

Phase state and viscosity of secondary organic aerosols over China simulated by WRF-Chem

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Abstract. Secondary organic aerosols (SOA) can exist in liquid, semi-solid or amorphous solid states. ~~Chemical transport models (CTMs), however, usually assume that SOA particles are homogeneous and well-mixed liquids, with rapid establishment of gas-particle equilibrium for simulations of SOA formation and partitioning, which are rarely accounted for in current chemical transport models (CTMs).~~ Missing the information of SOA phase state and viscosity in CTMs impedes accurate representation of SOA formation and evolution, affecting the predictions of aerosol effects on air quality and climate. We have previously developed a parameterization to estimate the glass transition temperature (T_g) of an organic compound based on volatility and to predict viscosity of SOA. In this study, we apply this method to predict the phase state of SOA particles over China in summer of 2018 using the Weather Research and Forecasting model coupled to Chemistry (WRF-Chem). ~~This is the first time that spatial distributions of the SOA phase state over China are investigated by a regional CTM. The simulated T_g of dry SOA ($T_{g,org}$) agrees well with the value estimated from ambient volatility measurements at an urban site in Beijing. For the spatial distributions of $T_{g,org}$.~~ Simulations show that ~~at the surface, T_g the values of $T_{g,org}$ dry SOA~~ range from ~287 K to 305 K, with higher values in the northwestern China where SOA particles have larger mass fractions of low volatility compounds. Considering water uptake by SOA particles, the SOA viscosity shows a prominent geospatial gradient that highly viscous or solid SOA particles are mainly predicted in the northwestern China. The lowest and highest SOA viscosity values both occur over the Qinghai-Tibet Plateau that the solid phase state is predicted over dry and high-altitude areas and the liquid phase state is predicted mainly in the south of the plateau with high relative humidity during the summer monsoon season. ~~We also calculate the characteristic mixing timescale of organic molecules in 200 nm SOA particles to evaluate kinetic limitations in SOA partitioning. Calculations show that during the simulated period the percent time of the mixing timescale longer than 1 h is > 70 % at the surface and at 500 hPa in most areas of the northern~~

33 ~~China, indicating that kinetic partitioning considering the bulk diffusion in viscous particles may be required for more~~
34 ~~accurate prediction of SOA mass concentrations and size distributions over these areas.~~ Sensitivity simulations show that
35 including the formation of extremely low-volatile organic compounds, the percent time that a SOA particle is in the liquid
36 phase state decreases by up to 12 % in the southeastern China during the simulated period. With an assumption that the
37 organic and inorganic compounds are internally mixed in one phase, we show that the water absorbed by inorganic species
38 can significantly lower the simulated viscosity over the southeastern China. This indicates that constraining the uncertainties
39 in simulated SOA volatility distributions and the mixing state of the organic and inorganic compounds would improve
40 prediction of viscosity in multicomponent particles in southeastern China. We also calculate the characteristic mixing
41 timescale of organic molecules in 200 nm SOA particles to evaluate kinetic limitations in SOA partitioning. Calculations
42 show that during the simulated period the percent time of the mixing timescale longer than 1 h is > 70 % at the surface and at
43 500 hPa in most areas of the northern China, indicating that kinetic partitioning considering the bulk diffusion in viscous
44 particles may be required for more accurate prediction of SOA mass concentrations and size distributions over these areas.

45 **1 Introduction**

46 Secondary organic aerosols (SOA) are major components of atmospheric fine particles, impacting air quality, climate and
47 public health (Jimenez et al., 2009; Pöschl and Shiraiwa, 2015). The formation and evolution of SOA involve both chemical
48 reactions and mass transport in the gas and particle phases (Ziemann and Atkinson, 2012). This complexity makes accurate
49 representation of SOA evolution in chemical transport models (CTMs) challenging, leading to a large uncertainty in
50 evaluating SOA impacts on air quality and climate (Kanakidou et al., 2005; Shrivastava et al., 2017).

51 Current CTMs usually assume that SOA particles are homogeneous and well-mixed liquids, with rapid establishment of
52 gas-particle equilibrium applied in simulations of SOA formation and partitioning (Pankow, 1994; Donahue et al., 2006). It
53 has been shown that SOA can exist in liquid (low dynamic viscosity η , $\eta < 10^2$ Pa s), semi-solid (10^2 Pa s $\leq \eta \leq 10^{12}$ Pa s) or
54 solid (amorphous or glassy solid; $\eta > 10^{12}$ Pa s) states, depending on particle chemical composition and atmospheric
55 conditions, such as ambient temperature (T) and relative humidity (RH) (Koop et al., 2011; Reid et al., 2018). Viscosities can
56 be converted to bulk diffusion coefficients via the Stokes-Einstein equation (Einstein, 1905; Seinfeld and Pandis, 2016) or
57 the fractional Stokes-Einstein equation (Price et al., 2016; Evoy et al., 2019; Evoy et al., 2020). The phase state, viscosity,
58 and bulk diffusivity of SOA are important in many aerosol processes. The semi-solid or solid phase state can prolong the
59 equilibration timescales in the gas-particle partitioning, indicating a need of considering kinetic limitations in simulating the
60 SOA partitioning into highly viscous particles (Shiraiwa and Seinfeld, 2012; Roldin et al., 2014; Zaveri et al., 2014; Li and
61 Shiraiwa, 2019). The viscosity of SOA can impact the rates of heterogeneous and multiphase reactions (Marshall et al., 2018;
62 Zhang et al., 2019a), photochemistry (Liu et al., 2018; Dalton and Nizkorodov, 2021; Baboosian et al., 2022), and the

63 uptake of gaseous pollutants (e.g., O₃, OH, N₂O₅, NO₂, NH₃, and SO₂) and water vapor (Abbatt et al., 2012; Kuwata and
64 Martin, 2012; Preston and Zuend, 2022), with implications for accurate predictions of atmospheric chemical composition
65 (Reid et al., 2018). The SOA phase state also affects particle size distribution evolution (Shiraiwa et al., 2013; Zaveri et al.,
66 2022) and ice nucleation pathways (Knopf and Alpert, 2023).

67 Accurate predictions of the viscosity need the information of molecular structures and functional groups (Song et al.,
68 2016; Rothfuss and Petters, 2017; Gervasi et al., 2020; Galeazzo and Shiraiwa, 2022); however, molecular specificity is
69 often unavailable in ambient measurements, leading to the prediction of the phase state of ambient SOA particles difficult.
70 Currently there are only a few methods developed to predict the phase state of ambient SOA particles, and successfully be
71 implemented in CTMs. Li, Shiraiwa and coauthors first developed a parameterization predicting the glass transition
72 temperature (T_g) based on the molar mass (M) and the atomic O/C ratio for carbon-hydrogen (CH) and
73 carbon-hydrogen-oxygen (CHO) compounds with their molar mass less than 450 g mol⁻¹ (Shiraiwa et al., 2017). T_g
74 characterizes the temperature at which a phase transition between amorphous solid and semi-solid states occurs (Koop et al.,
75 2011). When the ambient T is higher than T_g , a SOA particle is in a semi-solid or liquid phase state; otherwise, it behaves as
76 an amorphous solid. This parameterization has been successfully coupled into CTMs simulating the SOA phase state over
77 the globe (Shiraiwa et al., 2017) or the U.S. (Schmedding et al., 2020; Li et al., 2021b), showing that semi-solid and
78 amorphous solid phase states frequently occurred over dry lands and in the upper troposphere. Further parameterizations
79 were developed to predict T_g as a function of the saturation mass concentration (C^0) and the O/C ratio of organic compounds,
80 or as a function of C^0 solely, which indirectly included the effect of molecular structure on T_g estimations (Li et al., 2020).
81 This parameterization can be used in the volatility basis set (VBS) framework (Donahue et al., 2006), which is widely
82 adopted in CMTs simulating SOA formation (Lane et al., 2008a; Knote et al., 2015). Rasool et al. (2021) then coupled this
83 new method (Li et al., 2020) into the Weather Research and Forecasting Model coupled to chemistry (WRF-Chem) (Grell et
84 al., 2005; Fast et al., 2006), and the simulations showed that the viscosity of SOA particles could be reasonably predicted
85 during the dry-to-wet transition season in the Amazon rainforest. Li et al. (2020) was also applied in the WRF-Chem
86 simulating the effects of particle phase state on the multiphase chemistry of SOA formation in the Amazon rainforest
87 (Shrivastava et al., 2022; Rasool et al., 2023). Instead of predicting T_g , Maclean et al. (2021) developed parameterizations for
88 viscosity as a function of T and RH based on measured viscosity data of laboratory SOA, and applied the viscosity
89 parameterizations in CTMs to predict the mixing timescales of organic molecules and water molecules within SOA particles
90 on a global scale (Maclean et al., 2017; Maclean et al., 2021). It is needed to conduct more simulations to investigate the
91 SOA phase state varied with locations and the time. Simulations of the SOA phase state in China on a regional scale have not
92 been available.

93 Investigations in the particle phase state over China are currently focused on field observations and laboratory
94 experiments. Bounce factor measurements showed that submicrometer particles can be semi-solid in clear days and liquid in

95 hazy days in Beijing, China (Liu et al., 2017). The phase state of PM_{2.5} (particulate matter with an aerodynamic diameter ≤
96 2.5 μm) was found to be mostly semisolid to solid in winter Beijing based on the measurements using optical microscopy
97 combined with the poke-and-flow technique (Song et al., 2022). The RH-dependent viscosity of the proxies of actual
98 ambient particles in Beijing was also investigated based on dual optical tweezers (Tong et al., 2022). The phase state of
99 submicrometer particles in Beijing was retrieved from a polarization lidar that has the potential to infer the vertical profiles
100 of phase state (Tan et al., 2020). The phase state of traffic-related secondary aerosols in Beijing may have a distinguished
101 diurnal variation (Meng et al., 2021). The biomass burning aerosols, collected near a farmland in Yangtze River Delta, China,
102 were found to exist in the non-solid phase state at relatively dry conditions (Liu et al., 2021).

103 These measurements indicate that the particle phase state over China is highly variable under different atmospheric
104 conditions. It is important to know the spatial distributions and time variations of the SOA phase state and viscosity in
105 multicomponent particles to better quantify the aerosol effects on air quality, which, however, has not been investigated over
106 China with air quality models on a regional scale. Here we use the WRF-Chem model simulated SOA volatility distributions
107 to estimate the glass transition temperature and viscosity of SOA particles over China based on the parameterizations
108 developed in Li et al. (2020). We further calculate the diffusion coefficients and mixing timescales of organic molecules
109 within SOA, which has implications in how to properly treat the SOA partitioning (instantaneous equilibrium vs. kinetic
110 partitioning) in CTMs. As volatility and viscosity are closely related, we conduct a sensitivity calculation to evaluate the
111 effects of the simulated SOA volatility distributions on viscosity estimations. We also conduct a sensitivity calculation to
112 investigate how the water absorbed by inorganic components in PM_{2.5} affects viscosity estimations, which has implications
113 in predicting the viscosity of internally mixed ambient particles.

114 **2 Methods**

115 **2.1 WRF-Chem model configuration**

116 We use the WRF-Chem model version 3.7.1 (Grell et al., 2005; Fast et al., 2006) and simulate the period from 20 May to 23
117 June 2018 with a spin-up period of 7 days (May 13 – 19). We set up two domains (Fig. S1 in the Supplement) with the
118 horizontal resolutions of 81 km and 27 km, respectively, and 18 vertical layers are applied from the surface up to 100 hPa.
119 The meteorological initial and boundary conditions are from the National Centers for Environmental Prediction (NCEP)
120 Global Forecast System (GFS) final (FNL) reanalysis data. The outputs of a global chemical transport model MOZART-4
121 (Emmons et al., 2010) provide initial and boundary conditions of chemical species over the outer domain (Fig. S1 in the
122 Supplement). Anthropogenic emissions are from the MIX 2010 inventory for Asia (Li et al., 2017) and the MEIC 2016
123 inventory for China (<http://meicmodel.org.cn>) (Zheng et al., 2018). Biogenic emissions are calculated from the Model of
124 Emissions of Gases and Aerosols from Nature (MEGAN2.1) (Guenther et al., 2012).

125 The utilized physical and chemical schemes are given in Table S1. We use the MOZART-4 mechanism (Emmons et al.,
 126 2010) for the gas-phase chemistry. The MOSAIC (Model for Simulating Aerosol Interactions and Chemistry) aerosol module
 127 (Zaveri et al., 2008) is applied for the aerosol chemistry and we represent aerosol particles with 4-size sections having dry
 128 diameters ranging from 0.039 μm to 10 μm (Knote et al., 2015). SOA formation is treated with the 1-D volatility basis set
 129 (VBS) approach (Donahue et al., 2006) which has been implemented into the MOSAIC aerosol module (Lane et al., 2008a;
 130 Ahmadov et al., 2012). Five volatility bins are considered (effective saturation mass concentrations C^* of 10^{-4} , 1, 10, 100,
 131 and 1000 $\mu\text{g m}^{-3}$ at 298 K) in the official version 3.7.1 of the WRF-Chem model, with the enthalpy of vaporization (ΔH_{vap})
 132 values of 40, 131, 120, 109, and 98 kJ mol^{-1} used in each volatility bin (Knote et al., 2015). We follow Knote et al. (2015)
 133 with SOA mass yields adopted for four volatility bins (1, 10, 100, and 1000 $\mu\text{g m}^{-3}$). Further gas-phase aging is simulated
 134 through OH oxidation of SOA vapors with a fixed rate of $1.0 \times 10^{-11} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$, with products shifted down one
 135 volatility bin (Murphy and Pandis, 2009), e.g., the condensable vapors with C^* of 1 $\mu\text{g m}^{-3}$ react with OH forming surrogate
 136 species in the lowest volatility bin (C^* of $10^{-4} \mu\text{g m}^{-3}$). The partitioning of organic compounds between the gas and particle
 137 phases is simulated based on Pankow (1994) which is implemented in MOSAIC (Shrivastava et al., 2011). We apply glass
 138 transition temperature and viscosity calculations to WRF-Chem model output for traditional SOA formed from the oxidation
 139 of volatile organic compounds including alkanes, alkenes, aromatics, isoprene, and monoterpenes (Lane et al., 2008b).

140 2.2 Glass transition temperature and viscosity calculations

141 The glass-transition temperature of SOA products in each volatility bin at dry conditions ($T_{g,i}$) is calculated as a function of
 142 the saturation mass concentration at 298 K (C^0) using the parameterization (Eq. 1) developed in our previous study (Li et al.,
 143 2020). We assume ideal thermodynamic mixing that C^0 is equal to C^* , which is often applied in the VBS (Donahue et al.,
 144 2011).

$$T_{g,i} = 288.70 - 15.33 \times \log_{10}(C^0) - 0.33 \times [\log_{10}(C^0)]^2 \quad (1)$$

145 The T_g of mixtures of dry SOA compounds ($T_{g,\text{org}}$) is calculated by the Gordon–Taylor equation (Gordon and Taylor,
 146 1952), with the Gordon–Taylor constant (k_{GT}) assumed to be 1 (Dette et al., 2014):

$$T_{g,\text{org}} = \sum_i \omega_i T_{g,i} \quad (2)$$

147 where ω_i is the mass fraction of SOA products in each volatility bin simulated by the VBS module in WRF-Chem.

148 The particle phase state depends strongly on water content in particles, as water can act as a plasticizer to decrease
 149 viscosity (Mikhailov et al., 2009; Koop et al., 2011). The mass concentration of water absorbed by SOA particles under
 150 humid conditions is estimated using the effective hygroscopicity parameter (κ) (Petters and Kreidenweis, 2007) as:

$$m_{\text{H}_2\text{O}} = \left(\frac{a_w}{1 - a_w} \right) \frac{\kappa \rho_w m_{\text{SOA}}}{\rho_{\text{SOA}}} \quad (3)$$

151 where a_w is water activity calculated as $a_w = \text{RH}/100$ and ρ_w is the density of water. m_{SOA} is the simulated total mass
 152 concentrations of traditional SOA. The density of SOA particles (ρ_{SOA}) is assumed to be 1.5 g cm^{-3} (Knote et al., 2015). κ is
 153 assumed to be 0.1 based on previous studies (Gunthe et al., 2009; Duplissy et al., 2011; Wu et al., 2013) and consistent with
 154 the value used in our previous global SOA phase state simulations (Shiraiwa et al., 2017).

155 T_g of organic-water mixtures is also calculated by the Gordon–Taylor equation (Eq. 4) with k_{GT} suggested to be 2.5
 156 (Koop et al., 2011).

$$T_g(\omega_{\text{org}}) = \frac{(1 - \omega_{\text{org}})T_{g,w} + \frac{1}{k_{\text{GT}}}\omega_{\text{org}}T_{g,\text{org}}}{(1 - \omega_{\text{org}}) + \frac{1}{k_{\text{GT}}}\omega_{\text{org}}} \quad (4)$$

157 where ω_{org} is the mass fraction of the simulated SOA species in organic-water mixtures. The glass transition temperature of
 158 pure water ($T_{g,w}$) is 136 K (Kohl et al., 2005). Based on $T_g(\omega_{\text{org}})$, viscosity can be calculated with the
 159 Vogel–Tammann–Fulcher (VTF) equation (Angell, 1991): $\eta = \eta_{\infty} e^{\frac{T_0 D}{T - T_0}}$, where η_{∞} is the viscosity at infinite temperature
 160 (10^{-5} Pa s , Angell (1991)). D is the fragility parameter which is adopted to be 10 based on our previous study in DeRieux
 161 and Li et al. (2018). T_0 is the Vogel temperature calculated as $T_0 = \frac{39.17 T_g(\omega_{\text{org}})}{D + 39.17}$. We further calculate the bulk diffusion
 162 coefficient (D_b) of organic molecules with a radius of 0.4 nm (Maclean et al., 2021) and water molecules in SOA particles
 163 based on predicted viscosity and the fractional Stokes–Einstein equation (Price et al., 2016; Evoy et al., 2019; Evoy et al.,
 164 2020), which is detailed in the Supplement. The mixing timescales of molecules (τ_{mix}) within SOA particles is calculated as
 165 $\tau_{\text{mix}} = d_p^2 / (4\pi^2 D_b)$ (Seinfeld and Pandis, 2016), where d_p is the particle diameter. The d_p is assumed to be 200 nm
 166 (Maclean et al., 2021) when we calculate τ_{mix} .

167 2.3 Sensitivity simulations

168 Table 1 lists all the performed simulations. In the base case, we update the C^* in the lowest volatility bin from $10^{-4} \mu\text{g m}^{-3}$ in
 169 the official WRF-Chem v3.7.1 to $0.1 \mu\text{g m}^{-3}$ based on the ambient volatility observations (referring to Section 2.4 and Fig. 1),
 170 and calculate the ΔH_{vap} in the lowest volatility bin using the semi-empirical parameterization in Epstein et al. (2010), leading
 171 to a value of 142 kJ mol^{-1} . To evaluate the effects of simulated SOA volatility distributions on phase state estimations, we
 172 conduct a simulation (sensitivity case A) following the default setting in the model assuming that the lowest C^* is $10^{-4} \mu\text{g m}^{-3}$
 173 at 298 K, with ΔH_{vap} of 40 kJ mol^{-1} (Knote et al., 2015). A smaller ΔH_{vap} indicates less dependence of volatility on
 174 temperature variations. In the sensitivity case B, we increase the simulated RH by a factor of 10 % as we find that the
 175 simulated RH values are smaller than the observations (Section 3.1). In the base case and sensitivity cases A and B, we
 176 predict T_g for SOA-water mixtures accounting for the SOA-influenced water uptake solely, assuming that SOA particles are
 177 externally mixed with inorganic compounds such as sulfate and nitrate. In the sensitivity case C, we assume that the organic

178 and inorganic compounds are internally mixed in one phase and include the water absorbed by inorganic compounds in
179 viscosity calculations. The water associated with inorganics is calculated by the MOSAIC module coupled in WRF-Chem.

180 **2.4 Observation**

181 The observation data measured at an urban site in the Institute of Atmospheric Physics (IAP), Chinese Academy of Sciences
182 (39°58'28" N, 116°22'16" E) in Beijing (Fig. S1 in the Supplement) are used to compare with the simulation results of the
183 WRF-Chem model. The aerosol volatility was measured from 20 May to 23 June in 2018, using a thermodenuder coupled
184 with an Aerodyne high-resolution aerosol mass spectrometer (Xu et al., 2019). The volatility distributions of oxygenated
185 organic aerosols (OOA) resolved from positive matrix factorization (PMF) were estimated using a dynamic mass transfer
186 model (Riipinen et al., 2010). The volatility of OOA was found to be distributed in six logarithmically spaced C^* bins
187 including 0.001, 0.01, 0.1, 1, 10, and 100 $\mu\text{g m}^{-3}$, based on the best fits between the measured and predicted thermograms
188 using the methods in Karnezi et al. (2014). Chemical species including organics (Org), sulfate (SO_4^{2-}), nitrate (NO_3^-), and
189 ammonium (NH_4^+) in $\text{PM}_{2.5}$ were measured using an Aerodyne time-of-flight aerosol chemical speciation monitor (Fröhlich
190 et al., 2013) equipped with a capture vaporizer and $\text{PM}_{2.5}$ lens, with the details described in Li et al. (2023). The OOA factor
191 was identified with the PMF analysis. We obtain the mass concentrations of $\text{PM}_{2.5}$ from the Olympic Center observation site
192 (<http://zx.bjmemc.com.cn>) which is ~4 km from the IAP site (Fig. S1 in the Supplement). Meteorological parameters
193 including RH and T are from the Beijing meteorological tower at the IAP site.

194 **3 Results**

195 **3.1 Simulations in Beijing and the comparison with observations**

196 The comprehensive model evaluations were conducted in our previous studies, showing that the WRF-Chem model
197 reasonably captured the magnitudes and spatial distributions of concentrations of major air pollutants over China (Li et al.,
198 2011; Li et al., 2014; Li et al., 2015; Qu et al., 2019; Zhang et al., 2022). Here we focus on the comparison of simulations
199 and observations at the IAP site during 20 May – 23 June 2018 when the observed volatility distributions are available (Xu et
200 al., 2019).

201 Figure 1 shows the average volatility distributions of observed OOA and simulated SOA at the IAP site. The C^* of OOA
202 spans from 0.001 $\mu\text{g m}^{-3}$ to 100 $\mu\text{g m}^{-3}$, with an average value of 1.16 $\mu\text{g m}^{-3}$. The semi-volatile organic compounds (SVOC;
203 $0.3 < C^0 < 300 \mu\text{g m}^{-3}$) and the low-volatile organic compounds (LVOC; $3 \times 10^{-4} < C^0 < 0.3 \mu\text{g m}^{-3}$) (Donahue et al., 2012)
204 contribute 66.3 % and 33.7 % to OOA concentrations, respectively (Xu et al., 2019). The $T_{\text{g,org}}$ estimated from the observed
205 OOA volatility distributions is 286.7 K. Figure 1a shows the simulated volatility distributions of SOA with five C^* bins set to
206 be 0.1, 1, 10, 100, and 1000 $\mu\text{g m}^{-3}$ at 298 K, and ΔH_{vap} of 142, 131, 120, 109, and 98 kJ mol^{-1} used in the five C^* bins,
207 respectively (base case in Table 1). In this base case simulation, the SOA consists of 64.5 % SVOC and 35.3 % LVOC, and

208 most of the SVOC species are located in the C^* bin of $1 \mu\text{g m}^{-3}$. The simulated SOA in Fig. 1a has an average C^* of $0.64 \mu\text{g}$
209 m^{-3} and $T_{\text{g,org}}$ of 291.5 K , close to the values estimated from the volatility distributions of OOA. Figure 1b shows the
210 simulated volatility distributions of SOA with the lowest C^* bin set to be $0.0001 \mu\text{g m}^{-3}$ at 298 K with ΔH_{vap} of 40 kJ mol^{-1} ,
211 following the default option in the official WRF-Chem model 3.7.1 (Knote et al., 2015). In this sensitivity simulation (case A
212 in Table 1), the SOA consists of 40.4% extremely low-volatile organic compounds (ELVOC; $C^0 < 3 \times 10^{-4} \mu\text{g m}^{-3}$), which are
213 not determined in the observed OOA, leading to a much lower average C^* ($0.03 \mu\text{g m}^{-3}$) and a higher $T_{\text{g,org}}$ (309.0 K)
214 compared to the observations. In the following we estimate the $T_{\text{g,org}}$ and viscosity of SOA using the simulated volatility
215 distributions in the base case with the lowest C^* bin set as $0.1 \mu\text{g m}^{-3}$ at 298 K . The impacts of volatility distributions with
216 the incorporation of ELVOC (Fig. 1b) on viscosity estimations are evaluated in section 3.3.

217 Figure 2 shows that the model relatively well reproduces the observed hourly variations of RH, T , mass concentrations
218 of $\text{PM}_{2.5}$ and its major inorganic components (Figs. 2a – f), with the index of agreement (IOA, defined in the Supplement)
219 varied from ~ 0.70 for inorganic components to a higher value of 0.93 for T (Table S2 in the Supplement). The simulated
220 values of RH are constantly lower than the observations, with the mean bias (MB) being -10.97% (Table S2 in the
221 Supplement). The underestimation of RH observations was also found at other meteorological sites in the North China Plain
222 in our previous studies (Qu et al., 2019; Zhang et al., 2022), which would affect the SOA viscosity estimations. The effects
223 of RH on the viscosity estimations are evaluated in section 3.3. Figure 2g shows that the model could generally reproduce
224 the observed temporal variations of OOA concentrations, but largely underestimates the observation peaks (MB = $-5.88 \mu\text{g}$
225 m^{-3} , the normalized mean bias NMB = -53.28% , Table S2 in the Supplement). Incorporation of the SOA formed from
226 intermediate-volatile organic compounds (IVOCs) (Miao et al., 2021; Chang et al., 2022) would increase the simulated SOA
227 concentrations, which is beyond the scope of this study and will be considered in our future work. The simulated SOA mean
228 concentration is $5.15 \mu\text{g m}^{-3}$. Although it is lower than the observed value of $11.03 \mu\text{g m}^{-3}$, this difference in the simulated
229 and observed SOA concentrations would not affect the viscosity predictions significantly. The SOA viscosity has a much
230 closer relation with the volatility rather than its mass loadings (Champion et al., 2019; Li et al., 2020). In our previous study
231 we estimated $T_{\text{g,org}}$ and viscosities at different OA mass loadings varied from 1 to $1000 \mu\text{g m}^{-3}$, showing that the simulated
232 viscosities were very similar, particularly when RH was higher than 50% (DeRieux et al., 2018).

233 As the WRF-Chem model underestimates the observed RH at the IAP site, we calculate the SOA viscosity using the
234 simulated and observed RH, respectively. Figure 2h shows that the viscosities calculated at the two conditions are similar at
235 most times during the simulated period, ranging mainly from $\sim 10^2 \text{ Pa s}$ to 10^{10} Pa s , with a median value of $\sim 10^7 \text{ Pa s}$,
236 indicating that a semi-solid phase state frequently occurs. The underestimations of the observed RH by WRF-Chem mainly
237 impact the phase state estimations at relatively high RH. For example, SOA particles occur as liquid when the observed RH
238 is higher than $\sim 75 \%$; however, they remain in a semi-solid phase state at the simulated RH. The bulk diffusion coefficients
239 (D_b) of organic molecules range from 10^{-18} to $10^{-11} \text{ cm}^2 \text{ s}^{-1}$ at the simulated RH (Fig. 2i), leading to the mixing timescales

240 within 200 nm SOA particles being seconds to years, with 61 % of the time > 1 hour (Fig. S2 in the Supplement). We
241 highlight the mixing timescale of 1 hour as the time step adopted in CTMs is often ~ 0.5 – 1 hour (Maclean et al., 2021).
242 Current CTMs usually assume that the gas-particle partitioning of SVOCs reaches equilibrium quickly within the time step
243 (Pankow, 1994; Donahue et al., 2006). When the mixing timescales of organics within SOA particles are $\leq \sim 1$ hour, the
244 instantaneous equilibrium is a reasonable assumption. However, when the mixing timescales of organics are longer than ~ 1
245 hour, non-equilibrium between the gas phase and the particle phase, i.e., the kinetic partitioning may need to be considered in
246 simulating the SOA formation in CTMs (Shiraiwa and Seinfeld, 2012; Zaveri et al., 2018; Li and Shiraiwa, 2019; Zaveri et
247 al., 2020; He et al., 2021; Jathar et al., 2021; Maclean et al., 2021; Shiraiwa and Pöschl, 2021; Shrivastava et al., 2022).

248 The vertical profiles of SOA viscosity exhibit diurnal variations. Figure 3a shows the median diurnal and vertical
249 profiles of predicted SOA viscosity at the IAP site. The SOA particles remain highly viscous ($\sim 10^7 - 10^8$ Pa s) at the surface,
250 with a higher viscosity occurring from late afternoon to early evening, during which the RH is less than 20 %, lower than the
251 rest time of the day (Fig. S3 in the Supplement). The SOA particles become more viscous at higher altitudes than the surface
252 and adopt the phase transition from a semi-solid phase to a solid phase at ~4 km at the IAP site. The predicted altitude with
253 the phase transition is ~2 km higher than our previous global model prediction for the region of East China which was an
254 average of five years' simulations (Shiraiwa et al., 2017). Tan et al. (2020) inferred the phase state of submicrometer
255 particles in Beijing from the surface to an altitude of ~1 km using a polarization lidar and found that the particle phase state
256 exhibits a vertical variation. Further observations of SOA viscosity at high altitudes are needed to better understand the
257 viscosity vertical profiles and validate our predictions. Figure 3b shows that the mixing timescales for organic molecules
258 within 200 nm SOA particles are approximately 1 hour at the surface, and longer than ~10 hours at altitudes higher than 1
259 km, indicating that kinetic limitations in the gas-particle partitioning may be required to accurately predict SOA mass
260 concentrations in summer Beijing, particularly in the upper planetary boundary layer and the free troposphere.

261 **3.2 Simulated glass transition temperature and viscosity of SOA particles over China**

262 The glass transition temperature of the dry organic phase ($T_{g,org}$) shows a geospatial gradient over China. Figure 4a shows the
263 median surface values of $T_{g,org}$ calculated in the base simulation (Table 1). $T_{g,org}$ ranges from ~287 – 305 K over most areas of
264 China, with lower values occurring mainly over the southeast and higher values over the northwest. The $T_{g,org}$ range
265 simulated by the WRF-Chem model is consistent with our previous global simulations of $T_{g,org}$ that varied from ~285 K to
266 310 K at the surface over China (Shiraiwa et al., 2017). The geospatial variation in $T_{g,org}$ is related to the simulated SOA
267 volatility distributions. Figure 4b shows the mass fractions of SOA species distributed in the lowest volatility bin (SOAX
268 with $C^* = 0.1 \mu\text{g m}^{-3}$ at 298 K). The mass fractions of SOAX are mostly 20 – 35 % in the southeastern China, indicating that
269 the majority of the simulated SOA formed from VOCs is semi-volatile. In these areas, the simulated SOA mass
270 concentrations are higher than the other locations of China (Fig. S4 in the Supplement) (Li et al., 2022), which is favorable

271 for more SVOCs partitioning into the particle phase, leading to relatively low values of $T_{g,org}$ (Fig. 4a). LVOCs are more
272 frequently contained in aged SOA particles in remote areas, e.g., some areas in the northwestern China where the SOA mass
273 concentrations are very low, resulted in higher $T_{g,org}$ values.

274 The relative humidity plays an important role in regulating SOA viscosity (Koop et al., 2011). Considering the water
275 uptake by SOA particles in the phase state estimations, the predicted geospatial patterns in the viscosity (Fig. 5a) and RH
276 (Fig. S4 in the Supplement) are very similar with each other, particularly in southern and northeastern China. SOA particles
277 are predicted to mainly be liquid or with a low viscosity ($< 10^4$ Pa s) in the southeast. Figure 5b shows the frequency of
278 liquid phase state, which is calculated as the percent time that an organic aerosol particle is in the liquid phase state during
279 the simulated period. The frequency of liquid particles varies from $\sim 30\%$ to 70% in the southeastern China. The lowest
280 viscosity with the highest frequency of liquid particles occurs over the southern Tibetan Plateau where RH is very high (Fig.
281 S4), which is contributed by summer monsoons and regional moisture recycling (Dong et al., 2016). The SOA particles in
282 the central and northeastern China are predicted to be semi-solid, with the viscosity varied from 10^5 to 10^8 Pa s (Fig. 5a).
283 Highly viscous ($\eta > 10^8$ Pa s) or solid SOA particles are mainly found in the northwest, particularly over the northern Tibetan
284 Plateau where the ambient temperatures are lower than other areas of China (Fig. S4). The frequency of liquid SOA particles
285 in most areas with the latitude higher than 30°N is less than 20% (Fig. 5b).

286 The simulated geospatial pattern in SOA viscosity over China agrees with previous global simulations and ambient
287 measurements. Our previous global simulations predicted a lower viscosity ($\eta < 10^3$ Pa s) in SOA particles in southeastern
288 China and a higher viscosity ($\eta > 10^8$ Pa s) in northwestern China (Shiraiwa et al., 2017; Li et al., 2020), similar to the
289 WRF-Chem simulations in this study. Interestingly, the occurrence of liquid particles over the southern Tibetan Plateau in
290 summer simulated by the WRF-Chem was not found in our previous global predictions, which was an average of five years'
291 simulations (Shiraiwa et al., 2017). The semi-solid phase state of SOA particles simulated in Beijing is consistent with both
292 particle bounce measurements (Liu et al., 2017) and the $\text{PM}_{2.5}$ phase state determined by the poke-and-flow technique (Song
293 et al., 2022). The simulated viscosity of SOA particles is 0.15 Pa s in Shenzhen, a coastal urban city in southeastern China,
294 which also agrees with the findings in the previous bounce measurements indicating that the submicron particles in
295 Shenzhen are in the liquid state (Liu et al., 2019).

296 The phase state of SOA particles is affected by ambient conditions and the particle chemical composition (Koop et al.,
297 2011). Figure 6 shows the median values of viscosity as a function of RH, T and the mass fraction of low-volatility
298 compound (SOAX with C^* of $0.1 \mu\text{g m}^{-3}$ at 298 K) calculated for selected regions in the northern China, southern China,
299 northern Qinghai-Tibet Plateau, and southern Qinghai-Tibet Plateau as specified by white boxes in Fig. 5a. There is a strong
300 inverse relationship between SOA viscosity and relative humidity with high RH ($> \sim 60\%$) as the dominant factor
301 determining the phase state of SOA particles. When RH is lower than $\sim 60\%$, the predicted viscosity is affected by both RH
302 and T . For example, the SOA particles occur mainly as solid over the Northern Qinghai-Tibet Plateau while occur as

303 semi-solid over the Northern China within similar RH ranges ($20\% < \text{RH} < 60\%$); the reason is that the ambient T over the
304 Northern Qinghai-Tibet Plateau is much lower ($\sim 20\text{ K}$ lower) than the Northern China (Fig. 6b). When RH is relatively low,
305 the viscosity of SOA particles is also influenced by particle chemical composition, i.e., the SOA particles composed of high
306 mass fractions of low volatility compounds tend to have higher viscosity values (Fig. 6c). RH is the main factor driving the
307 diurnal variations of SOA viscosity in our simulations. Figure S5 in the Supplement shows that SOA particles have higher
308 viscosity in the daytime than the nighttime as RH in the daytime is lower than the nighttime (Fig. S6). Compared to the
309 northern China, the southern China exhibits stronger diurnal variations in SOA viscosity that SOA particles occur mainly as
310 semi-solid in the daytime and liquid in the nighttime. Highly viscous or solid SOA particles are found in the northern China
311 during both daytime and nighttime (Fig. S5).

312 The bulk diffusion coefficient is an important parameter determining the mass-transport and mixing rates, which can be
313 predicted by the particle viscosity through the fractional Stokes-Einstein relation (see the Method section and the
314 Supplement). The D_b of organic molecules is predicted to be $> \sim 10^{-10}\text{ cm}^2\text{ s}^{-1}$ in the southern China. The highest value is
315 $\sim 10^{-5}\text{ cm}^2\text{ s}^{-1}$ occurring in liquid SOA particles in the southern Tibetan Plateau (Fig. 7a) because of the very low viscosity
316 simulated over this region (Fig. 5a). The D_b of organic molecules within semisolid SOA particles is $\sim 10^{-18} - 10^{-10}\text{ cm}^2\text{ s}^{-1}$ in
317 the central and northeastern China, and lower than $\sim 10^{-18}\text{ cm}^2\text{ s}^{-1}$ in highly viscous and solid particles in most areas of the
318 northwestern China. Figure 7b shows the percent time that the mixing timescale of organic molecules in 200 nm particles is
319 less than 1 h in the entire simulation period. The mixing timescale is found nearly always less than 1 h in the southeastern
320 region of the “Hu Huanyong Line”. The “Hu Huanyong Line”, proposed by the Chinese geographer Huanyong Hu, divides
321 China into two parts based on contrasting population densities (Hu, 1935), which was found also useful characterizing the
322 drought conditions, with the northwestern region much dryer than the southeastern region (Zeng et al., 2021). The mixing
323 timescale of organic molecules in highly viscous or solid SOA particles in the northwest of the “Hu Huanyong Line” is often
324 longer than 1 h (the frequency $> 70\%$), indicating that in these areas kinetic limitations of bulk diffusion should be
325 considered in SOA partitioning. Compared to the diffusion coefficients of organic molecules, the D_b of water molecules in
326 SOA particles at the surface is several orders of magnitude larger, with the values higher than $10^{-10}\text{ cm}^2\text{ s}^{-1}$ in the southeast,
327 and as low as $\sim 10^{-13}\text{ cm}^2\text{ s}^{-1}$ in the northwestern China (Fig. 7c). The mixing timescales of water molecules in SOA particles
328 with a diameter of 200 nm are of the order of milliseconds in the southeast and seconds in the northwest of China (Fig. S7 in
329 the Supplement), indicating that the activation of cloud condensation nuclei would not be inhibited, in agreement with our
330 previous global simulations (Shiraiwa et al., 2017).

331 Figure 8 shows the simulated $T_{g,\text{org}}$ and the phase state of SOA particles, as well as the mixing timescale of organic
332 molecules in SOA particles at 500 hPa. The $T_{g,\text{org}}$ ranges from $\sim 285\text{ K} - 295\text{ K}$, lower than the $T_{g,\text{org}}$ simulated at the surface
333 (Fig. 8a). The reason is that the mass fractions of LVOCs (SOAX with C^* of $0.1\text{ }\mu\text{g m}^{-3}$ at 298 K) at 500 hPa (Fig. S8 in the
334 Supplement) are smaller than the surface values (Fig. 4b). The low temperature at 500 hPa is favorable for SVOCs

335 partitioned into the particle phase, thus compared to the surface conditions, there is less semi-volatile vapors undergoing the
336 further gas-phase aging forming SOAX species. The percent time that an organic aerosol particle is in the liquid phase state
337 (the frequency of liquid SOA particles) at 500 hPa in the southeastern China is 20 – 35 % (Fig. 8b), which is ~ 20 % lower
338 than the surface values (Fig. 5b). In the northern China, the frequency of liquid SOA particles at 500 hPa is similar to the
339 results at the surface, which is related to the RH spatial patterns (Fig. S9 in the Supplement). The mixing timescale of
340 organic molecules in 200 nm SOA particles is frequently longer than 1 h at 500 hPa, with the frequency > 70 % in the
341 northern China and ~ 40 % in the southeastern China (Fig. 8c). The τ_{mix} is relatively short (the frequency of $\tau_{\text{mix}} \leq 1$ h being ~
342 80 %) in some areas of the southwestern China at 500 hPa, where relatively high RH could occur (Fig. S9 in the Supplement)
343 in the season of summer monsoon (Huang et al., 1998).

344 3.3 Sensitivity simulations

345 3.3.1 Impacts of volatility distributions on phase state estimations

346 The volatility and viscosity of organic aerosols are closely related (Rothfuss and Petters, 2017; Shiraiwa et al., 2017;
347 Champion et al., 2019; Zhang et al., 2019b; Li et al., 2020). In this section we conduct sensitivity simulations (case A, Table
348 1) to evaluate how the simulated volatility distributions affect the phase state estimations. The lowest C^* bin in the base case
349 is $0.1 \mu\text{g m}^{-3}$ at 298 K, with ΔH_{vap} of 142 kJ mol^{-1} , which does not incorporate ELVOC species at the room temperature. The
350 sensitivity simulation (case A) adopts the default setting in the official WRF-Chem model v3.7.1, assuming that the lowest
351 C^* is $0.0001 \mu\text{g m}^{-3}$ at 298 K, with ΔH_{vap} of 40 kJ mol^{-1} (Knote et al., 2015). Figure 9a shows that including these ELVOCs at
352 298 K with a relatively small ΔH_{vap} mainly affect the $T_{\text{g,org}}$ simulated over remote areas, e.g., the northwestern China and the
353 marine areas, where the simulated $T_{\text{g,org}}$ is increased by 30 – 40 K. In other regions of China, the changes in $T_{\text{g,org}}$ are less
354 than 25 K. Although consideration of these ELVOCs could affect the simulated $T_{\text{g,org}}$ at the surface obviously in remote areas,
355 i.e., the northwestern China, it does not impact the predicted frequency of the occurrence of a liquid phase state (Fig. 9b), as
356 in these dry areas the SOA particles are highly viscous. Including ELVOC formation at 298 K (ΔH_{vap} of 40 kJ mol^{-1}) mainly
357 affects the phase state estimations in areas with a moderate humidity. For example, in some areas of the southeastern China
358 with ~ 70 % RH, the SOA particles are predicted to be more viscous, with the frequency of a liquid phase state decreasing by
359 up to 12 % (Fig. 9b). These results indicate that the SOA phase state estimations in the base and sensitivity case A are
360 generally in agreement in the simulated episode in this study. Ambient measurements of organic aerosol volatility
361 distributions are still sparse over China. A recent field study showed that the ELVOCs contributed more than half to the OA
362 mass observed at a regional background site near the Bohai Sea (Feng et al., 2023), which resulted in an estimated viscosity
363 much higher than our WRF-Chem simulations. More field volatility distribution measurements should be conducted over
364 China to further evaluate the effects of ELVOCs and how to choose reasonable values of the enthalpy of vaporization that
365 would affect phase state estimations.

366 3.3.2 Impacts of RH and the water absorbed by inorganics on phase state estimations

367 RH is an important parameter affecting the phase state estimations. We perform a sensitivity calculation (case B, Table 1)
368 with the simulated RH increased by a factor of 10 % to compensate for the fact that the current model underestimates the
369 observed RH as shown in Fig. 2 and found in our previous simulations (Qu et al., 2019; Zhang et al., 2022). The increases in
370 simulated RH lead to more occurrence (the liquid frequency increased by 10 – 20 %) of liquid SOA particles in southeastern
371 China where the predicted RH is ranged mainly from 70 to 80 % (Fig. S4 in the Supplement), with very limited effects in
372 phase state predictions in relatively dry areas, e.g., the northern China (Fig. 10a).

373 Besides RH, the mixing state of the organic and inorganic species in atmospheric particles also plays an important role
374 in the phase state of ambient particles. The SOA components are assumed to be phase separated from inorganic compounds
375 in particles in our base simulation, which is consistent to recent ambient observations showing that the phase separation with
376 an organic-rich shell and an inorganic core was a frequent phenomenon in individual particles (diameters > 100 nm)
377 collected over the North China Plain (Li et al., 2021a). To assess the potential effects of inorganic compounds on the phase
378 state of ambient particles, we perform a sensitivity calculation (case C, Table 1) assuming that the organic and inorganic
379 compounds are internally mixed in one phase. In this case the water absorbed by inorganic species can lower the particle
380 viscosity relative to the organic fraction alone. Figure 10b shows that the water associated with inorganic species can
381 significantly lower the viscosity over most areas of China, with the liquid frequency increased by 15 – 45 % in the southeast,
382 and 5 – 15 % in some areas of the northeast. The effects over dry lands in the northwestern China are relatively small.
383 Previous studies showed that at such mixing condition with one phase, on one hand, it is expected that the inorganic salts that
384 often have lower T_g compared to SOA compounds would further lower the particle viscosity (Dette and Koop, 2015). On the
385 other hand, the presence of divalent inorganic ions could increase the viscosity of mixed organic-inorganic particles,
386 enabling a humidity-dependent gel phase transition through cooperative ion-molecule interactions (Richards et al., 2020).
387 For complex mixtures of primary OA, SOA and inorganics, it was found that three distinct phases could occur (Huang et al.,
388 2021). The impacts of the mixing state of organic and inorganic compounds on the phase state of multicomponent particles
389 in ambient air warrant further investigations in future studies (Lilek and Zuend, 2022; Schervish and Shiraiwa, 2023).

390 4 Conclusions

391 We previously developed a new parameterization predicting the glass transition temperature of an organic compound as a
392 function of its volatility (Li et al., 2020). Based on this new parameterization, we use the WRF-Chem model and simulate the
393 T_g and viscosity of SOA particles over China in summer of 2018. This is the first time that spatial and temporal variations in
394 the SOA phase state over China are investigated on a regional scale. The main conclusions are summarized below.

395 (1) Simulations show that T_g values of dry SOA ($T_{g,org}$) range from ~287 K to 305 K over most areas of China at the surface,

396 consistent with our previous simulated results based on a global transport model (Shiraiwa et al., 2017). The $T_{g,org}$ is higher
397 in the northwestern China than the southeastern China. This geospatial variation in $T_{g,org}$ is related to the simulated SOA
398 volatility distributions that SOA particles in northwestern China have relatively low volatilities.

399 (2) Considering the water uptake by SOA particles, the SOA viscosity also shows a prominent geospatial gradient that highly
400 viscous or solid SOA particles are mainly found in the northwestern China. The frequency of liquid SOA particles in most
401 areas with the latitude higher than 30°N is less than 20 %. A very large spatial variation in SOA phase state over the
402 Qinghai-Tibet Plateau was found and we recommend measurements in ambient particle phase state to be conducted over this
403 area, one of the most sensitive regions to climate change.

404 ~~(3) The mixing timescale of organic molecules in 200 nm SOA particles is calculated based on the simulated particle~~
405 ~~viscosity and the bulk diffusion coefficients of organic molecules. Calculations show that at the surface and at 500 hPa, the~~
406 ~~percent time of τ_{mix} longer than 1 h is $> \sim 70\%$ in the northwest of the “Hu Huanyong Line”. The implication of this result is~~
407 ~~that when the τ_{mix} values are greater than roughly 1 h, which is longer than the typical time step in CTMs, the instantaneous~~
408 ~~equilibrium partitioning usually assumed in SOA formation simulations is subject to be re-evaluated. We recommend to test~~
409 ~~the effects of kinetic partitioning considering the bulk diffusion in viscous particles on the prediction of SOA mass~~
410 ~~concentrations and size distributions over the areas with long mixing timescale of organic molecules.~~

411 (43) The average volatility (C^*) and $T_{g,org}$ of the simulated SOA agree well with the values estimated from ambient
412 measurements of OOA volatilities at the IAP site in Beijing, where ELVOCs were not determined in the observed OOA (Xu
413 et al., 2019). The sensitivity simulation considering the formation of ELVOCs shows that compared to the base simulation,
414 the frequency of a liquid phase state does not change in most areas of the northern China. In some areas of the southeastern
415 China the SOA particles become more viscous with the percent time that a SOA particle is in the liquid phase state decreases
416 by up to 12 %. It needs more field volatility measurements to evaluate the effects of ELVOCs on OA phase state estimations
417 over China.

418 (54) Differed from the base simulation that SOA components are assumed to be phase separated from inorganic compounds
419 in particles, we conduct a sensitivity simulation assuming that the organic and inorganic compounds are internally mixed in
420 one phase. We show that the water absorbed by inorganic species has a significant impact lowering the simulated viscosity
421 over the southeastern China, with the liquid frequency increased by 15 – 45 %. Future work should consider the effects of
422 the mixing state of organic and inorganic compounds on the prediction of the phase state of multicomponent particles in
423 ambient air.

424 (35) The mixing timescale of organic molecules in 200 nm SOA particles is calculated based on the simulated particle
425 viscosity and the bulk diffusion coefficients of organic molecules. Calculations show that at the surface and at 500 hPa, the
426 percent time of τ_{mix} longer than 1 h is $> \sim 70\%$ in the northwest of the “Hu Huanyong Line”. The implication of this result is
427 that when the τ_{mix} values are greater than roughly 1 h, which is longer than the typical time step in CTMs, the instantaneous

428 equilibrium partitioning usually assumed in SOA formation simulations is subject to be re-evaluated. We recommend to test
429 the effects of kinetic partitioning considering the bulk diffusion in viscous particles on the prediction of SOA mass
430 concentrations and size distributions over the areas with long mixing timescale of organic molecules.

431
432 ~~In summary, our simulations demonstrate the spatial distributions of the glass transition temperature and viscosity of~~
433 ~~SOA particles over China on a regional scale for the first time. The further calculations of the mixing timescale of organic~~
434 ~~molecules in SOA particles have an implication of the need to evaluate the effects of the phase state on predictions of SOA~~
435 ~~gas particle partitioning, and thus the SOA mass concentrations and size distributions in CTMs.~~

436 *Data availability.* The simulation data in this study are available upon request from the corresponding author
437 (liying-iap@mail.iap.ac.cn).

438 *Supplement.* The supplement related to this article is available on-line.

439 *Author contributions.* YL and MS designed the research. ZZ, HR, and YQ performed the WRF-Chem modeling. ZZ, YL,
440 and HR analyzed the simulation data. WZ, WX, WH, and YS provided observation data. YL and ZZ wrote the manuscript.
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442 *Competing interests.* At least one of the (co-)authors is a member of the editorial board of Atmospheric Chemistry and
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776 **Table 1.** Sensitivity calculations for evaluating the effects of simulated SOA volatility distributions (sensitivity case A), RH (sensitivity
777 case B) and the water absorbed by inorganic components (sensitivity case C) on viscosity estimations.

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| Cases | C^* at 298 K and ΔH_{vap} in each volatility bin | The liquid water content considered in viscosity estimations | RH |
|--------------------|--|---|---|
| Base case | 0.1, 1, 10, 100, and 1000 $\mu\text{g m}^{-3}$ with ΔH_{vap} of 142, 131, 120, 109, and 98 kJ mol^{-1} | water absorbed by SOA particles with the assumption that SOA particles are externally mixed with inorganics | RH simulated by WRF-Chem |
| Sensitivity case A | 0.0001, 1, 10, 100, and 1000 $\mu\text{g m}^{-3}$ with ΔH_{vap} of 40, 131, 120, 109, and 98 kJ mol^{-1} | water absorbed by SOA particles with the assumption that SOA particles are externally mixed with inorganics | RH simulated by WRF-Chem |
| Sensitivity case B | 0.1, 1, 10, 100, and 1000 $\mu\text{g m}^{-3}$ with ΔH_{vap} of 142, 131, 120, 109, and 98 kJ mol^{-1} | water absorbed by SOA particles with the assumption that SOA particles are externally mixed with inorganics | RH simulated by WRF-Chem increased by a factor of 10% |
| Sensitivity case C | 0.1, 1, 10, 100, and 1000 $\mu\text{g m}^{-3}$ with ΔH_{vap} of 142, 131, 120, 109, and 98 kJ mol^{-1} | water absorbed by both SOA particles and inorganic components with the assumption that SOA particles are internally mixed with inorganics | RH simulated by WRF-Chem |

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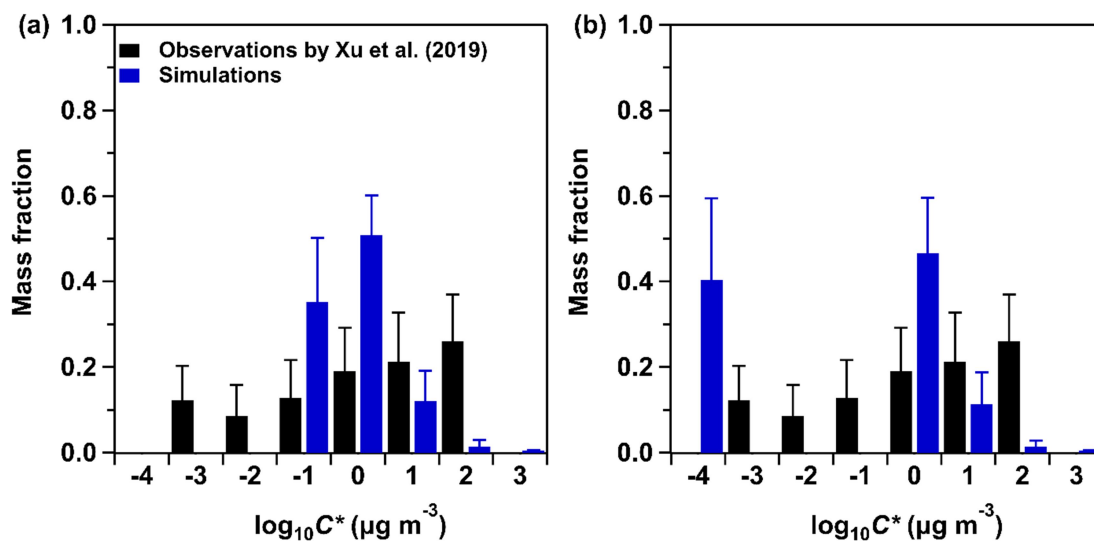
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806 **Figure 1.** Comparison of the average volatility distributions of observed OOA and simulated SOA at the IAP site during 20 May – 23 June
 807 2018. The black bars represent the volatility distributions of observed OOA adopted from Xu et al. (2019). The blue bars represent the
 808 volatility distributions of SOA simulated by WRF-Chem, with five C^* bins set to be 0.1, 1, 10, 100, and 1000 $\mu\text{g m}^{-3}$ at 298 K in (a), and
 809 0.0001, 1, 10, 100, and 1000 $\mu\text{g m}^{-3}$ at 298 K in (b). The blue error bars represent the one standard deviation.

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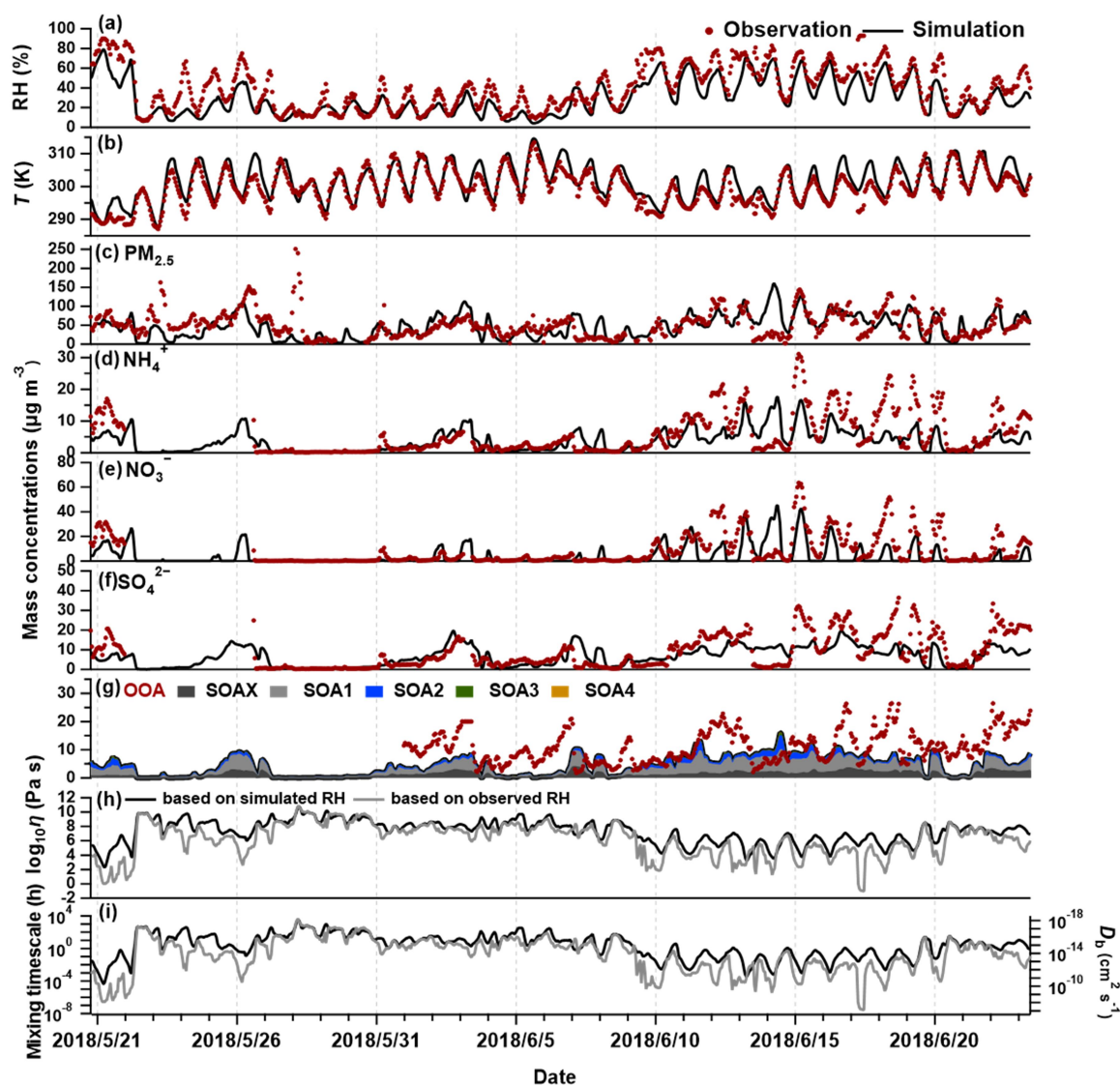
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 832 **Figure 2.** Observations and simulations of temporal variations of (a) RH, (b) T , (c) $\text{PM}_{2.5}$ concentrations, (d) NH_4^+ concentrations, (e)
 833 NO_3^- concentrations, and (f) SO_4^{2-} concentrations at the IAP site. (g) Observed OOA concentrations (red dots) and simulated SOA
 834 concentrations, with SOAX, SOA1, SOA2, SOA3, and SOA4 represent the SOA with C^* of 0.1, 1, 10, 100, and 1000 $\mu\text{g m}^{-3}$ at 298 K,
 835 respectively. (h) SOA viscosity, and (i) bulk diffusion coefficients and mixing timescale of organic molecules within 200 nm SOA particles
 836 calculated using the RH simulated by the WRF-Chem model or the RH observed at the IAP site.

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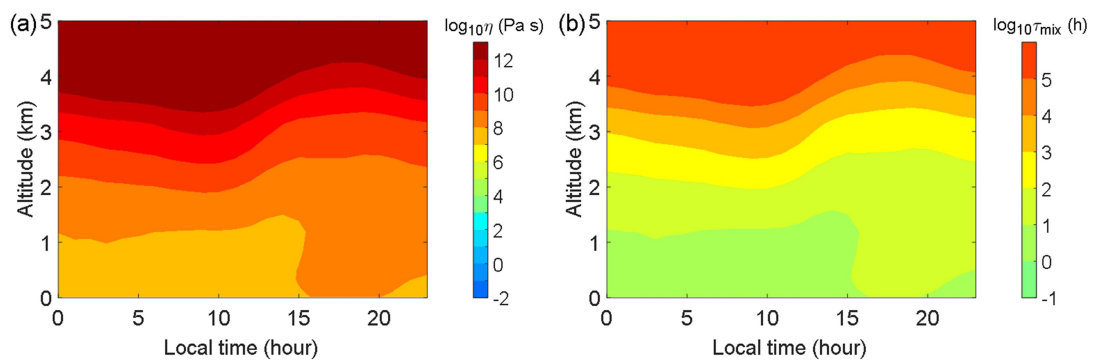
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844 **Figure 3.** Median diurnal and vertical profiles of estimated (a) SOA viscosity and (b) mixing timescales for organic molecules within 200
 845 nm SOA particles at the IAP site during May 20 – June 23 in 2018. Note: altitude is approximate and estimated from WRF pressure layers.

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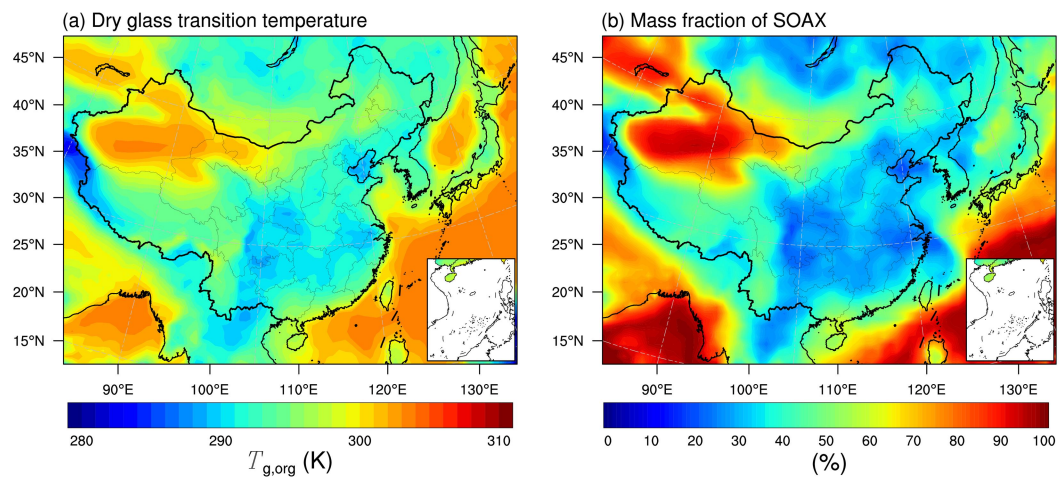
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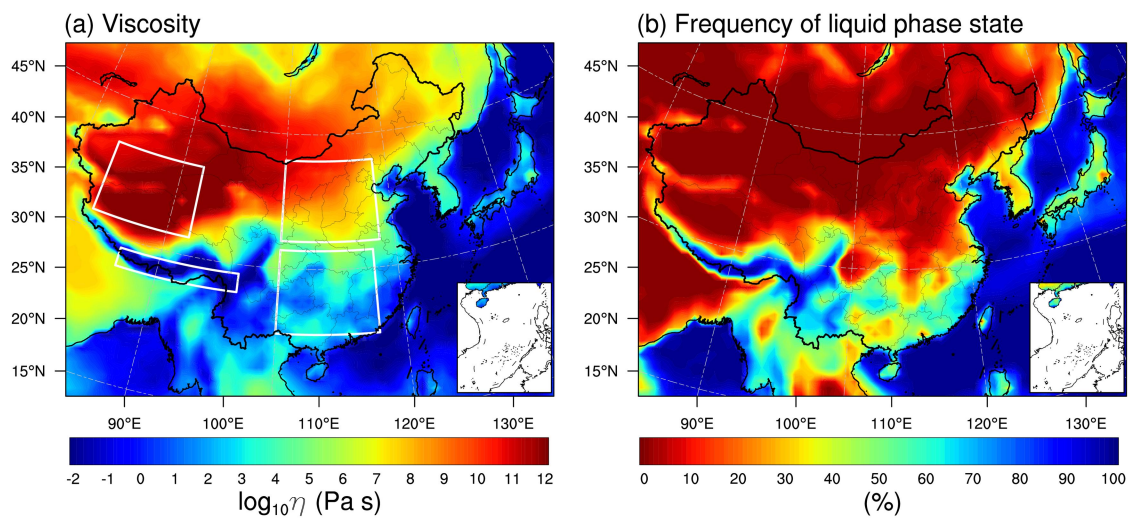
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 861 **Figure 4.** The predicted median surface values of (a) glass transition temperature of SOA particles at the dry condition and (b) mass
 862 fractions of SOAX ($C^* = 0.1 \mu\text{g m}^{-3}$ at 298 K) during May 20 – June 23 in 2018 simulated in the base case (Table 1).

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878 **Figure 5.** WRF-Chem predicted (a) median surface values of viscosity and (b) the percent time that an organic aerosol particle is in the
 879 liquid phase state during May 20 – June 23 in 2018. The white boxes in panel (a) specify the selected regions in the northern China,
 880 southern China, northern Qinghai-Tibet Plateau, and southern Qinghai-Tibet Plateau.

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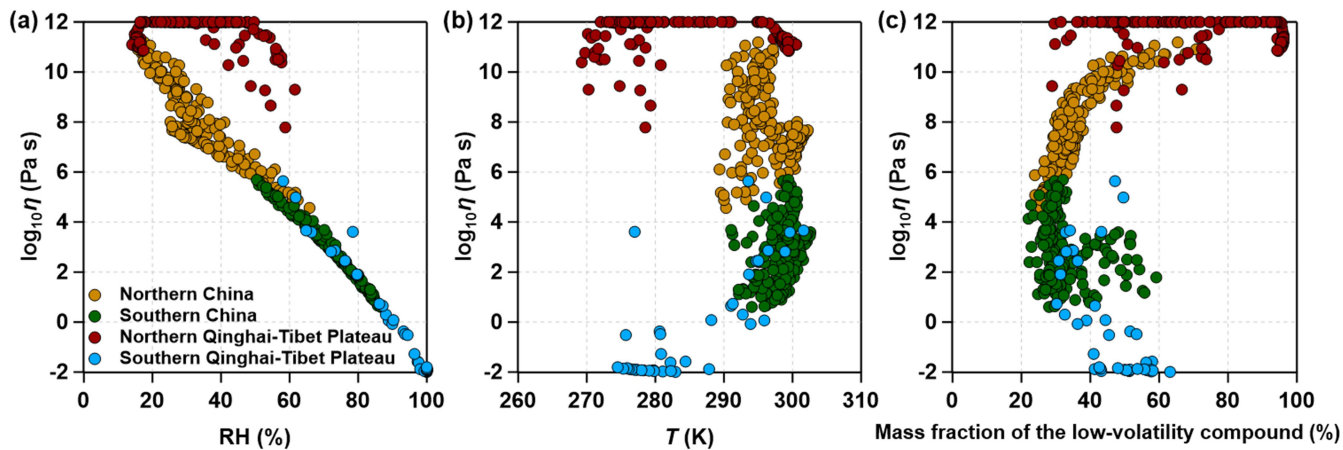
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898 **Figure 6.** The median values of viscosity as a function of (a) RH, (b) T and (c) the mass fraction of SOAX ($C^* = 0.1 \mu\text{g m}^{-3}$ at 298 K)
 899 calculated for selected regions in the northern China, southern China, northern Qinghai-Tibet Plateau, and southern Qinghai-Tibet Plateau
 900 as specified by white boxes in Fig. 5a during May 20 – June 23 in 2018.

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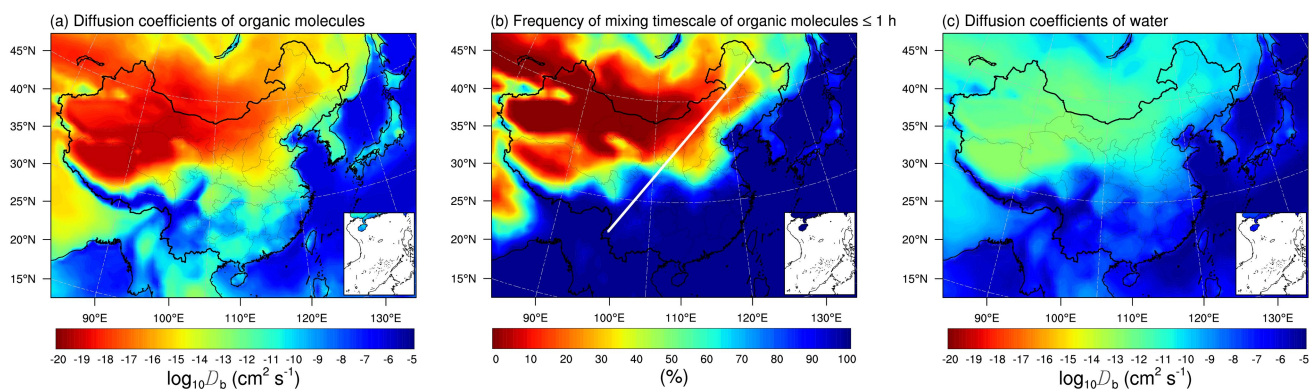
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920 **Figure 7.** WRF-Chem predicted median surface values of the diffusion coefficients of (a) organic molecules and (c) water molecules in
 921 SOA particles. (b) The percent time that the mixing timescale of organic molecules in a 200 nm particle is less than 1 h during May 20 –
 922 June 23 in 2018. The white line indicates the “Hu Huanyong Line” (Hu, 1935).

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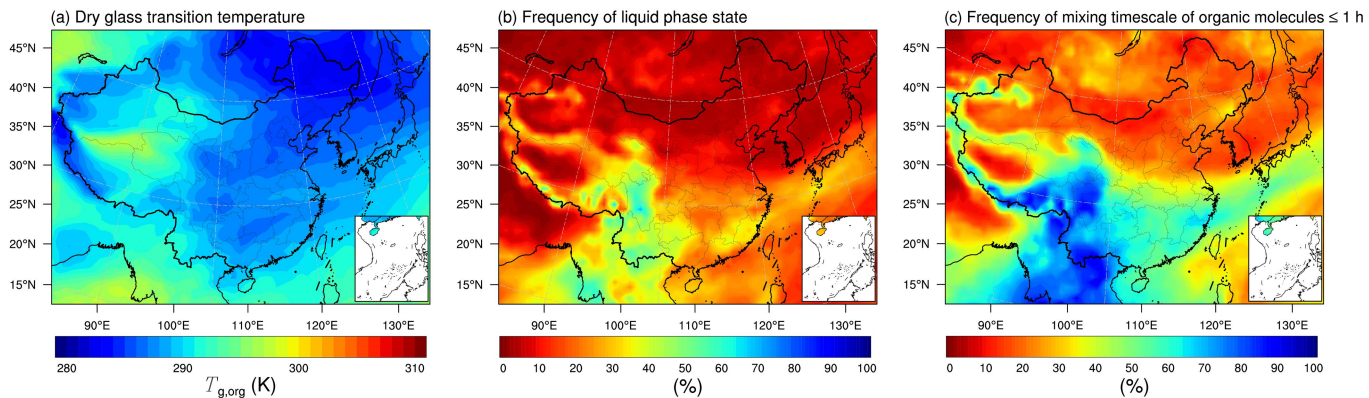
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937 **Figure 8.** WRF-Chem predicted median values of (a) glass transition temperature of SOA particles at the dry condition, (b) the percent
 938 time that an organic aerosol particle is in the liquid phase state, and (c) the percent time that the mixing timescale of organic molecules in a
 939 200 nm particle is less than 1 h at 500 hPa during May 20 – June 23 in 2018.

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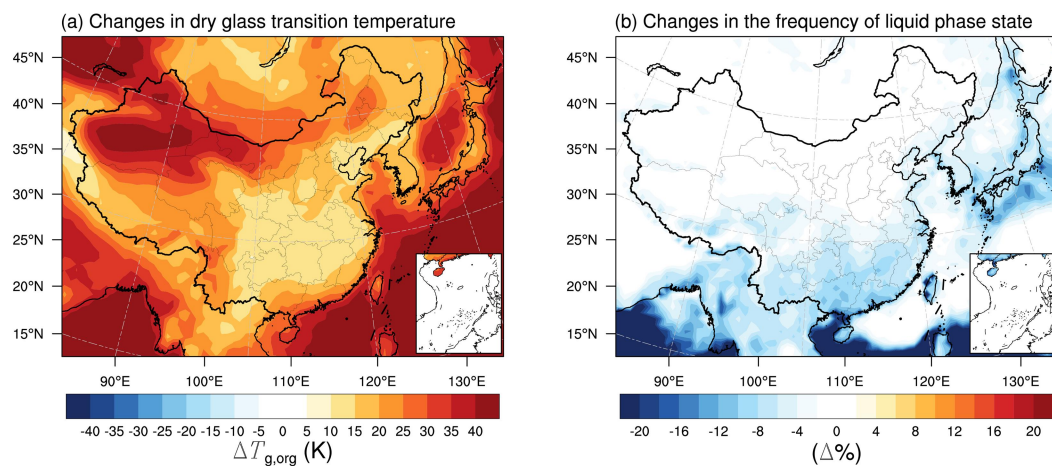
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953 **Figure 9.** Modelled median differences of (a) glass transition temperature of SOA particles at the dry condition, and (b) the percent time
 954 that an organic aerosol particle is in the liquid phase state between a sensitivity case with the lowest C^* of $0.0001 \mu\text{g m}^{-3}$ at 298 K (ΔH_{vap} of
 955 40 kJ mol^{-1} , case A in Table 1) and a base case with the lowest C^* of $0.1 \mu\text{g m}^{-3}$ at 298 K (ΔH_{vap} of 142 kJ mol^{-1} , Table 1).

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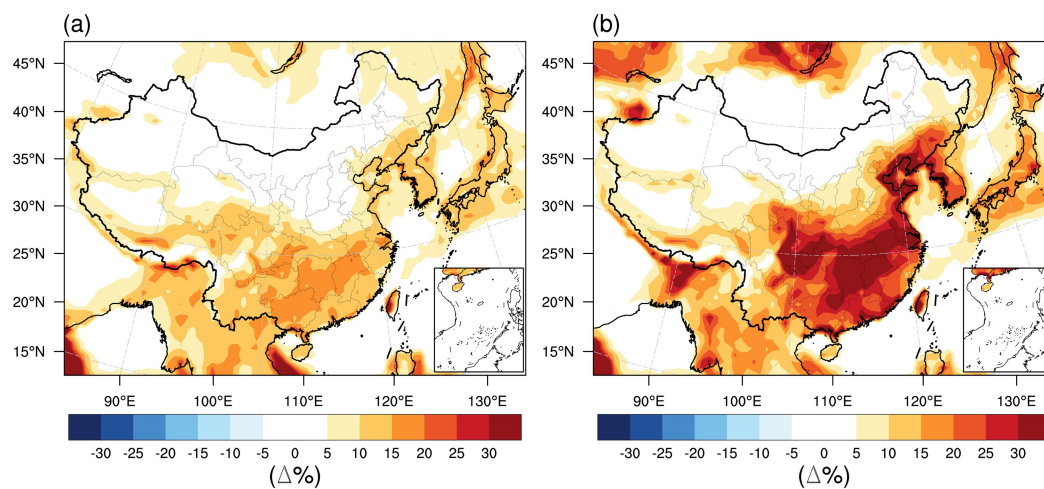
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 967 **Figure 10.** (a) Impacts of RH on the percent time that an organic aerosol particle is in the liquid phase state: modelled median differences
 968 between a case with the WRF-Chem simulated RH increased by a factor of 10 % (case B, Table 1) and a base case (Table 1) with the
 969 WRF-Chem simulated RH. (b) Impacts of the water absorbed by inorganics on the percent time that an organic aerosol particle is in the
 970 liquid phase state: modelled median differences between a case considering the water absorbed by both SOA particles and inorganics (case
 971 C, Table 1) and a base case (Table 1) considering the water absorbed by SOA particles solely.

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