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# HESS Opinions: Are soils overrated in hydrology?

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## Abstract

Traditional hydrological theories are based on the assumption that soil is key in determining water's fate in the hydrological cycle. According to these theories, soil hydraulic properties determine water movement in both saturated and unsaturated zones, described by matrix flow formulas such as the Darcy-Richards equations. They also determine plant available moisture and thereby control transpiration. Here we argue that these theories are founded on a wrong assumption. Instead, we advocate the reverse: the terrestrial ecosystem manipulates the soil to satisfy specific water management strategies, which are primarily controlled by its reaction to climatic drivers and by prescribed boundary conditions such as topography and lithology. According to this assumption, soil hydraulic properties are an “effect”, rather than a “cause” of water movement. We further argue that the integrated hydrological behaviour of an ecosystem can be inferred from considerations about ecosystem survival and growth, without relying on internal process descriptions. An important and favourable consequence of this climate and ecosystem-driven approach is that it provides a physical justification for catchment models that do not rely on soil information and on the complexity associated to the description of soil water dynamics. Another consequence is that modelling water movement in the soil, if required, can benefit from

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28 the constraints that are imposed by the embedding ecosystem. Here we illustrate our ecosystem  
29 perspective of hydrological processes and the arguments that support it. We suggest that advancing  
30 our understanding of ecosystem water management strategies is key to building more realistic  
31 hydrological theories and catchment models that are predictive in the context of environmental  
32 change.

## 33 **1 A change in perspective**

34 Soil is important in hydrology. Soil forms the substrate of the terrestrial ecosystem and hence it is  
35 a crucial element of the critical zone of life on Earth (Lin et al., 2006; Banwart et al., 2017).

36 Through its porous structure, exercising capillarity against gravity, it provides water storage  
37 against droughts and nutrients for plant growth.

38 It has been argued that the soil forms an ecosystem in itself, full of micro-biotic and macrobiotic  
39 life (Ponge, 2015; Weil and Brady, 2017). Fungi forming dense underground networks live in  
40 symbiosis with vegetation, exchanging nutrients for carbon, which makes them responsible for the  
41 larger part of subterranean carbon storage (Domeignoz-Horta et al., 2021). Soils are full of life.  
42 Above ground life cannot survive without sub-surface life; they are part of the same ecosystem.

43 Soils are embedded in the terrestrial ecosystems, which through evolution and natural selection,  
44 have found ways to make best use of its resources. The processes and structure of a terrestrial  
45 ecosystem are mainly controlled by external factors which are largely prescribed. Among them,  
46 climate plays a major role, as rainfall patterns and seasonal temperatures strongly affect the  
47 distribution of vegetation types; other external factors include topography, lithology, which  
48 determines parental material, and potential biota (Chapin et al., 2011). Given these boundary  
49 conditions, a terrestrial ecosystem adjust its internal behaviour to satisfy its needs, and it  
50 manipulates the substrate on which it grows.

51 In particular, the soil is the result of a long-term evolution of terrestrial ecosystems given their  
52 boundary conditions. The classic *clorpt* model presented by Hans Jenny's famous 1941 book "The  
53 Factors of Soil Formation" states that  $s = f(cl, o, r, p, t, \dots)$ , where soil properties ( $s$ ) are seen as  
54 a function of climate ( $cl$ ), biotic effects ( $o$  for organisms), topography ( $r$  for relief), parent material

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55 ( $p$ ), time ( $t$ ), additional factors such as fire (represented by the dots) (Huggett, 2023). This model  
56 suggests that soil properties are largely determined by the embedding ecosystems.

57 Managing water is an essential task of terrestrial ecosystems, as water is essential to life. And it is  
58 not a trivial task, as it implies bridging dry weather periods, but also avoiding troubles caused by  
59 sustained or heavy rainfall, such as water stagnation or soil erosion. We argue that terrestrial  
60 ecosystems achieve this balance by manipulating key hydrological characteristics such as  
61 interception capacity, infiltration capacity, moisture storage capacity, preferential pathways to  
62 replenish moisture stocks and recharge, and subsurface drainage. According to this view, a  
63 terrestrial ecosystem manipulates the soil hydraulic properties to satisfy specific water  
64 management strategies.

65 Yet, the most established hydrological theories parameterize water fluxes using soil attributes such  
66 as texture, porosity, moisture retention capacity, wilting point, plant available moisture, etc. (e.g.  
67 Drewniak, 2019, Lu et al., 2019). These theories assume that soil properties are controlling  
68 processes such as infiltration, drainage, or plant evaporation. But this is the wrong way round. Soil  
69 properties are the effect, rather than the cause of water movement, which itself, is governed by the  
70 behaviour of the embedding terrestrial ecosystem.

71 We therefore argue in favour of an ecosystem-based approach where the integrated hydrological  
72 behaviour of an ecosystem is inferred based on its water management strategies needed to survive  
73 and grow, without relying on internal process descriptions. As we shall see, this is not a  
74 prohibitive task. The very existence of an ecosystem already provides many indications about its  
75 ability to manage its water resources.

76 This ecosystem-based approach has several beneficial consequences for hydrology. First, it  
77 provides a physical justification for the development of catchment scale hydrological models that  
78 directly rely on the external factors that influence terrestrial ecosystems, such as climate,  
79 topography, lithology, etc. These models would be more realistic than soil-based models because  
80 based on the correct cause-effect relationships. Moreover, they would be less data demanding and  
81 simpler, as they would not require soil texture information and detailed description of soil water

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82 dynamics. Second, it would allow digging into the small scale, if this is deemed necessary,  
83 exploiting the constraints that are imposed by the behaviour of the larger scale system.

84 In the following, we first present the soil-centred hydrological perspective and its limitations  
85 (Section 2), we then argue that there is limited evidence that soil properties actually matter in  
86 catchment hydrology (Section 3), next we illustrate our terrestrial ecosystem perspective (Section  
87 4), and provide an interpretation of why the soil-based modelling tradition has proliferated in  
88 hydrology (Section 5), finally we illustrate the limitations of our approach (Section 6), and present  
89 our conclusions (Section 7).

## 90 **2 Limitations in the soil-centred hydrological perspective**

### 91 **2.1 Challenges in small-scale theories of soil water dynamics**

92 It is a deeply rooted perception in hydrology that small-scale soil water dynamics are key in  
93 determining the integrated catchment behaviour at larger scales such as the partitioning of rainfall  
94 between evaporation, drainage and storage (Vereecken et al., 2022). For example, soil is assumed  
95 to control plant evaporation, as plant available water content is often parameterized as a function  
96 of soil texture (Yang et al., 2016). Processes such as Hortonian overland flow, saturation excess  
97 overland flow, or percolation are often described in relation to water movement in the unsaturated  
98 zone, using laboratory-scale matrix flow theory developed by soil physicists. This theory describes  
99 flow in porous media based on equations that depend on soil hydraulic properties (e.g. porosity,  
100 hydraulic conductivity). Darcy's law describes matrix flow under saturated conditions through a  
101 porous medium under a head gradient. Richards' equation regards matrix flow under unsaturated  
102 conditions in the vadose zone, determining water flow direction and velocity. Numerous  
103 simplified semi-empirical soil infiltration equations were also derived to simulate the infiltration  
104 excess overland flow, such as the Philip and Horton equations (Schoener et al., 2021). The matrix  
105 flow theory is regarded as well-established, much like classical mechanics.

106 For a hydrological model to be considered "physically-based", it is generally assumed that it needs  
107 to be based on these small-scale theories. Land surface models (LSMs) are strongly based on these  
108 matrix flow equations (Freeze and Harlan, 1969; Lawrence et al., 2019), which determine soil

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109 water movement vertically and laterally (Duffy, 1996; Refsgaard et al., 2022). Even the  
110 representative elementary watershed (REW) approach (Reggiani et al., 1998), a physically based  
111 framework that describes catchment scale processes, is based on the integration of small-scale  
112 conservation equations developed for porous media.

113 This soil-centred perspective is highly rated in the hydrological community. Some of the most  
114 prestigious hydrology awards exemplify the tribute of the hydrological community to this  
115 perspective, such as the Henry Darcy medal of hydrological sciences in the European Geosciences  
116 Union (EGU), the Robert Horton American Geophysical Union (AGU) hydrological science  
117 medal, which are named after two hydrologists that pioneered the soil-centred approach.

118 Tracer field experiments, such as dye and isotope studies, have shown that matrix flow is rarely  
119 observed. Most soils contain crevices, preferential channels, and openings that transmit free water  
120 quite rapidly to the sub-surface, which is termed preferential flow (Beven and Germann, 2013;  
121 McDonnell et al., 2007; Beven, 2018; Zehe et al., 2021). Hence, natural conditions do not  
122 resemble well-prepared homogenous soil that can be recreated in a laboratory.

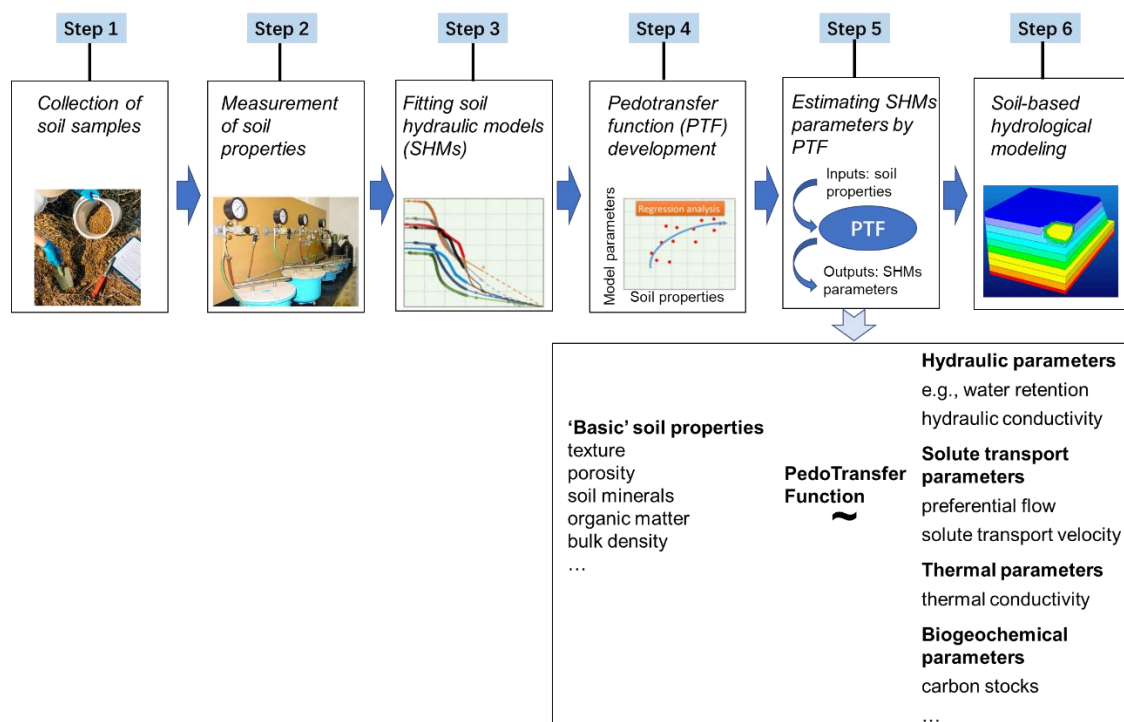
123 In response to this criticism, soil-water theories have become more complex, allowing for  
124 preferential flow, which required even more detailed soil characterizations. These challenges have  
125 stimulated the development of dual-continuum, dual-porosity, or dual-permeability modifications  
126 (Jarvis et al., 2016), most models are still based on matrix flow theory (Weiler, 2017). Because of  
127 the extreme complexity of soil preferential flow in nature, it is extremely hard to develop accurate  
128 models that describe it, even at the plot scale. The challenge is exponentially greater when  
129 upscaling preferential flow from plot-scale to hillslope or catchment scales (Davies et al., 2013;  
130 Germann, 2014; Or, 2020). At the global scale, hyper-resolution land surface models, which are  
131 deemed necessary to addressing critical water cycle science questions and applications, can have  
132 up to  $10^9$  unknowns (Wood et al, 2011)!

133 From its establishment, preferential flow theory was regarded as the main culprit challenging the  
134 foundation of “physically-based” hydrological models. This avenue has led to models that require  
135 many space and soil-dependent parameters that are difficult to measure, that require massive

136 computational resources, and that when calibrated are prone to equifinality. Arguably, the avenue  
 137 of building more complex models by increasingly detailed representation of soil water movement  
 138 is a steep one. But is it a necessary one, if the objective is to build a physically-based model of  
 139 catchment scale hydrological processes?

## 140 2.2 Limitations in the pedotransfer functions approach

141 Soil-centred bottom-up hydrological models rely on estimates of soil hydraulic properties (SHPs),  
 142 such as water retention characteristics and unsaturated and saturated hydraulic conductivity. As  
 143 these properties are difficult to measure at appropriate scales, soil pedotransfer functions (PTFs)  
 144 have been developed to express SHPs as a function of more accessible soil properties, such as soil  
 145 texture (i.e. sand-, silt-, clay- content), organic matter, and bulk density (Figure 1; van Looy et al.,  
 146 2017; Or, 2020; Haghverdi et al., 2020; Gupta et al., 2021; Hohenbrink et al., 2023).



147

148 Figure 1. Schematic illustration of data collection, laboratory measurements, fitting soil hydraulic  
 149 models (SHMs), pedotransfer function (PTF) development, and soil-based hydrological modelling  
 150 workflow (adapted from Van Looy et al., 2017 and Haghverdi et al., 2020).

151 There are several critical issues with the practicality and accuracy of this approach: 1) most soil

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152 property parameters are measured by pedologic surveys, at great expense and efforts (Van Looy et  
153 al., 2017). 2) PTFs are usually obtained by using measurements from uniform soil samples, and  
154 performed in laboratory-scale experiments, which merely reflect disturbed and therefore unnatural  
155 conditions; 3) the parameters obtained at the laboratory scale are not necessarily the same as at the  
156 model scale, which requires upscaling assumptions, which are difficult to verify or recalibrate,  
157 hampered by equifinality.

158 Unfortunately, readily-available soil information (e.g., texture, bulk density, organic matter)  
159 correlates poorly with soil hydraulic properties. Gutmann and Small (2007) have shown that soil  
160 textural classes, across a range of climates and vegetation covers, merely explained 5% of the  
161 variance of real SHPs. In another study, it was found that 95% of the default soil hydraulic  
162 parameters in a state-of-the-art land surface model, largely based on soil textural data, were  
163 significantly different from region-specific observations (Kishné et al., 2017).

164 Recent studies showed that in order to achieve more realistic estimates of soil hydraulic properties  
165 it is necessary to include information about vegetation or biophysical activity (Or, 2020). For  
166 example, Bonetti et al. (2021) proposed soil structure corrections into pedotransfer functions,  
167 informed by remote-sensing vegetation metrics and local soil texture. Additional studies  
168 "rebalance" the soil texture information and highlight the importance of soil structure, originated  
169 by soil biophysical activity (Or, 2020; Fatichi et al., 2020). Not only the "physically-based"  
170 models, but also the empirical soil-based models, for example the soil conservation service (SCS)  
171 method in the SWAT model (soil water assessment tool), involve land-use data to "rebalance" the  
172 soil-based curve number in catchment simulations (Arnold et al., 2012).

173 Building realistic pedotransfer functions requires detailed characterization of the soil, requiring a  
174 large number of parameters that are difficult to estimate. This approach, while feasible for a  
175 hillslope or a headwater catchment, becomes impractical at regional or global scales. For  
176 hydrological purposes, the ultimate goal is often to determine integrated fluxes of hydrological  
177 response at large scales. Hence, it is worth asking: can this integrated behaviour be determined  
178 directly from observations, without resorting to small-scale theories and upscaling assumptions?

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### 179 **3 Does soil variability matter in catchment hydrology?**

#### 180 **3.1 Do soil-centred models reproduce hydrological variability?**

181 A key objective of "physically based" models is to represent hydrological variability, such as  
182 spatial patterns of soil moisture, runoff or evaporation. Figure 2 is a revealing illustration of how a  
183 "physically based" model that relies on detailed soil information can make inconsistent predictions  
184 under extreme circumstances (Beekman et al., 2014). On average, these models may function  
185 adequately, as almost all hydrological models do under average conditions, but the example of  
186 Figure 2 shows how during a relatively extreme drought in The Netherlands the modelled  
187 evaporation is unrealistic.

188 The top panels in Figure 2 show remote sensing derived evaporation obtained by interpolation of  
189 eddy-covariance and lysimeter observations using ETLook, an energy balance-based evaporation  
190 product (Bastiaanssen et al., 2012). The bottom panels in Figure 2 show evaporation modelled  
191 with the Netherlands Hydrological Instrument (NHI) distributed model, which heavily relies on  
192 detailed soil data. The two methods for estimating evaporation are independent, and arguably, the  
193 ETLook approach is more realistic, as it is based on eddy-covariance observations. The  
194 comparison is presented for two dry summer months in 2006: June (left panel) and July (right  
195 panel).

196 Two aspects of this comparison are striking. First, in terms of temporal dynamics, ETLook  
197 evaporation estimates show an increase in response to increased evaporative demand, whereas the  
198 NHI evaporation estimates are decreasing, in response to water stress. Second, in terms of spatial  
199 patterns, ETLook estimates are more uniform in response to relatively uniform climatic  
200 conditions, whereas NHI estimates are highly variable, mimicking the variability of the soil maps  
201 used in the model, which are used to determine plant available storage. The July 2006 picture in  
202 the bottom panel, in fact, mimics the soil map. Red (high evaporation) is seen on clay soils and  
203 purple (almost no evaporation) on sand.

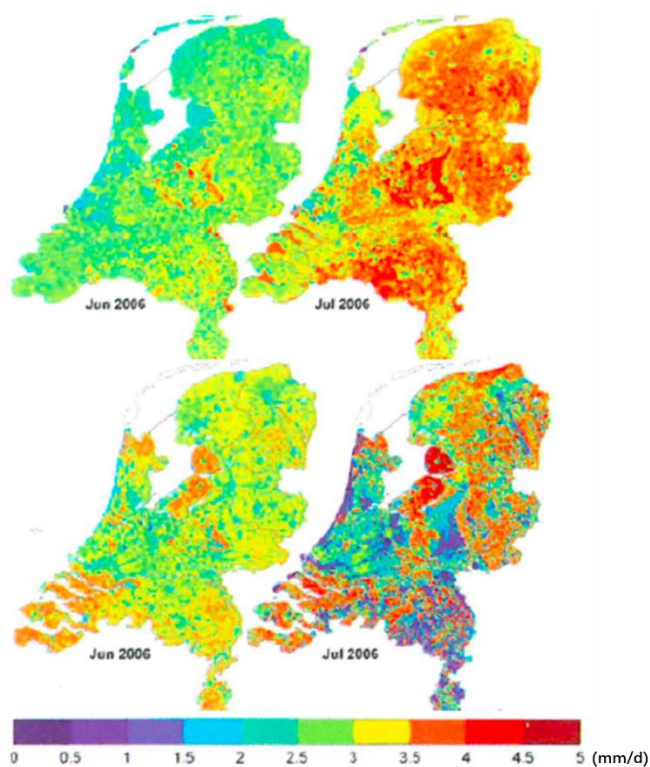
204 It is interesting to observe that according to ETLook (top right) the forested sandy part at the  
205 centre of The Netherlands was evaporating lushly, whereas according to the hydrological model



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206 (bottom right), this ecosystem appeared to be dead. Apparently, the ecosystems continued to  
207 evaporate well during July 2006, in spite of the dry weather conditions. Our interpretation is that  
208 the ecosystems had prepared for this eventuality and had created enough rootzone buffer to  
209 overcome this period of drought, compensating for the variability of soils.

210 Although such mismatch between distributed model outputs and remote sensing monitored  
211 patterns are not infrequent, they are typically not regarded as a challenge to the basic model  
212 assumptions, but rather, as a problem associated to the uncertainty in model inputs. Hence, such  
213 soil-centred hydrological models remain vivid under the hope that “novel, highly resolved soil  
214 information at higher resolutions than the grid scale of LSMs may help in better quantifying sub-  
215 grid variability of key infiltration parameters” (Vereecken et al., 2022). But is this a realizable  
216 hope?



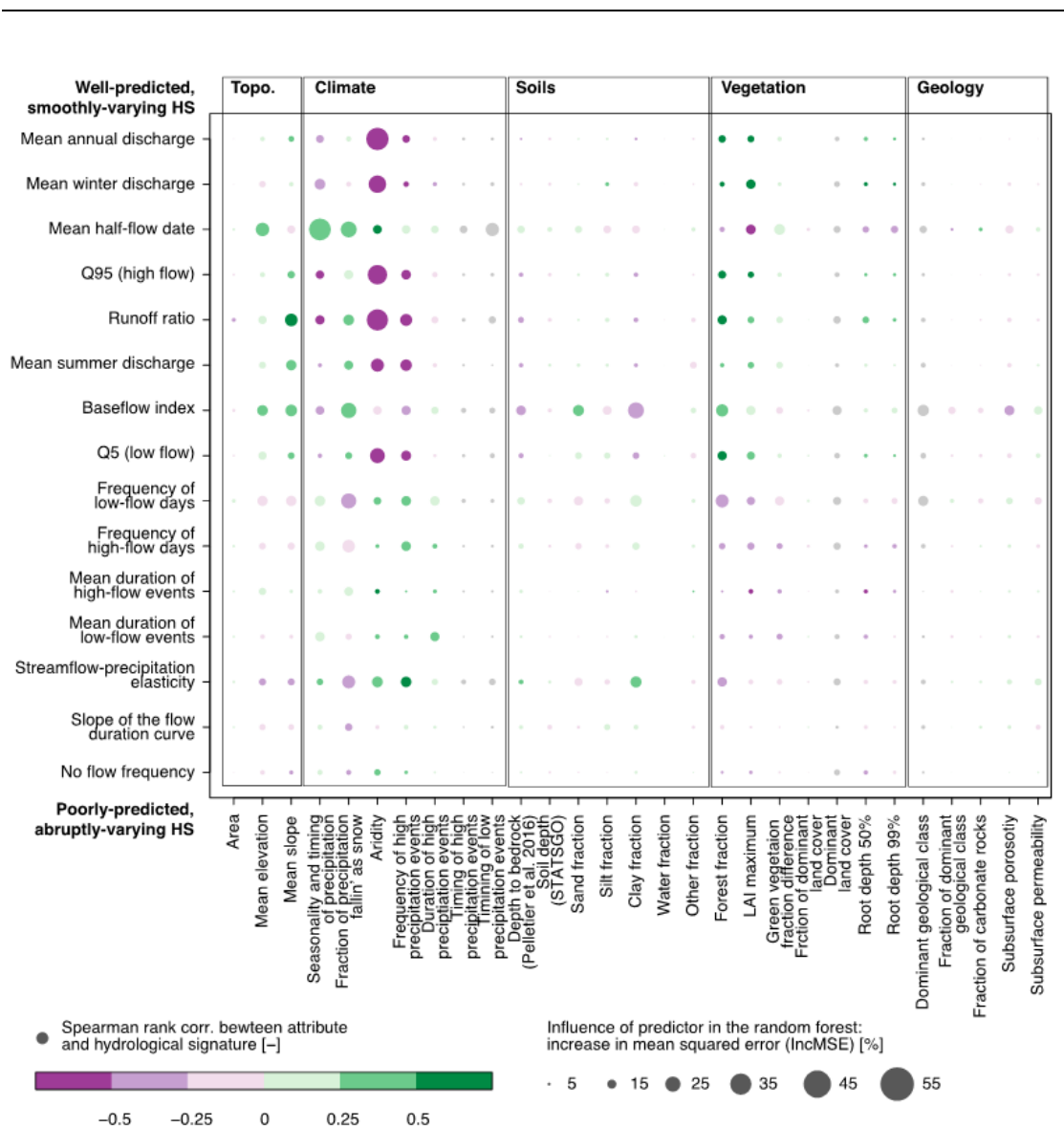
217  
218 Figure 2. Evaporation during June (left) and July (right) of 2006 in The Netherlands. Remote  
219 sensing derived above, modelled with a “physically-based” hydrological model below (from  
220 Beekman et al., 2014)

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221 **3.2 Is soil a good predictor for streamflow spatial variability?**

222 The top-down approach is a common way to infer internal catchment behaviour and its controlling  
223 factors from catchment response data (Sivapalan et al., 2003). For example, this approach has  
224 often been used to interpret spatial variability of streamflow based on controlling factors such as  
225 climate, vegetation, topography, geology and soils. Interestingly, in these applications, soil  
226 properties are often a poor predictor of streamflow variability. For example, Addor et al. (2018)  
227 used 671 catchments in the USA and found that, compared to soil properties, landscape features,  
228 i.e. vegetation and topography, have stronger correlations with hydrologic signatures, not only for  
229 average streamflow, but also for high-flow, low-flow, and streamflow seasonality (Figure 3).

230 One the arguments in favour of high resolution distributed models has been their ability of spatial  
231 extrapolation, such as capturing the spatial variability of streamflow. Such extrapolation ability  
232 cannot be achieved by lumped models that rely on calibration on each individual catchment.  
233 However, there are now several examples of catchment scale distributed models that describe the  
234 spatial variability of streamflow without relying on soil information (e.g. De Boer-Euser et al.,  
235 2016; Fenicia et al., 2016; Gao et al., 2019; Dal Molin et al., 2020; Fenicia et al., 2022). These  
236 models are clearly more complex than lumped models, but not orders of magnitude more complex,  
237 as they distribute parameters according to a small number of landscape units.



238

239 Figure 3. Comparison of the influence of catchments attributes and hydrological signatures for 671  
 240 U.S. watersheds (from Addor et al., 2018). Large, brightly coloured circles imply strong  
 241 correlations and high influence. (Reprinted by permission of John Wiley and Sons).

## 242 4 Putting the terrestrial ecosystem at the centre of hydrology

### 243 4.1 Ecosystem hierarchy

244 It has been shown that terrestrial ecosystems largely respond to external climate forcing, and to the  
 245 lower boundary conditions determined by topography and lithology (Chapin et al, 2011). With  
 246 time, terrestrial ecosystems organize themselves to make best use of the available solar energy and  
 247 resources. Hence, they adapt to the climate, by developing vegetation types in response to rainfall

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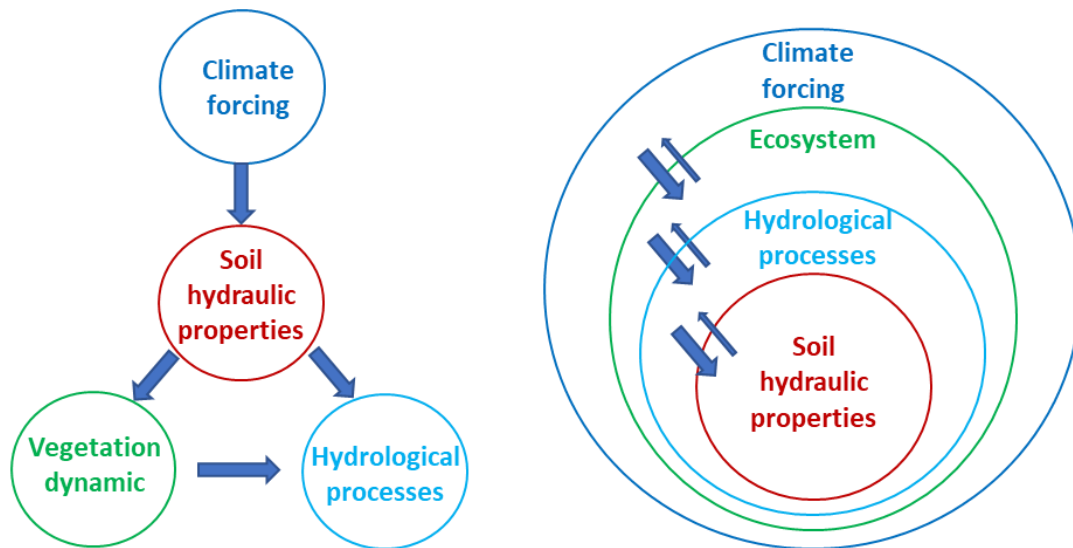
248 patterns and seasonal temperatures. They also develop the soil given the climate, organisms,  
249 topography and parental material as suggested by the *clorpt* model (see Section 1).

250 Our view, consistent with this perspective, is that an ecosystem adjusts the soil hydraulic  
251 properties to fulfil specific water management criteria. Hence, understanding the water  
252 management strategies of the ecosystem is a prerequisite to understanding and modelling soil  
253 processes. This perspective is opposite to the classical soil-centred hydrological perspective  
254 presented in Section 2, which sees water fluxes, such as evaporation and drainage, as a function of  
255 soil properties.

256 Figure 4 further illustrates our ecosystem view and how it differs from the classical approach in  
257 hydrological science. The traditional view is represented by the four isolated circles in the left  
258 panel of Figure 4. This view assumes that soil plays a central role in governing the terrestrial water  
259 cycle. In particular, depending on climate forcing, soil hydraulic properties will determine water  
260 availability for vegetation and water fluxes such as percolation and surface runoff. According to  
261 this view, the understanding of soil water processes is a prerequisite to simulate vegetation  
262 dynamics and water fluxes. The circles are isolated from each other, reflecting that in this view  
263 soil properties, vegetation cover and climate are seen as independent on each other, and can  
264 influence independently hydrological processes. Indeed, hydrological models typically  
265 parameterize soil and vegetation independently from each other and from climate forcing.

266 Our view is represented by the nested circles in the right panel of Figure 4. Climate sets the  
267 boundaries for the terrestrial ecosystem, and in turn, the ecosystem manages its water resources,  
268 determining hydrological processes. Soil hydraulic properties are a function of the ecosystem  
269 water management strategies. The circles are nested to reflect the hierarchy between them, in the  
270 sense that internal circles are dependent on the external ones. The double arrows indicate that  
271 there are feedbacks between these circles, but the influence of the external circles on the internal  
272 ones is much greater than vice versa. More specifically, local climate has a strong effect on an  
273 individual ecosystem, which intentionally adapts to it, developing strategies to grow, survive and  
274 reproduce. In turn, an individual ecosystem cannot change the local climate significantly  
275 according to its needs. Hence, the feedback of an ecosystem on the climate is smaller and less

276 “intentional” than the effect that the climate exerts on an ecosystem. Similarly, the control that the  
 277 terrestrial ecosystem exerts on soil hydraulic properties, mediated by its water management  
 278 strategies, is much greater and purposeful than the control of the soil on the embedding ecosystem.  
 279 In our perspective, such hierarchy and interactions can reduce rather than add complexity and  
 280 facilitate hydrological process understanding and modelling. For example, it provides a  
 281 justification for the level of detail of catchment models. In many applications of catchment  
 282 hydrology, the ‘ecosystem circle’ represents the necessary level of detail, and as the effect of soil  
 283 on the ecosystem is rather minor, it is unnecessary to dig into what happens within the ‘soil water  
 284 circle’.



285  
 286 Figure 4. The isolated circles (left) represent the traditional soil-centred hydrological perspective.  
 287 The nested circles (right) represent our view of ecosystem hierarchy and cause-effect  
 288 relationships.

## 289 **4.2 The ecosystem is the ultimate water manager**

290 An ecosystem that results from a process of evolution contains traits that are functional to its  
 291 survival. In this perspective, it is important to understand what the system is trying to achieve in  
 292 order to explain and predict its behaviour. In the context of hydrology, this approach requires to  
 293 understand (i) which water management strategies the ecosystem needs to adopt in order to sustain  
 294 itself and survive, (ii) how hydrological processes, such as interception, surface runoff (or lack  
 295 thereof), subsurface stormflow, contribute to satisfy the water needs of the ecosystem, and (iii)

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296 which physical characteristics the system needs to demonstrate to enable such processes. This  
297 evolutionary perspective considers the structure and internal processes of the ecosystem dependent  
298 on its overall behaviour, and it is contrary to the static approach which underlies typical  
299 description of soil hydrology, where the system structure is seen as prescribed, and small-scale  
300 processes are assumed to determine overall system behaviour.

301 So, what are the water strategies of the ecosystem, and how do they affect its structure and internal  
302 hydrological processes? Humans are well aware that water management is critical to their  
303 survival. For this reason, they have developed activities for optimum use of water resources such  
304 as flood control, water storage, water conservation, river regulation, irrigation, and water  
305 treatment. Similarly, a natural ecosystem can only survive if it organizes its water resilience. In  
306 other words, if an ecosystem had not organized its water resilience, it would not have survived and  
307 would no longer exist. The very existence of an ecosystem tells us several aspects of its water  
308 management strategies. In particular:

- 309 • An ecosystem needs to provide sufficient moisture storage in the rootzone, so that  
310 vegetation can overcome critical dry spells, but also sufficient infiltration capacity and  
311 subsurface drainage to maintain moisture levels between acceptable boundaries: not too  
312 wet and not too dry.
- 313 • Runoff ,the excess water after precipitation has replenished the ecosystem's water deficit  
314 needs to be drained as quickly and efficiently as possible.
- 315 • Preventing surface runoff is an essential need of an ecosystem, which serves to avoid soil  
316 erosion. Indeed, surface runoff is seldom observed on vegetated hillslopes. It does occur  
317 on bare rocks, where there is no vegetation, or on floodplains, where saturation overland  
318 flow does not cause significant erosion. Also, it occurs in disturbed ecosystems, such as  
319 urbanised areas, roads, paths and ploughed agricultural fields. In rare cases, such as the  
320 Loess Plateau in China, the failure of surface runoff prevention caused severe soil erosion  
321 at local scale and disastrous sediment deposition and flooding in the lower Yellow River.
- 322 • The ecosystem needs to retain nutrients and soil particles and to retain water for plants.

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323 For this reason, it creates preferential flow paths that facilitate infiltration while retaining  
324 moisture and nutrients in retention zones. If there is too much water, then excess water  
325 bypasses the rootzone where it can recharge the groundwater or is evacuated through  
326 preferential sub-surface drainage patterns on hillslopes. This type of drainage generates  
327 subsurface storm flow and recharges the groundwater system.

- 328 • Ecosystems will generally avoid catastrophic events such as death from drought,  
329 temperature stress, landslides, windthrows or fires. If such disruptive events occur, it is  
330 generally at time scales longer than ecosystem memory. If disturbances occur more  
331 frequently, ecosystem generally develop resilience to them, such as in the case of frequent  
332 fire, where ecosystems can develop fire resistant species, or vegetation that can recover  
333 biomass more quickly (Chapin et al., 2011).

334 Considering hydrological processes in the context of their purpose from an ecosystem perspective  
335 can clarify cause-effect relationships and therefore help their conceptualization and modelling. For  
336 example, it can constrain plausible values of SHPs, which can be determined based on  
337 considerations about overall system behaviour.

### 338 **4.3 The rootzone is the key element in hydrology**

339 From a catchment hydrology perspective, a key objective is to determine the partitioning of  
340 precipitation between evaporation, drainage and storage. This partitioning mostly takes place in  
341 the rootzone. The vertical profile of the critical zone can be divided into different layers, i.e.:  
342 canopy, litter layer, rootzone, water transition zone, unconfined groundwater, and confined  
343 groundwater. The most significant phase change of water happens in the canopy, litter layer, and  
344 rootzone. Once water overtakes these zones, evaporation is relatively small and water is routed to  
345 the stream through various pathways. Globally, the vegetation interception storage capacity of  
346 terrestrial ecosystems is about 1-2 mm, as estimated by remote sensing-based LAI data (De Roo et  
347 al., 1996). The litter layer storage capacity differs among ecosystems, but it is likely to increase  
348 the total interception storage capacity to around 2-5 mm (Shi et al., 2004; Gerrits et al., 2010).  
349 Global average rootzone storage capacity in vegetated regions is about 146-242 mm, as estimated  
350 by multiple approaches and datasets (Kleidon, 2004; Wang-Erlandsson et al., 2016), which is

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351 significantly larger than interception and litter layer water storage capacities. Therefore, the  
352 rootzone storage is the one with the longest memory, which influence how much precipitation  
353 eventually becomes streamflow.

354 Referring to common hydrological models, the rootzone storage can be assimilated to the  
355 “production” reservoir in the GR4J model (Perrin et al., 2003), the “upper zone” reservoir in HBV  
356 (Lindström et al., 1997), the “tension water” storage in Xinanjiang model (Zhao, 1992), or the  
357 “soil moisture” storage in probability distributed model (PDM) (Moore, 2007). In these models,  
358 the size of this reservoir is typically obtained by calibration. This approach is clearly  
359 unsatisfactory from a theoretical point of view as it makes these models not predictive under  
360 environmental change.

361 From a soil-based perspective, the rootzone storage is commonly estimated as a function of plant  
362 available moisture and rooting depth (Yang et al., 2016). In our view, this approach is also not  
363 satisfactory, as it considers plant available moisture and rooting depth as independent variables,  
364 and rootzone storage as the dependent variable. We argue the reverse: plant available moisture and  
365 rooting depth are a function of the rootzone storage that is created by the ecosystem to fulfil its  
366 water management strategies. Moreover, the classical approach is impractical, as obtaining the  
367 detailed spatio-temporal root and soil information at a global scale is virtually impossible (Or,  
368 2020).

369 So, how to determine rootzone storage without resorting to calibration, or in situ measurements?  
370 As mentioned in the previous section, our ecosystem approach would start with understanding the  
371 ecosystem water management strategies, and using this understanding to figure out how the  
372 ecosystem needs to organize its internal behaviour. Vegetation will try to maintain evaporation  
373 close to potential to maximise net carbon profit. It will therefore optimize its rootzone water  
374 storage so that it is sufficiently large to overcome typical dry spells, much like humans size dams  
375 to sustain droughts (Gao et al., 2014). An approach that appeared to work well locally and globally  
376 for estimating the rootzone storage capacity is the mass curve technique, originally developed for  
377 reservoir design at an acceptable probability of failure (Gao et al., 2014; Wang-Erlandsson et al.,  
378 2016). Here the supply is represented by precipitation, and the demand by potential evaporation.



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379 This technique is uniquely based on climate data. This technique has an important benefit over  
380 approaches based on calibration or field observations: it can also be used to describe how the  
381 rootzone would evolve in response to climate change.

382 It is worthwhile noting that rootzone and soil have a strong connection but are essentially different  
383 things. The soil profile can reach over hundreds of meters depth, e.g. the Loess Plateau in China  
384 (Zhang et al., 2014), of which only the rootzone is the active area, whereby the soil is merely the  
385 substrate of it. Rootzone storage can also be larger than soil water storage, for example in karst  
386 mountainous areas where soil is thin and discontinuous, bedrock storage serves as an important  
387 source of plant-available water (McCormick et al., 2021). In very dry climates, roots can even  
388 reach the deep groundwater, thus in this case, the rootzone also includes some part of the  
389 groundwater (see Singh et al., 2020). In cold regions, it is necessary to take account of snowmelt  
390 and soil freeze/thaw processes on rootzone water storage and resulting hydrologic connectivity  
391 (Gao et al., 2020; 2022). In cropland, where irrigation provides an extra water supply to rootzone  
392 during dry seasons, the rootzone water storage capacity is often smaller than under natural  
393 conditions with similar climate conditions (Xi et al., 2021).

#### 394 **4.4 Landscape-based model: the giant view of hydrology**

395 A soil-based model of catchment scale processes is like the ant's perspective, observing a complex  
396 world of heterogeneities and randomness (Savenije, 2010). According to this perspective, small-  
397 scale processes are the basis for integrated system behaviour. As a result, a model can be  
398 "physically-based" only if it relies on small scale physics.

399 Seeing the patterns of hillslope, landscape and catchment is rather the giant's perspective, as these  
400 patterns only become visible when we zoom out well beyond the microscale of the soil or the  
401 human scale (Savenije, 2010; Gao et al., 2018). Landscape, as the integration of topography and  
402 landcover, is seen as the long-term co-evolution of ecosystem, atmosphere, lithosphere,  
403 pedosphere, hydrosphere, and human activities (Wu, 2013; Troch et al., 2015). According to this  
404 perspective, a "physically-based" model needs to be based on large-scale system behaviour.

405 Both approaches can produce models that provide good results. However, in our perspective, for

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406 catchment hydrology applications it is the giant's perspective that wins. First, the giant's model  
407 captures the right cause-effect relationships, and is therefore more satisfactory from a theoretical  
408 point of view. For example, it is a tool to test how an ecosystem would adapt to changes in  
409 climatic drivers. Second, landscape-based catchment models will generally be simpler than  
410 fragmented catchment models, as natural system exhibits emergent properties, which effectively  
411 enable a description of large-scale processes independent on what happens at the smaller scale.  
412 Such emergent properties are often characterized by simple laws, such as the fill and spill bucket  
413 model with thresholds and associated time scales (McDonnell et al., 2021), and the linear reservoir  
414 for groundwater at hillslope and catchment scale (Savenije, 2010; Fenicia et al., 2011; Savenije  
415 and Hrachowitz, 2017). Interestingly, the groundwater system also appears to be self-organized  
416 and structured (Savenije, 2018). For example, the recession parameter  $k$  is around 45 days in  
417 worldwide catchments regardless of their climate, topography, soil, and geology (Brutsaert, 2008).  
418 Discovering these properties and related signatures benefit our understanding and prediction of the  
419 dynamic adaption of ecosystems to environmental change, and the subsequent impacts on  
420 hydrology (Gharari et al., 2014; Jackisch et al., 2021).

421 This ecosystem perspective provides a physical justification for catchment scale models that do  
422 not rely on small scale physics, as they are independent on what happens at the smaller scale.  
423 Moreover, they can provide a constraint to smaller scale processes, and therefore facilitate their  
424 representation. For example, the partitioning of water between evaporation, storage and release  
425 that characterize the larger scale system can be used to constrain plausible values of difficult to  
426 measure soil properties such as rooting depth, plant available moisture and hydraulic conductivity.  
427 This can favour more accurate descriptions of soil water dynamics, which, although often  
428 unnecessary for typical catchment scale applications, can may be important for other purposes.

#### 429 **4.5 Proposed modelling steps in poorly gauged catchments**

430 How can this approach be implemented in modelling an ungauged catchment? There are the  
431 following steps to be considered as a quick guide to model building.

432 The first thing is to classify the basin on landscape and geology. This determines model structure.  
433 It defines the proportion between the major three fast runoff mechanisms: rapid subsurface flow

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434 (for Hillslope), saturation overland flow (for Wetland) and Hortonian overland flow (for Plateau,  
435 bare rock). Theory and application of landscape-based modelling are presented in (Savenije, 2010;  
436 Gharari et al., 2011, Fenicia et al., 2011; Gao et al., 2014, 2018; De Boer-Euser et al., 2016;  
437 Hulsman et al., 2021a; Bouaziz et al.,2022).

438 Subsequently classify landscape units on ecosystem, land use and climate. The climate and the  
439 ecosystem determine hydrological parameters such as rootzone storage, interception capacity,  
440 infiltration capacity and subsurface drainage. Spatial variability of rootzone storage determines the  
441 Beta function of the non-linear rootzone reservoir (Gao et al., 2018). This results in hydrological  
442 response units based on landscape and geology (defining model structure), ecosystem and climate  
443 (defining parameter values), which can be grouped per sub-basin.

444 Recession time scales can be derived from limited observations, if available, or otherwise  
445 estimated; they do not affect the overall water balance. The longer time scales of groundwater  
446 recession may be derived from Gravity Recovery and Climate Experiment (GRACE) data, which  
447 can also be used to constrain groundwater dynamics (Winsemius et al., 2006; Hulsman et al.,  
448 2021b).

449 Minor calibration parameters remain, such as the splitter between fast subsurface runoff and  
450 recharge. These have a limited effect on the water balance and can be estimated if no observations  
451 are available.

## 452 **5 Why is the soil-based modelling tradition so rooted in** 453 **hydrology?**

### 454 **5.1 Agricultural bias**

455 Since hydrology was born from chapters of agricultural and hydraulics textbooks (Rodríguez-  
456 Iturbe and Rinaldo, 2004), the “agricultural bias” has probably played a major role in  
457 overemphasizing the importance of soils. In agriculture, the focus is on seasonal crops. A seasonal  
458 crop has limited time to develop a rootzone storage that can buffer for longer term variability. At  
459 best, it can buffer for average dry spells that may occur within an average year. This is why  
460 modern agriculture requires water management by the farmer to buffer for natural fluctuations. In

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461 agriculture, ploughing destroys preferential infiltration and sub-surface drainage. It also limits the  
462 rootzone storage capacity to the relatively small soil layer above the plough pan. In such cases it is  
463 indeed the moisture holding capacity of the soil that determines the rootzone storage capacity.

464 The widely used Penman-Monteith equation for estimating reference evaporation works well in  
465 agriculture, where the dominant evaporation is from crops. However, it is likely not appropriate to  
466 describe land-atmosphere interaction of natural ecosystems. Unfortunately, this “agricultural bias”  
467 only applicable in small proportion of terrestrial area has been dominant in most hydrological  
468 work. We argue that this deeply rooted soil-based perception may limit or even mislead the further  
469 development of hydrological science, especially for next generation professionals.

470 Even in the Anthropocene, where human impacts on essential planetary processes have become  
471 profound, and hydrological processes are affected by human activities such as agriculture,  
472 urbanization and deforestation, we believe it is still essential to emphasize the importance of  
473 ecosystem understanding. There are two reasons: 1) the majority of our earth, and particularly the  
474 uphill runoff generating parts of catchments, is still dominated by natural ecosystems, although  
475 human modification has modified 14.5% or 18.5 M km<sup>2</sup> of land (Theobald et al., 2020), and 2)  
476 also for human modified systems the ecological approach applies, provided that the ecosystem is  
477 given sufficient time to become self-sufficient and manage its own resources.

## 478 **5.2 Unreliable intuition**

479 Hydrologists intuitively see the soil as the critical agent. It may very well be the perspective of the  
480 agent that causes it. As people, we are biased by our perspective and the scale at which we observe  
481 processes. We are therefore just too small to perceive the larger scale processes that dominate  
482 landscape hydrology. We tend to dig holes in the Earth and try to infer larger scale behaviour from  
483 what we observe inside this hole. The human scale prevents us from seeing the larger picture. We  
484 need the giant's perspective to recognise the patterns present in the landscape.

485 At the human scale, assuming that soil properties, such as texture and porosity, matter makes  
486 intuitive sense. People tend to describe what they see, and if they see water flowing or  
487 disappearing in the ground they think that it is because of such soil properties. The role of the

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488 ecosystem as the driver of the system is much more difficult to recognize, especially within its  
489 evolutionary history. It requires seeing the environment as a living organism, which continuously  
490 evolves and adjusts to changing circumstances. It also implies that the hydrological properties are  
491 not constant over time. The rootzone storage, the most critical control on rainfall-runoff processes,  
492 is continuously changing in response to changing climatic and human drivers (Nijzink et al., 2016;  
493 Bouaziz et al., 2022). Instead of describing the 'now' as an invariant and static condition, with  
494 environmental properties as a given, we have to think of the history that determined these  
495 environmental conditions, which is much more difficult to realise.

## 496 **6 Limitations**

497 We stress that our ecosystem approach is subject to certain limitations. First, it applies at the so-  
498 called ecosystem scale. This spatial scale can vary depending on the environment. It can be a few  
499 square meters for grass, in the order of hectares for forest, even larger for sparse vegetation.  
500 Catchment scales are usually larger than the ecosystem scale. Therefore, our approach is generally  
501 suited for scales that are typical in hydrological modelling application. Second, we are talking  
502 about ecosystems that have reached a certain level of equilibrium and are self-sustained. We are  
503 not limiting ourselves to natural ecosystem. They can also be artificially induced. But they do not  
504 need to rely on artificial help for their survival, such as irrigation or fertilization. Third, our  
505 arguments are mostly related to water fluxes, and they do not pertain to water chemistry. The  
506 variability of soils can have a pronounced influence on predicting water quality, solute transport,  
507 and transit times (Weiler et al., 2017; Sternagel et al., 2021).

## 508 **7 Conclusions**

509 Traditional hydrological theories place soil physical properties at the heart of hydrology,  
510 considering them as the driver of water fluxes, which is misleading for both process understanding  
511 and model development. In contrast we need an ecosystem-based approach, where the structure of  
512 the terrestrial ecosystem and its internal processes are seen as a consequence of ecosystem water  
513 management strategies needed for its survival and growth. Hence, the ecosystem is the ultimate  
514 manager of the soil. We advocate a change in perspective that places the ecosystem and landscape

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515 at the heart of terrestrial hydrology and develop holistic and alive ecosystem-based hydrological  
516 models with a more realistic representation of hydrological processes.

517

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