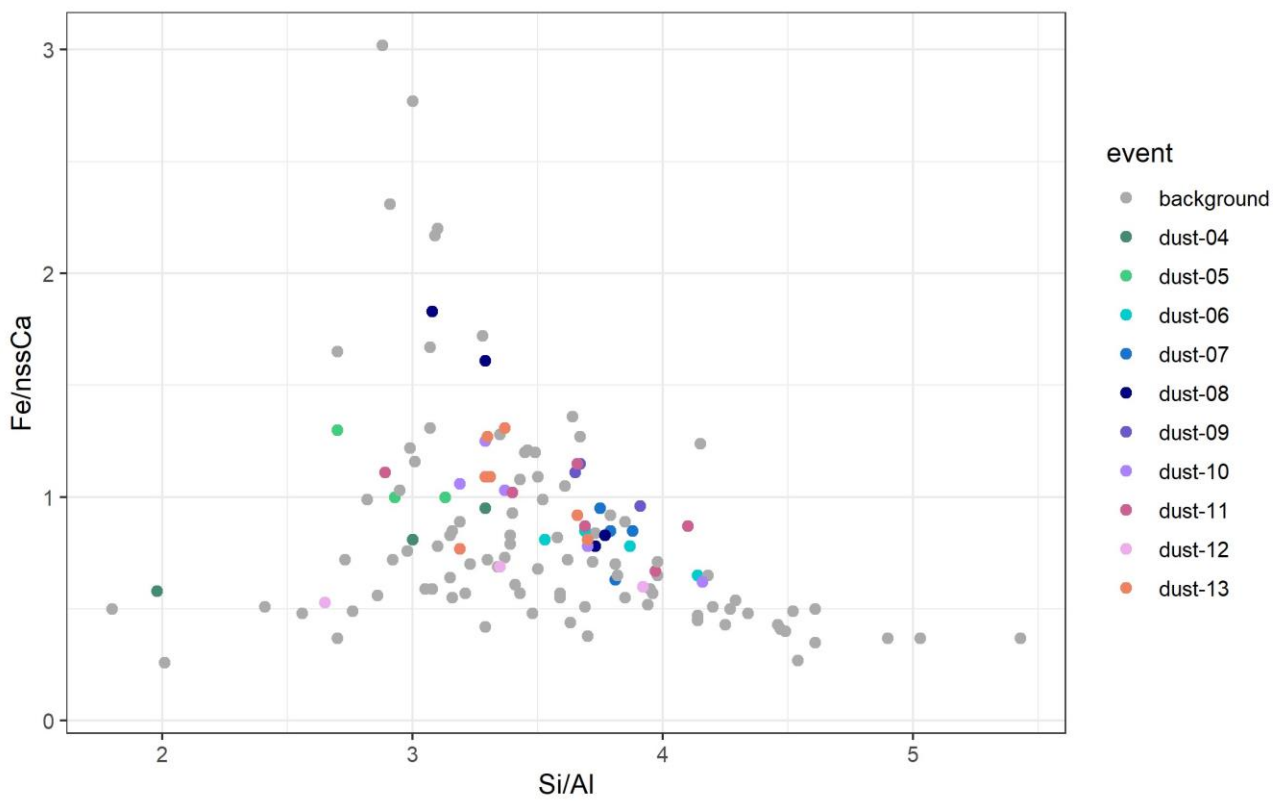
365 Several studies have showed that variations in aerosol Fe solubility could result from the source/com-
366 position of the aerosols. As a matter of fact, the Fe solubility has been linked to the iron mineralogy
367 (Journet et al., 2008) and has been shown being lower for African crustal sources than in continen-
368 tal/anthropogenic sources (Desboeufs et al., 2005; Sholkovitz et al., 2009; Shelley et al., 2018). The
369 iron fractional solubility in mineral dust is also affected by source mixing (Paris et al., 2010; Desboeufs
370 et al., 2005), by (photo)chemical processing with acids or organic ligands during atmospheric
371 transport (Paris et al., 2011, Paris et Desboeufs, 2013; Wozniak et al., 2013; Swan and Ivey, 2021)
372 and by the increase of surface area to volume ratio due to size changes during transport (Baker &
373 Jickells, 2006; Marcotte et al., 2020).

374 In the following sections, we discuss these possible factors to explain the seasonality and the ex-
375 tended range of variability of the fraction Fe solubility in HBAO samples. The possible increase of
376 surface area to volume ratio during transport (Baker and Jickells, 2006; Marcotte et al., 2020) will not
377 be discussed because of lack of appropriate observations of the size distribution. Because of the
378 similar transport time suggested by back trajectories (Fig. [S12](#)), it is likely that particle size distribution
379 would be similar from one event to the other.

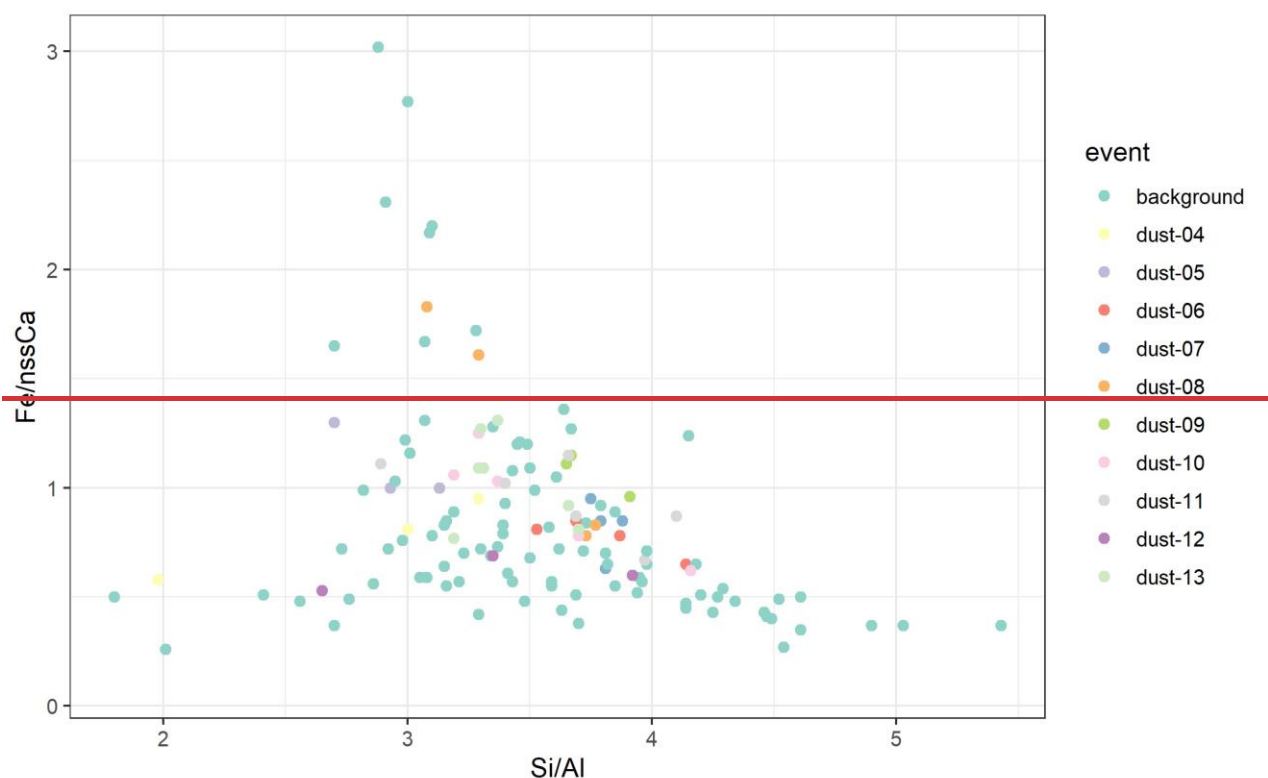
380 4.1. Influence of dust composition

381 Close to dust source, iron solubility could be mainly conditioned by the mineralogical composition of
382 dust (Journet et al., 2008, Formenti et al., 2014). Considering that soluble Fe-bearing aerosols were
383 issued from mineral dust for all the samples, the seasonality of dust emission sources (see 3.1) could
384 be a factor explaining the seasonality of %SFe (and other elements associated to mineral dust). **Fig.**
385 **4-5** shows the scatter plot of the elemental mass ratio of Fe/nss-Ca²⁺ and Si/Al, previously used for
386 northern Africa dust to distinguish aerosol dust from source areas enriched in clays or iron oxides to
387 soils rich in quartz or carbonates (Formenti et al., 2014). Specific to Namibia, because of the strong
388 link between nss-Ca²⁺ and fluorine, the Fe/nss-Ca²⁺ ratio may also to distinguish dust influenced by
389 fluorspar mining.

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Fig 45. Scatterplot of $Fe/nss-Ca^{2+}$ and Si/Al mass ratios for the samples collected at HBAO in period May-December 2017. Values obtained for samples collected during the dust events are represented as brown-colored dots. Values for samples collected outside those events (background) are represented as blue-grey dots.

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Figure 4-5 indicates that the range of variability of both $Fe/nss-Ca^{2+}$ and Si/Al ratios is small when considering dust events only. The elemental ratios of samples collected during the background periods are rather similar to dust events during a same sampling period, except for Si/Al for the period between 19-26 May and for $Fe/nss-Ca^{2+}$ for the samples of 18-25 September, when significant differences, not really explicable and not inducing a significant difference in the %SFe values are observed (Fig. S3).

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The values for ambient dust measured at HBAO are consistent with those of the previous field observations in Namibia (Annegarn et al., 1983; Eltayeb et al., 1993), but also with values reported by Caponi et al. (2017) for laboratory-aerosolised dust from two soils collected on the Namibian gravel plains. This is in agreement of the indications of the emission maps (Fig. S12), showing significant emissions in the gravel plains. The absence of seasonal cycle in the elemental composition illustrated in Fig. S3 suggests that the seasonal change from northern to southern sources does not induce a change in the composition of the aerosol dust sampled at HBAO, which is consistent with the fact that the northern and the southern gravel plains of Namibia have similar mineralogy (Heine and Vökel, 2010). This suggests that the mineralogical composition of mineral dust should not be a discriminating factor explaining the seasonality of the iron solubility observed at HBAO.

4.2. Evidence of processing by marine biogenic emissions

The atmospheric (in-cloud) processing associated with secondary aerosol production may increase the fractional solubility of Fe during transport (Takahashi et al., 2011; Rodríguez et al., 2021). This has also been shown for Al and Ti (Baker et al., 2020). The chemical processing could include both acidic and ligand-promoted dissolution (Desboeufs et al., 2001, Longo et al., 2016, Tao et al., 2019). Oxalic acid has previously been used as a proxy for organic ligand-mediated iron dissolution processes because it is the most abundant species in the atmosphere and is the most effective ligand in promoting iron dissolution (Baker et al., 2020; Hamilton et al., 2021). However, several secondary compounds, such as carboxylate ligands and marine secondary products derived from dimethyl sulfide (DMS) oxidation, have been identified as playing a role in increasing the ~~solubility~~ soluble fraction of iron from mineral aerosols (Johansen and Key, 2006; Paris et al., 2011; Paris and Desboeufs, 2013; Wozniak et al., 2013 and 2015). The increase of ligands-promoted dissolution is attributed to photochemical reduction of Fe(III) in Fe (II) (Siefert et al., 1994; Johansen and Key, 2006).

To investigate these aspects, the mass concentrations of the ionic compounds (oxalate, formate, MSA, NO_3^- , NH_4^+ and nss-SO_4^{2-}) implied in the secondary aerosol production, measured at HBAO during dust and background periods are reported in **Table 3**.

429

Table 3. Average and standard deviations of mass concentrations of water-soluble ions measured at HBAO during dust and background events from May to December 2017. Concentrations are expressed in ng m^{-3} . The number of samples pertaining to each occurrence is indicated in brackets.

	Dust	Background
nss-SO_4^{2-}	1795 ± 762 (N = 42)	1366 ± 505 (N=132)
Oxalate	155 ± 53 (N = 42)	127 ± 35 (N = 132)
Formate	18 ± 6 (N = 40)	16 ± 9 (N = 105)
MSA	64 ± 37 (N=36)	56 ± 36 (N=114)
NO_3^-	205 ± 79 (N=42)	200 ± 138 (N=132)
NH_4^+	192 ± 71 (N=42)	207 ± 98 (N=132)

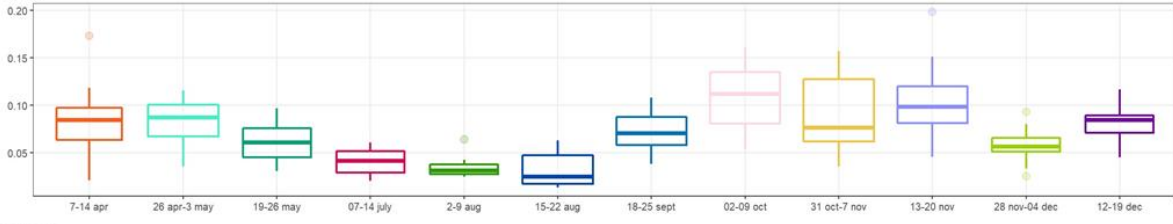
433

Oxalate was the most abundant organic compound, followed by MSA, a secondary product of DMS oxidation and a unique particulate tracer of the primary marine biogenic activity (Andreae et al., 1995). On average, organic compounds were equally concentrated in dust and background events. Amongst inorganic species, nss-SO_4^{2-} was the most concentrated compound, with higher values during the dust events than during the background period.

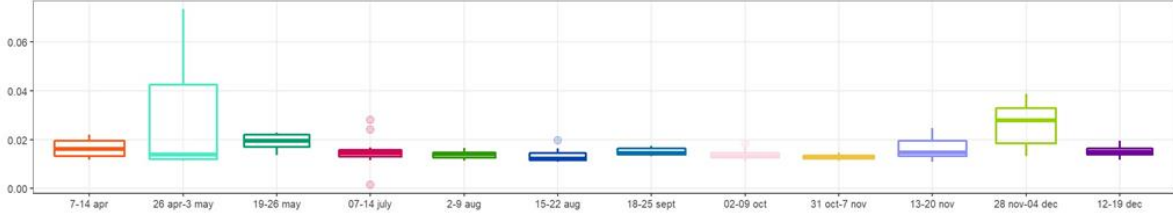
~~Their detailed time series are shown in Fig 5, where it is compared to that of the iron fractional solubility.~~

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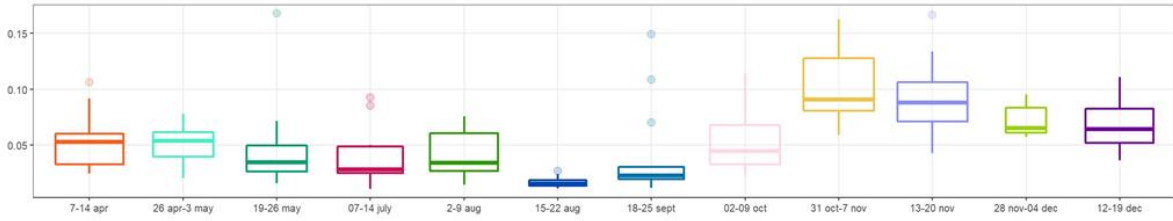
FS(Fe)



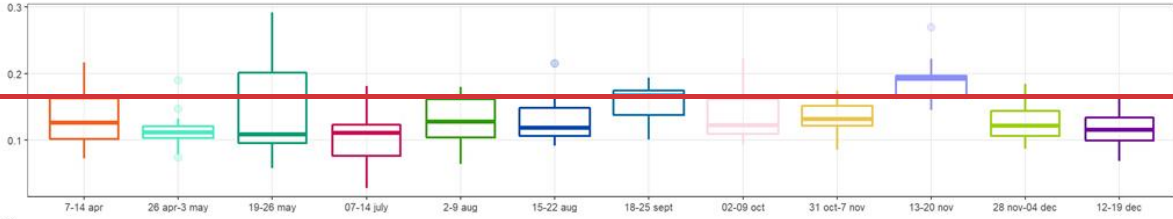
Formate



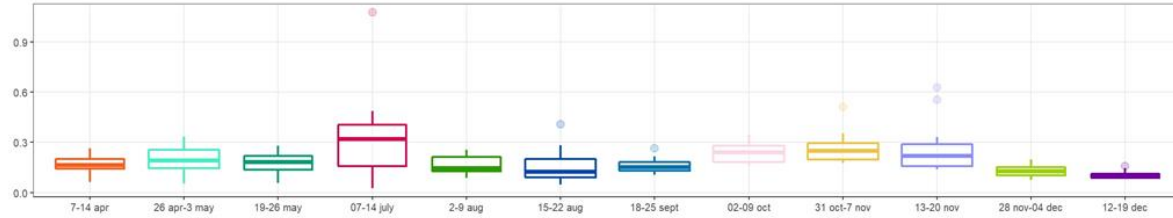
MSA



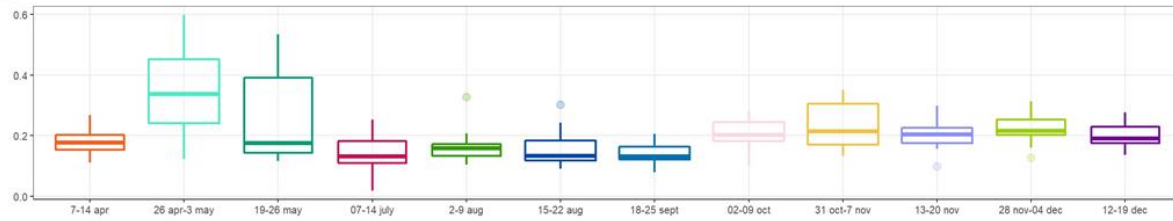
Oxalate



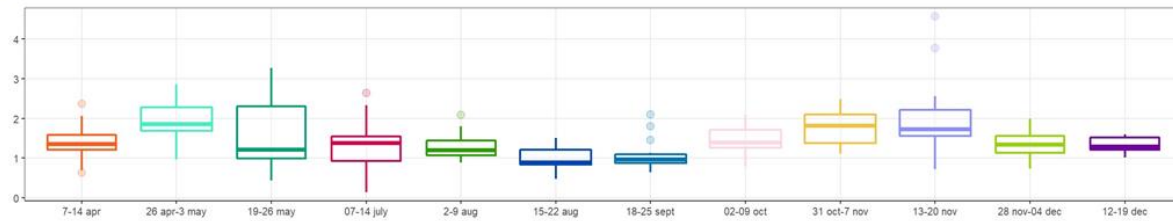
NO3



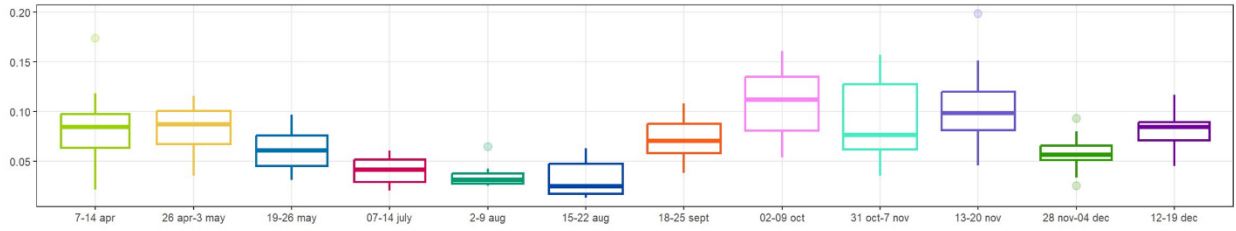
NH4



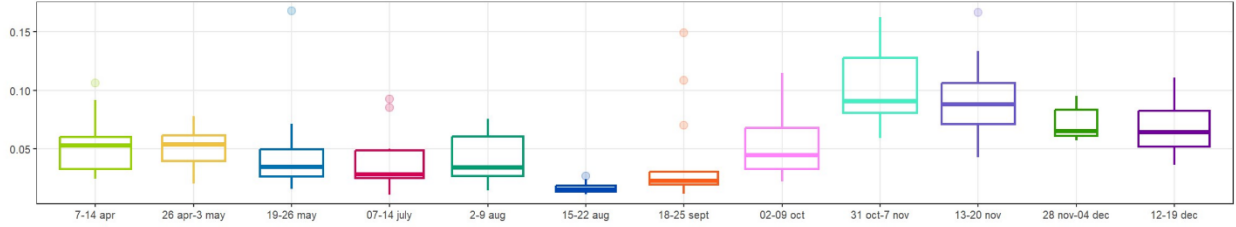
nssSO4



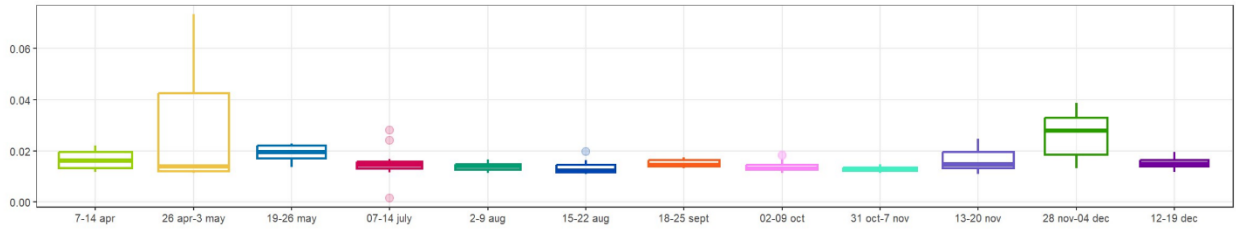
%SFe



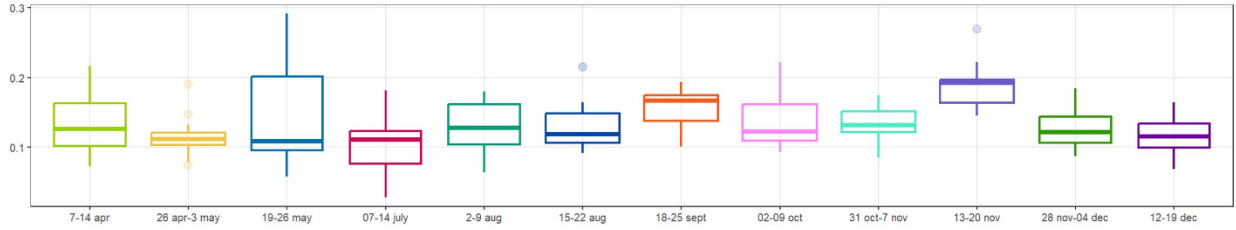
MSA



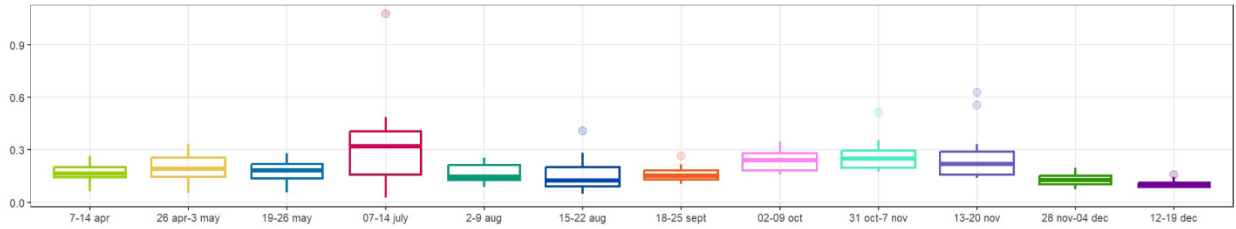
Formate



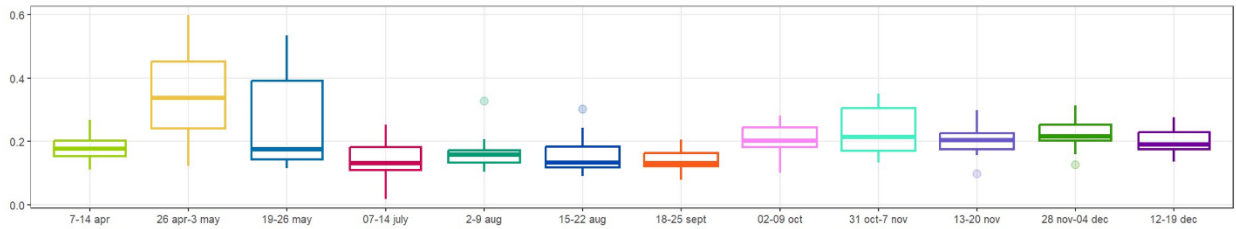
Oxalate



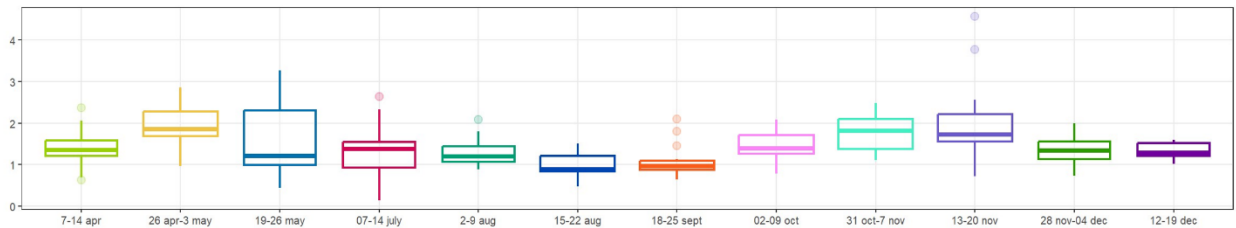
NO3



NH4



nssSO4



443 **Fig. 56.** Box-plots of the averages of %SFe and secondary organic and inorganic compounds mass concentrations ($\mu\text{g m}^{-3}$) for the sampling periods including all the samples (dust + background). Boxes and whiskers as in Fig. 2.

445

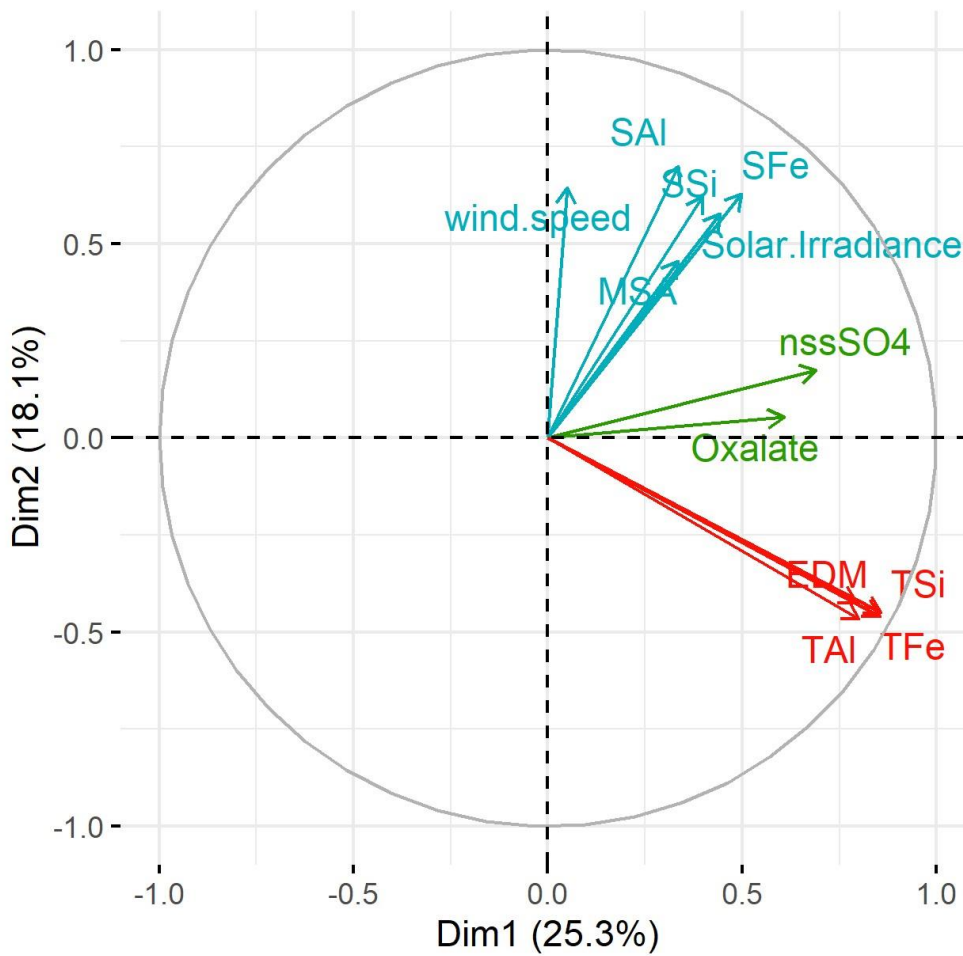
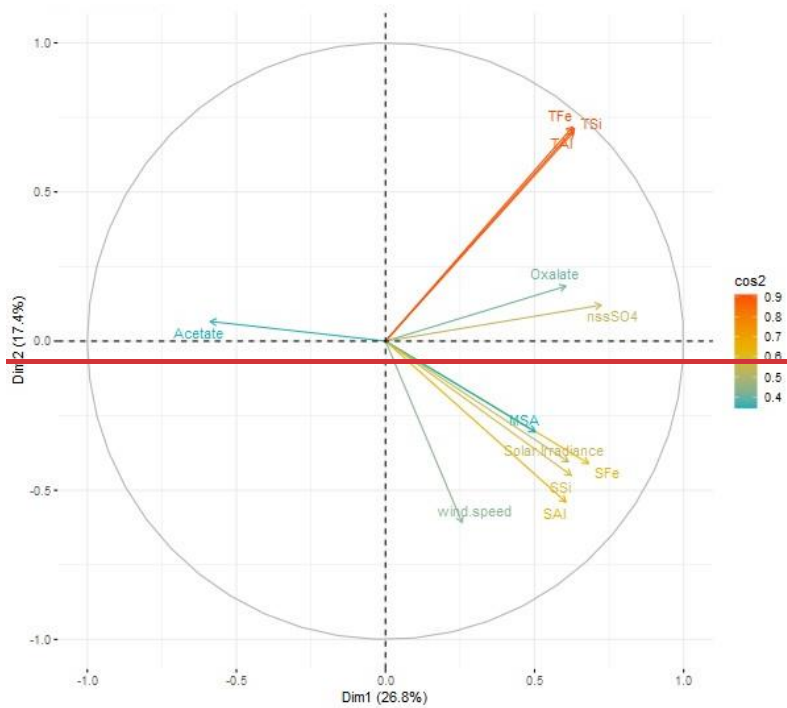
446 Their detailed time series are shown in Fig 56, where it is compared to that of the iron fractional
447 solubility.

448 There is no clear seasonal cycle for any of the ionic compounds, with the exception of MSA, which
449 shows a similar time variability than %SFe. MSA concentrations were lowest between May and Au-
450 gust (average $38.0 \pm 28.0 \text{ ng m}^{-3}$), while higher concentrations were measured from September to
451 December ($72.7 \pm 38.1 \text{ ng m}^{-3}$). These differences are also observed for the dust cases only. The
452 average MSA concentration was $40.6 \pm 23.4 \text{ ng m}^{-3}$ for Dust 04 to Dust 08 episodes. It increased to
453 $77.7 \pm 35.3 \text{ ng m}^{-3}$, almost a factor of 2 between episodes Dust 09 and Dust 13. The mass concen-
454 trations and the seasonal cycle of MSA are related with the proximity of the strong coastal upwelling
455 by the Benguela current (Formenti et al., 2019; KL20). The maximum concentration of MSA (106.2
456 ng m^{-3}) was measured during episode Dust 11, which is also the time of the highest SFe% observa-
457 tion. This episode was also characterised by the highest oxalate, nss-SO_4^{2-} and NO_3^- concentrations.

458 However, no clear correlation between the %SFe and the secondary compounds concentrations can
459 be found in our data (Fig. S4). In order to statistically explore the potential links

460 Based on their temporal variability between %SFe and various these concentrations, Fig. 6-7 shows
461 the correlation-correspondence plot between total Fe, Al and Si, and their respective fractional solu-
462 bility, the measured secondary compounds and the meteorological conditions during sampling ob-
463 tained from Principal Component Analysis (PCA) for all the samples. The variables correlated in time
464 are grouped together (the closer they are ~~to~~ in the circle, the stronger the correlation) whereas the
465 variables which are anti-correlated are situated on the opposite side of the plot origin.

466



467 **Fig 67.** PCA analysis performed from the database including %SX, TX, EDM, -and-secondary ions concentrations and me-
 468 teorological parameters. The colour of variables by groups is defined by a clustering algorithm, tending to find clusters of
 469 comparable spatial extent. Each colour corresponds to a cluster of parameters which evolve in the same way. The

470 ~~scale(\cos^2) gives the factor of correlation between the different parameters. Formate, nitrate, and ammonium, acetate,~~
471 ~~humidity, and wind speed are not visible in the plot showing that they are not significantly correlated with the other param-~~
472 ~~eters (i.e. their squared cosine < 0.4).~~

473

474 The PCA ~~correlation plot (Fig.7)~~ emphasizes 3 groups of dependent parameters: 1. a high correlation
475 between total Fe, Al and Si concentrations and dust loading (EDM), as previously identified in Table
476 1 and Fig.3, 2. a relation between oxalate and nss-SO₄²⁻ concentrations, suggesting a common chem-
477 ical process of formation, and 3. the dependence between %SFe (%SAI and %SSi), and the MSA
478 concentrations (correlation factor around 0.4), while indicating a weak dependence on oxalate, ace-
479 tate and nss-SO₄²⁻. Fig. 6 also shows that %SFe is correlated with both the wind speed and the solar
480 irradiance (correlation factor higher than 0.6) and to a lesser extent with the wind speed. While it is
481 expected that the emission of mineral dust occurs when the wind speed is high, the correlation of
482 %SFe with wind speed is rather surprising ~~as both Table 1 and Fig.2 show that since~~ the %SFe is
483 independent of the dust load ~~(Fig. 2 and 7).~~ **Fig. S4** in the supplementary material present the plots
484 between %SFe, MSA concentrations, solar irradiance and wind speed for background and dust
485 events. The correlation between shows that the wind speed is also correlated with and the MSA con-
486 centrations (Fig. S4). This is consistent with Andreae et al. (1995), who demonstrated how, in this
487 area due to persistent phytoplankton bloom, the atmospheric concentrations of dimethylsulphide
488 (DMS), the gaseous precursors of MSA, depend on the sea-to-air flux, in turn is determined by the
489 concentrations in the ocean water and the surface wind speed. ~~On the other hand, the MSA concen-~~
490 ~~trations do not correlate significantly with the average solar irradiance.~~

491 As previously mentioned, Johansen and Key (2006) showed an increase of dissolution of ferrihydrite,
492 a proxy of iron(oxy)hydroxide found in desert mineral dust, by photolysis of the Fe(III)-MSIA (methane
493 sulfinic acid) complex, producing MSA and soluble Fe. Zhuang et al. (1992) proposed an increase of
494 iron dissolution by the acidification of aerosol particles associated with dimethylsulphide (DMS) oxi-
495 dation. Here, the link between the Fe fractional solubility, solar irradiance and MSA is in agreement
496 with the photo-reduction dissolution of Fe by MSA condensation on Fe-bearing dust. Thus, we attrib-
497 ute the iron fractional solubility seasonality observed at HBAO both to solar irradiance and MSA tem-
498 poral evolution via this process. It is interesting to note that due to the high correlation between %SFe
499 and %SAI and %SSi, the photochemical processes could also impact the solubility of all element-
500 bearing dust.

501 **4.3. Link to other sources of iron and oxalate**

502 Formenti et al. (2018) have shown that in the Austral winter, when the synoptic circulation is domi-
503 nantly anti-cyclonic, air masses laden with light-absorbing aerosols either from ship pollution or bio-
504 mass burning can be transported to HBAO (Formenti et al., 2018). However, the lowest Fe solubility

505 (< 5%) was measured in July and August 2017, and no correlation between the %SFe and the percent
506 mass fraction of iron from sources other than dust can be found in our data (not shown).

507 The mass apportionment of iron reported by KL20 indicates that, during the dust events and the
508 background periods, respectively, 7% and 29% of the mass of total elemental Fe was not associated
509 to mineral dust, but rather to a factor indicated as “ammonium-neutralised component”, mostly char-
510 acterised by secondary species, and non-sea-salt potassium (nss-K⁺). The PMF analysis indicated
511 that Because of this association, previously reported by Andreae (1983), the “ammonium-neutralised
512 component” was associated to photo-oxidation of marine biogenic emission but also episodically to
513 biomass burning, which can be transported to HBAO during the Austral summertime, when the airflow
514 becomes anti-cyclonic, and the transport of air masses laden with light absorbing aerosols has been
515 documented (Formenti et al., 2018).

516 This component includes oxalate, the most concentrated organic species at HBAO, and the strongest
517 of the organic ligands promoting the photo-reduction of iron in mineral dust, henceforth the increase
518 of its fractional solubility (Paris and Desboeufs, 2013). Surprisingly, excepted individual cases (Dust
519 13), our analysis does not show this strong link (Fig. 7), and indeed, contrary to the SFe%, the oxalate
520 concentrations measured at HBAO was practically constant with time (in average 0.14±/ 0.04 µg.m
521 ³). The possible pathways of oxalate formation in this complex atmosphere are numerous through the
522 year, from natural and anthropogenic sources (marine, heavy-oil combustion, biomass burning) and
523 in-cloud and photo-oxidative processes (Baboukas et al., 2000; Myriokefalitakis et al., 2011).

524 ~~However, our data do not indicate any significant dependence of %SFe to the percent mass fraction~~
525 ~~of iron attributed to sources other than dust, notably combustion particles, which was expected in the~~
526 ~~light of previous research (e.g. Desboeufs et al., 2005; Sholkovitz et al., 2009; Shelley et al., 2018;~~
527 ~~Ito et al., 2021), and indeed the lowest Fe solubility (< 5%) was measured in July and August 2017,~~
528 ~~when the contribution of polluted air masses should be highest.~~

529 ~~The “ammonium-neutralised component” identified by KL20 included oxalate, the most concentrated~~
530 ~~organic species at HBAO, and the strongest of the organic ligands promoting the photo-reduction of~~
531 ~~iron in mineral dust, henceforth the increase of its fractional solubility (Paris and Desboeufs, 2013).~~
532 ~~Surprisingly, excepted individual cases (Dust 13), our analysis does not show this strong link (Fig. 6),~~
533 ~~which we explain by the fact that, contrary to the SFe%, the oxalate concentrations measured at~~
534 ~~HBAO was practically constant with time, the possible pathways of oxalate formation in this complex~~
535 ~~atmosphere being numerous and occurring through the year, from natural and anthropogenic sources~~
536 ~~(marine, heavy-oil combustion, biomass burning) and in-cloud and photo-oxidative processes (Ba-~~
537 ~~boukas et al., 2000; Myriokefalitakis et al., 2011).~~

539 **5. Conclusive-Concluding remarks**

540 For the first time, the fractional solubility of Fe in airborne atmospheric aerosols smaller than 10 μm
541 in diameter is investigated along the west coast of Namibia, in southern Africa, a critical region for the
542 global climate.

543 Ten intense episodes of transport of mineral dust from aeolian erosion were identified from the anal-
544 ysis of aerosol samples collected between May and December 2017 at the Henties Bay Aerosol Ob-
545 servatory (HBAO). Based on modelling and measurements, source regions were identified both in the
546 northern and southern gravel plains. Our data do not provide any evidence of the possible contribution
547 of dust from coastal riverbeds, which are considered to be frequent sources of atmospheric dust and
548 soluble iron in the region (Vickery et al., 2013; Von Holdt et al., 2017; Dansie et al., 2017a; 2017b).

549 ~~Our first measurement indicate that~~ †The total iron represents, on average, 5.8 % (\pm 0.6 %) of the total
550 dust mass, and that the average iron-water-soluble Fe fractional solubility is 6.9 % (\pm 3.3 %). These
551 values should be useful to atmospheric models estimating the dust-borne input of soluble Fe from the
552 gravel plains in Namibia to the surrounding oceans.

553 ~~The measured iron fractional solubility is comparable to values reported from shipborne measure-~~
554 ~~ments of transported dust in the remote southern oceanic regions (Baker et al., 2013; Chance et al.,~~
555 ~~2015, Gao et al., 2013) but significantly higher than obtained in a benchmark laboratory evaluation~~
556 ~~from the same soils and an identical dissolution protocol (unpublished data). The time series of frac-~~
557 ~~tional solubility of Fe shows an apparent seasonal cycle which is independent of dust composition.~~
558 ~~This is also the case for Al and Si.~~

559 ~~The observations presented in this paper exclude a major role of sources other than mineral dust to~~
560 ~~play on the values and the variability of %SFe, which might be due to the location of our sampling~~
561 ~~site, remote and only occasionally affected by polluted air masses (Formenti et al., 2018).~~

562 Conversely, †The seasonal increase of the iron fractional solubility is associated to that of the concen-
563 trations of MSA and correlated to meteorological parameters such as the wind speed and the surface
564 solar irradiance. Our observations support the role of photo-chemical processes in the dissolution of
565 Fe in our samples, and suggest that the oxidation of the marine biogenic emissions from the northern
566 Benguela upwelling, favoured under high wind speed conditions, could play a significant role in in-
567 creasing the solubility of elemental iron in mineral dust aerosols over coastal Namibia. This is in
568 agreement with the mechanism described by Zhuang et al. (1992), who proposed an increase of iron
569 dissolution by the acidification of aerosol particles associated with DMS oxidation, and Johansen and
570 Key (2006), who showed an increase of dissolution of ferrihydrite, a proxy of iron(oxy)hydroxide found
571 in desert mineral dust, by photolysis of the Fe(III)-MSIA (methane_sulfinic acid) complex, producing
572 MSA and soluble Fe. It is interesting to note that due to the high correlation between %SFe and %SAI
573 and %SSi (and %STi), the same photochemical processes could also impact the solubility of all ele-

574 ment-bearing dust. The possible mechanism suggested by this paper could be responsible for initiating a feedback loop whereby the input of dust of increased trace and major elements solubility would result in stronger marine biogenic emissions to the atmosphere. This possible mechanism could increase the iron solubility in mineral dust, maybe also initiating a feedback loop whereby the input of dust of increased solubility would result in stronger marine biogenic emissions to the atmosphere, including, Beside sulphur species, the role of Volatile Organic Compounds (VOCs), in particular butene, massively emitted by the organisms in the coastal marine foam (Giorio et al., 2022), should also be explored).

~~In conclusion, this paper describes the very first field observations suggesting that, while airborne, the atmospheric iron from mineral dust experiences a~~This complex and dynamic environment where the interplay between the input of atmospheric iron from transported dust and the marine biogenic emissions from the Benguela oceanic upwelling system should be further addressed by future research. ~~This possible mechanism could increase the iron solubility in mineral dust, maybe also initiating a feedback loop whereby the input of dust of increased solubility would result in stronger marine biogenic emissions to the atmosphere. Beside sulphur species, the role of Volatile Organic Compounds (VOCs), in particular butene, massively emitted by the organisms in the coastal marine foam (Giorio et al., 2022), should also be explored.~~

591

592 **Data availability.** Original and analysed data are available at the AERIS ([https://aerocio.aeris-](https://aerocio.aeris-data.fr/project/)
593 [data.fr/project/](https://aerocio.aeris-data.fr/project/), last accessed 20/07/2023). The statistical FactoMineR package is available in R (R
594 version 4.1.2, 2021; http://factominer.free.fr/index_fr.html, last accessed 20/07/2023). Meteorological
595 data from the Wlotzkasbaken station (22.31°S, 14.45°E, 73 m asl) are part of the Southern African
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598

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600 RT, KD, PF, SC, ST, and CMB performed the XRF, IC and ICP analysis of the collected samples. BL
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602 fluxes. JPC performed the model calculations of air mass back-trajectories. HA and JC provided with
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604 PF and KD wrote the paper with contributions from RT and SJP, and the remaining authors. PF and
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607

608 **Competing interests.** PF is guest editor for the ACP Special Issue “New observations and related
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611

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632 **References**

- 633 Andersen, H. and Cermak, J.: First fully diurnal fog and low cloud satellite detection reveals life cycle in the
634 Namib, *Atmos. Meas. Tech.*, 11, 5461–5470, doi: 10.5194/amt-11-5461-2018, 2018.
- 635 Andersen, H., Cermak, J., Solodovnik, I., Lelli, L. and Vogt, R.: Spatiotemporal dynamics of fog and low clouds
636 in the Namib unveiled with ground- and space-based observations, *Atmos. Chem. Phys.*, 1, 4383–4392, doi:
637 10.5194/acp-19-4383-2019, 2019.
- 638 Andersen, H., Cermak, J., Fuchs, J., Knippertz, P., Gaetani, M., Quinting, J., Sippel, S., and Vogt, R.: Synoptic-
639 scale controls of fog and low-cloud variability in the Namib Desert, *Atmos. Chem. Phys.*, 20, 3415–3438,
640 <https://doi.org/10.5194/acp-20-3415-2020>, 2020.
- 641 Andreae, M. O.: Soot Carbon and Excess Fine Potassium: Long-Range Transport of Combustion-Derived Aer-
642 osols, *Science*, 220, 1148-1151, doi:10.1126/science.220.4602.1148, 1983.
- 643 Andreae, M. O., Elbert, W., and de Mora, S. J.: Biogenic sulfur emissions and aerosols over the tropical South
644 Atlantic: 3. Atmospheric dimethylsulfide, aerosols and cloud condensation nuclei, *J. Geophys. Res.*, 100,
645 11335-11356, <https://doi.org/10.1029/94JD02828>, 1995.
- 646 Annegarn, H.J., van Grieken, R.E., Bibby, D.M. and von Blottnitz, F.: Background Aerosol Composition in the
647 Namib Desert, South West Africa (Namibia), *Atmos. Environ.*, 17, 2045–2053, doi: 10.1016/0004-
648 6981(83)90361-X, 1983.
- 649 Bhattachan, A., D'Odorico, P., Baddock, M.C., Zobeck, T.M., Okin, G.S., Cassar, N.: The Southern Kalahari: a
650 potential new dust source in the Southern hemisphere?, *Environ. Res. Lett.*, 7, 024001.
651 <http://dx.doi.org/10.1088/1748-9326/7/2/024001>, 2012.
- 652 Bhattachan, A., P. D'Odorico, and G. S. Okin, Biogeochemistry of dust sources in Southern Africa, *J. Arid En-
653 viron.*, 117, 18-27, <http://dx.doi.org/10.1016/j.jaridenv.2015.02.013>, 2015.
- 654 Baboukas, E. D., Kanakidou, M., and Mihalopoulos, N.: Carboxylic acids in gas and particulate phase above
655 the Atlantic Ocean, *J. Geophys. Res.*, 105, 14459–14471, <https://doi.org/10.1029/1999JD900977>, 2000.
- 656 Baker A.R., T. D. Jickells, M. Witt, and K. L. Linge, Trends in the solubility of iron, aluminium, manganese and
657 phosphorus in aerosol collected over the Atlantic Ocean., *Marine Chem.*, 98, 43-58,
658 <https://doi.org/10.1016/j.marchem.2005.06.004>, 2006.
- 659 Baker, A. R., and T. D. Jickells, Mineral particle size as a control on aerosol iron solubility, *Geophys. Res. Lett.*,
660 33, L17608. <https://doi.org/10.1029/2006GL026557>, 2006.
- 661 Baker, A. R., C. Adams C., T. G. Bell, T. D. Jickells, and L. Ganzeveld, Estimation of atmospheric nutrient inputs
662 to the Atlantic Ocean from 50°N to 50°S based on large-scale field sampling: iron and other dust-associated
663 elements, *Glob. Biogeochem. Cycles*, 27, 755–767, doi:10.1002/gbc.20062, 2013.
- 664 Baker, A. R., M. Thomas, H. W. Bange, and E. Plasencia Sánchez, E., Soluble trace metals in aerosols over
665 the tropical south-east Pacific offshore of Peru, *Biogeosciences*, 13, 817–825, [https://doi.org/10.5194/bg-13-](https://doi.org/10.5194/bg-13-817-2016)
666 817-2016, 2016.

667 Baker, Alex R., et T. D. Jickells, Atmospheric deposition of soluble trace elements along the Atlantic Meridional
668 Transect (AMT), in *The Atlantic Meridional Transect programme (1995-2016)* 158: 41-51.
669 <https://doi.org/10.1016/j.pocean.2016.10.002>, 2017.

670 Baker, A. R., M. Li, and R. Chance, Trace metal fractional solubility in size-segregated aerosols from the tropical
671 eastern Atlantic Ocean, *Global Biogeochemical Cycles*, 34, e2019GB006510.
672 <https://doi.org/10.1029/2019GB006510>, 2020.

673 Bryant, R. G., Bigg, G. R., Mahowald, N. M., Eckardt, F. D., and Ross, S. G., Dust emission response to climate
674 in southern Africa, *J. Geophys. Res.*, 112, D09207, doi:10.1029/2005JD007025, 2007.

675 Caponi, L., Formenti, P., Massabó, D., Di Biagio, C., Cazaunau, M., Pangui, E., Chevaillier, S., Landrot, G.,
676 Andreae, M. O., Kandler, K., Piketh, S., Saeed, T., Seibert, D., Williams, E., Balkanski, Y., Prati, P., and
677 Doussin, J.-F.: Spectral- and size-resolved mass absorption efficiency of mineral dust aerosols in the
678 shortwave spectrum: a simulation chamber study, *Atmos. Chem. Phys.*, 17, 7175–7191,
679 <https://doi.org/10.5194/acp-17-7175-2017>, 2017.

680 Chaboureau, J.-P., L. Labbouz, C. Flamant, and A. Hodzic. Acceleration of the southern African easterly jet
681 driven by radiative effect of biomass burning aerosols and its impact on transport during AEROCLO-sA, *At-
682 mos. Chem. Phys.*, 22, 8639-8658, <https://doi.org/10.5194/acp-22-8639-2022>, 2022.

683 Chance R, T. D. Jickells and A. R. Baker, Atmospheric trace metal concentrations, solubility and deposition
684 fluxes in remote marine air over the south-east Atlantic, *Marine Chemistry*, 177, 45–56, doi:10.1016/j.mar-
685 chem.2015.06.028, 2015.

686 Dansie, A. P., G. F. S. Wiggs, D. S. G. Thomas, and R. Washington, Measurements of windblown dust charac-
687 teristics and ocean fertilisation potential: The ephemeral river valleys of Namibia, *Aeolian Res.*, 29, 30–41,
688 doi:10.1016/j.aeolia.2017.08.002, 2017a.

689 Dansie, A. P., G. F. S. Wiggs, and D. S. G. Thomas, Iron and nutrient content of wind-erodible sediment in the
690 ephemeral river valleys of Namibia, *Geomorphology*, 290, 335-346, <https://doi.org/10.1016/j.geomorph.2017.03.016>, 2017b.

692 Dansie AP, Thomas DSG, Wiggs GFS, Baddock MC, Ashpole I. Plumes and blooms - Locally-sourced Fe-rich
693 aeolian mineral dust drives phytoplankton growth off southwest Africa. *Sci Total Environ.*, doi: 10.1016/j.sci-
694 totenv.2022.154562, 2022.

695 Desboeufs, K. V., R. Losno, et J. L. Colin, Factors influencing aerosol solubility during cloud processes, *Atmos.
696 Environ.*, 35, 3529-3537, [https://doi.org/10.1016/S1352-2310\(00\)00472-6](https://doi.org/10.1016/S1352-2310(00)00472-6), 2001.

697 Desboeufs, K.V., Sofikitis, A., Losno, R., Colin, J.L., Ausset, P. Dissolution and solubility of trace metals from
698 natural and anthropogenic aerosol particulate matter, *Chemosphere* 58, 195–203, 2005.

699 Desboeufs, K., Fu, Y., Bressac, M., Tovar-Sánchez, A., Triquet, S., Doussin, J.F., Giorio, C., Chazette, P.,
700 Disnaquet, J., Feron, A., Formenti, P., Maisonneuve, F., Rodríguez-Romero, A., Zapf, P., Dulac, F., and
701 Guieu, C., Wet deposition in the remote western and central Mediterranean as a source of trace metals to
702 surface seawater, *Atmos. Chem. Phys.*, 22, 2309–2332, <https://doi.org/10.5194/acp-22-2309-2022>, 2022.

703 Eltayeb, M.A.; Van Grieken, R.E., Maenhaut, W. and Annegarn, H.J.: Aerosol-Soil Fractionation for Namib De-
704 sert Samples. *Atmos. Environ.*, 27(5), [https://doi.org/10.1016/0960-1686\(93\)90185-2](https://doi.org/10.1016/0960-1686(93)90185-2), 1993.

705 Feuerstein, S., and Schepanski, K.: Identification of Dust Sources in a Saharan Dust Hot-Spot and Their Imple-
706 mentation in a Dust-Emission Model, *Remote Sensing*, 11, doi:10.3390/rs110100004, 2019.

707 Flamant, C., M. Gaetani, J.-P. Chaboureau, P. Chazette, S. J. Piketh, and P. Formenti. Smoke in the river: an
708 Aerosols, Radiation and Clouds in southern Africa (AEROCLO-sA) case study, *Atmos. Chem. Phys.*, 22,
709 5701–5724, <https://doi.org/10.5194/acp-22-5701-2022>, 2022.

710 [Formenti, P., Schütz, L., Balkanski, Y., Desboeufs, K., Ebert, M., Kandler, K., Petzold, A., Scheuven, D.,
711 Weinbruch, S., and Zhang, D.: Recent progress in understanding physical and chemical properties of African
712 and Asian mineral dust, *Atmos. Chem. Phys.*, 11, 8231–8256, <https://doi.org/10.5194/acp-11-8231-2011>,
713 2011.](#)

714 Formenti, P., S. Caquineau, K. Desboeufs, A. Klaver, S. Chevallier, E. Journet, J. L. Rajot, Mapping the phys-
715 ico-chemical properties of mineral dust in western Africa: mineralogical composition, *Atmos. Chem. Phys.*, 14,
716 10663-1068, 2014.

717 Formenti, P., Piketh, S. J., Namwoonde, A., Klopper, D., Burger, R., Cazaunau, M., Feron, A., Gaimoz, C.,
718 Broccardo, S., Walton, N., Desboeufs, K., Siour, G., Hanghome, M., Mafwila, S., Omoregie, E., Junkermann,
719 W., and Maenhaut, W.: Three years of measurements of light-absorbing aerosols over coastal Namibia: sea-
720 sonality, origin, and transport, *Atmos. Chem. Phys.*, 18, 17003-17016, [https://doi.org/10.5194/acp-18-17003-](https://doi.org/10.5194/acp-18-17003-2018)
721 2018, 2018.

722 Formenti, P., B. D'Anna, C. Flamant, M. Mallet, S.J. Piketh, K. Schepanski, F. Waquet, F. Auriol, G. Brogniez,
723 F. Burnet, J. Chaboureau, A. Chauvigné, P. Chazette, C. Denjean, K. Desboeufs, J. Doussin, N. Elguindi, S.
724 Feuerstein, M. Gaetani, C. Giorio, D. Klopper, M.D. Mallet, P. Nabat, A. Monod, F. Solmon, A. Namwoonde,
725 C. Chikwililwa, R. Mushi, E.J. Welton, and B. Holben, The Aerosols, Radiation and Clouds in Southern Africa
726 Field Campaign in Namibia: Overview, Illustrative Observations, and Way Forward, *Bull. Amer. Meteor. Soc.*,
727 100, 1277–1298, <https://doi.org/10.1175/BAMS-D-17-0278.1>, 2019.

728 Gao, Y., Xu, G., Zhan, J., Zhang, J., Li, W., Lin, Q., Chen, L., and Lin, H., Spatial and particle size distributions
729 of atmospheric dissolvable iron in aerosols and its input to the Southern Ocean and coastal East Antarctica,
730 *J. Geophys. Res.*, 118, 12,634–12,648, doi:10.1002/2013JD020367, 2013.

731 Gili, S., Vanderstraeten, A., Chaput, A., King, J., Gaiero, D. M., Delmonte, B., Vallelonga, Paola Formenti,
732 Claudia Di-Biagio, Mathieu Cazaunau, Edouard Pangui, Jean-Francois Doussin, Mattielli, N., South African
733 dust contribution to the high southern latitudes and East Antarctica during interglacial stages, *Communications*
734 *Earth & Environment*, 3, 129, <https://doi.org/10.1038/s43247-022-00464-z>, 2022.

735 Ginoux, P., Prospero, J.M., Gill, T.E., Hsu, N.C., Zhao, M.: Global-scale attribution of anthropogenic and natural
736 dust sources and their emission rates based on MODIS Deep Blue aerosols products, *Rev. Geophys.*, 50,
737 RG3005, doi:10.1029/2012RG000388, 2012.

738 Giorio, C., Doussin, J.F., D'Anna, B., Mas, S., Filippi, D., Denjean, C., Mallet, M.D., Bourriane, T., Burnet, F.,
739 Cazaunau, M., Chikwililwa, C., Desboeufs, K., Feron, A., Michoud, V., Namwoonde, A., Andreae, M.O., Piketh,

740 S.J. and Formenti, P.: Butene emissions from coastal ecosystems may contribute to new particle formation,
741 *Geophys. Res. Lett.*, 49, <https://doi.org/10.1029/2022GL098770>, 2022.

742 Grini, A., Tulet, P., and Gomes, L.: Dusty weather forecasts using the MesoNH mesoscale atmospheric model.
743 *J. Geophys. Res.*, 111, D19205, <https://doi.org/10.1029/2005JD007007>, 2006.

744 Hamilton, D. S., Perron, M.M.G., Bond, T.C., Bowie, A.R., Buchholz, R.R., Guieu, C., Ito, A., Maenhaut, W.,
745 Myriokefalitakis, S., Olgun, N., Rathod, S.D., Schepanski, K., Tagliabue, A., Wagner, R. and Mahowald, N.M.:
746 Earth, wind, fire, and pollution: Aerosol nutrient sources and impacts on ocean biogeochemistry, *Annual re-*
747 *view of Marine Science*, 14, pp. 303- 330, <https://doi.org/10.1146/annurev-marine-031921-013612>, 2021.

748 Heike, K and J. Volkel, Soil clay minerals in Namibia and their significance for the terrestrial and marine past
749 global change, *African Study Monographs, Suppl.40*, 2010.

750 Heimbürger, A., Losno, R., and Triquet, S.: Solubility of iron and other trace elements in rainwater collected on
751 the Kerguelen Islands (South Indian Ocean), *Biogeosciences*, 10, 6617–6628, [https://doi.org/10.5194/bg-10-](https://doi.org/10.5194/bg-10-6617-2013)
752 [6617-2013](https://doi.org/10.5194/bg-10-6617-2013), 2013.

753 Hooper, H., Mayewski, P., Marx, S., Henson, S., Potocki, M., Sneed, S., Handley, M., Gasso, S., Fischer, M.,
754 Saunders, K.M., Examining links between dust deposition and phytoplankton response using ice cores. *Aeo-*
755 *lian Res.*, 36, 45-60, <https://doi.org/10.1016/j.aeolia.2018.11.001>, 2019.

756 Ito, A., and Kok, J. F.: Do dust emissions from sparsely vegetated regions dominate atmospheric iron supply to
757 the Southern Ocean?, *J. Geophys. Res.*, 122, 3987-4002, <https://doi.org/10.1002/2016JD025939>, 2017.

758 Ito, A., Y. Ye, C. Baldo, and Z. Shi, Ocean Fertilization by Pyrogenic Aerosol Iron. *npj Climate Atmos. Sci.*, 4,
759 30, doi: 10.1038/s41612-021-00185-8, 2021.

760 Jickells, T., Andersen, K.K., Baker, A., Bergametti, G., Brooks, N., Cao, J., Boyd, P., Duce, R., Hunter, K.,
761 Global iron connections between desert dust, ocean biogeochemistry, and climate, *Science*, 308, 67-71, DOI:
762 10.1126/science.1105959, 2005.

763 Journet, E., Desboeufs, K., Caquineau, S. and Colin, J. L.: Mineralogy as a critical factor of dust iron solubility,
764 *Geophys. Res. Lett.*, 35, <https://doi.org/10.1029/2007GL031589>, 2008.

765 Johansen, A. M., and Key, J. M.: Photoreductive dissolution of ferrihydrite by methanesulfinic acid: Evidence of
766 a direct link between dimethylsulfide and iron-bioavailability, *Geophys. Res. Lett.*, 33, L14818,
767 doi:10.1029/2006GL026010, 2006.

768 Kaplan, J.O., Bigelow, N.H., Prentice, I.C., Harrison, S.P., Bartlein, P.J., Christensen, T.R., Cramer, W., Matve-
769 yeva, N.V., McGuire, A.D., Murray, D.F., Razzhivin, V.Y., Smith, B., Walker, D.A., Anderson, P.M., Andreev,
770 A.A., Brubaker, L.B., Edwards, M.E. and Lozhkin A.V.: Climate change and Arctic ecosystems: 2. Modeling,
771 paleodata-model comparison and future projections, *J. Atmos. Res.*, 108, 8171, doi: 10.1029/2002JD002559,
772 2003.

773 Kanguuehi, K. I., Southern African dust characteristics and potential impacts on the surrounding oceans, PhD
774 Thesis, Stellenbosch University, <http://hdl.handle.net/10019.1/123923>, 2021.

775 Klopfer, D., Formenti, P., Namwoonde, A., Cazaunau, M., Chevaillier, S., Feron, A., Gaimoz, C., Hease, P.,
776 Lahmidi, F., Mirande-Bret, C., Triquet, S., Zeng, Z. And Piketh, S.J.: Chemical composition and source apportionment of atmospheric aerosols on the Namibian Coast, *Atmos. Chem. Phys.*, 20, pp. 15811 – 15833,
777 <https://doi.org/10.5194/acp-20-15811-2020>, 2020.

779 Kok, J. F., Albani, S., Mahowald, N. M., and Ward, D. S.: An improved dust emission model – Part 2: Evaluation
780 in the Community Earth System Model, with implications for the use of dust source functions, *Atmos. Chem. Phys.*, 14, 13043–13061, <https://doi.org/10.5194/acp-14-13043-2014>, 2014.

782 Kok, J.F., Ridley, D.A., Zhou, Q., Miller, R.L., Zhao, C., Heald, C.L., Ward, D.S., Albani, S., Haustein, K.: Smaller
783 desert dust cooling effect estimated from analysis of dust size and abundance. *Nature Geoscience*, 10, 274–
784 278, <https://doi.org/10.1038/ngeo2912>, 2017.

785 Laurent, B., Marticorena, B., Bergametti, G., Chazette, P., Maignan, F. and Schmechtig C.: Simulation of the
786 mineral dust emission frequencies from desert areas of China and Mongolia using an aerodynamic roughness
787 length map derived from POLDER/ADEOS 1 surface products, *J. Geophys. Res.*, 110, D18, doi:
788 10.1029/2004JD005013, 2005.

789 Lide, D. R.: *CRC Handbook of Chemistry and Physics 1991–1992*, CRC Press, Boca Raton, Florida, 1992.

790 Liu, M., Matsui, H., Hamilton, D.S., Lamb, K.D., Rathod, S.D., Schwarz, J.P. and Mahowald, N.M.: The underap-
791 preciated role of anthropogenic sources in atmospheric soluble iron flux to the Southern Ocean, *Climate At-*
792 *mos. Sci.*, 5, 28, <https://doi.org/10.1038/s41612-022-00250-w>, 2022.

793 Longo, A., F. Y. Feng, B., W. M. Landing, R.U. Shelley, A. Nenes, N. Mihalopoulos, K. Violaki, E. D. Ingall.
794 Influence of Atmospheric Processes on the Solubility and Composition of Iron in Saharan Dust, *Environ. Sci.*
795 *Tech.*, 50, 13: 6912-20. <https://doi.org/10.1021/acs.est.6b02605>, 2016.

796 Mahowald, N., Luo, C., del Corral, J., Zender, C.S.: Interannual variability in atmospheric mineral aerosols from
797 a 22-year model simulation and observational data, *J. Geophys. Res.*, 108 (D12),
798 <https://doi.org/10.1029/2002JD002821>, 2003.

799 Marcotte, A.R., Anbar, A.D., Majestic, B.J., Herckes, P.: Mineral dust and iron solubility: Effects of composition,
800 particle size, and surface area, *Atmosphere*, 11, 533, doi:10.3390/atmos11050533, 2020.

801 Marticorena, B. and Bergametti, G.: Modelling the atmospheric dust cycle: 1. Design of a soil-derived dust
802 emission scheme, *J. Geochem. Res.*, 16415-16430, 1995.

803 Marticorena, B., Chazette, P., Bergametti, G., Dulac, F., Legrand, M.: Mapping the aerodynamic roughness
804 length of desert surfaces from the POLDER/ADEOS bi-directional reflectance product, *Int. J. Remote Sens.*,
805 25, 603– 626, 2004.

806 Myriokefalitakis, S., Tsigaridis, K., Mihalopoulos, N., Sciare, J., Nenes, A., Kawamura, K., Segers, A., and
807 Kanakidou, M.: In-cloud oxalate formation in the global troposphere: a 3-D modeling study, *Atmos. Chem.*
808 *Phys.*, 11, 5761-5782, doi:10.5194/acp-11-5761-2011, 2011.

809 Paris, R., Desboeufs, K. V., Formenti, P., Nava, S., and Chou, C.: Chemical characterisation of iron in dust and
810 biomass burning aerosols during AMMA-SOP0/DABEX: implication for iron solubility, *Atmos. Chem. Phys.*,
811 10, 4273–4282, <https://doi.org/10.5194/acp-10-4273-2010>, 2010.

812 Paris, R., K.V. Desboeufs, et E. Journet. Variability of dust iron solubility in atmospheric waters: Investigation of
813 the role of oxalate organic complexation, *Atmos. Environ.*, 45, 6510-17. <https://doi.org/10.1016/j.atmosenv.2011.08.068>, 2011.

815 Paris, R., and K. V. Desboeufs, Effect of atmospheric organic complexation on iron-bearing dust solubility. *At-*
816 *mos. Chem. Phys.*, 13, 4895-4905, <https://doi.org/10.5194/acp-13-4895-2013>, 2013.

817 [Perron, M. M. G., Strzelec, M., Gault-Ringold, M., Proemse, B. C., Boyd, P. W., and Bowie, A. R.: Assessment](https://doi.org/10.1016/j.talanta.2019.120377)
818 [of leaching protocols to determine the solubility of trace metals in aerosols, *Talanta*, 208, 120377,](https://doi.org/10.1016/j.talanta.2019.120377)
819 <https://doi.org/10.1016/j.talanta.2019.120377>, 2020.

820 Prospero, J.M., Ginoux, P., Torres, O., Nicholson S.E. and Gill, T.M.: Environmental characterization of global
821 sources of atmospheric soil dust identified with the Nimbus 7 total ozone mapping spectrometer (TOMS) ab-
822 sorbing aerosol product. *Reviews of Geophysics*, 40 (1): 1002, <https://doi.org/10.1029/2000RG000095>, 2002.

823 Redemann, J., Wood, R., Zuidema, P., Doherty, S. J., Luna, B., LeBlanc, S. E., Diamond, M. S., Shinozuka, Y.,
824 Chang, I. Y., Ueyama, R., Pfister, L., Ryoo, J.-M., Dobracki, A. N., da Silva, A. M., Longo, K. M., Kacenelen-
825 bogen, M. S., Flynn, C. J., Pistone, K., Knox, N. M., Piketh, S. J., Haywood, J. M., Formenti, P., Mallet, M.,
826 Stier, P., Ackerman, A. S., Bauer, S. E., Fridlind, A. M., Carmichael, G. R., Saide, P. E., Ferrada, G. A., Howell,
827 S. G., Freitag, S., Cairns, B., Holben, B. N., Knobelspiesse, K. D., Tanelli, S., L'Ecuyer, T. S., Dzambo, A. M.,
828 Sy, O. O., McFarquhar, G. M., Poellot, M. R., Gupta, S., O'Brien, J. R., Nenes, A., Kacarab, M., Wong, J. P.
829 S., Small-Griswold, J. D., Thornhill, K. L., Noone, D., Podolske, J. R., Schmidt, K. S., Pilewskie, P., Chen, H.,
830 Cochrane, S. P., Sedlacek, A. J., Lang, T. J., Stith, E., Segal-Rozenhaimer, M., Ferrare, R. A., Burton, S. P.,
831 Hostetler, C. A., Diner, D. J., Seidel, F. C., Platnick, S. E., Myers, J. S., Meyer, K. G., Spangenberg, D. A.,
832 Maring, H., and Gao, L.: An overview of the ORACLES (ObseRvations of Aerosols above CLouds and their
833 intEractionS) project: aerosol–cloud–radiation interactions in the southeast Atlantic basin, *Atmos. Chem.*
834 *Phys.*, 21, 1507–1563, <https://doi.org/10.5194/acp-21-1507-2021>, 2021.

835 Reichholf, J. H., Is Saharan Dust a Major Source of Nutrients for the Amazonian Rain Forest?, *Studies on*
836 *Neotropical Fauna and Environment*, 21:4, 251-255, DOI: 10.1080/01650528609360710, 1986.

837 Rodríguez, S., Prospero, J.M., Lopez-Darias, J., Garcia-Alvarez, M.I., Zuidema, P., Nava, S., Lucarelli, F., Gas-
838 ton, C.J., Galindo, L., Sosa, E.: Tracking the changes of iron solubility and air pollutants traces as African dust
839 transits the Atlantic in the Saharan dust outbreaks. *Atmos. Environ.*, 246, 118092, <https://doi.org/10.1016/j.atmosenv.2020.118092>, 2021.

841 [Rodríguez, S., Riera, R., Fonteneau, A., Alonso-Pérez, S., and López-Darias, J.: African desert dust influences](https://doi.org/10.1016/j.atmosenv.2023.120022)
842 [migrations and fisheries of the Atlantic skipjack-tuna, *Atmospheric Environment*, 312, 120022,](https://doi.org/10.1016/j.atmosenv.2023.120022)
843 <https://doi.org/10.1016/j.atmosenv.2023.120022>, 2023.

844 Shelley, R.U., Morton, P.L., Landing, W.M., Elemental ratios and enrichment factors in aerosols from the US-
845 GEOTRACES North Atlantic transects. *Deep-Sea Res. II*, 2014.

846 Shelley, R. U., Landing, W. M., Ussher, S. J., Planquette, H., & Sarthou, G. Regional trends in the fractional
847 solubility of Fe and other metals from North Atlantic aerosols (GEOTRACES cruises GA01 and GA03) follow-
848 ing a two-stage leach, *Biogeosciences*, 15(8), 2271–2288. <https://doi.org/10.5194/bg-15-2271-2018>, 2018.

849 Sholkovitz, E.R., Sedwick, P.N., Church, T.M. Influence of anthropogenic combustion emissions on the deposi-
850 tion of soluble aerosol iron to the ocean: empirical estimates for island sites in the North Atlantic, *Geochim.*
851 *Cosmochim. Acta*, 73, 3981–4003, <http://dx.doi.org/10.1016/j.gca.2009.04.029>, 2009.

852 Siefert, R. L., S. O. Pehkonen, Y. Erel, and M. R. Hoffman, Iron photochemistry of aqueous suspensions of
853 ambient aerosol with added organic acids, *Geochim. Cosmochim. Acta*, 58, 3271–3279, 1994.

854 Spirig, R., Vogt, R., Larsen, J. A., Feigenwinter, C., Wicki, A., Franceschi, J., Parlow, E., Adler, B., Kalthoff, N.,
855 Cermak, J., Andersen, H., Fuchs, J., Bott, A., Hacker, M., Wagner, N., Maggs-Kölling, G., Wassenaar, T. and
856 Seely, M.: Probing the fog life-cycles in the Namib desert, *Bull. Am. Met. Soc.*, 100, 2491-2508,
857 doi:10.1175/bams-d-18-0142.1, 2019.

858 Swan, H.B., and J. P. Ivey, Elevated particulate methanesulfonate, oxalate and iron over Sydney Harbour in the
859 austral summer of 2019-20 during unprecedented bushfire activity, *Atmos. Environ.*, 226, 118739,
860 <https://doi.org/10.1016/j.atmosenv.2021.118739>, 2021.

861 Swap, R., Garstang, M., Macko, S.A., et al., The long-range transport of southern African aerosols to the tropical
862 South Atlantic. *J. Geophys. Res.*, 101, 23777–23791, <https://doi.org/10.1029/95jd01049>, 1996.

863 Takahashi, Y., Higashi, M., Fukurawa, T., Mitsunobu, S.: Change of iron species and iron solubility in Asian
864 dust during the long-range transport from western China to Japan, *Atmos. Chem. Phys*, 11, 11237-11252,
865 doi:10.5194/acp-11-11237-2011, 2011.

866 Tao, Y., Murphy, J.G.: The mechanisms responsible for the interactions among oxalate, pH, and Fe dissolution
867 in PM_{2.5}. *Earth and Space Chemistry*, 3, 2259-2265, <https://doi.org/10.1021/acsearthspacechem.9b00172>,
868 2019.

869 Tegen, I., Schepanski, K.: The Global distribution of Mineral Dust, *IOP Conference Series: Earth and Environ-*
870 *mental Sciences*, 7, 012001, doi:10.1088/1755-1307/7/1/012001, 2009.

871 Tyson, P. D. and Preston-Whyte, R. A.: *The Weather and Climate of Southern Africa*, 2nd ed., Oxford University
872 Press Southern Africa, Cape Town, 2014.

873 Ventura, A., Simões, E.F.C., Almeida, A.S., Martins, R., Duarte, A.C., Loureiro, S., Duarte, R.M.B.O., Deposition
874 of aerosols onto upper ocean and their impacts on marine biota, *Atmosphere*, 12, 684,
875 <https://doi.org/10.3390/atmos12060684>, 2021.

876 Vickery, K.J., Eckardt, F.D.: Dust emission controls on the lower Kuseb River valley, Central Namib, *Aeolian*
877 *Res.*, 10, 125-133, <https://doi.org/10.1016/j.aeolia.2013.02.006>, 2013.

878 von Holdt, J.R., Eckardt, F.D., Wiggs, G.F.S.: Landsat identifies aeolian dust emission dynamics at the landform
879 scale, *Remote Sensing Environ.*, 198, 229-243, <https://doi.org/10.1016/j.rse.2017.06.010>, 2017.

880 Wozniak, A. S.; Shelley, R. U.; Sleighter, R. L.; Abdulla, H. A. N.; Morton, P. L.; Landing, W. M.; Hatcher, P. G.
881 Relationships among aerosol water soluble organic matter, iron and aluminium in European, North African,

882 and Marine air masses from the 2010 US GEOTRACES cruise, *Mar. Chem.*, 154, 24– 33 DOI: 10.1016/j.mar-
883 chem.2013.04.011, 2013.

884 Wozniak, A. S., R.U. Shelley, S.D. McElhenie, W.M. Landing, P. G. Hatcher. Aerosol water soluble organic
885 matter characteristics over the North Atlantic Ocean: Implications for iron-binding ligands and iron solubility.
886 SCOR WG 139: Organic Ligands – A Key Control on Trace Metal Biogeochemistry in the Ocean 173: 162-72.
887 <https://doi.org/10.1016/j.marchem.2014.11.002>, 2015.

888 Zhuang, G., Z. Yi, R. A. Duce, and Brown, P.R.: Link between iron and sulphur cycles suggested by detection
889 of Fe (II) in remote marine aerosols, *Nature*, 355, pp. 537–539, <https://doi.org/10.1038/355537a0>, 1992.

890

891