

# 1 **Tipping points in coupled human-environment system models: a**

## 2 **review**

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7 **Abstract.** Mathematical models that couple human behavior to environmental processes can offer valuable insights  
8 into how human behavior affects various types of ecological, climate, and epidemiological systems. This review  
9 focuses on human drivers of tipping events in coupled human-environment systems where changes to the human  
10 system can lead abruptly to desirable or undesirable new human-environment states. We use snowball sampling  
11 from relevant search terms to review the modeling of social processes—such as social norms and rates of social  
12 change—that are shown to drive tipping events, finding that many affect the coupled system depending on the system  
13 type and initial conditions. For example, tipping points can manifest very differently in human-extraction versus  
14 human-emission systems. Some potential interventions, such as reducing costs associated with sustainable behavior,  
15 have intuitive results. However, their beneficial outcomes via less obvious tipping events are highlighted. Of the  
16 models reviewed, we found that greater structural complexity can be associated with increased potential for tipping  
17 events. We review generic and state-of-the-art techniques in early warning signals of tipping events and identify  
18 significant opportunities to utilize digital social data to look for such signals. We conclude with an outline of  
19 challenges and promising future directions specific to furthering our understanding and informing policy that  
20 promotes sustainability within coupled human-environment systems.

21

22 **Non-technical summary.** Mathematical models that include interactions between humans and the environment can  
23 provide valuable information to further our understanding of tipping points. Many social processes such as social  
24 norms and rates of social change can affect these tipping points in ways that are often specific to the system being  
25 modeled. Higher complexity of social structure can increase the likelihood of these transitions. We discuss how data  
26 is used to predict tipping events across many coupled systems.

## 27 **1 Introduction to tipping points in coupled human-environment systems models**

28 Humans are facing environmental catastrophes of their own making, like climate change and biodiversity declines,  
29 at local and global scales, and yet avoiding these catastrophes still poses complex challenges for sustainable  
30 behavior and policy interventions (Steffen et al., 2017). Traditionally, mathematical models of environmental

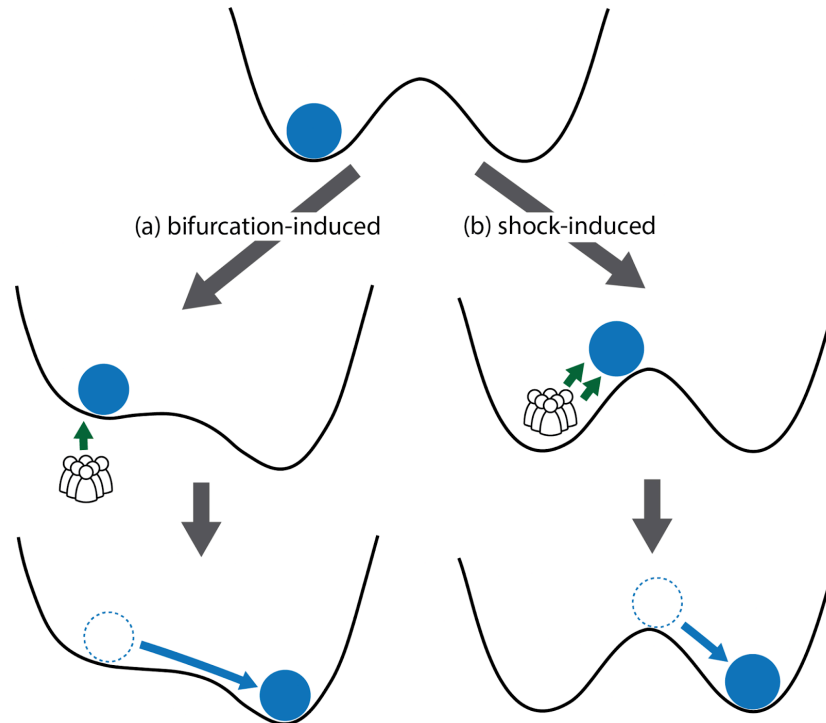
31 systems have represented human impacts through fixed, static parameters or functions independent of the  
32 environment's current state (Binford et al., 1987; Bosch, 1971; Chaudhuri, 1986; Getz, 1980), and these models can  
33 be useful to inform optimal levels of sustainable extraction for short timescales. However, for longer timescales,  
34 where human dynamics can evolve, it may be necessary to include human behavior endemically in the modeling  
35 framework to allow for human-environment feedback to occur (Bauch et al., 2016; Innes et al., 2013; Lade et al.,  
36 2013; Schlüter et al., 2012). Coupled human-environment system (CHES) models combine environmental (e.g.,  
37 ecological, epidemiological, and climate) models with human behavior and population dynamics (Bury et al., 2019;  
38 Carpenter et al., 2009; Farahbakhsh et al., 2022; Innes et al., 2013; Lade et al., 2013; Phillips et al., 2020; Sethi and  
39 Somanathan, 1996). For example, in Innes (2013), the amount of forest cover influences the proportion of the  
40 population that conserves forest ecosystems. The influence of each subsystem on one another often occurs as  
41 two-way (positive and/or negative) feedback loops. In a positive (self-reinforcing) feedback loop, variable 'A'  
42 causes an increase in variable 'B' which then causes an increase in 'A'. In a negative feedback loop, 'A' causes an  
43 increase (respectively, decrease) in 'B' which causes a decrease (respectively, increase) in 'A'. The inclusion of  
44 these feedbacks leads to increased diversity in the qualitative behavior of the system, such as whether the long-term  
45 dynamics converge to a sustainable or depleted environmental state, or cycle over time. Negative feedback promotes  
46 a return to equilibrium (Figure 2a) and can increase the system's capacity to respond to disturbances and adapt in  
47 ways that allow the system to maintain the function of social and ecosystem services, which is sometimes referred to  
48 as "resilience" (Folke, 2006).

49

50 Human-environment negative feedback loops via processes such as public concern pressuring governments to  
51 introduce environmental legislation can be powerful and there are many historical examples of it occurring (Dunlap,  
52 2014; Grier, 1982; Mather and Fairbairn, 2000; Stadelmann-Steffen et al., 2021). Forest cover in Switzerland  
53 doubled, following an all-time low in the first half of the 19th century. This was brought about by public concern  
54 responding to food shortages and floods, which triggered local regulation, the formation of the Swiss Forestry  
55 Society, and the first federal forestry law enacted in 1876 (Mather and Fairbairn, 2000). Similarly, the bald eagle  
56 population in North America recovered significantly after the banning of DDT by the EPA in 1972. This was  
57 instigated by public outcry following the publication of Rachel Carson's *A Silent Spring* in 1962 which linked DDT  
58 in the environment to low reproduction of birds and their declining population (Dunlap, 2014; Grier, 1982). In both  
59 cases, the gradual recovery of the population was not brought about simply by governmental legislation. There were  
60 strong movements in the public and scientific spheres, directly responding to perceived environmental risk which  
61 pressured governing bodies to enact immediate reform (Dunlap, 2014; Grier, 1982; Mather and Fairbairn, 2000). We  
62 interpret these two examples as negative feedback loops in a coupled human-environment system because a decline  
63 in forest/eagle abundance stimulated a response by humans which led to the recovery of the environmental system  
64 (Figure 2a). These negative feedback loops are pervasive in the CHES models that we examine here.

65

66 The historical examples above describe negative feedbacks promoting a return to a single environmentally beneficial  
67 equilibrium; however, in many cases, this does not happen and the system can persist in a depleted state. For  
68 example, the desertification of regions once rich in vegetation could become a positive feedback loop maintaining  
69 the new desert state (Hopcroft and Valdes, 2021; Pausata et al., 2020). When systems can persist in qualitatively  
70 different states (also referred to as “regimes”), we say that they exhibit alternative stable states (May, 1977; Lenton  
71 et al., 2008, Henderson et al. 2016). In mathematical models, alternative stable states are self-reinforcing for a range  
72 of parameters, for example, low harvest rates can promote a state of high biomass and high harvest rates can  
73 promote a state of low biomass in many extractive CHES (Farahbakhsh et al., 2021; Henderson et al., 2016; Richter  
74 and Dakos, 2015; Richter et al., 2013; Schlüter et al., 2016). Tipping points refer to critical points on this boundary  
75 between two alternative stable states. Near this boundary, small perturbations can be amplified through nonlinear  
76 self-reinforcing positive feedback loops. This leads to a qualitatively different system state and characteristic  
77 behavior, known as a “regime shift”, in a relatively short amount of time. When the system has entered a new  
78 regime, there are often positive or negative feedback loops that make it difficult to reverse this change. This  
79 self-perpetuating nature of some initial change through nonlinear feedbacks leading to qualitative and often  
80 long-term system change is a universal characteristic of many commonly studied tipping points. In many cases, a  
81 return to the system's previous state can be more difficult than anticipated, requiring additional effort rather than  
82 merely a return to parameters before the tipping point, a phenomenon known as hysteresis, which can make  
83 mitigation and adaptation efforts challenging. Systems near a tipping point can exhibit (often abrupt) regime shifts  
84 through gradual changes or noise in forcing parameters, which is a main focus of much of the bifurcation theory  
85 literature (Figure 1a, Box 1.1). The scope of models presented in this review will not include other types of tipping  
86 points such as those caused by a short sharp shock (s-tipping, or shock-tipping, where the system does not have to  
87 exist near this point for a regime shift to occur) (Figure 1b) (Boettiger and Batt, 2020; Halekotte and Feudel, 2020)  
88 or “rate-induced tipping”, which is a distinct phenomenon induced by the rate of change of parameters (Ashwin et  
89 al., 2012). Tipping events describe the crossing of a tipping point and can be used interchangeably with regime  
90 shifts.



**Figure 1: Two types of tipping events; bifurcation-induced tipping (a), where the drivers are gradual changes to system parameters leading to a tipping event, and shock-induced tipping (b), where a perturbation to the system causes it to enter an alternative stable state through the crossing of a tipping point. Many social tipping points are caused by a combination of both types of tipping events. The blue circle represents the current state of the system.**

91

92 Bifurcation theory has been applied to study tipping points in a vast number of environmental models (May and  
 93 Oster, 1976; Brovkin et al., 1998; Ghil and Tavantzis, 1983; Wollkind et al., 1988); however, more recently,  
 94 researchers have identified abrupt shifts in environmental systems for which bifurcation theory has yet to be  
 95 explicitly applied (Dakos et al., 2019; Lenton, 2020, 2013). For example, during the mid-Holocene, the Sahara was  
 96 much more humid than at present, showing evidence of shrub and savannah biomes as well as the expansion of  
 97 lakes, an alternative stable state to what we know as its current desert state. It is hypothesized that around 5,000  
 98 years ago, the gradual weakening of the North African Monsoon led to an abrupt decrease in vegetative cover, due to  
 99 positive feedback between reduced surface albedo and precipitation, bringing the Sahara into a stable desert state  
 100 (Hopcroft and Valdes, 2021; Pausata et al., 2020). In more dominantly human systems, many pivotal revolutions can  
 101 also be framed as tipping events where gradual changes are reinforced by positive feedback loops, leading to a new  
 102 political or technological stable state (Lenton et al., 2022). Social tipping also occurs in financial systems such as in  
 103 the 2008 financial crisis. Here, the bankruptcy of Lehman Brothers led to a rise in public panic around the stability  
 104 of markets, causing banks to increase their liquidity, amplifying the crisis in other economic sectors and leading to a

126 global recession (Van Nes et al., 2016). These are just two of many examples illustrating how important tipping  
127 points are as a phenomenon, in both human and environmental systems, and coupling these systems using  
128 mathematical models could lead to further insights.

129

130 Since the beginning of the Anthropocene and with our growing awareness of human impacts on the environment,  
131 tipping points are increasingly being conceptualized within the context of coupled human-environment systems  
132 (Bauch et al., 2016; Henderson et al., 2016; Lenton et al., 2022; Milkoreit et al., 2018). Tipping events can lead to  
133 highly beneficial or catastrophic outcomes for humans, especially when an environmental change occurs in the  
134 presence of social hysteresis. An example of detrimental tipping is in the forests of Kumaun and Garhwal in  
135 Northern India, where, prior to British colonization, wood harvest was sustainably regulated through social norms  
136 and strict rules enforced by local village councils. When the British colonial government imposed its own rules on  
137 the use of forests, these social norms broke down. Eventually, protests led to British lumber restrictions being  
138 removed, but the system subsequently experienced rapid deforestation rather than a return to its previous levels  
139 under local management. Here, the social system crossed a tipping point between a self-organized common property  
140 regime to one of open access devoid of self-regulating sanctions (Somanathan, 1991). This system has been modeled  
141 using a dynamical systems approach that allows for a quantitative understanding of the human drivers leading to  
142 these tipping events (Sethi and Somanathan, 1996). Contrasting this example, tipping events can also result in  
143 environmental change that is beneficial to humans and the environment. The rapid response of the international  
144 community to the hole in the ozone layer has been interpreted by some as an example of a CHES undergoing tipping  
145 events caused by self-perpetuating change through political, technological, and behavioral forces  
146 (Stadelmann-Steffen et al., 2021). In the 1970s, scientists demonstrated the detrimental effects of CFCs on the ozone  
147 layer, which could be viewed as the initial driver of the following socio-climate tipping events. This led to public  
148 concern, prompting several countries to ban the use of CFCs in aerosols. Through the enactment of national policies,  
149 public awareness increased, leading to more public pressure for national and international policy change, an example  
150 of a positive feedback loop. In parallel, these national bans of CFCs, especially in the US, led to the development of  
151 CFC alternatives, which prompted industries that could develop them to lobby for international policy. Increased  
152 public awareness also led to widespread shifts in social norms stigmatizing and boycotting the consumption of  
153 CFCs, which further pressured industry to offer alternatives, another positive feedback loop. The interaction of  
154 multiple tipping events at different scales led to the crossing of a global tipping point through the international  
155 banning of CFCs, bringing an alternative stable state of very low CFC emissions globally. (Andersen et al., 2013;  
156 Cook, 1990; Epstein et al., 2014; Haas, 1992; Stadelmann-Steffen et al., 2021).

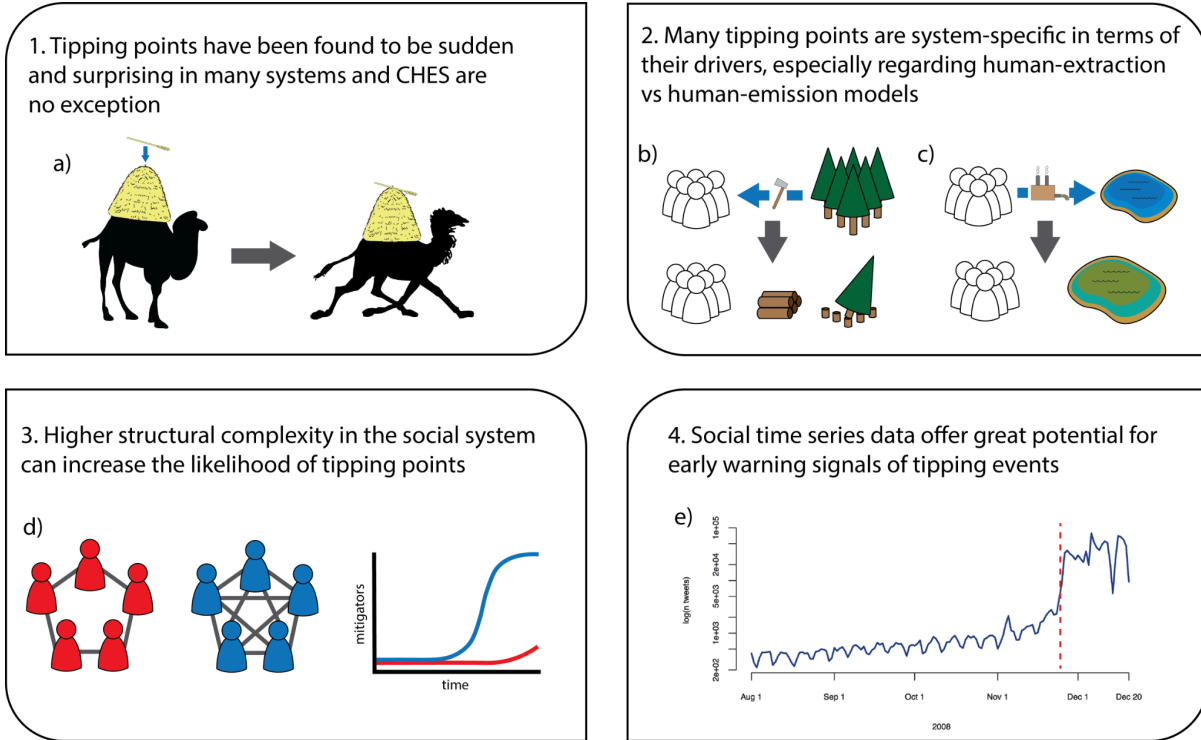
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158 Tipping events associated with social processes as described in the preceding paragraph can be conceptualized  
159 through positive feedback loops that capture a self-reinforcing process. In the case of social norms, this  
160 self-reinforcing process may correspond to peer pressure or conformism that reinforces the dominant opinion or

162 belief. Depending on whether pro- or anti-mitigation opinions are currently dominant, this could lead to hysteresis  
163 (Figure 2b). The negative feedback loop that might normally regulate the CHES to exist in a state of intermediate  
164 environmental health and public support for sustainability (Figure 2a) could be overpowered by the positive  
165 feedback of social norms, leading the population to a state where either sustainability (or anti-sustainability) is  
166 strongly entrenched. If the conditions governing social learning or social norms move beyond a tipping point, the  
167 population may flip between these two norms, or alternatively it may move into a regime where social norms are  
168 instead dominated by the negative feedback loop, causing the population to exist in an interior state of partial  
169 sustainability. As such, negative feedback and positive feedback may be characteristic of any CHES and should be  
170 systematically studied.

171

172 This review aims to deepen our understanding of human drivers of tipping events in CHES models by exploring  
173 three crucial topics: the feedback loops and interactions between the human and environmental systems, the  
174 structural characteristics of the human system that influence tipping points, and the identification of early warning  
175 signals within human systems. By “human drivers”, we refer to the changes in social parameters that elicit these  
176 non-linear tipping responses in either the environment, human system, or both. However, we also discuss aspects of  
177 social structure that may be conducive to tipping points. As most of the models reviewed are informed by dynamical  
178 systems and bifurcation theory, we primarily focus on systems that exist near tipping points and cross them through  
179 gradual changes in these drivers. In the following sections we review CHES model literature found using Google  
180 Scholar with the keywords: ‘human environment system’ OR ‘socio-ecological system’ OR ‘social ecological  
181 system’ OR ‘human ecological system’ OR ‘human natural system’ combined with ‘tipping’ OR ‘regime shift’ OR  
182 ‘bifurcation’. These results were filtered manually to include only dynamical models that showed clear tipping  
183 behavior. Additional literature was found through a snowball approach using references from the sources found in  
184 this search as well as papers referencing these sources (Wohlin, 2014). The findings in this review highlight  
185 commonalities between the CHES models surveyed; however, some trends may be a result of both the dynamical  
186 models chosen and the relatively low diversity and volume of these models. The body of this review is split into two  
187 parts; the first part synthesizes results from CHES models, organized into processes and structures that drive tipping  
188 behavior, and the second part introduces early warning signals describing how they can be used to predict tipping  
189 events.



**Box 1: Highlights of key findings from the synthesis of CHES models in this review. “The straw that broke the camel's back” illustrating bifurcation-induced tipping points (1a), in human-extraction systems (2b), increasing the speed of social change or the coupling strength leads to negative tipping points (i.e., ecological collapse), whereas in human-emission systems (2c), the effects of increasing the speed of social change or the coupling strength are model specific, higher connections in a social network leading to a positive tipping event, where the graph represents the proportion of mitigators in time (3d), time series data from Twitter showing an abrupt transition characteristic of a tipping event at the red dotted line (4e) from (Bollen et al., 2021).**

**198 2 Processes and structures in human systems that cause tipping events in CHES models**

199 In this section, we look at how social processes and structures cause tipping events. In order to have a better  
 200 understanding of how these human drivers affect tipping, it is important to understand the basics of modeling human  
 201 systems. Within CHES models, various factors, such as economic incentives, environmental considerations, and  
 202 social pressures determine how individuals make decisions and interact with the environment. In most of the current  
 203 modeling literature, individuals can choose between two behaviors (also referred to as opinions or strategies), one  
 204 that is environmentally sustainable (also referred to as mitigation or cooperation) and another that is detrimental to  
 205 the environment (also referred to as non-mitigation or defection). The perceived advantage of mitigation or  
 206 non-mitigation relative to the current state of the human and environmental system can be quantified through a

204 “utility function”. Common factors in the utility function are the rate of social learning, which determines the speed  
205 of human behavior change relative to environmental processes, social norms, which encourage the status quo or  
206 mitigation proportional to its frequency, cost of mitigation, which measures the economic cost of being a mitigator  
207 relative to a non-mitigator, and rarity-motivated valuation, which incentivizes mitigation as the environment  
208 approaches collapse (Bauch et al., 2016; Farahbakhsh et al., 2022; Tavoni et al., 2012). In most models that use  
209 social learning, individuals sample others in the population at a fixed rate and adopt a different behavior if the other  
210 behavior has a higher utility, with probability proportional to the difference in utility (Hofbauer and Sigmund, 1998;  
211 Schuster and Sigmund, 1983). This can also be formulated in a stochastic setting, where the probability of adopting  
212 a neighbor's behavior is a function of the difference in utility between behaviors (Schlag, 1998). Most of the models  
213 reviewed in this paper use social learning to represent human behavioral dynamics. There are also CHES models  
214 that do not include social learning such as Motesharrei (2014) and Dockstader (2019) where the human population is  
215 influenced by its current size and the state of the environment; however, these are outside the scope of this paper.

216

217 Many human behaviors, such as resource extraction and pollution, have direct detrimental impacts on the  
218 environment; however, the severity of these impacts is often hard to predict. In many CHES models, small changes  
219 in parameters governing human behavior and social processes can lead to the abrupt collapse of sustainable states  
220 through tipping events that can cascade between the human and environmental systems (Bauch et al., 2016; Lade et  
221 al., 2013; Richter and Dakos, 2015; Weitz et al., 2016). Additionally, structural elements of the human system (i.e.  
222 an individual's degree of choice, population diversity), as well as how the social system is organized (i.e. through  
223 social networks), can affect tipping. These heterogeneous model elements are often only accessible in agent-based  
224 models, where humans are represented as individual agents that follow a set of rules. CHES models do not always  
225 exhibit tipping points under realistic settings for the human system (Bury et al., 2019; Menard et al., 2021); however,  
226 in this review, we focus on models with tipping points.

## 227 **2.1 Coupling strength**

228 Coupling strength (how strongly the subsystems are coupled) can have a significant effect on the occurrence of  
229 tipping points in both systems, and the nature of these transitions often depends on whether systems are  
230 ‘human-extraction’ or ‘human-emission’ (Box 1.2). In human-extraction systems (Box 1.2b), humans extract from  
231 an environmental resource such as in forest and fishery models. Stronger coupling in human-extraction models often  
232 leads to negative environmental outcomes. A common social parameter representing the coupling strength in these  
233 systems is the extraction effort of humans, which when increased past a critical threshold, leads to abrupt  
234 environmental collapse (Farahbakhsh et al., 2021; Richter and Dakos, 2015; Richter et al., 2013; Schlüter et al.,  
235 2016). For human-emission systems (Box 1.2c), where human activity increases levels of harmful outputs, such as  
236 pollution and climate models, coupling strength is instead represented by pollution rates. The influence of this  
237 coupling is less intuitive in human-emission systems, for example, in lake eutrophication models as the pollution of



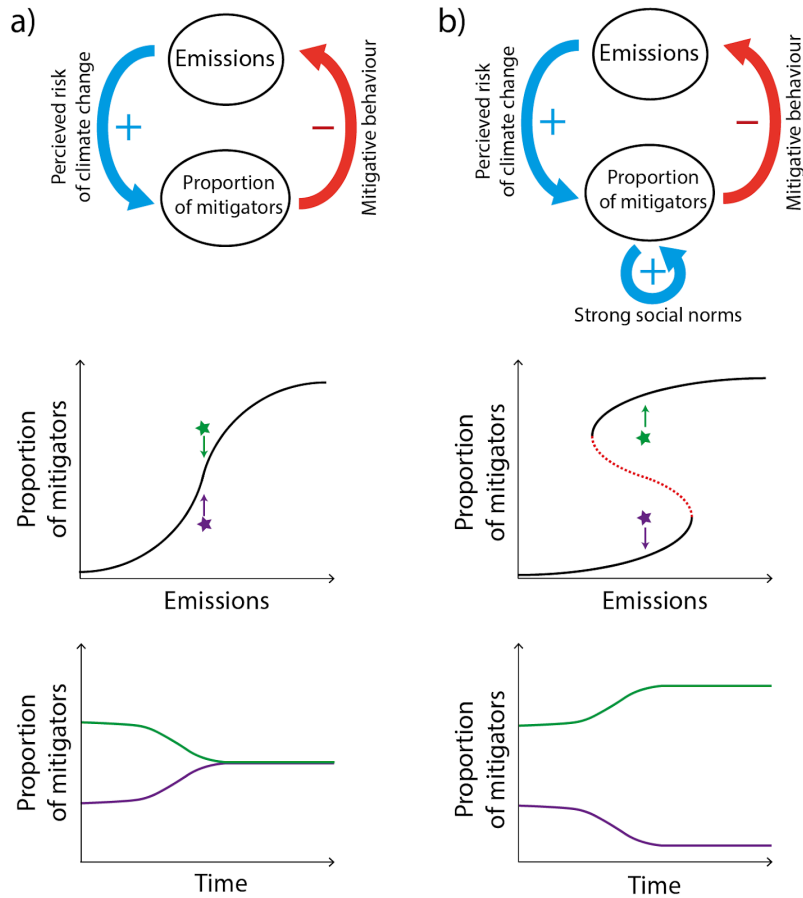
240 mitigators is decreased, pollution levels also decrease until a threshold is reached, heralding a detrimental tipping  
241 point where mitigation collapses and pollution then reaches a high level (Iwasa et al., 2010, 2007). This occurs  
242 because when the lake water is not very polluted, there is less incentive to be a mitigator and high-polluting behavior  
243 becomes a new norm. It is important to note that these models do not account for individuals valuing the  
244 environment in a healthy state, for example through the centering of ecosystem services, and the above example may  
245 be an artifact of this assumption. There is a need to shift both our relationship to the environment as well as the  
246 assumptions in our models so that inherent value in environmental systems is central in any decision-making, even  
247 when the environment is far from collapse. This fundamental valuing of the environment is present in many  
248 traditional indigenous belief systems, where relationships to the local natural environment are incorporated and  
249 prioritized in all aspects of life (Appiah-Opoku, 2007; Bavikatte and Bennett, 2015; Beckford et al., 2010; McMillan  
250 and Prosper, 2016).

## 251 **2.2 Rarity-motivated valuation**

252 Rarity-motivated valuation represents the extent to which humans increase their mitigative behavior in response to  
253 the environmental variable (e.g., forest cover, endangered species population size) nearing a depleted state. In CHES  
254 models, this sensitivity of human response to the abundance of the natural resource/population is represented by a  
255 ‘sensitivity’ parameter and there are often two critical thresholds in the sensitivity parameter that lead to tipping.  
256 Increasing the sensitivity parameter beyond the lower threshold induces a tipping point from a depleted to  
257 sustainable environmental equilibrium (Ali et al., 2015; Barlow et al., 2014; Bauch et al., 2016; Drechsler and  
258 Surun, 2018; Henderson et al., 2016; Lin and Weitz, 2019; Sun and Hilker, 2020; Thampi et al., 2018; Weitz et al.,  
259 2016). The second threshold exists at high values of the sensitivity parameter, which may be counterintuitive, as one  
260 might expect high sensitivity to resource depletion to lead to more sustainable outcomes. In this case, the sustainable  
261 equilibrium is destabilized by overshoot dynamics or a state of chaos in both the human and environmental systems.  
262 These dynamics are caused by the human system being too sensitive to changes in the environment, leading to  
263 extreme oscillations in both human behavior and the environment, which increases the likelihood of collapse in  
264 mitigation and the state of the environment (Bauch et al., 2016; Henderson et al., 2016).

265

266 Rarity-motivated valuation can also be represented by a threshold in the state of the environment, below which  
267 humans shift towards sustainable behavior. In a common-pool resource model, lowering this threshold led to a series  
268 of tipping points that surprisingly resulted in a higher biomass equilibrium, although the trajectory to this state  
269 comes close to environmental collapse. This is in contrast to a high threshold, which leads to lower final biomass;  
270 however, the trajectory remains much farther from a depleted environmental state (Mathias et al., 2020). Similarly to  
271 high coupling in pollution models, one should be very careful to not interpret these results as stating “too much  
272 conservation is detrimental to the environment”. They rest on model assumptions of a reactionary conservation  
273 paradigm, where there is less value in conserving when the environment is in a healthy state.



**Figure 2: Negative feedback between the human and environmental subsystems, supports convergence to the same equilibrium regardless of initial conditions (a). With strong majority-enforcing social norms, encouraging either mitigative or harmful behavior adds a positive feedback loop which makes the coupled system highly dependent on initial conditions (b). The top row shows the negative feedback loop between emissions and the proportion of mitigators, where (b) also includes the positive feedback of majority-enforcing social norms. In the middle row, equilibrium curves are plotted as a function of the maximum emissions of non-mitigators. Black solid lines represent stable equilibria and the red dotted line represents unstable equilibria. The green and purple curves in the bottom row are the trajectories for initial mitigation support and emission value given by the stars of the corresponding color in the upper row.**

### 277 2.3 Social norms

278 Introducing social norms can lead to alternative stable states and thus tipping points (Figure 2b), although the system  
 279 dynamics are highly dependent on both the type of social norms and initial conditions. Social norms are informal

312 rules emerging through social interaction that promote and discourage certain behaviors, especially around how  
313 humans relate to one another and the environment (Chung and Rimal, 2016). In models of small groups such as a  
314 community of fishers, they are often (rightly) assumed to support mitigative behavior by punishing those who  
315 violate norms by over-harvesting (Ostrom, 2000). However, at larger population scales, social norms can support  
316 either pro- or anti-mitigation behavior, on account of factors such as the politicization of actions relating to  
317 environmental, climate, and public health crises (Stoll-Kleemann et al., 2001; Van Boven et al., 2018; Latkin et al.,  
318 2022). Unlike a fisher in a small community, for instance, a climate denier may not acknowledge themselves as a  
319 ‘defector’ who is harming a public good, but rather view the climate activist as ‘defecting’ against a free society.  
320 Thereby, social norms have the ability to encourage behavior that is harmful to both human and environmental  
321 well-being, over larger spatial and temporal scales (Bury et al., 2019; Latkin et al., 2022; Menard et al., 2021;  
322 Stoll-Kleemann et al., 2001; Van Boven et al., 2018).

323

324 Social norms can be represented as majority-enforcing, incentivizing the behavior of the majority, or  
325 mitigation-enforcing, such as sanctions, which only incentivize mitigation, relative to the proportion of mitigators in  
326 the current state of the system. In CHES models, increasing the strength of majority-enforcing norms leads to an  
327 increased number of regimes as well as bistable (two stable states) regimes (Figure 2b), made up of a single  
328 dominant behavior, which is highly dependent on the initial proportion of behaviors in a population (Ali et al., 2015;  
329 Barlow et al., 2014; Bauch et al., 2016; Bury et al., 2019; Phillips et al., 2020; Sigdel et al., 2017; Thampi et al.,  
330 2018). This occurs because these norms are indifferent to the type of behavior they enforce (i.e. sustainable vs  
331 harmful actions), and they act as a double-edged sword that reinforces the status quo through a positive feedback  
332 loop, where the dominant behavior becomes more prevalent (Figure 2b). On the other hand, increasing  
333 mitigation-enforcing social norms lead to a transition of the environmental system into a sustainable equilibrium  
334 (Chen and Szolnoki, 2018; Iwasa et al., 2010; Lafuite et al., 2017; Moore et al., 2022; Schlüter et al., 2016; Tavoni et  
335 al., 2012), sometimes through an intermediate regime of oscillatory dynamics (Iwasa et al., 2007). In a lake pollution  
336 model, along with decreasing the likelihood of environmental collapse, this increase in mitigation-enforcing social  
337 norms also led to the appearance of alternate stable states (Sun and Hilker, 2020). These findings show that stronger  
338 social norms lead to a greater number of tipping points; however, the trajectories brought about by these tipping  
339 points are highly dependent on the type of social norms (mitigation- or majority-enforcing) as well as the current  
340 dominant social behavior.

#### 341 **2.4 Cost of mitigation**

342 Reducing the cost of mitigation often leads to beneficial tipping points; however, these tipping points can depend on  
343 the rate of social change as well as social norms. Although it is intuitive that reducing costs or increasing economic  
344 incentives associated with mitigative action will have beneficial impacts on the environment, CHES models also  
345 show that this beneficial change can occur through tipping points (Bauch et al., 2016; Drechsler and Surun, 2018;

348 Milne et al., 2021; Moore et al., 2022; Sigdel et al., 2017; Thampi et al., 2018). In coupled social-epidemiological  
349 models, where the environmental state is the proportion of infected individuals, mitigation cost is represented  
350 through the economic cost or perceived risk of vaccination. Decreasing this cost leads to beneficial tipping points  
351 from a state with low pro-vaccine opinion and vaccine coverage to high pro-vaccine opinion and vaccine coverage  
352 (Phillips et al., 2020). Conversely, increasing this cost leads to a state of high infection and low vaccination. This  
353 detrimental tipping point occurs in the human system at lower levels of vaccination cost when majority-enforcing  
354 social norms are low, leading to widespread anti-vaccine opinion before the infection becomes endemic again  
355 (Phillips and Bauch, 2021). Decreasing profits of individuals engaging in non-mitigative behavior can also lead to  
356 an abrupt shift to a state of pure mitigators (Shao et al., 2019; Wiedermann et al., 2015); however, this transition can  
357 be dependent on a low rate of social change (Wiedermann et al., 2015). Other models demonstrate tipping in the  
358 other direction where increasing non-mitigators' payoff brings about a regime shift to pure non-mitigation and  
359 environmental collapse (Richter et al., 2013; Tavoni et al., 2012). Similarly, a common-pool resource model that  
360 uses machine learning in a continuous strategy space shows tipping to a depleted resource regime when the costs  
361 associated with harvesting are too low (Osten et al., 2017). An analog to mitigation cost is taxation rates, which  
362 resource users pay towards public infrastructure mediating resource extraction. In a model where individuals can  
363 choose to work outside of the system, pushing taxation rates to high or low levels tips a sustainable regime where  
364 institutions are at full or partial capacity to a collapse of institutions (Muneepeerakul and Anderies, 2020). In another  
365 model, only individuals with high extractive effort are subject to taxation, and increasing this taxation rate brings  
366 about a beneficial tipping point to a sustainable regime. However, the size of this sustainable region in the parameter  
367 space is smaller with multiple governance nodes evolving through social learning compared to a single taxing entity  
368 (Geier et al., 2019). However the cost of mitigation is represented, increasing the relative economic incentive of  
369 mitigation behavior has the potential to bring about beneficial tipping to a sustainable regime.

## 370 2.5 Rates of social change and time horizons

371 Human and environmental change often occur on different timescales and their relative rates of change play a major  
372 role in the long-term dynamics of the coupled system and whether or not tipping points will occur. Increasing the  
373 rate of social change (in most cases, social learning) leads to collapse in human-extraction models due to overshoot  
374 dynamics, whereas, in human-emission models, the impacts of the rate of social change are more model-specific. In  
375 both types of models, increasing the time horizon in decision-making is beneficial. In CHES models, these rates of  
376 change can be controlled by the rate of social learning which determines how frequently individuals interact and  
377 consequently, the pace of behavioral change within a population. Changes in the speed of the human system can  
378 have very different outcomes depending on the nature of human-environment coupling (Box 1.2). In  
379 human-extraction models, increasing the speed of the human system relative to the environment often destabilizes  
380 sustainable equilibria, leading to oscillations in both systems and, in many cases, the abrupt collapse of the  
381 environmental system. These overshoot dynamics occur as humans change their behavior too quickly to allow for

375 the environment to stabilize. On the other hand, decreasing the relative speed of human dynamics usually brings  
376 about beneficial tipping events leading to a state of high forest cover (Figueiredo and Pereira, 2011), and supporting  
377 mitigators for a generalized resource (Hauert et al., 2019; Shao et al., 2019). These beneficial effects have also been  
378 observed in adaptive network models where individuals imitate their neighbors depending on the profitability of  
379 their strategies. In these models, the reduced speed of social change leads to beneficial outcomes as the resource is  
380 allowed more time to stabilize as decisions regarding extractive levels occur (Barfuss et al., 2017; Geier et al., 2019;  
381 Wiedermann et al., 2015). Other relative rates of change can also significantly influence the existence of a  
382 sustainable regime. For example, in an agricultural land use model, increasing the speed of agricultural expansion  
383 and intensification relative to human population growth leads to the collapse of both the natural land cover and  
384 human population (Bengochea Paz et al., 2022).

385

386 In human-emission models, increasing the speed of social interaction is more model-specific. In some cases, such as  
387 forest-pest and climate systems, increasing the speed of the human system leads to better mitigation of  
388 environmental harms in the short term. However, long-term sustainability often requires additional social  
389 interventions such as reducing mitigation costs and increasing levels of environmental concern (Ali et al., 2015;  
390 Barlow et al., 2014; Bury et al., 2019). In lake pollution models, higher relative speeds of social dynamics can  
391 destabilize low-pollution equilibria, leading to oscillations and eventually a polluted state with no mitigation (Iwasa  
392 et al., 2010, 2007; Sun and Hilker, 2020). This is a similar phenomenon to the overshoot dynamics that occur when  
393 the human system is extremely reactive to the environment discussed in the case of rarity-motivated valuation;  
394 however, these outcomes are highly dependent on other social parameters. In a related model, with no social  
395 hysteresis, represented by mitigation-enforcing social norms, and strong environmental hysteresis, represented by a  
396 high phosphorus turnover rate, fast social dynamics could stabilize oscillations, leading to a low-pollution  
397 equilibrium (Suzuki and Iwasa, 2009). The emergence of oscillations under low rates of social learning, which was  
398 not observed in similar models is likely due to the environmental system being in a bistable state under strong  
399 hysteresis, such that even slow changes in the human system could tip the lake system to an alternative stable state.

400

401 When looking at relative rates of change in human and environmental systems, it is clear that the pace of the human  
402 system can be more readily influenced by interventions. This suggests an urgent need to further study the  
403 relationship between social and ecological timescales across a wide range of coupled systems to aid in sustainable  
404 policy-making decisions (Barfuss et al., 2017). Additionally in many models, the length of time horizons that  
405 humans take into account when deciding how they interact with the environment has a significant beneficial effect  
406 on conserving natural states and mitigating harmful action (Barfuss et al., 2020; Bury et al., 2019; Henderson et al.,  
407 2016; Lindkvist et al., 2017; Müller et al., 2021; Satake et al., 2007). A high degree of foresight in decision-making  
408 is a fundamental basis for many indigenous belief systems across the world. One manner in which this shows up is

393 in land stewardship where care for the environment is prioritized as a means to ensure the health of many  
394 generations in the future (Appiah-Opoku, 2007; Beckford et al., 2010; Ratima et al., 2019).

## 395 **2.6 Social traits**

396 The inclusion and distribution of traits within agents can play a large role in determining the occurrence and types of  
397 tipping points within the coupled system, where increasing the modeled heterogeneity in social traits can lead to  
398 more tipping and also promote sustainable outcomes (Box 1.3). The majority of models discussed in the previous  
399 section only allow humans to choose between two strategies; mitigation and non-mitigation. The inclusion of  
400 additional strategies, determining how individuals interact with the environment and each other, can alter the  
401 potential for tipping points. For example, a common-pool resource model included a third strategy of conditional  
402 mitigation (Richter and Grasman, 2013). Under this additional strategy, agents act as mitigators until the number of  
403 non-mitigators reaches a certain threshold, where they then shift their behavior to non-mitigation. The addition of  
404 this third strategy alters tipping dynamics in opposite ways, depending on the value of maximum harvesting efforts.  
405 When efforts are high, the system is less prone to tipping; however, when they are low, tipping points are more  
406 likely to occur. This third strategy also affects tipping points by masking internal social dynamics, leading to more  
407 abrupt transitions, even when the system appears to be stable. This occurs when mitigators gradually change their  
408 strategy to conditional mitigators which can go unnoticed as their interaction with the environmental system does  
409 not change. However, when non-mitigation reaches high enough levels, there is a cascade of conditional mitigators  
410 choosing non-mitigation, in an example of herd behavior, which puts abrupt harvesting pressure on the resource.  
411 Another three-strategy model, where agents are partitioned by resource extraction rates, contrasts dynamics with and  
412 without the trait of environmental concern (Mathias et al., 2020). In the absence of this trait, the human system  
413 either tips to a state of high-extraction or low-extraction behavior, triggering either a detrimental or beneficial  
414 environmental tipping point, respectively. Including environmental concern leads to an increased number of  
415 cascading tipping points between both human and environmental systems. In a coupled agricultural model, where  
416 human traits include management strategies that respond to socio-economic and climate conditions, decreasing the  
417 diversity of these traits among agents in the system transitions the system from a sustainable state with high food  
418 production, landscape aesthetics, and habitat protection to a state with low habitat protection (Grêt-Regamey et al.,  
419 2019). As there are relatively few models that explicitly compare the complexity of social traits and their effect on  
420 tipping points, it is difficult to say with certainty whether higher complexity will increase the likelihood of tipping  
421 points in all CHES and whether this is due to a higher dimensionality of the system. However, the commonalities  
422 between models showing the effects of social trait complexity are worth highlighting and will be put to the test with  
423 future CHES models and empirical work.

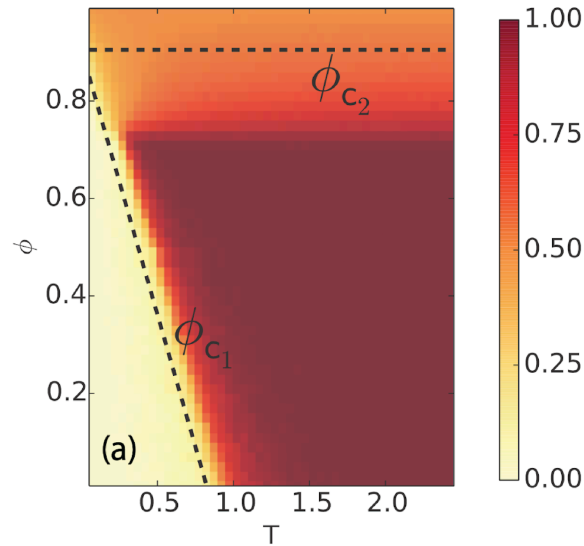
## 424 **2.7 Social networks**

420 In many agent-based CHES models, individuals are structured on a social network, where they interact with others  
421 whom they share a link with. These models demonstrate how a higher number of connections in social networks  
422 increases the potential for tipping points, often through the emergence and growth of bistable regimes (Holstein et  
423 al., 2021; Sugiarto et al., 2015, 2017a) (Box 1.3). Additionally, the distributions of these connections play an  
424 important role. For example, in networks with the same average number of connections, higher heterogeneity of  
425 connections among nodes leads to tipping points occurring earlier under certain social (Ising model) dynamics  
426 (Reisinger et al., 2022). The distribution of resources in human-environment networks also affects the potential for  
427 abrupt environmental collapse. This often occurs in CHES network models where both human and environmental  
428 dynamics occur on a multi-layer network, representing partitioned or private resources. Resource heterogeneity can  
429 be controlled through the distribution of carrying capacities or the amount of resource flow between nodes in the  
430 network, where higher flows lead to homogeneous resource distributions. In both cases, increasing this  
431 heterogeneity can tip the system to a state of low extraction and high sustainability. In one model, heterogeneity in  
432 carrying capacities increases the likelihood of sustainable harvesters extracting from a resource with a large capacity,  
433 which they can maintain at high levels (in contrast to non-sustainable harvesters who extract at a higher rate),  
434 eventually convincing neighboring nodes to imitate their strategy (Barfuss et al., 2017). In another model,  
435 heterogeneity through lower resource flows also leads to high-extraction nodes over-exploiting their resource and  
436 losing profits in the long run, de-incentivizing neighbors to imitate their behavior. Interestingly, optimal resource  
437 flow, which minimizes the likelihood of resource collapse is found to be close to the critical threshold of resource  
438 flow, above which the coupled system collapses. As optimal resource flow decreases the likelihood of collapse by  
439 supplementing resources harvested at high levels, this confers an advantage to high resource extraction. Increasing  
440 past optimal levels leads to similar resource levels among high and low-extraction nodes, resulting in higher profits  
441 from high-extraction nodes, incentivizing the entire human system to eventually choose the high-extraction strategy  
442 (Holstein et al., 2021).

443

444 Heterogeneity of human interaction can be quantified through homophily, the extent to which alike individuals  
445 interact. Homophily can play a large role in the occurrence and behavior of tipping points in CHES models  
446 occurring on social networks, often having a detrimental effect on the environmental system. In a common-pool  
447 resource model with two distinct communities, increasing segregation by lowering the probability that agents in  
448 separate communities will have a link, softens the abruptness of a single detrimental tipping point compared to when  
449 the communities are well-mixed. This is due to the occurrence of multiple intermediate tipping points within each  
450 segregated community; however, higher segregation adds more hysteresis to the system increasing the difficulty of  
451 reversing this transition and returning to a sustainable state (Sugiarto et al., 2017b). In a public goods game  
452 modeling climate change mitigation, where humans are partitioned into rich and poor agents, a transition to group  
453 achievement of mitigation goals occurs at a lower perceived risk when there is no homophily and agents are  
454 influenced by others from both economic classes equally (Vasconcelos et al., 2014). Another human-climate model

428 that included wealth inequality displayed an abrupt transition to lower peak temperature anomalies when homophily  
429 between economic classes approached zero (Menard et al., 2021).



**Figure 3: Mean proportion of nodes that are mitigators for network model (a) and ODE model (b).  $\phi$  is the rewiring probability and  $T$  is the time between social interactions.  $\phi_{c1}$  is the lower threshold and  $\phi_{c2}$  is the upper threshold, above which a fragmentation regime occurs. From (Wiedermann et al., 2015)**

430

431 Social networks are rarely static and their ability to evolve over time is represented in adaptive network models  
432 where agents can break existing social links and create new ones, a process called “rewiring”. Often this rewiring is  
433 homophilic, meaning that agents are more likely to create a new social connection with others who share a similar  
434 behavior. Common adaptive network CHES models have nodes representing renewable resource stocks with an  
435 associated extraction level which can adopt a high extraction or low extraction level through imitating neighbors.  
436 These models show that the level of homophilic rewiring can trigger regime shifts at both low and high levels,  
437 where intermediate ranges correspond to a sustainable equilibrium. As agents can either choose to rewire or imitate  
438 their neighbor, a low level of rewiring corresponds to a high speed of social interaction, which as discussed in  
439 Section 2.5 can lead to detrimental tipping points. On the other hand, although high-rewiring leads to slower social  
440 learning, it also brings about a fragmentation regime where social dynamics are dominated by homophily and the  
441 network fragments into components based on strategy type, which makes widespread mitigation infeasible (Barfuss  
442 et al., 2017; Geier et al., 2019; Wiedermann et al., 2015) (Figure 3). CHES models with social networks are still  
443 relatively new and lack diversity in how they are formulated. For example, regarding the tipping points related to  
444 rewiring social links, the lower threshold may be caused by increased social learning since in all models agents can



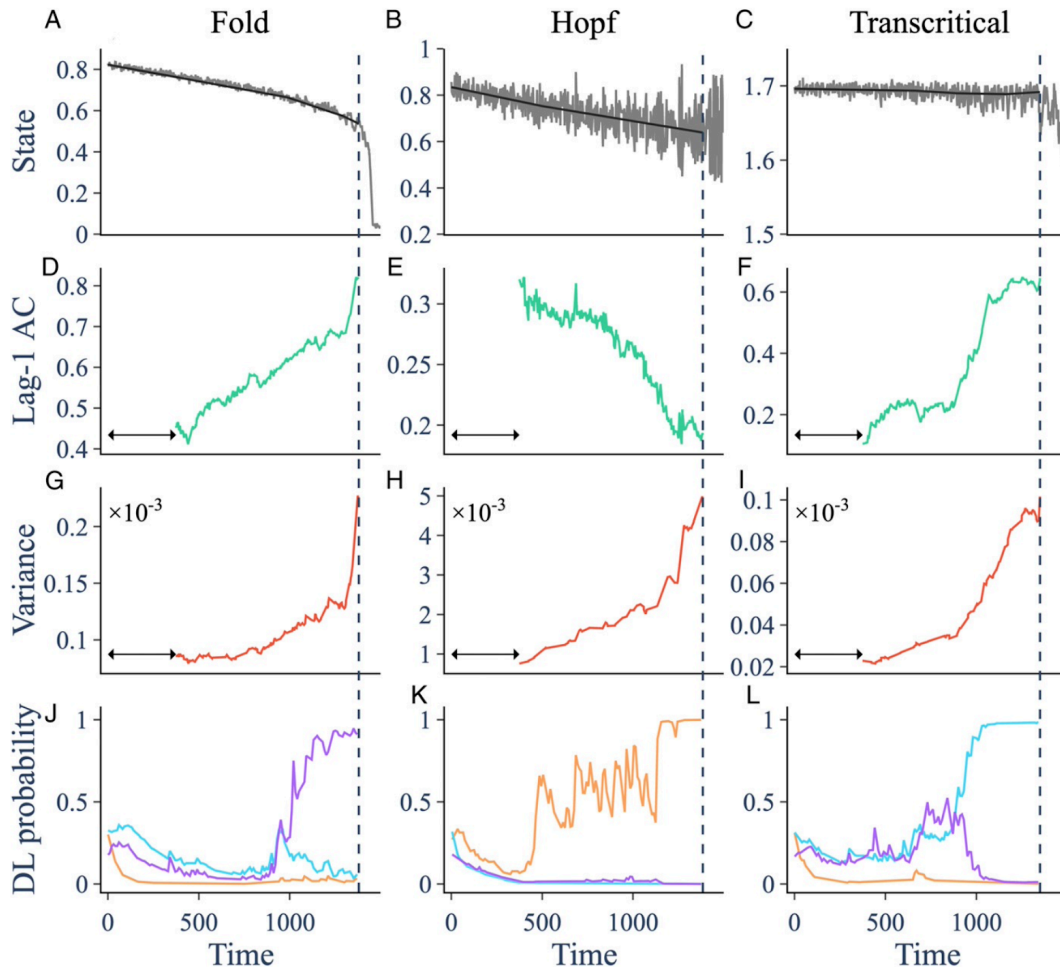
464 either rewire or imitate, but not both. There is still much to learn through isolating the effect of rewiring as well as  
465 exploring a wide array of different model formulations of CHES on social networks.

### 466 **3 Identifying early warning signals of tipping events in CHES**

467 Although dynamical models can offer qualitative insight into potential trajectories of CHES resulting from specific  
468 interventions, it is more difficult to use them to generate precise and reliable predictions. Given the potential for  
469 severe environmental tipping points in the coming decades, it is extremely useful to be able to predict these abrupt  
470 shifts without complete mechanistic knowledge of the system. The ability to predict tipping events with limited data  
471 can allow policymakers to have more time preparing for future disasters, and given enough warning and political  
472 will, an opportunity to avoid them or mitigate their severity. Rapidly growing research in early warning signals  
473 (EWS) offers tools to monitor empirical time series data and warn of future tipping events that are likely to occur  
474 (Bury et al., 2021; Dakos et al., 2012, 2015, 2008; Kéfi et al., 2014; Lapeyrolerie and Boettiger, 2021). Although  
475 much of the work has been conducted on synthetic data, there are many studies that successfully predict historical  
476 tipping events in both empirical human and environmental time series data such as the 1987 Black Monday financial  
477 crash (Diks et al., 2019) as well as abrupt temperature shifts from paleoclimate datasets (Dakos et al., 2008).

#### 478 **3.1 Recent advances for detecting early warning signals**

479 Much research has been done in the past few decades to develop tools for EWS using both empirical and synthetic  
480 time series data (Bury et al., 2021; Dakos et al., 2012, 2015, 2008; Kéfi et al., 2014; Lapeyrolerie and Boettiger,  
481 2021). Originally motivated by critical slowing down in bifurcation theory, where systems approaching a tipping  
482 point show a slower recovery to equilibrium under perturbations, generic EWS measure trends in this “slowing  
483 down” (Scheffer et al., 2009). The most commonly used methods compute the lag-1 autocorrelation and variance of  
484 the residuals from detrended time series data. Other widely used methods involve metrics such as skewness,  
485 measuring the asymmetry of fluctuations over time, and kurtosis, representing the likelihood of extreme values in  
486 the time series data. A phenomenon known as flickering occurs when there is sufficient noise to rapidly force the  
487 system between alternate stable states. In these cases, an increase in skewness and kurtosis is observed (Dakos et al.,  
488 2012). As lag-1 autocorrelation does not account for correlation beyond a single time step, power spectrum analysis  
489 has been used to look at changes in complete spectral properties, finding higher variations at low frequencies to  
490 commonly occur before a tipping point (Dakos et al., 2012; Scheffer et al., 2009). In spatial systems, many EWS are  
491 similar to those used in well-mixed systems, while also accounting for spatial variability. For example, Moran’s I is  
492 a spatial analog of lag-1 autocorrelation, which measures the correlation between neighboring nodes in a network  
493 (Kéfi et al., 2014).



**Figure 4: Generic EWS (second and third row) as well as deep learning EWS (bottom row) for time series generated by two ecological models exhibiting different types of bifurcations (top row); fold (left), Hopf (middle), and transcritical (right). As well as being more reliable, deep learning EWS can also distinguish between the type of bifurcation being approached. In the bottom row, the DL algorithm gives probabilities for the occurrence of fold (purple), Hopf (orange), or transcritical (blue) bifurcations. Image taken from (Bury et al., 2021).**

498 Numerous spatial ecological systems exhibit patterns of patchiness preceding a tipping point. For example, in  
 499 drylands, spotted vegetation patterns are hypothesized to be an EWS for the system approaching desertification  
 500 (Kéfi et al., 2014). Coupled human-epidemiological models also show that spatial properties in the distribution of  
 501 opinions on a social network offer potential EWS for the onset of disease outbreaks. Approaching this regime shift,  
 502 the number of anti-vaccine clusters increases, and very close to the transition point, these communities coalesce into  
 503 larger groups (Jentsch et al., 2018; Phillips et al., 2020). These clusters are quantified using a number of metrics,  
 504 such as an increase in modularity as well as the mean number, size, and maximum size of communities and

535 pro-vaccine echo chambers (Phillips and Bauch, 2021). This is also in agreement with previous work done in  
536 percolation theory showing that phase transitions follow a breakup of connected components on the network  
537 (Newman, 2010).

538

539 One downside to the generic metrics discussed above is that they have the potential to fail in the presence of large  
540 amounts of noise where transitions can occur far from their analytically derived tipping point. A technique called  
541 dynamical network markers increases the dimensionality of the time series by transforming it from state variables to  
542 probability distributions of the mean and variance over a given window of time. This reduces the magnitude of noise  
543 in each dimension and in approaching a tipping point, one dominant group of variables will show a drastic increase  
544 in variance and correlation between other variables within that group. At the same time, the correlation between one  
545 variable in this dominant group and others outside the group will decrease. This technique has shown success with  
546 empirical data, such as predicting critical transitions in time series data for a eutrophic lake as well as the bankruptcy  
547 of Lehman Brothers (Liu et al., 2015), and flu outbreaks (Chen et al., 2019). Dynamical network markers have also  
548 been used on spatial systems such as those occurring on social networks through the use of hierarchical network  
549 representations. Here, networks are transformed into binary trees where leaves are the nodes from the original  
550 network and branches group nodes together at multiple resolutions. Through this hierarchical model, dynamical  
551 network markers use these multi-scale communities as the groups of variables that are analyzed (Li et al., 2023).  
552 This spatial technique offers a novel method for predicting tipping events for CHES using human data occurring on  
553 complex social networks.

554

555 A very recent addition to the EWS toolkit uses concepts from statistical physics such as average flux, entropy  
556 production, generalized free energy, and time irreversibility to predict tipping points in a shallow lake model much  
557 earlier than generic methods such as autocorrelation and variance, showing promise for use in real-time monitoring  
558 (Xu et al., 2023). Additionally, the field of machine learning has motivated data-driven approaches to EWS which  
559 do not explicitly make use of any statistical metrics in the time series data. Instead, deep learning algorithms are  
560 trained on large synthetic datasets using models that have and have not approached tipping points. In the majority of  
561 cases, these algorithms have performed significantly better at predicting tipping events than generic EWS indicators  
562 when tested on empirical datasets that exhibit abrupt transitions (Bury et al., 2021; Deb et al., 2022) (Figure 4).  
563 Deep learning algorithms are also able to distinguish between different types of bifurcations as they are being  
564 approached which can offer vital information regarding the potential for catastrophic collapse in CHES.

### 565 **3.2 Social data for early warning signals**

566 In CHES models, the strength of EWS from environmental data has been shown to be muted compared to EWS  
567 from environmental systems not coupled to a human system (Bauch et al., 2016) or the same system with weak  
568 coupling between the human and environmental subsystems (Richter and Dakos, 2015). This is likely due to the

571 effects of human behavior acting to mitigate variability in the environmental system, for example, rarity-motivated  
572 valuation creates a negative feedback loop where incentives to mitigate increase as the environment becomes further  
573 depleted, serving as a mechanism to avoid collapse. The muting of EWS provides a unique challenge for monitoring  
574 tipping events in CHES using environmental data, especially as they occur more frequently in these coupled systems  
575 as discussed in Section 2. There are a small number of studies that have directly compared the strength and efficacy  
576 of EWS between various state or auxiliary variables in CHES models. In these studies, generic EWS from data in  
577 the human system were shown to be the only reliable indicators of the coupled system approaching a tipping point.  
578 Examples of human data used include the fraction of conservationists in a forest cover model (Bauch et al., 2016),  
579 average profits by resource harvesters, and catch per unit effort common-pool resource models (Lade et al., 2013;  
580 Richter and Dakos, 2015). In agreement with generic methods, a state-of-the-art machine learning algorithm for  
581 EWS showed higher success in detecting tipping events generated from a coupled epidemiological model using  
582 pro-vaccine opinion in the human system compared to total infectious in the epidemiological system (Bury et al.,  
583 2021). It is possible that the state variable most sensitive to the forcing parameter may exhibit the strongest EWS, as  
584 seen in experimental work on tipping points in a lake food web. In this system, data from the species that had a  
585 direct trophic linkage to a driver of the tipping event (predators added to the food web) exhibited EWS earlier than  
586 those that were farther removed from the driver (Carpenter et al., 2014). If this is the case, human drivers of tipping  
587 points would most directly affect the human system, and EWS should still be stronger using social data.

588

589 The improved reliability of EWS from social data demonstrated through CHES models shows a significant promise  
590 for monitoring resilience in CHES through the analysis of socio-economic data (Box 1.4). This confers a practical  
591 advantage as socio-economic data availability is growing faster than ecological data (and perhaps even  
592 environmental data despite the growth of publicly available satellite data) on account of the era of digital social data  
593 (Ghermandi and Sinclair, 2019; Hicks et al., 2016; Lopez et al., 2019; Salathé et al., 2012). Some examples of this  
594 are monitoring profits tied to resource extraction as well as using sentiment analysis on social media data, such as  
595 the number of tweets in a given area raising concern over the health of a coupled environmental system.  
596 Furthermore, citizen science not only generates environmental data but also provides social metadata through the  
597 participation of users who monitor specific areas. Leveraging existing platforms like CitSci.org, we can use this data  
598 to estimate trends in conservationist frequency over time (Wang et al., 2015). This approach allows for the  
599 implementation of real-time monitoring of environmental systems using data that is currently being generated,  
600 reducing the need for extensive knowledge or complex mechanistic models of the system. With the potential social  
601 data offers for use with EWS, it is important to note that much of the traditional social data, often conducted through  
602 national or regional surveys, do not provide fine-grained spatial or temporal resolution. On the other hand, novel  
603 methods that use social media data can solve the resolution issue, but may not accurately represent the population it  
604 is being used to model (Hargittai, 2020). These challenges may be addressed through a compound approach that  
605 uses hybrid time series generated from multiple types and sources of social data (Rosales Sánchez et al., 2017).

## 603 4 Conclusion and future directions

### 604 4.1 Summary of main points

605 From a wide range of examined theoretical models, we are able to gain insight into human drivers that lead to  
606 tipping events in CHES systems. Many social interventions, such as reducing mitigation costs and extractive effort,  
607 or increasing the time horizon in decision-making, lead to beneficial tipping events, regardless of the system  
608 modeled. The beneficial effect of these interventions is intuitive; however, non-linear responses manifested as  
609 tipping events may not be as evident. Mitigation costs can be reduced through subsidies for land preservation and  
610 green technology, and extraction effort through limits on land development and the expansion of protected natural  
611 areas (i.e. the Haudenosaunee-led protection of the Haldimand Tract) (Forester, 2021), and by increasing time  
612 horizons through passing long-term legislation that centers the well-being of human and environmental systems such  
613 as the Green New Deal (Galvin and Healy, 2020). These policy interventions become more difficult to implement at  
614 large scales, and models that are tailored to global coordination problems can give us insight into how institutions  
615 can work together to rapidly mitigate looming threats, such as the current climate crises we are facing (Karatayev et  
616 al., 2021).

617

618 Other human behaviors and social processes are much more nuanced and system-specific in how they affect tipping  
619 points. For example, models show that rarity-motivated valuation can act to detrimentally tip the environmental  
620 system into a depleted state when it crosses both a lower and (counterintuitively) an upper threshold value. This was  
621 illustrated most clearly in the example of forest cover in the paper by Bauch et al. (2016). Social norms, especially  
622 when majority-enforcing, increase the likelihood of tipping points through the emergence of bistable regimes that  
623 are made up of both sustainable and unsustainable environmental equilibria. The extent of coupling between the  
624 human and environmental system as well as the speed of social change relative to environmental change can have  
625 different effects depending on whether the model is human-extraction or human-emission. Interventions related to  
626 human valuation and social norms are much more difficult to implement as they require a deeper mechanistic  
627 understanding of how to influence social dynamics and may also have ethical considerations.

628

629 The models we reviewed also show that greater structural complexity via the number and diversity of human traits  
630 as well as the number of social connections can increase the potential for tipping points and mask social dynamics  
631 making these transitions much harder to predict. The modeling literature has only explored a small sliver of the  
632 space of possible choices regarding assumed social structure and the types of environmental models coupled to  
633 them. For example, the vast majority of models only allow for a binary choice in human behavior and adaptive  
634 social networks have only recently been incorporated, with limited mechanisms of re-wiring and types of coupled  
635 environmental systems. Consequently, we still have much to learn on how shifting underlying social structures acts  
636 as a driver of tipping events. This is especially true in human-emission models which are important to improving our

639 understanding of how our social structures affect pressing global issues such as pollution and climate change. Even  
640 if we include more diverse and realistic social structures and processes, CHES are composed of many non-linear  
641 feedbacks and contain high levels of uncertainty, and the reality is that we may not be able to have a complete  
642 mechanistic representation through models. EWS from empirical data show great potential in predicting tipping  
643 events without requiring a full understanding of the system being monitored. There have been many advances in  
644 using state-of-the-art machine learning algorithms to provide accurate EWS from 1-D time series (Bury et al., 2021;  
645 Deb et al., 2022), and very recent work is now developing similar techniques to predict tipping events from spatial  
646 data (Dylewsky et al., 2022). As synthetic data from models have shown the value of EWS from social data, it is  
647 likely that applying these techniques to diverse and hybrid empirical social datasets can vastly improve our ability to  
648 predict tipping events caused by human drivers in the future.

#### 649 **4.2 Future work in CHES modeling**

650 There are many social phenomena that are not commonly included in CHES models, yet may be important in  
651 furthering our understanding of tipping points within these systems. We know that inequality in human systems  
652 plays a large role in individuals' risk perception and ability to engage in pro-environmental behavior (Gibson-Wood  
653 and Wakefield, 2013; Pearson et al., 2017; Quimby and Angelique, 2011; Rajapaksa et al., 2018) and have  
654 mentioned two CHES models that incorporate wealth inequality in a human-climate system (Menard et al., 2021;  
655 Vasconcelos et al., 2014). However, more studies explicitly investigating the role of inequality could offer some  
656 valuable insight into interventions that can be more effective in benefiting both the environment and the most  
657 vulnerable in human systems. This could be complemented by incorporating social biases where perceptions of risk  
658 are linked to an individual's socio-economic status, and detrimental environmental outcomes are experienced  
659 disproportionately by vulnerable communities as is commonly observed globally (Banzhaf et al., 2019; Boyce, 2007).  
660 Future models could allow for alternatives to the common modeling assumption where individuals act in their own  
661 self-interest, for example by incorporating other-regarding preferences into utility functions so that individuals value  
662 their neighbors' well-being along with their own (Dimick et al., 2018). These models could also look at grassroots  
663 redistribution of wealth allowing us to explore the effects of alternative social value systems on the environment  
664 (Tilman et al., 2018).

665

666 Stochasticity (noise), especially regarding drivers of tipping points can significantly affect system dynamics  
667 including when tipping points occur. Although many CHES models are deterministic, recent work has shown that  
668 increasing noise can lead to earlier tipping (Willcock et al., 2023), or in other cases, increase the duration of time the  
669 environmental system can persist before becoming extinct (Jnawali et al., 2022). These contradictory results warrant  
670 further work in understanding how different types of noise and their magnitude within drivers of tipping events  
671 affect the resilience of these systems. With stochasticity comes uncertainty, and in real-world systems, it is  
672 impossible to know with precision the extent of social change required to bring about a beneficial or avoid a

676 detrimental tipping point. This uncertainty around our knowledge of system thresholds adds an additional challenge  
677 in both agreeing upon and following through with policy that promotes sustainable futures while taking into account  
678 potential tipping points. Experimental games have shown that high threshold uncertainty can promote the collapse of  
679 a shared resource, often through an increase in free-riding behavior (Barrett and Dannenberg, 2014, 2012). On the  
680 other hand, field experiments in fishing communities have shown that high uncertainty can promote cooperation and  
681 sustainable resource use (Finkbeiner et al., 2018; Rocha et al., 2020). Theoretical models show that increased  
682 uncertainty can lead to increased mitigative behavior if the shared resource is highly valued; however, for  
683 low-valued resources, increased uncertainty can deter mitigation, putting the persistence of the shared resource at  
684 risk (Jager et al., 2000; McBride, 2006). Uncertainty around thresholds is unavoidable, further motivating the need  
685 to offer additional incentives for mitigative action on institutional scales, rather than solely the threat of  
686 environmental collapse. In systems where uncertainty can promote mitigative action, increased communication and  
687 awareness campaigns around this threshold uncertainty could be useful to incorporate into policy.

688

689 This review has focused primarily on the effects of single drivers; however, research on multiple co-occurring  
690 human drivers of tipping events, while more analytically challenging, could offer a holistic understanding of how  
691 these drivers interact. A recent study has shown that multiple drivers can both reduce the time until tipping or lead to  
692 a tipping point that would not occur with a single driver (Willcock et al., 2023) and there is already a large body of  
693 empirical work exploring the diversity of these drivers which can be used to inform future CHES models  
694 (Jaureguiberry et al., 2022; Maciejewski et al., 2019; Millennium Ecosystem Assessment, 2005). Finally, as the  
695 majority of the studies in modeling tipping points have focused on slow gradual changes in the driver, fast changes  
696 require further research as they can exhibit very different tipping behavior (Ashwin et al., 2012). CHES models  
697 ubiquitously exemplify the phenomenon of tipping points, which often occur through drivers in the human system.  
698 Although these models offer valuable insight in understanding key feedbacks and qualitative behavior, their  
699 predictive power is limited. Additionally, as many model findings can depend on the type of system modeled as well  
700 as assumptions in the model formulation, translating this work into policy remains a significant challenge. However,  
701 further work in both diversifying model systems and assumptions paired with research in universal real-time  
702 indicators of EWS shows considerable promise in both improving our understanding and predicting human drivers  
703 of tipping events in the environment.

704

705 **Author contribution.** I.F.: visualization, writing—original draft, writing—review and editing; C.T.B.: visualization,  
706 writing—original draft, writing—review and editing; M.A.: conceptualization, funding acquisition, supervision,  
707 visualization, writing—original draft, writing—review and editing.

708

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710

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## 750 Appendix

Authors	Year	Title	System of study
Sethi & Somanathan	1996	The evolution of social norms in common property resource use	Common pool resource
Satake et al.	2007	Coupled ecological–social dynamics in a forested landscape: Spatial interactions and information flow	Land use
Iwasa et al.	2007	Nonlinear behavior of the socio-economic dynamics for lake eutrophication control	Lake eutrophication
Suzuki & Iwasa	2009	The coupled dynamics of human socio-economic choice and lake water system: the interaction of two sources of nonlinearity	Lake eutrophication
Iwasa et al.	2010	Paradox of nutrient removal in coupled socioeconomic and ecological dynamics for lake water pollution	Lake eutrophication
Figueiredo & Pereira	2011	Regime shifts in a socio-ecological model of farmland abandonment	Land use
Tavoni et al.	2012	The survival of the conformist: Social pressure and renewable resource management	Common pool resource
Lade et al.	2013	Regime shifts in a social-ecological system	Common pool resource
Iwasa & Lee	2013	Graduated punishment is efficient in resource management if people are heterogeneous	Fishery
Richter et al.	2013	Contagious cooperation, temptation, and ecosystem collapse	Common pool resource
Richter & Grasman	2013	The transmission of sustainable harvesting norms when agents are conditionally cooperative	Common pool resource
Barlow et al.	2014	Modelling interactions between forest pest invasions and human decisions regarding firewood transport restrictions	Pest
Vasconcelos et al.	2014	Climate policies under wealth inequality	Climate
Ali et al.	2015	Coupled human-environment dynamics of forest pest spread and control in a multipatch, stochastic setting	Pest
Sugiarto et al.	2015	Socioecological regime shifts in the setting of complex social interactions	Common pool resource
Wiedermann et al.	2015	Macroscopic description of complex adaptive networks coevolving with dynamic node states	Private resource
Richter & Dakos	2015	Profit fluctuations signal eroding resilience of natural resources	Common pool resource
Schlüter et al.	2016	Robustness of norm-driven cooperation in the commons	Common pool

			resource
Weitz et al.	2016	An oscillating tragedy of the commons in replicator dynamics with game-environment feedback	Common pool resource
Bauch et al.	2016	Early warning signals of regime shifts in coupled human–environment systems	Forest
Henderson et al.	2016	Alternative stable states and the sustainability of forests, grasslands, and agriculture	Land use
Sugiarto et al.	2017	Social cooperation and disharmony in communities mediated through common pool resource exploitation	Common pool resource
Barfuss et al.	2017	Sustainable use of renewable resources in a stylized social–ecological network model under heterogeneous resource distribution	Private resource
Lafuite et al.	2017	Delayed behavioral shifts undermine the sustainability of social–ecological systems	Land use
Lindkvist et al.	2017	Strategies for sustainable management of renewable resources during environmental change	Common pool resource
Osten et al.	2017	Sustainability is possible despite greed - Exploring the nexus between profitability and sustainability in common pool resource systems	Common pool resource
Sigdel et al.	2017	Competition between injunctive social norms and conservation priorities gives rise to complex dynamics in a model of forest growth and opinion dynamics	Forest
Sugiarto et al.	2017	Emergence of cooperation in a coupled socioecological system through a direct or an indirect social control mechanism	Common pool resource
Thampi et al.	2018	Socio-ecological dynamics of Caribbean coral reef ecosystems and conservation opinion propagation	Coral reef
Chen & Szolnoki	2018	Punishment and inspection for governing the commons in a feedback-evolving game	Common pool resource
Drechsler & Surun	2018	Land-use and species tipping points in a coupled ecological-economic model	Land use
Geier et al.	2019	The physics of governance networks: critical transitions in contagion dynamics on multilayer adaptive networks with application to the sustainable use of renewable resources	Private resource
Hauert et al.	2019	Asymmetric evolutionary games with environmental feedback	Common pool resource
Lin & Weitz	2019	Spatial interactions and oscillatory tragedies of the commons	Common pool resource
Sigdel et al.	2019	Convergence of socio-ecological dynamics in disparate ecological systems under strong coupling to human social systems	Common pool resource

Bury et al.	2019	Charting pathways to climate change mitigation in a coupled socio-climate model	Climate
Shao et al.	2019	Evolutionary dynamics of group cooperation with asymmetrical environmental feedback	Common pool resource
Barfuss et al.	2020	Caring for the future can turn tragedy into comedy for long-term collective action under risk of collapse	Common pool resource
Tilman et al.	2020	Evolutionary games with environmental feedbacks	Common pool resource
Muneepeerakul & Anderies	2020	The emergence and resilience of self-organized governance in coupled infrastructure systems	Water use
Sun & Hilker	2020	Analyzing the mutual feedbacks between lake pollution and human behavior in a mathematical social-ecological model	Lake eutrophication
Mathias et al.	2020	Exploring non-linear transition pathways in social-ecological systems	Common pool resource
Phillips et al.	2020	Spatial early warning signals of social and epidemiological tipping points in a coupled behavior-disease network	Epidemic
Menard et al.	2021	When conflicts get heated, so does the planet: coupled social-climate dynamics under inequality	Climate
Phillips & Bauch	2021	Network structural metrics as early warning signals of widespread vaccine refusal in social-epidemiological networks	Epidemic
Holstein et al.	2021	Optimization of coupling and global collapse in diffusively coupled socio-ecological resource exploitation networks	Private resource
Farahbakhsh et al.	2021	Best response dynamics improve sustainability and equity outcomes in common-pool resources problems, compared to imitation dynamics	Common pool resource
Yan et al.	2021	Cooperator driven oscillation in a time-delayed feedback-evolving game	Common pool resource
Müller et al.	2021	Anticipation-induced social tipping: can the environment be stabilised by social dynamics?	Climate
Milne et al.	2021	Local overfishing patterns have regional effects on health of coral, and economic transitions can promote its recovery	Coral reef
Moore et al.	2022	Determinants of emissions pathways in the coupled climate-social system	Climate
Bengochea Paz et al.	2022	Habitat percolation transition undermines sustainability in socioecological agricultural systems	Land use

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