

1 **Morphological and optical properties of carbonaceous aerosol particles from ship emissions and biomass burning**
2 **during a summer cruise measurement in the South China Sea**

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16 **Abstract.** Carbonaceous aerosols constitute a crucial component of atmospheric marine aerosols among which black carbon
17 (BC) and brown carbon (BrC) are important contributors to light absorption and hence the positive climatic radiative forcing
18 in the marine atmosphere. We conducted a month-long (May 05–June 09, 2021) onboard sample collections and online
19 measurements of carbonaceous aerosols to characterize their morphological and optical properties during a ship cruise in the
20 South China Sea (SCS), covering a marine region of 11.9–24.5 °N and 111.1–118.2 °E. Single particles were collected by a
21 single particle sampler and offline analyses were performed to investigate the mixing state and morphology using a
22 transmission electron microscope (TEM) coupled with energy dispersive X-ray spectroscopy (EDS). Online measurements of
23 BC in PM_{2.5} were made by a seven-wavelength aethalometer and organic carbon (OC)/elemental carbon (EC) mass
24 concentrations were measured by a semi-online OC/EC analyzer. ~~Single particle samples were classified into two modes: “stop”~~

25 ~~when the ship was anchored and “navigation” when the ship sailed at high speed.~~ Feret diameters of the single particles during
26 navigation and stop showed size distributions with the lognormal fitting peaks at 307 and 325 nm, respectively. The fresh
27 (without coating) and aged BC particles (after removal of coating by the electron beams in TEM) showed ~~comparable same~~
28 median fractal dimensions (~~1.65 vs 1.66~~1.61), in contrast to their different median lacunarities (0.53 vs 0.59). The aged BC
29 particles showed narrower Feret diameters (~~298229–1980–2557~~ nm) during navigation than those (~~30478–2982–2926~~ nm) of
30 freshly-emitted BC from the own ship during stop. Moreover, tar balls, as one important component of single particles from
31 ship emissions and as the tracer of biomass burning, were identified with geometrical diameters of 160–420 nm in the TEM
32 images. The ~~energy dispersive X-ray spectroscopy (EDS)~~EDS analyses showed those tar balls are mainly mixed with sea salt,
33 organics, BC, and sulfate. We also found a significant fraction of aged BC in various mixing states (core-shell, embedded)
34 with other components of the aerosol particles after long-range transport.

35 The campaign was further divided into several periods (before monsoon period, BMP; transition monsoon period, TMP;
36 after monsoon period, AMP; and ship pollution period, SPP) according to the wind direction during monsoon and the own ship
37 pollution. The median OC/EC ratios were 8.14, 5.20, 6.35, and 2.63 during BMP, TMP, AMP, and SPP, respectively, showing
38 higher OC/EC ratios for biomass burning emissions than for fossil fuel emissions. Additionally, the median absorption
39 Angström exponent (AAE) values derived from all wavelengths were 1.14, 1.02, 1.08, and 1.06 for BMP, TMP, AMP and SPP,
40 respectively. Particularly, a median AAE value of 1.93 was obtained during two significant biomass burning events. These
41 results showed that biomass burning (BB) and fossil fuel (FF) combustion contributed to 18–22% and 78–82% of all the BC
42 light absorption without the two intense biomass burning events, during which BB and FF accounted for 42% and 58%,
43 respectively. The two BB events originated from the Philippines and Southeast Asia before and after the summer monsoon.
44 Our results demonstrated that BC can serve as the core of aged particles but the fractal dimensions of BC aggregates were
45 subject to little variation; moreover, such BC particles become much more aggregated after aging in the marine atmosphere,
46 which further affects the light absorption of the BC particles in the SCS. ~~This study provides information about the morphology
47 and the optical properties of carbonaceous aerosols which can be used to evaluate their effects on light absorption and hence
48 the climatic radiative forcing in the SCS region.~~

49

50 1 Introduction

51 Carbonaceous aerosols (e.g., organic carbon (OC), elemental carbon (EC)/black carbon (BC), ~~and brown carbon (BrC)~~)
52 profoundly impact regional and global climate (Corbin et al., 2019; Lu et al., 2020; Rabha and Saikia, 2020). As an important
53 component of carbonaceous aerosols, BC can serve as a tracer of anthropogenic pollution once emitted from the incomplete
54 combustion of fossil fuels and biomass burning. Moreover, BC particles are generally soot-aggregated with ~~onion-like graphite~~
55 graphene-like layer microstructures which can be observed under electron microscopy (Adachi et al., 2019). BC and EC are
56 two components of carbonaceous aerosols that are measured differently. BC is typically quantified based on its light-absorbing
57 properties, while EC is measured using thermal-optical methods (Duarte et al., 2021). However, EC can also be referred to as
58 graphitic carbon or soot, with some overlap in their definitions. Another important component of carbonaceous aerosols, BrC,
59 represents a series of light-absorbing organic compounds, contributing significantly to the light absorption of atmospheric
60 aerosols (Wang et al., 2020b). BrC and BC show different light absorption patterns as a wavelength function. ~~BrC typically~~
61 ~~absorbs more strongly in the visible range (400–700 nm), while BC absorbs more strongly in the ultraviolet (UV) and near-~~
62 ~~infrared (NIR) ranges (200–400 nm and 700–1000 nm, respectively). Therefore,~~ BrC and BC can be distinguished by
63 measuring the absorption spectra of aerosol particles at different wavelengths (Andreae and Gelencsér, 2006; Bond et al., 2013;
64 Laskin et al., 2015; Li et al., 2020; Yus-Díez et al., 2021). Tar balls are commonly used as typical tracers of biomass and biofuel
65 burning due to their composition of amorphous carbon. These particles also belong to BrC ~~due to their light absorption~~
66 ~~properties over a broad range of solar spectra, from around 300 nm in the UV range to 2500 nm in the NIR range because they~~
67 are light absorbing organics (Adachi et al., 2019; Hand et al., 2005). Spherical tar balls emitted from biomass burning have
68 been observed in cases of both wild fire burning (Adachi et al., 2019) and laboratory generated tar ball particles (Tóth et al.,
69 2014). In the atmosphere, biomass burning produces a significant amount of tar balls, which are not deliquescent but can
70 absorb water at high relative humidity (RH = ~80%), thereby affecting their ability to scatter and absorb light (Hand et al.,
71 2005).

72 The optical properties of BC and BrC particles are affected by several factors including the emission source, coating
73 component, particle size, morphology, and mixing state of the particles (Wei et al., 2020). The BC configuration in the single

74 particles would influence their radiative effects (Luo et al., 2021). For example, core-shell BC particles show enhanced light
75 absorption compared to bare BC particles, especially when BC particles are coated with absorptive materials such as BrC
76 (Budhavant et al., 2020; Cappa et al., 2012; Shamjad et al., 2012; You et al., 2016). “Lensing effect” refers to the absorption
77 enhancement if BC is coated with non-absorbing organic or inorganic materials (Luo et al., 2021; Yang et al., 2009). In contrast,
78 if the BC coating materials are highly absorptive, no absorption enhancement may occur at shorter visible and UV wavelengths,
79 a phenomenon known as “shielding effect”. The shielding and lensing effects depend on the coating thickness over BC (Lack
80 and Cappa, 2010). When BC is well internally mixed with BrC, its total absorption enhancement becomes smaller than the
81 enhancements of not well mixed counterparts due to the absorptive coating that acts as a shield~~When BC is well internally~~
82 ~~mixed with BrC, its total absorption enhancement becomes smaller than those of not well mixed counterparts due to the~~
83 ~~absorptive coating that acts as a shield~~ (Feng et al., 2021). Moreover, it is impossible for BC and other materials to be
84 homogeneously distributed.

85 The extent to which ~~of~~ BC and BrC contribute to light absorption in atmospheric aerosols can be assessed using the
86 absorption Ångstrom exponent (AAE) (Wang et al., 2020a). AAE is a parameter used to quantify the spectral dependence of
87 aerosol light absorption. It is calculated by fitting a power-law relationship between the aerosol absorption and wavelengths
88 over a given spectral range. The AAE is used to identify sources and types of aerosols and a higher AAE value is associated
89 with sources such as biomass burning or urban pollution, while a lower AAE value suggests absorption by larger particles,
90 such as mineral dust or sea salt (Blanco-Donado, 2022; Duarte et al., 2021). However, many factors such as mixing state,
91 coating, particle size, refractive index, wavelength, and emission source, would affect the AAE values for BC and BrC aerosols,
92 leading to large variations among different studies (Moschos et al., 2021). For example, a previous study showed that the AAE
93 values derived from wavelengths of 405 and 781 nm are very sensitive to refractive index and particle diameter (Chylek et al.,
94 2019). The AAE values of 0.8–1.6 at 470 and 950 nm are attributed to traffic emissions and fuel combustion (Ezani et al.,
95 2021). Comparatively, those AAE values can be as large as 2.0 for ship emissions (Helin et al., 2021). Moreover, the
96 recommended AAE value for fossil fuel (FF) and biomass burning (BB) is 1 and 2 (or higher), respectively (Liu et al., 2023).
97 Other AAE values were also found in previous studies for FF (0.9) and BB (1.68) (Zotter et al., 2017) or FF (1.2) and BB (2.2)
98 as the mostly used optical pair (Milinković et al., 2021). The AAE values of 1.4 and 1.7 for BC and BrC were set to be the

99 lower and upper limits in the modelling study of biomass burning particles mixed with BC and BrC (Chylek et al., 2019).
100 However, the use of constant AAE values for calculating the BC fractions from BB and FF led to large uncertainties without
101 knowledge of the core size or coating thickness of the BC particles (Virkkula, 2021). Currently, the effect of the light absorption
102 is not well known for the carbonaceous particles in the marine atmosphere due to scarce ship-based measurements.

103 The optical properties of BC and BrC particles can also be investigated through fractal dimension (D_f) analysis based on the
104 fractal properties of BC aggregates. D_f illustrates how particles aggregate and grow and it can be determined through boxing
105 counting calculation, ensemble method, or soot parameter method with TEM images (Pang et al., 2022). The D_f values are
106 mainly related to emission sources and aging process of the particles. Previous studies showed that the D_f values of fresh BC
107 particles tend to be small but become larger after aging because the particles are more compact due to coatings (Luo et al.,
108 2022; Wang et al., 2017). D_f values of 1.8 and 2.6 were used respectively for fluffy and compact BC particles in a numerical
109 study to investigate the impact of the BC morphology on light absorption (Luo et al., 2021). Laboratory experiments simulating
110 wildfires showed that the D_f values of freshly emitted BC were in a range of 1.74–1.92 (Chakrabarty et al., 2006), compared
111 to the range of 1.67–1.83 from a field study of the Las Conchas fire (China et al., 2013). A similar range of the D_f values of
112 1.67–1.93 were found at a remote site in the southeastern Tibetan Plateau (Wang et al., 2017). The obtained D_f values for the
113 traffic emissions were as large as 3 (Wei et al., 2020). In addition to the emission sources and aging process, D_f is also dependent
114 on the particle size. A previous experimental study found that for polystyrene latexes (PSL) particles, the D_f values decreased
115 with the increase of particle size up to 200 nm (Wu et al., 2013b). Nevertheless, knowledge of the fractal dimension for
116 carbonaceous particles in the marine atmosphere is currently very limited, hindering our ability to understand the aging process
117 and the optical properties of these particles.

118 In the past years, carbonaceous aerosols in the marine atmosphere have been extensively studied on regional and global
119 scales, focusing on the transport of anthropogenic emissions to the sea areas. The BC background concentrations in Antarctic
120 and Arctic regions are below 20 ng m^{-3} (Fossum et al., 2022). The BC outflows from Asia to the Pacific Ocean exhibit seasonal
121 variations and originate from anthropogenic and biomass-burning sources in China, Siberia, and Southeast Asia (Matsui et al.,
122 2013). Ship-based BC and EC measurements reveal significant influence of continental transport on remote oceanic regions,
123 including the Bay of Bengal (Kedia et al., 2012), Indian Ocean (Kompalli et al., 2021), Southern Indian Ocean and the Southern

124 Ocean (Ueda et al., 2018), North Sea (Bencs et al., 2020), Antarctic (Chaubey et al., 2013; Schmale et al., 2019), North Pacific
125 (Taketani et al., 2016; Xing et al., 2014), Arctic (Pankratova et al., 2021; Sharma et al., 2019), Northeast Atlantic (Fossum et
126 al., 2022), the Yellow Sea (Kwak et al., 2022), and Western Pacific (Ma et al., 2022). However, to our knowledge, the BC mass
127 concentrations have been found to vary significantly across different oceans and seasons, with levels from 3 to 2800 ng m⁻³
128 and being influenced by anthropogenic activities and seasonal factors. The online BC measurements in the South China Sea
129 (SCS) region are limited. An early study reported BC concentrations on Yongxing Island during the rainy season (May 16–
130 June 20, 2008) and the dry season (Dec. 12, 2008–Jan. 8, 2009), with average concentrations of 0.54 and 0.67 μg m⁻³,
131 respectively (Wu et al., 2013a). Recent studies conducted at coastal sites in the SCS found that BC concentrations are strongly
132 impacted by the land anthropogenic emissions (Wang et al., 2022). The time-resolved BC concentration varies with the vertical
133 heights (Sun et al., 2020c) and the carbonaceous materials of OC and EC account for 31–62% in PM_{2.5} (Yan et al., 2018).
134 Nevertheless, quantification of the light absorption potential of BC and BrC aerosols remains challenging due to the limited
135 knowledge regarding the morphology, particle size, and mixing state of carbonaceous particles in the SCS (Kompalli et al.,
136 2021). Furthermore, the atmosphere in the SCS is typically influenced by the southwesterly monsoon from May to August
137 (Wang and Wu, 2020), which affects the air masses from Southeast Asia. In this study, we conducted ship-based measurements
138 of BC, OC/EC, and single particle sampling during summer (May 05–June 09, 2021) in the SCS. The morphology (i.e., the
139 fractal dimension and the size of the single BC particles) and light absorption properties of carbonaceous particles were
140 characterized. The source origins, relationships between the D_f and BC size, as well as the impact of summer monsoon on the
141 light absorption of the BC particles are discussed.

142 **2 Methods**

143 **2.1 Cruise route and instrumentation**

144 The cruise measurements were carried out from May 5 to June 9, 2021, covering a marine area of 11.9–24.5 °N and 111.1–
145 118.2 °E in the SCS. Single particles were collected on the TEM grids (3.05 mm I.D., copper meshed and covered with lacey
146 carbon film) located on the front deck during ship navigation and stop using a single-stage particle sampler (DKL-2, Genstar

147 Electronic Technology Co., Ltd., China) which is the same as other studies (Liu et al., 2021; Pang et al., 2022). The sampling
148 flow rate and time were set at 1 L min⁻¹ and 10 min, respectively, for each collection. The nozzle diameter of this single-cascade
149 impactor is 0.3 mm. The particles with aerodynamic diameters above 0.2 μm were collected with a collection efficiency of
150 50%, assuming a particle density of 1.5 kg m⁻³ (Marple and Olson, 2011). More details can be found in the supplementary
151 information (Section 1 of SI). The mixing state and morphology of the single particles were analyzed utilizing a transmission
152 electron microscope (TEM, FEI Tecnai G2 Spirit, Holland) operated at an accelerating voltage of 120 kV, in conjunction with
153 an energy dispersive spectrometer (EDS, Bruker Nano GmbH Berlin, Esprit 1.9, Germany) for elemental analysis. The
154 thickness of the EDS detector (type XFlash 5060) is 0.45 mm with a Si dead layer of 0.029 mm. Notably, in the EDS spectra,
155 when analyzing particle composition, Cu should be excluded, and a considerable level of C and Si should be observed in the
156 background signals due to the presence of Si in the detector, Cu and C in the TEM grid. The substrate holder of TEM was tilted
157 25° for thorough inspection during imaging and EDS analysis.

158 The sampling inlets were installed on the bow of the research vessel with a height of ~ 15 m above sea level. The own ship
159 emissions (e.g., engine, cooking, etc.) were exhausted from the chimney on the stern with a linear distance of ~ 22 m to the
160 inlets. The BC mass concentrations were measured by an aethalometer (Model AE33, Magee Scientific, USA) with a high
161 time resolution of ~~one second or~~ one minute (Drinovec et al., 2015). Note that the BC mass concentrations derived from AE33
162 are referred to as equivalent BC mass concentrations due to the light absorption of both BC and BrC at 880 nm. The sampling
163 air was regulated by a PM_{2.5} cyclone (BGI Inc., Waltham, MA, USA) and subsequently dried by a Nafion dryer (Model MD-
164 700 series, Perma Pure Inc., USA) with a relative humidity below 40% through the filter tape (type 8060) at a flow rate of 5 L
165 min⁻¹. Data corrections were made for the employed Aethalometer AE33, considering the multiple scattering parameters
166 (C(λ)=1.39) for the used filter type, the leakage factor (ζ=0.01), and the compensation parameters (K_{min}=-0.005, K_{max}=0.015).
167 The optical attenuation (ATN) was measured simultaneously from the two spots on the filter. The measured attenuation at
168 seven wavelengths (7 channels) is used to determine the wavelength-dependent absorption coefficient. The mass specific
169 absorption cross-sections (MAC_{g,air}) applied in the BC calculations were 18.47, 14.54, 13.14, 11.58, 10.35, 7.77, and 7.19 m²
170 g⁻¹ for wavelengths of 370, 470, 520, 590, 660, 880, and 950 nm, respectively (Ausmeel et al., 2020). The measured values at
171 880 nm (channel 6) are for black carbon concentration calculation, and at 370 nm (channel 1) for UV particulate matter (UVPM)

172 concentration (Drinovec et al., 2015). The detection limit of AE33 aethalometer is approximately $0.03 \mu\text{g m}^{-3}$ for 1-min
173 integration period and below $0.005 \mu\text{g m}^{-3}$ for 1-hour integration period. The instrument was automatically calibrated by zero
174 air every day. Notably, significant spikes were observed during periods when the ship was stationary, when it was travelling at
175 low speeds, and when the wind was blowing from the stern of the vessel.

176 The OC/EC concentrations were measured by a semi-continuous OC/EC analyzer (Model-4, Sunset Laboratory Inc., USA)
177 based on the optical attenuation and thermo-optical transmittance methods (Geron, 2009) under the NIOSH 5040 thermal-
178 optical protocol (Lappi and Ristimaki, 2017). Similarly, the sampling air passed through a $\text{PM}_{2.5}$ cyclone (BGI Inc., Waltham,
179 MA, USA) and was dried by a Nafion dryer (Model MD-700 series, Perma Pure Inc., USA) with a relative humidity below
180 40% at a flow rate of 8 L min^{-1} . The air then passed through a denuder for the removal of volatile organic compounds (VOCs)
181 and the particles were collected on the quartz filter with 45-min accumulation and 15-min analysis. The instrument was
182 calibrated with the standard sucrose solution as recommended. The manufacturer-claimed detection limits are 0.4 and $0.2 \mu\text{g}$
183 m^{-3} for OC and EC, respectively (Brown et al., 2019). However, several previous studies showed that these values may vary
184 substantially in a range of 0.04–2 and 0.001–0.5 $\mu\text{g m}^{-3}$ for OC and EC, respectively, due to the artifact of the quartz filters
185 (Bao et al., 2021; Bauer et al., 2012; Chen et al., 2017; Jung et al., 2011; Karanasiou et al., 2020; Park et al., 2018; Zhang et
186 al., 2021). Here, we estimated the detection limits of 0.15 and $0.012 \mu\text{g m}^{-3}$ for OC and EC based on 26 effective blank
187 measurements with 3 times standard deviation (3σ) during the campaign. The limit of detection (LOD) for OC and EC is 0.18
188 and $0.19 \mu\text{g m}^{-3}$, respectively, calculated as three times the standard deviation of replicate measurements of a standard sucrose
189 solution with a carbon content of $10.516 \mu\text{g m}^{-3}$. The Sunset OC/EC analyzer also measures optical EC based on the
190 transmission of 660 nm wavelength light through the quartz fiber filter employed for sampling, similar to the AE33 for optical
191 BC measurements. Optical EC is defined as the apparent EC on the filter based on the measured apparent absorbance and the
192 fixed absorption coefficient according to the user's manual of the Sunset OC/EC. Both our study and a previous study (Brown
193 et al., 2019) showed that the optical EC concentrations from Sunset were comparable with the BC concentrations from AE33.
194 Note that the resultant optical EC concentrations from the instrument output may be overestimated due to the limitation of the
195 filter-based optical measurements.

196

197 The measurements of solar radiation (SR), temperature (T), pressure (P), relative humidity (RH), relative wind direction
198 (RWD), and relative wind speed (RWS) were provided by the automatic weather station ([AWS430, Vaisala Inc., Finland](#)) ([Song
199 et al., 2022](#)) equipped on the front deck of the research vessel. This station comprises a range of integrated sensors, including
200 a wind speed and direction sensor (model WMT702), a temperature and humidity sensor (model HMP155), and an atmospheric
201 pressure sensor (model BARO-1). The cruise route for ship navigation is from the global positioning system (GPS) onboard
202 the ship (Scapath 330+, Kongsberg Inc., Norway). ~~equipped on the research vessel. The cruise route is from the global
203 positioning system (GPS) onboard the ship.~~

204 **2.2 Data analyses and processing**

205 ~~2.2 Data analyses and processing~~

206 **2.2.1 Analyses of single particles**

207 A total of 34 samples (15 during navigation and 19 at stop) were analyzed and more than 20 bright-field TEM images were
208 randomly captured for each sample except for those at the center of the grids where particles were easily overlapped. A total
209 of 15624 single particles were statistically analyzed to obtain morphology information (i.e., the Feret diameter, area, perimeter)
210 for each particle using the software Image-J (1.53q, National Institute of Health, USA) ([Cheng et al., 2021](#)). In the analysis of
211 particle size, the Feret diameter is defined as the distance between the parallel tangential lines that constrain the particle
212 perpendicularly. In this study, we applied the Feret diameter as the longest distance between any two points along the boundary
213 of the selected particles. Moreover, we utilized “geometrical diameter” to describe the size of tar balls with circular shape,
214 which signifies the distance between two points located on the surface of a geometric shape, with this line passing through the
215 shape's center. Using “Geometrical diameter” is suitable to quantify the size of the observed tar balls which excluded any
216 coatings or additional materials. Specifically, we employed TEM data acquisition software to measure the geometrical
217 diameters of observed tar balls. The D_f values of the BC particles were estimated using the boxing counting method using the
218 plugin Fraclac. An example was given in the SI (section 2, Fig. S1) to show how to calculate D_f using the software. A detailed
219 description of the procedure using the boxing counting method and the software ImageJ can be found in the SI. The D_f values
220 are very sensitive to the fill extent and sizes of the particles. A previous study showed low fractal dimensions when the particles

221 contain void volumes (Peyronel et al., 2010).

222 The own ship emissions can be identified using various measures, for example, high CO, NO_x concentrations (Sun et al.,
223 2020b), high BC concentrations (Alroe et al., 2019; Shank et al., 2012), regular cooking emissions (Cai et al., 2020), and wind
224 speeds/directions between the ship stop and start operation (Ausmeel et al., 2020; Kwak et al., 2022). The contribution of ship
225 emissions to BC sources on the marine atmosphere depends on engine types, operation modes, fuel types, and loadings (Gagne
226 et al., 2021; Jiang et al., 2018; Karjalainen et al., 2022; Lack and Corbett, 2012; Wu et al., 2021; Zhao et al., 2020). Here, we
227 classify two sampling modes (navigation vs stop) of single particle analyses according to ship operation, and relative wind
228 direction/speed. In this study, the relative wind direction/speed is relative to the ship heading. The navigation mode is
229 constrained by the relative wind direction of 0–80° or 280–360°, and the relative wind speed greater than 5 m s⁻¹, averaged
230 for every 10 minutes (consistent with the collection time of TEM samples). The stop mode is set with the relative wind direction
231 of greater than 80° and less than 280°, or the relative wind speed lower than 5 m s⁻¹. The navigation mode samples are mainly
232 from marine air and air masses of long-range transport while the stop mode collected air masses which are mixed with the own
233 ship emissions. The wind direction (speed) and relative wind direction (speed) are calculated by Eq. (1) (Aijjou et al., 2020).

$$234 \quad \underline{V_R = \sqrt{V_s^2 + V_w^2 + 2 * V_s * V_w * \cos \alpha}} \quad (1)$$

235 where V_R is the relative wind direction (speed), V_s is the ship direction (speed), V_w is the true wind direction (speed), α is the
236 angle between the ship heading and the true wind direction.

237 The temporal profiles of ship heading directions, and relative wind direction/speed are shown in the SI (section 3, Fig. S2).
238 Details of the two sampling modes (navigation vs stop) on a vector average of 10 minutes are listed in Table S1 and Fig. S3.

239 Here, we distinguished the own ship emissions (research vessel) from those of other ships or long-range transport based on
240 the following criteria: low relative wind speed (< 5 m s⁻¹), relative wind direction encompassing ship exhaust (80–280°), and
241 a substantial AE33-derived hourly averaged BC mass concentration (>2 μg m⁻³). Other ship emissions far from the research
242 vessel are treated as a part of the transported air masses in this study.

243 ~~Here, we distinguished the own ship emissions (ship pollution) from those of other ships or from long-range transport.~~
244 ~~In addition, an hourly averaged BC concentration above $2 \mu\text{g m}^{-3}$ was also considered as ship pollution based on the~~
245 ~~time series of BC concentration without significant interference of self emission.~~

246 2.2.2 BC, OC, EC and optical EC data

247 In this study, BC data obtained from the AE33 instrument are referred to as BC, while data from the OC/EC analyzer is
248 expressed as thermal OC, thermal EC, and optical EC. Here, we averaged the BC mass concentrations over one minute and
249 excluded those below the detection limit of $0.03 \mu\text{g m}^{-3}$ ~~for-to~~ minimizing the variations. We also removed the own ship
250 emissions which are characterized by spikes in particle number concentrations according to the wind directions (Fossum et al.,
251 2022). BC mass concentration was calculated using Eqs. (2, 3) which are cited from the AE33 aethalometer user's manual (Ver
252 1.54).

$$253 \text{-----} ATN = -100 * \ln (I/I_0) \text{-----} (2)$$

254 where ATN is optical attenuation, I_0 is reference signal, I is spot signal.

$$255 \text{-----} BC = \frac{S*(\Delta ATN_1/100)}{F_1(1-\zeta)*\sigma_{air}*C*(1-k*ANT_1)*\Delta t} \text{-----} (3)$$

256 where BC is black carbon concentration, S is spot area, F_1 is measured flow, ζ is leakage factor, σ_{air} is the mass absorption
257 cross-section (MAC), C is multiple scattering parameter, k is compensation parameter, and t is time.

258 AAE was calculated according to Eq. (4) using the light absorption at wavelengths of 470 and 950 nm, which are built-in
259 algorithms in the AE33 aethalometer as described elsewhere (Helin et al., 2021; Kang et al., 2022; Milinković et al., 2021;
260 Zotter et al., 2017). This method serves as a two-composition source apportionment for BC emitted from fossil fuels and
261 biomass burning (AAE model).

262 , which applied AAE=1 for fossil fuel and AAE=2 for biomass. The calculations for BC(BB) and BC(FF) are shown in Eqs.
263 (5–7) which are referred to the AE33 aethalometer user's manual and publication (Sandradewi et al., 2008). The optical
264 absorption coefficient is the sum of biomass burning and fossil fuel burning contributions. Basic equations are using Beer-
265 Lambert's Law.

$$AAE = - \frac{\ln \frac{\sigma_{abs}(\lambda_1)}{\sigma_{abs}(\lambda_2)}}{\ln \frac{\lambda_1}{\lambda_2}} \quad (4)$$

where σ_{abs} is aerosol absorption coefficient, σ_{air} is mass absorption cross-section (MAC), $\sigma_{abs} = BC * \sigma_{air}$. $\lambda_1 = 470$ nm and $\lambda_2 = 950$ nm.

$$\frac{\sigma_{abs}(470 \text{ nm})_{FF}}{\sigma_{abs}(950 \text{ nm})_{FF}} = \left(\frac{470}{950}\right)^{-AAE_{FF}} \quad (5)$$

$$\frac{\sigma_{abs}(470 \text{ nm})_{BB}}{\sigma_{abs}(950 \text{ nm})_{BB}} = \left(\frac{470}{950}\right)^{-AAE_{BB}} \quad (6)$$

$$\sigma_{abs}(\lambda) = \sigma_{abs}(\lambda)_{FF} + \sigma_{abs}(\lambda)_{BB} \quad (7)$$

where $\sigma_{abs}(470 \text{ nm})_{FF}$ and $\sigma_{abs}(950 \text{ nm})_{FF}$ are the aerosol absorption coefficients at wavelengths of 470 and 950 nm for fossil fuel (FF), $\sigma_{abs}(470 \text{ nm})_{BB}$ and $\sigma_{abs}(950 \text{ nm})_{BB}$ are the aerosol absorption coefficients at wavelengths of 470 and 950 nm for biomass burning (BB), AAE_{FF} and AAE_{BB} are equals to 1 and 2, respectively.

$$AAE = - \frac{\ln \frac{\sigma_{ap}(\lambda_1)}{\sigma_{ap}(\lambda_2)}}{\ln \frac{\lambda_1}{\lambda_2}} \quad (1)$$

where σ_{ap} is the aerosol absorption coefficient, $\sigma_{ap} = BC * MAC$. $\lambda_1 = 470$ nm and $\lambda_2 = 950$ nm.

Alternatively, AAE can be obtained from the negative slope of linear regression between the log-transformed $\sigma_{abs} \sigma_{ap}$ and all the wavelength spectra so that hourly AAE values (all λ) can be obtained following a similar method in Retama et al. (2022).

Details are shown in the SI (Section 4, Figure S4). Here, we define Delta-C as the difference between the concentration derived from the aforementioned UVPM data (at 370 nm) and BC concentration (at 880 nm). This Delta-C parameter was employed as an indicator of smoke from biomass burning in previous wood biomass burning studies (Harrison, 2020; Zhang et al., 2017).

The OC and EC (thermal) concentrations lower than the detection limits (0.15 and 0.012 $\mu\text{g m}^{-3}$ for OC and EC, respectively) were excluded. Additional data were removed for those with laser correction factors below 0.88 and calibration peak areas lower than the initial calibration levels (within 10%), and a total of ~~550~~ 551 h data were used for further analysis. In comparison, the Sunset optical EC (at 660 nm) is generally consistent with the Magee AE33 aethalometer derived BC (at 880 nm) within 9% (Brown et al., 2019) which is shown in Section 3.3. The EC concentration data from Sunset were considered as ship

288 pollution and were discarded when the BC concentrations from the AE33 aethalometer were higher than $2 \mu\text{g m}^{-3}$, in addition
289 to those with relative wind directions of $80\text{--}280^\circ$ regardless of the BC concentrations as mentioned before.

290 2.2.3 HYSPLIT backward trajectory and MODIS fire data

291 The backward trajectories were calculated using NOAA HYSPLIT (Hybrid Single-Particle Lagrangian Integrated Trajectory)
292 (Version 5) at heights of 100, 500 and 1000 m above ~~the~~ sea level (AGL). Daily meteorological data with $1.0^\circ \times 1.0^\circ$ spatial
293 resolution for trajectory calculation were downloaded from the global data assimilation system (GADS)
294 (<ftp://arlftp.arlhq.noaa.gov/pub/archives/gdas1/>). Here, we calculated the 72-h back trajectories of air masses arriving at the
295 single particles sampling sites along the cruise route.

296 Moderate Resolution Imaging Spectrometer (MODIS) data are available from the Near real-time MODIS Collection 6
297 products (<https://firms.modaps.eosdis.nasa.gov/download/>). Here, we selected a region of $102\text{--}127^\circ\text{E}$ and $0\text{--}30^\circ\text{N}$ fully
298 covering the campaign area. The number of fire hotspots was counted each day during the campaign with a confidence level
299 of higher or equal to 80% as recommended (Giglio et al., 2020). ~~Detailed~~ A detailed description of the fire detection algorithms
300 is available online (<https://earthdata.nasa.gov/what-is-new-collection-6-modis-active-fire-data>).

301 3 Results and discussion

302 3.1 Overview

303 **Figure 1** shows the time series of ship cruise route and the single particle sampling locations during ship navigation (marked
304 as solid triangles) and stop (marked as open squares) over the SCS during the campaign. The cruise sequences are $\text{AB} \rightarrow \text{C} \rightarrow$
305 $\text{D} \rightarrow \text{EB} \rightarrow \text{D} \rightarrow \text{A}$, with AB and EB being non-stop cruise, otherwise the ship stopped occasionally along the arrow routes for
306 other research tasks. **Figure 2** shows the time series of the meteorological variables (i.e., solar radiation (SR), temperature,
307 pressure, relative humidity (RH), wind direction (WD), and wind speed (~~WSWS~~)) during the whole campaign (May 05–June
308 09, 2021). The measurements were conducted mostly under-on sunny days prior to June 02 as shown by the SR data.
309 Subsequently, it became rainy and cloudy due to the summer monsoon in the SCS. One notable meteorological feature during

310 the campaign was the occurrence of the summer monsoon starting from May 27 close to the site *E*, during which (May 27–
311 June 01), an increase of RH (~9% from campaign-averaged 78.7% to monsoon period-averaged 85.6%) and a slight decrease
312 of pressure (~0.2% from 1007.4 to 1005.4 hPa) were observed. Meanwhile, the wind directions were mainly southerly during
313 this period and later changed to southwest.

314 It should be noted that Typhoon 202103 (CHOI-WAN) travelled across B→D, resulting in a bulge in the middle of the
315 cruise route ~~for avoiding to avoid~~ the typhoon during June 03–05, 2021. The typhoon track is available online with the last
316 accessed date Mar. 25, 2023: <http://agora.ex.nii.ac.jp/digital-typhoon/summary/wnp/s/202103.html.en>. The typhoon was
317 initiated at 02:00 local time on May 31 and dissipated at 14:00 on June 05, 2021 (Figure S5). It passed over our cruise route
318 from June 03 to June 05, 2021. While no significant increase of absolute wind speed was seen in Figure 2, a significant increase
319 of relative wind speed was shown in Figure S2, along with an obvious decrease of atmospheric pressure during the typhoon
320 period (Figure S5). The measured relative humidity increased from May 27 to June 01, prior to the presence of the typhoon,
321 which can be attributed to the decrease of ambient temperature during this period.

322 **Figure 3** shows the time series of the mass concentrations of carbonaceous aerosol components in PM_{2.5} (i.e., BC, UVPM,
323 OC, and EC) during the whole campaign. Frequently high spikes of the mass concentrations of carbonaceous aerosol
324 components were observed due to the ship pollution ~~from the research vessel~~. We notice that ship pollution was significantly
325 pronounced on the first two days after the ship left the harbor and on the last 3–4 days before the ship returned to the harbor,
326 during which the spikes of BC and UVPM concentrations were measured by the Magee AE33 aethalometer with the relative
327 wind direction of 80–280°. Before May 08 and after June 05, higher UVPM, OC, and EC concentrations were observed, which
328 can be attributed to significant fresh ship emissions from the research vessel, as evidenced by simultaneous higher BC
329 concentrations. Similar spikes in BC concentrations were observed during other measurement periods, either preceding or
330 following the monsoon period, which were caused by emissions from the frequent stops and starts of the ship. Note that no
331 significant diurnal trend for OC was observed during those aforementioned periods.

332 Figure 4 shows the time series of fire spots distribution and the 72-h backward trajectories at three AGLs (100, 500 and
333 1000 m) over the SCS during the campaign. Only several backward trajectories are shown to avoid massive overlapping.
334 Several fire spots located in the sea were attributed to oil or natural gas drilling processes which generate thermal energy,

335 combustion, and exhaust. Such processes included the prevalent hydrocarbon exploration and production activities in this
336 region. A comprehensive cartographic representation of these endeavors within the SCS can be accessed via the online
337 platforms (<https://amti.csis.org/south-china-sea-energy-exploration-and-development/>). Note that since the ship moved along
338 the cruise route, the air mass backward trajectories also changed with the movement of the ship. For example, significant
339 biomass burning was detected in Laos, northern Vietnam, and the Philippines during May 15–24, as indicated by the
340 corresponding fire spots. However, the back trajectories to the sampling route (C→D) during this period were mainly from the
341 Philippines.

342 Here, we classified the campaign period into several groups based on the cruise route, change of wind direction during
343 monsoon, backward trajectories, and ship pollution, as listed in Table 1: (1) BMP-1 (before monsoon period 1), AB route
344 mainly with northeast wind direction during May 05–09; (2) BMP-2, B→C→D route close to the Philippines primarily with
345 southeast wind direction during May 10–22; (3) BMP-3, D→E close to mainland China with the same wind direction as BMP-
346 2 during May 23–26; (4) TMP (transition monsoon period), EB route with south wind direction during May 27–Jun 01; (5)
347 AMP (after monsoon period), B→D→A route with southwest wind direction during June 02–09; (6) SPP (ship pollution
348 period), ~35% of the online measurement data could be attributed to this category in this study due to the interference from
349 the research vessel own emissions.

350 **3.2 Single particle analysis of BC and tar balls**

351 Particle size distribution, composition, and size-dependent BC fractals were investigated based on TEM images. The Feret
352 diameter is commonly used in microscopy for particle size analysis (Zefirov et al., 2018). The size distribution of all the single
353 particles from the analyzed TEM images is depicted in Fig. 4 Fig. 5. The distribution is represented with histograms starting
354 at 50 nm, a width interval of 20 nm, and a bin number of 200. The choice of bin width may vary depending on cases but it is
355 close to the quotient value of the square root of the measured particle number divided by the overall width of the distribution
356 (Pabst and Gregorova, 2007). Moreover, lognormal fitting is used for the peak size identification of particle size distribution
357 (Rice et al., 2013). Figure 5a~~Figure 4a~~ shows a fitted peak Feret diameter of 307 nm for a total of 6613 particles from 15
358 samples during navigation, while a fitted peak Feret diameter of 325 nm was obtained for a total of 9011 particles from 19

359 samples during stop ([Fig. 5b](#)~~Fig. 4b~~). Note that we could not successfully obtain a bimodal or multi-peak fit for the data of the
360 stop cases using multi-peak fitting function in the Igor Pro software, as shown in Figure S6. Hence, we believe that single peak
361 fitting best described the distribution in our stop cases, as illustrated in Figure 5. Particles collected during navigation were
362 predominantly aged at high wind speeds, while particles during stop were mixed significantly with freshly emitted particles
363 from the own ship and from other merchant ships or those from long-range transport, possibly leading to the variation of the
364 size distribution. Although the bimodal distribution was observed from particles in the indoor air, which was likely caused by
365 fresh emissions and secondary formation (Pipal et al., 2021), we did not obtain significant bimodal peaks for both navigation
366 and stop particles.

367 We obtained characteristic values for the particle shape descriptors such as circularity (0.7 ± 0.2) and aspect ratio (1.2 ± 0.3)
368 for all the particles collected during navigation and stop, implying that these particles are not perfectly spherical and may vary
369 in their mixing states. ~~Figure 6~~Figure 5 (top images) shows a comparison of the mixing states during navigation (a-c) and stop
370 (d-f) from typical BC TEM images. The BC particles collected during navigation are in the embedded (a), external (b), or core-
371 shell (c) mixing states classified with the methods which are based on single particle analysis of island and mountain samples
372 across East China Sea and Japan (Adachi et al., 2014; Sun et al., 2020a). More TEM images for the heavily coated internal
373 and external BC particles from navigation can be found in the SI (section 47, [Fig. S7](#)~~Fig. S3~~). The EDS analysis showed that
374 the single particles during navigation were predominantly composed of carbon (C), oxygen (O), sulfur (S), potassium (K),
375 sodium (Na), chloride (Cl), magnesium (Mg), and calcium (Ca) ([Fig. S8](#)~~Fig. S4~~), indicating that those BC particles were coated
376 with sulfate, sea salt, and organics. Furthermore, small externally mixed BC particles can be transported over the sea and easily
377 coated during long-range transport. Under the TEM electron beam, these coated volatile components were easily vaporized to
378 expose the BC fractal frame ([Fig. S7d-f](#)~~Fig. S3d-f~~).

379 Comparatively, a mixture of aged BC particles and much larger fresh BC particles as well as smaller scattered BC particles
380 during stop were found (Fig. 6d-f), which were likely emitted from other ships (Fig. 6d) and the research vessel (e, f). These
381 TEM images showed that the compressed BC particles are typically more aged and atmospherically processed, while the fractal
382 BC particles are fresh. Moreover, EDS analysis showed that sulfate formed from aqueous processes and less viscous organic
383 coating indicate an aging process. Those BC particles with Feret diameters larger than 2 μm during stop were fractal aggerates

384 ~~which could unlikely survive due to deposition during long-range transport. In addition, heavily coated internal BC particles~~
385 ~~were found during stop due to the mixing between ship pollution and the marine air (Fig. S9). Moreover, such particles could~~
386 ~~also be condensation of organics during the cooling process after they were emitted from the ship engine. Comparatively, much~~
387 ~~larger BC particles as well as smaller scattered BC particles during stop were found (Fig. 5d-f), among which were apparently~~
388 ~~emitted from the own ship (e, f). Those BC particles with Feret diameters greater than 2 μm during stop were fractal aggregates~~
389 ~~which unlikely survived during long-range transport due to deposition. In addition, heavily coated internal BC particles were~~
390 ~~found during stop due to the mixing between ship pollution and the marine air (Fig. S5). The bottom panels of Fig. 6-Fig. 5 (a-~~
391 ~~f) show the corresponding images obtained by boxing counting in fractal analysis with the resultant D_f , Feret diameter (D),~~
392 ~~Lacunarity (L), and sampling number underneath for each TEM sample image. Figure 7-Figure 6 shows D_f and L as a function~~
393 ~~of D for some representative BC particles during both navigation and stop. The BC particles showed narrower Feret diameters~~
394 ~~(229–2557 nm) during navigation than those (78–2926 nm) of BC from the own ship during stop. These particles were exposed~~
395 ~~to electron beam and volatile coatings were removed so that the morphology of bare BC was clearly shown regardless of the~~
396 ~~mixing state of the original BC. The D_f values during navigation were in a range of 1.281–1.77 with a median of 1.651.61,~~
397 ~~while the D_f values during stop were 1.43–1.761.48–1.75 with a median of 1.6661, indicating no significant differences of D_f~~
398 ~~($<4\%$) for the exposed BC particles during navigation and stop. Note that the particles in Figure 7 include pure BC and BC~~
399 ~~without thick coatings. These particles were exposed to the electron beam and volatile coatings were removed so that the~~
400 ~~morphology of BC was clearly shown regardless of the mixing state of the original BC particles (Figure S7). Most BC particles~~
401 ~~were below 1 μm in Feret diameter during navigation (Figure 7), while their sizes cover a wide range below 3 μm during stop,~~
402 ~~implying that the aged BC particles become smaller after long-range transport. Despite only a total of 134 BC data points~~
403 ~~shown in Figure 7, the results are still statistically meaningful due to the wide range of BC sizes covered in our analysis. Note~~
404 ~~that the size change of a BC particle cannot be determined because the original size of the particle is unknown before the~~
405 ~~removal of the coatings. Figure 6 also shows that most of the BC particles during navigation were below 1 μm in Feret diameter,~~
406 ~~implying that smaller BC particles were more susceptible to be coated and aged during transport. Comparatively, the~~
407 ~~lacunarities during navigation (0.3234–0.8290, median: 0.53) and stop (0.3534–0.9277, median: 0.59) were different, with the~~
408 ~~former being smaller than the latter, indicating that the lacunarity tended to become smaller ($\sim 10\%$) after coating or aging of~~

409 the BC particles.

410 ~~Tar balls were frequently observed during the campaign with an estimated sample fraction of about 11.8%. Fractal-like tar~~
411 ~~ball aggregates were usually found in wildfire smokes (Giroto et al., 2018); however, in this study, spherical tar ball particles~~
412 ~~were observed in the marine atmosphere and were mixed with sea salt (Fig. 8a and d for TEM image and EDS spectrum,~~
413 ~~respectively), organic carbon and sulfate (Fig. 8b and e) from the samples collected on May 27 during navigation. In contrast,~~
414 ~~the particles collected on June 01 were found to be amorphous carbon agglomerates (Fig. 8c and f) which were referred to OC.~~
415 ~~Tar balls were frequently observed during the campaign. The spherical tar ball particles in the marine atmosphere were mixed~~
416 ~~with sea salt (Fig. 7a and d for TEM image and EDS spectrum, respectively), organic carbon and sulfate (Fig. 6b and e) from~~
417 ~~the samples collected on May 27 during navigation. In contrast, the particles collected on June 01 were found to be amorphous~~
418 ~~carbon agglomerates which were mixed with sulfate (Fig. 6c and f). During these days, the wind directions were from [the](#)~~
419 ~~southwest, with air masses originating from both the Philippines and Southeast Asia. [The shape difference between the tar ball](#)~~
420 ~~spheres and the amorphous carbon agglomerates may be related to the type of biomass burning or the origin of the ship~~
421 ~~engines. The difference between tar ball spheres and amorphous carbon agglomerates may be related to biomass burning type~~
422 ~~or source origin.~~ Similar particle morphologies were found in other studies on brown carbon during aircraft measurements
423 over the Yellow Sea in 2001 (Zhu et al., 2013). Tar balls mixed with BC during stop were also observed ([Fig. S10](#)~~Fig. S6~~),
424 with geometrical diameters of 160–420 nm, much larger than nano-soot spheres (40–50 nm) ([Fig. S11](#)~~Fig. S7~~). In comparison,
425 the laboratory-generated tar balls were measured to have AAE values of 2.7–3.4, with an average of 2.9 at 467–652 nm (Hoffer
426 et al., 2016).

427 **~~3.3 Light absorption of carbonaceous aerosols~~[Air mass trajectories and mass concentrations of carbonaceous aerosols](#)**

428 ~~Figure 8 shows the time series of fire spots distribution and the 72-h backward trajectories at three AGLs (100, 500 and 1000~~
429 ~~m) over the SCS during the campaign. Only several backward trajectories are shown to avoid massive overlapping. Several~~
430 ~~fire spots located in the sea were attributed to oil or natural gas drilling processes which generate thermal energy, combustion,~~
431 ~~and exhaust. Such processes included the prevalent hydrocarbon exploration and production activities in this region. A~~
432 ~~comprehensive cartographic representation of these endeavors within the SCS can be accessed via the online platforms~~

433 (<https://amti.esis.org/south-china-sea-energy-exploration-and-development/>). Note that since the ship moved along the cruise
434 route, the air mass backward trajectories also changed with the movement of the ship. For example, significant biomass burning
435 was detected in Laos, northern Vietnam, and the Philippines during May 15–24, as indicated by the corresponding fire spots.
436 However, the back trajectories to the sampling route (C→D) during this period were mainly from the Philippines.

437 Here, we classified the campaign period into several groups based on the cruise route, change of wind direction during
438 monsoon, backward trajectories, and ship pollution, as listed in Table 1: (1) BMP 1 (before monsoon period 1), AB route
439 mainly with northeast wind direction during May 05–09; (2) BMP 2, B→C→D route close to the Philippines primarily with
440 southeast wind direction during May 10–22; (3) BMP 3, D→E close to mainland China with the same wind direction as BMP
441 2 during May 23–26; (4) TMP (transition monsoon period), EB route with south wind direction during May 27–Jun 01; (5)
442 AMP (after monsoon period), B→D→A route with southwest wind direction during June 02–09; (6) SPP (ship pollution
443 period), ~35% of the online measurement data could be attributed to this category in this study due to the interference of the
444 research vessel self-emission.

445 The BC concentrations measured by the Magee AE33 aethalometer agree excellently with the optical EC concentrations
446 obtained from the Sunset OC/EC analyzer, as evidenced by a linear regression coefficient of 0.97. The BC measurements
447 obtained from the AE33 instrument do not agree with the OC, EC values, yet their overall trends exhibit consistency. However,
448 the BC concentrations were considerably higher than the thermal EC concentrations, exhibiting linear regression coefficients
449 of 1.66 and 1.55, respectively. These findings, presented in Fig. S12 of Section 8 in the SI Fig. S8 of Section 5 in the SI, are in
450 line with previous research conducted by Brown et al. (2019). The OC/EC ratios can be used as an indicator for the source
451 origins of the air masses. Figure 9 shows the distribution of the OC/EC ratios and the corresponding EC concentrations. Figure
452 9 shows ~~†~~ The median OC/EC ratios ~~{are~~ 8.14, 5.20, 6.35 and 2.63 ~~}~~ for the classified periods ~~(BMP, TMP, AMP, and SPP,~~
453 ~~respectively)~~. Notably, EC median mass concentrations (0.24, 0.25 and 0.17 $\mu\text{g m}^{-3}$) for the marine air masses during BMP
454 (0.013–0.69 $\mu\text{g m}^{-3}$), TMP (0.015–0.60 $\mu\text{g m}^{-3}$), AMP (0.014–0.74 $\mu\text{g m}^{-3}$) were lower than the median concentration (1.70 μg
455 m^{-3}) during SPP. Compared with Figure 9d, the scattered higher OC/EC ratios in Figure 9a/b/c are caused by the very low EC
456 concentrations. The presence of extremely low EC concentrations, often falling below or near the detection limit, can introduce
457 discrepancies in the calculation of the OC/EC split, ultimately resulting in inaccurate EC concentrations (Bauer et al., 2009).

458 In addition, this study revealed a significant variation in EC concentrations during SPP, ranging from 0.15 to 22.8 $\mu\text{g m}^{-3}$.
459 Previous studies showed that OC/EC ratio could be characterized by various sources, ranging from 1.37–1.71 for residential
460 cookstoves, 1.63–2.23 for rural emissions, 1.05–1.24 for diesel exhaust, and 0.80–1.12 for urban environments (Khan et al.,
461 2012). A low OC/EC ratio (<3) corresponded to the predominant contribution of the primary OC in submicron particles
462 reported in a previous study in the Southern Indian Ocean, Northern Indian Ocean and Bay of Bengal (Neusüß, 2002). Here,
463 the median OC/EC ratio of 2.63 during SPP is much higher than the characteristic values of diesel combustion, most likely
464 because the sample air during SPP is composed of marine air and the own ship exhaust. Our results are consistent with a recent
465 study which showed that the diesel combustion from ships accounted for 15% of BrC in the total light absorption at a coastal
466 site in Shanghai during June–July, 2021 (Kang et al., 2022). In contrast, the OC/EC ratios during other periods (i.e., BMP,
467 TMP and AMP) were even much higher (5.20–8.13), indicating that the aerosols were highly aged during the long-range
468 transport of biomass burning aerosols. This is also consistent with our recent study in the SCS which showed that during
469 monsoon periods in the summer of 2019. The biomass burning organic aerosols became aged through atmospheric processes
470 during transport (Sun et al., 2023).

471 **3.5 Light absorption of carbonaceous aerosols**

472 The long-range biomass burning transport affects the air mass in the SCS. **Figure 10** illustrates the wavelength-dependent
473 mass concentration measured by the AE33 aethalometer during the campaign, showing (a) an example of a ship plume, and
474 (b, c) two significant biomass burning events during BMP (BB-1: 6:00–7:00 on May 15 and 15:00–22:00 on May 16) and
475 during TMP (BB-2: 15:00 on May 30–00:00 on May 31). The ship plumes, predominantly emitted from fossil fuel combustion,
476 showed similar absorption at all seven wavelengths. In contrast, significant absorption at low wavelengths was detected during
477 the biomass burning events, a phenomenon also observed in other field measurements in urban cities and towns where air
478 masses were susceptible to biomass burning (Zhang et al., 2017). A comparison of the two methods for AAE calculation is
479 presented in the SI ([Section 4, Fig. S4](#)[Section 6, Fig. S9](#)). The fitting results demonstrate that the AAE calculated for all
480 wavelengths was lower than the AAE calculated for only 470 and 950 nm wavelengths. The fitting slope is 0.78, and the
481 determination coefficient (R^2) is 0.98, indicating a strong correlation between the two methods.

482 **Figure 11** shows the hourly averaged AAE derived from all wavelengths as a function of the BC concentrations by AE33
483 aethalometer with the median (range) AAE values of 1.14 (0.57–1.48), 1.02 (0.51–1.36), 1.08 (0.54–1.42), and 1.06 (0.65–
484 1.37), respectively, for the classified periods (BMP, TMP, AMP, and SPP), and the corresponding BC median (range) mass
485 concentrations of 0.28 (0.033–1.17), 0.14 (0.042–2.86), 0.17 (0.055–1.08), and 3.01 (0.21–36.5) $\mu\text{g m}^{-3}$, respectively. Like
486 EC, ship pollution led to emissions of high BC concentrations, reaching as high as 36.5 $\mu\text{g m}^{-3}$. The median BC concentrations
487 decreased significantly during TMP and AMP, likely due to the increase of the RH which further increased the scavenging of
488 the BC particles during navigation as reported previously (Girach et al., 2014). Note that the biomass burning events were
489 excluded from the above calculations and are further discussed below.

490 During the biomass burning events, the correlations of AAE with AE33 BC and Delta-C concentrations are respectively
491 shown in **Figs. 11 and 12**, characterized by very high median AAE values (1.85 and 1.86, respectively for BB-1 and BB-2) and
492 BC concentrations (1.93 and 1.67 $\mu\text{g m}^{-3}$). The BC mass concentration ranged from 1.45 to 3.62 $\mu\text{g m}^{-3}$ during biomass burning
493 events based on light absorption at wavelength of 880 nm. The mass concentration in Figure 10 corresponds to BC mass
494 concentration obtained at each wavelength. We have emphasized that BC mass concentration in this study is equivalent BC at
495 individual wavelength. Notably, efficient light absorption of BrC in the range at 370–660 nm was observed during the biomass
496 burning events, while no significant wavelength-dependent BC concentrations were found during the own ship pollution (Fig.
497 10a). The AAE values below 1 in Figure 11 are not noises, in some cases due to aerosols from fossil fuel (Ezani et al., 2021)
498 and in other cases, they can be even lower than 0.5 when paired with wavelengths of 470 and 660 nm (Laing et al.,
499 2020). Notably, high BC (1.45–3.62 $\mu\text{g m}^{-3}$) mass concentrations at 370–660 nm were observed during the biomass burning
500 events due to more efficient light absorption of BrC in the range than at higher wavelengths, while no significant wavelength-
501 dependent BC concentrations were found during ship pollution (Fig. 10a). The higher AAE values imply much stronger
502 absorption of non-BC light-absorbing particles (BrC) at shorter (UV-vis) wavelengths, which mainly originated from the
503 biomass burning emissions (Ponczek et al., 2022). Moreover, the median OC/EC ratios were 5.03 and 5.29 respectively for the
504 two biomass burning events, even much higher than those for SPP (**Fig. 9**). The 72-h backward trajectories also showed that
505 the BB-1 air masses mainly originated from the Philippines while the BB-2 air masses were from the mainland Vietnam, both
506 with high densities of fire spots (**Fig. 4Fig-8**). The AAE values were also highly correlated with the Delta-C values with a

507 determination coefficient (R^2) of 0.92 (Fig. 12), further demonstrating a significant contribution of BrC to the AAE
508 enhancement. In addition, we further correlated the observed high AAE values with the Delta-C values for the two biomass
509 burning events and confirmed that these high AAE values (1.45–3.62) were indeed attributed to biomass burning rather than
510 ship emitted tar balls which covered an AAE range of 2.7–3.7 at 405 and 781 nm wavelengths in a previous wood burning
511 study (Chylek et al., 2019).

512 Our study found that the AAE values from all wavelengths for the marine atmosphere and ship pollution were 1.02–1.14
513 and 1.06, respectively, except for a higher AAE value (1.93) during the two biomass burning events. The AAE values for ship
514 pollution are dependent on the fuel types and loading conditions (Laskin et al., 2015). For example, heavy fuel oil operated at
515 high loads can result in AAE values (at 470/950 nm, and hereafter unless specified) of 2.0, while the intermediate fuel oil has
516 an AAE value of 1.3 at high loads (Helin et al., 2021). In addition, the presence of tar balls may contribute to the enhancement
517 of BrC absorption as mentioned in Section 3.2, as tar balls from ship emission have higher AAE values (2.5–6 depending on
518 the wavelengths) (Corbin et al., 2019). The occurrence of tar balls in this study was about 12% in the analyzed single samples.
519 These tar balls were likely aged during long-range transport from biomass burning and hence affected the light absorption of
520 BrC in the SCS.

521 **3.63.4-BC sources from fossil fuel vs biomass burning**

522 The source origins of BC particles can be investigated using the AAE model. In the model, we employed respectively the
523 characteristic AAE values of 1 and 2 for FF and BB to calculate their corresponding BC concentrations, namely BC(FF) and
524 BC(BB). Figure 13 shows the time series of BC(FF) and BC(BB) for different classified periods. The BC(FF) and BC(BB)
525 values were much higher before the monsoon than during/after the monsoon, except for the periods during BB-1 and BB-2
526 with significantly high BC(FF) and BC(BB) values (peaks $> 1 \mu\text{g m}^{-3}$), while high BC(FF) values were seen during SPP. Table
527 2 summarizes the average concentrations and the ranges of BC(FF) and BC(BB), along with their corresponding fractions. In
528 general, both the average BC(FF) and BC(BB) values were low during BMP, TMP, and AMP, compared to those during the
529 biomass burning events and SPP. BC(FF) contributed over 80% of ship pollution during SPP, whilst the BC(BB) could
530 contribute to more than 40% of the total black carbon during the two BB events. We hence conclude that fossil fuel combustion

531 is the major contributor to the light absorption of BC except during the seasonal biomass burning events and biomass burning
532 can have a profound contribution to the BC light absorption in the SCS.

533 Active biomass burning pollution during January–May in Southeast Asia occurs routinely because of crop residue and sugar
534 cane burning. A previous study showed that during dry and wet seasons, the annual contribution of BC(BB) was 11% and 30%
535 respectively in the Peninsular India (Soyam, 2021) based on the two-component AAE model (Drinovec et al., 2015; Yus-Díez
536 et al., 2021). Table 3 summarizes the BC concentrations, AAE values, and the corresponding fraction of biomass burning (or
537 fossil fuel) in previous and present studies for the marine atmosphere conducted at coastal sites or ship-based cruise
538 measurements using the AE series instrument. The BC(FF) and BC(BB) fractions of 58% and 42% were obtained respectively
539 during the two BB events, while they accounted for 78–83% and 18–22% during other periods, similar to those found at the
540 coastal site in Central Adriatic (79% and 21%), and significantly different from those reported at a coastal site in the East China
541 Sea (Yu et al., 2018). However, observation data are still lacking on the contribution of BC from fossil fuel vs biomass burning
542 in the sea regions which warrants more future studies during different seasons.

543 **3.7-3.5 Limitation of this study**

544 This cruise campaign for carbonaceous aerosols has several limitations which might need to be aware of due to the time and
545 area coverage constraints. The presence of other light-absorption aerosol components, polluted dusts, oil and gas drilling
546 emissions, as well as fishery policy may contribute to the uncertainties in the AAE model used for the BC source
547 apportionment in this study. Firstly, the composition of aerosols and refractive index may strongly affect the AAE calculation.
548 The source apportionment of BC for biomass burning and fossil fuel is based on the AAE two-component model which only
549 considered BC(BB) and BC(FF) as the light absorption materials. An AAE range of 0.9–1.4 is used for pure BC from fossil fuel
550 emissions, while it is 1.68–2.2 for biomass burning as mentioned earlier. The current AAE two-component model does not
551 include other potential light-absorbing materials, such as mineral dust and biological particles (Pileci et al., 2021). Interestingly,
552 two types of possible biological particles were observed during the campaign ([Fig. S13](#)[Fig. S10](#), in the SI, section 79). A similar
553 type of biological particles was observed and identified as brocosomes in another campaign near the East China Sea (Fu et al.,
554 2012). However, more future studies are needed to identify the types and species of biological particles and to evaluate their

555 contributions to the light-absorption.

556 Secondly, based on the Cloud-Aerosol Lidar & Infrared Satellite Observation (CALIPSO) data on May 15 and June 07 when
557 the orbit just passed over the SCS region and the Southeast Asia ([Fig. S14 and S15](#)~~Fig. S11 and S12~~, in the SI, section [810](#)),
558 we found the presence of polluted dust in the vertical profile over the Philippines, Indonesia, Thailand and Malaysia. Long-
559 range transport of dust may affect our measured AAE data. Thirdly, oil and natural gas drilling (Liu and Li, 2021) is active in
560 the SCS region and the distribution map is available online (<https://www.oilmap.xyz/>). These activities potentially contribute
561 to the BC emissions (Cordes et al., 2016), and these BC are similar to those of continental emitted BC from incomplete burning
562 of oil or natural gas.

563 Lastly, Chinese fishery policy enacts fishing prohibition for about three and a half months every year in the SCS during
564 summer which corresponds to May 01–Aug 16 in the year of 2021 in 12°N within our campaign region. Therefore, the cruise
565 measurements mainly captured ship emissions from the commercial ships in the SCS whose routes are available online
566 (<https://www.marinetraffic.com/en/ais/home/centerx:116.6/centery:20.5/zoom:4>). The average BC mass concentrations (~0.2
567 $\mu\text{g m}^{-3}$ for BC(FF) and 0.05–0.08 $\mu\text{g m}^{-3}$ for BC(BB)) are only limited to data of about a month and the coverage area. Hence,
568 more future measurements covering more seasons and wider areas are needed to better understand the morphology and optical
569 properties of the carbonaceous aerosols in the SCS.

570 **5. Conclusions**

571 As important components of carbonaceous aerosols, BC and BrC in the marine atmosphere may exert significantly positive
572 climatic radiative forcing through light absorption on the regional and global scales. However, quantification of their absorption
573 potential is tremendously challenging due to little knowledge on the microphysical properties, such as morphology, particle
574 size, and mixing state of BC or BrC in the marine region such as SCS. This ship-based study [is](#) intended to investigate the
575 morphological and optical properties of the BC particles in the SCS during summer using [a](#) combined online aethalometer,
576 semi-online OC/EC analyzer, and offline TEM/EDS analyses. The results showed that the lognormal fitted Feret diameter
577 distribution of the single particles peaks at 325 nm when the ship stopped, while it peaks at 307 nm when the ship navigated.
578 This minor difference in the size distribution could be attributed to the distinguishable air mass origins of the own ship

579 emissions for the former and the mixed other ship emissions and long-range transport for the latter. Furthermore, the Feret
580 diameters of the single particles spread much more narrowly during navigation (~~229-2557298~~–1980 nm) than those of the
581 freshly emitted particles during stop (~~78-2926304~~–2982 nm). In addition, the two types of single particles have ~~similar~~-same
582 median fractal dimension values (~~1.65~~ vs 1.661.61) but different lacunarity values (0.53 vs 0.59), indicating their different
583 aging degrees. The aged BC particles are present in various mixing states (core-shell, embedded, external) with other aerosol
584 components after long-range transport. Interestingly, a fraction of single particles ~~was~~-were identified as tar balls with
585 geometrical diameters of 160–420 nm which were primarily mixed with sea salt, organics, BC, and sulfate, and those tar balls
586 were found to originate from either ship emissions or long-range transport of biomass burning.

587 Since the marine atmosphere is mainly subject to the influence of biomass burning and fossil fuel combustion, a two-
588 component (BB and FF) AAE model was employed to evaluate the source contributions to the light absorption of the BC
589 particles. The modeling results indicated that BB and FF contributed respectively to 18–22% and 78–82% of all the BC light
590 absorption except for a substantial percentage of 42% for BB (hence 58% for FF) during the two observed significant biomass
591 burning events. The results from trajectory calculations showed that biomass burning was predominantly from the Philippines
592 and South East Asia before and after the summer monsoon during the cruise campaign. However, this highly simplified two-
593 component AAE model may have substantial uncertainties ~~on~~-in the evaluation of the source contributions when other sources
594 of BC particles were present and those included dust, biological materials, oil and gas drilling emissions during the
595 measurements. Nevertheless, this study demonstrates that emissions from ~~the~~-commercial ships and biomass burning from
596 Southeast Asia contribute to the enhanced light absorption of the BC particles in the SCS, especially during the crop harvest
597 seasons before monsoon, and the aged BC particles became more aggerated after long-range transport of air masses containing
598 biomass burning emissions.

599 **Author contributions**

600 JZ, CZS, and SZZ planned the cruise campaign. CZS, YYZ, BLL, MG, and XS performed the measurements. CZS performed
601 the data analysis and wrote the original draft. CZS and DXC performed funding acquisition. JZ and SZZ performed funding
602 acquisition and supervision. All authors reviewed and edited the manuscript.

603 **Declaration of competing interest**

604 The authors declare no conflict of interest.

605 **Data availability**

606 Data for figures and tables, along with raw data from online measurements and offline analyses of this study are available from
607 JZ via zhaojun23@mail.sysu.edu.cn upon request. The supplementary data are available online at xxx.

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617 **References**

618 Adachi, K., Zaizen, Y., Kajino, M., and Igarashi, Y.: Mixing state of regionally transported soot particles and the coating effect
619 on their size and shape at a mountain site in Japan, *J. Geophys. Res. Atmos.*, 119, 5386-5396,
620 <https://doi.org/10.1002/2013jd020880>, 2014.
621 Adachi, K., Sedlacek, A. J., III, Kleinman, L., Springston, S. R., Wang, J., Chand, D., Hubbe, J. M., Shilling, J. E., Onasch, T.
622 B., Kinase, T., Sakata, K., Takahashi, Y., and Buseck, P. R.: Spherical tarball particles form through rapid chemical and
623 physical changes of organic matter in biomass-burning smoke, *Proc. Natl. Acad. Sci. U.S.A.*, 116, 19336-19341,
624 <https://doi.org/10.1073/pnas.1900129116>, 2019.
625 Alroe, J., Cravigan, L. T., Miljevic, B., Johnson, G. R., Selleck, P., Humphries, R. S., Keywood, M. D., Chambers, S. D.,
626 Williams, A. G., and Ristovski, Z. D.: Marine productivity and synoptic meteorology drive summer-time variability in

627 Southern Ocean aerosols, *Atmos. Chem. Phys. Discuss.*, <https://doi.org/10.5194/acp-2019-1081>, 2019.

628 Andreae, M. O. and Gelencsér, A.: Black carbon or brown carbon? The nature of light-absorbing carbonaceous aerosols, *Atmos.*
629 *Chem. Phys.*, 6, 3131-3148, <https://doi.org/10.5194/acp-6-3131-2006>, 2006.

630 Ausmeel, S., Eriksson, A., Ahlberg, E., Sporre, M. K., Spanne, M., and Kristensson, A.: Ship plumes in the Baltic Sea sulfur
631 emission control area: Chemical characterization and contribution to coastal aerosol concentrations, *Atmos. Chem. Phys.*,
632 20, 9135-9151, <https://doi.org/10.5194/acp-20-9135-2020>, 2020.

633 Bao, M., Zhang, Y., Cao, F., Lin, Y., Wang, Y., Liu, X., Zhang, W., Fan, M. Y., Xie, F., Cary, R. A., Dixon, J., and Zhou, L.:
634 Highly time-resolved characterization of carbonaceous aerosols using a two-wavelength Sunset thermal-optical carbon
635 analyzer, *Atmos. Meas. Tech.*, 14, 4053-4068, <https://doi.org/10.5194/amt-14-4053-2021>, 2021.

636 Bauer, J. J., Yu, X.-Y., Cary, R., Laulainen, N., and Berkowitz, C.: Characterization of the Sunset semi-continuous carbon
637 aerosol analyzer, *J. Air Waste Manag. Assoc.*, 59, 826-833, <https://doi.org/10.3155/1047-3289.59.7.826>, 2012.

638 Bencs, L., Horemans, B., Buczyńska, A. J., Deutsch, F., Degraeuwe, B., Van Poppel, M., and Van Grieken, R.: Seasonality of
639 ship emission related atmospheric pollution over coastal and open waters of the North Sea, *Atmos. Environ.: X.*, 7,
640 <https://doi.org/10.1016/j.aeaoa.2020.100077>, 2020.

641 Blanco-Donado, E. P.: Source identification and global implications of black carbon, *Geosci. Front.*,
642 <https://doi.org/10.1016/j.gsf.2021.101149>, 2022.

643 Bond, T. C., Doherty, S. J., Fahey, D. W., Forster, P. M., Berntsen, T., DeAngelo, B. J., Flanner, M. G., Ghan, S., Kärcher, B.,
644 Koch, D., Kinne, S., Kondo, Y., Quinn, P. K., Sarofim, M. C., Schultz, M. G., Schulz, M., Venkataraman, C., Zhang, H.,
645 Zhang, S., Bellouin, N., Guttikunda, S. K., Hopke, P. K., Jacobson, M. Z., Kaiser, J. W., Klimont, Z., Lohmann, U., Schwarz,
646 J. P., Shindell, D., Storelvmo, T., Warren, S. G., and Zender, C. S.: Bounding the role of black carbon in the climate system:
647 A scientific assessment, *J. Geophys. Res. Atmos.*, 118, 5380-5552, <https://doi.org/10.1002/jgrd.50171>, 2013.

648 Brown, S., Minor, H., O'Brien, T., Hameed, Y., Feenstra, B., Kuebler, D., Wetherell, W., Day, R., Tun, R., Landis, E., and Rice,
649 J.: Review of Sunset OC/EC instrument measurements during the EPA's Sunset carbon evaluation project, *Atmosphere*
650 (Basel), 10, 287, <https://doi.org/10.3390/atmos10050287>, 2019.

651 Budhavant, K., Andersson, A., Holmstrand, H., Bikkina, P., Bikkina, S., Satheesh, S. K., and Gustafsson, Ö.: Enhanced light-
652 absorption of black carbon in rainwater compared with aerosols over the Northern Indian Ocean, *J. Geophys. Res. Atmos.*,
653 125, <https://doi.org/10.1029/2019jd031246>, 2020.

654 Cai, M., Liang, B., Sun, Q., Zhou, S., Chen, X., Yuan, B., Shao, M., Tan, H., and Zhao, J.: Effects of continental emissions on
655 cloud condensation nuclei (CCN) activity in the northern South China Sea during summertime 2018, *Atmos. Chem. Phys.*,
656 20, 9153-9167, <https://doi.org/10.5194/acp-20-9153-2020>, 2020.

657 Cappa, C. D., Onasch, T. B., Massoli, P., Worsnop, D. R., Bates, T. S., Cross, E. S., Davidovits, P., Hakala, J., Hayden, K. L.,
658 Jobson, B. T., Kolesar, K. R., Lack, D. A., Lerner, B. M., Li, S.-M., Mellon, D., Nuaaman, I., Olfert, J. S., Petäjä, T., Quinn,
659 P. K., Song, C., Subramanian, R., Williams, E. J., and Zaveri, R. A.: Radiative absorption enhancements due to the mixing
660 state of atmospheric black carbon, *Science*, 337, 1078, <https://doi.org/10.1126/science.1223447>, 2012.

661 Chakrabarty, R. K., Moosmüller, H., Garro, M. A., Arnott, W. P., Walker, J., Susott, R. A., Babbitt, R. E., Wold, C. E., Lincoln,
662 E. N., and Hao, W. M.: Emissions from the laboratory combustion of wildland fuels: Particle morphology and size, *J.*
663 *Geophys. Res.*, 111, <https://doi.org/10.1029/2005jd006659>, 2006.

664 Chaubey, J. P., Moorthy, K. K., Babu, S. S., and Gogoi, M. M.: Spatio-temporal variations in aerosol properties over the oceanic
665 regions between coastal India and Antarctica, *J. Atmos. Sol. Terr. Phys.*, 104, 18-28,
666 <https://doi.org/10.1016/j.jastp.2013.08.004>, 2013.

667 Chen, D., Cui, H., Zhao, Y., Yin, L., Lu, Y., and Wang, Q.: A two-year study of carbonaceous aerosols in ambient PM_{2.5} at a
668 regional background site for western Yangtze River Delta, China, *Atmos. Res.*, 183, 351-361,
669 <https://doi.org/https://doi.org/10.1016/j.atmosres.2016.09.004>, 2017.

670 [Chen, X., Ye, C., Wang, Y., Wu, Z., Zhu, T., Zhang, F., Ding, X., Shi, Z., Zheng, Z., and Li, W.: Quantifying evolution of soot
671 mixing state from transboundary transport of biomass burning emissions, *iScience*, 26, 108125, 10.1016/j.isci.2023.108125,
672 2023.](#)

673 [Cheng, Z., Sharma, N., Tseng, K. P., Kovarik, L., and China, S.: Direct observation and assessment of phase states of ambient
674 and lab-generated sub-micron particles upon humidification, *RSC Adv*, 11, 15264-15272, 10.1039/d1ra02530a, 2021.](#)

675 China, S., Mazzoleni, C., Gorkowski, K., Aiken, A. C., and Dubey, M. K.: Morphology and mixing state of individual freshly
676 emitted wildfire carbonaceous particles, *Nat. Commun.*, 4, 2122, <https://doi.org/10.1038/ncomms3122>, 2013.

677 Chylek, P., Lee, J. E., Romonosky, D. E., Gallo, F., Lou, S., Shrivastava, M., Carrico, C. M., Aiken, A. C., and Dubey, M. K.:
678 Mie scattering captures observed optical properties of ambient biomass burning plumes assuming uniform black, brown,
679 and organic carbon mixtures, *J. Geophys. Res. Atmos.*, 124, 11406-11427, <https://doi.org/10.1029/2019jd031224>, 2019.

680 Corbin, J. C., Czech, H., Massabò, D., de Mongeot, F. B., Jakobi, G., Liu, F., Lobo, P., Mennucci, C., Mensah, A. A., Orasche,
681 J., Pieber, S. M., Prévôt, A. S. H., Stengel, B., Tay, L. L., Zannata, M., Zimmermann, R., El Haddad, I., and Gysel, M.:
682 Infrared-absorbing carbonaceous tar can dominate light absorption by marine-engine exhaust, *npj Clim. Atmos. Sci.*, 2,
683 <https://doi.org/10.1038/s41612-019-0069-5>, 2019.

684 Cordes, E. E., Jones, D. O. B., Schlacher, T. A., Amon, D. J., Bernardino, A. F., Brooke, S., Carney, R., DeLeo, D. M., Dunlop,
685 K. M., Escobar-Briones, E. G., Gates, A. R., Génio, L., Gobin, J., Henry, L.-A., Herrera, S., Hoyt, S., Joye, M., Kark, S.,
686 Mestre, N. C., Metaxas, A., Pfeifer, S., Sink, K., Sweetman, A. K., and Witte, U.: Environmental impacts of the deep-water
687 oil and gas industry: A review to guide management strategies, *Front. Environ. Sci.*, 4,
688 <https://doi.org/10.3389/fenvs.2016.00058>, 2016.

689 [Dong, Z., Kang, S., Qin, D., Shao, Y., Ulbrich, S., and Qin, X.: Variability in individual particle structure and mixing states
690 between the glacier-snowpack and atmosphere in the northeastern Tibetan Plateau, *The Cryosphere*, 12, 3877-3890,
691 10.5194/tc-12-3877-2018, 2018.](#)

692 Drinovec, L., Močnik, G., Zotter, P., Prévôt, A. S. H., Ruckstuhl, C., Coz, E., Rupakheti, M., Sciare, J., Müller, T., Wiedensohler,
693 A., and Hansen, A. D. A.: The "dual-spot" Aethalometer: An improved measurement of aerosol black carbon with real-time
694 loading compensation, *Atmos. Meas. Tech.*, 8, 1965-1979, <https://doi.org/10.5194/amt-8-1965-2015>, 2015.

695 Duarte, R. M. B. O., Gomes, J. F. P., Querol, X., Cattaneo, A., Bergmans, B., Saraga, D., Maggos, T., Di Gilio, A., Rovelli, S.,
696 and Villanueva, F.: Advanced instrumental approaches for chemical characterization of indoor particulate matter, *Appl.*
697 *Spectrosc. Rev.*, 57, 705-745, <https://doi.org/10.1080/05704928.2021.2018596>, 2021.

698 Ezani, E., Dhandapani, S., Heal, M. R., Praveena, S. M., Khan, M. F., and Ramly, Z. T. A.: Characteristics and source
699 apportionment of black carbon (BC) in a suburban area of Klang Valley, Malaysia, *Atmosphere*, 12,
700 <https://doi.org/10.3390/atmos12060784>, 2021.

701 Feng, X., Wang, J., Teng, S., Xu, X., Zhu, B., Wang, J., Zhu, X., Yurkin, M. A., and Liu, C.: Can light absorption of black
702 carbon still be enhanced by mixing with absorbing materials?, *Atmos. Environ.*, 253, 118358,
703 <https://doi.org/10.1016/j.atmosenv.2021.118358>, 2021.

704 Fossum, K. N., Ovadnevaite, J., Liu, D., Flynn, M., O'Dowd, C., and Ceburnis, D.: Background levels of black carbon over
705 remote marine locations, *Atmos. Res.*, 271, <https://doi.org/10.1016/j.atmosres.2022.106119>, 2022.

706 Fu, H., Zhang, M., Li, W., Chen, J., Wang, L., Quan, X., and Wang, W.: Morphology, composition and mixing state of individual
707 carbonaceous aerosol in urban Shanghai, *Atmos. Chem. Phys.*, 12, 693-707, <https://doi.org/10.5194/acp-12-693-2012>, 2012.

708 Geron, C.: Carbonaceous aerosol over a Pinus taeda forest in Central North Carolina, USA, *Atmos. Environ.*, 43, 959-969,
709 <https://doi.org/10.1016/j.atmosenv.2008.10.053>, 2009.

710 Giglio, L., Schroeder, W., Hall, J. V., and Justice, C. O.: MODIS collection 6 active fire product user's guide revision C,
711 University of Maryland. National Oceanic and Atmospheric Administration, 63 pp., 2020.

712 Girach, I. A., Nair, V. S., Babu, S. S., and Nair, P. R.: Black carbon and carbon monoxide over Bay of Bengal during W_ICARB:
713 Source characteristics, *Atmos. Environ.*, 94, 508-517, <https://doi.org/10.1016/j.atmosenv.2014.05.054>, 2014.

714 [Giotto, G., China, S., Bhandari, J., Gorkowski, K., Scarnato, B. V., Capek, T., Marinoni, A., Veghte, D. P., Kulkarni, G., Aiken,](#)
715 [A. C., Dubey, M., and Mazzoleni, C.: Fractal-like tar ball aggregates from wildfire smoke, *Environ. Sci. Technol. Lett.*, 5,](#)
716 [360-365, 10.1021/acs.estlett.8b00229, 2018.](#) Hand, J. L., Malm, W. C., Laskin, A., Day, D., Lee, T., Wang, C., Carrico, C.,
717 Carrillo, J., Cowin, J. P., Collett, J., and Iedema, M. J.: Optical, physical, and chemical properties of tar balls observed during
718 the Yosemite Aerosol Characterization Study, *J. Geophys. Res.*, 110, <https://doi.org/10.1029/2004jd005728>, 2005.

719 Harrison, R. M.: Airborne particulate matter, *Philos. Trans. R. Soc. A*, 378, 20190319, <https://doi.org/10.1098/rsta.2019.0319>,
720 2020.

721 Helin, A., Virkkula, A., Backman, J., Pirjola, L., Sippula, O., Aakko-Saksa, P., Väätäinen, S., Mylläri, F., Järvinen, A., Bloss,
722 M., Aurela, M., Jakobi, G., Karjalainen, P., Zimmermann, R., Jokiniemi, J., Saarikoski, S., Tissari, J., Rönkkö, T., Niemi, J.
723 V., and Timonen, H.: Variation of absorption Ångström exponent in aerosols from different emission sources, *J. Geophys.*
724 *Res. Atmos.*, 126, <https://doi.org/10.1029/2020jd034094>, 2021.

725 Hoffer, A., Tóth, A., Nyirő-Kósa, I., Pósfai, M., and Gelencsér, A.: Light absorption properties of laboratory-generated tar ball
726 particles, *Atmos. Chem. Phys.*, 16, 239-246, <https://doi.org/10.5194/acp-16-239-2016>, 2016.

727 Jung, J., Kim, Y. J., Lee, K. Y., Kawamura, K., Hu, M., and Kondo, Y.: The effects of accumulated refractory particles and the
728 peak inert mode temperature on semi-continuous organic carbon and elemental carbon measurements during the
729 CAREBeijing 2006 campaign, *Atmos. Environ.*, 45, 7192-7200, <https://doi.org/10.1016/j.atmosenv.2011.09.003>, 2011.

730 Kang, H., Shang, X., Abdumutallip, M., Chen, Y., Li, L., Wang, X., Li, C., Ouyang, H., Tang, X., Wang, L., Rudich, Y., and
731 Chen, J.: Accurate observation of black and brown carbon in atmospheric fine particles via a versatile aerosol concentration
732 enrichment system (VACES), *Sci. Total Environ.*, 837, 155817, <https://doi.org/10.1016/j.scitotenv.2022.155817>, 2022.

733 Karanasiou, A., Panteliadis, P., Perez, N., Minguillon, M. C., Pandolfi, M., Titos, G., Viana, M., Moreno, T., Querol, X., and
734 Alastuey, A.: Evaluation of the Semi-Continuous OCEC analyzer performance with the EUSAAR2 protocol, *Sci. Total*
735 *Environ.*, 747, 141266, <https://doi.org/10.1016/j.scitotenv.2020.141266>, 2020.

736 Kedia, S., Ramachandran, S., Rajesh, T. A., and Srivastava, R.: Aerosol absorption over Bay of Bengal during winter:
737 Variability and sources, *Atmos. Environ.*, 54, 738-745, <https://doi.org/10.1016/j.atmosenv.2011.12.047>, 2012.

738 Khan, B., Hays, M. D., Geron, C., and Jetter, J.: Differences in the OC/EC ratios that characterize ambient and source aerosols
739 due to thermal-optical analysis, *Aerosol Sci. Technol.*, 46, 127-137, <https://doi.org/10.1080/02786826.2011.609194>, 2012.

740 Kompalli, S. K., Babu, S. N. S., Moorthy, K. K., Satheesh, S. K., Gogoi, M. M., Nair, V. S., Jayachandran, V. N., Liu, D.,
741 Flynn, M. J., and Coe, H.: Mixing state of refractory black carbon aerosol in the South Asian outflow over the northern
742 Indian Ocean during winter, *Atmos. Chem. Phys.*, 21, 9173-9199, <https://doi.org/10.5194/acp-21-9173-2021>, 2021.

743 Kwak, N., Lee, H., Maeng, H., Seo, A., Lee, K., Kim, S., Lee, M., Cha, J. W., Shin, B., and Park, K.: Morphological and
744 chemical classification of fine particles over the Yellow Sea during spring, 2015-2018, *Environ. Pollut.*, 305, 119286,
745 <https://doi.org/10.1016/j.envpol.2022.119286>, 2022.

746 [Lack, D. A. and Cappa, C. D.: Impact of brown and clear carbon on light absorption enhancement, single scatter albedo and](#)
747 [absorption wavelength dependence of black carbon, *Atmos. Chem. Phys.*, 10, 4207-4220, 10.5194/acp-10-4207-2010, 2010.](#)

748 [Laing, J. R., Jaffe, D. A., and Sedlacek, I. I. A. J.: Comparison of Filter-based Absorption Measurements of Biomass Burning](#)
749 [Aerosol and Background Aerosol at the Mt. Bachelor Observatory, *Aerosol Air Qual. Res.*, 20, 663-678,](#)
750 [10.4209/aaqr.2019.06.0298, 2020.](#)

751 Lappi, M. K. and Ristimäki, J. M.: Evaluation of thermal optical analysis method of elemental carbon for marine fuel exhaust,
752 *J. Air Waste Manag. Assoc.*, 67, 1298-1318, <https://doi.org/10.1080/10962247.2017.1335251>, 2017.

753 Laskin, A., Laskin, J., and Nizkorodov, S. A.: Chemistry of atmospheric brown carbon, *Chem. Rev.*, 115, 4335-4382,
754 <https://doi.org/10.1021/cr5006167>, 2015.

755 Li, J., Zhang, Q., Wang, G., Li, J., Wu, C., Liu, L., Wang, J., Jiang, W., Li, L., Ho, K. F., and Cao, J.: Optical properties and
756 molecular compositions of water-soluble and water-insoluble brown carbon (BrC) aerosols in northwest China, *Atmos.*
757 *Chem. Phys.*, 20, 4889-4904, <https://doi.org/10.5194/acp-20-4889-2020>, 2020.

758 Liu, J. and Li, X.: Recent advances on natural gas hydrate exploration and development in the South China Sea, *Energy Fuels*,
759 35, 7528-7552, <https://doi.org/10.1021/acs.energyfuels.1c00494>, 2021.

760 Liu, L., Zhang, J., Zhang, Y., Wang, Y., Xu, L., Yuan, Q., Liu, D., Sun, Y., Fu, P., Shi, Z., and Li, W.: Persistent residential
761 burning-related primary organic particles during wintertime hazes in North China: insights into their aging and optical
762 changes, *Atmos. Chem. Phys.*, 21, 2251-2265, <https://doi.org/10.5194/acp-21-2251-2021>, 2021.

763 Liu, X., Zhu, R., Jin, B., Zu, L., Wang, Y., Wei, Y., and Zhang, R.: Emission characteristics and light absorption apportionment
764 of carbonaceous aerosols: A tunnel test conducted in an urban with fully enclosed use of E10 petrol, *Environ. Res.*, 216,
765 <https://doi.org/10.1016/j.envres.2022.114701>, 2023.

766 Lu, Q., Liu, C., Zhao, D., Zeng, C., Li, J., Lu, C., Wang, J., and Zhu, B.: Atmospheric heating rate due to black carbon aerosols:
767 Uncertainties and impact factors, *Atmos. Res.*, 240, <https://doi.org/10.1016/j.atmosres.2020.104891>, 2020.

768 Luo, J., Zhang, Y., and Zhang, Q.: Effects of black carbon morphology on brown carbon absorption estimation: From numerical
769 aspects, *Geosci. Model Dev.*, 14, 2113-2126, <https://doi.org/10.5194/gmd-14-2113-2021>, 2021.

770 Luo, J., Li, Z., Zhang, C., Zhang, Q., Zhang, Y., Zhang, Y., Curci, G., and Chakrabarty, R. K.: Regional impacts of black carbon
771 morphologies on shortwave aerosol-radiation interactions: A comparative study between the US and China, *Atmos. Chem.*
772 *Phys.*, 22, 7647-7666, <https://doi.org/10.5194/acp-22-7647-2022>, 2022.

773 Ma, Y., Zhang, X., Xin, J., Zhang, W., Wang, Z., Liu, Q., Wu, F., Wang, L., Lyu, Y., Wang, Q., and Ma, Y.: Mass and number
774 concentration distribution of marine aerosol in the Western Pacific and the influence of continental transport, *Environ. Pollut.*,

775 298, 118827, <https://doi.org/10.1016/j.envpol.2022.118827>, 2022.

776 Marple, V. A. and Olson, B. A.: Sampling and measurement using inertial, gravitational, centrifugal, and thermal techniques,
777 in: Aerosol measurement: Principles, techniques, and applications, edited by: Kulkarni, P., Baron, P. A., and Willeke, K.,
778 John Wiley and Sons, Hoboken, New Jersey, USA, 129-151, <https://doi.org/10.1002/9781118001684.ch8>, 2011.

779 Matsui, H., Koike, M., Kondo, Y., Oshima, N., Moteki, N., Kanaya, Y., Takami, A., and Irwin, M.: Seasonal variations of Asian
780 black carbon outflow to the Pacific: Contribution from anthropogenic sources in China and biomass burning sources in
781 Siberia and Southeast Asia, *J. Geophys. Res. Atmos.*, 118, 9948-9967, <https://doi.org/10.1002/jgrd.50702>, 2013.

782 Milinković, A., Gregorič, A., Grgičin, V. D., Vidič, S., Penezić, A., Kušan, A. C., Alempijević, S. B., Kasper-Giebl, A., and
783 Frka, S.: Variability of black carbon aerosol concentrations and sources at a Mediterranean coastal region, *Atmos. Pollut.*
784 *Res.*, 12, <https://doi.org/10.1016/j.apr.2021.101221>, 2021.

785 Moschos, V., Gysel-Beer, M., Modini, R. L., Corbin, J. C., Massabò, D., Costa, C., Danelli, S. G., Vlachou, A., Daellenbach,
786 K. R., Szidat, S., Prati, P., Prévôt, A. S. H., Baltensperger, U., and El Haddad, I.: Source-specific light absorption by
787 carbonaceous components in the complex aerosol matrix from yearly filter-based measurements, *Atmos. Chem. Phys.*, 21,
788 12809-12833, <https://doi.org/10.5194/acp-21-12809-2021>, 2021.

789 Neusüß, C.: Carbonaceous aerosol over the Indian Ocean: OC/EC fractions and selected specifications from size-segregated
790 onboard samples, *J. Geophys. Res.*, 107, <https://doi.org/10.1029/2001jd000327>, 2002.

791 Pabst, W. and Gregorova, E.: Characterization of particles and particle systems, ICT Prague, 122, 122, 2007.

792 Pang, Y., Wang, Y., Wang, Z., Zhang, Y., Liu, L., Kong, S., Liu, F., Shi, Z., and Li, W.: Quantifying the fractal dimension and
793 morphology of individual atmospheric soot aggregates, *J. Geophys. Res. Atmos.*, 127,
794 <https://doi.org/10.1029/2021jd036055>, 2022.

795 Pankratova, N. V., Belikov, I. B., Belousov, V. A., Kopeikin, V. M., Skorokhod, A. I., Shtabkin, Y. A., Malafeev, G. V., and
796 Flint, M. V.: Concentration and isotopic composition of methane, associated gases, and black carbon over Russian Arctic
797 Seas (shipborne measurements), *Oceanology*, 60, 593-602, <https://doi.org/10.1134/s0001437020050197>, 2021.

798 Park, S., Yu, G.-H., and Lee, S.: Optical absorption characteristics of brown carbon aerosols during the KORUS-AQ campaign
799 at an urban site, *Atmos. Res.*, 203, 16-27, <https://doi.org/10.1016/j.atmosres.2017.12.002>, 2018.

800 Peyronel, M. F., Acevedo, N. C., and Marangoni, A. G.: Structural and mechanical properties of fats and their implications for
801 food quality, in: Chemical deterioration and physical instability of food and beverages, edited by: Skibsted, L. H., Risbo, J.,
802 and Andersen, M. L., Woodhead Publishing Limited, Abington Hall, Granta Park, Great Abington, Cambridge CB21 6AH,
803 UK, 216-259, <https://doi.org/10.1533/9781845699260.2.216>, 2010.

804 Pileci, R. E., Modini, R. L., Bertò, M., Yuan, J., Corbin, J. C., Marinoni, A., Henzing, B., Moerman, M. M., Putaud, J. P.,
805 Spindler, G., Wehner, B., Müller, T., Tuch, T., Trentini, A., Zanatta, M., Baltensperger, U., and Gysel-Beer, M.: Comparison
806 of co-located refractory black carbon (rBC) and elemental carbon (EC) mass concentration measurements during field
807 campaigns at several European sites, *Atmos. Meas. Tech.*, 14, 1379-1403, <https://doi.org/10.5194/amt-14-1379-2021>, 2021.

808 Pipal, A. S., Rohra, H., Tiwari, R., and Taneja, A.: Particle size distribution, morphometric study and mixing structure of
809 accumulation and ultrafine aerosols emitted from indoor activities in different socioeconomic micro-environment, *Atmos.*
810 *Pollut. Res.*, 12, 101-111, <https://doi.org/10.1016/j.apr.2021.02.015>, 2021.

811 Ponczek, M., Franco, M. A., Carbone, S., Rizzo, L. V., Monteiro dos Santos, D., Morais, F. G., Duarte, A., Barbosa, H. M. J.,

812 and Artaxo, P.: Linking the chemical composition and optical properties of biomass burning aerosols in Amazonia, *Environ.*
813 *Sci. Atmos.*, 2, 252-269, <https://doi.org/10.1039/d1ea00055a>, 2022.

814 Rabha, S. and Saikia, B. K.: An environmental evaluation of carbonaceous aerosols in PM10 at micro- and nano-scale levels
815 reveals the formation of carbon nanodots, *Chemosphere*, 244, 125519, <https://doi.org/10.1016/j.chemosphere.2019.125519>,
816 2020.

817 Retama, A., Ramos-Cerón, M., Rivera-Hernández, O., Allen, G., and Velasco, E.: Aerosol optical properties and brown carbon
818 in Mexico City, *Environ. Sci. Atmos.*, 2, 315-334, <https://doi.org/10.1039/d2ea00006g>, 2022.

819 Rice, S. B., Chan, C., Brown, S. C., Eschbach, P., Han, L., Ensor, D. S., Stefaniak, A. B., Bonevich, J., Vladar, A. E., Hight
820 Walker, A. R., Zheng, J., Starnes, C., Stromberg, A., Ye, J., and Grulke, E. A.: Particle size distributions by transmission
821 electron microscopy: An interlaboratory comparison case study, *Metrologia*, 50, 663-678, [https://doi.org/10.1088/0026-](https://doi.org/10.1088/0026-1394/50/6/663)
822 [1394/50/6/663](https://doi.org/10.1088/0026-1394/50/6/663), 2013.

823 Schmale, J., Baccarini, A., Thurnherr, I., Henning, S., Efraim, A., Regayre, L., Bolas, C., Hartmann, M., Welti, A., Lehtipalo,
824 K., Aemisegger, F., Tatzelt, C., Landwehr, S., Modini, R. L., Tummon, F., Johnson, J. S., Harris, N., Schnaiter, M., Toffoli,
825 A., Derkani, M., Bukowiecki, N., Stratmann, F., Dommen, J., Baltensperger, U., Wernli, H., Rosenfeld, D., Gysel-Beer, M.,
826 and Carslaw, K. S.: Overview of the Antarctic Circumnavigation Expedition: Study of Preindustrial-like aerosols and their
827 climate effects (ACE-SPACE), *Bull. Am. Meteorol. Soc.*, 100, 2260-2283, <https://doi.org/10.1175/bams-d-18-0187.1>, 2019.

828 Shamjad, P. M., Tripathi, S. N., Aggarwal, S. G., Mishra, S. K., Joshi, M., Khan, A., Sapra, B. K., and Ram, K.: Comparison
829 of experimental and modeled absorption enhancement by black carbon (BC) cored polydisperse aerosols under hygroscopic
830 conditions, *Environ. Sci. Technol.*, 46, 8082-8089, <https://doi.org/10.1021/es300295v>, 2012.

831 Shank, L. M., Howell, S., Clarke, A. D., Freitag, S., Brekhovskikh, V., Kapustin, V., McNaughton, C., Campos, T., and Wood,
832 R.: Organic matter and non-refractory aerosol over the remote Southeast Pacific: Oceanic and combustion sources, *Atmos.*
833 *Chem. Phys.*, 12, 557-576, <https://doi.org/10.5194/acp-12-557-2012>, 2012.

834 Sharma, S., Barrie, L. A., Magnusson, E., Brattström, G., Leaitch, W. R., Steffen, A., and Landsberger, S.: A factor and trends
835 analysis of multidecadal lower tropospheric observations of Arctic aerosol composition, black carbon, ozone, and mercury
836 at Alert, Canada, *J. Geophys. Res. Atmos.*, 124, 14133-14161, <https://doi.org/10.1029/2019jd030844>, 2019.

837 [Song, X., Xie, X., Qiu, B., Cao, H., Xie, S.-P., Chen, Z., and Yu, W.: Air-sea latent heat flux anomalies induced by oceanic](https://doi.org/10.3389/fmars.2022.850207)
838 [submesoscale processes: An observational case study, *Front. Mar. Sci.*, 9, 10.3389/fmars.2022.850207, 2022.](https://doi.org/10.3389/fmars.2022.850207)

839 Soyam, P. S.: Black carbon aerosols over a semi-arid rain shadow location in Peninsular India: Temporal variability and sources,
840 *J. Earth Syst. Sci.*, 130, <https://doi.org/10.1007/s12040-021-01610-5>, 2021.

841 Sun, C., Adachi, K., Misawa, K., Cheung, H. C., Chou, C. C. K., and Takegawa, N.: Mixing state of black carbon particles in
842 Asian outflow observed at a remote site in Taiwan in the spring of 2017, *J. Geophys. Res. Atmos.*, 125, 13,
843 <https://doi.org/10.1029/2020jd032526>, 2020a.

844 Sun, L., Chen, T., Jiang, Y., Zhou, Y., Sheng, L., Lin, J., Li, J., Dong, C., Wang, C., Wang, X., Zhang, Q., Wang, W., and Xue,
845 L.: Ship emission of nitrous acid (HONO) and its impacts on the marine atmospheric oxidation chemistry, *Sci. Total Environ.*,
846 735, 139355, <https://doi.org/10.1016/j.scitotenv.2020.139355>, 2020b.

847 Sun, Q., Liang, B., Cai, M., Zhang, Y., Ou, H., Ni, X., Sun, X., Han, B., Deng, X., Zhou, S., and Zhao, J.: Cruise observation
848 of the marine atmosphere and ship emissions in South China Sea: Aerosol composition, sources, and the aging process,

849 Environ. Pollut., 316, 120539, <https://doi.org/10.1016/j.envpol.2022.120539>, 2023.

850 Sun, T., Wu, C., Wu, D., Liu, B., Sun, J. Y., Mao, X., Yang, H., Deng, T., Song, L., Li, M., Li, Y. J., and Zhou, Z.: Time-resolved
851 black carbon aerosol vertical distribution measurements using a 356-m meteorological tower in Shenzhen, Theor. Appl.
852 Climatol., 140, 1263-1276, <https://doi.org/10.1007/s00704-020-03168-6>, 2020c.

853 Taketani, F., Miyakawa, T., Takashima, H., Komazaki, Y., Pan, X., Kanaya, Y., and Inoue, J.: Shipborne observations of
854 atmospheric black carbon aerosol particles over the Arctic Ocean, Bering Sea, and North Pacific Ocean during September
855 2014, J. Geophys. Res. Atmos., 121, 1914-1921, <https://doi.org/10.1002/2015jd023648>, 2016.

856 Tóth, A., Hoffer, A., Nyirő-Kósa, I., Pósfai, M., and Gelencsér, A.: Atmospheric tar balls: aged primary droplets from biomass
857 burning?, Atmos. Chem. Phys., 14, 6669-6675, <https://doi.org/10.5194/acp-14-6669-2014>, 2014.

858 Ueda, S., Osada, K., Hara, K., Yabuki, M., Hashihama, F., and Kanda, J.: Morphological features and mixing states of soot-
859 containing particles in the marine boundary layer over the Indian and Southern oceans, Atmos. Chem. Phys., 18, 9207-9224,
860 <https://doi.org/10.5194/acp-18-9207-2018>, 2018.

861 Virkkula, A.: Modeled source apportionment of black carbon particles coated with a light-scattering shell, Atmos. Meas. Tech.,
862 14, 3707-3719, <https://doi.org/10.5194/amt-14-3707-2021>, 2021.

863 Wang, G., Chen, J., Xu, J., Yun, L., Zhang, M., Li, H., Qin, X., Deng, C., Zheng, H., Gui, H., Liu, J., and Huang, K.:
864 Atmospheric processing at the sea-land interface over the South China Sea: Secondary aerosol formation, aerosol acidity,
865 and role of sea salts, J. Geophys. Res. Atmos., 127, <https://doi.org/10.1029/2021jd036255>, 2022.

866 Wang, Q., Liu, H., Ye, J., Tian, J., Zhang, T., Zhang, Y., Liu, S., and Cao, J.: Estimating absorption Ångström exponent of
867 black carbon aerosol by coupling multiwavelength absorption with chemical composition, Environ. Sci. Technol., 8, 121-
868 127, <https://doi.org/10.1021/acs.estlett.0c00829>, 2020a.

869 Wang, Q., Liu, H., Wang, P., Dai, W., Zhang, T., Zhao, Y., Tian, J., Zhang, W., Han, Y., and Cao, J.: Optical source
870 apportionment and radiative effect of light-absorbing carbonaceous aerosols in a tropical marine monsoon climate zone:
871 The importance of ship emissions, Atmos. Chem. Phys., 20, 15537-15549, <https://doi.org/10.5194/acp-20-15537-2020>,
872 2020b.

873 Wang, Y., Liu, F., He, C., Bi, L., Cheng, T., Wang, Z., Zhang, H., Zhang, X., Shi, Z., and Li, W.: Fractal dimensions and mixing
874 structures of soot particles during atmospheric processing, Environ. Sci. Technol., 4, 487-493,
875 <https://doi.org/10.1021/acs.estlett.7b00418>, 2017.

876 Wang, Y. L. and Wu, C. R.: Nonstationary El Niño teleconnection on the post-summer upwelling off Vietnam, Sci. Rep., 10,
877 13319, <https://doi.org/10.1038/s41598-020-70147-2>, 2020.

878 Wei, X., Zhu, Y., Hu, J., Liu, C., Ge, X., Guo, S., Liu, D., Liao, H., and Wang, H.: Recent progress in impacts of mixing state
879 on optical properties of black carbon aerosol, Curr. Pollut. Rep., 6, 380-398, <https://doi.org/10.1007/s40726-020-00158-0>,
880 2020.

881 Wu, D., Wu, C., Liao, B., Chen, H., Wu, M., Li, F., Tan, H., Deng, T., Li, H., Jiang, D., and Yu, J. Z.: Black carbon over the
882 South China Sea and in various continental locations in South China, Atmos. Chem. Phys., 13, 12257-12270,
883 <https://doi.org/10.5194/acp-13-12257-2013>, 2013a.

884 Wu, H., Lattuada, M., and Morbidelli, M.: Dependence of fractal dimension of DLCA clusters on size of primary particles,
885 Adv. Colloid Interface Sci., 195-196, 41-49, <https://doi.org/10.1016/j.cis.2013.04.001>, 2013b.

886 Xing, J., Bian, L., Hu, Q., Yu, J., Sun, C., and Xie, Z.: Atmospheric black carbon along a cruise path through the Arctic Ocean
887 during the fifth Chinese Arctic Research Expedition, *Atmosphere*, 5, 292-306, <https://doi.org/10.3390/atmos5020292>, 2014.

888 Yan, J., Chen, L., Zhao, S., Zhang, M., Lin, Q., and Li, L.: Impact of marine and continental sources on aerosol characteristics
889 using an on-board SPAMS over southeast sea, China, *Environ. Sci. Pollut. Res.*, 25, 30659-30670,
890 <https://doi.org/10.1007/s11356-018-2902-5>, 2018.

891 Yang, M., Howell, S. G., Zhuang, J., and Huebert, B. J.: Attribution of aerosol light absorption to black carbon, brown carbon,
892 and dust in China – interpretations of atmospheric measurements during EAST-AIRE, *Atmos. Chem. Phys.*, 9, 2035-2050,
893 <https://doi.org/10.5194/acp-9-2035-2009>, 2009.

894 You, R., Radney, J. G., Zachariah, M. R., and Zangmeister, C. D.: Measured wavelength-dependent absorption enhancement
895 of internally mixed black carbon with absorbing and nonabsorbing materials, *Environ. Sci. Technol.*, 50, 7982-7990,
896 <https://doi.org/10.1021/acs.est.6b01473>, 2016.

897 Yu, G.-H., Park, S., Shin, S.-K., Lee, K.-H., and Nam, H.-G.: Enhanced light absorption due to aerosol particles in ship plumes
898 observed at a seashore site, *Atmos. Pollut. Res.*, 9, 1177-1183, <https://doi.org/10.1016/j.apr.2018.05.005>, 2018.

899 Yus-Díez, J., Bernardoni, V., Močnik, G., Alastuey, A., Ciniglia, D., Ivančič, M., Querol, X., Perez, N., Reche, C., Rigler, M.,
900 Vecchi, R., Valentini, S., and Pandolfi, M.: Determination of the multiple-scattering correction factor and its cross-sensitivity
901 to scattering and wavelength dependence for different AE33 Aethalometer filter tapes: A multi-instrumental approach,
902 *Atmos. Meas. Tech.*, 14, 6335-6355, <https://doi.org/10.5194/amt-14-6335-2021>, 2021.

903 Zefirov, V. V., Elmanovich, I. V., Levin, E. E., Abramchuk, S. S., Kharitonova, E. P., Khokhlov, A. A., Kondratenko, M. S.,
904 and Gallyamov, M. O.: Synthesis of manganese oxide electrocatalysts in supercritical carbon dioxide, *J. Mater. Sci.*, 53,
905 9449-9462, <https://doi.org/10.1007/s10853-018-2242-3>, 2018.

906 Zhang, K., Allen, G., Yang, B., Chen, G., Gu, J., Schwab, J. J., Felton, D., and Rattigan, O.: Joint measurements of PM_{2.5} and
907 light-absorptive PM in woodsmoke-dominated ambient and plume environments, *Atmos. Chem. Phys.*, 17, 11441-11452,
908 <https://doi.org/10.5194/acp-17-11441-2017>, 2017.

909 Zhang, X., Trzepla, K., White, W., Raffuse, S., and Hyslop, N. P.: Intercomparison of thermal–optical carbon measurements
910 by Sunset and Desert Research Institute (DRI) analyzers using the IMPROVE_A protocol, *Atmos. Meas. Tech.*, 14, 3217-
911 3231, <https://doi.org/10.5194/amt-14-3217-2021>, 2021.

912 Zhu, J., Crozier, P. A., and Anderson, J. R.: Characterization of light-absorbing carbon particles at three altitudes in East Asian
913 outflow by transmission electron microscopy, *Atmos. Chem. Phys.*, 13, 6359-6371, <https://doi.org/10.5194/acp-13-6359-2013>, 2013.

914
915 Zotter, P., Herich, H., Gysel, M., El-Haddad, I., Zhang, Y., Močnik, G., Hüglin, C., Baltensperger, U., Szidat, S., and Prévôt,
916 A. S. H.: Evaluation of the absorption Ångström exponents for traffic and wood burning in the Aethalometer-based source
917 apportionment using radiocarbon measurements of ambient aerosol, *Atmos. Chem. Phys.*, 17, 4229-4249,
918 <https://doi.org/10.5194/acp-17-4229-2017>, 2017.

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921 **Table 1.** Classification of the campaign period during May 05–June 09, 2021.

Name	Date	Cruise route	Monsoon	Wind direction
BMP-1	May 05–09	AB	before	0–90°, northeast
BMP-2	May 10–22	B→C→D	before	90–180°, southeast
BMP-3	May 23–26	D→E	before	around 180°, southeast
TMP	May 27–Jun 01	EB	transition	around 180°, south
AMP	June 02–09	B→D→A	after	180–270°, southwest
SPP*	Screened	Screened	N/A	Screened

922 * Ship pollution period is screened based on BC concentration and relative wind direction as mentioned in the method section.

923

924 **Table 2.** Source apportionment of BC based on the two-component AAE model.

Period	a.BC(FF)* ($\mu\text{g m}^{-3}$)	a.BC(BB)* ($\mu\text{g m}^{-3}$)	r.BC(BB)* ($\mu\text{g m}^{-3}$)	r.BC(FF)* ($\mu\text{g m}^{-3}$)	f.BC(FF)* (%)	f.BC(BB)* (%)
BMP	0.2 ± 0.1	0.08 ± 0.06	0–0.9	0–0.3	77.9 ± 5.8	22.1 ± 5.8
TMP	0.2 ± 0.3	0.06 ± 0.1	0.02–1.8	0–1.1	82.2 ± 6.2	17.8 ± 6.2
AMP	0.2 ± 0.1	0.05 ± 0.05	0.01–0.9	0–0.3	80.8 ± 4.0	19.2 ± 4.0
SPP	4.4 ± 5.7	0.7 ± 0.9	0.2–32.8	0.04–10.3	83.0 ± 6.7	17.0 ± 6.7
Bio. **	0.8 ± 0.3	0.8 ± 0.4	0.05–1.2	0–1.5	58.1 ± 16.7	41.9 ± 16.7

925 * a represents average, r for range, f for fraction;

926 ** Bio. stands for the two biomass burning events as noted in the main text.

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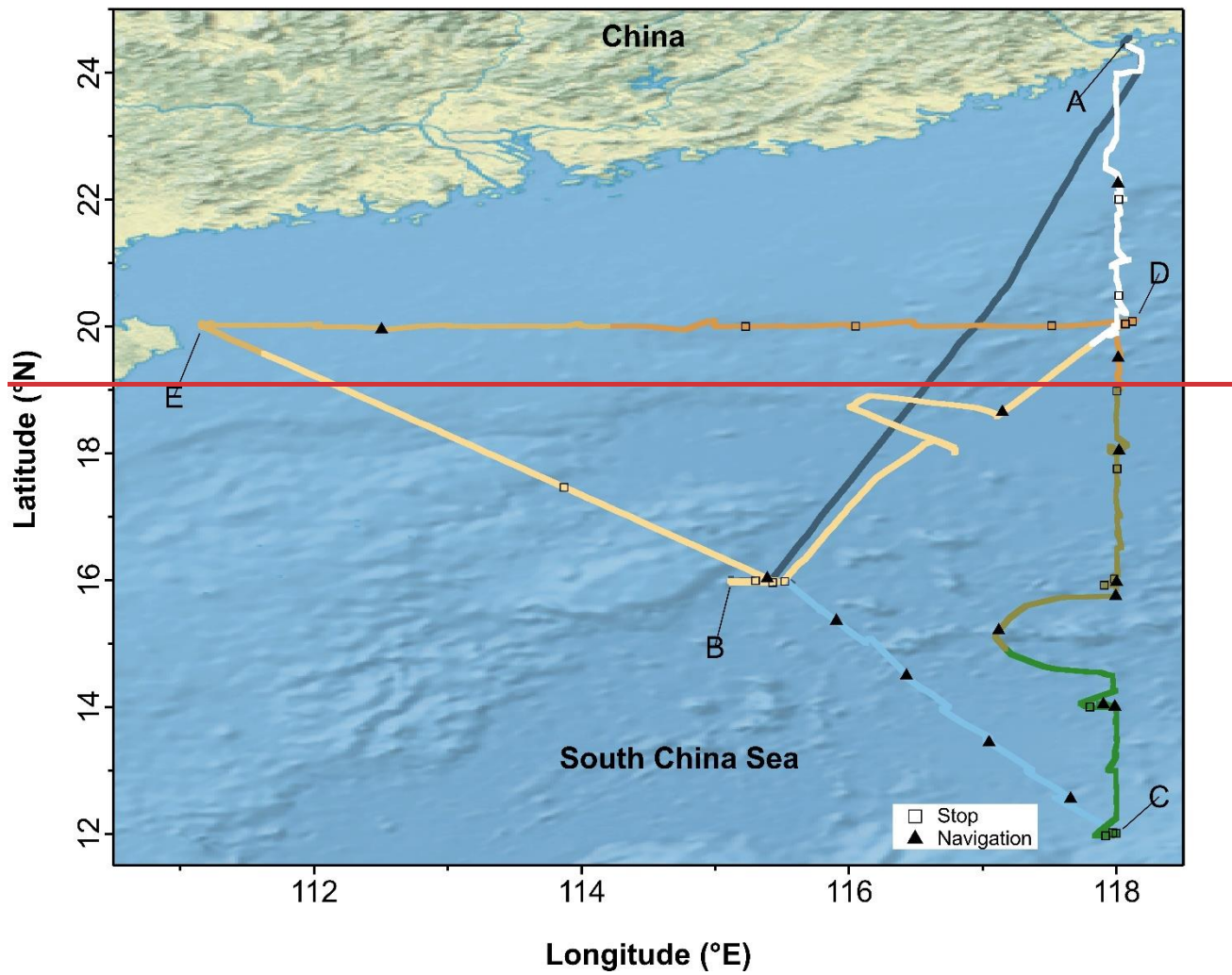
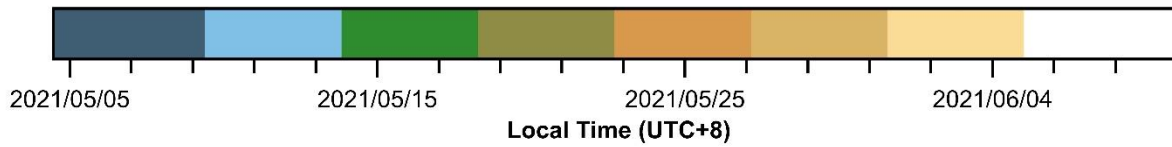
928 **Table 3.** Summary of AAE values, mass concentrations, and the relevant fractions for the BC particles in PM_{2.5} from coast site
 929 and cruise measurements using the AE series aethalometer.

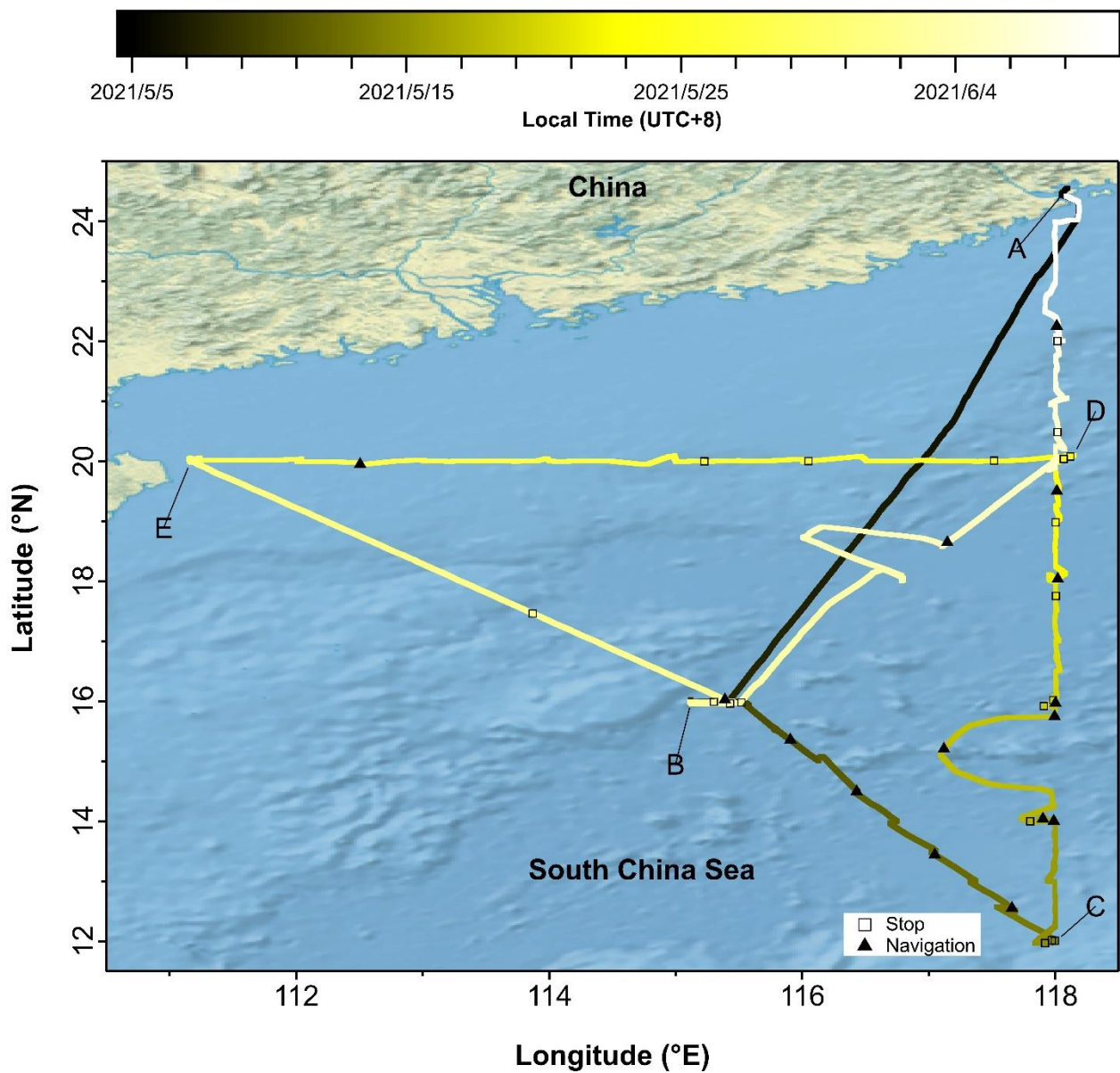
Region	Time	AAE values (at wavelengths, nm)	BC avg. conc. ($\mu\text{g m}^{-3}$)	Fraction (%)	Reference
East China Sea ^c	May, 2017	0.9–1.3 (370–950)	0.8–3.6	2.5–11 or 45–60 for BrC	(Yu et al., 2018)
Central Adriatic ^c	Feb.–Jul., 2019	1.25–1.49 (470/950)	0.57 ± 0.64	79 for BC(FF) 21 for BC(BB)	(Milinković et al., 2021)
Bay of Bengal ^{c, n}	Dec., 2008–Jan., 2009	1.81–1.98 (370–950)	5.1 ± 3.0^c ; 2.5 ± 1.4^n	<10 for BC(FF) >85 for BC(BB)	(Kedia et al., 2012)
South China Sea ⁿ	Sep.–Oct. 2019	-	1.9 ± 0.4	-	(Wang et al., 2022)
South China Sea ^c	Dec. 2017	1.2–1.5 (375/880)	6.6–4.9	-	(Sun et al., 2020c)
South China Sea ⁿ	May–Jun. 2021	1.02–1.14 or 1.93 (370–950)	0.33 ± 0.38	78–82 for BC(FF) 18–22 for BC(BB) BB events: 58 for BC(FF) 42 for BC(BB)	This study

930 ^c coastal site measurement;

931 ⁿ cruise (remote) measurement.

932





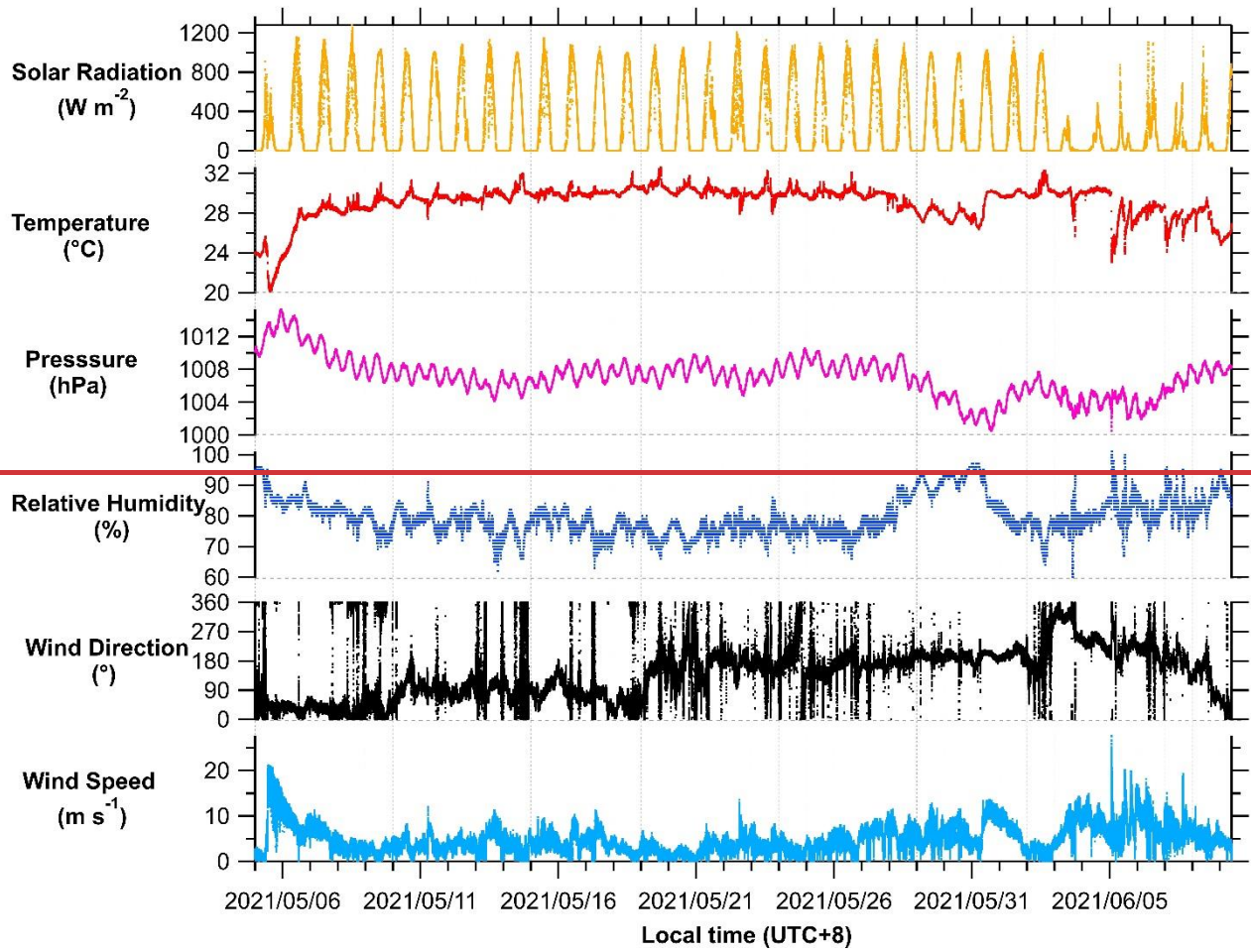
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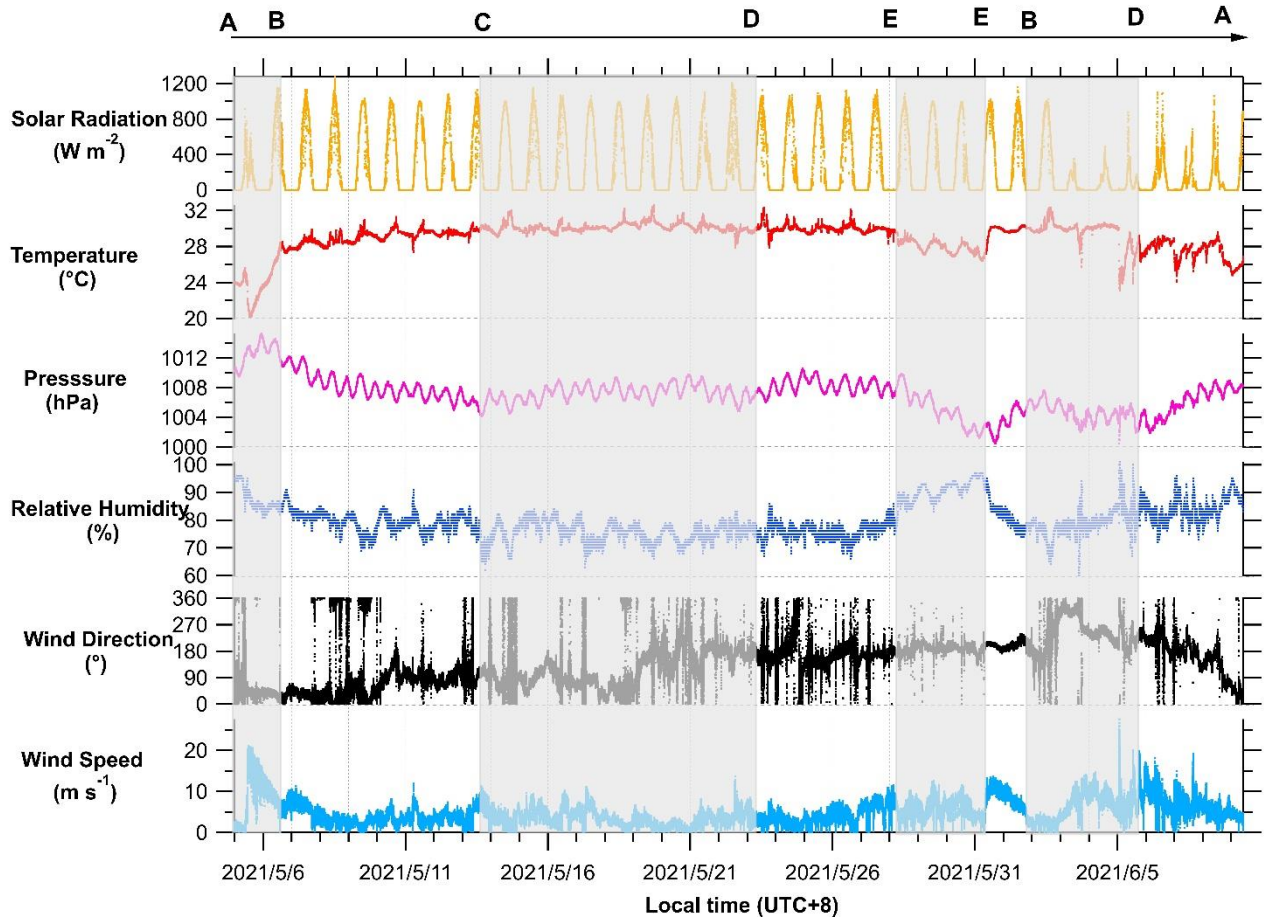
935 Figure 1. Map of the cruise route for the campaign in the South China Sea during May 05–Jun 9, 2021. The ship route is dated by
 936 the intensity bar at the top. The open squares and solid triangles indicate the single particle sampling location, collected during stop
 937 and navigation, respectively.

938 ~~Figure 1. Map of the cruise route for the campaign in the South China Sea during May 05–Jun 9, 2021. The ship route is dated by~~
939 ~~the intensity bar at the top. The open squares and solid triangles indicate the single particle sampling location, collected during stop~~
940 ~~and navigation, respectively.~~

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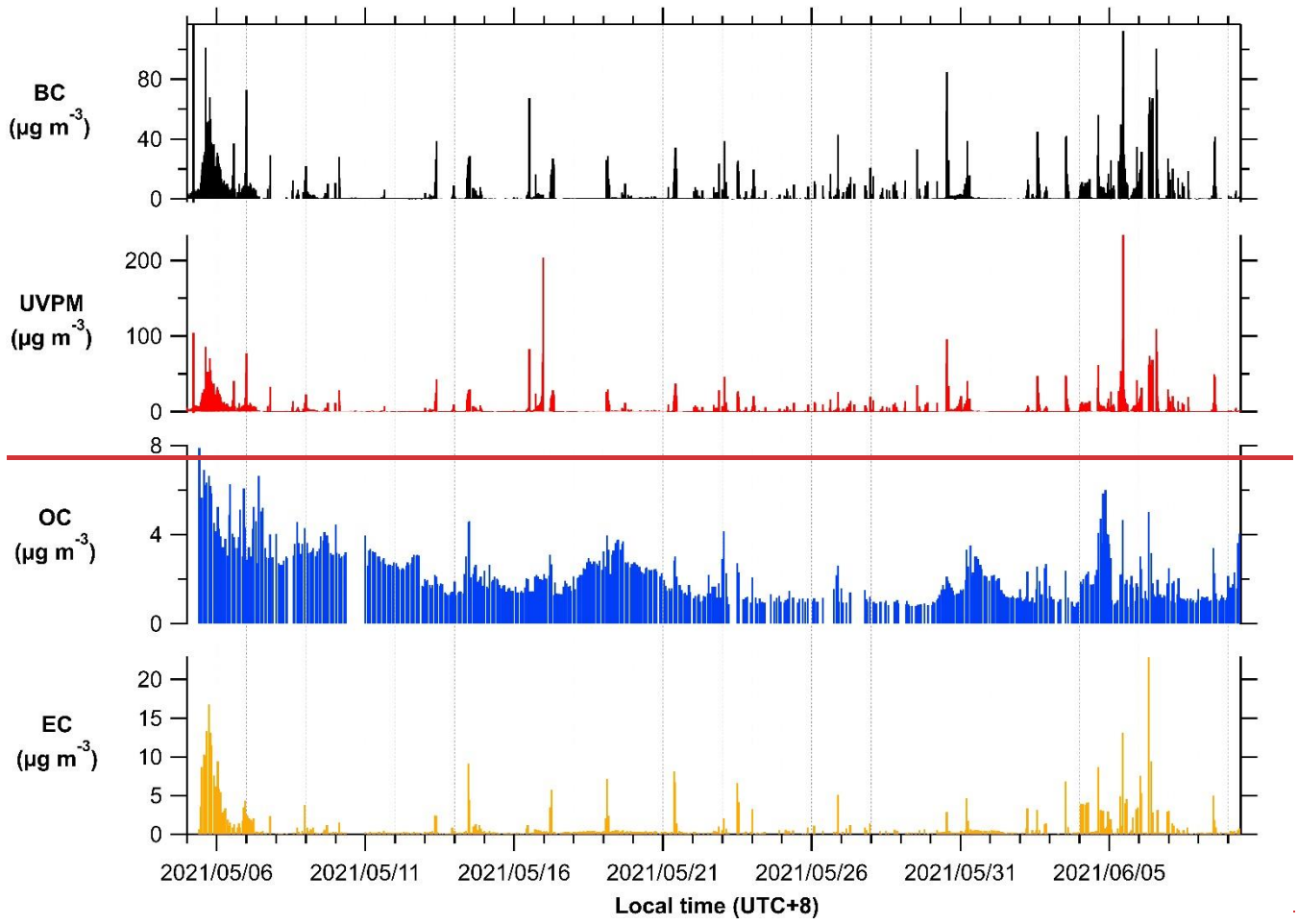
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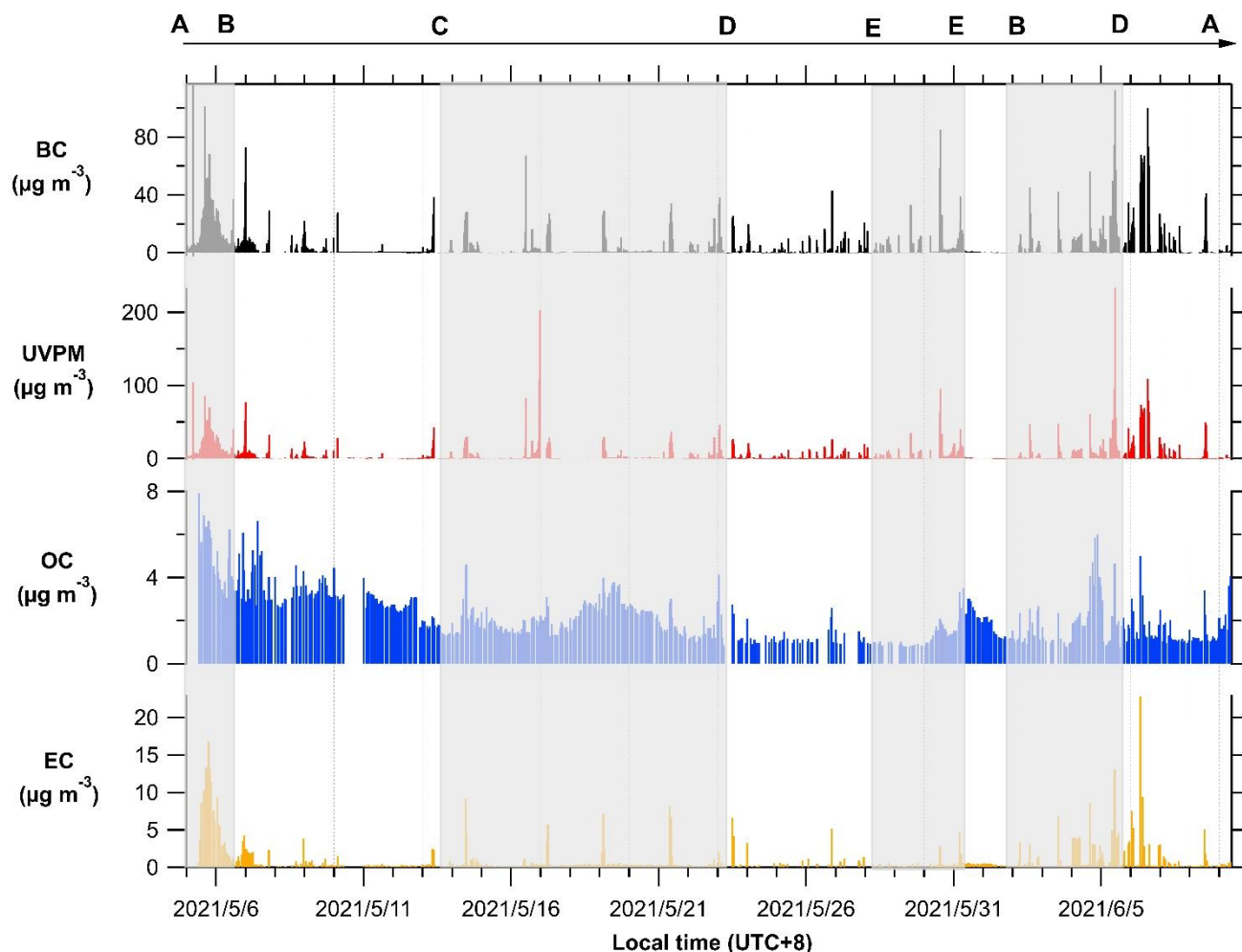
946 **Figure 2. Time series of meteorological variables of solar radiation (SR), temperature, pressure, relative humidity (RH), wind**
 947 **direction (WD), and wind speed (WS) during the campaign. The time resolution is 1 min for all the data except for WS and WD (3**
 948 **sec). All data points are shown in dots style. The shaded and unshaded areas sequentially indicate the cruise routes from AB, B to**
 949 **C, C to D, D to E, E to E (ship stop), E to B, B to D, and D to A, as marked in the Figure 1.**

950 **Figure 2. Time series of meteorological variables of solar radiation (SR), temperature, pressure, relative humidity (RH), wind**
 951 **direction (WD), and wind speed (WS) during the campaign. The time resolution is 1 min for all the data except for WS and WD (3**
 952 **sec). All data points are shown in dots style.**

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955

956 **Figure 3. Time series of the mass concentration of black carbon (BC), ultraviolet particle matter (UVPM), organic carbon (OC)**

957 **and elemental carbon (EC) in PM_{2.5} during the campaign. The time resolution is 1 min for all the data except for OC and EC (1 h).**

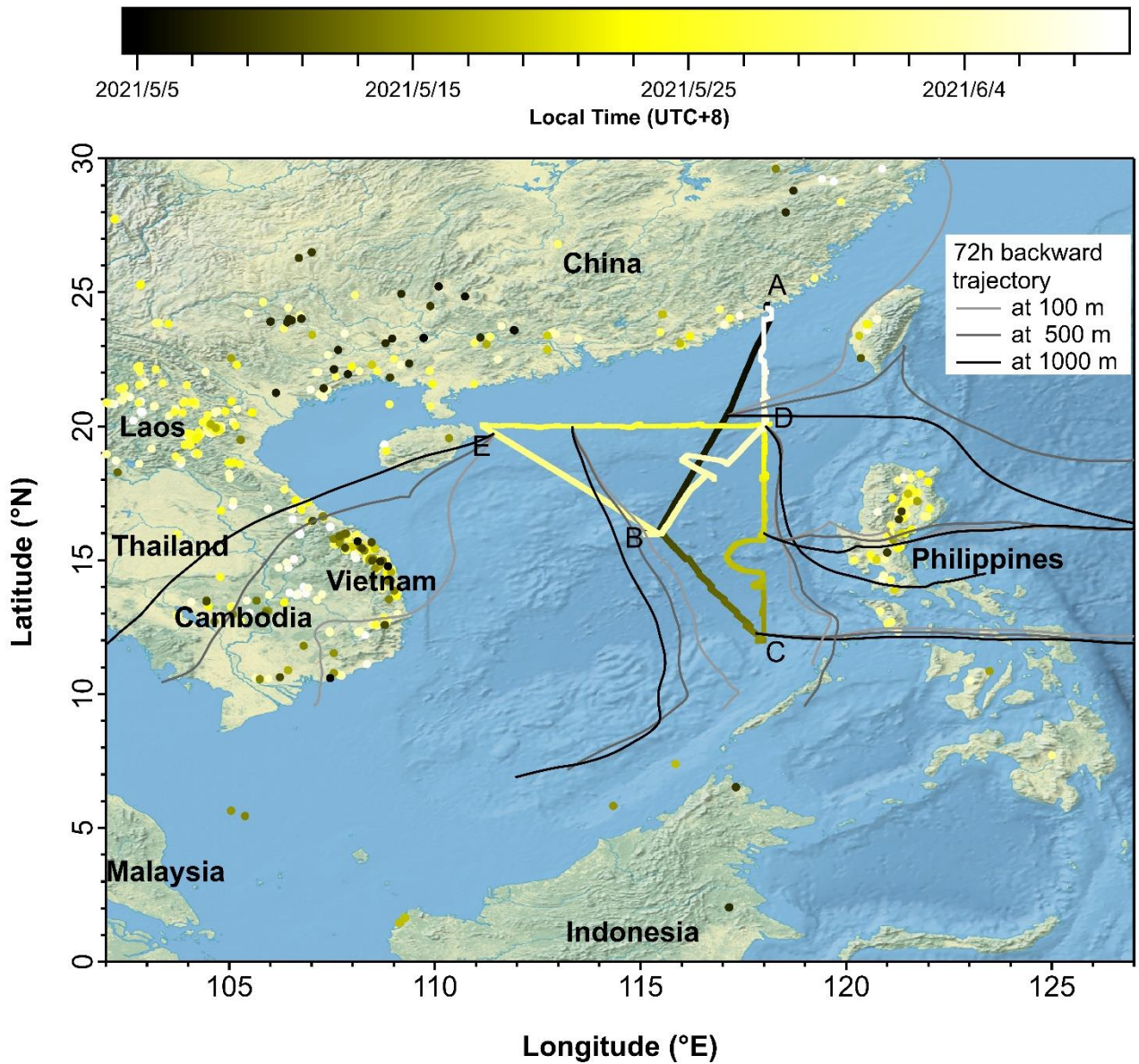
958 **All data points are shown in stick-to-zero style. The shaded and unshaded areas sequentially indicate the cruise routes from AB, B**

959 **to C, C to D, D to E, E to E (ship stop), E to B, B to D, and D to A, as marked in the Figure 1.**

960 **Figure 3. Time series of the mass concentration of black carbon (BC), ultraviolet particle matter (UVPM), organic carbon (OC)**

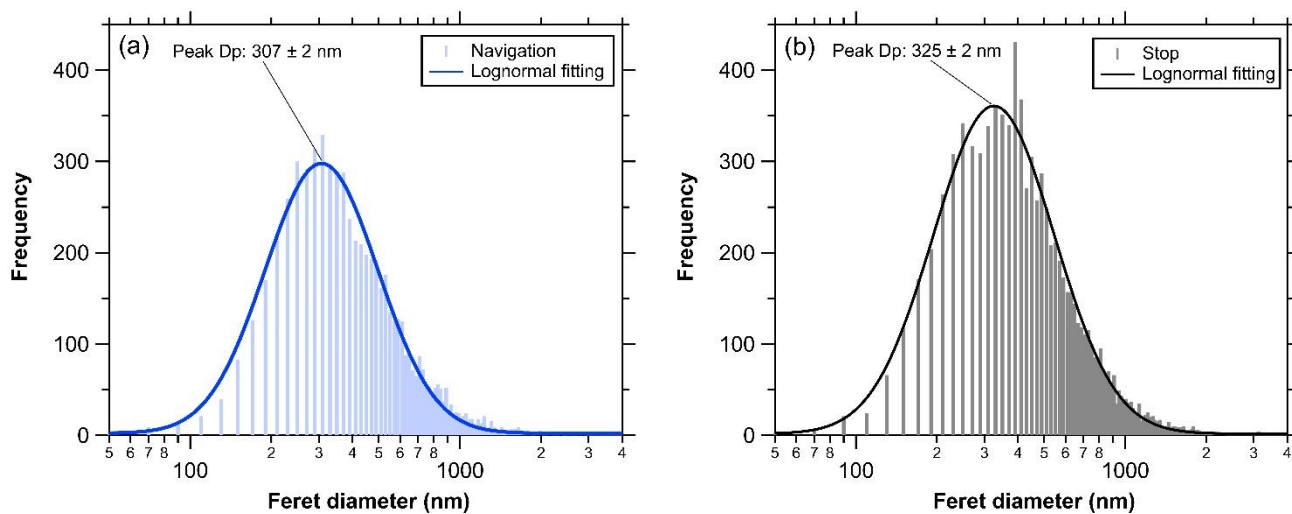
961 **and elemental carbon (EC) in PM_{2.5} during the campaign. The time resolution is 1 min for all the data except for OC and EC (1 h).**

962 **All data points are shown in stick-to-zero style.**



963

964 Figure 4. The time series of fire spots distribution and the 72-h backward trajectories over the South China Sea (SCS) during the
 965 campaign. The solid circles were dated by the intensity bar at the top, representing the detected fire spots using MODIS satellite
 966 with a confidence threshold of >80%. The light grey, grey and black 72-h backward trajectories were obtained at AGL heights of
 967 100, 500 and 1000 m using the HYSPLIT model.



969

970 **Figure 5** ~~Figure 4~~. Lognormal fitting of particle size distribution using Feret diameter determined from the TEM images with the
 971 ImageJ analysis during navigation (a) and stop (b). The histograms are set with a bin starting at 50 nm, a bin width of 20 nm, and a
 972 total bin number of 200.

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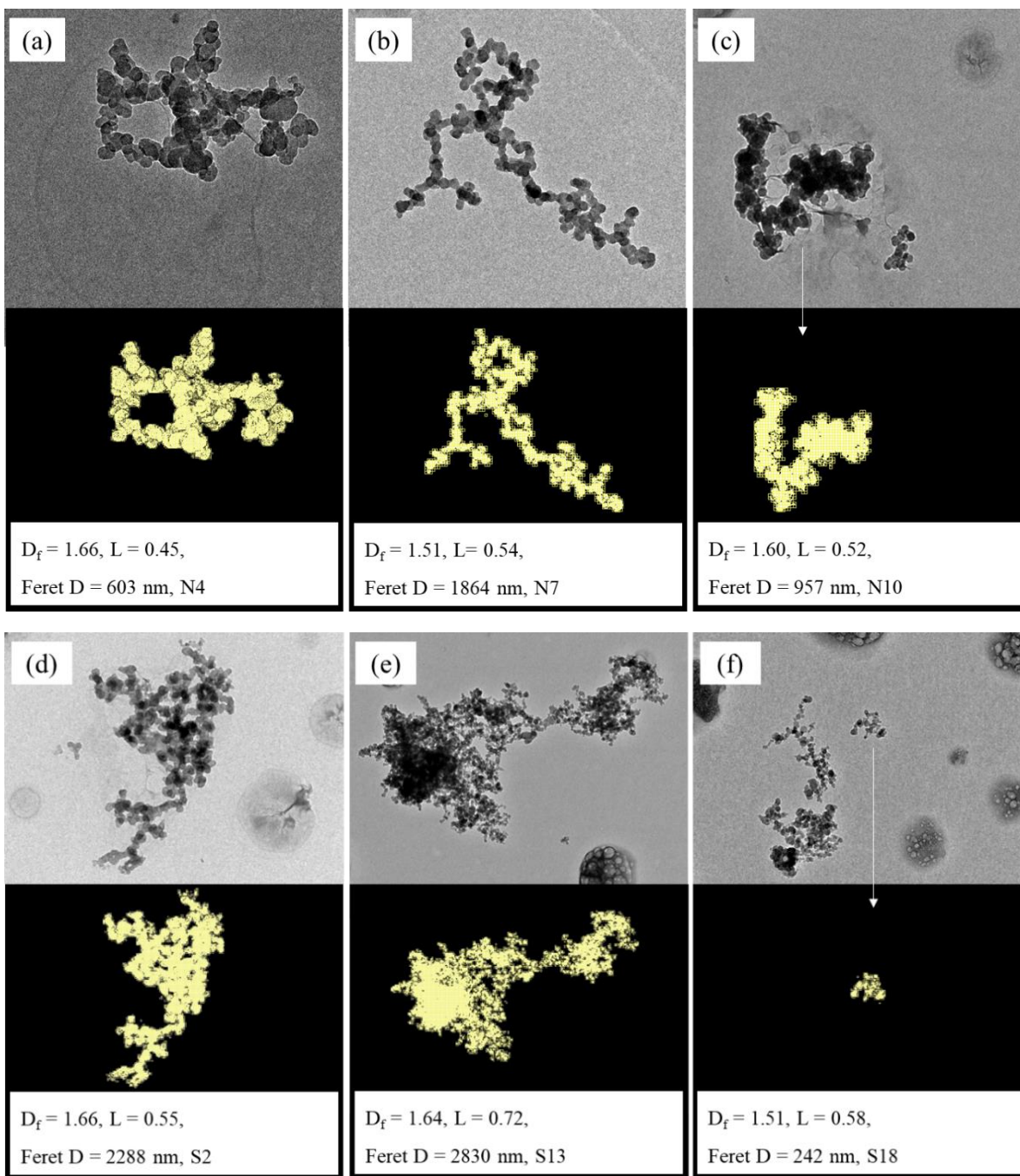
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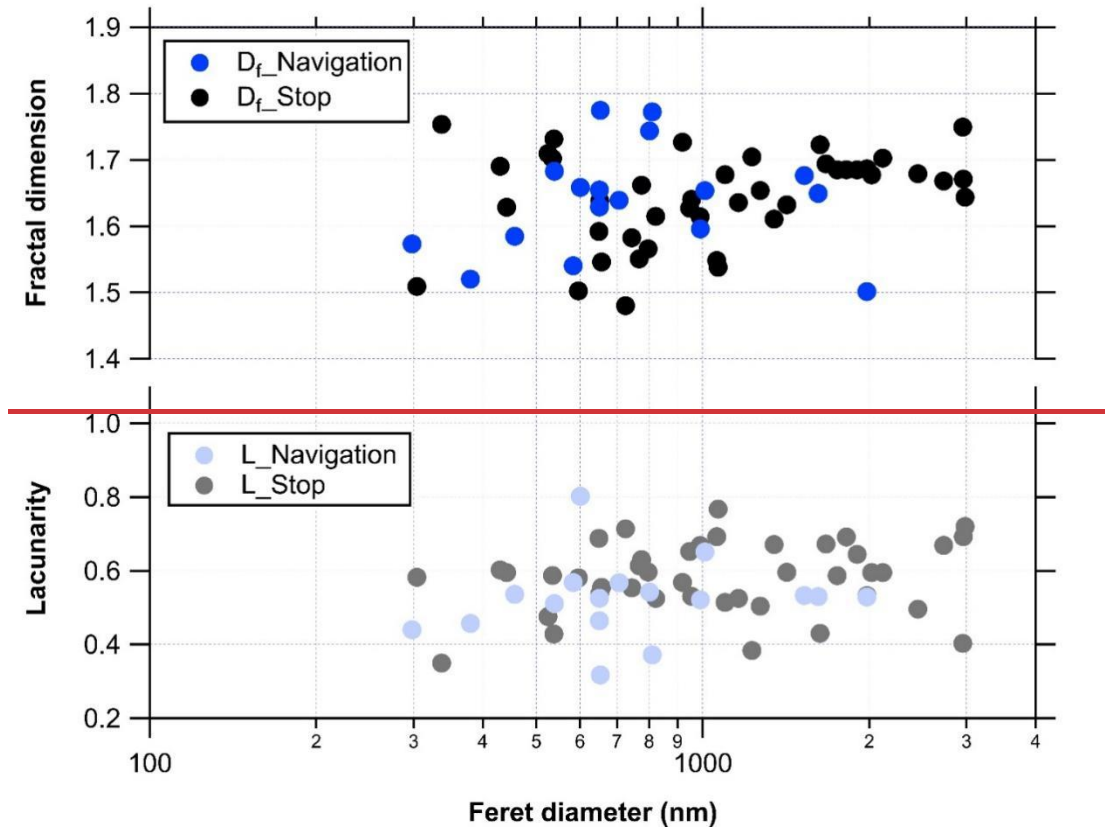
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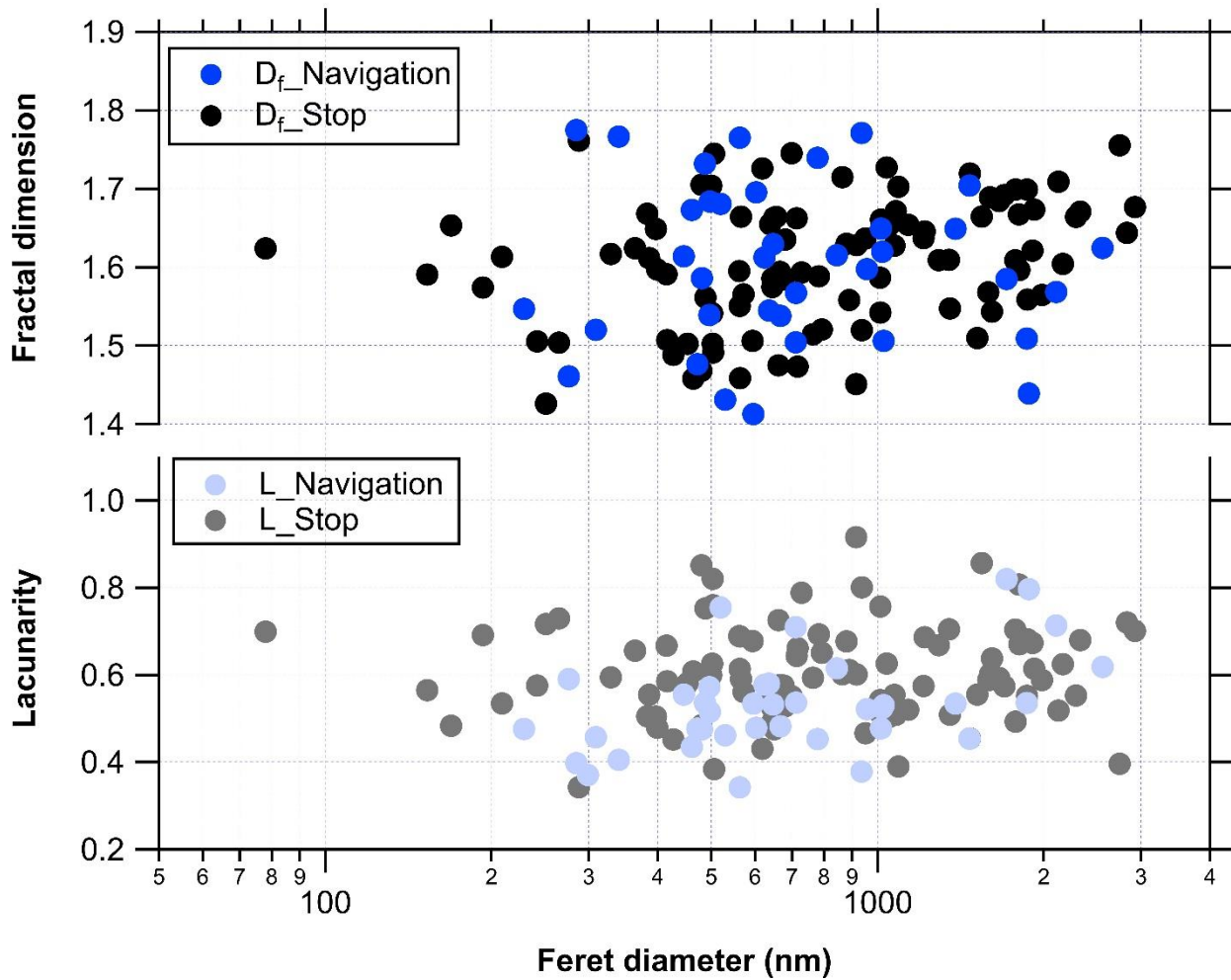
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980

981 **Figure 6Figure 5.** Examples of the BC TEM images and their corresponding Feret diameter (D), fractal dimension (D_f) and
 982 Lacunarity (L) based on the boxing counting method from the fractal analysis: (a-c) BC particles collected during navigation and
 983 (d-f) during stop. More sampling information can be found in Table S1 of SI (serial numbers N4, N7, N10, S2, S13, S18, etc.).



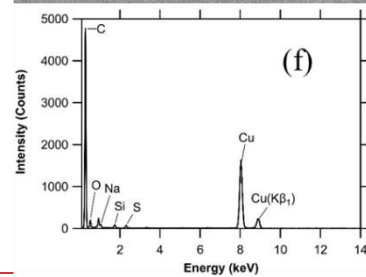
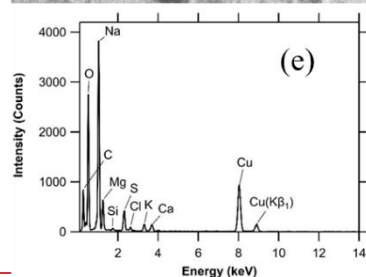
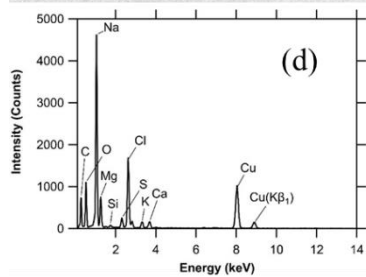
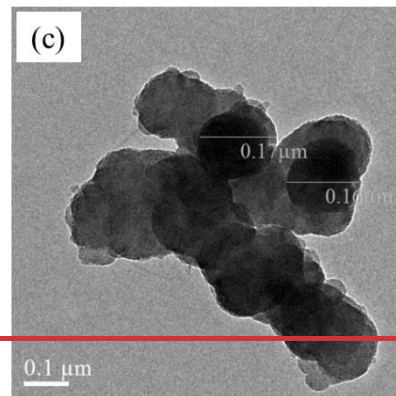
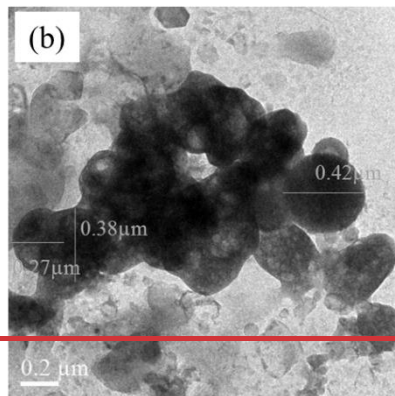
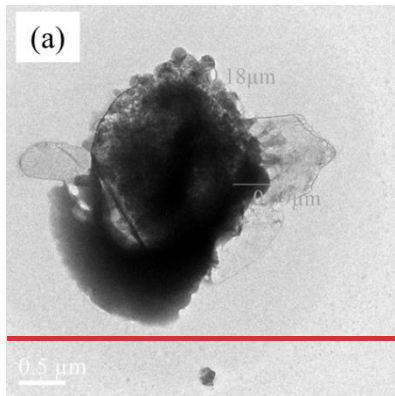
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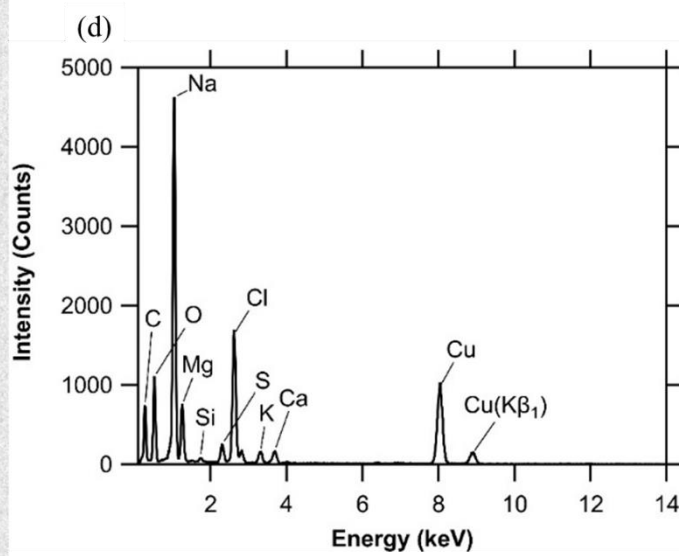
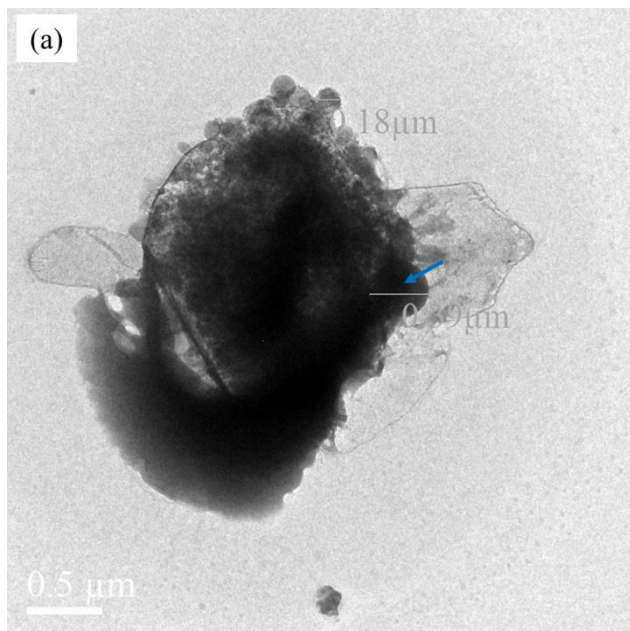
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986 Figure 7. The size-dependent fractal dimension (D_f) and lacunarity (L) for each BC particle during navigation and stop. A total
 987 number of 134 data points are shown in Figure 6.

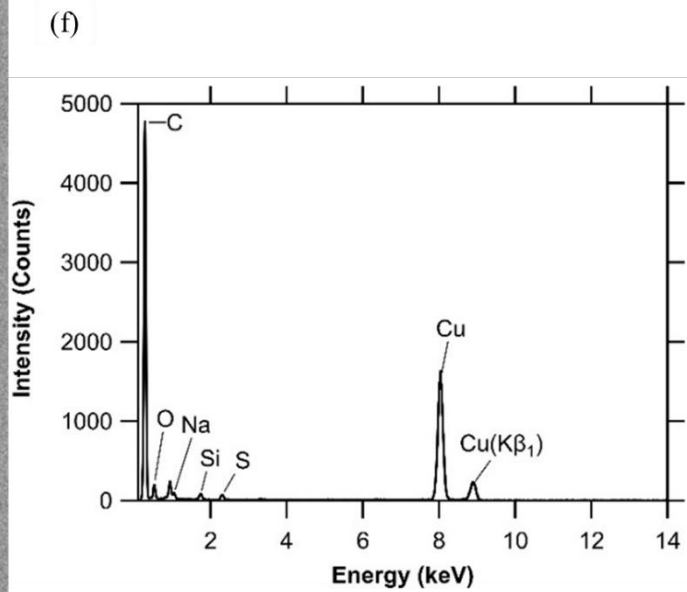
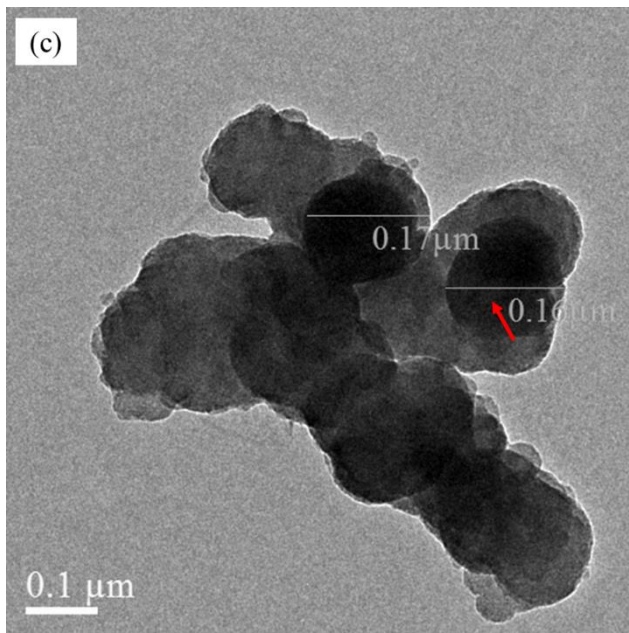
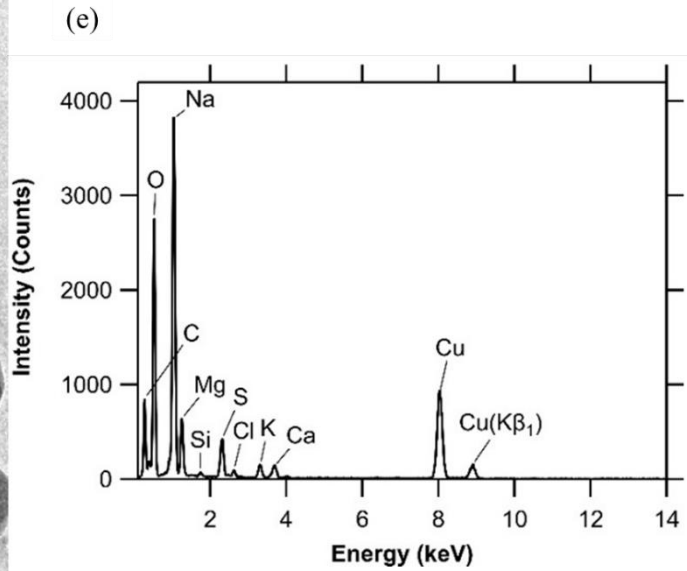
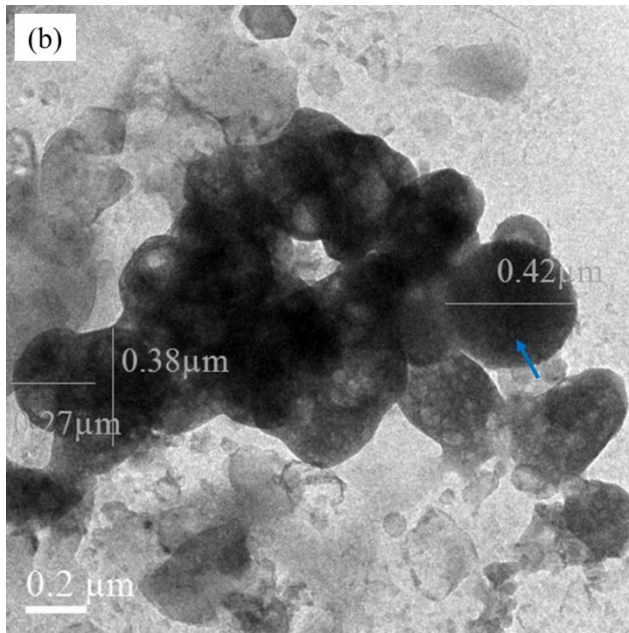
988 ~~Figure 6. The size-dependent fractal dimension (D_f) and lacunarity (L) for each BC particle during navigation and stop.~~



989



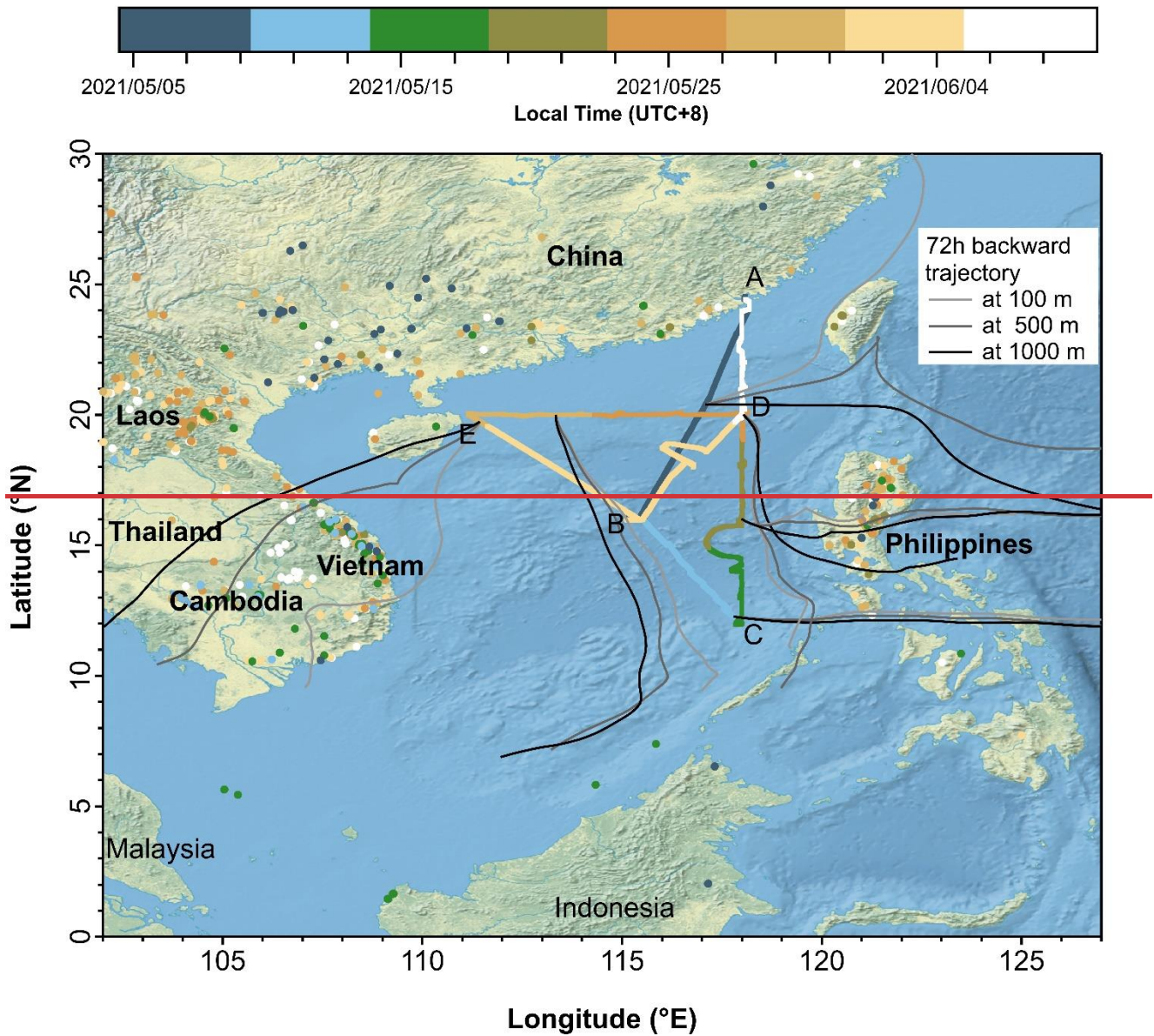
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993 **Figure 8. The example TEM images and their corresponding EDS spectra of particles collected during navigation: tar balls mixed**
994 **with sea salt (a, d) at 8:55 on May 27, tar balls mixed with OC, sulfate (b, e), and amorphous carbon agglomerates mixed with**
995 **sulfate (c, f) at 18:07 on Jun 01. The EDS spectra were collected by focusing the electron beam in the TEM and the illuminated area**

996 covers the center of the particle for elemental analysis. The blue arrows indicate tar balls and the red arrow indicates amorphous
997 carbon. The EDS is obtained from beam focus on the center of the particle.

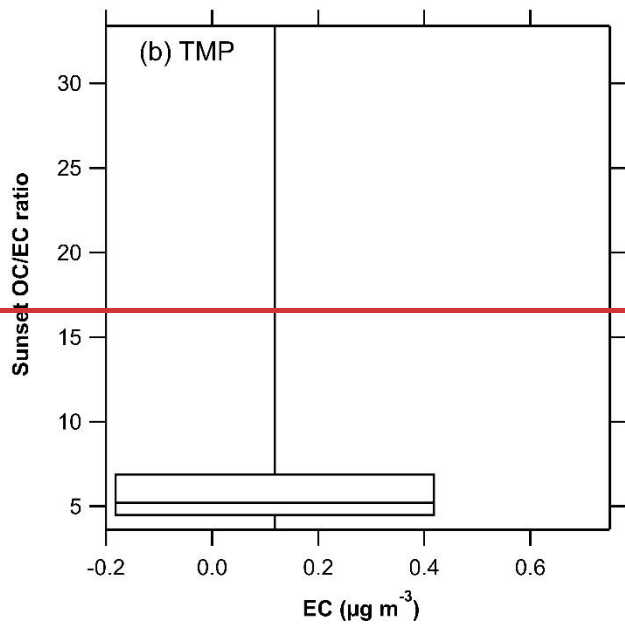
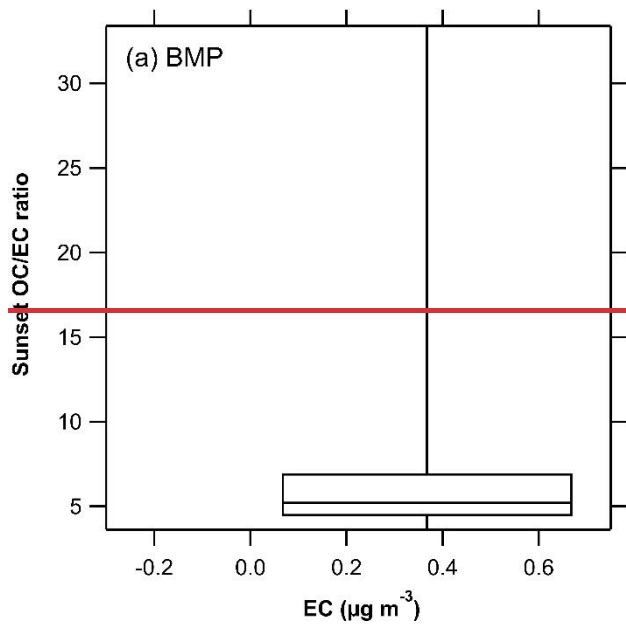
998 ~~Figure 7. The example TEM images and their corresponding EDS spectra of tar ball particles collected during navigation: tar balls~~
999 ~~mixed with sea salt (a, d) at 8:55 on May 27, tar balls mixed with OC, sulfate (b, c), and tar balls mixed with sulfate (e, f) at 18:07~~
1000 ~~on Jun 01. The EDS spectra were collected by focusing the electron beam in the TEM and the illuminated area covers the particle~~
1001 ~~for elemental analysis.~~



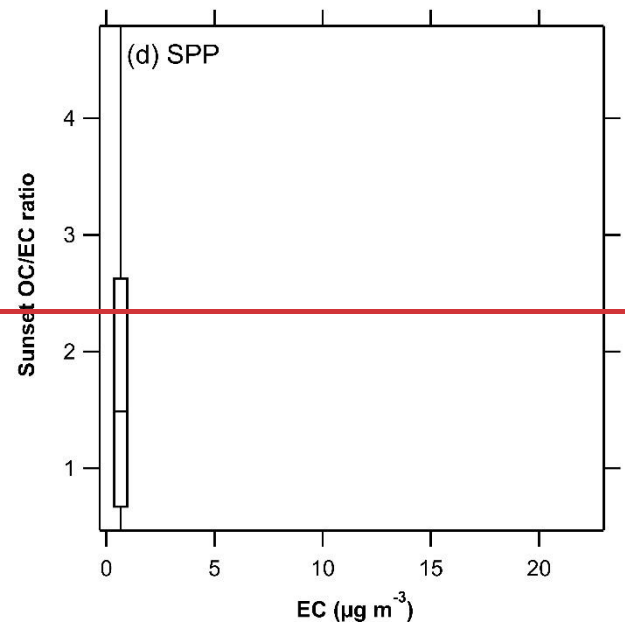
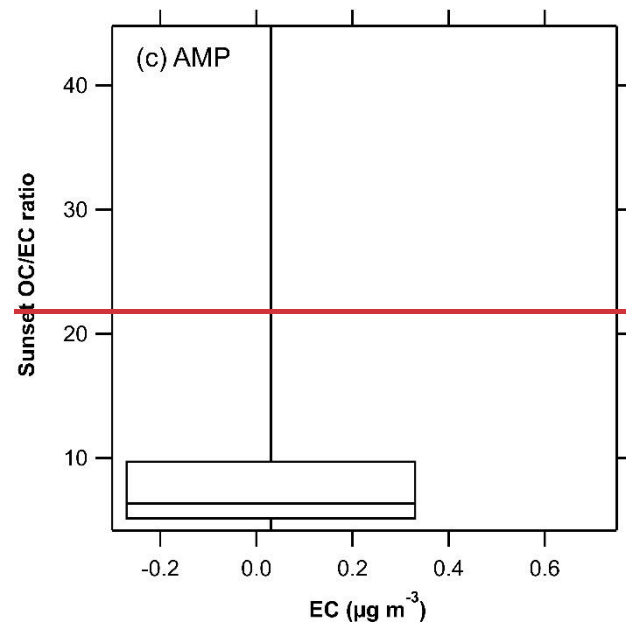
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Figure 8. The time series of fire spots distribution and the 72-h backward trajectories over the South China Sea (SCS) during the campaign. The solid circles were dated by the intensity bar at the top, representing the detected fire spots using MODIS satellite with a confidence threshold of >80%. The light grey, grey and black 72-h backward trajectories were obtained at AGL heights of 100, 500 and 1000 m using the HYSPLIT model.

1007

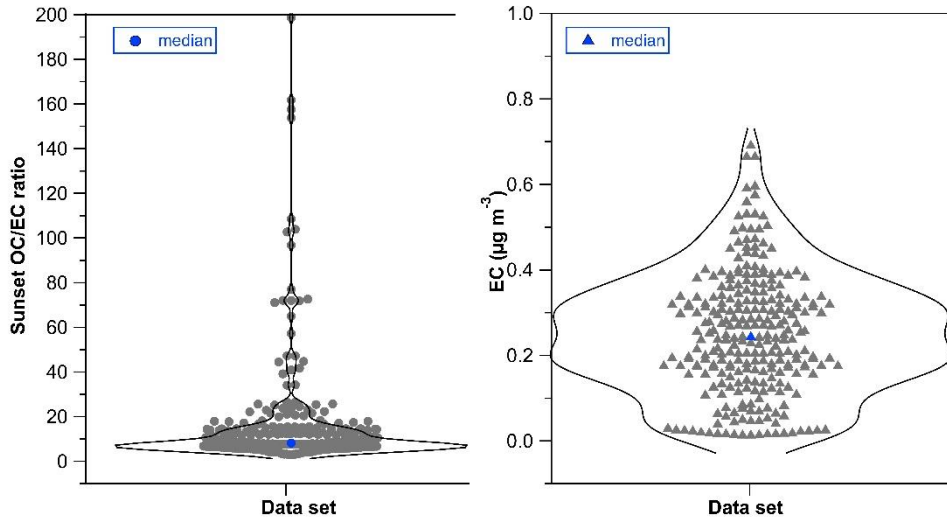


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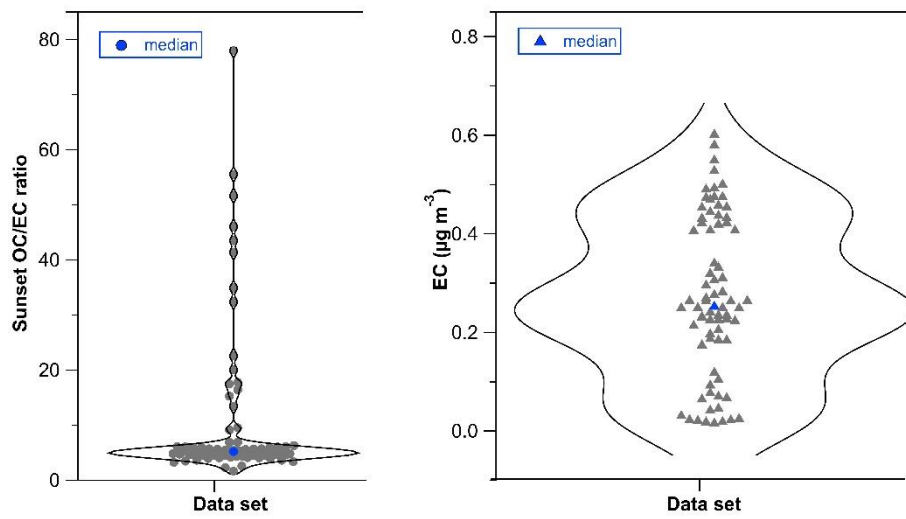
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(a) BMP



1010

(b) TMP



1011

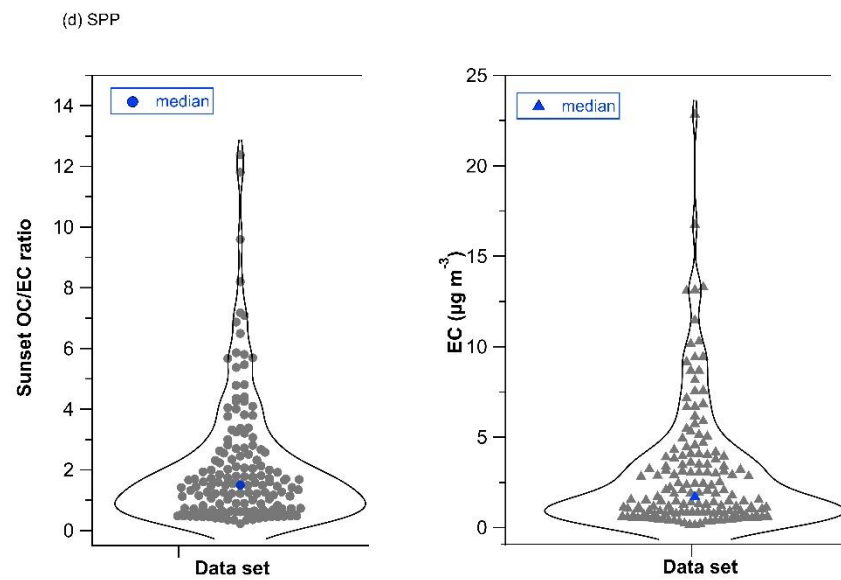
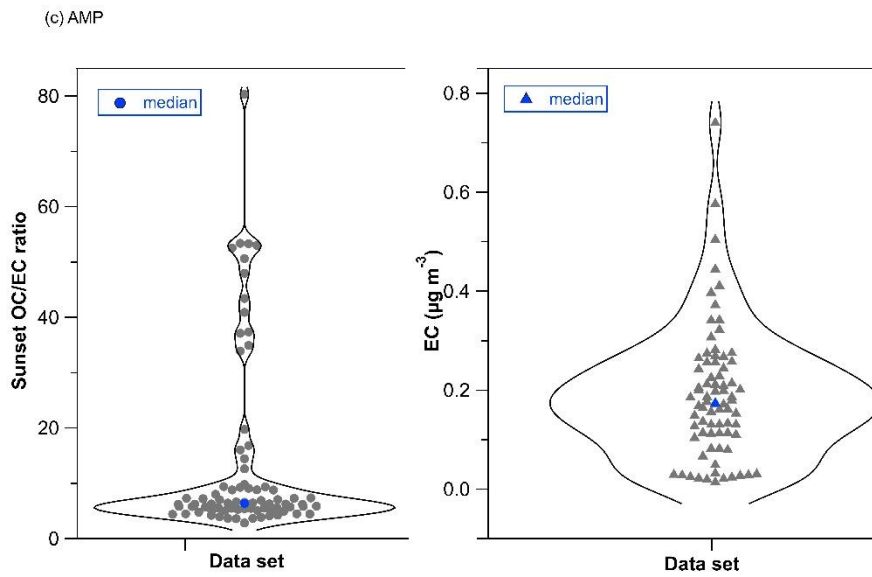
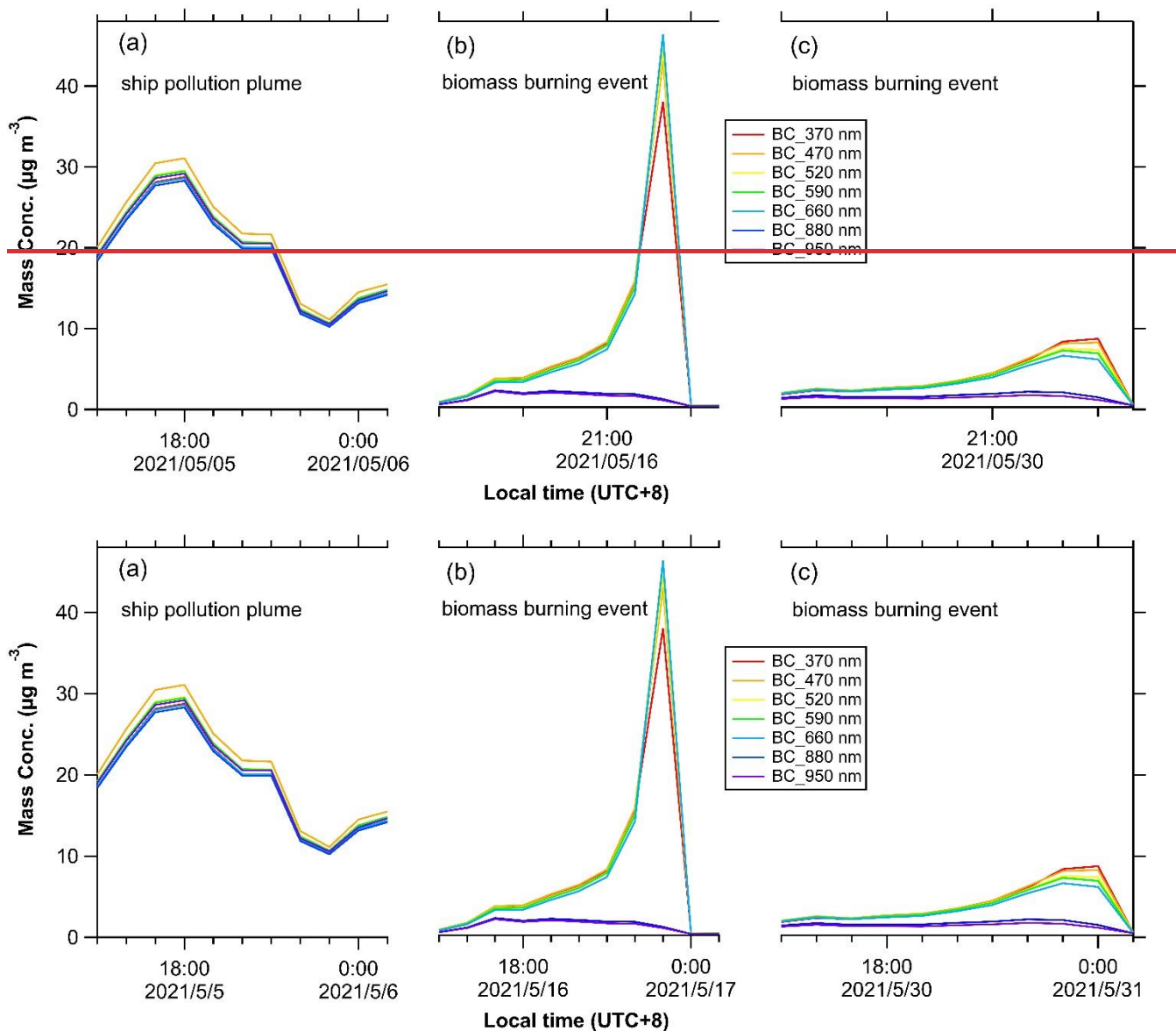


Figure 9. Violin plots of the OC/EC ratios and EC concentrations for (a) before the monsoon period (BMP), (b) transition monsoon period (TMP), (c) after the monsoon period (AMP), and (d) ship pollution period (SPP) based on the data from the Sunset OC/EC analyzer. The blue solid circles and triangles indicate median values of Sunset OC/EC ratios and EC mass concentrations.

1017 respectively. The total data points are 551 in the data set, with a concentration range of 0.76-7.90 $\mu\text{g m}^{-3}$ for OC and 0.013-22.84 for
 1018 EC $\mu\text{g m}^{-3}$. Particularly, all OC data is above LOD of 0.18 $\mu\text{g m}^{-3}$ while 29% of EC data is below LOD of 0.19 $\mu\text{g m}^{-3}$.
 1019 **Figure 9. Box and whisker plots of the OC/EC ratio vs the EC concentration for (a) before the monsoon period (BMP), (b) transition**
 1020 **monsoon period (TMP), (c) after the monsoon period (AMP), and (d) ship pollution period (SPP) based on the data from the Sunset**
 1021 **OC/EC analyzer. The whisker top, box top, box bottom, and whisker bottom are 90%, 75%, 25% and 10%, respectively.**



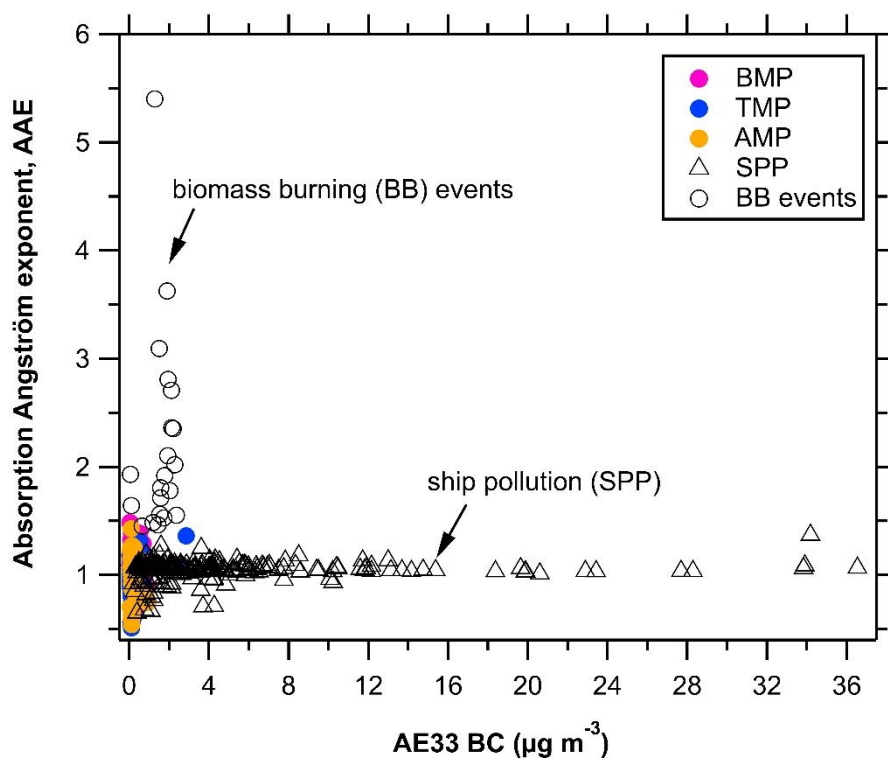
1024 Figure 10. The wavelength-dependent mass concentration from AE33 aethalometer for example spectra for (a) example of a ship
1025 pollution plume, and (b, c) two significant biomass burning events during this campaign.

1026 ~~Figure 10. The wavelength-dependent mass concentration from AE33 aethalometer for example spectra for (a) example of a ship~~
1027 ~~pollution plume, and (b, c) two significant biomass burning events during this campaign.~~

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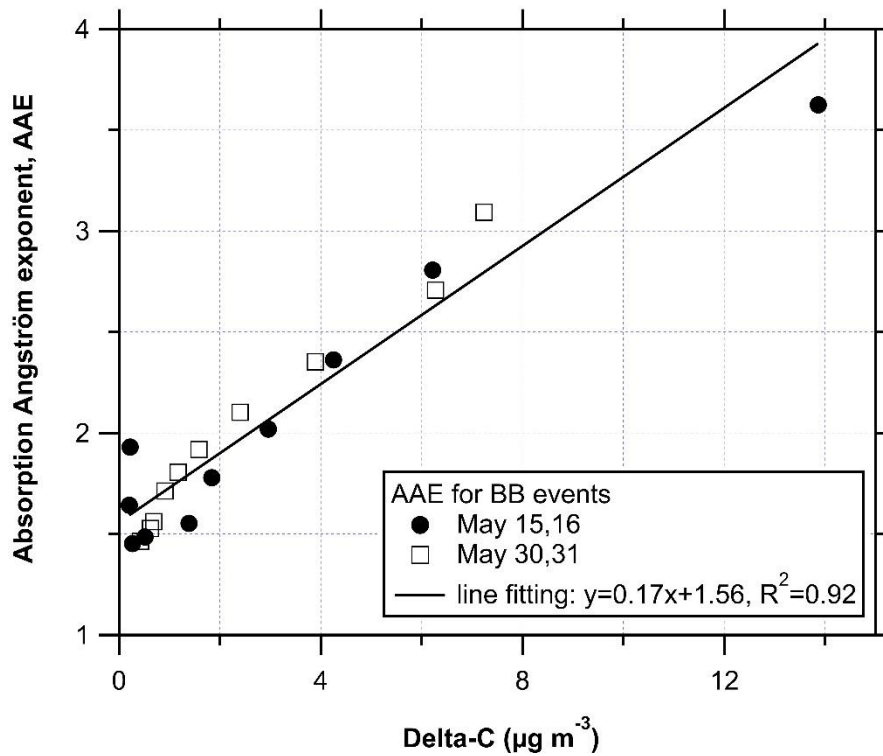
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1032 **Figure 11. The absorption Angström exponent (AAE, all wavelengths) vs AE33 BC concentration for before monsoon period (BMP),**
1033 **transition monsoon period (TMP), after monsoon period (AMP), and ship pollution period (SPP).**

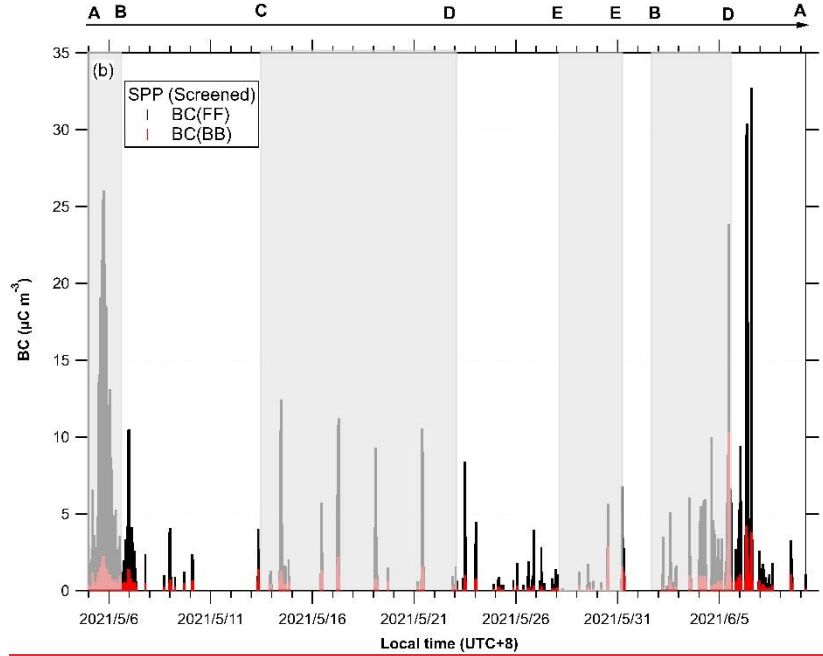
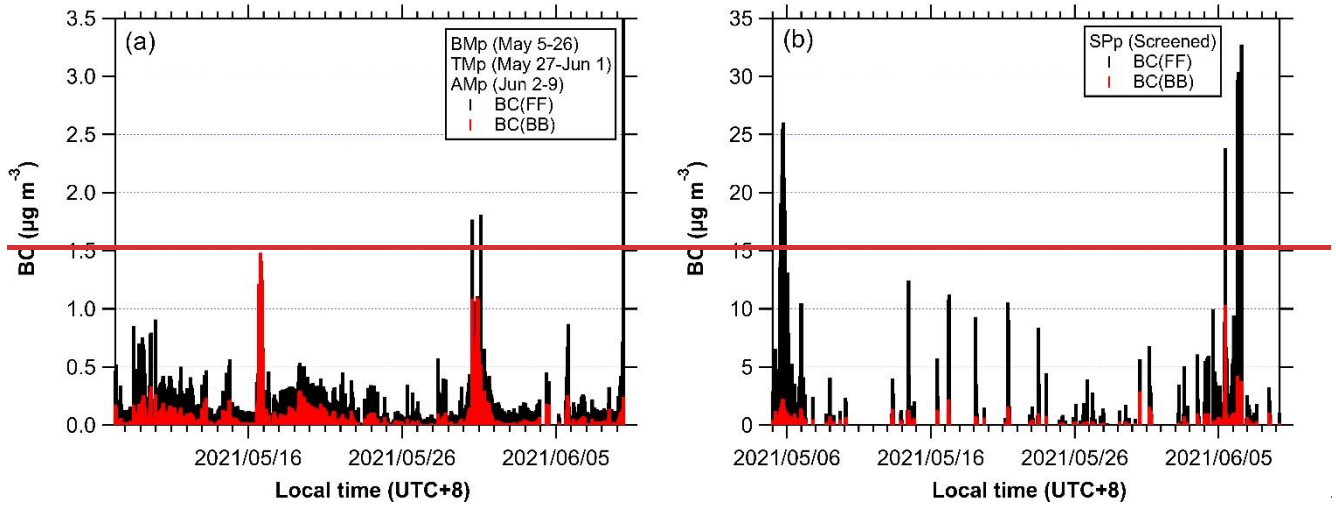
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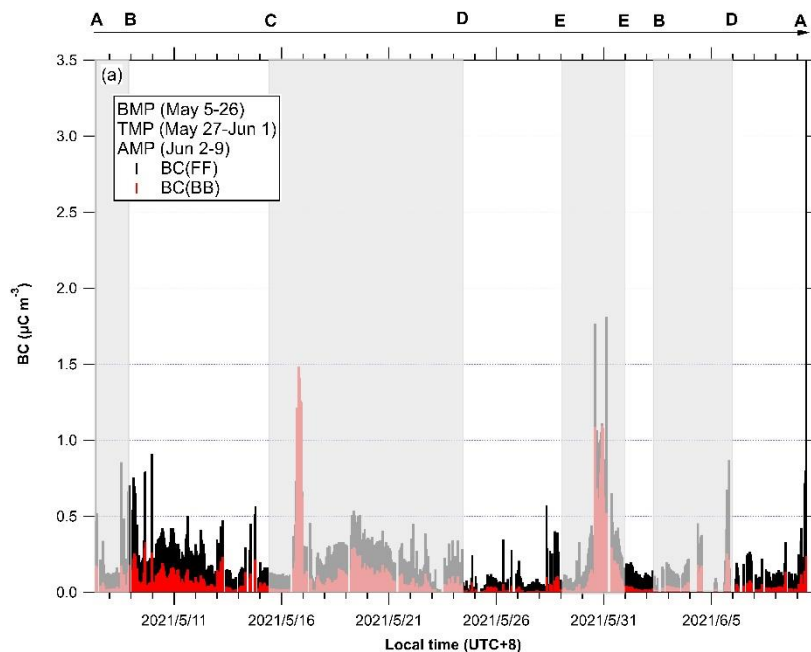


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1036 **Figure 12.** The absorption Angström exponent (AAE) vs the Delta-C concentration for the two biomass burning events: BB-1 at
 1037 6:00–7:00 on May 15 and 15:00–22:00 on May 16 during BMP, and BB-2 at 15:00–23:00 on May 30 and 00:00 on May 31 during
 1038 TMP. The BB-1 and BB-2 data points are marked in solid circles and open squares, respectively.

1039





1043

1044 **Figure 13. Source apportionment of the BC particles using the two-component AAE model (AAE=1 for foil fuel (FF) and AAE=2**
 1045 **for biomass burning (BB): (a) before monsoon period (BMP), transition monsoon period (TMP), after monsoon period (AMP), and**
 1046 **(b) ship pollution period (SPP). The shaded and unshaded areas sequentially indicate the cruise routes from AB, B to C, C to D, D**
 1047 **to E, E to E (ship stop), E to B, B to D, and D to A, as marked in Figure 1.**

1048 **Figure 13. Source apportionment of the BC particles using the two two-component AAE model (AAE=1 for foil fuel (FF) and**
 1049 **AAE=2 for biomass burning (BB): (a) before monsoon period (BMP), transition monsoon period (TMP), after monsoon period**
 1050 **(AMP), and (b) ship pollution period (SPP).**