

1 Tipping points in coupled human-environmental environment 2 system models: a review

3 Isaiah Farahbakhsh¹, Chris T. Bauch², Madhur Anand¹

4 ¹School of Environmental Sciences, University of Guelph, Guelph, N1G 2W1, Canada

5 ²Department of Applied Mathematics, University of Waterloo, Waterloo, N2L 3G1, Canada

6 *Correspondence to:* Madhur Anand (manand@uoguelph.ca)

7 **Abstract.** Mathematical models that couple human behavior to environmental processes can offer valuable insights
8 into how human behavior affects various types of ecological, climate, and epidemiological systems. ~~In many~~
9 ~~coupled human-environmental systems with tipping points, gradual~~ This review focuses on human drivers of tipping
10 ~~events in coupled human-environment systems where~~ changes to the human system can lead abruptly to desirable or
11 undesirable new human-environmental environment states. We use snowball sampling ~~from relevant search terms~~ to
12 review the ~~modelling~~ modeling of social processes—such as social norms and rates of social change—that are shown to
13 drive tipping ~~points~~ events, finding that many affect the coupled system depending on the system type and initial
14 conditions. For example, tipping points can manifest very differently in ~~input- versus~~
15 ~~output-limited~~ human-extraction versus human-emission systems. Some potential interventions, such as reducing
16 costs associated with sustainable behavior, have intuitive results. However, their beneficial outcomes via less
17 obvious tipping ~~point~~ behavior events are highlighted. Of the models reviewed, we found that greater structural
18 complexity can be associated with increased potential for tipping ~~points~~ events. We review generic and
19 state-of-the-art techniques in early warning signals of tipping ~~points~~ events and identify significant opportunities to
20 ~~utilise~~ utilize digital social data to look for such signals. We conclude with an outline of challenges and promising
21 future directions specific to furthering our understanding and informing policy that promotes sustainability within
22 coupled human-environmental environment systems.

23

24 **Non-technical summary.** Mathematical models that include interactions between humans and the environment can
25 provide valuable information to further our understanding of tipping points. Many social processes such as social
26 norms and rates of social change can affect these tipping points in ways that are often specific to the system being
27 ~~modelled~~ modeled. Higher complexity of social structure can increase the likelihood of these transitions. We discuss
28 how data is used to predict tipping ~~points~~ events across many ~~coupled~~ systems.

29 **1 Introduction to tipping points in coupled human-environmental environment systems models**

30 Humans are facing environmental catastrophes of their own making, like climate change and biodiversity declines,
31 at local and global scales, and yet avoiding these catastrophes still poses complex challenges for sustainable
32 behavior and policy interventions (Steffen et al., 2017). Traditionally, mathematical models of environmental
33 systems have represented human impacts through fixed, static parameters or functions independent of the
34 environment's current state (Binford et al., 1987; Bosch, 1971; Chaudhuri, 1986; Getz, 1980), and these models can
35 be useful to inform optimal levels of sustainable extraction for short timescales. However, for longer timescales,
36 where human dynamics can evolve, it may be necessary to include human behavior endemically in the
37 ~~modelling~~ modeling framework to allow for human-environmental environment feedback to occur (Bauch et al.,
38 2016; Innes et al., 2013; Lade et al., 2013; Schlüter et al., 2012). Coupled human-environmental environment system
39 (CHES) models combine environmental (e.g., ecological, epidemiological, and climate) models with human
40 behavior and population dynamics (Bury et al., 2019; Carpenter et al., 2009; Farahbakhsh et al., 2022; Innes et al.,
41 2013; Lade et al., 2013; Phillips et al., 2020; Sethi and Somanathan, 1996). ~~The human and environmental~~
42 ~~subsystems of the coupled system have two-way (positive and/or negative) feedback, such that changes in each~~
43 ~~subsystem influence one another. For example, in Innes (2013), the amount of forest cover influences the proportion~~
44 ~~of the population that conserves forest ecosystems.~~ For example, in Innes (2013), the amount of forest cover
45 influences the proportion of the population that conserves forest ecosystems. The influence of each subsystem on
46 one another often occurs as two-way (positive and/or negative) feedback loops. In a positive (self-reinforcing)
47 feedback loop, variable 'A' causes an increase in variable 'B' which then causes an increase in 'A'. In a negative
48 feedback loop, 'A' causes an increase (respectively, decrease) in 'B' which causes a decrease (respectively, increase)
49 in 'A'. The inclusion of these feedbacks leads to increased diversity in the qualitative behavior of the system, such as
50 whether the long-term dynamics converge to a sustainable or depleted environmental state, or cycle over time.
51 Negative feedback promotes a return to equilibrium (Figure 1a2a) and can increase the system's capacity to respond
52 to disturbances and adapt in ways that allow the system to maintain the function of social and ecosystem services,
53 which is sometimes referred to as "resilience" (Folke, 2006).
54
55 Human-environmental environment negative feedback loops via processes such as public concern pressuring
56 governments to introduce environmental legislation can be powerful and there are many historical examples of it
57 occurring (Dunlap, 2014; Grier, 1982; Mather and Fairbairn, 2000; Stadelmann-Steffen et al., 2021). Forest cover in
58 Switzerland doubled, following an all-time low in the first half of the 19th century. This was brought about by rapid
59 ~~population growth and early industrialisation. Wood shortages and floods led to public concern, triggering public~~
60 ~~concern responding to food shortages and floods, which triggered~~ local regulation, the formation of the Swiss
61 Forestry Society, and the first federal forestry law enacted in 1876 ~~that in turn caused a recovery of forest cover~~
62 (Mather and Fairbairn, 2000). Similarly, the bald eagle population in North America recovered significantly after the
63 banning of DDT by the EPA in 1972. This was instigated by public outcry following the publication of Rachel
64 Carson's *A Silent Spring* in 1962 which linked DDT in the environment to low reproduction of birds and their

65 declining population (Dunlap, 2014; Grier, 1982). In both cases, the gradual recovery of the population was not
66 brought about simply by governmental legislation. There were also strong movements in the public and scientific
67 spheres, directly responding to perceived environmental risk which pressured governing bodies to enact immediate
68 reform (Dunlap, 2014; Grier, 1982; Mather and Fairbairn, 2000). We interpret these two examples as negative
69 feedback loops in a coupled human-environmental environment system because a decline in forest/eagle abundance
70 stimulated a response by humans which led to the recovery of the environmental system (Figure 1a2a). These
71 negative feedback loops are pervasive in the CHES models that we reviewexamine here.

72

~~73 In contrast to negative feedback that promotes an eventual and often gradual return to equilibrium, tipping points
74 describe a phenomenon in complex systems near an equilibrium where gradual changes in external conditions lead
75 to abrupt and lasting shifts in the system state and characteristic behavior (also referred to as a “regime”). One way
76 tipping points may occur is through nonlinear self-reinforcing mechanisms known as positive feedback loops, which
77 amplify these gradual changes, propelling the system into a new stable state in ways that are often difficult to
78 reverse. Such transitions have been extensively modelled using dynamical systems theory, where they exemplify a
79 type of “bifurcation” (Ashwin et al., 2012; Crawford, 1991; Dakos. The historical examples above describe negative
80 feedbacks promoting a return to a single environmentally beneficial equilibrium; however, in many cases, this does
81 not happen and the system can persist in a depleted state. For example, the desertification of regions once rich in
82 vegetation could become a positive feedback loop maintaining the new desert state (Hopcroft and Valdes, 2021;
83 Pausata et al., 2020). When systems can persist in qualitatively different states (also referred to as “regimes”), we
84 say that they exhibit alternative stable states (May, 1977; Lenton et al., 2008, Henderson et al. 2016). In
85 mathematical models, alternative stable states are self-reinforcing for a range of parameters, for example, low
86 harvest rates can promote a state of high biomass and high harvest rates can promote a state of low biomass in many
87 extractive CHES (Farahbakhsh et al., 2021; Henderson et al., 2016; Richter and Dakos, 2015; Richter et al., 2008;
88 Lenton 2013; Schlüter et al., 2008). Additionally, many systems with tipping points exhibit alternative stable states,
89 where the system has the potential to persist over long periods of time in one of multiple states under the same
90 parameters (May, 1977; Lenton et al., 2008, Henderson et al. 2016). In many cases, a return to the system's previous
91 state can be more difficult than anticipated, requiring additional effort rather than merely a return to parameters
92 before the tipping point, a phenomenon known as hysteresis, which can make mitigation and adaptation efforts
93 challenging. ¶~~

94 2016). Tipping points refer to critical points on this boundary between two alternative stable states. Near this
95 boundary, small perturbations can be amplified through nonlinear self-reinforcing positive feedback loops. This
96 leads to a qualitatively different system state and characteristic behavior, known as a “regime shift”, in a relatively
97 short amount of time. When the system has entered a new regime, there are often positive or negative feedback
98 loops that make it difficult to reverse this change. This self-perpetuating nature of some initial change through
99 nonlinear feedbacks leading to qualitative and often long-term system change is a universal characteristic of many

100 commonly studied tipping points. In many cases, a return to the system's previous state can be more difficult than
 101 anticipated, requiring additional effort rather than merely a return to parameters before the tipping point, a
 102 phenomenon known as hysteresis, which can make mitigation and adaptation efforts challenging. Systems near a
 103 tipping point can exhibit (often abrupt) regime shifts through gradual changes or noise in forcing parameters, which
 104 is a main focus of much of the bifurcation theory literature (Figure 1a, Box 1.1). The scope of models presented in
 105 this review will not include other types of tipping points such as those caused by a short sharp shock (s-tipping, or
 106 shock-tipping, where the system does not have to exist near this point for a regime shift to occur) (Figure 1b)
 107 (Boettiger and Batt, 2020; Halekotte and Feudel, 2020) or “rate-induced tipping”, which is a distinct phenomenon
 108 induced by the rate of change of parameters (Ashwin et al., 2012). Tipping events describe the crossing of a tipping
 109 point and can be used interchangeably with regime shifts.

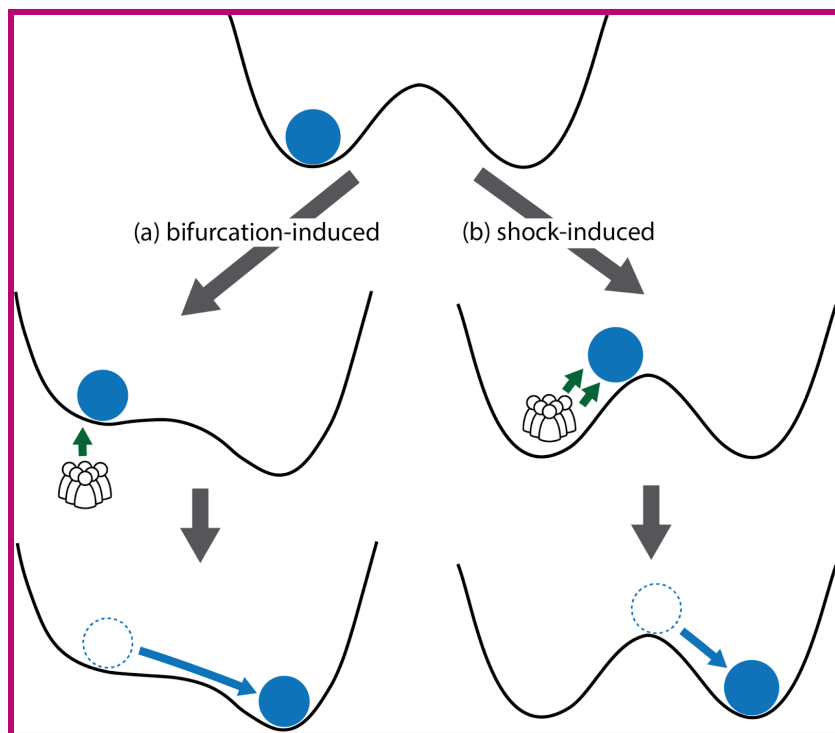


Figure 1: Two types of tipping events; bifurcation-induced tipping (a), where the drivers are gradual changes to system parameters leading to a tipping event, and shock-induced tipping (b), where a perturbation to the system causes it to enter an alternative stable state through the crossing of a tipping point. Many social tipping points are caused by a combination of both types of tipping events. The blue circle represents the current state of the system.

110

111 Bifurcation theory has been applied to study tipping points in a vast number of environmental models (May and
 112 Oster, 1976; Brovkin et al., 1998; Ghil and Tavantzis, 1983; Wollkind et al., 1988); however, more recently,
 113 researchers have identified abrupt shifts in environmental systems for which bifurcation theory has yet to be

114 explicitly applied (Dakos et al., 2019; Lenton, 2020, 2013). For example, during the mid–Holocene, the Sahara was
115 much more humid than at present, showing evidence of shrub and savannah biomes as well as the expansion of
116 lakes, an alternative stable state to what we know as its current desert state. It is ~~hypothesised~~ hypothesized that
117 around 5,000 years ago, the gradual weakening of the North African Monsoon led to an abrupt decrease in
118 vegetative cover, due to positive feedback between reduced surface albedo and precipitation, bringing the Sahara
119 into a stable desert state (Hopcroft and Valdes, 2021; Pausata et al., 2020). In more dominantly human systems,
120 many pivotal revolutions can also be framed as tipping ~~points~~ events where gradual changes are reinforced by
121 positive feedback loops, leading to a new political or technological stable state (Lenton et al., 2022). Social tipping
122 ~~points~~ also ~~occur~~ occurs in financial systems such as in the 2008 financial crisis. Here, the bankruptcy of Lehman
123 Brothers led to a rise in public panic around the stability of markets, causing banks to increase their liquidity,
124 amplifying the crisis in other economic sectors and leading to a global recession (Van Nes et al., 2016). These are
125 just two of many examples illustrating how important tipping points are as a phenomenon, in both human and
126 environmental systems, and coupling these systems using mathematical models could lead to further insights.

127

128 Since the beginning of the Anthropocene and with our growing awareness of human impacts on the environment,
129 tipping points are increasingly being ~~conceptualised~~ conceptualized within the context of coupled
130 ~~human-environmental~~ environment systems (Bauch et al., 2016; Henderson et al., 2016; Lenton et al., 2022;
131 Milkoreit et al., 2018). Tipping ~~points~~ events can lead to highly beneficial or catastrophic outcomes for humans,
132 especially when an environmental change occurs in the presence of social hysteresis. An example of detrimental
133 tipping is in the forests of Kumaun and Garhwal in Northern India, where, prior to British ~~colonisation~~ colonization,
134 wood harvest was sustainably regulated through social norms and strict rules enforced by local village councils.
135 When the British colonial government imposed ~~their~~ its own rules on the use of forests, these social norms broke
136 down. Eventually, protests led to British lumber restrictions being removed, but the system subsequently
137 experienced rapid deforestation rather than a return to its previous levels under local management. Here, the social
138 system crossed a tipping point between a self-organized common property regime to one of open access devoid of
139 self-regulating sanctions (Somanathan, 1991). This system has been ~~modelled~~ modeled using a dynamical systems
140 approach that allows for a quantitative understanding of the human drivers leading to ~~the tipping points~~ these tipping
141 events (Sethi and Somanathan, 1996). Contrasting this example, tipping ~~points~~ events can also result in
142 environmental change that is beneficial to humans and the environment. The rapid response of the international
143 community to the hole in the ozone layer has been interpreted by some as an example of a ~~system~~ CHES undergoing
144 tipping ~~points~~ events caused by ~~human-environmental feedback~~ (Stadelmann-Steffen et al., 2021). First, there was a
145 ~~shift in public opinion regarding the use of CFC products, causing a change in behavioral norms and pressure on~~
146 ~~political institutions to follow suit. Then when policy was passed, industry shifted abruptly to producing CFC~~
147 ~~alternatives, which led to a tipping point in CFC emissions bringing about a new self-perpetuating change through~~
148 political, technological, and behavioral forces (Stadelmann-Steffen et al., 2021). In the 1970s, scientists

149 demonstrated the detrimental effects of CFCs on the ozone layer, which could be viewed as the initial driver of the
150 following socio-climate tipping events. This led to public concern, prompting several countries to ban the use of
151 CFCs in aerosols. Through the enactment of national policies, public awareness increased, leading to more public
152 pressure for national and international policy change, an example of a positive feedback loop. In parallel, these
153 national bans of CFCs, especially in the US, led to the development of CFC alternatives, which prompted industries
154 that could develop them to lobby for international policy. Increased public awareness also led to widespread shifts in
155 social norms stigmatizing and boycotting the consumption of CFCs, which further pressured industry to offer
156 alternatives, another positive feedback loop. The interaction of multiple tipping events at different scales led to the
157 crossing of a global tipping point through the international banning of CFCs, bringing an alternative stable state of
158 relatively very low CFC emissions globally. (Andersen et al., 2013; Cook, 1990; Epstein et al., 2014; Haas, 1992;
159 Stadelmann-Steffen et al., 2021).

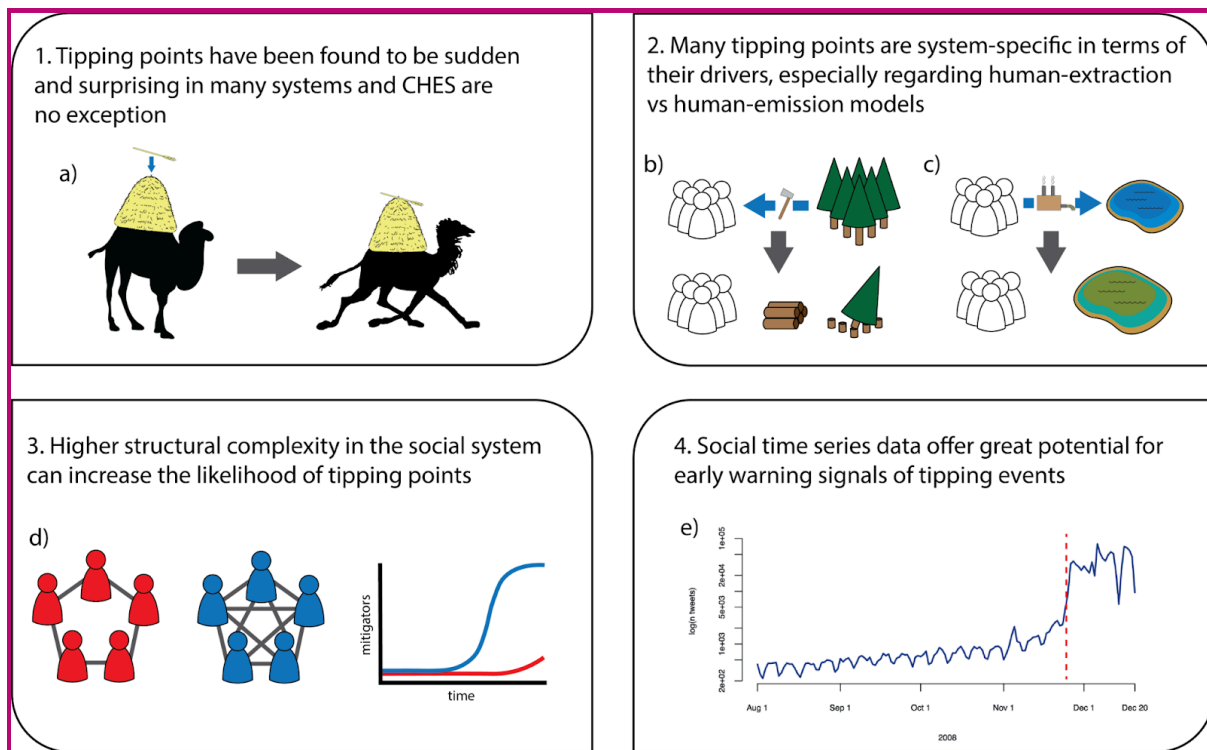
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161 Tipping ~~points~~ events associated with social processes as described in the preceding paragraph can be
162 ~~conceptualised~~ conceptualized through positive feedback loops that capture a self-reinforcing process. In the case of
163 social norms, this self-reinforcing process may correspond to peer pressure or conformism that reinforces the
164 dominant opinion or belief. Depending on whether pro- or anti-mitigation opinions are currently dominant, this
165 could lead to hysteresis (Figure 1a2b). The negative feedback loop that might normally regulate the CHES to exist
166 in a state of intermediate environmental health and public support for sustainability (Figure 1a2a) could be
167 overpowered by the positive feedback of social norms, leading the population to a state where either sustainability
168 (or anti-sustainability) is strongly entrenched. If the conditions governing social learning or social norms move
169 beyond a tipping point, the population may flip between these two norms, or alternatively it may move into a regime
170 where social norms are instead dominated by the negative feedback loop, causing the population to exist in an
171 interior state of partial sustainability. As such, negative feedback and positive feedback may be characteristic of any
172 CHES and should be systematically studied.

173

174 This review aims to deepen our understanding of human drivers of tipping ~~points~~ events in CHES models by
175 exploring three crucial topics: the feedback loops and interactions between the human and environmental systems,
176 the structural characteristics of the human system that influence tipping points, and the identification of early
177 warning signals within human systems. By “human drivers”, we refer to the ~~gradual~~ changes in social parameters
178 that elicit these non-linear tipping responses in either the environment, human system, or both. However, we also
179 discuss aspects of social structure that may be conducive to tipping points. **As most of the models reviewed are
180 informed by dynamical systems and bifurcation theory, we primarily focus on systems that exist near tipping points
181 and cross them through gradual changes in these drivers.** In the following sections we review CHES model literature
182 found using Google Scholar with the keywords: ‘human environment system’ OR ‘socio-ecological system’ OR
183 ‘social ecological system’ OR ‘human ecological system’ OR ‘human natural system’ combined with ‘tipping’ OR

184 ‘regime shift’ OR ‘bifurcation’. These results were filtered manually to include only dynamical models that showed
 185 clear tipping behavior. Additional literature was found through a snowball approach using references from the
 186 sources found in this search as well as papers referencing these sources (Wohlin, 2014). The findings in this review
 187 highlight commonalities between the CHES models surveyed; however, some trends may be a result of both the
 188 dynamical models chosen and the relatively low diversity and volume of these models.¶
 189 ~~2 Structures and processes~~ The body of this review is split into two parts; the first part synthesizes results from
 190 CHES models, organized into processes and structures that drive tipping behavior, and the second part introduces
 191 early warning signals describing how they can be used to predict tipping events.



Box 1: Highlights of key findings from the synthesis of CHES models in this review. “The straw that broke the camel’s back” illustrating bifurcation-induced tipping points (1a), in human-extraction systems (2b), increasing the speed of social change or the coupling strength leads to negative tipping points (i.e., ecological collapse), whereas in human-emission systems (2c), the effects of increasing the speed of social change or the coupling strength are model specific, higher connections in a social network leading to a positive tipping event, where the graph represents the proportion of mitigators in time (3d), time series data from Twitter showing an abrupt transition characteristic of a tipping event at the red dotted line (4e) from (Bollen et al., 2021).

192 **2 Processes and structures in human systems that cause tipping points** in CHES models

193 In this section, we look at how social processes and structures cause tipping ~~points~~ events. In order to have a better
194 understanding of how these human drivers affect tipping, it is important to understand the basics of
195 ~~modelling~~ modeling human systems. Within CHES models, various factors, such as economic incentives,
196 environmental considerations, and social pressures determine how individuals make decisions and interact with the
197 environment. In most of the current ~~modelling~~ modeling literature, individuals can choose between two behaviors
198 (also referred to as opinions or strategies), one that is environmentally sustainable (also referred to as mitigation or
199 cooperation) and another that is detrimental to the environment (also referred to as non-mitigation or defection). The
200 perceived advantage of mitigation or non-mitigation relative to the current state of the human and environmental
201 system can be quantified through a “utility function”. Common factors in the utility function are the rate of social
202 learning, which determines the speed of human behavior change relative to environmental processes, social norms,
203 which encourage the status quo or mitigation proportional to its frequency, cost of mitigation, which measures the
204 economic cost of being a mitigator relative to a non-mitigator, and rarity-motivated valuation, which incentivizes
205 mitigation as the environment approaches collapse (Bauch et al., 2016; Farahbakhsh et al., 2022; Tavoni et al.,
206 2012). In most models that use social learning, individuals sample others in the population at a fixed rate and adopt a
207 different behavior if the other behavior has a higher utility, with probability proportional to the difference in utility
208 (Hofbauer and Sigmund, 1998; Schuster and Sigmund, 1983). This can also be formulated in a stochastic setting,
209 where the probability of adopting a neighbor’s behavior is a function of the difference in utility between behaviors
210 (Schlag, 1998). Most of the models reviewed in this paper use social learning to represent human behavioral
211 dynamics. There are also CHES models that do not include social learning such as Motesharrei (2014) and
212 Dockstader (2019) where the human population is influenced by its current size and the state of the environment;
213 however, these are outside the scope of this paper.

214

215 Many human behaviors, such as resource extraction and pollution, have direct detrimental impacts on the
216 environment; however, the severity of these impacts is often hard to predict. In many CHES models, small changes
217 in parameters governing human behavior and social processes can lead to the abrupt collapse of sustainable states
218 through tipping ~~points~~ events that can cascade between the human and environmental systems (Bauch et al., 2016;
219 Lade et al., 2013; Richter and Dakos, 2015; Weitz et al., 2016). Additionally, structural elements of the human
220 system, ~~such as the~~ (i.e. an individual’s degree of choice ~~and individual~~, population diversity), as well as how the
221 social system is ~~organised~~ organized (i.e. through social networks), can affect tipping. These heterogeneous model
222 elements are often only accessible in agent-based models, where humans are represented as individual agents that
223 follow a set of rules. CHES models do not always exhibit tipping points under realistic settings for the human
224 system (Bury et al., 2019; Menard et al., 2021); however, in this review, we focus on models with tipping points.

225 2.1 Coupling strength

226 Coupling strength (how strongly the subsystems are coupled) can have a significant effect on the occurrence of
227 tipping points in both systems, and the nature of these transitions often depends on whether systems are
228 ~~‘input-limited human-extraction’ or ‘output-limited’~~. In ~~input-limited systems~~ **human-emission’ (Box 1.2)**. In
229 **human-extraction systems (Box 1.2b)**, humans extract from an environmental resource such as in forest and fishery
230 models. Stronger coupling in ~~input-limited~~ **human-extraction** models often leads to **negative** environmental
231 ~~collapse~~ **outcomes**. A common social parameter representing the coupling strength in these systems is the extraction
232 effort of humans, which when increased past a critical threshold, leads to abrupt environmental collapse
233 (Farahbakhsh et al., 2021; Richter and Dakos, 2015; Richter et al., 2013; Schlüter et al., 2016). For ~~output-limited~~
234 ~~systems~~ **human-emission systems (Box 1.2c)**, where human activity increases levels of harmful outputs, such as
235 pollution and climate models, coupling strength is instead represented by pollution rates. The influence of this
236 coupling is less intuitive ~~than extraction effort~~ **in human-emission systems**, for example, in lake
237 ~~pollution~~ **eutrophication** models as the pollution ~~output~~ of mitigators is decreased, pollution levels also decrease until
238 a threshold is reached, heralding a detrimental tipping point where mitigation collapses and pollution then reaches a
239 high level (Iwasa et al., 2010, 2007). This occurs because when the lake water is not very polluted, there is less
240 incentive to be a mitigator and high-polluting behavior becomes a new norm. It is important to note that these
241 models do not account for individuals valuing the environment in a healthy state, for example through the centering
242 of ecosystem services, and the above example may be an ~~artefact~~ **artifact** of this assumption. There is a need to shift
243 both our relationship to the environment as well as the assumptions in our models so that inherent value in
244 environmental systems is central in any decision-making, even when the environment is far from collapse. This
245 fundamental valuing of the environment is present in many traditional indigenous belief systems, where
246 relationships to the local natural environment are incorporated and ~~prioritised~~ **prioritized** in all aspects of life
247 (Appiah-Opoku, 2007; Bavikatte and Bennett, 2015; Beckford et al., 2010; McMillan and Prosper, 2016).

248 **2.2 Rarity-motivated valuation**

249 Rarity-motivated valuation represents the extent to which humans increase their mitigative behavior in response to
250 the environmental variable (e.g., forest cover, endangered species population size) nearing a depleted state. ~~Model~~
251 ~~systems with rarity-motivated valuation often exhibit two tipping points at high and low levels, with a sustainable~~
252 ~~regime for intermediate values. High levels of rarity-motivated valuation lead to overshoot dynamics, however, this~~
253 ~~may not be true in empirical systems. In models, the~~ **In CHES models, this** sensitivity of human response to the
254 abundance of the natural resource/population is represented by a ‘sensitivity’ parameter and there are often two
255 critical thresholds in the sensitivity parameter that lead to tipping. Increasing the sensitivity parameter beyond the
256 lower threshold induces a tipping point from a depleted to sustainable environmental equilibrium (Ali et al., 2015;
257 Barlow et al., 2014; Bauch et al., 2016; Drechsler and Surun, 2018; Henderson et al., 2016; Lin and Weitz, 2019;
258 Sun and Hilker, 2020; Thampi et al., 2018; Weitz et al., 2016). The second threshold exists at high values of the
259 sensitivity parameter, ~~where~~ **which may be counterintuitive, as one might expect high sensitivity to resource**

260 depletion to lead to more sustainable outcomes. In this case, the sustainable equilibrium is ~~destabilised~~destabilized
261 by overshoot dynamics or a state of chaos in both the human and environmental systems. These dynamics are caused
262 by the human system being too sensitive to changes in the environment, leading to extreme oscillations in both
263 human behavior and the environment, which increases the likelihood of collapse in mitigation and the state of the
264 environment (Bauch et al., 2016; Henderson et al., 2016).

265

266 Rarity-motivated valuation can also be represented by a threshold in the state of the environment, below which
267 humans shift towards sustainable behavior. In a common-pool resource model, lowering this threshold led to a series
268 of tipping points that surprisingly resulted in a higher biomass equilibrium, although the trajectory to this state
269 comes close to environmental collapse. This is in contrast to a high threshold, which leads to lower final biomass;
270 however, the trajectory remains much farther from a depleted environmental state (Mathias et al., 2020). Similarly to
271 high coupling in pollution models, one should be very careful to not interpret these results as stating “too much
272 conservation is detrimental to the environment”. They rest on model assumptions of a reactionary conservation
273 paradigm, where there is less value in conserving when the environment is in a healthy state.

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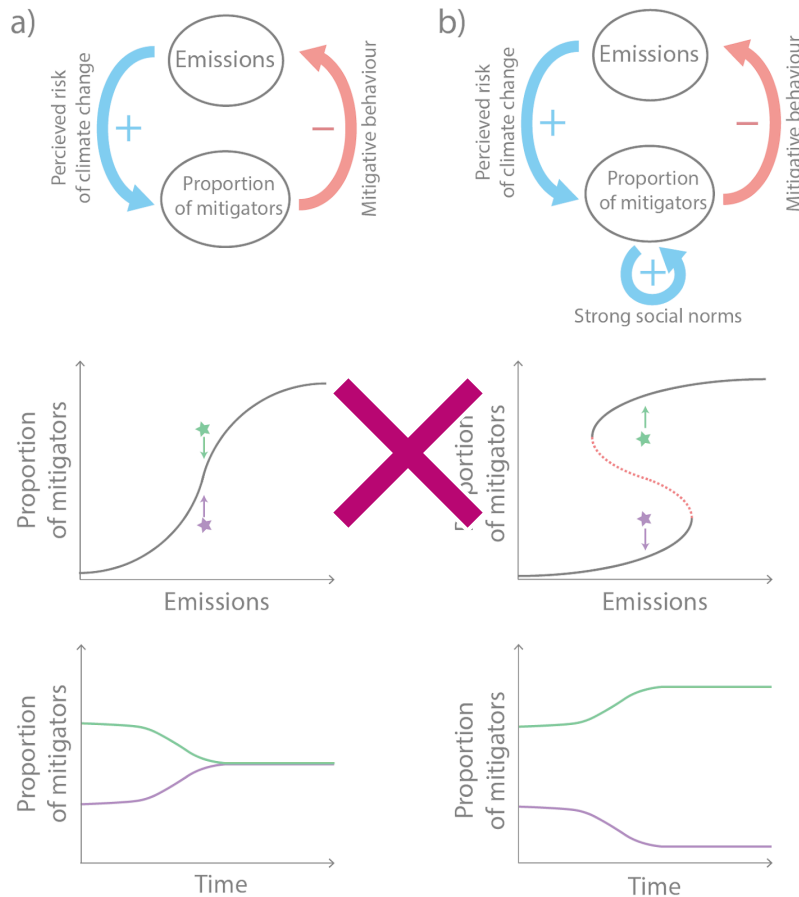


Figure 42: Negative feedback between the human and environmental subsystems, **supports** convergence to the same equilibrium regardless of initial conditions (a). With strong majority-enforcing social norms, encouraging either mitigative or harmful behavior adds a positive feedback loop which makes the coupled system highly dependent on initial conditions (b). The top row shows the negative feedback loop between emissions and the proportion of mitigators, where (b) also includes the positive feedback of majority-enforcing social norms. In the middle row, equilibrium curves are plotted as a function of the maximum emissions of non-mitigators. Black solid lines represent stable equilibria and the red dotted line represents unstable equilibria. The green and purple curves in the bottom row are the trajectories for initial mitigation support and emission value given by the stars of the corresponding color in the upper row.

275 2.3 Social norms

276 Introducing social norms can lead to alternative stable states and thus tipping points (Figure 42b), although the
 277 system dynamics are highly dependent on both the type of social norms and initial conditions. Social norms are
 278 informal rules emerging through social interaction that promote and discourage certain behaviors, especially around

279 how humans relate to one another and the environment (Chung and Rimal, 2016). In models of small groups such as
280 a community of fishers, they are often (rightly) assumed to support mitigative behavior by punishing those who
281 violate norms by over-harvesting (Ostrom, 2000). However, at larger population scales, social norms can support
282 either pro- or anti-mitigation behavior, on account of factors such as ~~politicisation~~the politicization of actions
283 relating to environmental, climate, and public health crises (Stoll-Kleemann et al., 2001; Van Boven et al., 2018;
284 Latkin et al., 2022). Unlike a fisher in a small community, for instance, a climate denier may not acknowledge
285 themselves as a ‘defector’ who is harming a public good, but rather view the climate activist as ‘defecting’ against a
286 free society. Thereby, social norms have the ability to encourage behavior that is harmful to both human and
287 environmental well-being, over larger spatial and temporal scales (Bury et al., 2019; Latkin et al., 2022; Menard et
288 al., 2021; Stoll-Kleemann et al., 2001; Van Boven et al., 2018).

289

290 Social norms can be represented as majority-enforcing, incentivizing the behavior of the majority, or
291 mitigation-enforcing, such as sanctions, which only incentivize mitigation, relative to the proportion of mitigators in
292 the current state of the system. In CHES models, increasing the strength of majority-enforcing norms leads to an
293 increased number of regimes as well as bistable (~~more than one stable state~~two stable states) regimes (Figure 4b2b),
294 made up of a single dominant behavior, which is highly dependent on the initial proportion of behaviors in a
295 population (Ali et al., 2015; Barlow et al., 2014; Bauch et al., 2016; Bury et al., 2019; Phillips et al., 2020; Sigdel et
296 al., 2017; Thampi et al., 2018). This occurs because these norms are indifferent to the type of behavior they enforce
297 (i.e. sustainable vs harmful actions), and they act as a double-edged sword that reinforces the status quo through a
298 positive feedback loop, where the dominant behavior becomes more prevalent (Figure 4b2b). On the other hand,
299 increasing mitigation-enforcing social norms lead to a transition of the environmental system into a sustainable
300 equilibrium (Chen and Szolnoki, 2018; Iwasa et al., 2010; Lafuite et al., 2017; Moore et al., 2022; Schlüter et al.,
301 2016; Tavoni et al., 2012), sometimes through an intermediate regime of oscillatory dynamics (Iwasa et al., 2007).
302 In a lake pollution model, along with decreasing the likelihood of environmental collapse, this increase in
303 mitigation-enforcing social norms also led to the appearance of alternate stable states (Sun and Hilker, 2020). These
304 findings show that stronger social norms lead to a greater number of tipping points; however, the trajectories brought
305 about by these tipping points are highly dependent on the type of social norms (mitigation- or majority-enforcing) as
306 well as the current dominant social behavior.

307 2.4 Cost of mitigation

308 Reducing the cost of mitigation often leads to beneficial tipping points; however, these tipping points can depend on
309 the rate of social change as well as social norms. Although it is intuitive that reducing costs or increasing economic
310 incentives associated with mitigative action will have beneficial impacts on the environment, CHES models also
311 show that this beneficial change can occur through tipping points (Bauch et al., 2016; Drechsler and Surun, 2018;
312 Milne et al., 2021; Moore et al., 2022; Sigdel et al., 2017; Thampi et al., 2018). In coupled social-epidemiological

313 models, where the environmental state is the proportion of infected individuals, mitigation cost is represented
314 through the economic cost or perceived risk of vaccination. Decreasing this cost leads to beneficial tipping points
315 from a state with low pro-vaccine opinion and vaccine coverage to high pro-vaccine opinion and vaccine coverage
316 (Phillips et al., 2020). Conversely, increasing this cost leads to a state of high infection and low vaccination. This
317 detrimental tipping point occurs in the human system at lower levels of vaccination cost when majority-enforcing
318 social norms are low, leading to widespread anti-vaccine opinion before the infection becomes endemic again
319 (Phillips and Bauch, 2021). Decreasing profits of individuals engaging in non-mitigative behavior can also lead to
320 an abrupt shift to a state of pure mitigators (Shao et al., 2019; Wiedermann et al., 2015); however, this transition can
321 be dependent on a low rate of social change (Wiedermann et al., 2015). Other models demonstrate tipping in the
322 other direction where increasing non-mitigators' payoff brings about a regime shift to pure non-mitigation and
323 environmental collapse (Richter et al., 2013; Tavoni et al., 2012). Similarly, a common-pool resource model that
324 uses machine learning in a continuous strategy space shows tipping to a depleted resource regime when the costs
325 associated with harvesting are too low (Osten et al., 2017). An analog to mitigation cost is taxation rates, which
326 resource users pay towards public infrastructure mediating resource extraction. In a model where individuals can
327 choose to work outside of the system, pushing taxation rates to high or low levels tips a sustainable regime where
328 institutions are at full or partial capacity to a collapse of institutions (Muneepeerakul and Anderies, 2020). In another
329 model, only individuals with high extractive effort are subject to taxation, and increasing this taxation rate brings
330 about a beneficial tipping point to a sustainable regime. However, the size of this sustainable region **in the parameter**
331 **space** is smaller with multiple governance nodes evolving through social learning compared to a single taxing entity
332 (Geier et al., 2019). However the cost of mitigation is represented, increasing the relative economic incentive of
333 mitigation behavior has the potential to bring about beneficial tipping to a sustainable regime.

334 **2.5 Rates of social change and time horizons**

335 Human and environmental change often occur on different timescales and their relative rates of change play a major
336 role in the long-term dynamics of the coupled system and whether or not tipping points will occur. Increasing the
337 rate of social change (in most cases, social learning) leads to collapse in **input-limited human-extraction** models due
338 to overshoot dynamics. ~~Whereas,~~ **whereas,** in **output-limited human-emission** models, the impacts of the rate of
339 social change are more model-specific. In both types of models, increasing the time horizon in decision-making is
340 beneficial. In CHES models, these rates of change can be controlled by the rate of social learning which determines
341 how frequently individuals interact and consequently, the pace of behavioral change within a population. Changes in
342 the speed of the human system can have very different outcomes depending on the nature of
343 ~~human-environmental~~ **environment** coupling (Box 1.2). In **input-limited human-extraction** models, increasing the
344 speed of the human system relative to the environment often ~~destabilises~~ **destabilizes** sustainable equilibria, leading
345 to oscillations in both systems and, in many cases, the abrupt collapse of the environmental system. These overshoot
346 dynamics occur as humans change their behavior too quickly to allow for the environment to ~~stabilise~~ **stabilize**. On

347 the other hand, decreasing the relative speed of human dynamics usually brings about beneficial tipping ~~points~~ events
348 leading to a state of high forest cover (Figueiredo and Pereira, 2011), and supporting mitigators for a
349 ~~generalised~~generalized resource (Hauert et al., 2019; Shao et al., 2019). These beneficial effects have also been
350 observed in adaptive network models where individuals imitate their neighbors depending on the profitability of
351 their strategies (~~Barfuss et al., 2017; Geier et al., 2019; Wiedermann et al., 2015~~). ~~The~~. In these models, the reduced
352 speed of social change leads to beneficial outcomes as the resource is allowed more time to ~~stabilise~~stabilize as
353 decisions regarding extractive levels occur (Barfuss et al., 2017; Geier et al., 2019; Wiedermann et al., 2015). Other
354 relative rates of change can also significantly influence the existence of a sustainable regime. For example, in an
355 agricultural land use model, increasing the speed of agricultural expansion and intensification relative to human
356 population growth leads to the collapse of both the natural land cover and human population (Bengochea Paz et al.,
357 2022).

358

359 In ~~output-limited~~human-emission models, increasing the speed of social interaction is more model-specific. In some
360 cases, such as forest-pest and climate systems, increasing the speed of the human system leads to better mitigation of
361 environmental harms in the short term. However, long-term sustainability often requires additional social
362 interventions such as reducing mitigation costs and increasing levels of environmental concern (Ali et al., 2015;
363 Barlow et al., 2014; Bury et al., 2019). In lake pollution models, higher relative speeds of social dynamics can
364 ~~destabilise~~destabilize low-pollution equilibria, leading to oscillations and eventually a polluted state with no
365 mitigation (Iwasa et al., 2010, 2007; Sun and Hilker, 2020). This is a similar phenomenon to the overshoot dynamics
366 that occur when the human system is extremely reactive to the environment discussed in the case of rarity-motivated
367 valuation; however, these outcomes are highly dependent on other social parameters. In a related model, with no
368 social hysteresis, represented by mitigation-enforcing social norms, and strong environmental hysteresis, represented
369 by a high phosphorus turnover rate, fast social dynamics could ~~stabilise~~stabilize oscillations, leading to a
370 low-pollution equilibrium (Suzuki and Iwasa, 2009). The emergence of oscillations under low rates of social
371 learning, which was not observed in similar models is likely due to the environmental system being in a bistable
372 state under strong hysteresis, such that even slow changes in the human system could tip the lake system ~~into~~ an
373 alternative stable state.

374

375 When looking at relative rates of change in human and environmental systems, it is clear that the pace of the human
376 system can be more readily influenced by interventions. This suggests an urgent need to further study the
377 relationship between social and ecological timescales across a wide range of coupled systems to aid in sustainable
378 policy-making decisions (Barfuss et al., 2017). Additionally in many models, the length of time horizons that
379 humans take into account when deciding how they interact with the environment has a significant beneficial effect
380 on conserving natural states and mitigating harmful action (Barfuss et al., 2020; Bury et al., 2019; Henderson et al.,
381 2016; Lindkvist et al., 2017; Müller et al., 2021; Satake et al., 2007). A high degree of foresight in decision-making

382 is a fundamental basis for many indigenous belief systems across the world. One manner in which this shows up is
383 in land stewardship where care for the environment is prioritized as a means to ensure the health of many
384 generations in the future (Appiah-Opoku, 2007; Beckford et al., 2010; Ratima et al., 2019).

385 2.6 Social traits

386 The inclusion and distribution of traits within agents can play a large role in determining the occurrence and types of
387 tipping points within the coupled system, where ~~increasing the modelled~~ **increasing the modeled** heterogeneity in
388 social traits can lead to more tipping and also promote sustainable outcomes (Box 1.3). The majority of models
389 discussed in the previous section only allow humans to choose between two strategies; mitigation and
390 non-mitigation. The inclusion of additional strategies, determining how individuals interact with the environment
391 and each other, can alter the potential for tipping points. For example, a common-pool resource model included a
392 third strategy of conditional mitigation (Richter and Grasman, 2013). Under this additional strategy, agents act as
393 mitigators until the number of non-mitigators reaches a certain threshold, where they then shift their behavior to
394 non-mitigation. The addition of this third strategy alters tipping dynamics in opposite ways, depending on the value
395 of maximum harvesting efforts. When efforts are high, the system is less prone to tipping; however, when they are
396 low, tipping points are more likely to occur. This third strategy also affects tipping points by masking internal social
397 dynamics, leading to more abrupt transitions, even when the system appears to be stable. This occurs when
398 mitigators gradually change their strategy to conditional mitigators which can go unnoticed as their interaction with
399 the environmental system does not change. However, when non-mitigation reaches high enough levels, there is a
400 cascade of conditional mitigators choosing non-mitigation, in an example of herd behavior, which puts abrupt
401 harvesting pressure on the resource. Another three-strategy model, where agents are partitioned by resource
402 extraction rates, contrasts dynamics with and without the trait of environmental concern (Mathias et al., 2020). In the
403 absence of this trait, the human system either tips to a state of high-extraction or low-extraction behavior, triggering
404 either a detrimental or beneficial environmental tipping point, respectively. Including environmental concern leads
405 to an increased number of cascading tipping points between both human and environmental systems. In a coupled
406 agricultural model, where human traits include management strategies that respond to socio-economic and climate
407 conditions, decreasing the diversity of these traits among agents in the system transitions the system from a
408 sustainable state with high food production, landscape aesthetics, and habitat protection to a state with low habitat
409 protection (Grêt-Regamey et al., 2019). As there are relatively few models that explicitly compare the complexity of
410 social traits and their effect on tipping points, it is difficult to say with certainty whether higher complexity will
411 increase the likelihood of tipping points in all CHES and whether this is due to a higher dimensionality of the
412 system. However, ~~these~~ **the commonalities between models showing the effects of social trait complexity** are worth
413 highlighting and will be put to the test with future CHES models and empirical work.

414 2.7 Social networks

415 In many agent-based CHES models, individuals are structured on a social network, where they ~~usually only~~ interact
416 with others whom they share a link with. These models demonstrate how a higher number of connections in social
417 networks increases the potential for tipping points, often through the emergence and growth of a ~~bistable~~
418 ~~regime~~ regimes (Holstein et al., 2021; Sugiarto et al., 2017a, 2015, 2017a) (Box 1.3). Additionally, the
419 distributions of these connections play an important role. For example, in networks with the same average number of
420 connections, higher heterogeneity of connections among nodes leads to tipping points occurring earlier under certain
421 social (Ising model) dynamics (Reisinger et al., 2022). The distribution of resources in
422 human-environmental environment networks also affects the potential for abrupt environmental collapse. This often
423 occurs in CHES network models where both human and environmental dynamics occur on a multi-layer network,
424 representing partitioned or private resources. Resource heterogeneity can be controlled through the distribution of
425 carrying capacities or the amount of resource flow between nodes in the network, where higher flows lead to
426 homogeneous resource distributions. In both cases, increasing this heterogeneity can tip the system to a state of low
427 extraction and high sustainability. Heterogeneity In one model, heterogeneity in carrying capacities increases the
428 likelihood of sustainable harvesters extracting from a resource with a large capacity, which they can maintain at high
429 levels (in contrast to non-sustainable harvesters who extract at a higher rate), eventually convincing neighboring
430 nodes to imitate their strategy (Barfuss et al., 2017). Heterogeneity In another model, heterogeneity through lower
431 resource flows also leads to high-extraction nodes over-exploiting their resource and losing profits in the long run,
432 de-incentivizing neighbors to imitate their behavior. Interestingly, optimal resource flow, which ~~minimises~~ minimizes
433 the likelihood of resource collapse is found to be close to the critical threshold of resource flow, above which the
434 coupled system collapses. As optimal resource flow decreases the likelihood of collapse by supplementing resources
435 harvested at high levels, this confers an advantage to high resource extraction. Increasing past optimal levels leads to
436 similar resource levels among high and low-extraction nodes, resulting in higher profits from high-extraction nodes,
437 incentivizing the entire human system to eventually choose the high-extraction strategy (Holstein et al., 2021).
438
439 Heterogeneity of human interaction can be quantified through homophily, the extent to which alike individuals
440 interact. Homophily can play a large role in the occurrence and behavior of tipping points in CHES models
441 occurring on social networks, often having a detrimental effect on the environmental system. In a common-pool
442 resource model with two distinct communities, increasing segregation by lowering the probability that agents in
443 separate communities will have a link, softens the abruptness of a single detrimental tipping point compared to when
444 the communities are well-mixed. This is due to the occurrence of multiple intermediate tipping points within each
445 segregated community; however, ~~increased~~ higher segregation adds more hysteresis to the system increasing the
446 difficulty of reversing this transition and returning to a sustainable state (Sugiarto et al., 2017b). In a public goods
447 game ~~modelling~~ modeling climate change mitigation, where humans are partitioned into rich and poor agents, a
448 transition to group achievement of mitigation goals occurs at a lower perceived risk when there is no homophily and
449 agents are influenced by others from both economic classes equally (Vasconcelos et al., 2014). Another

450 human-climate model that included wealth inequality displayed an abrupt transition to lower peak temperature
 451 anomalies when homophily between economic classes approached zero (Menard et al., 2021).

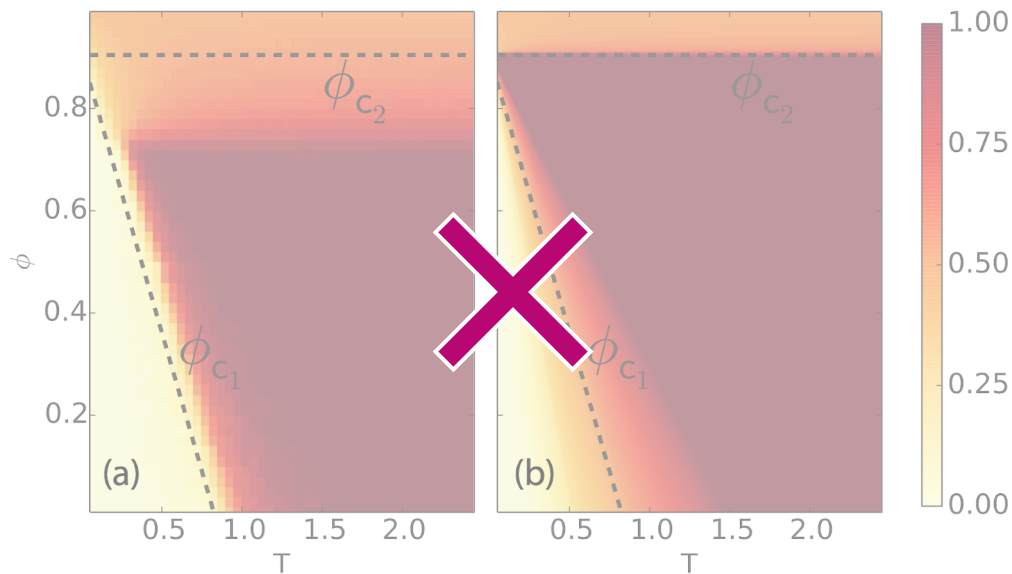


Figure 23: Mean proportion of nodes that are mitigators for network model (a) and ODE model (b). ϕ is the rewiring probability and T is the time between social interactions. ϕ_{c_1} is the lower threshold and ϕ_{c_2} is the upper threshold, above which a fragmentation regime occurs. From (Wiedermann et al., 2015)

452

453 Social networks are rarely static and their ability to evolve over time is represented in adaptive network models
 454 where agents can break existing social links and create new ones, a process called “rewiring”. Often this rewiring is
 455 homophilic, meaning that agents are more likely to create a new social connection with others who share a similar
 456 behavior. Common adaptive network CHES models have nodes representing ~~a renewable resources stock~~ **renewable**
 457 **resource stocks** with an associated extraction level which can adopt a high extraction or low extraction level through
 458 imitating neighbors. These models show that the level of homophilic rewiring can trigger regime shifts at both low
 459 and high levels, where intermediate ranges correspond to a sustainable equilibrium. As agents can either choose to
 460 rewire or imitate their neighbor, a low level of rewiring corresponds to a high speed of social interaction, which as
 461 discussed in Section 2.5 can lead to detrimental tipping points. On the other hand, although high-rewiring leads to
 462 slower social learning, it also brings about a fragmentation regime where social dynamics are dominated by
 463 homophily and the network fragments into components based on strategy type, which makes widespread mitigation
 464 infeasible (Barfuss et al., 2017; Geier et al., 2019; Wiedermann et al., 2015) (Figure 23). CHES models with social
 465 networks are still relatively new and lack diversity in how they are formulated. For example, regarding the tipping
 466 points related to rewiring social links, the lower threshold may be caused by increased social learning since in all

467 models agents can either rewire or imitate, but not both. There is still much to learn through isolating the effect of
468 rewiring as well as exploring a wide array of different model formulations of CHES on social networks.

469 **3 Identifying early warning signals in the of tipping events in CHES**

470 Although dynamical models can offer qualitative insight into potential trajectories of CHES resulting from specific
471 interventions, it is more difficult to use them to generate precise and reliable predictions. Given the potential for
472 severe environmental tipping points in the coming decades, it is extremely useful to be able to predict these abrupt
473 shifts without complete mechanistic knowledge of the system. The ability to predict tipping ~~points~~events with
474 limited data can allow policymakers to have more time preparing for future disasters, and given enough warning and
475 political will, an opportunity to avoid them or mitigate their severity. Rapidly growing research in early warning
476 signals (EWS) offers tools to monitor empirical time series data and warn of future tipping ~~points~~events that are
477 likely to occur (Bury et al., 2021; Dakos et al., 2012, 2015, 2008; Kéfi et al., 2014; Lapeyrolerie and Boettiger,
478 2021). Although much of the work has been conducted on synthetic data, there are many studies that successfully
479 predict historical tipping ~~points~~events in both empirical human and environmental time series data such as the 1987
480 Black Monday financial crash (Diks et al., 2019) as well as abrupt temperature shifts from paleoclimate datasets
481 (Dakos et al., 2008).

482 **3.1 Recent advances for detecting early warning signals**

483 Much research has been done in the past few decades to develop tools for EWS using both empirical and synthetic
484 time series data (Bury et al., 2021; Dakos et al., 2012, 2015, 2008; Kéfi et al., 2014; Lapeyrolerie and Boettiger,
485 2021). Originally motivated by critical slowing down in bifurcation theory, where systems approaching a tipping
486 point show a slower recovery to equilibrium under perturbations, generic EWS measure trends in this “slowing
487 down” (Scheffer et al., 2009). The most commonly used methods compute the lag-1 autocorrelation and variance of
488 the residuals from detrended time series data. Other widely used methods involve metrics such as skewness,
489 measuring the asymmetry of fluctuations over time, and kurtosis, representing the likelihood of extreme values in
490 the time series data. A phenomenon known as flickering occurs when there is sufficient noise to rapidly force the
491 system between alternate stable states. In these cases, an increase in skewness and kurtosis is observed (Dakos et al.,
492 2012). As lag-1 autocorrelation does not account for correlation beyond a single time step, power spectrum analysis
493 has been used to look at changes in complete spectral properties, finding higher variations at low frequencies to
494 commonly occur before a tipping point (Dakos et al., 2012; Scheffer et al., 2009). In spatial systems, many EWS are
495 similar to those used in well-mixed systems, while also accounting for spatial variability. For example, Moran’s I is
496 a spatial analog of lag-1 autocorrelation, which measures the correlation between neighboring nodes in a network
497 (Kéfi et al., 2014).

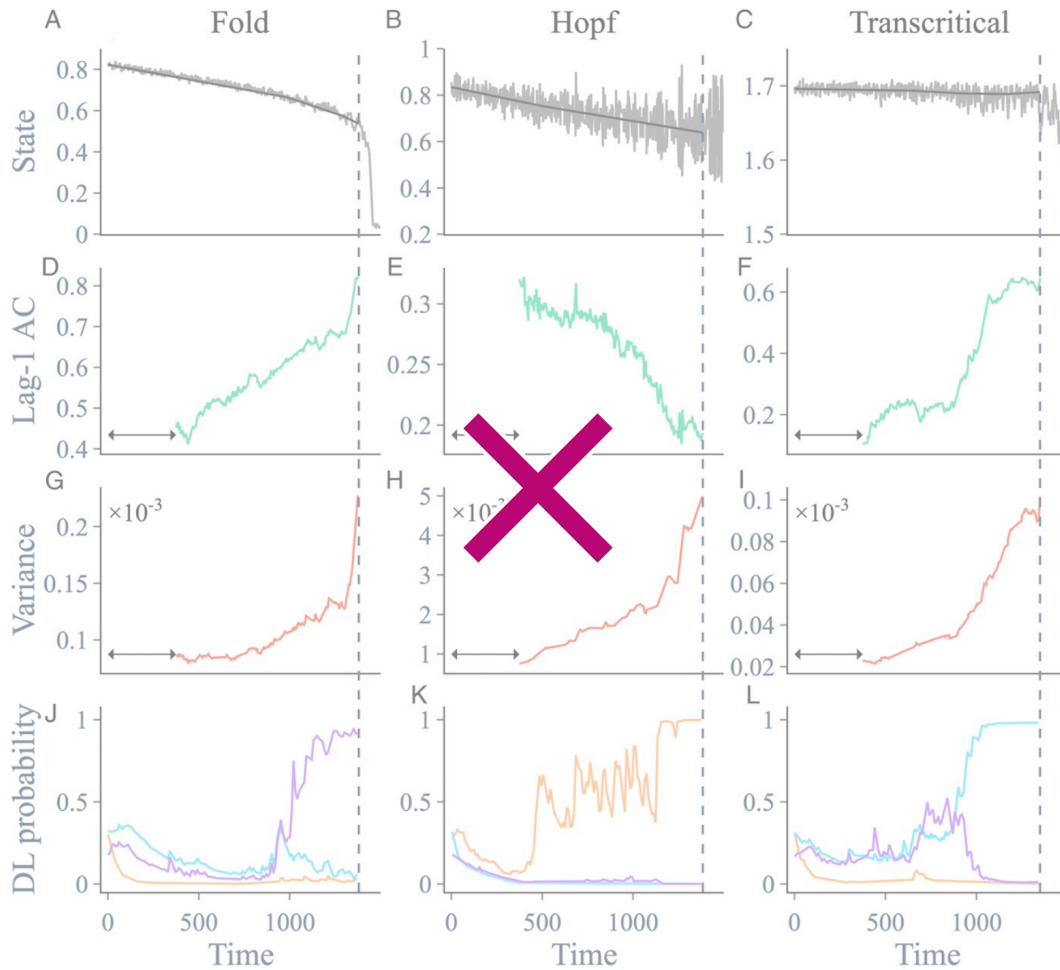


Figure 34: Generic EWS (second and third row) as well as deep learning EWS (bottom row) for time series generated by two ecological models exhibiting different types of bifurcations (top row); fold (left), Hopf (middle), and transcritical (right). As well as being more reliable, deep learning EWS can also distinguish between the type of bifurcation being approached. In the bottom row, the DL algorithm gives probabilities for the occurrence of fold (purple), Hopf (orange), or transcritical (blue) bifurcations. Image taken from (Bury et al., 2021).

498 Numerous spatial ecological systems exhibit patterns in of patchiness preceding a tipping point. For example, in
 499 drylands, spotted vegetation patterns are hypothesised hypothesized to be an EWS for the system approaching
 500 desertification (Kéfi et al., 2014). Coupled human-epidemiological models also show that spatial properties in the
 501 distribution of opinions on a social network offer potential EWS for the onset of disease outbreaks. Approaching this
 502 regime shift, the number of anti-vaccine clusters increases, and very close to the transition point, these communities
 503 coalesce into larger groups (Jentsch et al., 2018; Phillips et al., 2020). These clusters are quantified using a number
 504 of metrics, such as an increase in modularity as well as the mean number, size, and maximum size of communities

505 and pro-vaccine echo chambers (Phillips and Bauch, 2021). This is also in agreement with previous work done in
506 percolation theory showing that phase transitions follow a breakup of connected components on the network
507 (Newman, 2010).

508

509 One downside to the generic metrics discussed above is that they have the potential to fail in the presence of large
510 amounts of noise where transitions can occur far from their analytically derived tipping point. A technique called
511 dynamical network markers increases the dimensionality of the time series by transforming it from state variables to
512 probability distributions of the mean and variance over a given window of time. This reduces the magnitude of noise
513 in each dimension and in approaching a tipping point, one dominant group of variables will show a drastic increase
514 in variance and correlation between other variables within that group. At the same time, the correlation between one
515 variable in this dominant group and others outside the group will decrease. This technique has shown success with
516 empirical data, such as predicting critical transitions in time series data for a eutrophic lake as well as the bankruptcy
517 of Lehman Brothers (Liu et al., 2015), and flu outbreaks (Chen et al., 2019). Dynamical network markers have also
518 been used on spatial systems such as those occurring on social networks through the use of hierarchical network
519 representations. Here, networks are transformed into binary trees where leaves are the nodes from the original
520 network and branches group nodes together at multiple resolutions. Through this hierarchical model, dynamical
521 network markers use these multi-scale communities as the groups of variables that are analysedanalyzed (Li et al.,
522 2023). This spatial technique offers a novel method for predicting tipping pointsevents for CHES using human data
523 occurring on complex social networks.

524

525 A very recent addition to the EWS toolkit uses concepts from statistical physics such as average flux, entropy
526 production, generalisedgeneralized free energy, and time irreversibility to predict tipping points in a shallow lake
527 model much earlier than generic methods such as autocorrelation and variance, showing promise for use in real-time
528 monitoring (Xu et al., 2023). Additionally, the field of machine learning has motivated data-driven approaches to
529 EWS which do not explicitly make use of any statistical metrics in the time series data. Instead, deep learning
530 algorithms are trained on large synthetic datasets using models that have and have not approached tipping points. In
531 the majority of cases, these algorithms have performed significantly better at predicting tipping pointsevents than
532 generic EWS indicators when tested on empirical datasets that exhibit abrupt transitions (Bury et al., 2021; Deb et
533 al., 2022) (Figure 34). Deep learning algorithms are also able to distinguish between different types of bifurcations
534 as they are being approached which can offer vital information regarding the potential for catastrophic collapse in
535 CHES.

536 3.2 Social data for early warning signals

537 In CHES models, the strength of EWS from environmental data has been shown to be muted compared to EWS
538 from environmental systems not coupled to a human system (Bauch et al., 2016) or the same system with weak

539 coupling between the human and environmental subsystems (Richter and Dakos, 2015). This is likely due to the
540 effects of human behavior acting to mitigate variability in the environmental system, for example, rarity-motivated
541 valuation creates a negative feedback loop where incentives to mitigate increase as the environment becomes further
542 depleted, serving as a mechanism to avoid collapse. The muting of EWS provides a unique challenge for monitoring
543 tipping ~~points~~ events in CHES using environmental data, especially as they occur more frequently in these coupled
544 systems as discussed in Section 2. There are a small number of studies that have directly compared the strength and
545 efficacy of EWS between various state or auxiliary variables in CHES models. In these studies, generic EWS from
546 data in the human system were shown to be the only reliable indicators of the coupled system approaching a tipping
547 point. Examples of human data used include the fraction of conservationists in a forest cover model (Bauch et al.,
548 2016), average profits by resource harvesters, and catch per unit effort common-pool resource models (Lade et al.,
549 2013; Richter and Dakos, 2015). In agreement with generic methods, a state-of-the-art machine learning algorithm
550 for EWS showed higher success in detecting tipping ~~points~~ events generated from a coupled epidemiological model
551 using pro-vaccine opinion in the human system compared to total infectious in the epidemiological system (Bury et
552 al., 2021). It is possible that the state variable most sensitive to the forcing parameter may exhibit the strongest
553 EWS, as seen in experimental work on tipping points in a lake food web. In this system, data from the species that
554 had a direct trophic linkage to a driver of the tipping ~~point~~ event (predators added to the food web) exhibited EWS
555 earlier than those that were farther removed from the driver (Carpenter et al., 2014). If this is the case, human
556 drivers of tipping points would most directly affect the human system, and EWS should still be stronger using social
557 data.

558

559 The improved reliability of EWS from social data demonstrated through CHES models shows a significant promise
560 for monitoring resilience in CHES through the analysis of socio-economic data (Box 1.4). This confers a practical
561 advantage as socio-economic data ~~is often more frequently collected and readily available than environmental data~~
562 ~~(Hicks et al., 2016)~~ availability is growing faster than ecological data (and perhaps even environmental data despite
563 the growth of publicly available satellite data) on account of the era of digital social data (Ghermandi and Sinclair,
564 2019; Hicks et al., 2016; Lopez et al., 2019; Salathé et al., 2012). Some examples of this are monitoring profits tied
565 to resource extraction as well as using sentiment analysis on social media data, such as the number of tweets in a
566 given area raising concern over the health of a coupled environmental system. Furthermore, citizen science not only
567 generates environmental data but also provides social metadata through the participation of users who monitor
568 specific areas. Leveraging existing platforms like CitSci.org, we can use this data to estimate trends in
569 conservationist frequency over time (Wang et al., 2015). This approach allows for the implementation of real-time
570 monitoring of environmental systems using data that is currently being generated, reducing the need for extensive
571 knowledge or complex mechanistic models of the system. With the potential social data offers for use with EWS, it
572 is important to note that much of the traditional social data, often conducted through national or regional surveys, do
573 not provide fine-grained spatial or temporal resolution. On the other hand, novel methods that use social media data

574 can solve the resolution issue, but may not accurately represent the population it is being used to model (Hargittai,
575 2020). These challenges may be addressed through a hybridcompound approach that uses hybrid time series
576 generated from multiple types and sources of social data (Rosales Sánchez et al., 2017).

577 4 Conclusion and future directions

578 4.1 Summary of main points

579 ¶

580 From a wide range of examined theoretical models, we are able to gain insight into human drivers that lead to
581 tipping pointsevents in CHES systems. Many social interventions, such as reducing mitigation costs and extractive
582 effort, or increasing the time horizon in decision-making, lead to beneficial tipping pointsevents, regardless of the
583 system modelledmodeled. The beneficial effect of these interventions is intuitive; however, non-linear responses
584 manifested as tipping pointsevents may not be as evident. Mitigation costs can be reduced through subsidies for land
585 preservation and green technology, and extraction effort through limits on land development and the expansion of
586 protected natural areas (i.e. the Haudenosaunee-led protection of the Haldimand Tract) (Forester, 2021), and by
587 increasing time horizons through passing long-term legislation that centers the well-being of human and
588 environmental systems such as the Green New Deal (Galvin and Healy, 2020). These policy interventions become
589 more difficult to implement at large scales, and models that are tailored to global coordination problems can give us
590 insight into how institutions can work together to rapidly mitigate looming threats, such as the current climate crises
591 we are facing (Karatayev et al., 2021).

592

593 Other human behaviors and social processes are much more nuanced and system-specific in how they affect tipping
594 points. For example, models show that rarity-motivated valuation can act to detrimentally tip the environmental
595 system into a depleted state when it crosses both an uppera lower and (counterintuitively) a loweran upper threshold
596 value. This was illustrated most clearly in the example of forest cover in the paper by Bauch et al. (2016). Social
597 norms, especially when majority-enforcing, increase the likelihood of tipping points through the emergence of
598 bistable regimes that are made up of both sustainable and unsustainable environmental equilibria. The extent of
599 coupling between the human and environmental system as well as the speed of social change relative to
600 environmental change can have different effects depending on whether the model is input-or
601 output-limitedhuman-extraction or human-emission. Interventions related to human valuation and social norms are
602 much more difficult to implement as they require a deeper mechanistic understanding of how to influence social
603 dynamics and may also have ethical considerations.

604

605 The models we reviewed also show that greater structural complexity via the number and diversity of human traits
606 as well as the number of social connections can increase the potential for tipping points and mask social dynamics

607 making these transitions much harder to predict. The ~~modelling~~ modeling literature has only explored a small sliver
608 of the space of possible choices regarding assumed social structure and the types of environmental models coupled
609 to them. For example, the vast majority of models only allow for a binary choice in human behavior and adaptive
610 social networks have only recently been incorporated, with limited mechanisms of re-wiring and types of coupled
611 environmental systems. Consequently, we still have much to learn on how shifting underlying social structures acts
612 as a driver of tipping ~~points~~ events. This is especially true in ~~output-limited~~ human-emission models which are
613 important to improving our understanding of how our social structures affect pressing global issues such as pollution
614 and climate change. Even if we include more diverse and realistic social structures and processes, CHES are
615 composed of many non-linear feedbacks and contain high levels of uncertainty, and the reality is that we may not be
616 able to have a complete mechanistic representation through models. EWS from empirical data show great potential
617 in predicting tipping ~~points~~ events without requiring a full understanding of the system being monitored. There have
618 been many advances in using state-of-the-art machine learning algorithms to provide accurate EWS from 1-D time
619 series (Bury et al., 2021; Deb et al., 2022), and very recent work is now developing similar techniques to predict
620 tipping ~~points~~ events from spatial data (Dylewsky et al., 2022). As synthetic data from models have shown the value
621 of EWS from social data, it is likely that applying these techniques to diverse and hybrid empirical social datasets
622 can vastly improve our ability to predict tipping ~~points~~ events caused by human drivers in the future.

623 4.2 Future work in CHES ~~modelling~~ modeling

624 There are many social phenomena that are not commonly included in CHES models, yet may be important in
625 furthering our understanding of tipping points within these systems. We know that inequality in human systems
626 plays a large role in individuals' risk perception and ability to engage in pro-environmental behavior (Gibson-Wood
627 and Wakefield, 2013; Pearson et al., 2017; Quimby and Angeli, 2011; Rajapaksa et al., 2018) and have
628 mentioned two CHES models that incorporate wealth inequality in a human-climate system (Menard et al., 2021;
629 Vasconcelos et al., 2014). However, more studies explicitly investigating the role of inequality could offer some
630 valuable insight into interventions that can be more effective in benefiting both the environment and the most
631 vulnerable in human systems. This could be complemented by ~~incorporating~~ social biases where perceptions of risk
632 are linked to an individual's socio-economic status, and detrimental environmental outcomes are experienced
633 disproportionately by vulnerable communities as is commonly observed globally (Banzhaf et al., 2019; Boyce, 2007).
634 Future models could allow for alternatives to the common ~~modelling~~ modeling assumption where individuals act in
635 their own self-interest, for example by incorporating other-regarding preferences into utility functions so that
636 individuals value their neighbors' well-being along with their own (Dimick et al., 2018). These models could also
637 look at grassroots redistribution of wealth allowing us to explore the effects of alternative social value systems on
638 the environment (Tilman et al., 2018).

639

640 Stochasticity (noise), especially regarding drivers of tipping points can significantly affect system dynamics
641 including when tipping points occur. Although many CHES models are deterministic, recent work has shown that
642 increasing noise can lead to earlier tipping (Willcock et al., 2023), or in other cases, increase the duration of time the
643 environmental system can persist before becoming extinct (Jnawali et al., 2022). These contradictory results warrant
644 further work in understanding how different types of noise and their magnitude within drivers of tipping
645 ~~points~~events affect the resilience of these systems. With stochasticity comes uncertainty, and in real-world systems,
646 it is impossible to know with precision the extent of social change required to bring about a beneficial or avoid a
647 detrimental tipping point. This uncertainty around our knowledge of system thresholds adds an additional challenge
648 in both agreeing upon and following through with policy that promotes sustainable futures while taking into account
649 potential tipping points. Experimental games have shown that high threshold uncertainty can promote the collapse of
650 a shared resource, often through an increase in free-riding behavior (Barrett and Dannenberg, 2014, 2012). On the
651 other hand, field experiments in fishing communities have shown that high uncertainty can promote cooperation and
652 sustainable resource use (Finkbeiner et al., 2018; Rocha et al., 2020). Theoretical models show that increased
653 uncertainty can lead to increased mitigative behavior if the shared resource is highly valued; however, for
654 low-valued resources, increased uncertainty can deter mitigation, putting the persistence of the shared resource at
655 risk (Jager et al., 2000; McBride, 2006). Uncertainty around thresholds is unavoidable, further motivating the need
656 to offer additional incentives for mitigative action on institutional scales, rather than solely the threat of
657 environmental collapse. In systems where uncertainty can promote mitigative action, increased communication and
658 awareness campaigns around this threshold uncertainty could be useful to incorporate into policy.
659

660 This review has focused primarily on the effects of single drivers; however, research on multiple co-occurring
661 human drivers of tipping ~~points~~events, while more analytically challenging, could offer a holistic understanding of
662 how these drivers interact. A recent study has shown that multiple drivers can both reduce the time until tipping or
663 lead to a tipping point that would not occur with a single driver (Willcock et al., 2023) and there is already a large
664 body of empirical work exploring the diversity of these drivers which can be used to inform future CHES models
665 (Jaureguiberry et al., 2022; Maciejewski et al., 2019; Millennium Ecosystem Assessment, 2005). Finally, as the
666 majority of the studies in ~~modelling~~modeling tipping points have focused on slow gradual changes in the driver, fast
667 changes require further research as they can exhibit very different tipping behavior (Ashwin et al., 2012). CHES
668 models ubiquitously exemplify the phenomenon of tipping points, which often occur through drivers in the human
669 system. Although these models offer valuable insight in understanding key feedbacks and qualitative behavior, their
670 predictive power is limited. Additionally, as many model findings can depend on the type of system
671 ~~modelled~~modeled as well as assumptions in the model formulation, translating this work into policy remains a
672 significant challenge. However, further work in both diversifying model systems and assumptions paired with
673 research in universal real-time indicators of EWS shows considerable promise in both improving our understanding
674 and predicting human drivers of tipping ~~points~~events in the environment.

675

676 **Author contribution.** I.F.: visualization, writing—original draft, writing—review and editing; C.T.B.: visualization,
677 writing—original draft, writing—review and editing; M.A.: conceptualization, funding acquisition, supervision,
678 visualization, writing—original draft, writing—review and editing.

679

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681

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686 **Appendix**

Authors	Year	Title	System of study
Sethi & Somanathan	1996	The evolution of social norms in common property resource use	Common pool resource
Satake et al.	2007	Coupled ecological–social dynamics in a forested landscape: Spatial interactions and information flow	Land use
Iwasa et al.	2007	Nonlinear behavior of the socio-economic dynamics for lake eutrophication control	Lake eutrophication
Suzuki & Iwasa	2009	The coupled dynamics of human socio-economic choice and lake water system: the interaction of two sources of nonlinearity	Lake eutrophication
Iwasa et al.	2010	Paradox of nutrient removal in coupled socioeconomic and ecological dynamics for lake water pollution	Lake eutrophication
Figueiredo & Pereira	2011	Regime shifts in a socio-ecological model of farmland abandonment	Land use
Tavoni et al.	2012	The survival of the conformist: Social pressure and renewable resource management	Common pool resource
Lade et al.	2013	Regime shifts in a social-ecological system	Common pool resource
Iwasa & Lee	2013	Graduated punishment is efficient in resource management if people are heterogeneous	Fishery
Richter et al.	2013	Contagious cooperation, temptation, and ecosystem collapse	Common pool resource
Richter & Grasman	2013	The transmission of sustainable harvesting norms when agents are conditionally cooperative	Common pool resource
Barlow et al.	2014	Modelling interactions between forest pest invasions and human decisions regarding firewood transport restrictions	Pest
Vasconcelos et al.	2014	Climate policies under wealth inequality	Climate
Ali et al.	2015	Coupled human-environment dynamics of forest pest spread and control in a multipatch, stochastic setting	Pest
Sugiarto et al.	2015	Socioecological regime shifts in the setting of complex social interactions	Common pool resource
Wiedermann et al.	2015	Macroscopic description of complex adaptive networks coevolving with dynamic node states	Private resource
Richter & Dakos	2015	Profit fluctuations signal eroding resilience of natural resources	Common pool resource
Schlüter et al.	2016	Robustness of norm-driven cooperation in the commons	Common pool

			resource
Weitz et al.	2016	An oscillating tragedy of the commons in replicator dynamics with game-environment feedback	Common pool resource
Bauch et al.	2016	Early warning signals of regime shifts in coupled human-environment systems	Forest
Henderson et al.	2016	Alternative stable states and the sustainability of forests, grasslands, and agriculture	Land use
Sugiarto et al.	2017	Social cooperation and disharmony in communities mediated through common pool resource exploitation	Common pool resource
Barfuss et al.	2017	Sustainable use of renewable resources in a stylized social-ecological network model under heterogeneous resource distribution	Private resource
Lafuite et al.	2017	Delayed behavioral shifts undermine the sustainability of social-ecological systems	Land use
Lindkvist et al.	2017	Strategies for sustainable management of renewable resources during environmental change	Common pool resource
Osten et al.	2017	Sustainability is possible despite greed - Exploring the nexus between profitability and sustainability in common pool resource systems	Common pool resource
Sigdel et al.	2017	Competition between injunctive social norms and conservation priorities gives rise to complex dynamics in a model of forest growth and opinion dynamics	Forest
Sugiarto et al.	2017	Emergence of cooperation in a coupled socioecological system through a direct or an indirect social control mechanism	Common pool resource
Thampi et al.	2018	Socio-ecological dynamics of Caribbean coral reef ecosystems and conservation opinion propagation	Coral reef
Chen & Szolnoki	2018	Punishment and inspection for governing the commons in a feedback-evolving game	Common pool resource
Drechsler & Surun	2018	Land-use and species tipping points in a coupled ecological-economic model	Land use
Geier et al.	2019	The physics of governance networks: critical transitions in contagion dynamics on multilayer adaptive networks with application to the sustainable use of renewable resources	Private resource
Hauert et al.	2019	Asymmetric evolutionary games with environmental feedback	Common pool resource
Lin & Weitz	2019	Spatial interactions and oscillatory tragedies of the commons	Common pool resource
Sigdel et al.	2019	Convergence of socio-ecological dynamics in disparate ecological systems under strong coupling to human social systems	Common pool resource

Bury et al.	2019	Charting pathways to climate change mitigation in acoupled socio-climate model	Climate
Shao et al.	2019	Evolutionary dynamics of group cooperation with asymmetrical environmental feedback	Common pool resource
Barfuss et al.	2020	Caring for the future can turn tragedy into comedy for long-term collective action under risk of collapse	Common pool resource
Tilman et al.	2020	Evolutionary games with environmental feedbacks	Common pool resource
Muneepeerakul & Anderies	2020	The emergence and resilience of self-organized governance in coupled infrastructure systems	Water use
Sun & Hilker	2020	Analyzing the mutual feedbacks between lake pollution and human behavior in a mathematical social-ecological model	Lake eutrophication
Mathias et al.	2020	Exploring non-linear transition pathways in social-ecological systems	Common pool resource
Phillips et. al	2020	Spatial early warning signals of social and epidemiological tipping points in a coupled behavior-disease network	Epidemic
Menard et al.	2021	When conflicts get heated, so does the planet: coupled social-climate dynamics under inequality	Climate
Phillips & Bauch	2021	Network structural metrics as early warning signals of widespread vaccine refusal in social-epidemiological networks	Epidemic
Holstein et al.	2021	Optimization of coupling and global collapse in diffusively coupled socio-ecological resource exploitation networks	Private resource
Farahbakhsh et al.	2021	Best response dynamics improve sustainability and equity outcomes in common-pool resources problems, compared to imitation dynamics	Common pool resource
Yan et al.	2021	Cooperator driven oscillation in a time-delayed feedback-evolving game	Common pool resource
Müller et al.	2021	Anticipation-induced social tipping: can the environment be stabilised by social dynamics?	Climate
Milne et al.	2021	Local overfishing patterns have regional effects on health of coral, and economic transitions can promote its recovery	Coral reef
Moore et al.	2022	Determinants of emissions pathways in the coupled climate–social system	Climate
Bengochea Paz et al.	2022	Habitat percolation transition undermines sustainability in socialecological agricultural systems	Land use

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