

1 **Insights into carbonate environmental conditions in the Chukchi Sea**

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16 **Abstract**

17 Healthy Arctic marine ecosystems are essential to the food security and sovereignty, culture,
18 and wellbeing of Indigenous Peoples in the Arctic. At the same time, Arctic marine ecosystems
19 are highly susceptible to impacts of climate change and ocean acidification. While increasing
20 ocean and air temperatures and melting sea ice act as direct stressors on the ecosystem, they also
21 indirectly enhance ocean acidification, accelerating the associated changes in the inorganic
22 carbon system. Yet, much is to be learned about the current state and variability of the inorganic
23 carbon system in remote, high-latitude oceans. Here, we present time-series (2016-2020) of pH
24 and the partial pressure of carbon dioxide ($p\text{CO}_2$) from the northeast Chukchi Sea continental
25 shelf. The Chukchi Ecosystem Observatory includes a suite of subsurface year-round moorings
26 sited amid a biological hotspot that is characterized by high primary productivity and a rich
27 benthic food web that in turn supports coastal Iñupiat, whales, ice seals, walrus (*Odobenus*
28 *rosmarus*), and Arctic cod (*Boreogadus saida*). Our observations suggest that near-bottom
29 waters (33 m depth, 13 m above the seafloor) are a high carbon dioxide and low pH and
30 aragonite saturation state (Ω_{arag}) environment in summer and fall, when organic material from the
31 highly productive summer remineralizes. During this time, Ω_{arag} can be as low as 0.4. In winter,
32 when the site was covered by sea ice, pH was < 8 and Ω_{arag} remained undersaturated under the
33 sea ice. There were only two short seasonal periods with relatively higher pH and Ω_{arag} , which
34 we term ocean acidification relaxation events. In spring, high primary production from sea ice
35 algae and phytoplankton blooms led to spikes in pH (pH > 8) and aragonite oversaturation. In
36 late fall, strong wind-driven mixing events that delivered low CO_2 surface water to the shelf also
37 led to events with elevated pH and Ω_{arag} . Given the recent observations of high rates of ocean
38 acidification, and sudden and dramatic shift of the physical, biogeochemical, and ecosystem

39 conditions in the Chukchi Sea, it is possible that the observed extreme conditions at the Chukchi
40 Ecosystem Observatory are deviating from carbonate conditions to which many species are
41 adapted.

42

43 **1. Introduction**

44 The quickly changing Arctic Ocean has climatic, societal, and geopolitical implications for
45 the peoples of the Arctic and beyond (Huntington et al., 2022). Arctic Indigenous Peoples are at
46 the forefront of this change and their food security, food sovereignty, culture, and ways of life
47 depend on healthy Arctic marine ecosystems (ICC, 2015). The Arctic is warming at a rate that is
48 up to four times that of the rest of the globe (Serreze and Barry, 2011; Serreze and Francis, 2006;
49 Rantanen et al., 2022). This phenomenon, called Arctic Amplification, is observed in air and sea
50 temperatures, has accelerated in recent years, and is expected to continue in the future (Rantanen
51 et al., 2022; Shu et al., 2022). Warming exerts a toll on sea ice extent, ice thickness, and the
52 duration of seasonal sea ice cover: ice is forming later in fall and retreating earlier in spring,
53 thereby increasing the length of the open water period (Stroeve et al., 2011; Serreze et al., 2016;
54 Wood et al., 2015; Stroeve et al., 2014). The lowest Arctic wide minimum sea ice extents were
55 recorded during the last 16 years of the 44 year-long satellite time-series (National Snow and Ice
56 Data Center, DiGirolamo et al. (2022)).

57 At the same time, the Arctic Ocean is vulnerable to ocean acidification. Although oceanic
58 uptake of anthropogenic carbon dioxide (CO₂) increases oceanic CO₂ and decreases pH and
59 calcium carbonate (CaCO₃) saturation states of calcite (Ω_{calc}) and aragonite (Ω_{arag}) globally,
60 climate induced changes to riverine input, temperature, sea ice, and circulation are accelerating
61 the rate of ocean acidification in the Arctic Ocean like nowhere else in the world (Woosley and

62 Millero, 2020; Qi et al., 2022a; Yamamoto-Kawai et al., 2009; Orr et al., 2022; Semiletov et al.,
63 2016; Qi et al., 2017). Recent observational studies propose that freshening of the Arctic Ocean
64 due to increased riverine input may play an even greater role in acidifying the Arctic Ocean than
65 the uptake of anthropogenic CO₂ (Woosley and Millero, 2020; Semiletov et al., 2016). In
66 addition, the cold Arctic waters have naturally low concentrations of carbonate ions (CO₃²⁻) and
67 are therefore closer to aragonite undersaturation ($\Omega_{\text{arag}} < 1$) than more temperate waters (Orr,
68 2011; Sarmiento and Gruber, 2006), which leads to the chemical dissolution of free aragonitic
69 CaCO₃ structures (Bednaršek et al., 2021). Because of the naturally low concentrations of CO₃²⁻,
70 such high latitude waters have a lower capacity to take up anthropogenic CO₂ and buffer these
71 changes (Orr, 2011). As a result, concentrations of hydrogen ions (H⁺) increase and pH decreases
72 faster in the Arctic than in the tropics, for example.

73 In the Pacific Arctic, the Chukchi shelf waters have warmed by 0.45 °C decade⁻¹ since 1990,
74 triple the rate since the beginning of the data record in 1922 (Danielson et al., 2020). Direct
75 observations of the inorganic carbon dynamics of the Chukchi Sea are mostly limited to June
76 through November because of the region's remoteness and accessibility during sea ice covered
77 months. Summertime profiles across the Chukchi Sea show steep vertical gradients in inorganic
78 carbon chemistry (Bates, 2015; Bates et al., 2009; Pipko et al., 2002; Mathis and Questel, 2013).
79 Surface waters have a low partial pressure of carbon dioxide ($p\text{CO}_2$) as a result of high primary
80 production after sea ice retreat, leading to aragonite supersaturated conditions, with $\Omega_{\text{arag}} > 2$
81 (Bates, 2015; Bates et al., 2009). In areas with sea ice melt or riverine freshwater influence, Ω_{arag}
82 tends to be lower and at times undersaturated (Bates et al., 2009; Yamamoto-Kawai et al., 2009).
83 At the same time, $p\text{CO}_2$ values near the seafloor are around 1000 μatm as a result of organic
84 matter remineralization, leading to summertime aragonite undersaturation (Mathis and Questel,

85 2013; Pipko et al., 2002; Bates, 2015). Between September and November, continuous
86 measurements from within a few meters of the surface suggest a mosaic of $p\text{CO}_2$ levels between
87 ~ 200 to $600 \mu\text{atm}$, likely due to patchy wind-induced mixing entraining high- CO_2 waters from
88 depth into the surface mixed layer (Hauri et al., 2013). Yamamoto-Kawai et al. (2016) used
89 mooring observations of S, T, and apparent oxygen utilization to estimate dissolved inorganic
90 carbon (DIC), total alkalinity (TA), and Ω_{arag} in bottom waters at their mooring site in the Hope
91 Valley in the southwestern Chukchi Sea to give first insights into year round variability of the
92 inorganic carbon system. They found slightly less intense aragonite undersaturation in spring and
93 winter compared to summer, with a net undersaturation duration of 7.5-8.5 months per year.

94 The Chukchi Ecosystem Observatory (CEO) is situated in a benthic hotspot (Figure 1) where
95 high primary production supports rich and interconnected benthic and pelagic food webs
96 (Grebmeier et al., 2015; Moore and Stabeno, 2015). The benthos is dominated by calcifying
97 bivalves, polychaetes, amphipods, sipunculids, echinoderms and crustaceans (Grebmeier et al.,
98 2015; Blanchard et al., 2013). Benthic foraging bearded seals (*Erignathus barbatus*), walrus
99 (*Odobenus rosmarus divergens*), gray whale (*Eschrichtius robustus*), and seabirds feed on these
100 calcifiers during the open water season (Kuletz et al., 2015; Jay et al., 2012; Moore et al., 2022).
101 The CEO site, located on the southern flank of Hanna Shoal, is a region of reduced stratification
102 (relative to other sides of the shoal) that likely alternately feels the effects of differing flow
103 regimes located to the west and to the east (Fang et al., 2020). Consequently, the site exhibits
104 relatively weaker currents (Tian et al., 2021) and so is conducive to deposition of sinking organic
105 matter that in turn feeds the local benthos (Grebmeier et al., 2015). Prolonged open-water
106 seasons during periods of high solar irradiance, in combination with an influx of new nutrients
107 and wind mixing, are likely enhancing primary and secondary production as well as advection of

108 zooplankton (Lewis et al., 2020; Arrigo and van Dijken, 2015; Wood et al., 2015). These
109 physical processes in turn fuel keystone consumers such as Arctic cod (*Boreogadus saida*) and
110 upper trophic level ringed seals (*Phoca hispida*), beluga (*Delphinapterus leucas*) and bowhead
111 whales (*Balaena mysticetus*) as well as predatory polar bears (*Ursus arctos*) and Indigenous
112 People who rely on the marine ecosystem for traditional and customary harvesting (Huntington
113 et al., 2020).

114 Perturbation of the seawater carbonate system associated with ocean acidification and
115 climate change can have significant physiological and ecological consequences for marine
116 species and ecosystems (Doney et al., 2020). All parameters of the carbonate system (pH, $p\text{CO}_2$,
117 Ω_{arag} , concentrations of HCO_3^- , CO_3^{2-} , etc.) have the potential to affect the physiology of marine
118 organisms while a change in the saturation state (Ω) can lead to the dissolution of unprotected or
119 “free” CaCO_3 structures. Recent work has highlighted the importance of local adaptation to the
120 present environmental variability as a key factor driving species sensitivity to ocean acidification
121 (Vargas et al., 2017, 2022). As carbonate chemistry conditions vary enormously between
122 regions, marine organisms are naturally exposed to different selective pressures and can evolve
123 different strategies to cope with low pH or Ω , or high $p\text{CO}_2$. For example, the deep-sea mussel
124 *Bathymodiolus brevior* living around vents at 1600 m depths is capable of precipitating calcium
125 carbonate at pH ranging between 5.36 and 7.30 and highly undersaturated waters (Tunnicliffe et
126 al., 2009). The response to changes in the carbonate chemistry is also modulated by other
127 environmental drivers such as temperature or food availability (e.g. Thomsen et al., 2013;
128 Breitberg et al., 2015). Consequently, no absolute or single threshold is expected for ocean
129 acidification (e.g., Bednaršek et al., 2021) and a pre-requisite to assessing the impact on any
130 biota is the monitoring at a short temporal scale to characterize the present environmental niche.

131 When it comes to future impacts, the more intense and faster the changes associated with ocean
132 acidification, the more adverse associated biological impacts are expected (Vargas et al. 2017,
133 2022). As a result, it is anticipated that Arctic marine waters that are experiencing widespread
134 and rapid ocean acidification will potentially undergo severe negative ecosystem impacts
135 (AMAP 2018).

136 Here, we present satellite sea ice coverage data and four years of nearly continuous salinity,
137 temperature, and $p\text{CO}_2$ data, accompanied by pH, nitrate (NO_3), dissolved oxygen (O_2), and
138 chlorophyll fluorescence data for some of the time (Table 1, Figures 2 and 3). We developed an
139 empirical equation for estimating pH from moored $p\text{CO}_2$, temperature, and salinity and evaluated
140 it using discrete samples collected across the Chukchi Sea, Bering Sea, and Beaufort Sea. Our
141 timeseries allow us to assess the seasonal and interannual variability and controls of the
142 inorganic carbon system in the Chukchi Sea between 2016 and 2020 and characterize the
143 chemical conditions experienced by organisms. We discuss our observations in terms of
144 progressing acidification and implications to organisms in the Chukchi Sea region.

145

146 **2. Materials and Methods**

147 **2.1 The Chukchi Ecosystem Observatory (CEO)**

148 The Chukchi Sea is a shallow shelf sea with maximum depths < 50 m. It is largely a
149 unidirectional inflow shelf system with Pacific origin water entering the Chukchi Sea through the
150 Bering Strait and advecting north into the Arctic Ocean (Carmack and Wassmann, 2006). The
151 CEO ($71^\circ 36'$ N, $161^\circ 30'$ W, Figure 1, Hauri et al., 2018) is located along the pathway of waters
152 flowing through Bering Strait (Fang et al., 2020) and thence from the west of Hanna Shoal
153 toward Barrow Canyon to the south, although the wind can also drive waters from the east over

154 the observatory site (Fang et al., 2020). From both shipboard and moored acoustic Doppler
155 current profiler records, the south side of Hanna Shoal mean flow is characterized by a weak
156 southward-directed current (Tian et al., 2021).

157 The observatory consists of oceanographic moorings that sample year-round, equipped with a
158 variety of sensors that measure sea ice cover and thickness (Sandy et al., 2022), light, currents,
159 waves, salinity, temperature, concentrations of dissolved oxygen, nitrate, and particulate matter,
160 pH, $p\text{CO}_2$, chlorophyll fluorescence, zooplankton abundance and vertical migration (Lalande et
161 al., 2021, 2020), the presence of Arctic cod and zooplankton (Gonzalez et al., 2021), and the
162 vocalizations of marine mammals. During some years, the observatory included a third mooring,
163 an experimental “freeze-up detection mooring”, which transmitted real-time data of conductivity
164 and temperature throughout the water column until sea ice formation. The primary moorings
165 stretch from the seafloor at 46 m to about 33 m depth, designed to avoid collisions with ice keels.
166 Pressure sensors at the top of the moorings show less than ± 1 m of excursion of the moored
167 sensor package from its deployment mean depth in any given year, indicating that mooring blow-
168 over or diving is not the cause of any observed large variability. Description of the CEO and lists
169 of sensors deployed at the site can be found in Danielson et al. (2017) and Hauri et al., (2018).
170 For this study we focus on the inorganic carbon system and its controlling mechanisms.

171

172 **2.2 $p\text{CO}_2$**

173 We used a CONTROS HydroC CO_2 sensor (4H-Jena Engineering GmbH, Kiel, Germany) to
174 measure $p\text{CO}_2$. The Contros HydroC CO_2 sensor was outfitted with a pump (SBE 5M, Sea-Bird
175 Electronics) that flushes ambient seawater against a thin semi permeable membrane, which
176 serves as equilibrator for dissolved CO_2 between the ambient seawater and the headspace of the

177 sensor. Technical details about the sensor and its performance are described in Fietzek et al.
178 (2014), who estimated sensor accuracy to be better than 1% with postprocessing.

179 A HydroC CO₂ sensor has been deployed at the CEO site since 2016. In all deployments,
180 except 2016, HydroC CO₂ sensors were post-calibrated. The lack of post-calibration in 2016 is
181 not expected to negatively affect data quality because a battery failure resulted in data returns
182 only over the first 3 months (August through November). Following a zero interval where the
183 gas was pumped through a soda lime cartridge to create a zero-signal reference with respect to
184 CO₂, and subsequent flush interval to allow CO₂ concentrations to return to ambient conditions,
185 measurements were taken in a burst fashion every 12 or 24 hours depending on deployment year
186 (Table 1). Average *p*CO₂ values are reported as the mean of the measure interval (Table 1) with
187 standard uncertainty (Equation 1) defined following best practices (Orr et al., 2018) and where
188 the random component is the standard deviation of the mean, and the systematic components
189 include sensor accuracy and estimated error of the regression during calibration.

$$190 \quad u = \sqrt{u_{\text{systematic}}^2 + u_{\text{random}}^2} \quad (1)$$

191 More than 96% of the time, the relative uncertainty of the *p*CO₂ data met the weather data
192 quality goal, defined as 2.5% by the Global Ocean Acidification Observing Network (GOA-ON,
193 Newton et al., 2015).

194 HydroC CO₂ data were processed using Jupyter notebook scripts developed by 4H-Jena
195 Engineering GmbH using pre- and post-calibration coefficients interpolated with any change in
196 the zero-signal reference over the deployment (Fietzek et al., 2014). Further processing using in-
197 house MATLAB scripts included removal of outliers, calculation of the average *p*CO₂, and
198 calculation of uncertainty estimates for each measure interval.

199

200 2.3 pH

201 A SeapHOx sensor (Satlantic SeaFET™ V1 pH sensor integrated with Sea-Bird Electronics
202 SBE 37-SMP-ODO) was used to concurrently measure pH, salinity, temperature, pressure, and
203 oxygen (Martz et al., 2010). A SeapHOx was deployed at CEO in 2016, 2017, and 2018. No
204 SeapHOx was deployed in 2019 or 2020 due to supply chain delays and communication issues at
205 sea. Unfortunately, measured pH ($\text{pH}_{\text{SeaFET}}$) from the 2016 and 2018 SeapHOx deployments
206 were unusable due to high levels of noise in both the internal and external electrodes. In short,
207 we only have usable pH data between August 2017 and August 2018.

208 $\text{pH}_{\text{SeaFET}}$ data were excluded during a 14-day conditioning period following deployment and
209 were processed with post-calibration corrected temperature and salinity from the SBE37
210 following Bresnahan et al. (2014) using voltage from the external electrode (V_{ext}), and $\text{pH}_{V_{\text{ext}}}$
211 (pH calculated from the external electrode of the SeaFET) from an extended period of low
212 variability (18 February 2018). Despite the availability of discrete data from one calibration cast
213 (Cross et al., 2020b; Table 2), $\text{pH}_{V_{\text{ext}}}$ was used as the single calibration point (Bresnahan et al.,
214 2014) for a variety of reasons: 1) high variability of $\text{pH}_{\text{SeaFET}}$ (0.0581 pH units) straddling a 12
215 hour window around the discrete sample collection time, 2) high temporal and spatial variability
216 often seen in the Chukchi Sea, and 3) the discrete pH sample was within the published SeaFET
217 accuracy of 0.05 (Table 2, Figure S1). $\text{pH}_{\text{SeaFET}}$ values are reported as the mean of the measure
218 interval (Table 1) and standard uncertainty is calculated with Equation 1 with the standard
219 deviation of the average (random), and the SeaFET accuracy (systematic). Data handling and
220 processing were done using in-house MATLAB scripts. pH is reported in total scale and at *in*
221 *situ* temperature and depth for the entirety of this paper.

222

223 2.4 Nitrate

224 NO₃ measurements were from a Submersible Ultraviolet Nitrate Analyzer (SUNA) V2 by
225 Sea-Bird Scientific. The SUNA is an *in situ* ultraviolet spectrophotometer designed to measure
226 the concentration of nitrate ions in water. SUNA V2 data were processed using a publicly
227 available toolbox (Hennon et al., 2022; Irving, 2021) with QA/QC steps that included thermal
228 and salinity corrections (Sakamoto et al., 2009), assessment of spectra and outlier removal based
229 on spectral counts (Mordy et al., 2020), and concentration adjustments (absolute offset and linear
230 drift) based on pre-deployment and post-recovery reference measurements of zero concentration
231 (DI) water and a nitrate standard and, when available, nutrient samples taken from Niskin bottles
232 near the mooring site (e.g. Daniel et al., 2020).

233

234 2.5 CTD and Oxygen

235 Two CTDs were deployed on the CEO mooring near the HydroC CO₂ depth. The main
236 pumped Sea-Bird SeaCAT (SBE16) has been deployed on the CEO mooring around 33 m depth
237 since 2014. A pumped SBE43 oxygen sensor was deployed with the SBE16 during the 2015-
238 2016, 2017-2018, and 2019-2020 deployments but only data returns from the 2017-2018
239 deployment is discussed briefly in this manuscript (Figure S2).

240 The other pumped CTD was a Sea-Bird MicroCAT (SBE37-SMP-ODO), which was
241 integrated with an optical dissolved oxygen sensor (SBE63; Figure S2), and the SeaFET pH
242 sensor within the SeapHOx instrument. The SeapHOx was deployed in fall 2016, 2017, and
243 2018. The SBE37-SMP-ODO did not record any CTD or oxygen data during the 2016
244 deployment and only recorded CTD and oxygen data between August and November 3 in 2018
245 due to battery failure.

246 Processing of these data included temperature and conductivity correction using pre- and
247 post-calibration data following Sea-Bird Application Note 31 and oxygen correction using pre-
248 and post-calibration data following Sea-Bird Module 28. Oxygen was converted from ml/l to
249 $\mu\text{mol/kg}$ following Bittig et al. (2018). Density and practical salinity were calculated using the
250 TEOS-10 GSW Oceanographic Toolbox (McDougall and Baker, 2011).

251 Differences between the two oxygen sensors (SBE43 and SBE63) of approximately 145 to
252 265 $\mu\text{mol/kg}$ were observed over the 2017-2018 deployment, and both moored sensors had
253 varying offsets compared to nearby casts (Figure S2). Therefore, only relative oxygen values
254 from the freshly calibrated SBE63 are discussed in this paper.

255 The freeze-up detection mooring (Figure 6) consisted of four Sea-Bird SBE 37 inductive
256 modem CTD sensors that transmitted in real time hourly temperature, salinity, and pressure data
257 via the surface float from four subsurface depths (8, 20, 30, and 40 m; Hauri et al., 2018).

258

259 **2.6 Development of empirical relationship to estimate pH**

260 Empirical relationships for estimating water column pH have been developed for regions
261 spanning southern, tropical, temperate and Arctic biomes, using a variety of commonly measured
262 parameters (e.g., $\text{pH}(\text{S}, \text{T}, \text{NO}_3, \text{O}_2, \text{Si})$ Carter et al 2018; $\text{pH}(\text{O}_2, \text{T}, \text{S})$ Li et al., 2016; $\text{pH}(\theta, \text{O}_2)$
263 Watanabe et al., 2020; $\text{pH}(\text{NO}_3, \text{T}, \text{S}, \text{P})$ and $\text{pH}(\text{O}_2, \text{T}, \text{S}, \text{P})$ Williams et al., 2016; $\text{pH}(\text{O}_2, \text{T})$
264 Alin et al., 2012; $\text{pH}(\text{O}_2, \text{T})$ and $\text{pH}(\text{NO}_3, \text{T})$ Juranek et al., 2009). Given the tight coupling
265 between the concentration of H^+ and concentration of CO_2 solution, an empirical relationship for
266 estimating surface pH from $p\text{CO}_2$ was developed by the National Academies of Sciences,
267 Engineering and Medicine (2017) appendix F. Licker et al. (2019) used this empirical
268 relationship to calculate the global average surface ocean pH and found it represented the

269 relationship for surface water temperatures spanning 5°C to 45°C. Here, we take a similar
270 approach but extend it to water column pH in our cold region using temperature (T) and salinity
271 (S) as additional proxy parameters (Equation 2).

$$272 \quad pH^{est} = \alpha_0 + \alpha_1 \log(pCO_2) + \alpha_2 T + \alpha_3 S \quad (2)$$

273 Where pH^{est} is the estimated value of water column pH, pCO_2 is from the HydroC, and T and S
274 are from the SBE16, and all α ($\alpha_0 = 10.4660$, $\alpha_1 = -0.4088$, $\alpha_2 = 0.0013$, $\alpha_3 = -0.0001$) terms are
275 model-estimated coefficients determined using MATLAB's multiple linear regression algorithm
276 *regress.m* (Chatterjee and Hadi, 1986). After interpolating pH_{SeaFET} (Figure 4, red dots) to the
277 pCO_2 timestamp, the algorithm was trained over an arbitrarily chosen 180-day period
278 (15/9/2017-14/3/2018, Figure 4, dashed box). An uncertainty of 0.0525 for pH^{est} (Figure 3 and
279 Figure S1, gray shading) was determined with Equation 1, where the RMSE (the uncertainty in
280 the estimation) over the entire pH_{SeaFET} timeseries is the random component and the published
281 accuracy of the SeaFET is the systematic component (since the algorithm was trained with
282 pH_{SeaFET}). The algorithm cross-validation and evaluation are discussed in section 3.1. Unless
283 explicitly defined otherwise, observations of pH refer to pH^{est} for the remainder of this paper.
284

285 **2.7 Carbonate system calculations**

286 Moored data were collected at different sample intervals (Table 1) and were linearly
287 interpolated to the HydroC CO_2 timestamp to enable further calculations. TA, DIC, and Ω_{arag}
288 (Figure 11 a & b and Figure 3d) were calculated based on measured pCO_2 , S, T, and pressure (P)
289 and algorithm-based pH (pH^{est}). Due to a lack of data, nutrient concentrations (Si, PO_4 , NH_4 ,
290 H_2S) were assumed to be negligible in the CO2SYS calculations (e.g. deGrandpre et al., 2019;
291 Vergara-Jara et al., 2019; Islam et al., 2017). pH^{est} was used in lieu of pH_{SeaFET} to allow for

292 calculations over the whole $p\text{CO}_2$ record and due to erroneously large variability of DIC and TA
293 when $\text{pH}_{\text{SeaFET}}$ was used as an input parameter (Raimondi et al., 2019; Cullison-Gray et al.,
294 2011). The pH - $p\text{CO}_2$ input pair leads to large, calculated errors in DIC and TA (Raimondi et al.,
295 2019; Cullison-Gray et al., 2011) due to strong covariance between the two parameters (both
296 temperature and pressure dependent). Cullison-Gray et al. (2011) attributed unreasonably large
297 short-term variability in calculated TA and DIC to temporal or spatial measurement mismatches
298 between input pH and $p\text{CO}_2$ parameters and found that appropriate filtering alleviated noise
299 spikes. By using pH^{est} , which by the nature of its definition is well correlated to $p\text{CO}_2$, we are
300 eliminating some of these spurious noise spikes. We show Ω_{arag} calculated from $\text{pH}_{\text{SeaFET}}$ - $p\text{CO}_2$
301 (Figure 3d, red line) because it is less sensitive to calculated errors as it accounts for a small
302 portion of the total CO_2 in seawater (Cullison-Gray et al., 2011).

303 All inorganic carbon parameters were calculated using CO2SYSv3 (Sharp et al., 2023; Lewis
304 and Wallace, 1998) with dissociation constants for carbonic acid of Lueker et al. (2000),
305 bisulfate of Dickson (1990), hydrofluoric acid of Perez and Fraga (1987), and the boron-to-
306 chlorinity ratio of Lee et al. (2010). Sulpis et al. (2020) found that the carbonic acid dissociation
307 constants of Lueker et al. (2000) may underestimate $p\text{CO}_2$ in cold regions (below $\sim 8^\circ\text{C}$), and
308 therefore overestimate pH and CO_3^{2-} . However, we choose to use Lueker et al. (2000) because
309 they are recommended (Dickson et al., 2007; Woosley, 2021), continue to be the standard (Jiang
310 et al., 2021; Lauvset et al., 2021), and are commonly used at high latitudes (Duke et al., 2021;
311 Raimondi et al., 2019; Woosley et al., 2017). Furthermore, the difference between DIC
312 calculated from pH^{est} and $p\text{CO}_2$ and discrete samples interpolated to moored instrument depth
313 ranged from 266 to -195 $\mu\text{mol}/\text{kg}$ using the K_1^* and K_2^* of Sulpis et al. (2020), compared to -38
314 to -7 $\mu\text{mol}/\text{kg}$ using Lueker et al. (2000).

315

316 **2.8 Sea ice concentration**

317 Sea ice concentration at the observatory site was taken from the National Snow and Ice Data
318 Center (NSIDC; DiGirolamo et al., 2022). Latitude and longitude coordinates were converted to
319 NSIDC's EASE grid coordinate system (Brodzik and Knowles, 2002) and the 25-km gridded
320 data were bilinearly interpolated to calculate sea ice concentration at the CEO site. Low sea ice is
321 defined by < 15 % sea ice coverage per grid cell.

322

323 **2.9 Estimation of model-based ocean acidification trend**

324 Model results were obtained from historical simulations of five different global Earth System
325 Models: 1) GFDL-CM4 (Silvers et al., 2018), 2) GFDL-ESM4 (Horowitz et al., 2018), 3) IPSL-
326 CM6A-LR-INCA (Boucher et al., 2020), 4) CNRM-ESM2-1 (Seferian, 2019), and 5) Max Plank
327 Earth System Model 1.2 (MPI-ESM1-2-LR, Wieners et al., 2019) that are part of the Coupled
328 Model Intercomparison Project Phase 6 (CMIP6). Each simulation was used to calculate the
329 annual trend of aragonite saturation state and pH at the closest depth and grid cell to the CEO
330 mooring.

331

332 **3. Results**

333 In the following, we will evaluate the pH algorithm (section 3.1), analyze the large
334 variability patterns (sections 3.2 and 3.3), and then take a closer look at the data from 2020 since
335 the seasonal cycle was different in 2020 than in previous years (section 3.4).

336

337 **3.1 pH algorithm**

338 The algorithm estimated pH data from the CEO site reasonably well and within the weather
339 uncertainty goal as defined by Newton et al. (2015) most of the time. As a first step, pH^{est}
340 consistency was assessed through cross-validation (Figure 5) using the test dataset (outside the
341 training period, $r^2 = 0.9666$, $\text{RMSE} = 0.166$) and across the whole timeseries ($r^2 = 0.9598$, RMSE
342 $= 0.0161$, $p < 0.0001$, Figure 5). Observed high frequency spikes in $\text{pH}_{\text{SeaFET}}$ (Figure 4, red dots;
343 Figure 5d, red line) were not captured by the HydroC $p\text{CO}_2$ sensor (sampling frequency of 12 h)
344 and as a result, are not reproduced in the pH^{est} timeseries. Throughout the $\text{pH}_{\text{SeaFET}}$ timeseries,
345 pH^{est} overestimates $\text{pH}_{\text{SeaFET}}$ by a mean of 0.0008 and median of 0.0039. Since pH^{est} generally
346 overestimates $\text{pH}_{\text{SeaFET}}$, we assume that Ω_{arag} is also somewhat overestimated throughout this
347 manuscript. Discrete water samples were used as reference values to evaluate the algorithm at
348 the CEO site (Table 2) and were found to be within the pH^{est} uncertainty (Figure S1).

349 An independent verification of our algorithm was done using discrete data collected from the
350 Bering Sea to the Arctic Ocean on four research cruises in 2020, 2019, 2018, and 2017 (Figure
351 6d; Monacci et al., 2022; Cross et al., 2021; 2020a; 2020b), henceforth called the DBO dataset.
352 Samples collected from deeper than 500 m below the surface or flagged as questionable or bad
353 were excluded from this analysis. pH and $p\text{CO}_2$ were calculated from 1275 discrete samples
354 analyzed for TA, DIC, silicate, phosphate, and ammonium (except when silicate, phosphate, and
355 ammonium were assumed to be negligible for the 327 samples from cruise SKQ202014S;
356 Monacci et al., 2022) using CO2SYSv3 (Sharp et al., 2023; section 2.7 for details) and are
357 referred to as $\text{pH}^{\text{disc}}_{\text{calc}}$ and $p\text{CO}_2^{\text{disc}}_{\text{calc}}$, respectively. $\text{pH}^{\text{disc}}_{\text{est}}$ was based on discrete water samples
358 and calculated using Equation 2 and was fit to $\text{pH}^{\text{disc}}_{\text{calc}}$ using a linear regression ($r^2 = 0.9975$,
359 $\text{RMSE} = 0.0078$, $p\text{-value} < 0.0001$; Figure 6 a – c). Mean and median differences between
360 $\text{pH}^{\text{disc}}_{\text{calc}}$ and $\text{pH}^{\text{disc}}_{\text{est}}$ were zero and 0.0022, respectively, with largest anomalies observed at

361 lower salinities (Figure 6c). Absolute differences between $\text{pH}^{\text{disc}}_{\text{est}}$ and $\text{pH}^{\text{disc}}_{\text{cal}}$ over the salinity
362 range observed at the CEO site (30.87 to 33.93) fall within the weather data quality goal
363 (Newton et al., 2015) 98.7% of the time with maximum absolute differences < 0.03 . The
364 uncertainty of 0.0154 for $\text{pH}^{\text{disc}}_{\text{est}}$ was determined using Equation 1, where the mean combined
365 standard uncertainty (u_c) for $\text{pH}^{\text{disc}}_{\text{calc}}$ (0.0133; Orr et al., 2018) was the systematic component,
366 and the regression RMSE was the random component.

367 Empirical relationships for estimating water column pH that rely on dissolved oxygen often
368 ignore surface waters to limit biases due to decoupling the stoichiometry of the $\text{O}_2:\text{CO}_2$
369 relationship due to air-sea gas exchange (e.g. Juranek et al., 2011; Alin et al., 2012; Li et al.,
370 2016). We see evidence of this bias in our algorithm at low salinity (Figure 6c) and low $p\text{CO}_2$
371 (not shown) when compared with the DBO dataset samples collected across the Arctic and from
372 the surface to 500 m, with $\text{pH}^{\text{disc}}_{\text{est}}$ overestimating $\text{pH}^{\text{disc}}_{\text{calc}}$ by a maximum of 0.049. If depth is
373 restricted to between 30 and 500 m when evaluating the algorithm with the DBO dataset,
374 algorithm performance improves ($r^2 = 0.9990$, $\text{RMSE} = 0.0055$, $p\text{-value} < 0.0001$; not shown)
375 and the maximum $\text{pH}^{\text{disc}}_{\text{est}}$ overestimates $\text{pH}^{\text{disc}}_{\text{calc}}$ by 0.022.

376

377 **3.2 Relaxation events**

378 The sub-surface waters at the CEO site comprise a high $p\text{CO}_2$, low pH, and low Ω_{arag}
379 environment, with mean values of $p\text{CO}_2^{\text{mean}} = 538 \pm 7 \mu\text{atm}$, $\text{pH}^{\text{mean}} = 7.91 \pm 0.05$, $\Omega_{\text{arag}}^{\text{mean}} =$
380 0.94 ± 0.23 across the full data record (Figure 3 b - d). In the following we will focus on spikes
381 of high pH and Ω_{arag} and low $p\text{CO}_2$ that occur in spring (May-June) and fall (September-
382 December); we define these spikes as relaxation events (see discussion for justification of term).

383 *Spring:* Springtime relaxation events at 33 m depth that exhibit relatively higher pH and
384 Ω_{arag} and lower $p\text{CO}_2$ compared to the overall mean, are likely consequences of photosynthetic
385 activity during sea ice break-up (Figures 2 and 3). In June of 2018 and 2019, near bottom pH and
386 Ω_{arag} spiked to > 8.17 and > 1.5 , respectively, while $p\text{CO}_2$ dropped to $< 286 \mu\text{atm}$. Ω_{arag} remained
387 oversaturated and pH was greater than 8.0 for nearly all of June in 2018. In 2019, the relaxation
388 event was less sustained, with only four short (2-6 day-long) events of relatively higher pH and
389 $\Omega_{\text{arag}} > 1$ in June. In both years, chlorophyll fluorescence spiked and either O_2 increased (in
390 2018) or NO_3 decreased (in 2019), which are signs of photosynthetic activity and primary
391 production.

392 *Fall:* The relaxation events in fall were characterized by large and sudden drops in $p\text{CO}_2$,
393 abrupt increases in pH and Ω_{arag} , and considerable interannual variability in their timing. Unlike
394 the relaxation events observed in spring, we attribute these fall relaxation events to wind-induced
395 physical mixing. To examine the controlling mechanisms causing these abrupt relaxation events
396 in fall, we will start with using water column salinity and temperature data from a freeze-up
397 detection buoy (Hauri et al., 2018) that was deployed in summer 2017 approximately 1 km away
398 from the biogeochemical mooring. The freeze-up detection mooring provided temperature and
399 salinity measurements every 7 meters throughout the water column from the time of its
400 deployment in mid-August until freeze-up. Data from the freeze-up detection mooring suggest
401 that warmer and fresher water from the upper water column gets periodically entrained down to
402 the location of the biogeochemical sensor package at 33 m depth, leading to enhanced variability
403 of density in August and September (Figure 7). Fluctuations of the pycnocline associated with
404 the passage of internal waves could also elevate signal variances. During this time $p\text{CO}_2$ often
405 decreased to or below atmospheric levels and pH sporadically reached values > 8 . At the end of

406 September, a strong mixing event (with coincident strong surface winds) homogenized the water
407 column from the surface down to the location of the sensor package and caused a sudden
408 temperature increase from 0.4 °C to 3.9 °C (Figure 7c and 8a). At the same time, $p\text{CO}_2$ (Figure
409 7b and 8) decreased from 590 to 308 μatm . This suggests that warm and low CO_2 surface water
410 mixed with CO_2 -rich subsurface water and led to a sustained relaxation period that subsequently
411 lasted until mid-November. Another mixing event further eroded the water column stratification
412 and replaced subsurface water with colder and fresher water (ice melt) from the surface at the
413 end of October. This second large mixing event did not lead to large changes in $p\text{CO}_2$, pH, and
414 Ω_{arag} .

415 Salinity and temperature records from the biogeochemical mooring at 33 m depth also
416 suggest fall season mixing events in all other years, when increases in temperature coincide with
417 decreases in $p\text{CO}_2$ (Figure 2b and c, 3a and 8). For example, two mixing events shaped the
418 carbonate chemistry evolution in fall 2018. $p\text{CO}_2$ decreased from 915 μatm to around 565 μatm
419 and Ω_{arag} increased to 0.9 as temperature increased and salinity decreased in early September
420 (Figures 2 and 8). $p\text{CO}_2$ then increased to 1160 μatm in late October, before decreasing to 385
421 μatm at the beginning of November, causing a spike in Ω_{arag} to 1.34. At the same time, salinity
422 decreased by 1 unit, suggesting a strong mixing event. Throughout November 2018, $p\text{CO}_2$
423 oscillated between 344 and 757 μatm and salinity between 31.01 and 32.97, hinting at additional
424 mixing.

425 Similarly, an early mixing event in 2019 decreased $p\text{CO}_2$ to 352 μatm at the beginning of
426 September. Short-term variability in $p\text{CO}_2$ with maximum levels of up to 855 μatm and
427 minimum values below 300 μatm , variable temperature and salinity, and sporadic aragonite
428 oversaturation events point to mixing through mid-September. At the end of October, a large

429 mixing event homogenized the water column, accompanied by a decline of salinity by >1 unit,
430 increase of temperature to 4 °C, and decrease of $p\text{CO}_2$ from 565 μatm to below 400 μatm . In a
431 similar fashion to 2018, this fall mixing event was followed by a month-long period of large
432 variability of $p\text{CO}_2$, salinity, pH, and Ω_{arag} , leading to short and sporadic aragonite oversaturation
433 events in November, and sustained oversaturation in December.

434

435 **3.3 Sustained periods of low pH and Ω_{arag} , and high $p\text{CO}_2$**

436 Waters at 33 m depth at the CEO site were most acidified during the sea ice free periods
437 until mixing events entrained surface waters to the sensor depth (section 3.2). pH and Ω_{arag}
438 started to gradually decrease from their maximum levels ($\Omega_{\text{arag_max}} = 1.65$, $\text{pH}_{\text{max}} = 8.19$) at the
439 beginning of June in 2018 to their annual low at the beginning of November ($\Omega_{\text{arag_min}} = 0.47$,
440 $\text{pH}_{\text{min}} = 7.58$, Figure 3 d and c). In November, the waters were also undersaturated with regards
441 to calcite (not shown) and $p\text{CO}_2$ peaked at 1159 μatm (Figure 3b). Dissolved oxygen decreased
442 by about 400 $\mu\text{mol kg}^{-1}$ between July and October, when the sensor stopped working properly.
443 The decrease of dissolved oxygen suggests remineralization of organic material. The decrease of
444 pH, Ω_{arag} , O_2 and increase of $p\text{CO}_2$ was briefly interrupted by a strong mixing event in
445 September, which entrained warmer, fresher, and CO_2 -poorer water down to 33 m depth (section
446 3.2, Figure 8). The 2019 observations paint a similar picture of remineralization during the
447 summer months, as the $p\text{CO}_2$ increase and pH and Ω_{arag} decreases were accompanied by an NO_3
448 increase (Figure 2d and 3b-d).

449 $p\text{CO}_2$ steadily increased and pH and Ω_{arag} decreased during the sea ice covered periods
450 (Figures 8). pH was < 8 and Ω_{arag} remained undersaturated under the sea ice. At the same time,
451 NO_3 slowly increased and O_2 decreased, which points to slow organic matter remineralization

452 (Figure 9). Short-term variability in $p\text{CO}_2$, especially in January of all three observed years, was
453 also reflected in salinity, O_2 and NO_3 (Figure 9) and could be attributed to advection, as the CEO
454 site is adjacent to contrasting regimes of flow and hydrographic properties (Fang et al., 2020).

455

456 **3.4 Spring and summer of 2020 were different**

457 The seasonal cycle in 2020 strongly contrasted with the previous observed years. $p\text{CO}_2$
458 gradually increased by roughly 200 μatm throughout the sea ice covered months to 650 μatm
459 when sea ice started to retreat at the beginning of July. By the end of July, $p\text{CO}_2$ doubled and
460 increased to 1389 μatm , which is the highest $p\text{CO}_2$ level recorded in this timeseries. The peak of
461 $p\text{CO}_2$ was accompanied by an increase in salinity of 0.5 while temperature did not change,
462 suggesting the influence of advection. At the beginning of August, $p\text{CO}_2$ dropped to 536 μatm
463 and then oscillated around 600 μatm through much of August before returning to around 900
464 μatm for the next month. Similarly, pH decreased to 7.5 at the end of July and then oscillated
465 around 7.85, while Ω_{arag} dropped to 0.37, and oscillated around 0.85. The steep drop and
466 oscillation of $p\text{CO}_2$ was reflected in NO_3 , suggesting that primary production and
467 remineralization played a role. When $p\text{CO}_2$ and NO_3 decreased at the beginning of August,
468 temperature simultaneously increased by 0.7 $^{\circ}\text{C}$ and salinity decreased by 0.12, suggesting that
469 entrainment of shallower water masses may have played a role too. Comprehensive analyses of
470 the factors that resulted in the 2020 differing conditions are beyond the scope of this paper, but
471 deserve attention in a future effort.

472

473 **4. Discussion**

474 CEO data provide new insights into the synoptic, seasonal and interannual variability of
475 the inorganic carbon system in a time when ocean acidification and climate change have already
476 started to transform this area. The observations suggest that the CEO site is a high-CO₂ and low-
477 pH and low- Ω_{arag} environment most of the time, except during sea ice break-up when the effects
478 of photosynthetic activity remove CO₂ from the system, and later in fall, when strong storm
479 events entrain low $p\text{CO}_2$ surface waters to the seafloor. Lowest pH and CaCO₃ saturation states
480 and highest $p\text{CO}_2$ occur in summer through late fall when organic matter remineralization
481 dominates the carbonate system balance. During this time, Ω_{arag} can fall below 0.5 and even Ω_{calc}
482 becomes sporadically undersaturated ($\Omega_{\text{calc}} < 1$).

483

484 **4.1 pH algorithm**

485 Deploying oceanographic equipment in remote Arctic locations is challenging. The data
486 return from the SeapHOx sensors was disappointingly minimal, despite annual servicing and
487 calibration by the manufacturer. Our new pH algorithm is therefore even more important as it
488 fills pH data gaps in the CEO timeseries and can be applied with confidence from the Bering to
489 the western Beaufort seas (Figure 6). While another successful year of moored pH data return at
490 the CEO site is needed to fully evaluate our algorithm throughout the year, comparison with
491 single discrete water samples nearby the CEO site and the DBO dataset (section 3.1, Table 2,
492 Figures 6 and S1) suggest that our algorithm-derived pH meets the weather quality uncertainty
493 goal of ± 0.02 (Newton et al., 2015) much of the time.

494 The combination of our new algorithm with recent progress in monitoring $p\text{CO}_2$ with
495 Seagliders (Hayes et al., 2022) will further increase our ability to study the inorganic carbon
496 dynamics at times and locations when shipboard or mooring based measurements may not be

497 practical. Additional assessment is needed to determine to what degree the algorithm needs
498 adjustments beyond the region evaluated in this work.

499

500 4.2 Uncertainty

501 Inherent spatial and temporal variability of the inorganic carbon parameters in the
502 Chukchi Sea make the use of discrete water samples for evaluating sensor-based measurements
503 difficult. Historic continuous surface measurements from the area suggest that surface $p\text{CO}_2$ can
504 be as low $< 250 \mu\text{atm}$ in early fall (Hauri et al., 2013), at a time of year when subsurface $p\text{CO}_2$
505 reaches its max of $>800 \mu\text{atm}$ at the CEO site. This suggests a steep $p\text{CO}_2$ gradient of $> 17 \mu\text{atm}$
506 per meter. High-resolution pH data from the 2017/2018 deployment suggests high temporal
507 variability as well, further complicating the collection of discrete water samples to adequately
508 evaluate the sensors. The HydroC's zeroing function, in addition to our pre- and post-calibration
509 routines that factor into the post-processing of the data, gives us confidence in the accuracy of
510 the $p\text{CO}_2$ data, and further confidence in pH derived from $p\text{CO}_2$.

511 The pH^{est} uncertainty of 0.0525 is likely a conservative estimate based on our validation
512 of pH^{est} (section 3.1, Table 2). Consequently, propagated uncertainties in the calculated
513 parameters are high. As discussed in section 2.7, the pH- $p\text{CO}_2$ input pair exacerbates these larger
514 uncertainties. Mean $\text{TA}(\text{pH}^{\text{est}}, p\text{CO}_2)$, $\text{DIC}(\text{pH}^{\text{est}}, p\text{CO}_2)$, and $\Omega_{\text{arag}}(\text{pH}^{\text{est}}, p\text{CO}_2)$, $\pm u_c$ (Orr et al.,
515 2018) are $2173 \pm 281 \mu\text{mol kg}^{-1}$, $2111 \pm 263 \mu\text{mol kg}^{-1}$, and 0.94 ± 0.23 , respectively, when
516 input uncertainties are the standard uncertainty (Equation 1). When the input uncertainty for
517 pH^{est} is only the RMSE of 0.0161 (section 3.1), uncertainties decrease to $\pm 98 \mu\text{mol kg}^{-1}$, ± 93
518 $\mu\text{mol kg}^{-1}$, and ± 0.09 , respectively. When input uncertainties are only the random component of
519 the input parameters (i.e. standard deviation for $\text{pH}_{\text{SeaFET}}$ and $p\text{CO}_2$ and instrument precision for

520 T and S), $TA(pH_{\text{SeaFET}}, pCO_2)$, $DIC(pH_{\text{SeaFET}}, pCO_2)$, and $\Omega_{\text{arag}}(pH_{\text{SeaFET}}, pCO_2)$ μc drops to ± 38
521 $\mu\text{mol kg}^{-1}$, $\pm 37 \mu\text{mol kg}^{-1}$, and ± 0.06 , respectively. Given the above uncertainties and that we
522 do not see significant biofouling at the CEO site, we believe that short term variability can be
523 discussed with confidence with this dataset. In other words, wiggles in the data represent real
524 events, despite the high uncertainty in the precise value of the calculated parameters.

525

526 **4.3 Subsurface biogeochemical drivers of pH, Ω_{arag} , and pCO_2**

527 Inorganic carbon chemistry can be influenced by advection and vertical entrainment of
528 different water masses, temperature, salinity, biogeochemistry, and conservative mixing with TA
529 and DIC freshwater endmembers. Here, we followed Rheuban et al. (2019) and separated the
530 drivers of the observed large pH, Ω_{arag} , and pCO_2 variability to provide additional insights into
531 our timeseries (Figure 10) using CO2SYS by altering input parameters temperature, salinity, TA,
532 and DIC. Anomalies (black) relative to the reference values $pH(T_0, S_0, DIC_0, TA_0)$, $\Omega_{\text{arag}}(T_0, S_0,$
533 $DIC_0, TA_0)$, and $pCO_2(T_0, S_0, DIC_0, TA_0)$, were calculated using a linear Taylor series
534 decomposition, adding up the thermodynamic effects of temperature and salinity, and the
535 perturbations due to biogeochemistry, and conservative mixing with freshwater DIC and TA
536 endmembers (Rheuban et al., 2019). Reference values T_0 , S_0 , DIC_0 , and TA_0 , are the mean of the
537 CEO timeseries. Freshwater from sea ice melt and meteoric sources (precipitation and rivers)
538 may influence the CEO site. TA and DIC concentrations of $450 \mu\text{mol kg}^{-1}$ and $400 \mu\text{mol kg}^{-1}$,
539 respectively, have been measured in Arctic sea ice (Rysgaard et al., 2007). Riverine input along
540 the Gulf of Alaska tends to have lower TA ($366 \mu\text{mol kg}^{-1}$) and DIC ($397 \mu\text{mol kg}^{-1}$)
541 concentrations (Stackpoole et al., 2016, 2017) than rivers draining into the Bering, Chukchi, and
542 Beaufort Seas (TA = $1860 \mu\text{mol kg}^{-1}$, DIC = $2010 \mu\text{mol kg}^{-1}$, Holmes et al., 2021) all of which

543 can influence the CEO site to some extent (Asahara et al., 2012; Jung et al., 2021). In this Taylor
544 decomposition we used sea ice TA and DIC endmembers (Rysgaard et al., 2007) but want to
545 emphasize that using Arctic river endmembers did not meaningfully change the results (not
546 shown). Figure 10 shows the effects of biogeochemical processes, temperature, salinity, and
547 conservative mixing with TA and DIC freshwater endmembers on pH, Ω_{arag} , and $p\text{CO}_2$. The
548 effects of salinity (turquoise) and conservative mixing with TA and DIC freshwater endmembers
549 (green) are negligible for pH, Ω_{arag} , and $p\text{CO}_2$. Temperature varied between -1.7 °C during the
550 sea ice covered months and up to 4 °C in late fall, when wind events mixed the whole water
551 column and entrained warm and low $p\text{CO}_2$ surface waters to the instrument depth at 33 m (see
552 section 3.2 for a more in-depth discussion of these mixing events). During this time, the increase
553 in temperature counteracted the effect of biogeochemistry slightly and increased $p\text{CO}_2$ and
554 decreased pH (Figure 10 a,c). Temperature did not affect Ω_{arag} .

555 Biogeochemistry (photosynthesis, respiration, calcification, dissolution) is the most
556 important driver of the inorganic carbon dynamics at 33 m depth at the CEO site. The springtime
557 relaxation events in 2018 and 2019 with relatively higher pH and Ω_{arag} , and lower $p\text{CO}_2$, were
558 mainly driven by biogeochemistry (Figure 10, magenta). During these events O_2 increased and
559 NO_3 decreased, suggesting photosynthetic activity (Figure 2d, e and 3a). Near bottom
560 photosynthetic activity by phytoplankton or sea ice algae has been observed at different locations
561 across the Chukchi Sea (Arrigo et al., 2017; Ouyang et al., 2022; Stabeno et al., 2020; Koch et
562 al., 2020). Sediment trap data from a CEO deployment prior to the start of this $p\text{CO}_2$ and pH
563 time-series suggest that export of the exclusively sympagic sea ice algae *Nitzschia frigida* peaked
564 in May and June, during snow and ice melt events (Lalande et al., 2020), further supporting the
565 hypothesis that sea ice algae contributed to the CO_2 draw down. Interestingly, TA also increased

566 significantly during these events in 2018 and 2019, which cannot be solely attributed to organic
567 matter production. Specifically, TA increased by 23 $\mu\text{mol kg}^{-1}$ in 2019 (Figure 11a). However,
568 with an observed NO_3 decrease of 7.6 $\mu\text{mol kg}^{-1}$, we would expect an increase of TA by 7.6 μmol
569 kg^{-1} . This is assuming that NO_3 is the primary source of nitrogen during organic matter
570 formation, and that assimilation of 1 μmol of NO_3 leads to an increase of TA of 1 μmol (Wolf-
571 Gladrow et al., 2007). The TA increase of 23 $\mu\text{mol kg}^{-1}$ is therefore larger than expected from
572 organic matter formation alone and is likely due to CaCO_3 mineral dissolution. While direct
573 evidence is missing, the strong TA increase suggests that CaCO_3 mineral dissolution during sea
574 ice break up also plays an important role at the CEO site. As observed in other Arctic areas, it is
575 possible that ikaite crystals that were trapped in the ice matrix dissolved in the water column
576 when sea ice melted (Rysgaard et al., 2012, 2007).

577

578 **4.4 Progression of ocean acidification in the Chukchi Sea**

579 Organisms living at the CEO site may have always been exposed to large seasonal
580 variability and low pH and Ω_{arag} (high $p\text{CO}_2$), but the combined and cumulative effects of
581 climate change and ocean acidification have rapidly made these conditions more extreme and
582 longer lasting. Ocean acidification serves as a gradual environmental press by increasing the
583 system's mean and extreme $p\text{CO}_2$ and decreasing mean and extreme pH and Ω_{arag} . Climate
584 induced changes to other important controls of the inorganic carbon system, such as sea ice,
585 riverine input, temperature, and circulation can act as sudden pulses and further modulate the
586 inorganic carbon system to a less predictable degree and cause extreme events (Woosley and
587 Millero, 2020; Orr et al., 2022; Hauri et al., 2021; Qi et al., 2017). Huntington et al. (2020)
588 describe a sudden and dramatic shift of the physical, biogeochemical and ecosystem conditions

589 in the Chukchi and Northern Bering seas in 2017. For example, satellite data for the CEO site
590 illustrate that the longest open water seasons on record occurred between 2017 and 2020. Before
591 2017, the open water season was on average 81 (\pm 40) days long (i.e., below 15 %
592 concentration), of which 60 (\pm 44) days were ice free, whereas between 2017 and 2020, the low
593 sea ice period was 157 (\pm 30) days long, of which 152 (\pm 24) days were ice free (Figure 12). Sea
594 ice decline and increased nutrient influx has also promoted increased phytoplankton primary
595 production in the area (Lewis et al., 2020; Arrigo and van Dijken, 2015; Payne et al., 2021).
596 Since our inorganic carbon timeseries started after the “dramatic shift” that was observed in the
597 Chukchi Sea in 2017 (Huntington et al., 2020) and given the uncertainty in model output in this
598 region, we can only speculate about how the changes in sea ice, temperature and biological
599 production may have affected seasonal variability and extremes of the inorganic carbon
600 chemistry at the CEO site. However, since the summertime low pH and Ω_{arag} and high $p\text{CO}_2$ are
601 tightly coupled to the length of the ice-free period and intensity of organic matter production, it
602 is possible that the observed summertime period of extreme conditions may have been
603 previously unexperienced at this site. We therefore think it is justified to call the spikes of pH
604 and Ω_{arag} “ocean acidification relaxation events”, since the long-lasting summertime period of
605 extremely low pH and Ω_{arag} may be a new pattern.

606

607 **4.5 Relevance for ecosystem**

608 Marine organisms are exposed to a wide range of naturally fluctuating environmental
609 conditions such as temperature, salinity, carbonate chemistry and food concentrations that
610 together constitute their ecological niche. As evolution works toward adaptation, the tolerance
611 range of species and ecosystems to such parameters varies between locations and is often closely

612 related to niche status (Vargas et al., 2022). Stress can be defined as a condition evoked in an
613 organism by one or more environmental and biological factors that bring the organism near or
614 over the limits of its ecological niche (after Van Straalen, 2003). The consequence of the
615 exposure to a stressor will depend on organismal sensitivity, stress intensity (how much it
616 deviates from present conditions) and stress duration. In a synthesis of the global literature on the
617 biological impacts of ocean acidification, Vargas et al. (2017, 2022) showed that the extreme of
618 the present range of variability of carbonate chemistry is a good predictor of species sensitivity.
619 In other words, larger deviations from present extreme high $p\text{CO}_2$ or extreme low pH, would be
620 expected to exert more negative biological impacts. Organismal stress and niche boundaries have
621 implications for the definition and understanding of controls and future ocean acidification
622 conditions in experiments aimed at evaluating future biological impacts.

623 Our data provide insights on conditions that affect and determine local species'
624 ecological niches, and a necessary key is to evaluate or re-evaluate their sensitivity to present and
625 future carbonate chemistry conditions, particularly for the sessile benthic calcifiers that constitute
626 prey for mobile and upper trophic level taxa. For example, an experimental study on three
627 common Arctic bivalve species (*Macoma calcarean*, *Astarte montagui* and *Astarte borealis*)
628 collected in the CEO concluded that these species were generally resilient to decreasing pH
629 (Goethel et al., 2017). However, only two pH were compared (a “control” (pH of 8.1) and an
630 “acidified” treatment (pH of 7.8) and our results show that organisms are already experiencing
631 more extreme conditions today than have been experimentally manipulated. While these data
632 provide insights on these species' plasticity to present pH conditions, they cannot be used to infer
633 sensitivity to future ocean acidification or extremes of current conditions. Based on the local
634 adaptation hypothesis (Vargas et al. 2017, 2022), stress and associated negative effect on species

635 fitness can be expected when pH deviates from the extreme of the present range of variability
636 (pH<7.5) as shown in other regions (e.g. echinoderms: Dorey et al., (2013); crustaceans: Thor
637 and Dupont, (2015); bivalves: Ventura et al., (2016)).

638 At the CEO, our results show sustained periods of remarkably low pH (e.g., 7.5; summer
639 to fall, winter). Higher pH values are observed in spring and late fall. While we are lacking the
640 local biological data to sufficiently evaluate past and future ecosystem changes, a high rate of
641 ocean acidification as observed in the Chukchi Sea (Qi et al., 2022b, a), associated with potential
642 temperature-induced shifts in the carbonate chemistry cycle (e.g. Orr et al. 2022), have the
643 potential to impact species and ecosystems. Exposure to low pH increases organismal energy
644 requirements for maintenance (e.g. acid-base regulation: Stumpp et al., 2012, compensatory
645 calcification: Ventura et al., 2016). Organisms can cope with increased energy costs using a
646 variety of strategies, ranging from individual physiological to behavioral responses, depending
647 on trophic level, mobility, and other ecological factors. For example, they can use available
648 stored energy to compensate for increased costs or they can decrease their metabolism to limit
649 costs (AMAP 2018). At the CEO, the low pH period observed during the summer and fall is
650 associated with elevated temperature and an elevated food supply for herbivores (Lalande et al.,
651 2020). The high availability of food may then foster compensation for the higher energetic costs
652 associated with exposure to low pH. However, a longer period of low pH as suggested by our
653 data could lead to a mismatch between the low pH and food availability, with cascading negative
654 consequences for the ecosystem (Kroeker et al., 2021). In winter, the low pH conditions are
655 associated with low temperature, no light, and low food level concentrations. These conditions
656 are likely to keep metabolisms low and limit the negative effects of exposure to low pH
657 (Gianguzza et al., 2014). As food availability is limited by the absence of light, this strategy may

658 be compromised by an increase in temperature that could also lead to increased metabolism.
659 Additional work is needed to understand impacts of acidification conditions and variability on
660 the marine biota of the Chukchi Sea, including field and laboratory experiments that evaluate
661 biological response under realistic scenarios. The characterization of the environmental
662 conditions at the CEO, including the variability in time, can be used to design single and multiple
663 stressor experiments (carbonate chemistry, temperature, salinity, food, oxygen; Boyd et al.
664 2018).

665 Indigenous communities are at the forefront of the changing Arctic, including changes in
666 accessibility, availability, and condition of traditional marine foods (Buschman and Sudlovenick,
667 2022; Hauser et al., 2021). Several marine species are critical to the food and cultural security of
668 coastal Inupiat who have thrived in Arctic Alaska for millenia. While it is not possible to resolve
669 the consequences of the seasonal and interannual variations in carbonate chemistry documented
670 in this manuscript without a proper sensitivity evaluation, the seasonally low pH conditions have
671 the potential to impact organisms like bivalves in a foraging hotspot for walrus (Jay et al., 2012;
672 Kuletz et al., 2015). Walrus, as well as their bivalve stomach contents, are important nutritional,
673 spiritual, and cultural components, raising concerns for food security in the context of ecosystem
674 shifts associated with the variability and multiplicity of climate impacts within the region (ICC,
675 2015).

676

677 **5. Concluding Thoughts**

678 The Chukchi Sea is undergoing a rapid environmental transformation with potentially
679 far-reaching consequences across the ecosystem. While we are lacking a long-term time-series,
680 we used this dataset to investigate the drivers of extreme pH, Ω_{arag} , and $p\text{CO}_2$ and document

681 conditions that could affect the ecological niches of organisms, including a fast rate of ocean
682 acidification, elongated sea ice free periods, increased primary productivity and elevated
683 temperature. While a combination of experimental and monitoring approaches is needed for an
684 understanding of the ecological consequences of these changes, our results also highlight the
685 urgency to mitigate CO₂ emissions and simultaneously support Indigenous-led conservation
686 measures to safeguard an ecosystem in transition. Indigenous People in the Arctic have
687 established strategies to monitor, adapt to, and conserve the ecosystems upon which they depend.
688 Ethical and equitable engagement of Indigenous Knowledge and the communities at the forefront
689 of climate impacts can help guide research and conservation action by centering local priorities
690 and traditional practices, thereby supporting self-determination and sovereignty (Buschman and
691 Sudlovenick, 2022).

692

693 **Data availability**

694 The inorganic carbon data used in this manuscript are publicly available (Hauri and
695 Irving, 2023a; Hauri and Irving, 2023b).

696

697 **Author contributions**

698 CH and BI managed and serviced the HydroC CO₂ and SeapHOx sensors, analyzed and
699 published the data, and wrote the manuscript. SD and Peter Shipton carried out the CEO mooring
700 deployments and recoveries and managed and serviced the CTD and NO₃ sensors. RP, DH, SD,
701 and SLD contributed to the manuscript.

702

703 **Competing interests**

704 The authors have no competing interests.

705

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1404 **Tables**

1405 **Table 1.** Chukchi Ecosystem Observatory location and instrument sampling frequency. Sensor
 1406 type and parameter measured (*italicized*) shown in top row. Values in parenthesis indicate the
 1407 number of measurements averaged over the measurement interval window.

Deployment	Latitude	Longitude	SUNA <i>NO₃</i>	HydroC CO₂ <i>pCO₂</i>	SBE16 <i>CTD+</i>	SBE37 <i>CTD</i>	SeaFET <i>pH</i>	SBE63 <i>O₂</i>
2016-2017	71°35'58.5600"N	161°31'06.2400"W	1 h	12 h (300/5 min)*	1 h	-	-	-
2017-2018	71°35'58.9200"N	161°31'08.0400"W	1 h	12 h (5/5 min)	2 h	2 h	2 h (30/5 min)	2 h
2018-2019	71°35'59.6400"N	161°31'41.1600"W	1 h	24 h (5/5 min)	1 h	2 h*	-	2 h*
2019-2020	71°35'58.9200"N	161°31'39.0000"W	1 h	12 h (5/5 min)	2 h	-	-	-

* indicates the sensor did not return data over the whole year due to battery failure

CTD+ indicates ancillary data was available with the SBE16 file (e.g., chlorophyll fluorescence)

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1411 **Table 2.** Evaluation of pH_{SeaFET} and pH^{est} using reference pH from nearby discrete samples

1412 ($pH^{\text{disc}}_{\text{calc}}$). Uncertainty, u_c , is the propagated combined standard uncertainty from *errors.m* (Orr

1413 et al., 2018). $\text{pH}_{\text{SeaFET}}$ and pH^{est} were interpolated to the discrete timestamp. Figure S1 for

1414 visualization of reference values.

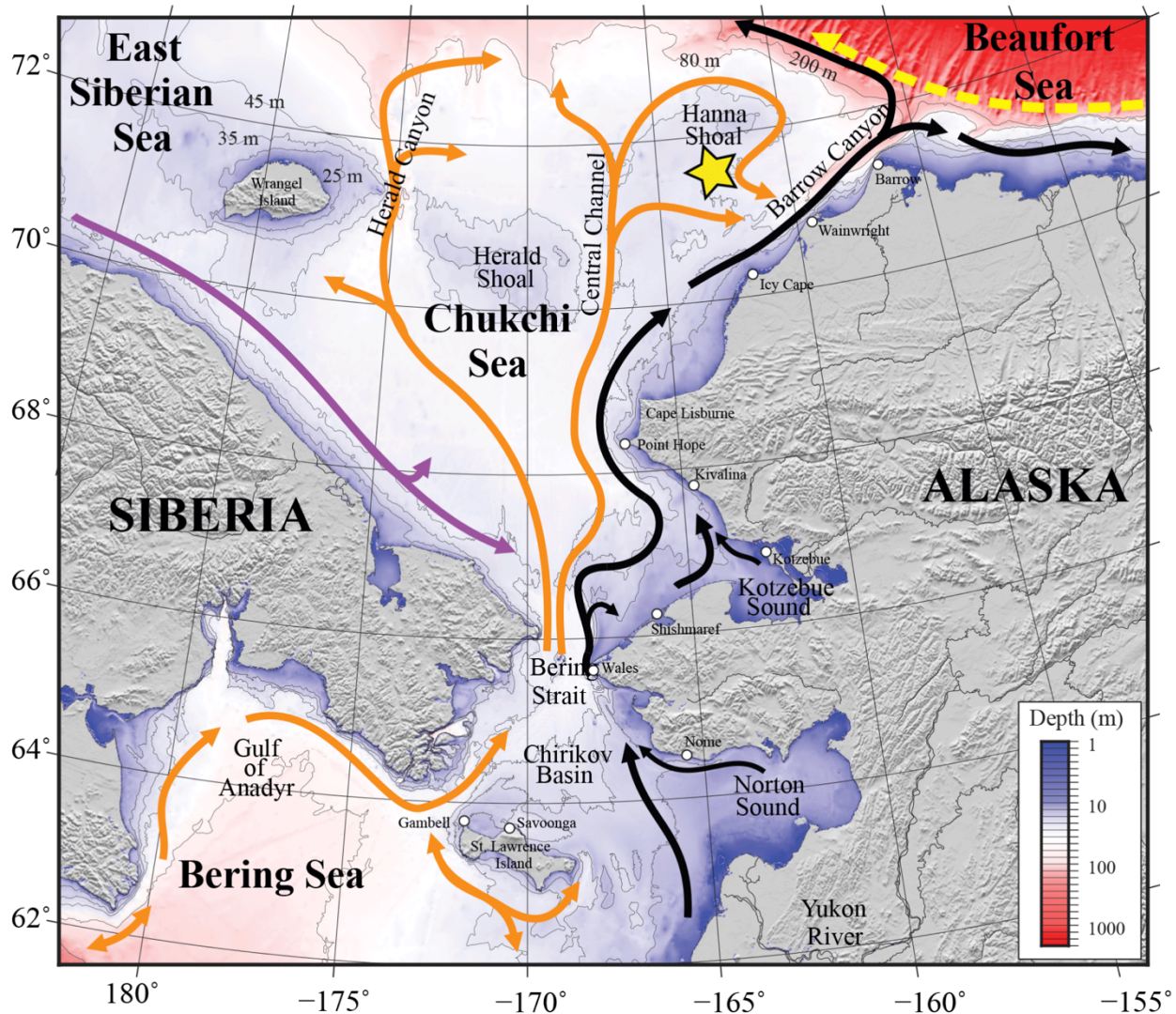
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Date	Cruise	Cast No.	Distance (km)	$\text{pH}_{\text{calc}}^{\text{disc}} \pm u_c$	Anomaly ($\text{pH}^{\text{est}} - \text{pH}_{\text{calc}}^{\text{disc}}$)	Anomaly ($\text{pH}_{\text{SeaFET}} - \text{pH}_{\text{calc}}^{\text{disc}}$)	Source
2017-09-10	HLY1702	127	0.52	8.0123 ± 0.0166	-0.0450*	-0.0354	Cross et al., 2020a
2019-08-11	HLY1901	39	3.75	7.6423 ± 0.012	0.0079*	-	Cross et al., 2021
2019-08-19	OS1901	33	0.27	7.7367 ± 0.0145	-0.0200	-	unpublished

* indicates $\text{pH}_{\text{calc}}^{\text{disc}}$ was interpolated to mooring depth

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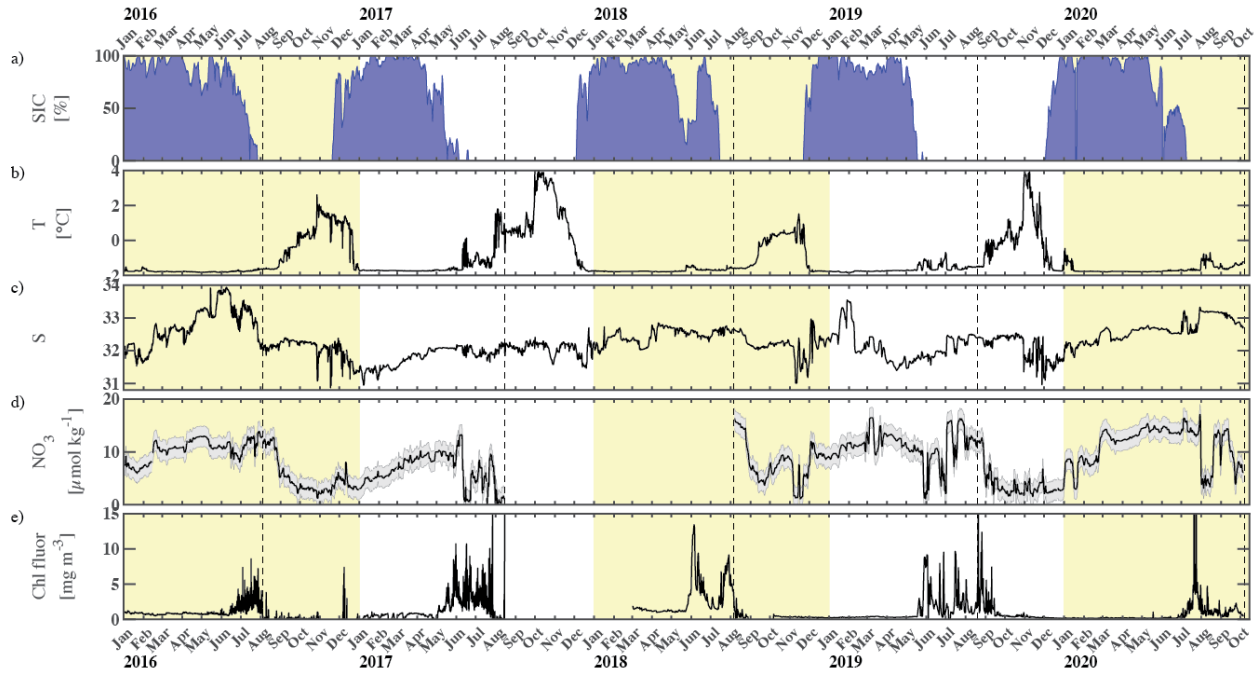
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 1420 **Figure 1. Map of the study area.** Bathymetry of the Chukchi, northern Bering, East Siberian
 1421 and eastern Beaufort seas is shown in color. The Chukchi Ecosystem Observatory (CEO)
 1422 location near Hanna Shoal is marked with a yellow star. General circulation patterns are shown
 1423 with arrows: black – Alaskan Coastal Water and Alaskan Coastal Current, dividing into the
 1424 Shelf-break Jet (right) and Chukchi Slope Current (left, Corlett and Pickart, (2017)); orange –
 1425 Anadyr, Bering, and Chukchi Seawater; purple – Siberian Coastal Current; yellow – Beaufort
 1426 Gyre boundary current. Figure is from Hauri et al. (2018).

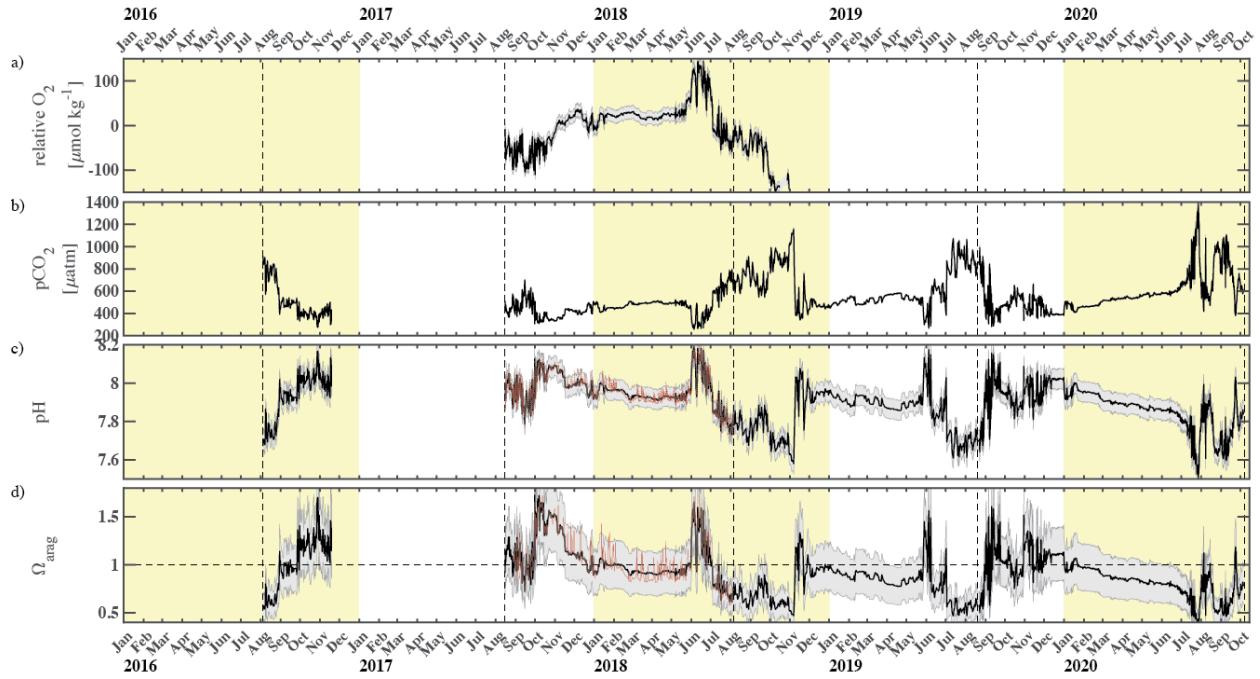
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1430 **Figure 2. Chukchi Ecosystem Observatory timeseries from 2016 through 2020.** a) sea ice
1431 concentration (blue shading to highlight coverage, %; DiGirolamo et al., 2022), b) temperature
1432 ($^{\circ}\text{C}$), c) salinity, d) NO_3 with uncertainty envelope ($\mu\text{mol kg}^{-1}$), and e) chlorophyll fluorescence
1433 (mg m^{-3}). Years are indicated by alternating yellow and white background shading. The vertical
1434 dotted gray lines indicate the mooring turn around timing.



1435

1436 **Figure 3. Chukchi Ecosystem Observatory timeseries from 2016 through 2020, part 2. a)**

1437 relative dissolved oxygen with uncertainty envelope (relative to the mean; $\mu\text{mol kg}^{-1}$), b) $p\text{CO}_2$

1438 with uncertainty envelope (μatm ; Hauri and Irving, 2023a), c) pH with uncertainty envelope

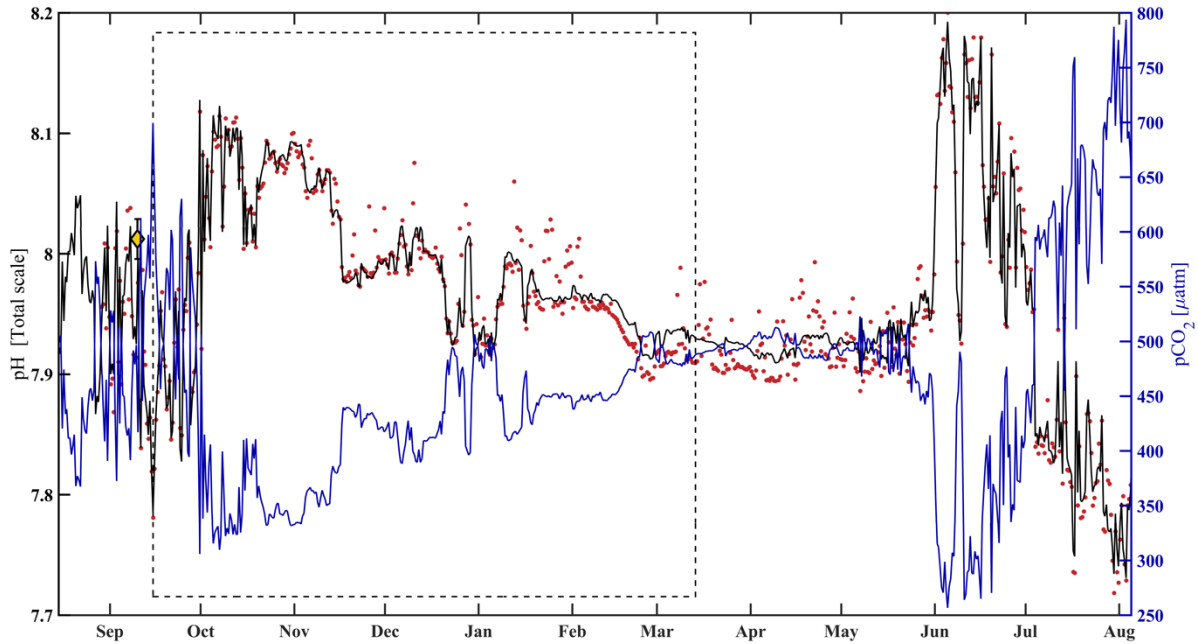
1439 (pH^{est} in black, $\text{pH}_{\text{SeaFET}}$ in red; Hauri and Irving 2023b), and d) aragonite saturation state with

1440 uncertainty envelope ($\Omega_{\text{arag}}(p\text{CO}_2, \text{pH}^{\text{est}})$ in black; $\Omega_{\text{arag}}(p\text{CO}_2, \text{pH}_{\text{SeaFET}})$ in red). Years are

1441 indicated by alternating yellow and white backgrounds. The vertical dotted gray lines indicate

1442 the mooring turn around timing.

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1446 **Figure 4. HydroC $p\text{CO}_2$ and pH highlighting mirrored trend from mid-August 2017 to**

1447 **beginning of August 2018.** Measured pH ($\text{pH}_{\text{SeaFET}}$, red dots) is interpolated onto the HydroC

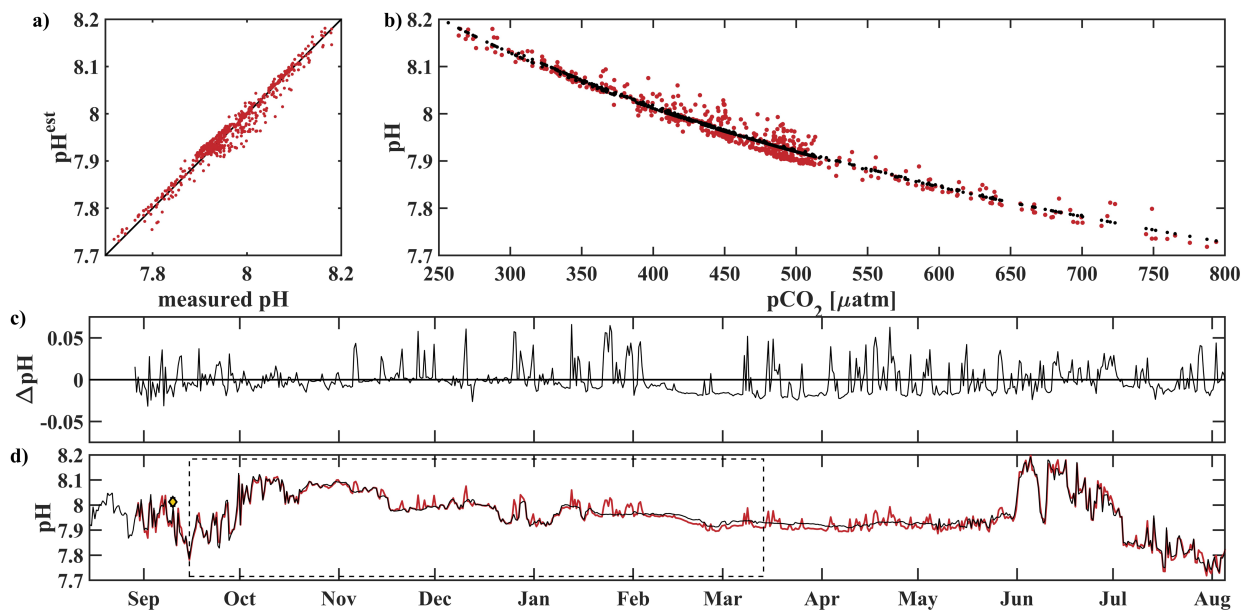
1448 $p\text{CO}_2$ timestamp (blue), and pH^{est} is shown as the solid black line. The dashed box shows the

1449 period over which pH^{est} was trained. The yellow faced diamond with error bars show reference

1450 $\text{pH}^{\text{disc}}_{\text{calc}} \pm u_c$ (Table 2; Cross et al., 2020a; Orr et al., 2018).

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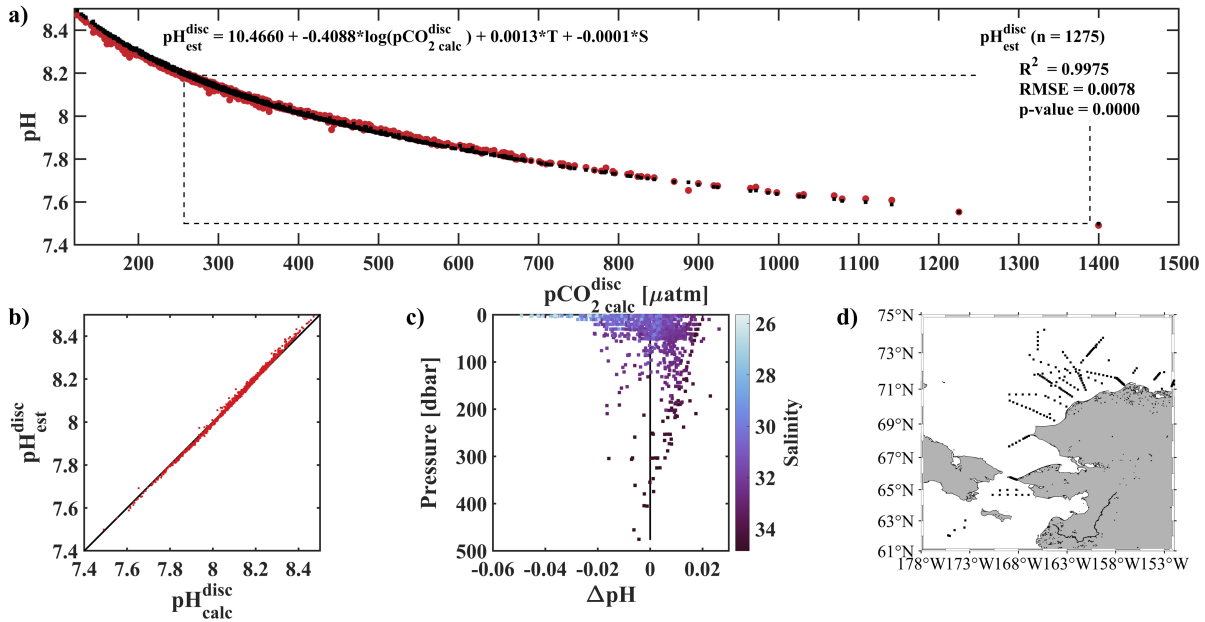
1454 **Figure 5. Performance of the pH algorithm.** (a) pH_{SeaFET} vs pH^{est} with black line highlighting
1455 1:1 ratio, (b) pCO_2 vs pH_{SeaFET} (red) and pCO_2 vs pH^{est} (black), (c) residual pH ($pH_{SeaFET} -$
1456 pH^{est}), and (d) pH_{SeaFET} (red) and pH^{est} (black) vs. time, with dashed box highlighting the period
1457 over which pH^{est} was trained (15 September - 14 March 2017), and the yellow faced diamond
1458 with error bars showing reference $pH^{disc}_{calc} \pm u_c$ (Table 2; Cross et al., 2020).

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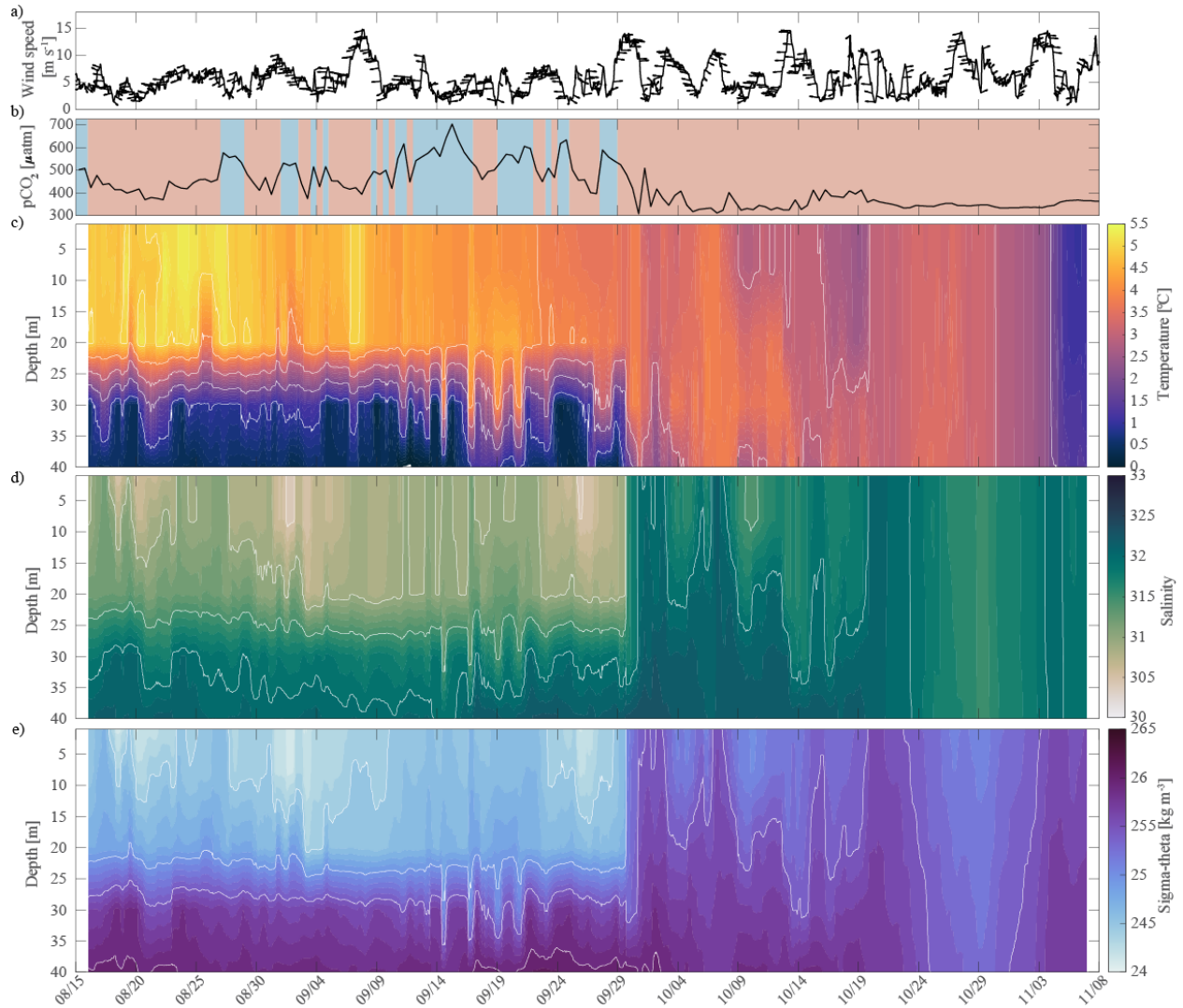


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1464 **Figure 6. Evaluation of the pH algorithm.** pH^{est} evaluation with $\text{pH}^{\text{disc}}_{\text{calc}}$ from discrete
 1465 samples collected during 4 cruises in the fall or early winter (August - November) of 2017-2020
 1466 and $\text{pH}^{\text{disc}}_{\text{est}}$ from our linear regression model (Equation 2). (a) $p\text{CO}_2^{\text{disc}}_{\text{calc}}$ (TA, DIC) vs pH (red
 1467 $\text{pH}^{\text{disc}}_{\text{calc}}$ and black $\text{pH}^{\text{disc}}_{\text{est}}$) with dashed black box showing the range of pH and $p\text{CO}_2$ observed
 1468 at the CEO at 33 m depth, (b) $\text{pH}^{\text{disc}}_{\text{calc}}$ vs $\text{pH}^{\text{disc}}_{\text{est}}$ with black 1:1 ratio, (c) residual pH ($\text{pH}^{\text{disc}}_{\text{calc}} -$
 1469 $\text{pH}^{\text{disc}}_{\text{est}}$) vs depth with color shading by salinity and black vertical line at 0, and (d) map showing
 1470 the locations of the 1275 discrete water samples used for evaluation (Monacci et al., 2022; Cross
 1471 et al., 2021; 2020a; 2020b).

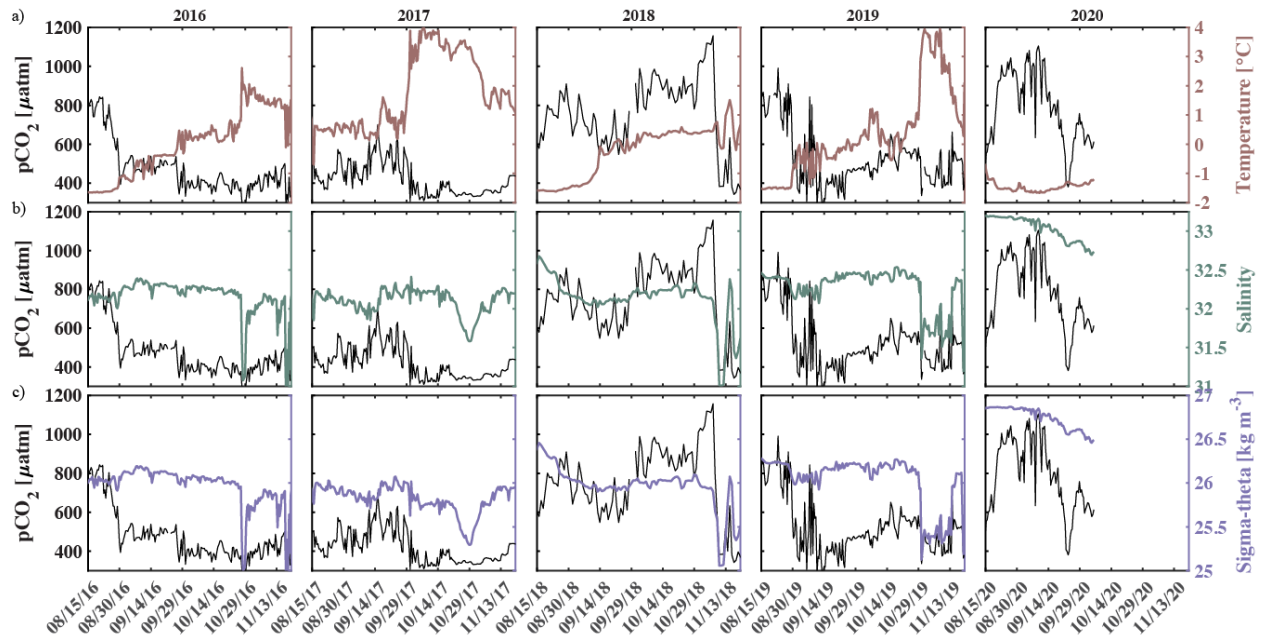
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1475 **Figure 7. Water column structure from late summer 2017 to freeze up.** Profiles of a) wind
 1476 speed and direction (arrows pointing downwind) from the NOAA-operated Wiley Post-Will
 1477 Rogers Memorial Airport, b) $p\text{CO}_2$ (μatm) with blue background indicating the water was
 1478 undersaturated regarding aragonite ($\Omega_{\text{arag}} < 1$) and red shading indicating aragonite
 1479 oversaturation ($\Omega_{\text{arag}} \geq 1$), c) temperature ($^{\circ}\text{C}$), d) salinity, and e) sigma-theta (kg m^{-3}).
 1480 Temperature (c) and salinity (d) were measured at 8, 20, 30, and 40 m by the Chukchi Ecosystem
 1481 Observatory freeze-up detection mooring deployed in fall 2017. Density was calculated with the
 1482 TEOS-10 GSW Oceanographic Toolbox (McDougall and Baker, 2011).



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1484 **Figure 8. Impact of water column mixing on $p\text{CO}_2$.** Timeseries of $p\text{CO}_2$ (black, left axis) and
 1485 a) temperature (maroon, right axis), b) salinity (green, right axis), and c) density (purple, right
 1486 axis) for 15 August to 1 December in 2016 -2020 measured at ~33m septh at the Chukchi Sea
 1487 Ecosystem Observatory.

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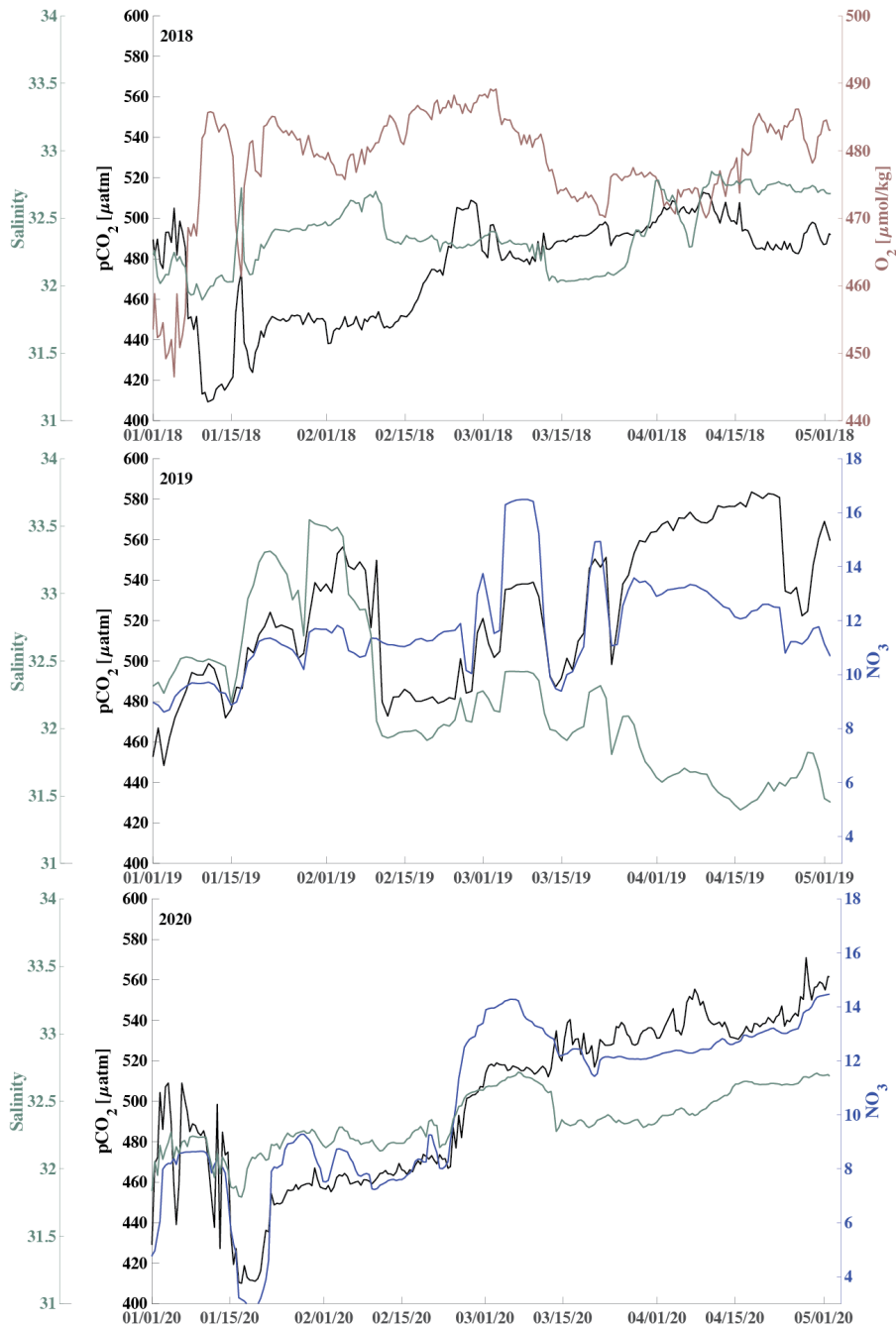
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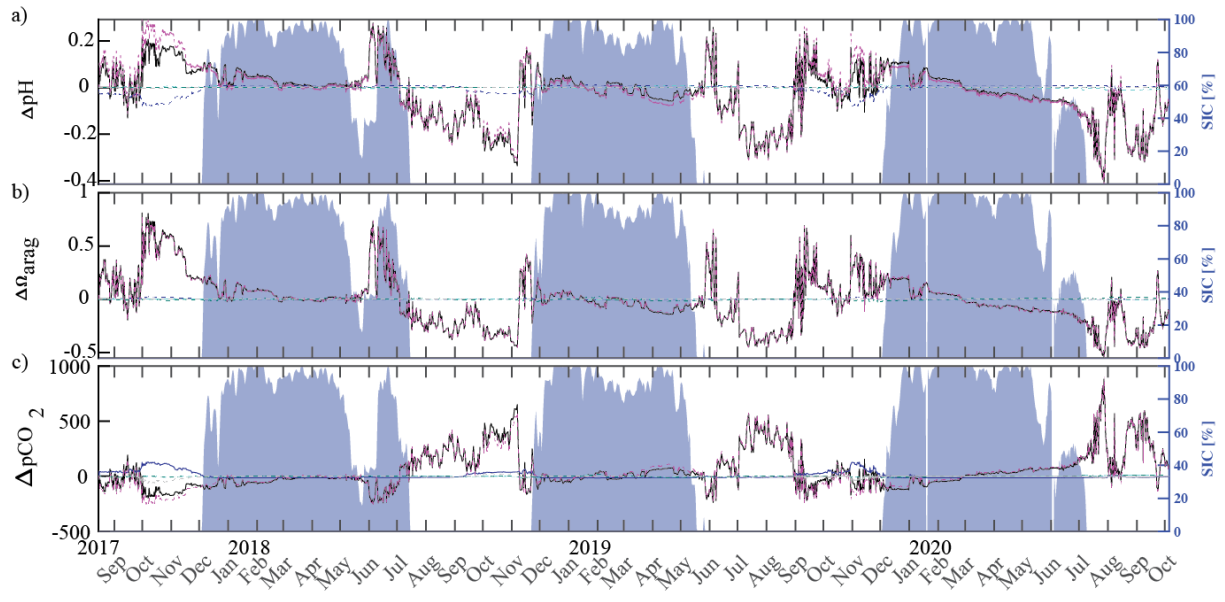
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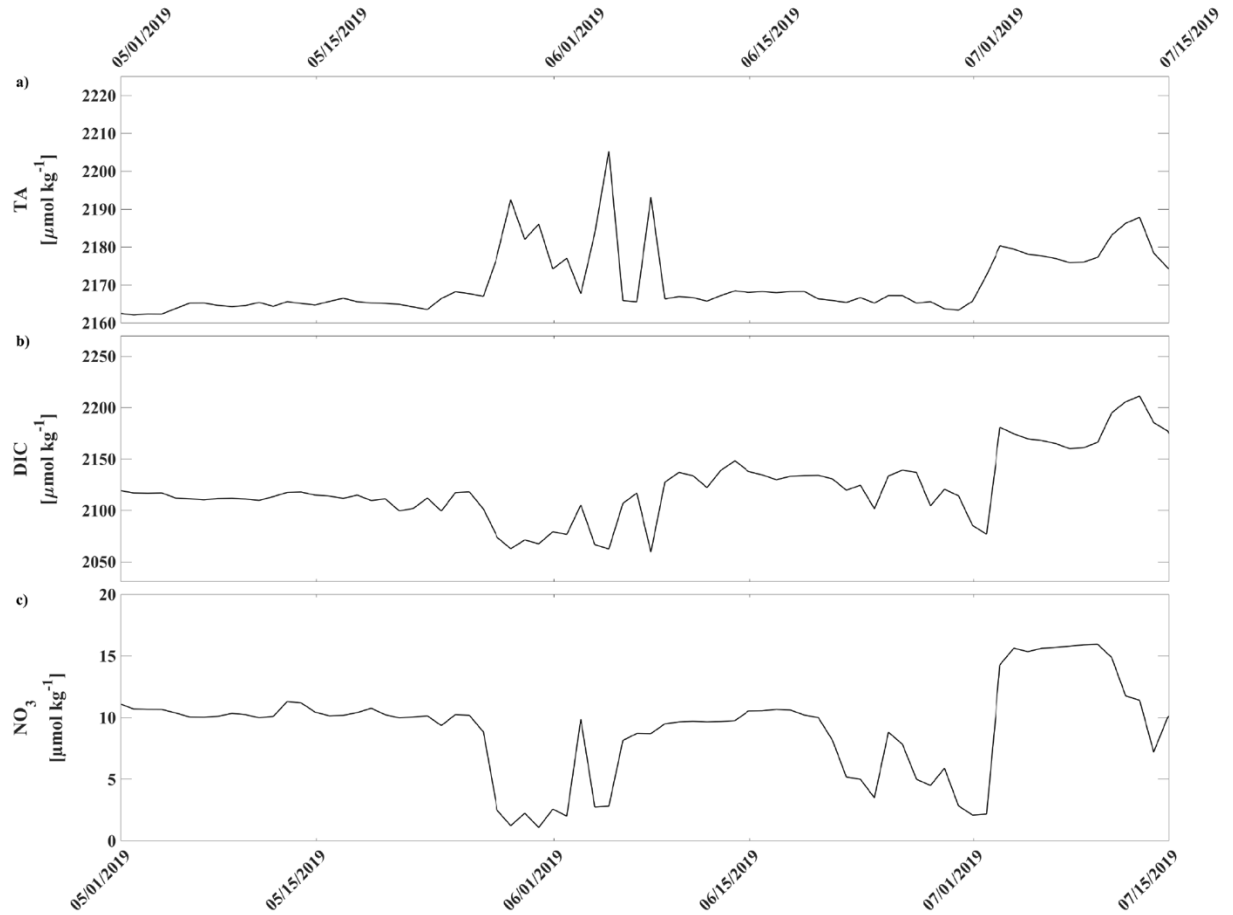
1496 **Figure 9. Respiration under the sea ice.** Timeseries of $p\text{CO}_2$ (black) and salinity (green, left
 1497 axis), and oxygen (O_2 , $\mu\text{mol kg}^{-1}$, maroon, top) and nitrate (NO_3 , $\mu\text{mol kg}^{-1}$, blue, middle and
 1498 bottom) concentration (right axis during January through April for 2018 (top), 2019 (middle) and
 1499 2020 (bottom)).



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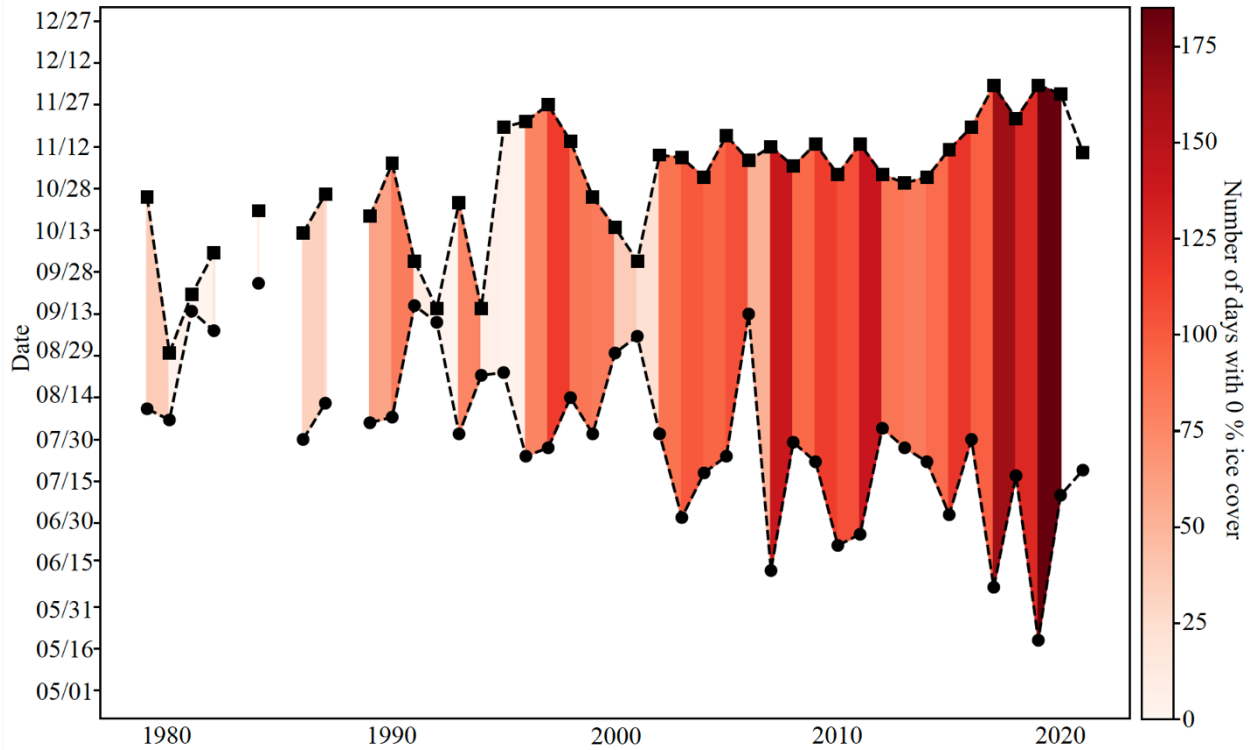
1501 **Figure 10. Drivers of the inorganic carbon system.** Component timeseries of the linear Taylor
 1502 decomposition of a) pH, b) Ω_{arag} , and c) $p\text{CO}_2$. Contributions of changes in salinity (red),
 1503 temperature (blue), biogeochemistry (pink), and freshwater mixing (green) to changes (black,
 1504 relative to the mean of the timeseries) in pH, Ω_{arag} , and $p\text{CO}_2$ were computed following Rheuban
 1505 et al. (2019). The grey dotted line illustrates an estimated residual term. Sea ice concentration
 1506 (blue shading, %; DiGirolamo et al., 2022) is shown on the right axes.

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1509 **Figure 11. Spring 2019 relaxation event.** Timeseries of a) total alkalinity (TA, $\mu\text{mol kg}^{-1}$), b)
 1510 dissolved inorganic carbon (DIC, $\mu\text{mol kg}^{-1}$), and c) nitrate (NO_3 , $\mu\text{mol kg}^{-1}$) from May 1st, 2019
 1511 through July 15th, 2019.



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1514 **Figure 12. Low sea ice period at the Chukchi Sea Observatory.** Timeseries of start (circle)

1515 and end (square) of low sea ice (< 15 % per grid cell) period from 1982-2021. Shades of red

1516 illustrate number of days with 0 % sea ice cover. The satellite sea ice cover at the observatory

1517 site was taken from the NSIDC (DiGirolamo et al., 2022).

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