# Developing a tile drainage module for <u>the</u> Cold Regions Hydrological Model: Lessons from a farm in Southern Ontario, Canada

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## 12 Abstract

13 Systematic tile drainage is used extensively in poorly drained agricultural lands to remove excess 14 water and improve crop growth; however, tiles can also transfer nutrients from farmlands to 15 downstream surface water bodies, leading to water quality problems. Thus, there is a need to 16 simulate the hydrological behaviour of tile drains to understand the impacts of climate or land 17 management change on agricultural surface and subsurface runoff. The Cold Regions 18 Hydrological Model (CRHM) is a physically based, modular modeling system developed for cold 19 regions. Here, a tile drainage module is developed for CRHM. A multi-variable, multi-criteria 20 model performance evaluation strategy was deployed to examine the ability of the module to 21 capture tile discharge under both winter and summer conditions (NSE>0.29, RSR<0.84 and PBias 22 <20 for tile flow and saturated storage water table simulations). Initial model simulations run at a

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23	15-min interval did not satisfactorily represent tile discharge; however, model simulations
24	improved when the time step was lengthened to hourly but also with the explicit representation of
25	capillary rise for moisture interactions between the rooting zone and groundwater, demonstrating
26	the significance of capillary rise above the saturated storage layer water table in the hydrology of
27	tile drains in loam soils. Novel aspects of this module include the sub-daily time step, which is
28	shorted shorter than most existing models, and which may enable future water quality modules to
29	be added, and the use of field capacity and its corresponding pressure head to provide estimates of
30	drainable water and the thickness of the capillary fringe, rather using than using detailed soil
31	retention curves that may not always be available. An additional novel aspect is the demonstration
32	that flows in some tile drain systems can be better represented and simulated when related to
33	shallow saturated storagewater table dynamics.
34	
35	Keywords: tile drainage, cold regions, hydrological model, capillary fringe, drainable water,
36	saturated storagewater table fluctuations
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39	1. Introduction
40	
41	Harmful algal blooms and eutrophication in large freshwater lakes surrounded by agricultural
42	lands are major environmental challenges in Canada and globally. The transport of nutrients,

43 particularly phosphorus, in runoff from agricultural fields into rivers, ponds and eventually

44 lakessurface water is an important contributor to the increased frequency of algal blooms being

45 experienced in North America and elsewhere (Sharpley et al., 1995; Correll, 1998; Filippelli, 2002;

46 Ruttenberg, 2005; Schindler, 2006; Quinton et al., 2010; Costa et al., 2022). Although nNutrient 47 transport from agricultural fields can occur via both surface runoff and tile drainage (Radcliffe et 48 al., 2015), and recent increases in the frequency and magnitude of algal blooms in Lake Erie in 49 North America have been attributed to tile drainage (King et al., 2015; Jarvie et al., 2017). Tile drain systems lower the seasonally high waterhigh-water tables in poorly drained fields, reduce 50 51 the retention time of soil water, lessening waterlogging in fields and improving both crop growth 52 and field trafficability for farmers (Cordeiro and Ranjan, 2012; Kokulan et al., 2019a). However, 53 they are also important pathways for dissolved nutrients and particulate nutrients material 54 (Kladivko et al., 1999; Tomer et al., 20152003). In Alberta, tile drains have also been used to 55 address salinity issues (Broughton and Jutras, 2013). It has been estimated that 14% of farmlands 56 in Canada (ICID, 2018) and 45% of fields in Southern Ontario, Canada (ICID, 2018; Kokulan, 57 2019) are drained by tile systems. In Alberta, tile drains have also been used to address salinity 58 issues (Broughton and Jutras, 2013). Given their importance in hydrological budgets and 59 biogeochemical transport, there is a need to understand the controlling mechanisms of water and 60 nutrient export from tile systems as an integral part of the broader, modified hydrological system. 61 The ability to integrate a dynamic quantification of tile drainage from fields in hydrological models 62 can help understand the relative importance of this human-induced process as it interplays with an 63 array of other phenomena, including energy and physical mass balance hydrological processes, 64 climate change, and the impacts of modified land management practices on runoff and nutrient 65 export.

66 There are several models that can represent tile drainage, <u>controlled tile drainage and</u>
67 <u>surface runoff in different soil types at the small basin scale, which mainlytypically calculate the</u>
68 <u>amount of gravitational drainage from the soil</u>, such as HYPE (Lindstrom et al., 2010; Arheimer

69 et al., 2015), DRAINMOD (Skaggs, 1978, 1980a; Skaggs et al., 2012), MIKE SHE (Refsgaard 70 and Storm, 1995) and SWAT (Arnold et al., 1998; Koch et al., 2013; Du et al., 2005; Du et al., 71 2006; Green et al., 2006; Kiesel et al., 2010). These models include conceptual components for 72 many key hydrological processes, and but research shows that they have been primarily designed 73 and tested for temperate regions (Costa et al., 2020a). In Canada and other cold regions, some 74 unique hydrological processes such as frozen soil, snowmelt, rain on snow, and runoff over and 75 infiltration into frozen or partially-frozen soils may also be very important (Rahman et al., 2014; 76 Cordeiro et al., 2017; Pomeroy et al., 1998, 2007; Fang et al., 2010, 2013). Many hydrological 77 processes, such as the sublimation of snow, energy balance snowmelt, and infiltration into frozen 78 soils, are strongly affected by temperature and the phase changes of water, which make many 79 existing models developed for warm regions less appropriate for regions with cold seasons 80 (Pomeroy et al., 2007; Pomeroy et al., 2013; Pomeroy et al., 2016; Fang et al., 2010, 2013). Even 81 for temperate regions, the representation of cold season processes is often underrepresented in 82 models (Costa et al., 2020a).

83 Since the use of tile drainage is becoming popular increasing in many cold regions (Kokulan et al., 2019a; OMAFRA, 2023), it has become important to integrate such human-induced 84 85 processes in the specialized hydrological modelling tools that have been developed for these 86 regions, such as the Cold Regions Hydrological Modelling platform (CRHM, Pomeroy et al., 2007; 87 2013; 2022). CRHM was initially developed in 1998 to assemble and explore the hydrological 88 understanding developed from a series of research basins spanning Canada and elsewhere into a 89 flexible, modular, object-oriented, multiphysics platform for simulating hydrological processes 90 and basin response in cold regions (Pomeroy et al., 2007; 2022). The modular CRHM platform allows for multiple representations of forcing data interpolation and extrapolation, hydrological
model spatial and physical process structure and parameter values.

Many existing models typically operate at default daily or monthly time intervals, which is inadequate for the prediction of many short-duration "flashy" hydraulic <u>responses events</u> often observed in tiles (Pluer et al., 2020; Vivekananthan, 2019; Vivekananthan et al., 2019; Lam et al., 2016a, 2016b; Macrae et al., 2019). Indeed, the ability to simulate shorter time intervals (e.g., hourly) facilitates the ability to capture both the rising and falling limbs of tile flow hydrographs, as well as the magnitude of peak flows, both of which are important to tile drain chemistry and export (Rozemeijer et al., 2016; Williams et al., 2015, 2016; Macrae et al., 2019).

100 Hydrological process models such as DRAINMOD, MIKE SHE and SWAT use a 101 combination of empirical and physically based formulations for the simulation of tile flow derived 102 by Hooghoudt (1940), Kirkham (1957), van Schilfgaarde (1974), Bouwer and van Schilfgaarde 103 (1963) and Skaggs et al., (1978). Such formulations contemplate both cases where the water table 104 is below and above the ground surface (Kirkham, 1957). In contrast, simulations of tile drainage 105 in other models such as HYPE use empirically derived recession curves (Eckersten et al., 1994) to 106 simulate tile flow and soil hydrological storage (typically represented as water table). In cases 107 where there is a need for more focus on soil matrix hydrology and less need for understanding 108 hydrological processes at the catchment scale and the relative contribution of tiles (and its 109 interplay), modellers tend to use specialised porous-media PDE-based (partial differential 110 equation-based) numerical models such as HYDRUS (Simunek et al., 2011) and MACRO (Larsbo 111 and Jarvis, 2003).

The amount of water transported by tiles depends on soil moisture dynamics, <u>hydraulic</u>
 <u>gradients</u> and the positioning of the <u>saturated storage layerwater table</u>, which are in turn affected

114 by many factors, including soil type, surface topography and morphology, as well as the local 115 climate and the hydrological characteristics of the field (Frey et al. 2016; Klaiber et al., 2020; 116 Coelho et al., 2012; King et al., 2015). Thus, to provide reliable estimations of water loss from 117 farmland via surface runoff and tile flow, models must be able to predict soil moisture and 118 saturated layer storage and the water table elevation accurately (Brockley, 1976; Rozemeijer et al., 119 2016; Javani-Jouni et al., 2018). Early Many Ssome studies have shown that in some soil types, 120 including silty loam and clay loam soils, the drainable water is less than expected based on the 121 effective porosity (e.g., Skeggs-Skaggs et al., 1978; Raats and Gardner, 1974). Raats and Gardner 122 (1974) have argued that the calculation of drainable porosity requires knowledge of water table 123 elevation and the distribution of soil moisture above the saturated storage layer-water table. Skaggs 124 et al. (1978) added that the calculation of drainable porosity should consider "the unsaturated zone 125 drained to equilibrium with the water table". However, because the soil column is often composed 126 of different soil layers with varying physical characteristics, drainable porosity varies with 127 evapotranspiration rate, soil water dynamics and the depth of saturated water (Logsdon et al., 2010; 128 Moriasi et al., 2013). In a sandy loam soil, Lam et al. (2016a, 2016b) demonstrated that tile 129 drainage was not initiated until soil was at or above field capacity. Williams et al. (2019) observed 130 in the American Midwest that tile drainage was not initiated until the field storage capacity had 131 been exceeded. It has also been shown that despite the presence of tile drains, the soil above the 132 tile may did not always drain all the gravitational water following a rainfall/snowmelt event and 133 the soil may remain at or above field capacity (Skaggs et al., 1978; Lam et al., 2016a). 134 Therefore This means that, the soil drainable water content may be considerably smaller than the 135 storage capacity. This is related to matric potential within the vadose zone, which is driven by the 136 soil characteristics but can also be due to the development of a capillary fringe that reduces the

rate of vertical percolation through the unsaturated zone, reducing tile flow (Youngs, 2012). 137 138 Despite this evidence, some saturated flow models that simulate tile flow overlook the effect of 139 capillary rise and over-estimate the soil drainable water. Other models that represent unsaturated 140 flow (i.e., HYDRUS 3D, Simunek et al., 2011) using Richard's Equation (Richards, 1931) capture 141 the effect of capillary rise and saturation-pressure variation within the soil profile and assess the 142 soil drainable water more accurately. Although the effect of capillary rise is considered in 143 DRAINMOD through the concept of drainable porosity (represented as a "water yield") (Skaggs, 144 1980b), and is calculated for layered soil profiles (Badr, 1978), it requires detailed information 145 surrounding the soil water characteristic curve (Skaggs, 1980b). Although it is indeed optimal to 146 use soil-specific water characteristic curves, Twarakawi-Twarakavi et al. (2009) found that it was 147 is possible to employ average representative values from the soil water characteristic curve to 148 represent soil drainable water where a-soil-specific curves are-was- not available, with some 149 reduction in. and They they found in this case that the model performance was reduced.

150 In this study, a new Tile Drainage Module (TDM) was developed and incorporated within 151 the physically based, modular Cold Regions Hydrological Modelling (CRHM) platform (Pomeroy 152 et al., 2022) to enable hydrological simulations in tile-drained farm fields in cold agricultural 153 regions. As a first iteration, the new module was developed for a field with sloping ground and 154 loam soil with imperfect drainage. Such landscapes are common in the Great Lakes Region (e.g., 155 Michigan and Vermont, USA and Ontario, Canada) and tile drainage in such landscapes has not 156 been as widely studied as it has been in clay-dominated soil. In this module, considerations were 157 explicitly included for the effects of capillary rise and annual fluctuations in saturated storage 158 groundwater water table fluctuations on drainable soil water storage. The use of field capacity and groundwater/soil saturated storage water elevation head (Twarakawi-Twarakawi et al., 2009) to 159

160 modulate soil drainable water across the soil profile, including the capillary fringe region, is an 161 innovative aspect of the model that has been demonstrated to circumvent the need for water 162 characteristic curves. The development of this physically based module provides insight into 163 hydrological processes in tile drainage from sloping landscapes with imperfect drainage, which 164 are increasingly being artificially drained (Cordeiro and Ranjan, 2012; Kokulan et al., 2019a; 165 <u>OMAFRA, 2023</u>).

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- 167 2. Materials and Methods

168 2.1 Study area

169 The study site is a ~10 ha farm field located near Londesborough, Ontario at UTM 17T 466689m 170 E, 4832203m N, shown as LON in Fig. 1a. Mean annual precipitation recorded in this region is 1247 mm (ECCC, 2020). Mean air temperature is 7.2 °C, with annual maxima in July (25.9 °C) 171 172 and minima in January (-10.2 °C), (ECCC, 2020). Soil texture-type has been identified as Perth 173 clay loam (Gr. Br. Luvisolic), with a slope between 0.2 and 3.5%. The field is systematically 174 drained with a tile depth of 0.9 m and a spacing of 14 m (laterals). The tile network collects 175 infiltrated water from about 75% of the field (~ 7.6 ha) but may also receive lateral groundwater 176 flow from neighbouring fields. Water yields from the tile drain laterals (10 cm diameter) are 177 discharged via a common tile outlet (main, 15 cm diameter) below ground. Surface runoff from 178 the field is directed toward a common outlet on the surface using plywood berms installed along 179 the field edge (see van Esbroeck et al., 2016). The tile and surface runoff outlets do not join into a 180 common outlet and are fully separated from one another, even during surface ponding events. The 181 field is a corn-soy-winter wheat rotation with cover drops and rotational conservation till (shallow 182 vertical tillage every three years). Additional details related to farming practices are provided in

183 Plach et al. (2019), soil characteristics are provided in Plach et al. (2018a) and Plach et al. (2018b) 184 and equipment and monitoring are provided in van Esbroeck et al., (2016). The outlets for both 185 surface and tile flow are located at the edge of the field and drain into an adjacent field (Fig. 1b). 186 Water tends to accumulate in a topographic low in the field, in front of the field outlet during 187 snowmelt or high-intensity rainfall events, presumably due to either surface runoff or return flow 188 (see ponded area, Fig. 1b). However, surface water or elevated soil moisture conditions are not 189 observed in this topographic low during smaller events or dry periods of the year, suggesting that 190 this saturated ponding is not in a perennial groundwater discharge zone. Although surface ponding 191 is observed in the topographic depression within the field, water discharges freely at the opposite 192 end of the culvert, facilitating the measurement of flow.

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#### 199 2.2 CRHM: The modelling platform

200 CRHM is a modular hydrological process modelling platform that allows users to select relevant 201 process submodules and apply them as needed to their study. For example, The modular the 202 CRHM platform includes options for empirical and physically based calculations of precipitation 203 phase, snow redistribution by wind, snow interception, sublimation, sub-canopy radiation, 204 snowmelt, infiltration into frozen and unfrozen soils, hillslope water movement, actual 205 evapotranspiration, wetland fill and spill, soil water movement, groundwater flow and streamflow 206 (Pomeroy et al., 2007; 2022). Where appropriate, it is able to calculates runoff from rainfall and 207 snowmelt as generated by infiltration excess and/or saturated overland flow, flow over partially 208 frozen soils, detention flow, shallow subsurface flow, preferential flow through macropores and 209 groundwater flow (Pomeroy et al., 2007; 2022). Water quality nNutrients losses in water can also 210 be simulated in CRHM (Costa et al., 2021), although this was not done in the current study. 211 Modules of a CRHM model can be customized to basin setup, such as delineating and discretizing 212 the basin, conditioning observations for extrapolation and interpolation in the basin, or are process-213 support algorithms such as for estimating longwave radiation, complex terrain wind flow, or 214 albedo dynamics, but most modules commonly address hydrological processes such as 215 evapotranspiration, infiltration, snowmelt, and streamflow discharge. CRHM discretizes basins 216 into hydrological response units (HRU) for mass and energy balance calculations, each with 217 unique process representations, parameters, and position along flow pathways in the basin. HRU 218 are connected by blowing snow, surface, subsurface and groundwater flow and together generate 219 streamflow which is routed to the basin outlet. The size of TDM-HRUs is flexible and can be as 220 small as the size of a single tile pipe (e.g., 1 m) times the pipe spacing (which was 14 m in our case study region), and as large as entire tile networks within a given farm or study area. CRHM
does not require a stream within a modelled basin. The feature allows CRHM to model the
hydrology of cold regions dominated by storage and episodic runoff, such as agricultural fields.

224 Although CRHM has the capability to represent many hydrological and thermodynamic processes, not all processes need/must be represented in all situations. The modular design of the 225 226 CRHM platform enables the user to activate or inactive inactivate specific processes to optimize 227 the model for a particular situation. This is a modelling approach that enables testing different 228 modelling hypotheses and has been pioneered by CRHM and other models, which has inspired a 229 range of hydrological (e.g., SUMMA, Clark et al., 2015a, 2015b), hydrodynamic (e.g., mizuRoute, 230 Mizukami et al., 20152016) and biogeochemical (e.g., OpenWQ, Costa et al., 2023a, 2023b) 231 modelling tools. For example, in the current study, blowing snow was not employed in CRHM as 232 it does not appear to be significant at the study site (periodic snow surveys showed relatively 233 uniform snow cover). Similarly, Ppreferential flow into tile drains was not included in developed 234 for the current simulation. as a Although it it can be is a key process in some clay loam soils, 235 previous studies at the study site have shown that it is not as it does not appear to be a significant the 236 case here driver of preferential flow into tile drainsat the study site, which is a combination of clay-237 loam and silt-loam soils in coarse textured soil (Pluer et al., 2020; Macrae et al., 2019), and was 238 consequently it was not included in our model at this stage. We used hHydrograph analysis 239 (Macrae et al., 2019) and conservative tracer (electrical conductivity and major ions, as well as temperature) over multiple years (Pluer et al., 2020) showed that and found minimal preferential 240 241 flow was minimal at this site as well as other similar sites. For this reason, preferential flow was 242 not included in this study. However, we will certainly continue exploring this transport mechanism 243 in future studies. We did not model preferential flow in this study, and it will be assessed in future

244 studies. Freeze-thaw of soil can occur in the study region, leading to partially frozen soils. 245 However, the extent of freezing varies<del>can differ</del> with snowpack development, winter temperatures and other radiation tive factors. Data collected over an 8-year period at this site found soil freezing 246 247 was restricted to brief periods and such freezing never extended below 10 cm depth (Macrae, 248 unpublished data) which is insufficient for soils to behave as frozen ground for infiltration 249 calculations. Consequently, freeze-thaw processes were also-not deployed in the CRHM model of 250 this siteemployed in the current study. As the model is improved further in the future, Sthese and 251 other model simplifications should be addresseduch processes can be included in future studies to 252 further develop the model. Freeze-thaw processes in soil were also not employed here as there is very little seasonal soil frost in the temperate Great Lakes region due to the persistent snow cover, 253 254 and where soil frost occurs, it is restricted to brief periods and shallow depths (above 10 cm depth) 255 (Macrae unpublished data).

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## 257 2.3 *Observations and input data for the model*

258 Tile flow, water table elevation (saturated storage water table elevation head) and surface flow 259 were measured at the site between Oct. 2011 and Sept. 2018 at 15-minute intervals. It was not 260 possible to install more than one measuring station for water table elevation and soil moisture at 261 the site due to farming activity; consequently, water table elevation head and soil moisture were 262 measured at the approximate midpoint of the field at the edge-of-field. Both tile flow rates and 263 surface runoff were determined using simultaneous measurements of flow velocity and water depths in each of the pipes at the edge-of-field using Hach Flo-tote sensors and an FL900 data 264 265 logger- (Onset Ltd.) (Table A1, Appendix A). Continuous measurements of velocity were included 266 due to the potential for impeded drainage under very wet conditions or caused by the accumulation

of snow and ice around the surface culvert in winter. An additional barometrically-corrected
pressure transducer (U20, Onset Ltd.) (Table A1) was also used for periods when the flow sensors
did not function using a rating curve developed from the depth-velocity sensors; however, it should
be noted that these were for brief periods and the depth-velocity sensor functioned for the majority
of the study. The water table elevation was measured using a barometric pressure-corrected
pressure transducer (U20, Onset Ltd.).

273 Air temperature, wind speed, air relative humidity, incoming solar irradiance and rainfall 274 were also measured at the site at 15-minute intervals and used to forcewere implemented in the 275 model. Variable names and their symbols in CRHM are listed in Appendix B. The air temperature, 276 wind speed and incoming solar radiance measurements were collected 1 m above ground using a 277 Temperature Smart Sensor S-THB-M002, Wind Smart Sensor Set S-WSET-M002 and a Solar 278 Radiation Sensor (Table A1). Rainfall and relative humidity were measured via a tipping bucket 279 rain gauge (Table A1) and an RH Smart Sensor (Table A1). These observations were continuously 280 recorded throughout the study period, except for brief periods of instrument failure and 281 maintenance, when data from nearby stations (Table T1, Supplementary Material) was substituted 282 using the double mass analysis method (Searcy and Hardison, 1960).

Although rainfall was recorded continuously at the field site, snowfall data was not. Snowfall data was obtained from nearby stations (Wroxeter-Davis and Wroxeter, Environment Canada, 20212020), located 31.7 km from the field site. Periodic snow surveys done at the site throughout the study period found that data from the nearby stations was a close approximation of snow at the field site (Plach et al., 2019). Hourly snowfall observations from Wroxeter-Geonor were used for the period between 2015 and 2018, whereas daily data from the Wroxeter-Geonor were used for the 2011 to 2014 period, reconstructed to hourly snowfall time series based on themethod presented by Waichler and Wigmosta (2003).

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#### 92 2.4 Development of the new tile module

293 A Tile Drainage Module (TDM) was developed within CRHM (Figures 2, 3) with the goal of 294 adding the ability to simulate tile flow and the resulting saturated storage (water table) at an hourly 295 time step. CRHM was forced with hourly precipitation, air temperature, solar radiation, wind speed 296 and relative humidity to calculate hydrological states and fluxes in HRUs and the basin. The model 297 requires parameterizations that specify the hydraulic and hydrological properties of the soil, 298 including its thickness, saturated hydraulic conductivity (K), and surface cover. CRHM calculates 299 water storage and fluxes between HRUs, as well as vertical fluxes amongst different hydrological 300 compartments (within each HRU) that include snow, depressional storage, different soil layers, 301 and groundwater.

302 Using the simulation of soil moisture (including both saturated and unsaturated soil 303 moisture) performed by the original CRHM "Soil" module, TDM calculates the dynamic tile flow 304 rate that, in turn, feeds back to soil moisture at each time step. The presence of a capillary fringe 305 (sometimes referred to as the tension-saturated zone within the soil profile) and its effects are 306 considered by limiting the amount of drainable soil water. TDM uses site-specific information 307 regarding the tile network, such as tile depth, diameter and spacing. Information regarding site-308 specific details regarding tile depth, diameter and spacing may be obtained directly from 309 landowners or can be estimated based on standard design and installation guidelines for the region. 310 This information was used to set up the model together with parameterization to translate the 311 hydrological effects of the soil capillary fringe (CF), if present, through two variables, CF

thickness and CF drainable water (discussed in Section 2.5, Figures 2, 3). These two variables are

313 used to limit the fraction of the soil moisture that can freely drain to the tiles.

314

#### B15 2.4.1 Soil moisture and <u>saturated storage</u>water table elevation

316 The TDM uses the water quality soil module or soil module (WQ\_soil or Soil), which divides the 317 soil column into threewo- layers: a recharge layer where evapotranspiration and root uptake 318 generally take place, and a deeper layer that connects to the groundwater system and a deeper 319 groundwater layer that is always saturated. Since CRHM's state variable for soil moisture in the 320 upper two layers is soil water storage volume (Fig. 2), the model results were converted into water 321 table elevation above the semi-permeable layer (Table B1, Appendix B; see-Fig. 2b) for 322 comparison with water table observations,) by dividing volumetric soil moisture content (Table 323 B1) by soil porosity (Table B1) for the cases with no capillary fringe above the water table. 324 Additional steps were taken for periods when a capillary fringe developed (discussed below).

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# 326 2.4.2 Capillary fringe and drainable water

327 Soil moisture in the capillary fringe is equal to the average volumetric water content at capillary fringe  $(\theta_c)$  which is usually greater than the field capacity  $(\theta_{fc})$  (Bleam, 2017, Sect. 2.4). 328 329 Therefore, while the positioning of the capillary fringe responds dynamically to the matric 330 potential, the saturation profile within the capillary fringe remains constant, as well as its thickness 331 because it only depends on the pressure head (capillary forces) that are related to the grain size distribution and field capacity  $(h_{fc})$  as introduced by Twarakawi Twarakawi et al. (2009). 332 333 Therefore, the drainable water in the capillary fringe becomes the difference between saturation  $(\theta_s)$ , computed dynamically in CRHM, and  $\theta_c$ , which corresponds to the water held by capillary 334

335	forces at the capillary fringe moisture content (Fig. 2). Accordingly, Fig. 2 shows the schematic
336	soil characteristic curve for the three water level conditions contemplated in the model.
337	1. Condition $1$ is when the water table is at the surface and the soil is completely saturated
338	(matric potential $= 0$ );
339	2. <i>Condition 2</i> is when the water table drops but the upper boundary of the capillary fringe
340	is at the soil surface; and
341	3. <i>Condition 3</i> is when the water table drops further, and the upper boundary of the capillary
342	fringe drops beneath the surface.
343	In essence, the soil is completely saturated ( $\theta_s$ ) in <i>Condition 1</i> . Between <i>Conditions 1</i> and 2, the
344	capillary fringe occupies the entire soil column above the water level; thus, it can only release the
345	volume of water corresponding to $\theta_s$ - $\theta_c$ or $\varphi_c$ (dimensionless). Between <i>Conditions 2</i> and <i>3</i> , two
346	layers with distinct hydraulic characteristics develop: (1) the top one at $\theta_{fc}$ that releases water up
347	to $\theta_C - \theta_{fc}$ , and (2) the lower one that corresponds to the capillary fringe and can release up to the
348	volume of water corresponding to $\theta_s$ - $\theta_c$ or $\varphi_c$ .
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Drained water when the water table position is changed:

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Figure 2. Schematic representation of the capillary fringe above the water table assuming a 1-m thickness (for demonstration purposes). The soil characteristic curves are shown for the three water level conditions considered: water level at the (1) surface, (2) intermediate depth, and (3) deeper depth. Two transitional drops can be seen in the characteristic curves, one from saturation  $\theta_{s}$  to capillary fringe water content ( $\theta_{c}$ ) (between *Conditions 1* and 2) and one from  $\theta_{c}$  to field capacity ( $\theta_{fc}$ ) (between *Conditions 2* and 3). The coloured areas (green and blue) of the right panel correspond to the amount of water that can be released between *Conditions 1* and 2 (blue) and between *Conditions 2* and 3 (green).

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#### 358 2.4.3 Tile flow calculation

A modified version of the Hooghoudt equation was used to calculate tile flow in the TDM (Smedema et al., 2004). This, which presumes no surface ponding, an assumption that generally holds at the study site (Eq. 1), where water ponds only during very wet periods and on a small portion of the study site (see Fig. 1b). Hooghoudt's equation (Hooghoudt, 1940) is a steady state, physically based equation for saturated flow toward the tile drain. Flow estimates are provided based on the hydraulic conductivity of the soil and water table elevation above the tile pipe. It allows different saturated hydraulic conductivities for the layers above (AL) and below (BL) the tile (Fig. S1). At the study site, soil surveys have reported almost the same soil type (Loam) down to the depth of 90 cm (*e.g.*, Van Esbroeck et al., 2016; Plach et al., 2018b), which was parameterized in the model set up as,

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B70 
$$q = \frac{8 \times K_2 \times d \times h}{L^2} + \frac{4 \times K_1 \times h^2}{L^2},$$
 (1)  
371

where  $K_1$  and  $K_2$  are respectively the saturated hydraulic conductivity in the upper and lower layers in mm h<sup>-1</sup>; *L* is the tile spacing in mm; *h* is the water table elevation above the tile in mm, *d* is the lower layer thickness in mm (Fig. S1), and *q* is the predicted tile flow in mm h<sup>-1</sup>. The only variable that is dynamically updated by CRHM is *h*. Equation (1) is was used to estimate the tile flow rates in TDM, using saturated storage to estimate *h*.

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#### 378 2.4.4 Calculation of the effect of tile flow on soil moisture and water levels

The simulated tile flows (see Sect. 2.3.3) weare subtracted from the soil moisture. To calculate saturated storage (water table or groundwater elevation head level) from soil moisture calculated by the model, a threshold soil moisture content  $(sm_t)$  is defined, which consists of drainable water in the soil ( $\varphi_c$ ) when the upper boundary of the capillary fringe is at the surface (*Condition 2*, Fig. 2) and was calculated as:

$$885 \qquad sm_t = sm_{max} - (C_t \times \varphi_c) \quad , \tag{2}$$

where  $sm_{max}$  is the maximum soil moisture and  $C_t$  is the capillary fringe thickness in mm. However, since the hydrological conditions of the soil are markedly different between the two transitional situations described in Sect. 2.3.2 and Fig. 2 (*Condition 1* to 2 and *Condition 2* to 3), a step function was deployed for determination of <u>saturated storage</u> the water table elevation: 391

$$B92 \quad SSWT = \begin{cases} \frac{sm_t - \left(C_t \times \left((\varphi_s - \varphi_c) + \theta_{fc}\right)\right)}{\varphi_s + \theta_{fc}} + \frac{sm - sm_t}{\varphi_c} & \text{, if between Conditions 1 and 2} \\ \frac{sm_{max}}{\varphi_s + \theta_{fc}} - \left(\left(\frac{sm_t - sm}{\varphi_s}\right) + C_t\right) & \text{, if between Conditions 2 and 3} \end{cases}$$
(3)

393

where *WT* is water table elevation (or soil-<u>SS</u> is saturated storage, <u>SSS</u>)\_-in mm from the bottom of the soil, and *sm* is soil moisture (both saturated and unsaturated storage) in the given time step in mm. <u>Water table observations were used to estimate SS from the field.</u> Equation (3) is determined based on soil moisture curves in Fig. 2 and water level *Conditions 1-3* discussed in Sect. 2.3.2. In Fig. 2, the first and second parts of Eq. (3), which refer to *Conditions 1* to 2 and 2 to 3, respectively, correspond to the volumes of soil water highlighted in "blue" and "green."

400

# 401 2.4.5 Lower semi-permeable soil layer and periodicity in annual groundwater levels

This model application focused on the study site field without including other adjacent areas. This was possible because years of field monitoring at this site have demonstrated that there is no observable surface flow into the site from adjacent <u>farmsfields</u>. The tile network is restricted to the field and is not connected to tile drains or surface inlets in adjacent fields. However, field soil water table observations show evidence of annual groundwater level periodicity/fluctuation (Rust 407 et al., 2019) that are sinusoidal in nature and cannot be neglected. Some studies predict the annual 408 groundwater oscillations or the annual responses of groundwater to precipitation by using sine and 409 cosine functions (De Ridder et al., 1974; Malzone et al., 2016; Qi et al., 2018). De Ridder et al. 410 (1974) studied the design of the drainage systems and described the seasonal groundwater 411 fluctuations observed in wells using sinusoidal curves. Malzone et al. (2016) used a sine function 412 to predict annual groundwater fluctuations in the hyporheic zone. Qi et al. (2018) and Rust et al 413 (2019) used a cross-wavelet transform, consisting of the superposition of sine and cosine curves, 414 to predict shallow groundwater response to precipitation at the basin scale. This approach, using 415 the sine function, was used in this application to simulate annual fluctuations in saturated 416 storage<del>groundwater water table</del>, in Eq. (4), over a period of 1 year, with minimums around the 417 middle of the growing season (mid-July), and maximums in the cold season (early February). This 418 translates into the lower higher greater matric potential, with soil moisture depletion, during the growing season, coinciding with soil moisture depletion, and lower matric potential, with soil 419 420 moisture increases, then during the non-growing season, an elevated matric potential coinciding 421 with an increase in the soil moisture, consistent with field observations. Thus, a sine function representing the annual fluctuations in percolation rate from soil to groundwater  $(G_{y,i})$  layers in 422 CRHM, through the lower soil semi-permeable layer (in mm hr<sup>-1</sup>) is defined as: 423

424

425 
$$G_{y,i} = \left[A \times sin\left(\frac{(T_s - D_d \times 24) \times 360}{24 \times 365.25}\right) - B\right] \times f_{y,i}$$
 (4)  
426

where  $T_s$  is the time step number,  $D_d$  is a time delay in days, A is the amplitude of the water tablesaturated storage (WTSS) fluctuation, and B is an intercept factor.  $f_{y,i}$  is a seasonal factor. The sine function coefficient ( $D_d$ , A, and B) and seasonal factor were adjusted for the whole period and for each year through model verification and shown in Table 1. Appendix C provides more
details on the implementation of Eq. (4). <u>Although this is a simplification of the entire groundwater</u>
<u>system dynamics, it was needed here to provide a more controlled basis for testing the new module</u>
<u>at the field scale before expanding it to larger areas in future work.</u>

434

## 435 2.5 Model application and multi-variable, multi-metric validation

436 The study site is a relatively small field, and 2 HRUs were sufficient to capture its hydrological 437 dynamics in CRHM. The HRUs represent (1) the area immediately upstream of the outlet where 438 surface ponding occurs (depression storage); and (2) the remaining field (Fig. 3). The maximum 439 ponding capacity of HRU 1 was estimated using the spatially distributed hydrodynamic model 440 FLUXOS-OVERFLOW (Costa et al., 2016, 2020b). The CRHM model with its new TDM module 441 wereas set up using the information described in Table 1. Soil textures at the LON site measured 442 in a 25 m grid across three soil depths (0-25 cm, 25-50 cm, and 50-100 cm) averaged 29% sand, 443 48% silt, and 23% clay (Ontario Ministry of Agriculture, Food and Rural Affairs Soil Team, 444 unpublished data). This soil grain size distribution corresponds with a soil-soil-saturated hydraulic conductivity of ~ 0.56 cm h<sup>-1</sup> (~ $10^{-2.5}$ ) (Garcia-Gutierrez et al., 2018), which was implemented 445 446 in CRHM (0.5 cm h<sup>-1</sup>), corresponding to a field capacity of 0.04 (volumetric water content) and  $h_{fc}$  of ~0.8 m (<del>Twarskawi <u>Twarskavi</u> et al., 2009</del>, based on a drainage flux of 0.1 cm d<sup>-1</sup>). 447

448

A robust multi-variable, multi-metric model evaluation strategy was deployed to verify the
capacity of the model to predict tile flow and its impact on the local hydrology. The outflows
examined were tile flow, surface flow, and <u>saturated storagewater table depth</u>. The multi-metric
approach contemplated five different methods, namely the Nash-Sutcliffe efficiency (*NSE*), Root-

453 Mean-Square Error (RMSE), Model Bias (Bias), Percentage Bias (Pbias), and RMSE-observation 454 standard deviation ratio (RSR). These methods were used to assess model accuracy. See Appendix 455 C for more details about the methodology used. It is generally assumed that NSE > 0.50,  $RSR \leq$ 456 0.70, and *PBias* in the range of  $\pm 25\%$  are satisfactory for hydrological applications (Moriasi et 457 al., 2007). Five different metrics were used to evaluate model accuracy in order to describe 458 different aspects of the discrepancies between simulated and observed values. For example, Bias 459 reveals the positive or negative general deviations of simulated values from the observed values, while RMSE shows the average absolute differences between them (Moriasi et al., 2007). Hourly 460 461 values were used in these calculations, which departs from the daily and monthly analyses typically 462 reported for these types of models. Although the hourly timestep is challenging for this sort of 463 simulation, it is an important advance forward toward more detailed, accurate, and advanced 464 models for tile-tile-drained agricultural fields. For example, Costa et al., (2021) noted that the 465 successful extension of hydrological models to water quality studies relies on their ability to 466 operate at small time scales in order to capture intense, short-duration storms that may have a 467 disproportional impact on the runoff transport of some chemical species such as phosphorus – in 468 essence, to capture hot spots and hot moments for flux generation.

469

#### 470 Table 1. Key model parameters in CRHM for representation of the LON site.

Model Parameter	Value	Unit Source	Adjusted/Calibrated	Comment
Soil depth or Soil thickness, $T_{SL}$	2	m	No	Assumed
Semipermeable layer depth	3	m	No	Assumed
Tile depth	0.9	m	No	Farmer/Blueprints
				of the field
Corn root depth	0.5	m	No	Online sources

Soil recharge zone thickness	0.5	m	No	Based on the root
				depth
Tile spacing	14	m	No	Farmer/Blueprints
				of the field
Soil porosity (soil drainable water)	0.045		Yes	Adjusted
$arphi_s$				
Saturated Hydraulic conductivity,	5	mm h <sup>-1</sup>	Yes	Adjusted
$K_s K$ in lower soil layer				
$K_{s}$ <i>K</i> in upper soil layer	5	mm h <sup>-1</sup>	Yes	Adjusted
Capillary fringe thickness, $T_{CF}$	0.8	m	Yes	Adjusted
Capillary fringe drainable water, $arphi_{\mathcal{C}}$	0.03		Yes	Adjusted
Surface depression close to farm	35	mm	Yes	Calculated
surface flow outlet (HRU2)				
Surface depression in rest of the field	0	mm	No	Calculated
(HRU1)				
Surface area of HRU1	79000	m <sup>2</sup>	No	Field observations
				and DEM
Surface area of HRU2	1000	m <sup>2</sup>	No	Field observation
				and DEM
Soil module name in CRHM	WQ_soil		No	
Infiltration module name in CRHM	GreenAmpt		No	
Soil type in GreenAmpt module	5		Yes	Adjusted
Saturated K in GreenAmpt module	6	mm h <sup>-1</sup>	Yes	Adjusted
Soil wilting point	0.025		Yes	Adjusted
A, in sine function	0.025	mm h <sup>-1</sup>	Yes	Adjusted
<i>B</i> , in sine function	-0.005	mm h <sup>-1</sup>	Yes	Adjusted
$D_d$ , in sine function	15	d	Yes	Adjusted
f2012,2 (Seasonal factor, sine function)	2.0		Yes	Adjusted
$f_{2015,2}$ (Seasonal factor, sine function)	1.8		Yes	Adjusted
f2016,2 (Seasonal factor, sine function)	2		Yes	Adjusted



Figure 3. a) Schematic conceptual view of the CRHM model configuration, including soil layers, water tablesaturated storage
(WTSS), groundwater, and tile flow.; and b) soil profile, including the capillary fringe and its location relative to the soil and tile.

# 479 **3. Results**

A multi-variable, multi-metric model evaluation approach was deployed to verify the capacity of
 the model to predict not only tile flow but also the effects it has on the local hydrology, from
 surface to sub-surface processes. The outflows examined were tile flow (Section 3.1), saturated
 storagewater table depth (Section 3.2), and surface flow (Section 3.3). The multi-metric approach

484 <u>contemplated five different methods, namely the Nash-Sutcliffe efficiency (NSE), Root-Mean-</u>
 485 <u>Square Error (RMSE), Model Bias (Bias), Percentage Bias (Pbias), and RMSE-observation</u>
 486 <u>standard deviation ratio (RSR).</u>

487 *3.1 Tile flow* 

The model was able to capture most tile flow events, both in terms of the timing and magnitude of peak flows and the most important seasonal patterns (Fig. 4). For example, the almost complete absence of tilenear absence of flow during the growing season (May to September) was captured. The simulated flow peaks generally had a good agreement with observations, as well as the low flow or base flows during cold periods (December-March). The ascending and descending limbs of the response signal were also adequately predicted.

494

495 Results show that tile flows generally occurred during snowmelt events, as indicated by the 496 synchrony between snow water equivalent (SWE) depletion and tile flow. The maximum 497 snowpacks (or snow water equivalent, SWE) were markedly smaller during the winters of 2016 498 and 2017 when compared with those of 2013 to 2015. However, this did not necessarily translate 499 into lower tile flows as precipitation also occurred as rain during these seasons. Although the 500 magnitude of peak tile drainage flow peaks was not always predicted accurately, the model was 501 able to capture the annual trends of both an absence of tile flow during the summer months 502 (growing season) and the ascending and descending limbs of the tile hydrograph during events 503 (Figure 4).



tile flow, groundwater flow, and matric potential that affect the drainable water from the capillary

517 fringe. This caused flashier storage responses above the tile that were captured well by the model. 518 In contrast, tiles did not withdraw water from the soil layer below the tile pipe and thus did not 519 control water table fluctuations in saturated storage when levels were below the drain pipe, and 520 tile drains did not flow during such periods. During the growing season, both the observed and 521 simulated water table (or saturated storage) dropped abruptly because of the seasonal lowering of 522 the regional groundwater water table. In the growing seasons of 2012, 2015 and 2016, which were 523 dry years, large declines in the water table and saturated storage were observed, whereas in wetter 524 years such as 2013 and 2014, seasonal saturated storage water level declines were smaller. The 525 seasonal declines in saturated storage water level during the growing season led to a cessation in 526 tile flow in most years (Fig. 4, 5), even following rainfall events. For example, there was a large 527 precipitation event (~35 mm) in the growing season of 2016 that did not produce tile flow (apparent 528 in both model and observations).



Figure 5. Time series of the simulated <u>and observed</u> saturated storage <u>and (or observed water table)</u> in the soil or groundwater layers of the model along with the observed temperature and precipitation. Given that tiles do not flow when the WT is below them, the WT = 0 when the water table elevation is at the depth of the tile drainpipe. In the figure, the water table is measured as the elevation above (+) or below (-) the tile pipe.

#### 535 3.3 Surface flow and total flow

536 The model was not always able to capture the observed surface flow as satisfactorily as it captured 537 tile drainage (Fig. 6a). Some possible reasons are uncertainties in the measurements of surface 538 flow due to ponding in surface depressions on the field, which impeded the drainage of some of 539 the surface runoff prior to when it exitinged the field through the culvert (see Fig. 1), or uncertainty 540 in field estimates of SWE. However, the model performance improvesd considerably when both 541 runoff and tile flow weare combined (referred to as total flow, Fig. 6b). Indeed, most of the flow 542 from the field was through tile drains (80% in 5-year average) rather than surface runoff (20% in 543 5-year average, Plach et al., 2019). The underestimation of both cumulative total and surface flows 544 during 2017 and 2018 is possibly due to the removal of the blockage in the tile pipe in early 2017, 545 which may have affected both surface and tile flow. The differences in timing of the simulated and 546 observed surface flow for many of the main events (Figure 6) shows that there remain systematic issues in simulation of surface flow by CRHM which should be addressed in future research. 547



548 Figure 6. Observed and simulated cumulative surface flow (a) and total flow (b).

#### 550 3.4 Overall model performance

551 The model performance was calculated based on hourly data for various model outputs (Table 2). 552 To compare the performance of the model in different seasons we calculated the coefficient for 553 wholeentire year as well as separately for the growing and non-growing seasons separately. The 554 results confirm that for the whole year the model is robust over an annual cycle in the sense that it 555 can capture the main patterns of tile flow, surface flow, and saturated storagewater table elevation. The Pbias values are below 25% for most of the fluxes and cumulative fluxes. The RSR values are 556 557 also generally below 1.0. The NSE values are positive and above 0.3 for most fluxes, except for 558 surface flow, where the model exhibited some difficulties. The weaker performance of the model 559 in the simulation of surface flow, which is illustrated by the NSE coefficient, can be partly related 560 to difficulties in measurement of surface flow during flooding, ponding, and freeze-and thaw on 561 the surface. The performance coefficients were calculated for, the growing season, May-September (Table 2b) and non-growing season, October-April (Table 2c). The results shows that 562 563 surface flow biases are significantly larger and negative in May-September and are smaller and 564 positive during October-April. For tile flow the Biases biases are a bitslightly higher in May-September while whereas for saturated storageoil water table and total flow the biases are a 565 566 bitslightly lower in May-September. The NSEs are more acceptable in October to April for surface 567 flow, tile flow and total flow, but the NSE for <u>WT-SS</u> is more acceptable in May-September. The 568 overall performance of the model for both tile and surface flow is more reliable in the non-growing 569 season, when the regional water table was above the tile and saturated storage the water level fluctuations was were mainly controlled by tile flow rather than regional groundwater oscillations. 570 571

572 Table 2. Performance coefficients for surface flow, tile flow and water tablesaturated storage (WTSS/SSS), as well as total (tile +
573 surface) flow, for the simulation period of October 2011 to January 2018. The coefficients were calculated for both hourly and

574 daily flow rates, for the whole year (a) for May to September (b) and for October to April (c). (Green and red color show the

575 seasonal coefficients improved and worsened, respectively, -compared to their seasonal values).

576

a) Coefficients for whole year

Performance	Surface	Tile flow	₩T	Total	
coefficients	flow		<del>(m)<u>SS</u></del>	flow	
NSE*	-2.29	0.31	0.49	-1.38	(ir fo C
RMSE <sup>^</sup>	0.27	0.08	0.26	0.30	pefficie r hou 1m h <sup>-1</sup> )
Bias <sup>#</sup>	0.54	0.24	0.14	0.28	nts c rly fl
Pbias <sup>\$</sup>	21.77	17.91	10.46	18.63	calcula ow re
RSR <sup>&amp;</sup>	1.82	0.83	0.71	1.54	ited
NSE	-0.73	0.29	0.50	0.01	for (m
RMSE	2.04	1.72	0.24	2.92	efficie · daily m d <sup>-1</sup> )
Bias	0.35	0.20	0.09	0.22	ents of flow I
Pbias	35.11	19.63	9.33	21.73	alcula ates
RSR	1.31	0.84	0.70	0.99	ited

577

578 Б79

b) coefficients for May to September

Performance coefficients	Surface flow	Tile flow	<del>WT</del> ( <del>m)<u>SS</u></del>	Total flow	
NSE*	-18.98	0.19	0.40	-11.76	Co for (m
RMSE <sup>^</sup>	0.26	0.03	0.12	0.26	efficie hou m h <sup>-1</sup> )
Bias <sup>#</sup>	-1.43	0.49	0.03	0.11	nts c rly fl
Pbias <sup>\$</sup>	-142.79	48.88	3.44	10.96	ow r
RSR <sup>&amp;</sup>	2.85	0.57	0.39	2.27	ited
NSE	-3.89	0.21	0.41	-1.08	Co for (m:
RMSE	1.39	0.73	0.11	1.66	efficie daily m d <sup>-1</sup> )
Bias	-1.43	0.49	0.02	0.11	nts c flow r
Pbias	-142.79	48.88	2.07	10.96	alcula ates
RSR	1.41	0.56	0.39	0.92	ted

Performance coefficients	Surface flow	Tile flow	<del>WT</del> ( <del>m)<u>SS</u></del>	Total flow	
--------------------------	-----------------	-----------	---	---------------	--

NSE*       -0.37       0.24       0.20       -0.04         RMSE^       0.11       0.07       0.21       0.14         Bias#       0.87       0.14       0.11       0.24         Pbias\$       86.59       13.56       11.00       24.11         RSR&       0.90       0.67       0.77       0.79	Coefficients ca for hourly flov (mm h <sup>-1</sup> )
RMSE^       0.11       0.07       0.21       0.14         Bias#       0.87       0.14       0.11       0.24         Pbias <sup>\$</sup> 86.59       13.56       11.00       24.11         RSR <sup>&amp;</sup> 0.90       0.67       0.77       0.79	hourly flov n h <sup>-1</sup> )
Bias#       0.87       0.14       0.11       0.24         Pbias <sup>\$</sup> 86.59       13.56       11.00       24.11         RSR <sup>&amp;</sup> 0.90       0.67       0.77       0.79	nts ca ly flov
Pbias <sup>\$</sup> 86.59         13.56         11.00         24.11           RSR <sup>&amp;</sup> 0.90         0.67         0.77         0.79	ia کر
RSR <sup>&amp;</sup> 0.90 0.67 0.77 0.79	v ra
	ites
NSE -0.11 0.26 0.24 0.18	(m for G
RMSE 1.50 1.56 0.21 2.40	daily : n d <sup>-1</sup> )
Bias 0.87 0.14 0.11 0.24	nts c flow ra
Pbias 86.59 13.56 10.58 24.11	ates
RSR 0.81 0.67 0.75 0.70	led

582

583

584

\* Nash-Sutcliffe efficiency, 'Root-Mean-Square Error, \*Model Bias, \*Percentage Bias, \*RMSE-observation standard deviation ratio
 587

# 588 3.5 Presence of capillary fringe: effects and hypotheses

589 Results show that the thickness and vertical positioning of the capillary fringe have-had a strong 590 impact on the amount of drainable soil water that can-flowed into the tiles. To investigate this 591 effect further, the response of tile flow and soil moisture to changes in the capillary fringe was 592 examined. It should be noted that although this thickness may change slightly depending on the 593 soil type and water retention curves (Skaggs et al., 1978), the model assumed a constant value 594 given the field-scale nature of the simulations and myriad of processes contemplated. However, 595 despite the simplification, the vertical positioning of the capillary fringe was still calculated and 596 enabled a dynamic (time-dependent) calculation of the drainable soil water that was available for 597 tile drainage over time.

#### 599 *Effect of capillary fringe on tile flow*

Figure 7a relates the simulated normalized total cumulative tile flow ( $Q_{tR}$ , total tile flow divided 600 601 by the total tile flow when there is no influence of capillary fringe) to capillary fringe drainable water ( $\varphi_{cR} = \varphi_c/\varphi_s$ ) for two different  $\varphi_s$  values (0.045 and 0.125). The values were normalized 602 603 (0 - 1 scale) for comparison purposes. As expected, the model indicates that tile flow increases 604 with drainable water, but the relationship is non-linear, likely because as tile carrying capacity is 605 exceeded more frequently, there is more opportunity for groundwater seepage and 606 evapotranspiration. The direct effect of  $\varphi_s$  (comparing the solid and dashed lines) on tile flow is 607 small because the amount of water that can effectively drain to the tile is controlled by the capillary 608 fringe and the associated drainable soil water. Figure 7b looks at the impact of the capillary fringe 609 thickness on tile flow. Here, the values are also normalized. Results show that  $Q_{tR}$  decreases with increasing normalized thickness of the capillary fringe,  $T_{CFR}$  ( $\frac{T_{CF}}{D_t}$ , capillary fringe thickness 610 divided by tile depth), but only while the  $T_{CFR}$  is less than 1 that is when the capillary fringe 611 612 position is above the tile but has not reached the soil surface. Beyond this point, increments in the 613 capillary fringe thickness have no impact on tile flow because *Condition 1* has been reached (see 614 Fig. 2), which essentially means that the capillary fringe has reached the soil surface. The match between the curves for two different  $\varphi_s$  values shows that the changes in  $\varphi_s$  does not influence the 615 616 effect of normalized capillary fringe thickness and drainable water on normalized tile flow. In 617 Appendix D the sensitivity of cumulative tile flow and mean saturated storage soil water table 618 elevation to different parameters are shown along with general approaches for evaluation of the 619 model parameters for new sites, the site with no tile flow and water table observations.



Figure 7. Comparison between normalized tile flow  $(Q_{tR})$  and (a) normalized drainable soil water  $(\varphi_c/\varphi_s)$  and (b) capillary fringe thickness  $(T_{CFR})$  for different maximum soil saturation values  $(\varphi_s)$ , by drawing the model prediction lines.

#### 623 *Effect of capillary fringe on soil moisture*

624 Observations and simulations of saturated storage model results of WT reveal a bimodal frequency 625 distribution (Fig. 8 and 9, respectively) with peaks at 0.85 m and 1.25 m depth, with the former 626 corresponding to the influencedepth of the tile pipe and the second peak reflecting that from the 627 capillary fringe thickness. In I simulated soil saturated storage (SSS as a measure of WT) frequency 628 distributions (Fig. 9), show a the first peak that highlights again the efficiency of the tile in 629 removing soil moisture. In contrast, the second peak indicates a strong model response to 630 differences in the capillary fringe thickness. It shows that when there is near-constant percolation 631 from the bottom of the soil layer, the matric potential varies the greatest while it remains between 632 the tile depth and the soil surface. While the saturated storagewater table fluctuates faster and is 633 more unstable within this range, it also remains there for shorter periods. This bimodal response 634 tends to push the saturated storage layer water table depth below the tile. In Figure 9, we can see 635 that the first peak happens at 0.9 m depth where the tile pipe is located, and the second peak happens at the depth equal to capillary fringe thickness. In Figure 9 the second peak is more clearer 636

for the capillary fringe thickness of more than 1000 mm. The first peak in the observed water table
saturated storage frequency plot (Figure 8) happened around 0.8 m which almost matches with the
tile depth. And the second peak happened at the depth of ~1.2 m which shows that the capillary
fringe thickness should be around 1.2 m. But, to have a more reliable estimate for the capillary
fringe, based on Figure 8, data is needed at depths greater than 1.5 m.

642

The bimodal behaviour of the observed water table and and simulated saturated storage (observed saturated storage is the observed water table ) demonstrated here provides the opportunity to quantify the thickness of the capillary fringe using continuously monitored saturated storagewater table elevations. The capillary fringe thickness determined using this method can then be used as an input to the TDM module.



Figure 8. Histogram of the observed <u>saturated storage (or observed water table)</u> distribution for the period pf 2011 to 2018 in LON

651

<sup>650 (</sup>Londesborough).



652

Figure 9. Histograms of the simulated soil saturated storages versus saturated storage depth (SSS or WT) for the capillary fringe thicknesses of 0 (a,b), 400 (c,d), 800 (e,f), 1000 (g,h) and 1400 (I, j) mm and for the  $\varphi_s$  and  $\varphi_c$  of 0.125 and 0.025 (left column)as well as 0.045 and 0.009 (right column).

# 657 4. Discussion

658

659 4.1 Insights into key control mechanisms of tile flow for catchment scale for model simulations 660 The model suggests that tile flow may not be accurately predicted exclusively based on the soil's 661 saturated storage the water table depth and soil saturated hydraulic conductivity as suggested by 662 the steady-state flow assumptions of the Hooghoudt's equation (Hooghoudt, 1940). These results 663 indicate two additional controls: (1) the amount of drainable soil water in the soil, which has also 664 been identified in some field studies (e.g., Skaggs et al., 1978; Moriasi et al., 2013) and (2) 665 fluctuations in saturated storage the groundwater table (GWRD) are important to account for in 666 catchment scale simulations. However, the relationship between drainable water and tile flow rates 667 is non-linear, as demonstrated in Fig. 7a. This is because the residence time for groundwater 668 seepage and evapotranspiration increases when the hydraulic tile carrying capacity is exceeded. 669 Comparatively, the effect of soil drainable water,  $\varphi_s$  (see also Fig. 7a) on tile flow is small because 670 the capillary fringe and associated drainable soil water control the amount of water that can 671 effectively flow to the tile.

672

The verification of the model also indicated that the slopes of the rising and falling limbs of tile flow hydrographs and <u>saturated storage</u> WT were very sensitive to (1) the ratio between  $K_s$ K and drainable soil water; and (2) the net outflow in the soil through tile flow and groundwater level fluctuations <u>in saturated storage (GWRD)</u>. This is supported by previous studies showing rapid responses of tile flow to precipitation events (Gentry et al., 2007; Smith et al., 2015) and others that have related rapid responses in tile discharge to antecedent moisture conditions (Macrae et al., 679 2007; Vidon and Cuadra, 2010; Lam et al., 2016a; Macrae et al., 2019), which can be affected by680 the development of a capillary fringe and its non-drainable water.

681

682 Results show that large fluctuations in saturated storage WT SS (or SSSWT) and tile flow during 683 the cold season, when the water table tends to be above the tile, are primarily triggered by the 684 development of a capillary fringe that reduces the amount of drainable soil water. Model sensitivity 685 tests showed that a small amount of drainable soil water produces steeper rising and falling 686 responses (and with larger fluctuation amplitudes) in both the water table (saturated storage) and 687 the tile flow. Indeed, this pattern can be observed by exploring differences in tile drain responses 688 in clay loam soils with larger field capacities (and correspondingly smaller drainable water) and 689 smaller hydraulic conductivity which are more likely to experience pronounced oscillations (e.g., 690 steeper rising and falling response curves) compared to tile drain responses of sandy soil, which is 691 characterized by reduced capillary forces, lower field capacities (but correspondingly larger 692 drainable water) and higher hydraulic conductivity. Notably, both model and observations of 693 saturated storage WT SS (as a proxy for soil moisture) reveal a bimodal (*i.e.*, two peaks) frequency 694 distribution when examined in relation to the tile depth and capillary fringe thickness (Fig. 8 and 695 9, respectively). The two peaks (i.e. most frequently observed saturated storage WT or SSS 696 conditions) correspond with the (1) depth of the tile pipe (0.75 m), which demonstrates the efficacy 697 of the tile at rapidly removing excess soil water, and the (2) the capillary fringe thickness (for the 698 depths of 1.0 and 1.4 m, Figs. G, h, I and j) beyond which the amount of drainable water above the 699 water table significantly increases.

701 These findings align well with studies such as Lam et al. (2016a) that recorded soil moisture near 702 saturation after tile flow had ceased, suggesting the development of a capillary fringe. Combined 703 experimental and modeling works, such as in Moriasi et al. (2013) and Logsdon et al. (2010), also 704 discuss the impact of drainable soil water ("drainable porosity" or "specific water yield") on tile 705 flow and note that the drainable water is, in turn, dependent on the soil type, soil-water dynamic 706 and water table depth. However, these studies did not explore the dynamic nature of the capillary 707 fringe and its thickness relative to the soil column above in determining the transient amount of 708 drainage soil water that will impact the saturated storage frequency WT distribution and tile flow 709 differently over time (Conditions 1 to 3, see Fig. 2). Herein, while a capillary fringe with a fixed 710 thickness that is generally related to the soil properties was assumed, its vertical positioning was 711 simulated dynamically, which allowed determining the drainable soil water based on the evolution 712 of pressure head corresponding to field capacity. Thus, the development of the TDM has provided 713 a step forward in the modeling of tile drainage and suggests that in loam soils such as those at the 714 study site, the effects of a capillary fringe on tile flow should be included. Soil moisture (soil 715 unsaturated storage) measurements from the study site by Van Esbroeck et al., (2017) between 716 November 2011 and May 2014 from depths of 10, 30, and 50 cm (using EC-5 Soil Moisture Smart 717 Sensor) showed that almost 90% of the gravitational soil moisture drains out with 0.5 to 2.5 h. 718 ThisIt suggests that the saturated storage-water table and capillary fringe can reach an equilibrium 719 condition within one hour at this field site, enabling the us to use of a steady state equation 720 (Hooghoudt, 1940) to predict the dynamic behaviour of the water table fluctuations.

722 4.2 Importance of capturing seasonal patterns in <u>saturated storage groundwater</u> to improve
723 tile flow predictions

724 The saturated storage GWRD changed dramatically between seasons affecting soil 725 moisture (both saturated and unsaturated storage inof the soil) and tile flow patterns. Both 726 observations and model results show that low precipitation and higher evapotranspiration rates 727 tend to produce little tile flow during the growing season. These seasonal patterns in precipitation 728 and evapotranspiration are accompanied by a reduction in soil moisture (both unsaturated storage 729 and saturated storage ) that leads to a substantial storage capacity in fields. Even following 730 moderate and high-intensity storms during the growing season, rapid soil moisture increases awere 731 observed (both saturated and unsaturated soil storage); however, tile flow rarely developeds due 732 to higher evapotranspiration and a seasonal decrease in the saturated storage GWRD, suggesting 733 that the soil is able to hold the water (Lam et al., 2016a; Van Esbroeck et al., 2016). In contrast, 734 tile flow wais often observed during the cold season, with significantly smaller evapotranspiration 735 fluxes, even during smaller rainfall-runoff and snowmelt events because of reduced soil storage 736 but also a seasonal increase in regional groundwater tableGWRD (Lam et al., 2016a; Macrae et 737 al., 2007, 2019; Van Esbroeck et al., 2016). This concurs with several studies throughout the Great 738 Lakes and St. Lawrence region that have reported stronger tile responses during the non-growing 739 season, with the summer months often showing little to no tile flow (Lam et al., 2016a, 2016b; 740 Jamieson et al., 2003; Macrae et al., 2007; Hirt et al., 2011; King et al., 2016; Van Esbroeck et al., 741 2016; Plach et al., 2019).

These results (the controlling effect of soil drainable water and <u>saturated storage</u> groundwater level-fluctuations on tile flow) suggest that while soil moisture (both saturated and unsaturated storage) is largely controlled by tile flow rather than <u>saturated storage</u> GWRD in the

745 cold season, this reverses in the growing season (*i.e.*, soil moisture controls tile flow), with soil 746 moisture (both saturated and unsaturated storage) being also impacted by evapotranspiration. The 747 controlling effect of groundwater fluctuations in the growing season has also been studied by 748 Hansen et al., (2019). The model indicated that the rapid drops in observed saturated storage WT 749 during the growing season could not be explained by evapotranspiration alone as well as the crop 750 root depths, thus pointing to the role of <u>saturated storage GWRD</u>. Johnsen et al. (1995) and Akis 751 (2016) also showed that the effect of groundwater accretion was more effective on tile flows than 752 surface runoff. Also, Vaughan et al. (1999) found that tile drain flows in their study site in San 753 Joaquin Valley of California were better explained and related to nonlocal groundwater appearance than to local variations in irrigation amount, evapotranspiration, variation in water storage or tile 754 755 drain blockage. Thus, it was determined that in that in addition to soil saturated hydraulic 756 conductivity and soil thickness, the seasonal groundwater fluctuations in saturated storage and 757 capillary fringe drainable water are other important controlling factors on tile flow rates.

758

# 759 **5.** Conclusions

760 A new tile drain module within the modular Cold Regions Hydrological Modelling (CRHM) 761 platform has been created and tested at the field scale to support the management of agricultural 762 basins with seasonal snow covers. The model was tested and validated for a small working farm 763 in southern Ontario, Canada, and presents a step forward in the dynamic simulation of tile flow 764 and its effects on the hydrological cycle in cold climates. Observations and model results showed 765 that the dynamic prediction of tile flow and soil moisture at catchment scales needs to account for 766 (1) the amount of drainable soil water that can be affected by the development of a capillary fringe 767 and (2) fluctuations in <u>saturated storagethe groundwater water table</u>, in addition to (3) the typical saturated storage near the tile pipe depth, (3) water table elevation above the tile pipe and (4) the soil saturated hydraulic conductivity considered by the steady-state flow Hooghoudt's equation. The <u>saturated storage groundwater table</u> and matric potential changed dramatically between seasons, affecting patterns of <u>overall</u> soil moisture and tile flow. Observations and model results showed that low precipitation and higher evapotranspiration rates caused minimal tile flows during the crop-growing season. Conversely, tile flow was often observed during the cold season, even during small rainfall-runoff and snowmelt events, due to a seasonal increase in the groundwater

775 table and soil-saturated storage.

776 Model sensitivity tests showed that the capillary fringe strongly affected the amount of drainable 777 soil water flowing into the tile. Tile flow increased with drainable water, but the relationship is 778 highly non-linear likely because, as the tile carrying capacity is exceeded more frequently, there is 779 more opportunity time for groundwater seepage and evapotranspiration. Finally, observations and 780 model results reveal a bimodal soil-saturated storage response in the presence of tiles, which is 781 controlled by the relative positioning of the capillary fringe in relation to the soil surface and the 782 depth of tile drains below the soil surface. Capturing these dynamics is a critical advance enabling 783 the accurate prediction of the swift hydrological changes caused by the presence of tiles in models. 784 The TDM was developed as a first approximation from a single field site with the goal of providing 785 insight into control mechanisms of tile flow. Given this limitation, it is not yet widely applicable 786 across multiple field sites and for larger areas. However Yet, the development of this module 787 provides critical insights into its potential and performance for hourly time-step simulations, as 788 well as the importance of saturated storage regional groundwater table fluctuations and simplifying 789 the capillary fringe parameters within models in some landscape types. Future work will-should 790 include building on the <u>current</u> model and adapting it to for different soil textures, such as those in clay loam soils, where preferential flow can have a strong impact on soil-ssaturated storage and tile flow. Also, explicit representation of unsaturated flow will-may be needed to enable the use of the model in regions where groundwater is disconnected from surface water, as commonly happens in arid and semi-arid regions. Subsequent steps should include also the integration of the new TDM model with CRHM's frozen soil and water quality modules.

796

# 797 Code/Data availability

The tile flow and soil water table data are not publicly available and will be provided upon request to the data owner, Merrin Macrae. TDM code is not completely implemented in the main version of the Cold Regions Hydrological Model platform and is provided <u>only uupon</u> request to the corresponding author.

802

# **803** Author contribution

MK and DC developed the <u>new</u> model code and performed the simulations. MM prepared the data and supported the field work. <u>JP developed CRHM</u>. MK, DC and MM prepared the manuscript with contributions from JP and RP. All authors edited the manuscript.

807

# 808 **Competing interests**

809 The contact author has declared that none of the authors has any competing interests.

810

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- 1288
- 1289
- 1290 Appendix A
- 1291 Table A1. Instrument names and descriptions

Instrument name	Description
Hach Flo-tote and FL900 logger	Flow velocity and water level measurement
U20, Onset Ltd.	Barometrically-corrected pressure transducer
Temperature Smart Sensor S-THB-M002	Air temperature measurement
Wind Smart Sensor S-WSET-M002	Wind speed measurement
(Silicon Pyranometer)-S-LIB-M003	Solar radiation sensor
Tipping bucketrain gauge, 0.2 mm Rainfall	Rainfall measurement
Smart Sensor – SRGB-M002	
RH Smart Sensor(S-THB-M002)	Relative Humidity measurement

- 1293
- 1294
- 1295 Appendix B
- 1296 Table B1. Parameter names and their symbols in CRHM platform

Parameter symbol	Parameter name
Tair	Air temperature

Wspeed	Wind speed
RH	Relative Humidity
Qsi	Incoming solar irradiance
R	Rainfall
WQ_soil	Water Quality soil module
WT	Water table elevation above the semipermeable layer
SS <del>S</del>	Soil saturated storage or the saturated part of the soil moisture
soil_moist	Soil moisture
Poro_soil	Soil porosity
AL	Above layer
BL	Below layer
GWRDK <sub>s</sub>	Groundwater level fluctuations, groundwater recharge and
	dischargeSaturated hydraulic conductivity

1300 Appendix C

1802 Here, it was shown how seasonal factors  $(f_{y,i})$  is assessed for different years. Equation (4) can be

1303 written as:

$$1305 \qquad G_{y,i} = G \times f_{y,i} \tag{C1}$$

|  For each year (y),  $f_{y,i}$  for the first  $(f_{y,1})$  and second  $(f_{y,2})$  part of the sine function (G) were assessed individually. It should be note that in first and second part of the sine function for each year G is larger than zero  $(G \ge 0)$  and smaller than zero (G < 0), respectively. *G* can be defined for the two parts as:

$$\begin{cases} if \ G \ge 0 \ [i = 1] then \ f_{y,1} = x \\ if \ G < 0 \ [i = 2] then \ f_{y,2} = y \end{cases}$$
(C-2)

1313

1B14 G is the sine function representing the annual fluctuations in water tablesaturated storage (WT/SSS) or it can be simply defined as the percolation rate (in mm hr<sup>-1</sup>) of soil water to 1815 groundwater through lower semi-permeable layer. So, for *n* years there are  $n \times 2 f_{y,i}$  values. The 1**B**16 default values for  $f_{y,i}$  are 1 and the default values can be changed for each year and for first and 1β17 second parts in each year independently. Calculated  $G_{y,i}$  in each time step add or subtracted to or 1818 from the total soil moisture depend on its sign. The  $f_{y,i}$  values for the sine function parameters are 1<u></u>319 1320 presented in Fig. C1. The verified sine function time series along with time series of temperature, 1321 precipitation and calculated evapotranspiration are shown in Fig. C1. In this figure it is obvious 1322 that in years 2012 and 2015 to 2017 the warm season amplitudes are larger. The ET values are 1323 happened more in the warm seasons (growing seasons). Also, the seasonal oscillation in sine 1324 function is very similar to the temperature general oscillations.



Figure C1. Time series of the adjustable sine function along with the time serioes of calculated evapotranspiration, temperature and
precipitation during the study period from Oct 2011 to Sept 2018.

1325

#### 1330 Appendix D

A sensitivity analysis was conducted for the cumulative tile flow  $(Q_{tc})$ , mean soil saturated storage (SS) (it is equal to water table elevation,  $(WT_m, as it is mentioned in Eq. 3)$  and cumulative outflow rate from the semi-permeable layer at the bottom of the soil to groundwater  $(G_c)$  (see section 2.4.5, Eq. 4) with respect to six module parameters. Additionally, an approach for assessing model parameters at a new sites, potentially lacking water table elevation and tile flow observations is proposed.

1337

#### 1338 D.1 Sensitivity analysis

1839 In this section, the sensitivity of  $Q_{tc}$ , *SS* and  $G_c$  to six distinct module parameters, namely capillary 1840 fringe thickness ( $T_{CF}$ ), capillary fringe drainable water ( $\varphi_c$ ), soil saturated hydraulic conductivity 1841 (*K*), soil thickness ( $T_{SL}$ ), sine function amplitude (*A*) and sine function (*B*) was examined.  $Q_{tc}$ ,  $G_c$ 1842 and *SS* were computed over the entire simulation period, expressed in units of mm, mm and m,



# 1343 respectively. Figures D-1a to f illustrate these sensitivities, with each parameter's impact discussed

in dedicated sections.

Figure D-1 Sensitivity of cumulative tile flow,  $Q_{tc}$ , cumulative soil to groundwater percolation rate,  $G_c$ , and mean soil water tablesaturated storage elevation, *SS*, to capillary fringe thickness,  $T_{CF}$  (a) capillary fringe drainable water,  $\varphi_c$  (b), soil hydraulic conductivity, *K* (c), soil thickness,  $T_{SL}$  (d), sine function amplitude, *A* (e) and sine function intercept, *B* (f).

1355 D.1.1 Sensitivity to capillary fringe thickness

To gauge sensitivity to capillary fringe thickness  $T_{CF}$ , flow rates and the *SS* were analyzed for  $T_{CF}$ ranging 0 to 1600 mm. Figure D-1a indicates that as  $T_{CF}$  increases, both cumulative tile flow  $(Q_{tc})$ and mean soil water tablesaturated storage (*SS*) decline. The *SS* drop is sharper for  $T_{CF}$  beyond 900 mm. Beyond this thickness,  $Q_{tc}$  stabilizes at a minimal value. A negative *SS* indicates its position below the tile pipe.  $G_c$  remains consistent despite  $T_{CF}$  variations.

- 1361
- 1362 D.1.2 Sensitivity to capillary fringe drainable water

1363 With rising  $\varphi_c$  both  $Q_{tc}$  and *SS* surge (Figure D-1b). As  $\varphi_c$  ascends from 0.005 to 0.45,  $Q_{tc}$  jumps 1364 from 1300 mm to 1900 mm and *SS* from -0.45 m to -0.25 m (Figure D-1b).  $G_c$  stays constant, 1365 irrespective of  $\varphi_c$  fluctuations.

1366

1367 D.1.3 Sensitivity to soil hydraulic conductivity

Increasing soil hydraulic conductivity (*K*) from 0 to 10 mm hr<sup>-1</sup>leads to a surge in  $Q_{tc}$  and a drop in *SS* (Figure D-1c). However, adjusting *K* from 10 to 50 mm hr<sup>-1</sup> results in leveling off slopes for  $Q_{tc}$  and *SS*, especially when K > 20mm hr<sup>-1</sup>. Both metrics are acutely responsive to *K* when *K* is below 10 mm hr<sup>-1</sup> but become non-responsive beyond 20mm hr<sup>-1</sup>.  $G_c$ 's response to *K* remains neutral.

1373

1374 D.1.4 Sensitivity to soil thickness

1375 Similar to *K*, a rise in  $T_{SL}$  from 1500mm to 15000 mm causue  $Q_{tc}$  to rise and *SS* to decline (Figure 1376 D-1d). The most significant rate of change for both metrics occurs between 1500 to 5000 mm  $T_{SL}$ . 1377 Beyond 5000 mm, changes flatten.  $G_c$  shows no response to  $T_{SL}$  variations.

- 1378
- 1379 D.1.5 Sensitivity to sine function amplitude

Increasing the sine function amplitude, *A*, from -0.03 to 0 mm hr<sup>-1</sup> pushes both  $Q_{tc}$  and *SS* increase and reach to their maximum at *A*=0 (Figure D-1e). But as *A* rises from 0 to 0.06 mm hr<sup>-1</sup>, they both decline. In contrast,  $G_c$  descends to its lowest (400 mm) when *A* shifts from -0.03 to 0 and then increases to 900 mm as *A* hits 0.063.

1384

# 1385 D.1.6 Sensitivity to sine function intercept

Both  $Q_{tc}$  and *SS* ascend with the growth in sine function's intercept, *B*. Increasing *B* from -0.015 to 0.005 mm hr<sup>-1</sup>sees  $G_c$  descend. During this *B* increase,  $Q_{tc}$  expands from 1100 to 2400 mm, while  $G_c$  shrinks from 1400 to 0 mm. It seems the sum of  $Q_{tc}$  and  $G_c$  might be constant. This suggests that water either drains through the tile pipe or percolates through the soil bottom.  $Q_{tc}$ , and *SS* appear sensitive to all six module parameters, but  $G_c$  only to *A* and *B*.

1391

#### 1392 **D.2 Module parameter evaluation for new sites**

As discussed in section 2.5, initial values for K,  $T_{CF}$  and  $\varphi_c$  can be determined by soil grain-size distribution. Parameters less explored in past research for new sites include the sine function's amplitude (*A*), intercept (*B*), and time delay ( $D_d$ ).

- 1396
- 1397 D.2.1 Evaluating sine function's A and B

1398 If no percolation exists from the soil's bottom to groundwater and  $G_{y,i}$  is zero, both *A* and *B* should 1399 be zero. However, if percolation or interactions between soil and groundwater occurs, *A* and *B* 1400 need calibration assessment. Before this, reasonable initial values and bounds must be set.

From this study's findings, *A* and *B* should fall between the mean hourly difference of infiltration and observed tile flow rates. For instance, observed hourly rates for infiltration and tile flow at our site are 0.07 and 0.03 mm hr<sup>-1</sup>. Thus, *A*'s and *B*'s initial values should range from -0.04 to 0.04 mm hr<sup>-1</sup>. Negative *A* and *B* values indicate outflow from soil to groundwater and vice versa. Initial values were set at 10% of the range limits: -0.004 for *B* and 0.004 for *A*. Eventually, *B* and *A* were adjusted to -0.005 and 0.025 mm hr<sup>-1</sup>.

1407

#### 1408 D.2.2 Assessment of sine function's time delay

The sine function begins on the first Julian day. If its peak occurs around 91<sup>st</sup> Julian day ( three months later), its minimum should be on the 274<sup>th</sup> day. If the peak comes later, say the 111<sup>th</sup> day, a 20-day delay is present. This delay should mirror in both function's minima and maxima. In this case the minimum would be on day 294. This delay aligns with the soil water table's peak annual fluctuations. When no observed fluctuations exist, the delay can be calibrated. A sensible initial delay can be ascertained by examining the study site's water table elevations, fitting a sine function, and noting the peak's Julian day annually.