



# 1 **Modelling water quantity and quality for integrated water** 2 **cycle management with the WSIMOD software**

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## 7 **Abstract**

8 Problems of water system integration occur when a model's boundaries are too narrow to capture interactions and  
9 feedbacks across the water cycle. We propose that integrated water systems models are required to overcome  
10 them, and are necessary to understand emergent system behaviour, to expand model boundaries, to evaluate  
11 interventions, and to ensure simulations reflect stakeholder goals. We present the Water Systems Integrated  
12 Modelling Framework (WSIMOD) software as one such approach and describe its theoretical basis, covering the  
13 node and arc nature of simulations, the integration framework that enables communication between model  
14 elements, and the model orchestration to customise interactions. We highlight data requirements for creating such  
15 a model and the potential for future development and refinement. WSIMOD offers a flexible and powerful  
16 approach to represent water systems, and we hope it will encourage further research and application into using  
17 model integration towards achieving sustainable and resilient water management.

## 18 **Plain Language Summary**

19 Water management is challenging when models don't capture the entire water cycle. We propose using integrated  
20 models facilitates management and improves understanding. Thus, we introduce a software tool designed for this  
21 task. We discuss its foundation, how it simulates water system components and their interactions, and its  
22 customization. We provide a flexible way to represent water systems, and we hope it will inspire more research  
23 and practical applications for sustainable water management.

## 24 **1 Introduction**

25 Water fluxes and their pollution concentration are influenced by the interactions and management of all  
26 components that make up the human-altered water cycle, including but not limited to upstream rivers, reservoirs,  
27 freshwater treatment, distribution networks, residential and non-residential water consumers, foul sewers, urban  
28 drainage, storm sewers, wastewater treatment works, groundwater, agriculture, and hydrological catchments (and  
29 the many physical and operational processes within each component). The importance of this interconnectedness  
30 is most evident in rivers, which reflect the overall condition of the catchment system since they aggregate  
31 behaviour over such large areas (Dobson and Mijic 2020; Kirchner 2009). River catchments are rarely dominated  
32 by the behaviour of any specific component or any individual stakeholder's decisions. For example, 60% of  
33 English catchments that do not achieve a 'good' status in the Water Framework Directive (WFD) do so because



34 of multiple different pollution sources, with wastewater infrastructure and agriculture being the most prevalent  
35 drivers, each affecting 50% of catchments. (Environment Agency 2020b). The implication is that a modelled  
36 representation focussing on any individual component is unlikely to give accurate estimates of impacts beyond  
37 that component (Beven 2007; Blair et al. 2019; Dobson, Wagener, and Pianosi 2019; Schmitt and Huber 2006).  
38 Furthermore, estimates within a subsystem representation may be inaccurate if sensible boundary conditions  
39 cannot be defined, something that water managers are highly sensitive to (Höllermann and Evers 2017). We term  
40 these problems of water systems integration, and it follows that understanding the water cycle as a whole is needed  
41 to address them. Models that take such an approach will better capture component boundaries and the wider  
42 impacts of stakeholder decisions, ultimately enabling more accurate representations of water quality in rivers,  
43 which is essential to effectively manage, for example, water supply (Mortazavi-Naeini et al. 2019) and  
44 biodiversity (Dobson, Barry, et al. 2022).

45 In this paper we introduce the theoretical underpinning behind a novel method for modelling integrated water  
46 systems to address these challenges. Firstly, the need for integrated water cycle simulation models is explained,  
47 including their current coverage. The importance of parsimonious representations within an integrated model is  
48 then discussed, along with the methods used to achieve integration.

49 The environmental modelling research community has responded to problems of water systems integration  
50 primarily through computer simulation models (Bach et al. 2014; Best et al. 2011; Douglas-Mankin, Srinivasan,  
51 and Arnold 2010; Rauch et al. 2017; Tscheikner-Gratl et al. 2019; Whitehead, Wilson, and Butterfield 1998). We  
52 distinguish an integrated water system modelling approach from a system dynamics approach (see Zomorodian  
53 et al., (2018)) by further specifying that component representations must have a physical basis, which is needed  
54 to link observational data to model behaviour and to capture interventions (e.g., new infrastructure or changes to  
55 operations). We define integrated water system models as those which link component representations to capture  
56 and understand the complex interactions and feedbacks that occur between components. We categorise the four  
57 key goals of these models: (1) to understand which fundamental processes drive emergent behaviour at a whole-  
58 water system scale; (2) to avoid simulation inaccuracies caused by narrow boundary conditions; (3) to test  
59 interventions to the physical system or operational behaviour in order to understand their water cycle wide impacts  
60 or interactions and (4) to capture impacts that align more closely with desired water system outcomes, in addition  
61 to performance indicators of individual components. For example, in-river pollutant concentration is a better  
62 indicator of wastewater system performance than the more typically monitored number of sewer spills (Giakoumis  
63 and Voulvoulis 2023).

64 In addressing problems of water systems integration, existing modelling approaches have made significant  
65 progress. Bach et al. (2014) set out a comprehensive typology for integrated urban water systems modelling.  
66 However, among the reviewed models, only CityDrain3 (Burger et al. 2016), WEST (Vanhooren et al. 2003), and  
67 SIMBA (IFAK 2007) can represent receiving water bodies (i.e., rivers), which is where the importance of an  
68 integrated representation is most pronounced. Furthermore, due to the urban focus of these models, the ability to  
69 simulate pollution concentrations in receiving waters impacted by upstream catchments is highly limited yet is  
70 central to quantifying in-river impacts (Liu, Dobson, and Mijic 2022). A more recent effort to characterize  
71 integrated water systems modelling places importance on in-river conditions (Tscheikner-Gratl et al. 2019) and  
72 present a comprehensive review of urban and rural water cycles and their impacts on rivers. However, the  
73 reviewed modelling approaches omit some key factors: the importance of water resources infrastructure, which



74 play a significant role in concentrating pollution during low flows if abstractions take place; the relevance of  
75 groundwater, which provides baseflow to dilute pollution during critical low flow periods; and consideration of  
76 agricultural processes and associated pollution that results from them, which is a critical source of water pollution  
77 worldwide (Mateo-Sagasta et al. 2017; Tang et al. 2021), and the second most common catchment pollution source  
78 in England (Environment Agency 2020b). Integrated models that capture groundwater and agricultural processes  
79 are present in the modelling literature, such as INCA (Whitehead et al. 1998) and HYPE (Lindström et al. 2010),  
80 however, in contrast, these are limited by their ability to capture urban systems. Thus, while water systems  
81 integration is well-served from a rural or urban modelling perspective, we identify that there is not yet an approach  
82 that offers a self-contained representation to capture all key processes required to model in-river water quality at  
83 a whole-water cycle scale.

84 A further critical factor in creating an integrated water systems model is how components are represented. In  
85 general, current approaches have favoured identifying pre-existing detailed component representations which are  
86 then integrated (Schmitt and Huber 2006). For example, DAnCE4Water (Rauch et al. 2017), which is the most  
87 comprehensive application to date, includes high resolution and sophisticated models for a wide range of urban  
88 components. However, as more and more components are captured by integrated modelling, it becomes  
89 increasingly difficult parameterising such detailed models. Simply combining separately calibrated models  
90 provides no guarantee of performance as a whole (Lee 1973). Meanwhile, integrated models typically have many  
91 parameters that may compensate each other, thus making calibration a challenging and risky process (Voinov and  
92 Shugart 2013). An alternative approach is to forego calibration altogether by adopting parsimonious models with  
93 fewer parameters and ideally deriving those parameters from best available data (Dobson et al. 2021). Although  
94 complicated modelling approaches are needed for tasks such as design, these approaches are also more difficult  
95 to apply widely and thus may hinder the goals of integrated water systems modelling. For example, building  
96 scientific understanding requires repeated testing of an approach in various locations, and customising the model  
97 to match local conditions is essential when representing interventions. Therefore, a modelling approach that can  
98 be easily deployed on a wider scale is of significant benefit to problems of water systems integration.

99 To ensure that a range of water system configurations can be accommodated, flexibility or customisability must  
100 be incorporated into the approach for integration. Integration approaches vary broadly between tightly coupled  
101 and loosely coupled. In a tightly coupled approach equations and interactions are pre-defined to create a self-  
102 contained integrated representation, such as with JULES (Best et al. 2011) or INCA (Whitehead et al. 1998).  
103 While, in a loosely coupled approach, component representations are self-contained, and the integration occurs  
104 by facilitating their interactions, filling the role as a message passing interface, such as with OpenMI (Harpham,  
105 Hughes, and Moore 2019) and DAnCE4Water (Rauch et al. 2017). Belete et al. (2017) describe the arrangement  
106 of components and their interactions as integrated model orchestration, highlighting that different orchestrations  
107 are suitable for different applications. While looser coupling provides greater control over orchestration, and thus  
108 greater ability to customise and capture a wide variety of systems, it also creates a higher user burden to set up  
109 and understand many subsystems, considered to be a key barrier to the uptake of such approaches (Zomorodian  
110 et al. 2018). Conversely, a tightly coupled model that represents the same components as a loosely coupled one  
111 may be easier to set up but typically offers less control over orchestration. In the middle ground is an integrated  
112 representation that gives flexibility around orchestration but comes with self-contained components that do not  
113 need to be onerously setup by a user, such as the CityDrain3 software for modelling urban drainage systems



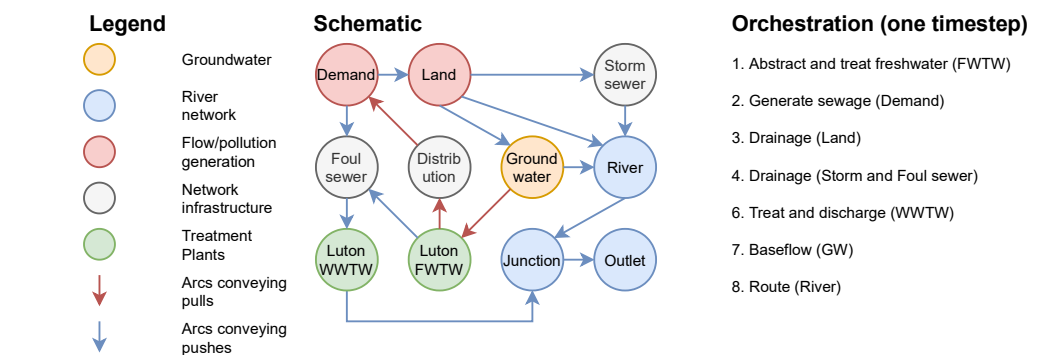
114 (Burger et al. 2016). We propose that this middle ground is the most beneficial for a modeller and believe that  
115 such an approach to integration is the most productive avenue towards creating highly flexible, user-friendly  
116 models of the integrated water cycle.

117 The concepts introduced above suggest that, for many problems of water systems integration, capturing a broad  
118 representation of the water cycle and interactions between its components is equally important as detailed  
119 component representations. We have created a tool to implement this modelling philosophy, the Water Systems  
120 Integrated Modelling framework (WSIMOD), which is an open-source Python package for flexible and  
121 customisable simulations of the water cycle that treats the physical components of the water cycle as nodes  
122 connected by arcs that convey water and pollutant flux between them. The software source code and online  
123 tutorials are published by Dobson, Liu, and Mijic (2023a), in contrast, this paper presents WSIMOD's theoretical  
124 underpinning with a discussion on model setup and of integrated water system modelling in general. To address  
125 the difficulties in application associated with integrated modelling mentioned above, WSIMOD contains a library  
126 of built-in component representations covering a more complete water cycle coverage than any identified  
127 integrated models, and a default but customisable orchestration adjudged to be suitable for many catchments and  
128 regional water systems coordination. Where possible these representations are based on parsimonious and peer-  
129 reviewed models. Extensive model documentation with worked examples is provided online (Dobson, Liu, and  
130 Mijic 2023b), enabling users to gain confidence and become familiar with using WSIMOD.

## 131 **2 WSIMOD**

132 WSIMOD is an integrated modelling framework that provides ready-to-use objects (nodes, arcs, water stores, and  
133 model orchestration) that are suitable for a wide range of water systems and described in greater detail in the  
134 following sections. However, WSIMOD is not intended to be a one-size-fits-all solution, indeed, the ubiquity of  
135 non-textbook water systems led us to create a more customisable modelling approach in the first place. This paper  
136 describes the theory behind WSIMOD in general and user-friendly terms, avoiding the use of equations and  
137 technical details, while further documentation can be found online (Dobson, Liu, and Mijic 2023c). The WSIMOD  
138 framework is implemented in Python 3, which is widely practiced in the environmental modelling community and  
139 facilitates quick setup and easy customisation. WSIMOD is the combined effort of many studies conducted as  
140 part of the CAMELLIA (Community Management for a Liveable London) project  
141 (<https://www.camelliawater.org/>), which are linked to relevant sections of the model description to highlight the  
142 range of possible applications.

143 An example WSIMOD model is shown in Figure 1, demonstrated for Luton, UK, selected for illustration because  
144 its water cycle is reasonably self-contained. WSIMOD uses object-oriented programming (OOP), which classifies  
145 components by common attributes and behaviours (classes), thus facilitating customisation or the introduction of  
146 new behaviours. All objects in WSIMOD are a subclass of WSIObj, which predefines efficient arithmetic  
147 operations for water quality and volume, however users will typically instead interact with the subclasses  
148 described in the following sections. Additionally, users may customise a model's high-level control over how  
149 interactions take place within a timestep, or the model's orchestration, which is a unique feature of WSIMOD  
150 (Belete et al. 2017). Thus, while Figure 1 depicts one possible arrangement and selection of nodes and arcs, a  
151 wide variety of water systems can be represented.



152

153 **Figure 1: An example WSIMOD model for Luton, UK. Orchestration is shown demonstrating the high-level functions**  
154 **called for each timestep. Nodes are shown as circles and arcs as arrows. WWTW stands for Wastewater Treatment**  
155 **Works and FWTW stands for Freshwater Treatment Works.**

## 156 2.1 Nodes represent water cycle components

157 Physical representations of the different components in the water cycle are typically implemented as WSIMOD  
158 nodes, Figure 1. In the software implementation, all nodes are instances of the Node class or its subclasses.  
159 Formulation of components as nodes using OOP draws heavily on the CityDrain3 software (Burger et al. 2016).  
160 Our generic definition allows nodes to represent diverse entities, for example, a collection of manholes  
161 representing a region of sewer network or individual manholes that can be connected to represent a sewer network;  
162 as demonstrated in (Dobson, Watson-Hill, et al. 2022). In this section we describe the Node class, summarise the  
163 existing node subclasses currently implemented in WSIMOD, and describe how to customise them.

### 164 2.1.1 The Node class

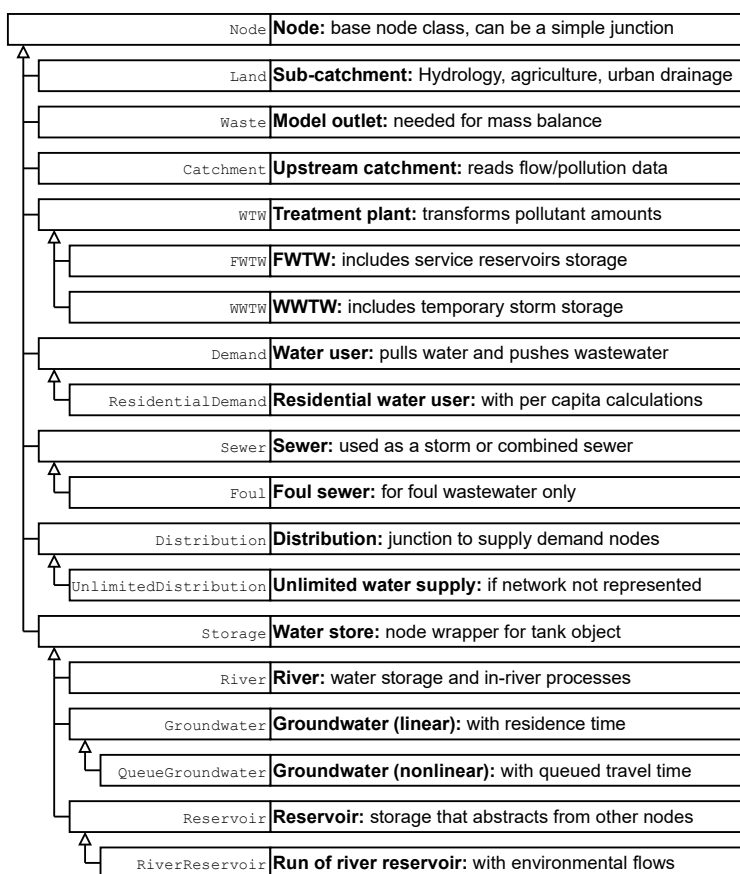
165 The Node class in WSIMOD is a generic class that is expected to be the parent of any component represented.  
166 The class captures three key behaviours. Firstly, it predefines defaults for much of the interaction functionality  
167 later described in Section 2.2. Secondly, it contains a variety of useful functions to interact with other nodes via  
168 arcs (see Section 2.2.1). Thirdly, it enables mass balance checking (see Section 2.3.2). For these reasons, physical  
169 components should generally be implemented in WSIMOD as a subclass or variation of a Node, even if the  
170 computational implementation is simply a wrapper for another, pre-existing, model. Although the base Node class  
171 does not implement any physical processes, it may serve as a junction for branches or convergences in the water  
172 system, such as river bifurcations and confluences.

### 173 2.1.2 Node subclasses

174 To ensure WSIMOD is as easy to be implemented as possible, a variety of water cycle components have been  
175 developed in Python as Node subclasses. We summarise these components in Figure 2, and recommend viewing  
176 the online documentation for full details, which provides a library of documented components as well as tutorials  
177 on their use (Dobson et al. 2023c). As anticipated in the introduction, these components are designed for  
178 parsimony and can be instantiated with as few parameters as possible, thus minimising data requirements and



179 maximising utility. We note that this summary is up to date as of time of writing, however the Imperial College  
 180 London Water Systems Integration research group will be continually upgrading and adding new functionality to  
 181 WSIMOD.  
 182



183  
 184 **Figure 2: An inheritance diagram of components implemented as nodes in WSIMOD, arrows indicate that a node is a**  
 185 **subclass of another node. The courier text is the name of the node in WSIMOD.**

186 Many node types include water stores, which are sufficiently prevalent in water systems to warrant their own class  
 187 in WSIMOD, referred to as a Tank, with further details provided in A1.1 Tank object to generalise water stores.  
 188 It's important to note that a tank is not a node, but a node may have a tank. The number of tanks a node subclass  
 189 may have varies depending on what it represents, with some having none (e.g., Demand), one (e.g., Reservoir),  
 190 or multiple tanks (e.g., Land). If a user requires a tank that can act as a node, they can utilize the Storage class to  
 191 achieve this functionality.

### 192 2.1.3 Customising nodes

193 We do not expect that the provided WSIMOD components will be sufficient to cover every water system or be  
 194 suitable in cases where detailed representations are necessary. Where possible, users should consider customising

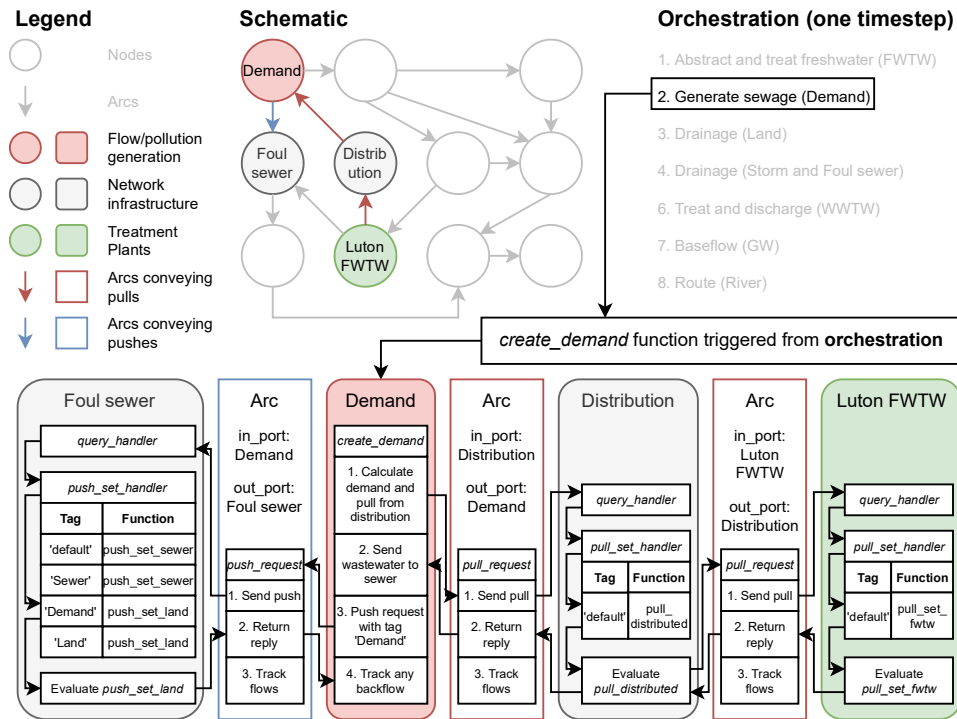


195 model orchestration (Section 2.3) or component interactions (Section 2.2.3) to represent these cases. However,  
196 new physical subsystems may also be represented based on those defined in Figure 2. The variety of techniques  
197 to customise subclass behaviour in OOP are outside the scope of this paper but have been discussed extensively  
198 elsewhere (Gamma et al. 1995). In general, these techniques aim to avoid duplication of effort, because most  
199 functionality will be predefined in an existing class that can maintain specific interactions between existing  
200 subclasses. We enthusiastically encourage users to contribute to the open-source package to create a better  
201 software for the wider environmental modelling community, though request that contributors our read online  
202 guidelines before doing so to avoid highly duplicative and bloated components that are likely to confuse others.  
203 If a pre-existing model aims to be integrated and it would be too significant a programming effort to reimplement  
204 under this philosophy, then we would request that wrappers are used to treat the model as a node and interface  
205 with the existing software.

206 To highlight the flexibility offered by WSIMOD, we briefly discuss some examples of complex behaviour being  
207 captured through node customisation. In Dobson et al., (2021), demand nodes were assigned time varying  
208 population, water use behaviour, and pollutant generation to capture changing commuter patterns in London that  
209 resulted from the COVID-19 pandemic. In Liu et al., (2023a), the hydrological processes in land nodes were  
210 customised in a variety of ways to represent nature-based solutions. We note that these examples are provided  
211 with open-source code but are not included in the WSIMOD repository because they have significant data  
212 requirements to set up that are context specific, and thus are unlikely to be generalisable to a wide range of cases.  
213 In the software documentation we also provide a how-to guide for node customisation (Dobson et al. 2023c).

## 214 **2.2 Integration framework**

215 The WSIMOD integration framework facilitates interactions between nodes by serving as a message passing  
216 interface that transfers information relating to water quality and quantity. It was developed in an application to  
217 London's water cycle at a wastewater catchment scale (Dobson et al. 2021). It draws significantly on the OpenMI  
218 (Harpham et al. 2019) and Open MPI (Graham et al. 2006) interfaces but has been tailored to water systems by  
219 providing a variety of built-in behaviours. The integration framework consists of three key concepts: arcs, pushes  
220 and pulls, and requests and checks (demonstrated in Figure 3 and described in the following sections). Arcs are a  
221 class that facilitate interactions between nodes but can also represent physical entities (e.g., pipes). Arcs convey  
222 both water quality and quantity fluxes, which are discretised and packaged together, based on concepts from  
223 CityDrain3 (Burger et al. 2016). Pushes and pulls differentiate between the directionality of an interaction.  
224 Requests and checks differentiate between information passing that simulates the movement of water (requests)  
225 and that which does not (checks). While we do not recommend changing the integration framework itself, we  
226 provide a generic method to accommodate a wider variety of interactions in Section 2.2.3.



227

228 **Figure 3: An example of the WSIMOD integration framework, illustrated through the automatic behaviour triggered**  
 229 **when the *create\_demand* function is called by a Demand node during orchestration. The node pulls water via an arc**  
 230 **from a distribution network and pushes foul wastewater via an arc to a sewer. The figure further illustrates the use of**  
 231 **handlers and tags to customise interactions between nodes. Italicized text indicates that it is a function. Nodes are shown**  
 232 **as circles or rounded squares, while arcs are shown as coloured arrows or sharp coloured squares.**

233 **2.2.1 The Arc class and fluxes**

234 Arcs are a class to establish connections between nodes. They transmit all message passing and track fluxes when  
 235 requests are made. Arcs can have a capacity property that limits the flow in a given timestep, which can also be  
 236 customised to be dynamically calculated, for example, to implement Manning’s equation along pipes (Dobson,  
 237 Watson-Hill, et al. 2022). Additionally, if multiple arcs are linked to a single node, they can be assigned a  
 238 preference attribute. This enables the node to prioritize certain arcs over others. For example, a sewer node may  
 239 connect to a wastewater treatment plant (WWTW) and to a river via a sewer spill. In this case, the spill arc could  
 240 be assigned a low preference to ensure that it is only utilized if the WWTW cannot accept any water.

241 Fluxes in WSIMOD are described in discrete packages called Volume-Quality Information Packages (VQIP). A  
 242 VQIP is a dictionary that contains entries for volume and all simulated pollutants. Calculations in WSIMOD are  
 243 typically performed on VQIPs rather than simply flows, thus ensuring simulation of both water quantity and  
 244 quality. The core parent class (WSIObj) of all WSIMOD classes provides functions to perform basic operations  
 245 with VQIPs in place of the normal arithmetical operations that would typically be only performed on flows or  
 246 volumes. VQIPs track water quality as mass rather than as concentration to accommodate cases where pollutants





247 are being moved with no associated water quantity, except for non-additive variables such as temperature or pH.  
248 WSIMOD refers to anything tracked in a VQIP that is not water volume as a pollutant, however this should be  
249 interpreted as a water quality constituent, and does not imply that everything simulated (e.g., temperature or  
250 dissolved oxygen) is a pollutant from an environmental/management perspective.

251 WSIMOD can simulate any number of pollutants in a mass balance approach, also referred to as conservatively,  
252 provided their sources into the water cycle can be identified and quantified. However, bio-chemical changes for  
253 pollutant decay can also be represented, see Appendix 1, A1.2 Non-flux pollutant changes. Other more  
254 complicated pollutant transformation can be captured on a case-by-case basis, for example to capture nutrient  
255 cycling in the soil pool, using on equations from Liu et al., (2022).

256 A further consideration commonly required in water systems is that of travel time of water, which requires its own  
257 specific implementation due to the discrete nature of flux and VQIPs in WSIMOD, further details are provided in  
258 A1.3 Travel time of water.

## 259 **2.2.2 Types of interactions: pushes/pulls and requests/checks**

260 In order for a simulation to occur in a non-tightly coupled integrated representation, something must trigger the  
261 interactions that are conveyed via arcs. High-level controls that govern these behaviours are described as model  
262 orchestration (see Section 2.3), however a key benefit to using this integration framework is that the user does not  
263 have to predefine all possible interactions in advance. Because information transmitted by arcs automatically  
264 triggers further information transmission in connected nodes, a user may customise their nodes, arcs, and  
265 orchestration, without onerously updating every possible interaction that may take place in the model, as  
266 visualised in Figure 3.

267 To represent a wide variety of behaviours, WSIMOD categorises interactions based on directionality of intent and  
268 whether they represent flux or not. Directionality of intent refers to either cases when a node has water that must  
269 be sent somewhere, called a push, or when a node needs water from somewhere, called a pull. Interactions between  
270 nodes, whether pushes or pulls, may convey flux of water and pollutants, and simulate the movement of water, or  
271 they may convey non-flux information necessary for achieving realistic simulations. An interaction conveying  
272 flux is referred to as a request, while a non-flux interaction is a check.

273 Pushes occur when a node needs to discharge water. For example, when a wastewater treatment works (WWTW)  
274 must discharge effluent to a river, or a catchment must discharge runoff downstream. In general, push scenarios  
275 are more common in water systems because water travels from upstream to downstream. Meanwhile, pulls occur  
276 when a node requires water. For example, when a farmer abstracts water from a borehole to fill their irrigation  
277 reservoir. Pulls typically represent human-related effort to move water in a non-natural way.

278 A request occurs when a node intends to push or pull a certain amount of water to or from another node, regardless  
279 of the connected node's current state. For example, in the pull case, a demand node will intend to satisfy its entire  
280 water needs by pulling water from the distribution network, and this intention is independent of the availability of  
281 water in the distribution network. For example, in the push case, a demand node will always intend to send the  
282 entire volume of its foul water to a sewer system, even if the sewer cannot accommodate the full amount. All  
283 water flux in WSIMOD is ultimately simulated by requests, however most cases are not as straightforward as the  
284 above examples, and nodes often require additional information about the state of giving or receiving nodes to  
285 calculate their requests. Interactions passing such non-flux information are referred to as checks.



286 Generically a check is any kind of non-flux information passing between nodes, it enables a node to use the state  
 287 of the nodes that it interacts with to calculate its requests. For example, when a freshwater treatment works  
 288 (FWTW) can draw water from multiple viable reservoirs collectively containing more water than needs to be  
 289 treated, a calculation is required before requesting water. The FWTW will send checks to the connected reservoirs  
 290 to determine their available water capacity and calculate the appropriate ratio to satisfy its treatment demand. Only  
 291 then will the FWTW send requests for the required water.

292 **2.2.3 Default and customised interactions**

293 During simulation, a node needs to make responses when it receives a push/pull request/check, which we term as  
 294 a reply. We formulate four types of predefined replies that are widely observed in node interactions in the water  
 295 systems. These default replies allow nodes to interpret the responses of requests/checks sent to the interacting  
 296 nodes without the need to understand their detailed behaviour. These defaults are set out in Table 1 with examples.

297 **Table 1: The default reply from each kind of component interaction with an example.**

	Push		Pull	
	Reply	Example	Reply	Example
Request	Amount not received	A sewer sends a push request to a WWTW. The WWTW calculates the available capacity, updates its state variables to represent the increased throughput, and replies with how much water from the request that could not be received.	Amount sent	A FWTW sends a pull request to a reservoir. The reservoir calculates how much of the request can be met, updates its state variables to decrease the current volume, and replies with the amount of water abstracted.
Check	Maximum volume available to push	A sewer sends two push checks to two downstream sewers (required to calculate what proportion to discharge water to them). The two sewers reply with total amount of water that they can each receive.	Maximum volume available to pull	A FWTW sends two pull checks to two reservoirs (required to determine what proportion to pull water from them). The reservoirs reply with each of their current abstractable volumes.

298 While these default interactions are likely to accommodate much of the information passing required to simulate  
 299 an integrated water cycle, further customisation may be necessary to allow specific nodes to respond differently  
 300 to others. For example, a sewer node might respond differently to a push request from another sewer than from a  
 301 land node because sewer to sewer travel time may be calculated differently than land runoff to sewer travel time.  
 302 Furthermore, the default check behaviour which transmits capacity information may not be sufficient for a  
 303 component to calculate where to push/pull water. For example, due to the importance of head in determining flow,  
 304 the amount of water that a floodplain can discharge would not be based on the receiving river's capacity but  
 305 instead its head (Liu et al. 2023a).

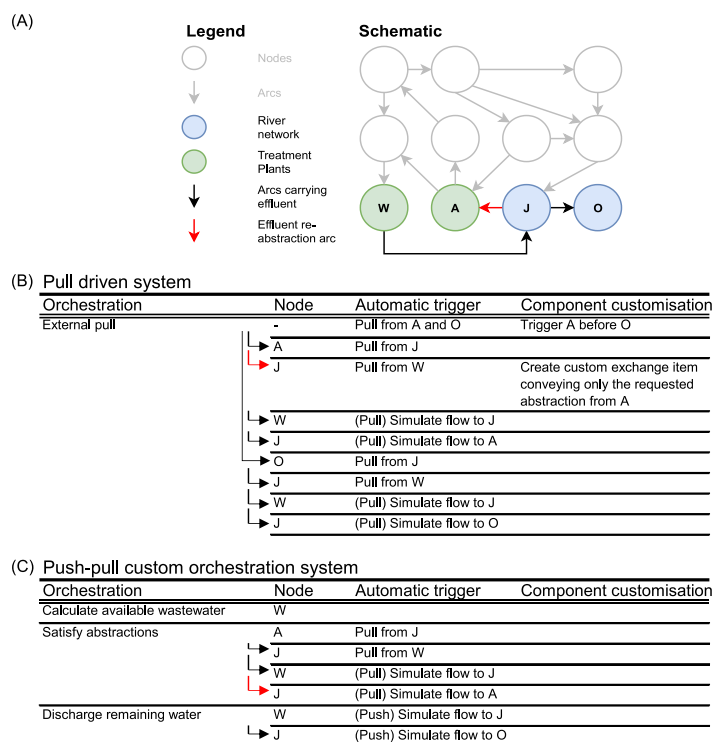


306 To define what a node does in reply to an interaction, all WSIMOD interactions pass through ‘handlers’ that are  
307 associated with a ‘tag’. A handler is a Python dictionary belonging to a node, and the tag is a key to the handler  
308 that determines which function is called during a given interaction (see Figure 3). All nodes must have handlers  
309 which contain a ‘default’ tag, thus enabling any node to interact with any other node. However, additional tags  
310 and replies can be added to enable different behaviours based on the type of node that is interacting with it. An  
311 example interaction for a Demand node draining to a foul Sewer node is given in Figure 3. The generation of  
312 wastewater in the Demand node is triggered by orchestration and sent as a push request, via an arc, to a sewer  
313 node, with a tag ‘Demand’ to indicate how the sewer node should reply. The sewer node has predefined behaviour  
314 for tags ‘default’, ‘Sewer’, ‘Demand’, and ‘Land’, with different functions associated with these tags. The sewer  
315 node uses the handler to identify that it should evaluate the ‘push\_set\_land’ function (which calculates the travel  
316 time through the sewer using the same equation as is used for calculating travel time of runoff arriving from a  
317 land node) and returns the reply via the arc. The base Node predefines simple default handlers that enable it to  
318 convey interactions to connected nodes, for example, a push request to a Node will trigger push requests to  
319 connected nodes, in this way the Node can behave as a simple junction. A how-to guide to explain this behaviour  
320 in greater detail, with examples, is available in the documentation (Dobson et al. 2023c).

### 321 **2.3 Orchestration to manage and enable simulation**

322 As defined by Belete et al. (2017), orchestration is how nodes and their interactions are managed and enabled.  
323 The default steps that we consider to be important to include in orchestration for many water systems are shown  
324 in Figure 1. In integration frameworks such as OpenMI, all interactions in a simulation occur by nodes triggering  
325 pulls to other nodes, orchestrated by a single external pull each timestep (Harpham et al. 2019). Meanwhile, in  
326 WSIMOD, a finer level of control of orchestration is given to a user. We argue that the WSIMOD approach is  
327 better at capturing complex within-timestep behaviours.

328 In Figure 4, building on the system shown in Figure 1, we provide a motivating example of how it is more efficient  
329 to customise the behaviour of water systems by customising orchestration, rather than customising/adding nodes  
330 or arcs, as would be required in other integration frameworks. We consider a common case for an integrated water  
331 system of requiring the simulation to perform downstream re-abstraction of wastewater effluent, visualised by the  
332 red arc in Figure 4. Wastewater re-abstraction (water from W reaching A) requires available wastewater at W to  
333 be matched against demand for re-abstracted wastewater at A, both calculations involving the interconnected  
334 node, J.



335

336 **Figure 4:** (A) an example water system based on the Luton example from Figure 1 that includes re- abstraction of  
 337 treated effluent (red arc). (B) the model steps to achieve re- abstraction of effluent when using a pull-driven integration  
 338 framework, of the kind used in OpenMI, without customisable orchestration, (C) and when using a push-driven  
 339 integration framework, of the kind used in WSIMOD, with customisable orchestration.

340 To capture such an interaction without customising orchestration and using only pulls, as would be required with  
 341 an OpenMI approach (Figure 4 (B)), customisation of J would need to specify whether the pull is originating from  
 342 A or O, and this information would need to be conveyed to W. This is because, if the pull is intended to route the  
 343 system for that timestep (i.e., it is originating from O via J), then W should release all its effluent, while if it is  
 344 intending to satisfy the demand (i.e., it is originating from A via J), it should only release enough to meet A's  
 345 requirements. Additionally, the external pull that initially triggers all interactions must be customised to pull from  
 346 A before O, so that W does not fully route before abstractions can be made.

347 Instead, a user with a broad overview of the water cycle may more easily accommodate such behaviour if given  
 348 high-level control over the model orchestration (Figure 4 (C)). During a timestep the orchestration can specify  
 349 that W first calculates its treated effluent without discharging it into a receiving river, abstractions are triggered  
 350 at A, which draws water from W via J, and then W discharges any remaining effluent into its receiving water.  
 351 This flexibility in orchestration enables representations of a wide variety of water systems while minimizing  
 352 changes to the behaviour of underlying components.



353 **2.3.1 The Model class**

354 A Model class contains all nodes, arcs and forcing data, and provides a default orchestration adjudged to represent  
355 a wide variety of water systems. Whether the WSIMOD default orchestration is used or whether an entirely new  
356 orchestration is defined, we recommend the use of a Model class for a variety of reasons. Firstly, built-in load and  
357 save functionality enables easy sharing and editing of a specific model. Secondly, it contains a run function that  
358 can carry out orchestration, perform mass balance checking (see following section), store simulation results, and  
359 trigger end of timestep functions. Thirdly, it enables easy collection and grouping of nodes to facilitate  
360 orchestration, for example, all WWTWs are referenced by dictionary belonging to the Model class so that they  
361 can easily be triggered during a timestep. A tutorial in use of the Model class is provided in the online  
362 documentation (Dobson et al. 2023c).

363 **2.3.2 Software quality control**

364 Due to the integrated nature of the water systems that WSIMOD simulates, it can be easy to introduce errors that  
365 are difficult to spot. We provide extensive unit testing in line with best software development practice, which  
366 enables ensuring that changes to code do not introduce unintended behaviour changes. However, a further  
367 safeguard against this is a unified method for mass balance error checking. Both nodes and arcs have predefined  
368 lists of functions to calculate the total inflows, total outflows and change in any storage. Any newly defined  
369 behaviour for nodes and arcs must then also consider how mass balance checks will be impacted and thus update  
370 the lists of functions associated with inflows/outflows/change in storage.

371 Because the core functionality of WSIMOD that performs mass balance checking is indifferent to units (see  
372 Section 2.3.3), and because some pollutants exist in far smaller quantities than others, it is possible that both  
373 incredibly small and large numbers may be present in VQIPs in the model. Therefore, mass balance checking  
374 compares at the magnitude of the largest value of inflows/outflows/storage for a given pollutant or water volume  
375 for a given model element in each timestep. Any discrepancy that is larger than a user specified value is reported,  
376 although because of floating point accuracy some discrepancies will be unavoidable. We encourage users to  
377 exercise common sense to not chase down incredibly small discrepancies while seeking to understand larger  
378 discrepancies, which are usually indicative of some implementation error. The user control over orchestration  
379 typically makes debugging WSIMOD models easier than most integrated models, as a user may step through each  
380 set of triggers in the orchestration and recheck mass balances until the error occurs.

381 **2.3.3 Units and timesteps**

382 While the core of WSIMOD is agnostic of units, many nodes are parameterised and assume input data is in SI  
383 units, thus we recommend use of SI units throughout. If extensive work would be required to re-implement  
384 equations in new units, we recommend converting before and after the calculations from/to SI units.

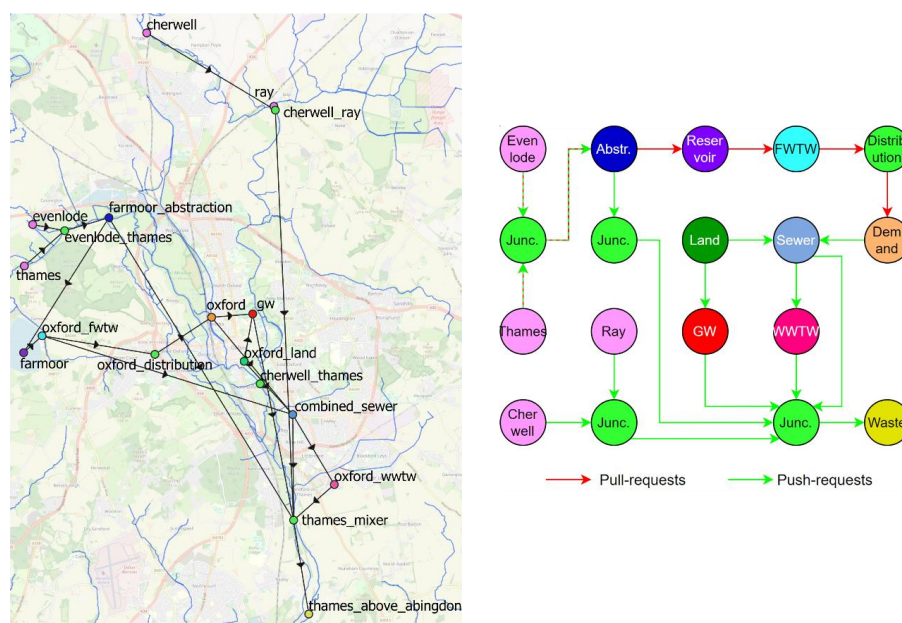
385 Because the timestep size will vary depending on the application of WSIMOD, nodes do not necessarily make  
386 assumptions about timestep, instead requiring the user to define the timestep that is consistent with their  
387 parameters and input data. This enables significant flexibility in representations that can enable studies  
388 investigating the impact of timestep size on simulations (Dobson, Watson-Hill, et al. 2022). We note that the



389 detailed bio-chemical processes used in agricultural surfaces and rivers assume a daily timestep. In addition, these  
390 processes are developed to focus on pollutants associated with crop nutrient cycles.

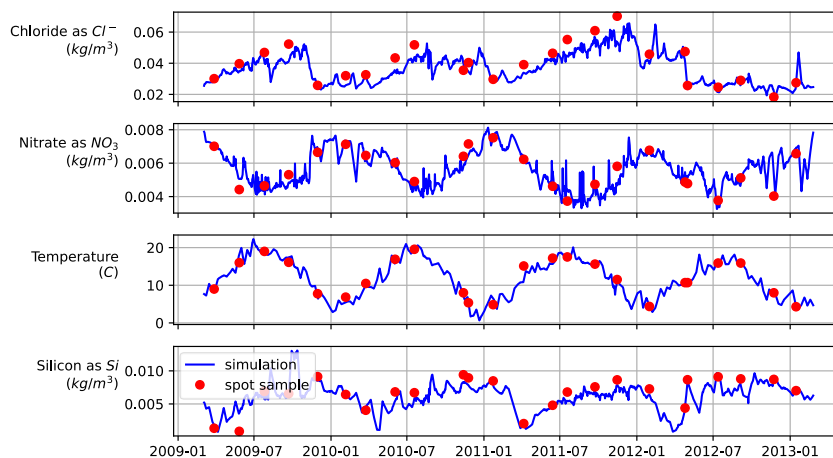
### 391 3 Demonstration

392 Throughout this paper, numerous WSIMOD case studies are referenced. However, to illustrate the scope and  
393 purpose of WSIMOD, we present a demonstration case study. We have chosen a model featured as a tutorial in  
394 the online documentation so that readers will be able to easily reproduce, run and edit it  
395 ([https://barneydobson.github.io/wsi/demo/scripts/oxford\\_demo/](https://barneydobson.github.io/wsi/demo/scripts/oxford_demo/)). The demonstration covers the area of Oxford,  
396 UK, depicted in Figure 5 both as a map and schematic.



397  
398 **Figure 5: (a, left) Map of the model nodes and arcs over the city of Oxford, map base layer © OpenStreetMap**  
399 **contributors, licensed under the Open Database License (ODbL), rivers data from Open Rivers © 2023 Ordnance**  
400 **Survey Limited (b, right) schematic demonstrating the node-arc connectivity of the model.**

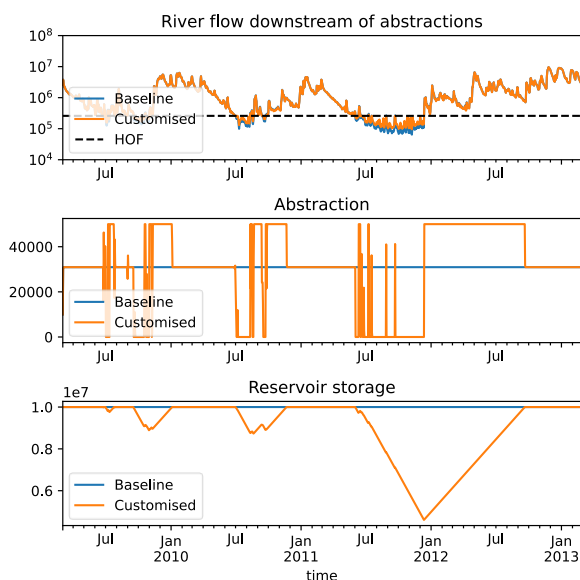
401 The model includes four upstream catchments, a reservoir-based water supply system, and a combined sewage  
402 system, all of which ultimately combine and drain into the River Thames. Because the model is primarily  
403 demonstrative, the parameters given in the model are estimated based on local knowledge. However, high-  
404 resolution (weekly sampling) water quality data are available in the area (Bowes et al. 2018) thus enabling accurate  
405 boundary conditions at upstream catchments and water quality representations in the model. In Figure 6 we plot  
406 simulations against the weekly sampling, water quality indicators to plot were selected based on those that had  
407 data for all locations for the simulation duration.



408

409 **Figure 6: Demonstration of simulated levels of chemical water quality indicators against spot samples.**

410 In addition to WSIMOD's capabilities as a general-purpose simulator of hydrological and water quality, it is also  
 411 a valuable tool for management and interventions. The software's design prioritizes customization, making it easy  
 412 to incorporate specific operational preferences which are not typical to capture due to the limited modelling scope  
 413 of most water simulators. For example, in the online customisation guides  
 414 (<https://barneydobson.github.io/wsi/how-to/>), we show how the behaviour of nodes and arcs can be altered to  
 415 accommodate changing abstraction licencing and environmental flow requirements for the Oxford case study,  
 416 with results shown in Figure 7. Liu et al., (2023b), demonstrate how a user can implement highly sophisticated  
 417 water quality management strategies in WSIMOD based on pollution load allocation.



418

419 **Figure 7: Example of implementing a 'hands-off flow' (HOF), which abstractions cannot draw water below, for the**  
 420 **reservoir abstraction node in the Oxford case study model. The 'customised' simulation are with HOF implemented,**  
 421 **and 'baseline' simulation without.**



## 422 **4 Discussion**

### 423 **4.1 Why are integrated water systems models necessary?**

424 We defined four key goals for integrated models of water systems. In this section we will discuss how WSIMOD  
425 helps to meet these goals.

#### 426 **4.1.1 Integration to understand emergent behaviour in water systems**

427 Emergent behaviours in human and environmental systems arise from the interactions between multiple  
428 components and are difficult to predict and understand without addressing the complexity that occurs from this  
429 interconnectedness (Liu et al. 2007). The human-altered water cycle covers a diverse set of components and so  
430 any attempt to understand the fundamental (physical and operational) drivers behind emergent behaviours must  
431 acknowledge the interactions and feedbacks between humans and hydrological processes (Wada et al. 2017).  
432 WSIMOD represents many components within water cycle needed to capture these interactions to reveal and  
433 quantify fundamental processes, for example, that in-river water quality during wetter periods is driven by  
434 agricultural processes, and during drier periods by urban processes (Liu et al. 2022). Without integrated modelling,  
435 we risk oversimplifying the system and omitting feedbacks that could have significant implications for water  
436 management decisions (Dobson and Mijic 2020).

437 Understanding and reducing uncertainty behind behaviour in hydrological systems requires intercomparison of  
438 hydrological process representations and thus a flexible modelling framework (Knoben et al. 2019). We suggest  
439 the same is true for the wider water cycle, and so anticipate that the flexible approach to integrated modelling in  
440 WSIMOD is well suited for these purposes, for example, by comparing different assumptions around water  
441 consuming behaviour (Dobson et al. 2021). Furthermore, customisability enables accommodating unconventional  
442 water systems that may stress the assumptions of underlying process representations. This approach can help  
443 identify any weaknesses or gaps in the model and refine our understanding of the both the process in question and  
444 the behaviour of the wider system. For example, the poor simulations of hydraulic structures identified in Dobson  
445 et al. (2022) are a result of reduced accuracy in capturing water head, in turn caused by the discrete modelling  
446 scheme inherent to a non-tightly coupled integrated framework. Identifying this weakness has implications beyond  
447 simulating hydraulic structures and provides guidance for future work.

448 Finally, research that contributes to knowledge must be reproducible, and if it is not then it can hardly be  
449 considered science (Hutton et al. 2016). To ensure that WSIMOD applications are reproducible the software is  
450 open source with a permissive licence, and it is provided with significant documentation and worked examples to  
451 ensure that it is used as intended. This documentation transparently lists the assumptions made for each model  
452 component, both in the source code of that component and in a self-contained library page (Dobson et al. 2023c).  
453 Furthermore, the ability to save the Model class in a self-contained and human readable file enables publishing  
454 WSIMOD applications in an easy to reproduce way.

#### 455 **4.1.2 Integration to expand model boundaries**

456 The WSIMOD modelling framework offers a versatile approach to representing various processes that are  
457 commonly treated as boundary conditions in water cycle models. One example is the agricultural-hydrological





458 Land node, which utilizes the CatchWat model to estimate the pollution and flow in rivers upstream of the model  
459 region (Liu et al. 2022, 2023a). Accounting for upstream pollution and dilution is crucial in accurately assessing  
460 the impact of urban processes on in-river conditions. By incorporating the Land node, WSIMOD provides a means  
461 to capture the typically neglected upstream agricultural system in urban wastewater studies, which is often viewed  
462 as a boundary. WSIMOD also facilitates the representation of boundary conditions within urban systems. For  
463 instance, it allows for the inclusion of abstractions, which are critical in understanding in-river impacts of  
464 wastewater effluent because of their influence on dilution during low flows (Dobson and Mijic 2020).

465 Besides achieving accurate simulations, we further propose that a modelled representation of boundary conditions  
466 is essential towards assessment of the system under future scenarios. For example, by treating the upstream river  
467 as a model rather than as a fixed boundary condition, the in-river impact of the wastewater system for future  
468 scenarios can be quantified, which would not be possible if the changes to upstream river behaviour under the  
469 future scenario were fixed (Mijic et al. 2022).

#### 470 **4.1.3 Integration to evaluate intervention impacts at whole-water system scale**

471 The OOP deployed in WSIMOD enables flexibility to incorporate a variety of different physical and management  
472 interventions to the water cycle. As demonstrated in Liu et al., (2023a), the included WSIMOD nodes may be  
473 customised both in terms of their parameters and in terms of their physical processes to represent a range of Nature  
474 Based Solutions (NBS) such as flood plains, runoff attenuation features, regenerative farming, urban green space,  
475 and urban wetlands. Because all WSIMOD nodes can communicate with all other nodes, this is a reasonably  
476 straightforward exercise as new rules for interactions with other model components do not need to be redefined  
477 when additional components are added to the model.

478 Due to the flexibility and coverage provided by WSIMOD models, it is often straightforward to include decision-  
479 making, policy constraints, and operational rules (Dobson and Mijic 2020). The high-level control over  
480 orchestration provides an ideal place to represent these stakeholder actions that require information on a multitude  
481 of states across the wider system. For example, the amount available to be abstracted from the River Thames, UK,  
482 is based off both reservoir levels and river flow, meaning no specific node has access to all of the information  
483 required to determine abstraction amount (Environment Agency 1991). Rather than creating a complicated  
484 interaction between London reservoirs and the River Thames, a modeller may more simply inspect the state of  
485 these two systems in the orchestration to dynamically set the abstraction capacity, filling the role of an operator  
486 (Dobson and Mijic 2020). Thus, the nationally important question of modelling changes to abstraction policies  
487 may easily be implemented in WSIMOD, and the impacts of these, both on water supply and in-river water quality,  
488 may be quantified (Mijic et al. 2022).

489 The parsimonious methods selected to represent physical components in WSIMOD ensure that simulations are  
490 computationally efficient. Further to this, if computational speed is of critical interest, then we have demonstrated  
491 that varying timestep and spatial resolution is possible while still achieving accurate simulations (Dobson,  
492 Watson-Hill, et al. 2022). The result is an integrated water cycle model that can be used for purposes that are not  
493 typically computationally tractable. Dobson et al., (2022), demonstrate exploration of uncertainty in sewer model  
494 parameters (roughness and runoff coefficient) is possible with WSIMOD, this study is the first of its kind because  
495 pipe network models are typically too computationally expensive to perform the numerous simulations required  
496 for uncertainty analysis. Meanwhile, Liu et al., (2023a), demonstrate that regional portfolios of 5-year catchment-



497 scale nature-based solutions can be created by applying optimisation to a WSIMOD model spanning 32  
498 catchments.

#### 499 **4.1.4 Integration to align simulations with systems level outcomes**

500 A key benefit we have found in applications of WSIMOD is that, due to the breadth of systems that are represented,  
501 performance metrics that are relevant to different stakeholders can usually be included in the model in a physically  
502 based way. The most obvious example, which we have drawn on throughout this paper, is the ability to place in-  
503 river impacts central to decision making. This enables better alignment with policy goals, such as the Water  
504 Framework Directive chemical water quality classifications, which are defined based on in-river average pollutant  
505 concentration (Environment Agency 2020b). However, a variety of other metrics can be conjunctively included  
506 such as the reliability of water supply. For example, Dobson & Mijic (2020) examine both the water quality and  
507 water supply benefits of a variety of water cycle interventions (leakage reduction, rainwater harvesting, etc.). Such  
508 an approach is particularly beneficial to stakeholder engagement because non-water facing stakeholders can still  
509 be represented and understand their interactions with the water cycle and system-wide goals. For example, Puchol-  
510 Salort et al., (2022), demonstrate how developers can measure the impacts of new developments on flooding,  
511 water quality and water supply. Furthermore, this study demonstrates how integrated modelling can enable them  
512 to quantify how to ensure developments achieve a net-zero impact on these metrics, interestingly revealing that  
513 retrofitting households outside of the development area is typically required to offset changes.

#### 514 **4.2 Data to support water systems integration**

515 A key challenge to the application of an integrated approach to water systems modelling is the difficulty of setting  
516 up models and associated availability of data. In WSIMOD, we provide parsimonious but physically based  
517 representations to ensure parameterisation is possible with widely available data and we propose model evaluation  
518 with observed in-river flow and water quality. The catalogue of documented nodes in WSIMOD (Dobson et al.  
519 2023c) presents data requirements to help users understand how feasible it will be to apply the approach in their  
520 study area. The data requirements will be entirely dependent on what nodes and at what resolution a model user  
521 chooses to represent. However, we provide a broad overview of data requirements for a generic catchment scale  
522 WSIMOD case study in Table 2, which may help to develop future automatic model setup at national and global  
523 scale to facilitate applicability.

524 **Table 2: List of datasets typically required in catchment-scale WSIMOD applications with references for data sources**  
525 **or further information.**

Category	WSIMOD input	Availability	Further information (scale)
Climate	Evapotranspiration	Global datasets available	(Khan et al. 2018), (global)
	Temperature	Global datasets available	(Morice et al. 2021) (global)
	Precipitation	Global datasets available	(Sun et al. 2018) (global)
Rural	Catchment outlines	Global datasets available	(Lin et al. 2019a) (global)
	Catchment connectivity	Global datasets available	(Lin et al. 2019b) (global)
	Catchment hydrologic parameters	Global datasets available for some hydrological models	(Zhang and Schaap 2018) (global)
	Crop surfaces	Global datasets available	(Thenkabail et al. 2016) (global)
	Crop properties	Lookup tables available	(Allen et al. 1998) (-)



	Pollutants/nutrients	National datasets may be available with high uncertainty (Liu et al. 2022) (UK)
Urban	Population	Global datasets available (Leyk et al. 2019) (global)
	Garden area	National datasets may be available, otherwise rule of thumb may be acceptable (Office for National Statistics (2021) (UK)
	Wastewater treatment plants	European dataset available through Urban Wastewater Treatment Directive (European Commission (2016) (Europe)
	Foul catchments	National datasets may be available, otherwise contacting wastewater companies required (Hoffmann et al. 2022) (UK)
Water use	Irrigation water use	Global datasets available (Thenkabail et al. 2009) (global)
	Water resources system	Not typically available, contacting - water supply companies required
Evaluation	Flow observations	National datasets typically available (Fry 2010) (UK)
	Water quality observations (river and WWTW)	National datasets may be available (Environment Agency 2020a) (England)
	Reservoir levels	Not typically available, contacting - water supply companies required

#### 526 4.3 Future work and research direction

527 One of the main concerns that other modellers have expressed regarding a WSIMOD-like approach is the level of  
 528 detail with which components are represented. While hydrologists and agricultural modellers are often  
 529 comfortable with parsimony and aggregation in their catchment modelling, most other parts of the water cycle  
 530 tend towards more complexity in their representations. This desire for complexity is likely due to the detailed  
 531 application context in which different models have been developed. For instance, designing a new process in a  
 532 wastewater treatment plant inevitably requires a highly detailed model (Hreiz, Latifi, and Roche 2015). However,  
 533 questions are being raised in many fields with a tradition of complex modelling about the need for such  
 534 complexity. Models of in-river phosphorus (Jackson-Blake et al. 2017), urban flooding (Li and Willems 2020),  
 535 and sewer flow (Dobson, Watson-Hill, et al. 2022; Thrysoe, Arnbjerg-Nielsen, and Borup 2019) have shown that  
 536 good results can be achieved with simpler approaches. While practical modellers typically question what level of  
 537 complexity is necessary to answer their questions, scientific modellers examine the impacts of assumptions to  
 538 build evidence around whether they are suitable and under which circumstances. WSIMOD prioritises integration  
 539 of the whole-water cycle, which is enabled by reduced-complexity modelling of the system components. In the  
 540 examples provided, we demonstrate that sacrificing complexity in terms of detail should be viewed as an  
 541 opportunity to better accommodate and contextualise components in the wider water cycle, as well as highlighting  
 542 the importance of interactions between components.

543 We see the WSIMOD platform as an ideal opportunity for the environmental modelling community to implement  
 544 and compare (or benchmark) different modelling assumptions and examine water cycle impacts. We plan to  
 545 continue developing the representation of different components and testing whether more complex representations  
 546 can improve simulations. Our focus will be on representations of a wider range of NBS, treatment plants, and  
 547 urban sewer network hydraulic structures. Furthermore, we believe that complementing WSIMOD with machine  
 548 learning representations of components that are too complex to be captured in a physical way can be a promising  
 549 approach, thus implementing a "surrogate" strategy (Razavi et al. 2022; Razavi, Tolson, and Burn 2012).



550 A key opportunity for improving the accessibility of WSIMOD will be in the development of a graphical user  
551 interface (GUI). The current implementation as a Python package makes the software well suited to customisable  
552 and flexible simulations for programmers, but inaccessible to a wide range of potential users. We see a variety of  
553 different approaches towards greater interactivity and visualisation that are not mutually exclusive. A “virtual  
554 decision room” approach may provide an ideal environment for non-technical stakeholders to explore simulation  
555 results and to highlight integrated system-wide impacts (Schouten, van den Hooff, and Feldberg 2016).  
556 Meanwhile, incorporation into GIS-based frameworks such as 3DNet (Todorović et al. 2019) or Google Earth  
557 Engine (Gorelick et al. 2017) would enable more seamless incorporation of pre-processing and provide a suite of  
558 streamlined tools to help users create, edit, and run WSIMOD models.

## 559 **5 Conclusion**

560 We have presented the theoretical underpinning of WSIMOD, which is an open-source software for simulating a  
561 range of urban and rural processes and operations in the integrated water cycle. WSIMOD represents different  
562 components of the water cycle as nodes that are connected by arcs. The nodes that we have discussed throughout  
563 the paper are parsimonious implementations that are conducive towards easy parameterisation and setup. Arcs  
564 convey interactions between nodes that fall under four key categories: pushes (a node has water to go somewhere),  
565 pulls (a node needs water from somewhere), requests (interaction represents flux of water and/or pollutants), and  
566 checks (interaction does not represent flux). This integration framework allows all nodes to communicate with  
567 each other, thus facilitating a flexible method that can accommodate a wide variety of water systems. Because  
568 this approach uses object-oriented programming, WSIMOD enables customisation to capture unconventional  
569 behaviours and implementation of a wide variety of physical and management interventions.

570 In summary, our early case studies show WSIMOD to be a useful and versatile tool for water systems modelling.  
571 We hope to have persuaded other modellers of the importance of an integrated approach and believe the design  
572 philosophy behind WSIMOD can serve as a helpful starting point for understanding integration in their respective  
573 contexts.

## 574 **6 Competing interests**

575 The contact author has declared that none of the authors has any competing interests.

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581 improved the paper.

582



583 **8 Code availability**

584 WSIMOD is provided open-source under the terms of the BSD-3-Clause license. The code can be accessed at  
585 <https://github.com/barneydobson/wsi> (last access: 2023-07-19), and documentation at  
586 <https://barneydobson.github.io/wsi/> (last access: 2023-07-19), with further technical details in Appendix 1. The  
587 code has been tested up to Python 3.10 and requires minimal dependencies (see website).

588 **9 Author contributions**

589 BD and LL created and tested all model code and documentation. All authors were involved in theoretical  
590 development and writing.

591 **10 References**

- 592 Allen, Richard G., Luis S. Pereira, Dirk Raes, Martin Smith, and others. 1998. "Crop Evapotranspiration-  
593 Guidelines for Computing Crop Water Requirements-FAO Irrigation and Drainage Paper 56." *Fao, Rome*  
594 300(9):D05109.
- 595 Almeida, M. C., D. Butler, and E. Friedler. 1999. "At-Source Domestic Wastewater Quality." *Urban Water*  
596 1(1):49–55. doi: 10.1016/s1462-0758(99)00008-4.
- 597 Bach, Peter M., Wolfgang Rauch, Peter S. Mikkelsen, David T. McCarthy, and Ana Deletic. 2014. "A Critical  
598 Review of Integrated Urban Water Modelling - Urban Drainage and Beyond." *Environmental Modelling*  
599 *and Software* 54:88–107. doi: 10.1016/j.envsoft.2013.12.018.
- 600 Belete, Getachew F., Alexey Voinov, and Gerard F. Laniak. 2017. "An Overview of the Model Integration  
601 Process: From Pre-Integration Assessment to Testing." *Environmental Modelling and Software* 87:49–63.  
602 doi: 10.1016/j.envsoft.2016.10.013.
- 603 Best, M. J., M. Pryor, D. B. Clark, G. G. Rooney, R. ... L. H. Essery, C. B. Ménard, J. M. Edwards, M. A. Hendry,  
604 A. Porson, N. Gedney, L. M. Mercado, S. Sitch, E. Blyth, O. Boucher, P. M. Cox, C. S. B. Grimmond, and  
605 R. J. Harding. 2011. "The Joint UK Land Environment Simulator (JULES), Model Description – Part 1:  
606 Energy and Water Fluxes." *Geoscientific Model Development* 4(3):677–99. doi: 10.5194/gmd-4-677-2011.
- 607 Beven, K. 2007. "Towards Integrated Environmental Models of Everywhere: Uncertainty, Data and Modelling as  
608 a Learning Process." *Hydrology and Earth System Sciences* 11(1):460–67. doi: 10.5194/hess-11-460-2007.
- 609 Blair, Gordon S., Keith Beven, Rob Lamb, Richard Bassett, Kris Cauwenberghs, Barry Hankin, Graham Dean,  
610 Neil Hunter, Liz Edwards, Vatsala Nundloll, Faiza Samreen, Will Simm, and Ross Towe. 2019. "Models  
611 of Everywhere Revisited: A Technological Perspective." *Environmental Modelling & Software*  
612 122(August):104521. doi: 10.1016/j.envsoft.2019.104521.
- 613 Bowes, Michael J., Linda K. Armstrong, Sarah A. Harman, Heather D. Wickham, David J. E. Nicholls, Peter M.  
614 Scarlett, Colin Roberts, Helen P. Jarvie, Gareth H. Old, Emma Gozzard, Nuria Bachiller-Jareno, and Daniel  
615 S. Read. 2018. "Weekly Water Quality Monitoring Data for the River Thames (UK) and Its Major  
616 Tributaries (2009–2013): The Thames Initiative Research Platform." *Earth System Science Data*  
617 10(3):1637–53. doi: 10.5194/essd-10-1637-2018.
- 618 Burger, Gregor, Peter M. Bach, Christian Ulrich, Günther Leonhardt, Manfred Kleidorfer, and Wolfgang Rauch.



- 619 2016. “Designing and Implementing a Multi-Core Capable Integrated Urban Drainage Modelling  
620 Toolkit:Lessons from CityDrain3.” *Advances in Engineering Software* 100:277–89. doi:  
621 10.1016/j.advengsoft.2016.08.004.
- 622 Dobson, Barnaby, Saoirse Barry, Robin Maes-Prior, Ana Mijic, Guy Woodward, and William D. Pearse. 2022.  
623 “Predicting Catchment Suitability for Biodiversity at National Scales.” *Water Research* 221:118764.
- 624 Dobson, Barnaby, Tijana Jovanovic, Yuting Chen, Athanasios Paschalis, Adrian Butler, and Ana Mijic. 2021.  
625 “Integrated Modelling to Support Analysis of COVID-19 Impacts on London’s Water System and in-River  
626 Water Quality.” *Frontiers in Water* 3(April):26. doi: 10.3389/frwa.2021.641462.
- 627 Dobson, Barnaby, Leyang Liu, and Ana Mijic. 2023a. “Water Systems Integrated Modelling Framework,  
628 WSIMOD: A Python Package for Integrated Modelling of Water Quality and Quantity across the Water  
629 Cycle.” *Journal of Open Source Software* 8(83):4996. doi: 10.21105/joss.04996.
- 630 Dobson, Barnaby, Leyang Liu, and Ana Mijic. 2023b. “WSIMOD Documentation.” *GitHub Pages*. Retrieved  
631 March 24, 2023 (<https://barneydobson.github.io/wsi>).
- 632 Dobson, Barnaby, Leyang Liu, and Ana Mijic. 2023c. “WSIMOD Documentation.” *GitHub Pages*.
- 633 Dobson, Barnaby, and Ana Mijic. 2020. “Protecting Rivers by Integrating Supply-Wastewater Infrastructure  
634 Planning and Coordinating Operational Decisions.” *Environmental Research Letters* (December 2016):0–  
635 31. doi: 10.1088/1748-9326/abb050.
- 636 Dobson, Barnaby, Thorsten Wagener, and Francesca Pianosi. 2019. “How Important Are Model Structural and  
637 Contextual Uncertainties When Estimating the Optimized Performance of Water Resource Systems?” *Water  
638 Resources Research* (2017):1–24. doi: 10.1029/2018WR024249.
- 639 Dobson, Barnaby, Hannah Watson-Hill, Samer Muhandes, Morten Borup, and Ana Mijic. 2022. “A Reduced  
640 Complexity Model With Graph Partitioning for Rapid Hydraulic Assessment of Sewer Networks.” *Water  
641 Resources Research* 58(1). doi: 10.1029/2021WR030778.
- 642 Douglas-Mankin, K. R., R. Srinivasan, and J. G. Arnold. 2010. “Soil and Water Assessment Tool (SWAT) Model:  
643 Current Developments and Applications.” *Transactions of the ASABE* 53(5):1423–31. doi:  
644 10.13031/2013.34915.
- 645 Environment Agency. 1991. “Lower Thames Operating Agreement, 28/39/M/2.” Retrieved  
646 (<http://www.thames21.org.uk/wp-content/uploads/2013/08/LTOA-Description-and-Operation.pdf>).
- 647 Environment Agency. 2020a. “Open Water Quality Archive Datasets (WIMS).” Retrieved March 19, 2020  
648 (<https://environment.data.gov.uk/water-quality/view/download>).
- 649 Environment Agency. 2020b. “WFD Classification Status Cycle 2.”
- 650 European Commission. 2016. “Urban Wastewater Treatment Directive - Treatment Plants.” Retrieved March 17,  
651 2020 (<https://uwwtdeu/United-Kingdom/download>).
- 652 Fry, M. J. 2010. “Hydrological Data Management Systems within a National River Flow Archive.” in *Role of  
653 hydrology in managing consequences of a changing global environment*. British Hydrological Society.
- 654 Gamma, Erich, Ralph Johnson, Richard Helm, Ralph E. Johnson, and John Vlissides. 1995. *Design Patterns:  
655 Elements of Reusable Object-Oriented Software*. Pearson Education.
- 656 Giakoumis, T., and N. Voulvoulis. 2023. “Combined Sewer Overflows: Relating Event Duration Monitoring Data  
657 to Wastewater Systems’ Capacity in England.” *Environmental Science: Water Research & Technology*  
658 9(3):707–22. doi: 10.1039/D2EW00637E.



- 659 Gorelick, Noel, Matt Hancher, Mike Dixon, Simon Ilyushchenko, David Thau, and Rebecca Moore. 2017.  
660 "Google Earth Engine: Planetary-Scale Geospatial Analysis for Everyone." *Remote Sensing of Environment*  
661 202:18–27. doi: 10.1016/j.rse.2017.06.031.
- 662 Graham, Richard, Galen Shipman, Brian Barrett, Ralph Castain, George Bosilca, and Andrew Lumsdaine. 2006.  
663 "Open MPI: A High-Performance, Heterogeneous MPI." Pp. 1–9 in *2006 IEEE International Conference*  
664 *on Cluster Computing*. Barcelona: IEEE.
- 665 Harpham, Q. K., A. Hughes, and R. V. Moore. 2019. "Introductory Overview: The OpenMI 2.0 Standard for  
666 Integrating Numerical Models." *Environmental Modelling and Software* 122(April):104549. doi:  
667 10.1016/j.envsoft.2019.104549.
- 668 Hoffmann, Till, Sarah Bunney, Barbara Kasprzyk-Hordern, and Andrew Singer. 2022. "Wastewater Catchment  
669 Areas in Great Britain." *Authorea Preprints*.
- 670 Höllermann, Britta, and Mariele Evers. 2017. "Perception and Handling of Uncertainties in Water Management—  
671 A Study of Practitioners' and Scientists' Perspectives on Uncertainty in Their Daily Decision-Making."  
672 *Environmental Science & Policy* 71:9–18. doi: 10.1016/j.envsci.2017.02.003.
- 673 Hreiz, Rainier, M. A. Latifi, and Nicolas Roche. 2015. "Optimal Design and Operation of Activated Sludge  
674 Processes: State-of-the-Art." *Chemical Engineering Journal* 281:900–920. doi: 10.1016/j.cej.2015.06.125.
- 675 Hutton, Christopher, Thorsten Wagener, Jim Freer, Dawei Han, Chris Duffy, and Berit Arheimer. 2016. "Most  
676 Computational Hydrology Is Not Reproducible, so Is It Really Science?" *Water Resources Research*  
677 52(10):7548–55. doi: 10.1002/2016WR019285.
- 678 IFAK. 2007. "SIMBA (Simulation of Biological Wastewater Systems): Manual and Reference."
- 679 Jackson-Blake, L. A., J. E. Sample, A. J. Wade, R. C. Helliwell, and R. A. Skeffington. 2017. "Are Our Dynamic  
680 Water Quality Models Too Complex? A Comparison of a New Parsimonious Phosphorus Model, SimplyP,  
681 and INCA-P." *Water Resources Research* 53(7):5382–99. doi: 10.1002/2016WR020132.
- 682 Khan, Muhammad Sarfraz, Umar Waqas Liaqat, Jongjin Baik, and Minha Choi. 2018. "Stand-Alone Uncertainty  
683 Characterization of GLEAM, GLDAS and MOD16 Evapotranspiration Products Using an Extended Triple  
684 Collocation Approach." *Agricultural and Forest Meteorology* 252:256–68. doi:  
685 10.1016/j.agrformet.2018.01.022.
- 686 Kirchner, James W. 2009. "Catchments as Simple Dynamical Systems: Catchment Characterization, Rainfall-  
687 Runoff Modeling, and Doing Hydrology Backward." *Water Resources Research* 45(2). doi:  
688 10.1029/2008WR006912.
- 689 Knoben, Wouter J. M., Jim E. Freer, Keirnan J. A. Fowler, Murray C. Peel, and Ross A. Woods. 2019. "Modular  
690 Assessment of Rainfall–Runoff Models Toolbox (MARRMoT) v1. 2: An Open-Source, Extendable  
691 Framework Providing Implementations of 46 Conceptual Hydrologic Models as Continuous State-Space  
692 Formulations." *Geoscientific Model Development* 12(6):2463–80.
- 693 Lee, Douglass B. 1973. "Requiem for Large-Scale Models." *Journal of the American Institute of Planners*  
694 39(3):163–78. doi: 10.1080/01944367308977851.
- 695 Leyk, Stefan, Andrea E. Gaughan, Susana B. Adamo, Alex de Sherbinin, Deborah Balk, Sergio Freire, Amy Rose,  
696 Forrest R. Stevens, Brian Blankespoor, Charlie Frye, Joshua Comenetz, Alessandro Soricchetta, Kytt  
697 MacManus, Linda Pistoletti, Marc Levy, Andrew J. Tatem, and Martino Pesaresi. 2019. "The Spatial  
698 Allocation of Population: A Review of Large-Scale Gridded Population Data Products and Their Fitness for



- 699 Use.” *Earth System Science Data* 11(3):1385–1409. doi: 10.5194/essd-11-1385-2019.
- 700 Li, Xiaohan, and Patrick Willems. 2020. “A Hybrid Model for Fast and Probabilistic Urban Pluvial Flood  
701 Prediction.” *Water Resources Research* 1–26. doi: 10.1029/2019wr025128.
- 702 Lin, Peirong, Ming Pan, Hylke E. Beck, Yuan Yang, Dai Yamazaki, Renato Frasson, Cédric H. David, Michael  
703 Durand, Tamlin M. Pavelsky, George H. Allen, Colin J. Gleason, and Eric F. Wood. 2019a. “Global  
704 Reconstruction of Naturalized River Flows at 2.94 Million Reaches.” *Water Resources Research*  
705 55(8):6499–6516. doi: 10.1029/2019WR025287.
- 706 Lin, Peirong, Ming Pan, Hylke E. Beck, Yuan Yang, Dai Yamazaki, Renato Frasson, Cédric H. David, Michael  
707 Durand, Tamlin M. Pavelsky, George H. Allen, Colin J. Gleason, and Eric F. Wood. 2019b. “Global  
708 Reconstruction of Naturalized River Flows at 2.94 Million Reaches.” *Water Resources Research*  
709 55(8):6499–6516. doi: 10.1029/2019WR025287.
- 710 Lindström, Göran, Charlotta Pers, Jörgen Rosberg, Johan Strömqvist, and Berit Arheimer. 2010. “Development  
711 and Testing of the HYPE (Hydrological Predictions for the Environment) Water Quality Model for Different  
712 Spatial Scales.” *Hydrology Research* 41(3–4):295–319. doi: 10.2166/nh.2010.007.
- 713 Liu, Jianguo, Thomas Dietz, Stephen R. Carpenter, Marina Alberti, Carl Folke, Emilio Moran, Alice N. Pell, Peter  
714 Deadman, Timothy Kratz, Jane Lubchenco, Elinor Ostrom, Zhiyun Ouyang, William Provencher, Charles  
715 L. Redman, Stephen H. Schneider, and William W. Taylor. 2007. “Complexity of Coupled Human and  
716 Natural Systems.” *Science* 317(5844):1513–16. doi: 10.1126/science.1144004.
- 717 Liu, Leyang, Barnaby Dobson, and Ana Mijic. 2022. “Hierarchical Systems Integration for Coordinated Urban-  
718 Rural Water Quality Management at a Catchment Scale.” *Science of the Total Environment* 806:150642.  
719 doi: 10.1016/j.scitotenv.2021.150642.
- 720 Liu, Leyang, Barnaby Dobson, and Ana Mijic. 2023a. “Optimisation of Urban-Rural Nature-Based Solutions for  
721 Integrated Catchment Water Management.” *Journal of Environmental Management* 329:117045. doi:  
722 10.1016/j.jenvman.2022.117045.
- 723 Liu, Leyang, Barnaby Dobson, and Ana Mijic. 2023b. “Water Quality Management at a Critical Checkpoint by  
724 Coordinated Multi-Catchment Urban-Rural Load Allocation.” *Journal of Environmental Management*  
725 340:117979. doi: 10.1016/j.jenvman.2023.117979.
- 726 Mateo-Sagasta, Javier, S. Marjani Zadeh, Hugh Turrall, and Jacob Burke. 2017. “Water Pollution from  
727 Agriculture: A Global Review.” *Executive Summary* 35.
- 728 Mijic, Ana, B. Bromwich, D. Crilly, B. Dobson, D. Forrow, J. Giambona, S. Kirk, L. Liu, and R. MacDonald.  
729 2022. *Systems Approach to Regional Water Planning*.
- 730 Morice, C. P., J. J. Kennedy, N. A. Rayner, J. P. Winn, E. Hogan, R. E. Killick, R. J. H. Dunn, T. J. Osborn, P.  
731 D. Jones, and I. R. Simpson. 2021. “An Updated Assessment of Near-Surface Temperature Change From  
732 1850: The HadCRUT5 Data Set.” *Journal of Geophysical Research: Atmospheres* 126(3). doi:  
733 10.1029/2019JD032361.
- 734 Mortazavi-Naeini, Mohammad, Gianbattista Busi, J. Alex Elliott, Jim W. Hall, and Paul G. Whitehead. 2019.  
735 “Assessment of Risks to Public Water Supply from Low Flows and Harmful Water Quality in a Changing  
736 Climate.” *Water Resources Research* 2018WR022865. doi: 10.1029/2018WR022865.
- 737 Office for National Statistics. 2021. “Access to Gardens and Public Green Space in Great Britain.” Retrieved  
738 April 13, 2023





- 739 (https://www.ons.gov.uk/economy/environmentalaccounts/datasets/accesstogardensandpublicgreenspacei  
740 ngreatbritain).
- 741 Puchol-Salort, Pepe, Stanislava Boskovic, Barnaby Dobson, Maarten van Reeuwijk, and Ana Mijic. 2022. “Water  
742 Neutrality Framework for Systemic Design of New Urban Developments.” *Water Research* 219:118583.  
743 doi: 10.1016/j.watres.2022.118583.
- 744 Rauch, W., C. Urich, P. M. Bach, B. C. Rogers, F. J. de Haan, R. R. Brown, M. Mair, D. T. McCarthy, M.  
745 Kleidorfer, R. Sitzenfrei, and A. Deletic. 2017. “Modelling Transitions in Urban Water Systems.” *Water  
746 Research* 126:501–14. doi: 10.1016/j.watres.2017.09.039.
- 747 Razavi, Saman, David M. Hannah, Amin Elshorbagy, Sujay Kumar, Lucy Marshall, Dimitri P. Solomatine, Amin  
748 Dezfuli, Mojtaba Sadegh, and James Famiglietti. 2022. “Coevolution of Machine Learning and Process-  
749 based Modelling to Revolutionize Earth and Environmental Sciences: A Perspective.” *Hydrological  
750 Processes* 36(6). doi: 10.1002/hyp.14596.
- 751 Razavi, Saman, Bryan A. Tolson, and Donald H. Burn. 2012. “Review of Surrogate Modeling in Water  
752 Resources.” *Water Resources Research* 48(7). doi: 10.1029/2011WR011527.
- 753 Schmitt, T. G., and W. C. Huber. 2006. “The Scope of Integrated Modelling: System Boundaries, Sub-Systems,  
754 Scales and Disciplines.” *Water Science and Technology* 54(6–7):405–13. doi: 10.2166/wst.2006.595.
- 755 Schouten, Alexander P., Bart van den Hooff, and Frans Feldberg. 2016. “Virtual Team Work.” *Communication  
756 Research* 43(2):180–210. doi: 10.1177/0093650213509667.
- 757 Sun, Qiaohong, Chiyan Miao, Qingyun Duan, Hamed Ashouri, Soroosh Sorooshian, and Kuo-Lin Hsu. 2018.  
758 “A Review of Global Precipitation Data Sets: Data Sources, Estimation, and Intercomparisons.” *Reviews of  
759 Geophysics* 56(1):79–107. doi: 10.1002/2017RG000574.
- 760 Tang, Fiona H. M., Manfred Lenzen, Alexander McBratney, and Federico Maggi. 2021. “Risk of Pesticide  
761 Pollution at the Global Scale.” *Nature Geoscience* 14(4):206–10. doi: 10.1038/s41561-021-00712-5.
- 762 Thenkabail, P., P. Teluguntla, J. Xiong, A. Oliphant, and R. Massey. 2016. “NASA MEaSURES Global Food  
763 Security Support Analysis Data (GFSAD) Crop Mask 2010 Global 1 Km V001.” *NASA EOSDIS Land  
764 Processes DAAC*. doi: 10.5067/MEaSURES/GFSAD/GFSAD1KCD.001.
- 765 Thenkabail, Prasad S., Chandrashekar M. Biradar, Praveen Noojipady, Venkateswarlu Dheeravath, Yuanjie Li,  
766 Manohar Velpuri, Muralikrishna Gumma, Obi Reddy P. Gangalakunta, Hugh Turrall, Xueliang Cai, Jagath  
767 Vithanage, Mitchell A. Schull, and Rishiraj Dutta. 2009. “Global Irrigated Area Map (GIAM), Derived from  
768 Remote Sensing, for the End of the Last Millennium.” *International Journal of Remote Sensing*  
769 30(14):3679–3733. doi: 10.1080/01431160802698919.
- 770 Thrysoe, Cecilie, Karsten Arnbjerg-Nielsen, and Morten Borup. 2019. “Identifying Fit-for-Purpose Lumped  
771 Surrogate Models for Large Urban Drainage Systems Using GLUE.” *Journal of Hydrology* 568(July  
772 2018):517–33. doi: 10.1016/j.jhydrol.2018.11.005.
- 773 Todorović, Andrijana, Miloš Stanić, Željko Vasilić, and Jasna Plavšić. 2019. “The 3DNet-Catch Hydrologic  
774 Model: Development and Evaluation.” *Journal of Hydrology* 568:26–45. doi:  
775 10.1016/j.jhydrol.2018.10.040.
- 776 Tscheikner-Gratl, Franz, Vasilis Bellos, Alma Schellart, Antonio Moreno-Rodenas, Manoranjan Muthusamy,  
777 Jeroen Langeveld, Francois Clemens, Lorenzo Benedetti, Miguel Angel Rico-Ramirez, Rita Fernandes de  
778 Carvalho, Lutz Breuer, James Shucksmith, Gerard B. M. Heuvelink, and Simon Tait. 2019. “Recent Insights



- 779 on Uncertainties Present in Integrated Catchment Water Quality Modelling.” *Water Research* 150:368–79.  
780 doi: 10.1016/j.watres.2018.11.079.
- 781 Vanhooren, Henk, Jurgen Meirlaen, Youri Amerlinck, Filip Claeys, Hans Vangheluwe, and Peter A.  
782 Vanrolleghem. 2003. “WEST: Modelling Biological Wastewater Treatment.” *Journal of Hydroinformatics*  
783 5(1):27–50. doi: 10.2166/hydro.2003.0003.
- 784 Voinov, Alexey, and Herman H. Shugart. 2013. “‘Integronsters’, Integral and Integrated Modeling.”  
785 *Environmental Modelling and Software* 39:149–58. doi: 10.1016/j.envsoft.2012.05.014.
- 786 Wada, Y., J. Liu, M. F. P. Bierkens, A. De Roo, N. Wanders, P. A. Dirmeyer, J. S. Famiglietti, N. Hanasaki, M.  
787 Konar, M. Sivapalan, H. M. Schmied, T. Oki, Y. Pokhrel, T. J. Troy, A. I. J. M. Van Dijk, T. Van Emmerik,  
788 M. H. J. Van Huijgevoort, H. A. J. Van Lanen, C. J. Vörösmarty, and H. Wheatler. 2017. “Human-Water  
789 Interface in Hydrological Modelling: Current Status and Future Directions.” *Hydrology and Earth System*  
790 *Sciences* 21(8):4169–93. doi: 10.5194/hess-21-4169-2017.
- 791 Whitehead, P. G., E. J. Wilson, and D. Butterfield. 1998. “A Semi-Distributed Integrated Nitrogen Model for  
792 Multiple Source Assessment in Catchments (INCA): Part I — Model Structure and Process Equations.”  
793 *Science of The Total Environment* 210–211:547–58. doi: 10.1016/S0048-9697(98)00037-0.
- 794 Zhang, Yonggen, and Marcel G. Schaap. 2018. “A High-Resolution Global Map of Soil Hydraulic Properties  
795 Produced by a Hierarchical Parameterization of a Physically-Based Water Retention Model.”
- 796 Zomorodian, Mehdi, Sai Hin Lai, Mehran Homayounfar, Shaliza Ibrahim, Seyed Ehsan Fatemi, and Ahmed El-  
797 Shafie. 2018. “The State-of-the-Art System Dynamics Application in Integrated Water Resources  
798 Modeling.” *Journal of Environmental Management* 227(June 2017):294–304. doi:  
799 10.1016/j.jenvman.2018.08.097.  
800

## 801 **11 Appendix 1 – technical details in WSIMOD**

### 802 **11.1 A1.1 Tank object to generalise water stores**

803 A particularly important concept in developing new and understanding existing nodes is the Tank object. The  
804 concept of a water store is so common in water systems that a generic Tank object is provided. Tanks streamline  
805 a variety of uses for stores and have a range of ‘children’ that implement travel time (A1.2) and pollutant decay  
806 (A1.3). A node that represents a water store should be a subclass of the Storage class (see Figure 2), which itself  
807 is a generic node wrapper for the Tank object. The simplest case of a tank would be a water supply reservoir,  
808 demonstrated for WSIMOD at a lumped London scale in (Dobson and Mijic 2020). However, many nodes use  
809 stores but in an auxiliary fashion, for example WWTWs have temporary storage tanks while FWTWs have service  
810 reservoir tanks.

### 811 **11.2 A1.2 Non-flux pollutant changes**

812 By default, everything tracked in a VQIP follows mass balance. However, in the water cycle, many pollutants  
813 undergo transformations due to biological, physical, or chemical processes, and thus preservation of mass may be  
814 insufficient to simulate them. WSIMOD represents the nitrogen/phosphorus cycles in soil (see documentation of



815 Land nodes) and denitrification/mineralisation/production/macrophyte uptake in rivers (see documentation of  
816 River nodes), based on the equations from (Lindström et al. 2010; Liu et al. 2022).

817 While the transformations that act on chemicals in soils and rivers are well studied in the literature, there are  
818 difficulties in conceptualising bio-chemical that take place in groundwater and sewers, despite agreement that  
819 these are chemically active (Almeida, Butler, and Friedler 1999). As a result, WSIMOD provides a generic two-  
820 parameter method to implement temperature sensitive chemical decay, given by:

$$M_t = M_{t-1}(1 - cd^{T_t - T_{ref}}) \quad (1)$$

821 , where M is the mass of a chemical in each timestep, c is a parameter that determines non-temperature sensitive  
822 decay, d is a parameter that determines temperature sensitive decay, T is the temperature with a reference  
823 temperature ( $T_{ref}$ ) assumed to be 20C. We do not intend that equation (1) can be a substitute for well-researched  
824 and verified process representations, however in our experience using WSIMOD it is an easy and useful option  
825 to improve water quality representations.

826 Wastewater and freshwater treatment processes are well-studied fields. However, simulation models of these  
827 systems require detailed information describing the different treatment technologies and processes that are present  
828 in a specific plant. While we plan to include these types of models in WSIMOD in the future, we have opted to  
829 take a parsimonious approach to treatment modelling under the assumption that most users will not have detailed  
830 information about the plants they model. This approach assumes that the plant performs a single operation, based  
831 on equation (1), to transform influent, which is then split into three streams of effluent, liquor, and solids.  
832 Depending on whether freshwater or wastewater treatment, these streams go to different places, see documentation  
833 of WTW for further details.

### 834 11.3 A1.3 Travel time of water

835 Arcs are the key model element to implement travel time of water. Two arc subclasses that provide alternate  
836 methods to implement travel time are provided in WSIMOD. The first, more simple approach, formulates the  
837 travel time of the arc as a dictionary object where each key is the number of timesteps remaining; when water is  
838 sent along the arc, it is combined with the any existing water for the key that matches the specified travel time.  
839 These travel times are updated at the end of each timestep. This method is computationally efficient because the  
840 number of operations each timestep is limited by the maximum number of timesteps the arc takes to traverse.  
841 However, this approach cannot represent a dynamic flow capacity, as is the case in, for example, sewer networks,  
842 where hydraulic head governs flow. Thus WSIMOD also contains a less computationally efficient arc to  
843 accommodate this behaviour, described and demonstrated in (Dobson, Watson-Hill, et al. 2022). Arcs can also  
844 implement pollutant changes associated with decay over this travel time, using equation (1).