## Supplement of

# Influences of sources and weather dynamics on atmospheric deposition of Se species and other trace elements 

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## S1: Additional information on ICP-MS/MS analysis

Table S1. ICP-MS/MS acquisition parameters used for each analysed elements in aerosol digests and water extracts, cloud water samples and precipitation samples.

| Element | Collision/reaction cell gas and flow | MS Mode | $m / z$ | Acquisition time (ms) |
| :---: | :---: | :---: | :---: | :---: |
| Li | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}^{-1}$ | MS/MS | 7->7 | 50 |
| Na | $\mathrm{He} ; 5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | Single Quad | 23 | 10 |
| Mg | $\mathrm{He} ; 5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | Single Quad | 24 | 10 |
| Al | $\mathrm{He} ; 5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | Single Quad | 27 | 10 |
| Si | He ; $5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | MS/MS | $28->28$ | 100 |
| K | $\mathrm{He} ; 5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | Single Quad | 39 | 10 |
| P | $\mathrm{O}_{2} ; 30 \%$ | MS/MS | $31->47$ | 100 |
| S | $\mathrm{O}_{2} ; 30 \%$ | MS/MS | $34->50$ | 50 |
| Ti | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}^{-1}$ | MS/MS | $47->47$ | 50 |
| V | $\mathrm{He} ; 5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | Single Quad | 51 | 100 |
| Cr | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}^{-1}$ | MS/MS | 52->52 | 100 |
| Mn | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | MS/MS | 55->55 | 50 |
| Fe | $\mathrm{He} ; 5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | Single Quad | 56 | 50 |
| Co | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | MS/MS | 59->59 | 100 |
| Ni | $\mathrm{He} ; 5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | Single Quad | 60 | 50 |
| Cu | $\mathrm{He} ; 5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | Single Quad | 63 | 100 |
| Zn | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | MS/MS | 66->66 | 100 |
| As | $\mathrm{O}_{2} ; 30 \%$ | MS/MS | 75->91 | 300 |
| Se | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}^{-1}$ | MS/MS | $\begin{aligned} & 78->78 \\ & 80->80 \end{aligned}$ | 300 |
| Br | $\mathrm{O}_{2} ; 30 \%$ | MS/MS | $81->97$ | 100 |
| Rb | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | MS/MS | 85->85 | 50 |
| Sr | $\mathrm{He} ; 5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | Single Quad | 88 | 50 |
| Nb | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | MS/MS | 93->93 | 100 |
| Mo | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}^{-1}$ | MS/MS | 98->98 | 100 |
| Ag | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}^{-1}$ | MS/MS | 107->107 | 100 |
| Cd | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | MS/MS | 111->111 | 100 |
| I | $\mathrm{O}_{2} ; 30 \%$ | MS/MS | $127->127$ | 100 |
| Cs | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | MS/MS | 133->133 | 100 |
| Ba | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}^{-1}$ | MS/MS | 135->135 | 50 |
| Pb | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}^{-1}$ | MS/MS | 208->208 | 100 |

Table S2. Elements analysed as internal standards during total element analysis along with information on used ICP-MS/MS acquisition parameters.

| Element | Collision/reaction cell gas and flow | MS Mode | $m / z$ | Acquisition time (ms) |
| :---: | :---: | :---: | :---: | :---: |
| Sc | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}^{-1}$ | MS/MS | 45->45 | 50 |
| Sc | O2; 30\% | MS/MS | $45->61$ | 50 |
| Sc | He; $5 \mathrm{~mL} \mathrm{~min}^{-1}$ | Single Quad | 45 | 50 |
| Y | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}^{-1}$ | MS/MS | 89->89 | 50 |
| Y | O2; 30\% | MS/MS | 89->105 | 50 |
| Y | He; $5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | Single Quad | 89 | 50 |
| In | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}^{-1}$ | MS/MS | $\begin{aligned} & 115->115 \\ & 118->118 \end{aligned}$ | 50 |
| In | O2; 30\% | MS/MS | 115->131 | 50 |
| In | He; $5 \mathrm{~mL} \mathrm{~min}^{-1}$ | Single Quad | $\begin{aligned} & 115 \\ & 118 \end{aligned}$ | 50 |
| Lu | $\mathrm{H}_{2} ; 5 \mathrm{~mL} \mathrm{~min}^{-1}$ | MS/MS | 175->175 | 50 |
| Lu | O2; 30\% | MS/MS | 175->191 | 50 |
| Lu | $\mathrm{He} ; 5 \mathrm{~mL} \mathrm{~min}{ }^{-1}$ | Single Quad | 175 | 50 |

Table S3. Recoveries (\%) and error (\%) obtained for measured elements in two certified reference materials (CRMs) which were analysed with the atmospheric samples. The CRMs were analysed after their dilution ( d 10 or d100) in the corresponding sample matrix, i.e., $1 \% \mathrm{HNO}_{3}$ for precipitation, cloud water and aerosol water extracts, and $16-26 \% \mathrm{HNO}_{3}$ for aerosol digests. Analysed replicates for each CRM are indicated by N.

|  |  | Precipitation \& Cloud water $\left(1 \% \mathrm{HNO}_{3}\right)$ |  |  |  |  | Aerosol Extract ( $16-26 \% \mathrm{HNO}_{3}$ ) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Certified element concentration ( $\mu \mathrm{g} \mathrm{L}^{-1}$ ) |  | $v^{2}{ }^{\text {a }}$ <br> (\%) <br> $s d^{d}$ | Error ${ }^{\text {b }}$ (\%) | $\begin{gathered} \text { Used } \\ \text { Dilution } \end{gathered}$ | N | Recove $\mathrm{av}^{\mathrm{c}}$ |  |  | $\begin{gathered} \text { Errorb } \\ (\%) \end{gathered}$ | $\begin{gathered} \text { Used } \\ \text { Dilution } \\ \text { d } \end{gathered}$ | N |
| CRM NIST 1643f, Trace Elements in Water |  |  |  |  |  |  |  |  |  |  |  |  |
| Li | $16.6 \pm 0.4$ | 103 | $\pm 9$ | 3 | d10, d100 | 4 | 99.5 | $\pm$ | 4 | -0.5 | d10, d100 | 6 |
| Na | $18830 \pm 250$ | 102 | $\pm 2$ | 2 | d10, d100 | 4 | 98 | $\pm$ | 2 | -2 | d10, d100 | 12 |
| Mg | $7454 \pm 60$ | 103 | $\pm 2$ | 3 | d10, d100 | 4 | 102.7 | $\pm$ | 1.1 | 2.7 | d10, d100 | 12 |
| Al | $133.8 \pm 1.2$ |  | $\pm$ |  |  | 4 | 105 | $\pm$ | 3 | 5 | d10, d100 | 12 |
| K | $1933 \pm 9$ | 100.4 | $\pm 6.7$ | 0.4 | d10 | 4 | 98.5 | $\pm$ | 0.7 | -1.5 | d10, d100 | 12 |
| V | $36.1 \pm 0.3$ | 105 | $\pm 0$ | 5 | d10 | 4 | 103 | $\pm$ | 2 | 3 | d10, d100 | 6 |
| Cr | $18.5 \pm 0.1$ | 101 | $\pm 8$ | 1 | d10, d100 | 4 | 105 | $\pm$ | 3 | 5 | d10, d100 | 6 |
| Mn | $37.1 \pm 0.6$ | 103 | $\pm 4$ | 3 | d10, d100 | 4 | 98.7 | $\pm$ | 2.3 | -1.3 | d10, d100 | 6 |
| Fe | $93.4 \pm 0.8$ | 105 | $\pm 3$ | 5 | d10 | 4 | 101.8 | $\pm$ | 0.5 | 1.8 | d10, d100 | 12 |
| Co | $25.3 \pm 0.2$ | 98 | $\pm 4$ | -2 | d10, d100 | 4 | 103.0 | $\pm$ | 0.1 | 3.0 | d10, d100 | 6 |
| Ni | $59.8 \pm 1.4$ | 99.6 | $\pm 3.0$ | -0.4 | d10, d100 | 4 | 101.9 | $\pm$ | 0.4 | 1.9 | d10, d100 | 6 |
| Cu | $21.7 \pm 0.7$ | 102 | $\pm 7$ | 2 | d10, d100 | 4 | 102.3 | $\pm$ | 1.6 | 2.3 | d10, d100 | 6 |
| Zn | $74.4 \pm 1.7$ | 102 | $\pm 4$ | 2 | d10 | 4 | 105 | $\pm$ | 3 | 5 | d10 | 6 |
| As | $57.4 \pm 0.4$ | 99 | $\pm 0.7$ | -1 | d10, d100 | 4 | 103.7 | $\pm$ | 0.7 | 3.7 | d10, d100 | 6 |
| ${ }^{78} \mathrm{Se}$ | $11.7 \pm 0.1$ | 99 | $\pm 1$ | -1 | d10, d100 | 4 | 102.9 | $\pm$ | 0.5 | 2.9 | d10, d100 | 6 |
| ${ }^{80} \mathrm{Se}$ | $11.7 \pm 0.1$ | 99.2 | $\pm 0.4$ | -0.8 | d10, d100 | 4 | 101.2 | $\pm$ | 0.1 | 1.2 | d10, d100 | 6 |
| Rb | $12.6 \pm 0.1$ |  | $\pm 7$ | -3 | d10 | 4 | 105.5 | $\pm$ | 1.0 | 5.5 | d10, d100 | 6 |
| Sr | $314 \pm 19$ | 101 | $\pm 5$ | 1 | d10, d100 | 4 | 103.2 | $\pm$ | 1.4 | 3.2 | d10, d100 | 6 |
| Mo | $115.3 \pm 1.7$ | 99 | $\pm 5$ | -1 | d10, d100 | 4 | 103.4 | $\pm$ | 0.4 | 3.4 | d10, d100 | 6 |
| Ag | $1.0 \pm 0.0$ |  | $\pm 7$ | -3 | d10 | 4 | 107 | $\pm$ | 4 | 7 | d10 | 6 |
| Cd | $5.9 \pm 0.1$ | 102 | $\pm 4$ | 2 | d10, d100 | 4 | 102 | $\pm$ | 3 | 2 | d10, d100 | 6 |
| Ba | $518.2 \pm 7$ | 102 | $\pm 2$ | 2 | d10, d100 | 4 | 102.2 | $\pm$ | 0.2 | 2.2 | d10, d100 | 6 |
| Pb | $18.5 \pm 0.1$ | 99.7 | $\pm 1.7$ | -0.3 | d10 | 4 | 103.1 | $\pm$ | 1.5 | 3.1 | d10, d100 | 6 |
| CRM TMDA 51.2, Trace Elements in Surface Water (Lake Ontario) |  |  |  |  |  |  |  |  |  |  |  |  |
| Al | $96 \pm 19$ |  |  |  |  |  | 99.2 | $\pm$ | 7.8 | -0.8 | d10 | 6 |
| V | $48 \pm 8$ | 102 | $\pm \quad 2$ | 2 | d10, d100 | 4 | 102.7 | $\pm$ | 0.5 | 2.7 | d10, d100 | 6 |
| Cr | $63 \pm 7$ | 97 | $\pm 5$ | -3 | d10, d100 | 4 | 103.8 | $\pm$ | 1.3 | 3.8 | d10, d100 | 6 |
| Mn | $82 \pm 10$ | 98 | $\pm 2$ | -2 | d10, d100 | 4 | 102.0 | $\pm$ | 1.0 | 2.0 | d10, d100 | 6 |
| Fe | $111 \pm 26$ | 108 | $\pm 6$ | 8 | d10 | 4 | 103 | $\pm$ | 3 | 3 |  | 6 |
| Co | $72 \pm 6$ | 94 | $\pm 8$ | -6 | d10, d100 | 4 | 104.8 | $\pm$ | 1.0 | 4.8 | d10, d100 | 6 |
| Ni | $67 \pm 7$ | 97.7 | $\pm 0.9$ | -2.3 | d10, d100 | 4 | 104.9 | $\pm$ | 0.3 | 4.9 | d10, d100 | 6 |
| Cu | $91 \pm 10$ | 99 | $\pm 3$ | -1 | d10, d100 | 4 | 104.4 | $\pm$ | 1.2 | 4.4 | d10, d100 | 6 |
| Zn | $106 \pm 15$ | 102 | $\pm 3$ | 2 | d10 | 4 | 105.6 | $\pm$ | 0.4 | 5.6 | d10, d100 | 6 |
| As | $15 \pm 3$ | 96.0 | $\pm \quad 1.2$ | -4 | d10, d100 | 4 | 103.1 | $\pm$ | 0.9 | 3.1 | d10, d100 | 6 |
| ${ }^{78} \mathrm{Se}$ | $12 \pm 3$ | 96.7 | $\pm 0.3$ | -3.3 | d10, d100 | 4 | 97.6 | $\pm$ | 0.3 | -2.4 | d10, d100 | 6 |
| ${ }^{80} \mathrm{Se}$ | $12 \pm 3$ | 96.0 | $\pm \quad 1.2$ | -4 | d10, d100 | 4 | 98.4 | $\pm$ | 2.0 | -2 | d10, d100 | 6 |
| Sr | $121 \pm 12$ | 96 | $\pm 4$ | -4 | d10 | 4 | 101.7 | $\pm$ | 0.4 | 1.7 | d10, d100 | 6 |
| Mo | $59 \pm 6$ | 97 | $\pm 5$ | -3 | d10, d100 | 4 | 106.3 | $\pm$ | 1.0 | 6.3 | d10, d100 | 6 |
| Cd | $25 \pm 3$ | 99.2 | $\pm \quad 1.0$ | -0.8 | d10, d100 | 4 | 103 | $\pm$ | 2 | 3 | d10, d100 | 6 |
| Ba | $73 \pm 6$ | 99 | $\pm 3$ | -1 | d10, d100 | 4 | 104.3 | $\pm$ | 0.8 | 4.3 | d10, d100 | 6 |
| Pb | $73 \pm 11$ | 99 | $\pm \quad 4$ | -1 | d10, d100 | 4 | 102.0 | $\pm$ | 0.5 | 2.0 | d10, d100 | 6 |

${ }^{\text {a }}$ Recoveries were calculated as follows: (element concentration measured)/(element concentration certified) x 100; ${ }^{\text {b }}$ Errors were calculated as follows: [(element concentration measured)-(element concentration certified)] / (element concentration certified) x100; ${ }^{c}$ av.: average; ${ }^{d}$ sd: standard deviation

## S2: Extraction and pre-concentration of Se species in atmospheric samples

## S2.1 Extraction of Se species from aerosol filter samples

We developed a method for extraction of selenium (Se) species from aerosol samples. For a better comparison to sampled precipitation, our focus was on the water-soluble fraction of aerosols. The extraction of Se species from aerosol samples was done as follows: $11.404 \mathrm{~cm}^{2}$ of aerosol filter (corresponding to $7197 \pm 1816 \mathrm{~m}^{3}$ of sampled air) was placed in a polyporopylene tube with 15 mL of ultrapure water, the mixture was then sonicated twice for 20 min at $20^{\circ} \mathrm{C}$, and finally filtered after extraction $(0.22 \mu \mathrm{~m}$, Nylon). The sonication was done twice for 20 min instead of full 40 min to avoid increases of the sample temperature, which is known to potentially cause species transformation (Gómez-Ariza et al., 1999).

Stability of $\mathrm{Se}^{\mathrm{IV}}, \mathrm{Se}^{\mathrm{VI}}$, SeMet and $\mathrm{SeCys}_{2}$ was determined by measuring recoveries of initial species concentration of $200 \mathrm{ng} \cdot \mathrm{L}^{-1}$ after sonication by HPLC-ICP-MS/MS (specification on methods in S3). $\mathrm{Se}^{\mathrm{IV}}, \mathrm{Se}^{\mathrm{VI}}$ and $\mathrm{SeCys}_{2}$ showed high recoveries, i.e., $101 \pm 2 \%, 101 \pm 2 \%$ and $96 \pm 0.4 \%$, respectively. However, recoveries of SeMet were lower, i.e., of $65 \pm 3 \%$, indicating that sonication caused a partial transformation of SeMet species. Prior to lyophilisation of the aerosol extracts, ammonium citrate was added (to match mobile phase concentration of LC method after lyophilisation).

Overall extraction efficiencies of the water-soluble fraction of aerosols were determined by comparison between total Se concentrations in the water extract and the total acid digestion by ICP-MS/MS (method specification in S1 for total analysis and S3 for speciation analysis). Extraction efficiencies in the water extracts of the 2015-2020 aerosol time series $(n=70)$ were $85 \pm 18 \%$ for Se and $100 \pm 10 \%$ for sulfur (S) species.

## S2.2 Pre-concentration of Se species in atmospheric samples

We developed a method to pre-concentrate atmospheric samples based on lyophilisation using inorganic Se species (i.e., $\mathrm{Se}^{\mathrm{IV}}$ and $\mathrm{Se}^{\mathrm{VI}}$ ) that have been previously identified in rainwater (Suess et al., 2019; Roulier et al., 2021) as well as potentially present organic species for which standards are commercially available, i.e., $\mathrm{SeCys}_{2}$ and SeMet. Different parameters were tested including the effects of ionic strength, sampling containers, as well as lyophilisation to complete dryness versus lyophilisation to a residual volume. For each test, Se species recovery and stability were assessed.

The first experiments consisted of lyophilisation to complete dryness of:

- Solutions of individual Se species ( $200 \mathrm{ng} \cdot \mathrm{L}^{-1}$ ) in ultrapure water
- Solutions of individual Se species ( $200 \mathrm{ng} \cdot \mathrm{L}^{-1}$ ) in a $2 \mathrm{mmol} \cdot \mathrm{L}^{-1}$ ammonium citrate solution to match mobile phase concentration of LC method (described in S3) after lyophilisation
- Three different rainwater samples collected at Pic du Midi Observatory (previously stored at $4^{\circ} \mathrm{C}$ ) and spiked with individual Se species ( $50 \mathrm{ng} \cdot \mathrm{L}^{-1}$ ).

Recoveries were determined by comparison between the initial concentration of spiked Se species and the final recovered Se species (i.e., after lyophilisation) determined using HPLC-ICP-MS/MS. Se species recoveries are shown in Fig. S1 and demonstrate that lyophilisation to complete dryness leads to transformation and losses of Se species in all different tested matrices. The lowest recoveries were obtained for Se species in ultrapure water ( $\mathrm{Se}^{\mathrm{IV}}$ : $64 \pm 8 \%$, Se ${ }^{\text {VII }}: 78 \pm 0.3 \%$, SeMet: $12 \pm 0.2 \%$, SeCys $_{2}: 8 \pm 1 \%$ ) followed by Se species in rainwater ( $\mathrm{Se}^{\mathrm{IV}}: 85 \pm 16 \%$, Se ${ }^{\text {VI: }}: 106 \pm 12 \%$, SeMet: $58 \pm 10 \%$, $\mathrm{SeCys}_{2}: 22 \pm 8 \%$ ) and finally those for Se species in $2 \mathrm{mmol} \cdot \mathrm{L}^{-1}$ ammonium citrate (Se ${ }^{\mathrm{IV}}: 90 \pm 1 \%, \mathrm{Se}^{\mathrm{VI}}: 100 \pm 0.3 \%$, SeMet: $80 \pm 0.3 \%$, $\mathrm{SeCys}_{2}: 103 \pm 1 \%$ ). We particularly observed losses of the organic Se species SeMet and $\mathrm{SeCys}_{2}$ in ultrapure water and rainwater (Fig. S1). Lyophilisation of tested organic Se species in relatively low ionic strength solutions might lead to transformation to other organic Se species that are not retained by anion exchange. Overall, the results suggest a significant influence of ionic strength on Se species recoveries, with higher recoveries of Se species during lyophilisation at increased ionic strength.


Fig. S1. Recoveries of individual Se species during lyophilisation to complete dryness in different matrices. Tested matrices included ultrapure water, a $2 \mathrm{mmol} \cdot \mathrm{L}^{-1}$ ammonium citrate solution (eluent of LC-method) and three different rainwater samples collected at Pic du Midi Observatory. Recoveries of single Se species are displayed as bar plots (\% of initial Se species concentrations). The sum of concentrations of all recovered Se species are displayed as diamonds for each tested solution. The error bars represent the standard deviation values resulting from quantification by LC-ICP-MS/MS in duplicate.

The following tests aimed at investigating the effect of different types of containers used for the lyophilisation on the Se species recoveries. The tested containers for lyophilisation included:

- Greiner polypropylene centrifuge tubes "PP-Greiner" (Huber)
- metal free-centrifuge tubes (VWR)
- Perfluoralkoxy (PFA) tubes (AHF, Analysentechnik AG)

Compared to when using lyophilisation of sample to complete dryness, all Se species were entirely recovered with lyophilisation to a residual volume of $<1.5 \mathrm{~mL}$, and this was true for all tested containers (Fig. S2). The Greiner centrifugation tubes were thus chosen for collection of field samples and their processing, i.e., lyophilisation.


Fig. S2. Recoveries Se species during lyophilisation to a residual volume in different tested containers. Lyophilisation was done with addition of $2 \mathrm{mmol} \cdot \mathrm{L}^{-1}$ ammonium citrate solution (eluent of LC-method). Tested lyophilisation containers included polypropylene (PP) Greiner centrifuge tubes, metal free-centrifuge tubes and perfluoralkoxy (PFA) tubes. Recoveries of Se species are displayed as bar plots (\% of initial Se species concentrations). The error bars represent the standard deviation values resulting from three independent replicates and subsequent quantification by LC-ICP-MS/MS in duplicate.

In addition to above tested parameters, the ratio of initial to residual volume should be considered. Our optimized pre-concentration method involves the lyophilisation of frozen precipitation samples from an initial volume of

12 mL (for precipitation) or 9 mL (for water extract of aerosol filter) to a residual volume of 1.5 mL (preconcentration factor of 8 or 6 , respectively). These pre-concentration factors were found to be sufficient to determine Se speciation in atmospheric samples collected at remote sites with limited available samples volumes. Different ratios of initial to residual volume, especially much higher initial samples volumes compared to low residual volumes ( 1.5 mL ) will likely lead to poorer Se species recoveries and would need to be tested in future studies.

## S3: LC-ICP-MS/MS method optimization and detailed procedures

The optimization of the LC-ICP-MS/MS method to determine Se speciation in atmospheric samples was aimed at optimizing a method previously used for Se speciation in soil extracts and water samples (e.g., Tolu et al. (2011); Darrouzès et al. (2008)) in order to achieve:

1) lower solvent and sample consumption
2) lower running time (i.e. analysis time per sample)
3) lower detection limits for different Se species (i.e., $\mathrm{Se}^{\mathrm{IV}}, \mathrm{Se}^{\mathrm{VI}}, \mathrm{SeMet}$, and $\mathrm{SeCys}_{2}$ )

In the previous method by Tolu et al. (2011), an anion exchanges chromatography column (Hamilton PRP-X100, $250 \times 4.1 \mathrm{~mm}, 10 \mu \mathrm{~m}$ ) was used with an isocratic elution of ammonium citrate ( $5 \mathrm{mM}, 2 \% \mathrm{MeOH}, \mathrm{pH} 5.2$ ) delivered at $1 \mathrm{~mL} \cdot \mathrm{~min}^{-1}$, injection volumes of $100-400 \mu \mathrm{~L}$ and a total measurement time of 15 min .
Here, we used the same column type, however with smaller inner column diameter (i.e., 2.1 mm vs 4.1 mm ) to reduce solvent and sample consumption (reduction of mobile phase flow rate and sample injection volume), smaller column length (i.e., 150 vs 250 mm ) to reduce measurement time, as well as smaller particle size (i.e., 5 vs $10 \mu \mathrm{~m}$ ) to increase retention of Se species. Furthermore, we optimized the gradient elution (Table S4 and Fig. S3) of tested Se species to reduce measurement time and to improve peak shapes and thus lower detection limits of Se species.

Table S4. Gradient of the LC-ICP-MS/MS method used for the speciation analysis of Se. Eluents were as follows: A = ultrapure water, and $\mathrm{B}=20 \mathrm{mmol} \cdot \mathrm{L}^{-1}$ ammonium citrate.

| Time | Ammonium Citrate <br> concentration <br> $\left(\mathrm{mmol} \cdot \mathrm{L}^{-1}\right)$ | Percentage of <br> eluent A <br> $(\%)$ | Percentage of eluent <br> $\mathbf{B}$ |
| :---: | :---: | :---: | :---: |
| 0.9 | 5.2 | 74 | $(\%)$ |
| 1.5 | 13 | 35 | 26 |
| 1.8 | 13 | 35 | 65 |
| 1.95 | 5.2 | 74 | 65 |
| 5.25 | 5.2 | 74 | 26 |
|  |  |  | 26 |

Fig. S3 shows the LC-ICP-MS/MS chromatogram obtained with a mixed standard solution of $1 \mu \mathrm{~g} \cdot \mathrm{~L}^{-1}$ using our optimized separation (orange chromatogram) and using the method of Tolu et al. (blue chromatogram). In comparison to the previous method by Tolu et al. (2011), peak shape (both height and width) was greatly improved, and analysis time ( 5.4 min ), consumption of solvent $0.5 \mathrm{~mL} \cdot \mathrm{~min}^{-1}$ and injection volume ( $20 \mu \mathrm{~L}$ ) was significantly reduced.


Fig. S3. Comparison of Se intensity chromatogram obtained for a multi-Se species standard using the initial LC-ICP-MS method described by Tolu et al. (blue line) and with the optimized LC-ICP-MS/MS method (orange line). The multi-Se species standard contained $1 \mu \mathrm{~g}(\mathrm{Se}) \cdot \mathrm{L}^{-1}$ of (1) $\mathrm{SeCys}_{2}$, (2) $\mathrm{Se}^{\mathrm{IV}}$, (3) SeMet and (4) $\mathrm{Se}^{\mathrm{VI}}$. The initial LC-ICP-MS method described by Tolu et al. (blue line) involves a PRP-X100 column ( $250 \times 4.1 \mathrm{~mm}, 10 \mu \mathrm{~m}$ ), an isocratic elution using a mobile phase consisting of $5 \mathrm{mmol} \cdot \mathrm{L}^{-1}$ ammonium citrate and $2 \% \mathrm{MeOH}$ at pH 5.2 , a flow of $1 \mathrm{~mL} \cdot \mathrm{~min}^{-1}$ and an injection volume of $100 \mu \mathrm{~L}$. Our optimized LC-ICP-MS/MS method (orange line) involves a PRP-X100 column ( $150 \times 2.1 \mathrm{~mm}, 5 \mu \mathrm{~m}$ ), a gradient elution from 5.2 to $13 \mathrm{mmol} \cdot \mathrm{L}^{-1}$ ammonium citrate with $2 \% \mathrm{MeOH}$, at pH 5.2 , a flow of $0.5 \mathrm{~mL} \cdot \mathrm{~min}^{-1}$ and an injection volume of $20 \mu \mathrm{~L}$. Shown chromatograms were obtained without adding tetramethylammonium hydroxide (TMAH) post-column.

Finally, $14 \% \mathrm{v} / \mathrm{v}$ tetramethylammonium hydroxide (TMAH) was added to the internal standard solution, which was continuously supplied post-column through a T-piece by using the peristaltic pump of the ICP-MS/MS, to further increase the sensitivity of Se analyses via the well-known carbon enhancement effect (Larsen and Stürup, 1994). The addition of TMAH resulted in an increase of sensitivity of approximately 3.5. With this set-up and considering the pre-concentration step, detection limits of 1-2 ng. $\mathrm{L}^{-1}$ were reached.

Table S5. LC-ICP-MS/MS operating conditions for speciation analysis of Se.

| LC system | Agilent 1260 Infinity II Bio-Inert high performance LC System |
| :---: | :---: |
| Column | Hamilton PRPX-100, 150x2.1mm, $5 \mu \mathrm{~m}$ |
| Injection Volume | $20 \mu \mathrm{~L}$ |
| Mobile Phase | Ammonium Citrate (optimized) - $2 \% \mathrm{MeOH}, \mathrm{pH} 5.2$ |
| Flow Rate | $0.5 \mathrm{~mL} \mathrm{~min}^{-1}$ |
| ICP-MS/MS system | 8900 Agilent ICP-MS/MS |
| Configuration | Quartz torch ( 2.5 mm id ), Ni cones, x -lenses |
| RF power | 1550 W |
| Spray Chamber | Scott type, $+2{ }^{\circ} \mathrm{C}$ |
| Nebulizer Gas | $1.09 \mathrm{~L} \mathrm{~min}^{-1}$ |
| Make up Gas | $0.12 \mathrm{~L} \mathrm{~min}^{-1}$ |
| Cell Gas | $5 \mathrm{ml} \mathrm{min}^{-1}$ of $\mathrm{H}_{2}$ |
| Acquisition time | $\begin{aligned} & 100 \mathrm{~ms} \text { for }{ }^{74} \mathrm{Se},{ }^{76} \mathrm{Se},{ }^{77} \mathrm{Se},{ }^{78} \mathrm{Se},{ }^{80} \mathrm{Se},{ }^{82} \mathrm{Se}, \\ & 50 \mathrm{~ms} \text { for } \mathrm{Br} \text { and } 10 \mathrm{~ms} \text { for } \mathrm{Y} \\ & \hline \end{aligned}$ |

Table S6. LC-ICP-MS/MS operating conditions for speciation analysis of sulfur (S) as published previously by Müller et al. (2019).

| LC system | Agilent $\mathbf{1 2 6 0}$ Infinity I Bio-Inert high performance LC System |
| :--- | :--- |
| Column | Thermofisher Hypercarb, 100x4.6 $\mathrm{mm}, 5 \mu \mathrm{~m}$ |
| Injection Volume | $50 \mu \mathrm{~L}$ |
| Mobile Phase | Formic acid $24-240 \mathrm{mmol} \mathrm{L}^{-1}-1 \% \mathrm{MeOH}, \mathrm{pH} 2.1$ |
| Flow Rate | $1 \mathrm{~mL} \mathrm{~min}^{-1}$ |
| ICP-MS/MS system | $\mathbf{8 8 0 0}$ Agilent ICP-MS/MS |
| Configuration | Quartz torch $(2.5 \mathrm{~mm} \mathrm{id}), \mathrm{Ni}$ cones, x-lenses |
| RF power | 1550 W |
| Spray Chamber | Scott type, $+2^{\circ} \mathrm{C}$ |
| Nebulizer Gas | $1.10 \mathrm{~L} \mathrm{~min}^{-1}$ |
| Make up Gas | $0.12 \mathrm{~L} \mathrm{~min}^{-1}$ |
| Cell Gas | $30 \%$ of $\mathrm{O}_{2}$ with $2 \mathrm{ml} \mathrm{min} \mathrm{mof}^{-1} \mathrm{H}_{2}$ |
| Acquisition time | $50 \mathrm{~ms} \mathrm{for}^{32} \mathrm{~S}$ |
|  | and $20 \mathrm{~ms} \mathrm{for} \mathrm{Sc}, \mathrm{Y}$ |

## S4: Identification of organic compounds by Py-GC-MS

The analysis of organic compounds in aerosols was performed on a FrontierLab pyrolyzer equipped with a FrontierLab AS-1020E autosampler and connected to a Thermoscientific Trace 1310 GC coupled to a Thermoscientific ISQ 7000 MS. The operating conditions and subsequent data processing method were followed as described in Tolu et al. (2015). Briefly, the pyrolysis was performed at $450^{\circ} \mathrm{C}$ and the temperatures of the PyGC interface, GC injector and GC-MS interface were set to $320^{\circ} \mathrm{C}$ and $300^{\circ} \mathrm{C}$, respectively. The injector was operated with He as the carrier gas (at $1.2 \mathrm{~mL} \cdot \mathrm{~min}^{-1}$ ) and a split ratio of $16: 1$. After one minute the gas-saver mode was used with a flow rate of $10 \mathrm{~mL} \cdot \mathrm{~min}^{-1}$ to vent away the pyrolysate bleed of the sample remaining in the pyrolyzer oven. The pyrolysate was separated on a DB-5MS capillary column ( $30 \mathrm{~m} \times 0.25 \mathrm{~mm}, 0.25 \mathrm{~mm}$ film thickness; J\&W, Agilent Technologies AB, Sweden) and the GC temperature program increased from $40^{\circ} \mathrm{C}$ with a rate of $10 \mathrm{C} \cdot \mathrm{min}^{-1}$ to $320^{\circ} \mathrm{C}$, which was held for 5 min . The mass spectrometer with a quadrupole type analyser was operated at unit mass resolution and scanned the mass range from $\mathrm{m} / \mathrm{z} 45$ to 650 at $3.1 \mathrm{scan} \cdot \mathrm{s}^{-1} .70 \mathrm{eV}$ electron bombardment was used for ionization. Peak integration was done using a data processing pipeline under the "R" computational environment. Peak identification was then made using the software "NIST MS Search 2" containing the library "NIST/EPA/NIH 2011" and additional spectra from published studies (Tolu et al., 2015). In total, 105 Py-compounds from 11 classes (e.g. carbohydrates, N compounds, (poly)aromatics) were identified in warm season aerosol samples, respectively (Table S7), along with information on their molecular mass and chemical formula and references for the theoretical mass spectra. For each compound, relative abundance was calculated by setting the total identified peak area for each sample to $100 \%$. The individual Py-products were grouped based on similarities in the molecular structure and origin, which were used for statistical analysis (see defined compound groups in Table S7).

Table S7. List of organic compounds identified by Py-GC-MS in the aerosols from the 2015-2020 time series together with information on their formula, molecular mass (M), specific mass fragments, references for the mass spectra and the compound sub-groupings used for the statistical analyses.

| Name | Formula | $\mathbf{M}$ | Specific mass fragments | Ref |
| :--- | :--- | :--- | :--- | :--- |
| Carbohydrates |  |  | Compound sub- <br> groups |  |
| 4,4-Dimethyl-2- <br> cyclopenten-1-one | C 7 H 10 O | 110 | $41+67+\mathbf{9 5 + 1 1 0}$ | NIST (Cyclo)pentenone/- |
| 4-Cyclopentene-1,3-dione | C 5 H 4 O 2 | 96 | $26+42+54+68+96$ | NIST |


| Alkylamide 1 | unknown |  | 59+72 | NIST |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Alkylamide2 | unknown |  | 59+72 | NIST |  |
| Alkylamide3 | unknown |  | 59+72 | NIST |  |
| Docosenamide | C22H43NO | 337 | 59+72+240+277+294+320+337 | NIST |  |
| Diketopiperazine (DKP) <br> Pro-Pro <br> Diketopiperazine (DKP) <br> Pro-Lys-NH3 | C10H14N2O2 unknown | 194 208 | $70+96+138+166+194$ $70,125,154,166,208$ | $[1,2]$ $[1,2]$ | Proteins |
| Octadecanenitrile | C18H35N | 265 | 57+70+97+110+124 | NIST |  |
| alkanenitrile1 | unknown | x | 57+70+97+110+124 | NIST |  |
| alkanenitrile2 | unknown | x | 57+70+97+110+124 | NIST | Alkylnitriles |
| alkanenitrile3 | unknown | x | 57+70+97+110+124 | NIST |  |
| alkanenitrile4 | unknown | x | 57+70+97+110+124 | NIST |  |
| $n$-alkenes |  |  |  |  |  |
| n-C13:1 | C13H26 | 182 | $56+69+\ldots+182$ | NIST | C13 (odd ${ }^{\circ}$ ) |
| n-C14:1 | C14H28 | 196 | $56+69+\ldots+196$ | NIST |  |
| n-C16:1 | C16H32 | 224 | $56+69+\ldots+224$ | NIST | C14 to C20 (even ${ }^{\circ}$ ) |
| n-C18:1 | C18H36 | 252 | $56+69+\ldots+252$ | NIST |  |
| n-C20:1 | C20H40 | 280 | $56+69+\ldots+280$ | NIST |  |
| n-C15:1 | C15H30 | 210 | $56+69+\ldots+210$ | NIST |  |
| n-C17:1 | C17H34 | 238 | $56+69+\ldots+238$ | NIST | C15 to C19 (odd n ${ }^{\circ}$ ) |
| n-C19:1 | C19H38 | 266 | $56+69+\ldots+266$ | NIST |  |
| n-C21:1 | C21H42 | 294 | $56+69+\ldots+294$ | NIST |  |
| n-C23:1 | C23H46 | 322 | $56+69+\ldots+322$ | NIST |  |
| n-C25:1 | C25H50 | 350 | $56+69+\ldots+350$ | NIST | C21 C33 (odd n ${ }^{\circ}$ ) |
| n-C29:1 | C29H58 | 406 | $56+69+\ldots+406$ | NIST |  |
| n-C31:1 | C31H62 | 434 | $56+69+\ldots+434$ | NIST |  |
| n-C33:1 | C33H66 | 462 | $56+69+\ldots+462$ | NIST |  |
| n-C22:1 | C22H44 | 308 | 56+69+... 308 | NIST |  |
| n-C24:1 | C24H48 | 336 | $56+69+\ldots+336$ | NIST | C22 to C26 (even ${ }^{\circ}$ ) |
| n-C26:1 | C26H52 | 364 | $56+69+\ldots+364$ | NIST |  |
| $n$-alkanes |  |  |  |  |  |
| n-C14:0 | C14H30 | 198 | $57+71+85+\ldots+198$ | NIST |  |
| n-C18:0 | C18H38 | 254 | $57+71+85+\ldots+254$ | NIST | C14 to C20 (even ${ }^{\circ}$ ) |
| n-C20:0 | C20H42 | 282 | $57+71+85+\ldots+282$ | NIST |  |
| n-C15:0 | C15H32 | 212 | $57+71+85+\ldots+212$ | NIST |  |
| n-C17:0 | C17H36 | 240 | $57+71+85+\ldots+240$ | NIST | C15 to C19 (odd n ${ }^{\circ}$ ) |
| n-C19:0 | C19H40 | 268 | $57+71+85+\ldots+268$ | NIST |  |
| n-C21:0 | C21H44 | 296 | $57+71+85+\ldots+296$ | NIST |  |
| n-C23:0 | C23H48 | 324 | $57+71+85+\ldots+324$ | NIST |  |
| n-C25:0 | C25H52 | 352 | $57+71+85+\ldots+352$ | NIST |  |
| n-C27:0 | C27H56 | 380 | $57+71+85+\ldots+380$ | NIST | C21 to C37 (odd n ${ }^{\circ}$ ) |
| n-C29:0 | C29H60 | 408 | $57+71+85+\ldots+408$ | NIST |  |
| n-C31:0 | C31H64 | 436 | $57+71+85+\ldots+436$ | NIST |  |
| n-C37:0 | C37H76 | 521 | $57+71+85+\ldots+521$ | NIST |  |
| n-C22:0 | C22H46 | 311 | $57+71+85+\ldots+310$ | NIST | C 22 to C 38 (even $\mathrm{n}^{\circ}$ ) |


| n-C24:0 | C24H50 | 338 | $57+71+85+\ldots+338$ | NIST |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| n-C32:0 | C32H66 | 450 | $57+71+85+\ldots+450$ | NIST |  |
| n-C34:0 | C34H70 | 480 | $57+71+85+\ldots+480$ | NIST |  |
| n-C36:0 | C36H74 | 506 | $57+71+85+\ldots+506$ | NIST |  |
| n-C38:0 | C38H78 | 535 | $57+71+85+\ldots+535$ | NIST |  |
| Alkanones |  |  |  |  |  |
| 2-Pentadecanone | C15H30O | 226 | $43+58+71$ | NIST | Alkanones |
| Heptadecanone | C17H34O | 254 | $43+58+71+254$ | NIST |  |
| Alkanone | Unknown | x | $43+58+71$ | NIST |  |
| 2-Pentanone | C5H10O | 86 | 43+86 | NIST |  |
| Other aliphatics |  |  |  |  |  |
| 2-Pentene, 2,4-dimethyl- | C7H14 | 98 | $41+55+83+98$ | NIST | Other aliphatics |
| Nonanal | C9H18O | 142 | $29+43+43+56+57+70+98$ | NIST |  |
| Phenols |  |  |  |  |  |
| Phenol | C6H6O | 94 | $39+65+66+94$ | NIST | Phenols |
| Phenol, 4-methyl- | C7H8O | 108 | 77+79+107+108 | NIST |  |
| Phenol, -dimethyl- | C7H8O | 122 | $77+91+\mathbf{1 0 7}+121+\mathbf{1 2 2}$ | NIST |  |
| S compounds |  |  |  |  |  |
| 2(3H)-Benzothiazolone | C7H5NOS | 151 | $96+123+151$ | NIST | S compounds |
| Benzenesulfonamide, N -butyl- | C10H15NO2S | 213 | $51+77+141+170$ | NIST |  |
| Steroids |  |  |  |  |  |
| Stigmasta-3,5-dien-7-one | C29H46O | 411 | $\mathbf{1 6 1 + 1 7 4 + 1 8 7 + 3 9 5 + 4 1 0}$ | NIST | Steroids |
| (Poly)aromatics |  |  |  |  |  |
| Toluene | C7H8 | 92 | 91+92 | NIST | Toluene |
| Benzoic acid | C7H6O2 | 122 | $51+77+\mathbf{1 0 5}+122$ | NIST | Benzoic acid |
| Naphthalene, 2-phenyl- | C16H12 | 204 | 101+204 | NIST | Polyaromatics |
| Phenanthrene, 3,6-dimethyl- | C16H14 | 206 | 102+191+206 | NIST |  |
| Retene or Phenanthrene, 3,4,5,6-tetramethyl- | Both C18H18 | 234 | $165+178+\mathbf{2 0 4}+\mathbf{2 1 9}+234$ | NIST |  |
| 1,4-Diphenyl-1,3- butadiene | C16H14 | 206 | $91+128+191+206$ | NIST | (Alkyl)benzenes |
| Alkyl-Benzene C12 | C18H30 | 246 | 91+92+... 246 | NIST |  |
| Alkyl-Benzene C16 | C22H38 | 302 | $\mathbf{9 1 + 9 2 +} . . .302$ | NIST |  |
| Alkyl-Benzene C17 | C 23 H 40 | 316 | $\mathbf{9 1 + 9 2}+\ldots 316$ | NIST |  |
| Alkyl-Benzene C18 | C24H42 | 330 | 91+92+... 330 | NIST |  |
| Alkyl-Benzene C22 | C28H50 | 386 | 91+92+... 386 | NIST |  |

[1] Chen et al., (2009) Journal of Food Science, 74: 100-105; [2] Fabbri et al., (2012) Journal of Analytical and Applied Pyrolysis, 95: 145-155

S5: Variability of elemental concentrations in digested aerosol filters, cloud water and precipitation


Fig. S4. Concentrations of total Se (shown in blue) and total sulfur ( S ; shown in yellow) in samples of digested aerosol filters collected during the 2015-2020 aerosol time series, as well as in aerosol filter digests, cloud water and precipitation (total deposition) samples collected during the campaign in 2019. Sample size of each sample type set is indicated with N .


Fig. S5. Total concentrations of sulfur (S; panel a), iron (Fe; panel b) and lead ( Pb ; panel c) in acid digests (grey) and water extracts (yellow) of the aerosol filter from the 2015-2020 aerosol time series. The error bars represent the standard deviation values resulting from element quantification by ICP-MS/MS in triplicate.

## S6: Source contribution of Se in aerosols by comparison between aerosol filter digest measurements and

 modelled Se concentration by SOCOL-AERv2

Fig. S6. Relationship between the Se concentrations in aerosol filter digests from the 2015-2020 aerosol time series obtained with the SOCOL-AERv2 model and those measured in acid digests by ICP-MS/MS analysis. The Se concentration values are shown as filled circles, the overall regression of values as a continuous line (correlation coefficient Spearman, $r_{s}=0.822$; $p<0.01)$, and the $1: 1$ line $(\mathrm{y}=\mathrm{x})$ as a dashed line.


Fig. S7. Relationship between the Se concentrations in aerosol filter digests from the 2015-2020 aerosol time series obtained with the SOCOL-AERv2 model and those measured in acid digests by ICP-MS/MS analysis organized according to Se source contributions: a) terrestrial, b) marine, c) volcanic, and d) anthropogenic. The size of the data points correspond to proportions of respective modelled source contributions (in \% total sources). The 1:1 line $(y=x)$ is shown as a dashed line.

S7: Variability of moistures sources and water isotopes in precipitation and cloud water


Fig. S8. Distribution of contributing moistures source of precipitation samples (panel a) and cloud water samples (panel b) collected during the campaign in 2019. In panel a, the x axis shows the 18 collected precipitation events ( $\mathrm{P} 1-\mathrm{P} 18$ ), as well as their sub-samples for events P3, P5, P9, P11, and P17. In panel b, the x axis shows the 11 collected cloud events ( $\mathrm{C} 1-\mathrm{C} 11$ ), as well as their sub-samples (for all events, except for event C 5 for which only one sub-sample was taken). The data are shown as proportion of total moistures sources (\%). Modelled regional moistures sources discussed in the manuscript include sources from France, Spain and Local. Modelled moistures sources with average values <3 \% are not shown (e.g., moisture sources from Eastern Europe, America, Atlantic Subtropics, Portugal, UK, Ireland).


Fig. S9. Total Se concentrations and isotopic composition of deuterium ( $\delta^{2} \mathrm{H}$ : isotopic ratio relative to Vienna Standard Mean Ocean Water, expressed in per mill) in cloud water samples collected during the campaign in 2019. Each data point is coloured according to the dominant moisture sources of the cloud water samples, i.e., samples with dominant regional moisture sources (including moisture sources from France, Spain and Local) are shown in yellow, those with dominant moisture source from North Africa in brown, those with dominant moisture source from the Atlantic Ocean in blue, and those with dominant moisture source from other areas in grey. Sub-samples potentially influenced by precipitation are highlighted by a black symbol outline. Regional moistures sources discussed in the manuscript include sources from France, Spain and Local (local source, in France south of $43.6^{\circ} \mathrm{N}$ ). The overall regression of values as a continuous line (correlation coefficient Spearman, $r s=0.758 ; p<0.01$ ). The error bars for the $y$ variable, i.e. Se concentration, represent the standard deviation values resulting from element quantification by ICP-MS/MS in triplicate. The error bars for the $x$ variable, i.e. $\delta^{2} H$, represent the standard deviation values resulting from stable water isotope analysis by Picarro in triplicate.


Fig. S10. Total selenium ( Se ) deposition and isotopic composition of deuterium ( $\delta^{2} \mathrm{H}$ : isotopic ratio relative to Vienna Standard Mean Ocean Water, expressed in per mill) in precipitation samples collected during the campaign in 2019. Each data point is coloured according to the dominant moisture sources of the precipitation samples, i.e., samples with dominant regional moisture sources (including moisture sources from France, Spain and Local) are shown in yellow, those with dominant moisture source from North Africa in brown, and those with dominant moisture source from the Atlantic Ocean in blue. Regional moistures sources include sources from France, Spain and Local (local source, in France south of $43.6^{\circ} \mathrm{N}$ ). The error bars for the y variable, i.e. Se concentration, represent the standard deviation values resulting from element quantification by ICP-MS/MS in triplicate. The error bars for the x variable, i.e. $\delta^{2} \mathrm{H}$, represent the standard deviation values resulting from stable water isotope analysis by Picarro in triplicate.

Table S8. Additional information on sampled precipitation (sub-)events. The precipitation types include rain (RW), hail (H), snow/sleet (S/S), and light rain/cloud water (LR/CW).

| Date | Sub-event label | Sub-sampling time | Duration (min) | Precipitation type | $\begin{gathered} \text { Amount } \\ (\mathrm{mL}) \end{gathered}$ | pH | Rain rate ( $\mathrm{mm} \cdot \mathrm{h}^{-1}$ ) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 01.09.2019 | P1 | 04:15:00-5:15 | 60 | LR/CW | 190 |  | 0.4 |
| 05.09.2019 | P2 | 03:45-10:00 | 375 | LR/CW | 264 |  | 0.1 |
| 10.09.2019 | P3.1 | 02:10-09:10 | 420 | S/S | 1000 | 5.6 | 0.3 |
|  | P3.2 | 02:10-09:10 | 420 | S/S | 1000 |  | 0.3 |
| 10.09.2019 | P4 | 09:30-11:45 | 135 | S/S | 450 | 5.6 | 0.4 |
| 10.09.2019 | P5.1 | 17:45-20:00 | 135 | S/S | 759 | 5.5 | 0.7 |
|  | P5.2 | 20:00-21:00 | 60 | S/S | 653 | 5.6 | 1.4 |
|  | P5.3 | 21:00-22:30 | 90 | S/S | 555 | 5.6 | 1.2 |
|  | P5.4 | 22:30-01:30 | 180 | S/S | 391 | 5.6 | 0.4 |
| 14.09.2019 | P6 | 19:20-19:50 | 30 | RW | 348 |  | 1.5 |
| 15.09.2019 | P7 | 17:30-18:20 | 50 | RW/H | 1700 | 6.5 | 4.3 |
| 16.09.2019 | P8 | 21:25-22:15 | 50 | RW | 1196 | 6.3 | 4.5 |
| 17.09 .2019 | P9.1 | 15:08-15:58 | 50 | RW | 657 | 7.2 | 1.7 |
|  | P9.2 | 15:58-16:08 | 10 | RW/H | 472 | 7.2 |  |
| 18.09.2019 | P10 | 17:00-8:00 | 900 | RW | 2108 | 6.8 | 2.2 |
| 18.09.2019 | P11.1 | 15:14-16:10 | 56 | RW/H | 3101 | 6.0 | 9.5 |
|  | P11.2 | 15:14-16:10 | 71 | RW/H | 1138 | 6.0 | 9.5 |
| 19.09.2019 | P12 | 22:00-24:00 | 120 | RW | 110 |  | 0.1 |
| 22.09.2019 | P13 | 6:05-11:50 | 345 | LR/CW | 418 | 5.8 | 0.2 |
| 22.09.2019 | P14 | 11:15-15:15 | 240 | LR/CW | 280 | 5.7 | 0.2 |
| 01.10.2019 | P15 | 16:00-17:46 | 106 | RW/H | 1004 | 5.7 | 1.2 |
| 01.10.2019 | P16 | 18:38-18:50 | 12 | RW/H | 973 | 5.4 | 10.2 |
| 02.10.2019 | P17.1 | 1:00-5:30 | 270 | S/S | 960 | 5.4 | 0.5 |
|  | P17.2 | 5:30-11:50 | 380 | S/S | 2730 | 5.5 | 0.9 |
|  | P17.3 | 11:50-13:40 | 110 | LR/CW | 752 | 5.7 | 0.9 |
| 02.10.2019 | P18 | 13:40-20:00 | 500 | LR/CW | 624 | 5.7 | 0.2 |

## S8: High Se and other elements deposition associated with deep convective activity during thunderstorms

## S8.1 Principle component analysis of precipitation samples

Variability in elemental concentrations in precipitation (given in $\mu \mathrm{g} \cdot \mathrm{L}^{-1}$ ) were explored using principal component analysis (PCA). The first two principal components (PC1-2) are shown in Fig. S11a. The precipitation volume was included passively in the PC-loading plots by using bivariate correlation coefficients between these variables and the PC scores of each PC. PC1 shows positive loadings of many major and trace elements (e.g. $\mathrm{Na}, \mathrm{K}, \mathrm{P}, \mathrm{Fe}, \mathrm{Cu}$, $\mathrm{Zn}, \mathrm{As}, \mathrm{Pb}$ ), which significantly correlate with the collected precipitation amount ( $p<0.05$ ). Elemental concentration in precipitation are expected to be influenced by the dilution effect, which describes lower occurring concentrations at increasing rain volumes. PC scores of sub-events classified by their precipitation type, including events with rainwater/hail, light rain/cloud water or snow/sleet, show clear groupings on PC1 and PC2 (Fig. S11b). Particularly events that included light rain (light rain/cloud water) plot on PC1, which indicates that these events were primarily affected by the dilution effect.


Fig. S11. Loading plot (panel a) and score plot (panel b) for principal components (PCs) 1-2 resulting from the principal component analysis performed with the element concentrations in precipitation samples from the 2019 campaign. PC1-2 account together for $63 \%$ of the total variance. For the PC-loadings, filled circles correspond to active variables, and the precipitation amount (yellow diamond) was added passively. Significance levels (PC-loading values $>0.39$ ) are indicated in loading plots by horizontal and vertical dashed line.

## S8.2 Precipitation chemistry



Fig. S12. Comparison between black carbon (BC; panel a) concentration (indicated by optical absorbing suspended particles at 880 nm ) and the total column ice cloud water content (TCIW, in mm ; panel b) during the dust period and all other events. The letters x and y denote that the two sample set are significantly different (Mann-Whitney- U test; $p<0.01$ ).


Fig. S13. Comparison between the total Se and S deposition during thunderstorms (yellow) and all other precipitation events (grey). Panel a) shows the data for Se and panel b) shows the data for $S$. The letters $x$ and $y$ denote that the two sample sets are significantly different in terms of total Se deposition in panel a and total $S$ deposition in panel $b$ (Mann-Whitney-U test; $p<0.01$ ).

## S8.3 Aerosol chemistry and meteorological characteristics



Fig. S14. Comparison between the total Se concentrations in aerosol filter digests from the 2015-2020 time series which were associated with (yellow) and without (grey) thunderstorms. Panel a) shows the Se concentrations in aerosol filter digests considering the entire 2015-2020 aerosol time series dataset. Panel b) shows the Se concentrations in aerosol filter digests considering only the samples taken during summer months. The letters $x$ and $y$ denote that the two sample sets are significantly different (Mann-Whitney-U test; $p<0.01$ ).

Table S9. $p$ values associated to Mann-Whitney-U significant difference test performed between total concentrations of various elements in aerosol filter digests from the 2015-2020 aerosol time series associated with or not associated with thunderstorms when i) considering the entire 2015-2020 aerosol time series dataset ( p values given in row "full series"); and ii) considering only the samples taken during summer months ( $p$ values given in row "only summer"). $p$ values $<0.01$ are highlighted in green, while $p$ values $p<0.05$ are highlighted in yellow. Elements with substantial periods of missing data were excluded from statistical analysis (i.e., $\mathrm{Na}, \mathrm{Ag}, \mathrm{Cd}$ and Ba ).

$$
\begin{array}{lllllllllllllllllllll}
\mathrm{Se} & \mathrm{As} & \mathrm{~S} & \mathrm{Mg} & \mathrm{Al} & \mathrm{~K} & \mathrm{Ti} & \mathrm{~V} & \mathrm{Cr} & \mathrm{Mn} & \mathrm{Fe} & \mathrm{Co} & \mathrm{Ni} & \mathrm{Cu} & \mathrm{Zn} & \mathrm{Rb} & \mathrm{Sr} & \mathrm{Nb} & \mathrm{Mo} & \mathrm{Cs} & \mathrm{~Pb}
\end{array} \mathrm{U}
$$

$\begin{array}{llllllllllllllllllllll}\text { full series } & 0.000 & 0.000 & 0.000 & 0.000 & 0.000 & 0.000 & 0.008 & 0.000 & 0.000 & 0.000 & 0.000 & 0.000 & 0.000 & 0.000 & 0.016 & 0.000 & 0.000 & 0.000 & 0.167 & 0.000 & 0.000 \\ 0.000\end{array}$

| only summer | 0.003 | 0.001 | 0.011 | 0.007 | 0.000 | 0.004 | 0.232 | 0.005 | 0.009 | 0.001 | 0.004 | 0.001 | 0.063 | 0.029 | 0.818 | 0.001 | 0.006 | 0.033 | 0.388 | 0.006 | 0.003 | 0.002 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |



Fig. S15. Concentrations of measured elements in aerosol filter digests of the 2015-2020 aerosol time series. Additional chemical parameters shown include black carbon (BC) and particle number (PN). Aerosol samples associated with thunderstorms are shown at the bottom of the diagram with yellow diamonds.


Fig. S16. Comparison of modelled SOCOL-AERv2 results with measurements of 2015-2020 aerosol time series. Se measurements in aerosol filter digests sampled during weeks with thunderstorms (shown as yellow circles, based on reported lightning activity by Blitzortung network within 3 km of Pic du Midi Observatory). All other aerosol measurements without detected thunderstorms during sampling are shown as black filled circles ("others"). The overall regression of all values is shown as a continuous line and the $1: 1$ line as a dashed line.


Fig. S17. Comparison between the absolute humidity for sampling periods in the 2015-2020 aerosol time series which were associated with (yellow) and without (grey) thunderstorms. Absolute humidity was calculated as (RH $\times \mathrm{Ps}$ ) $/(\mathrm{Rw} \times \mathrm{T} \times 100$, with relative humidity ( $\mathrm{RH} ; \%$ ), Saturation vapour pressure ( $\mathrm{Ps} ; \mathrm{Pa}$ ), the specific gas constant for water vapour ( $\mathrm{Rw} ; 461.5 \mathrm{~J} \cdot \mathrm{~kg}^{-1} \cdot \mathrm{~K}^{-}$ ${ }^{1}$ ) and the air temperature ( $\mathrm{T}, \mathrm{K}$ ). The letters x and y denote that the two sample sets are significantly different (Mann-WhitneyU test; $\mathrm{p}<0.01$ ).

S9: Additional information on inorganic and organic Se and S speciation
(a) Example of Se chromatograms

## (b) Example of S chromatograms



Fig. S18. Examples of intensity chromatograms of Se (panel a) and S (panel b) obtained for Pic du Midi aerosol water extracts (Se: S18_2, S: S18_12), cloud water (Se: C1.2, S: C12.1) and precipitation (Se: P8, S: P11.2). Identified Se and S species are indicated in respective chromatograms


Fig. S19. Intensity chromatograms of Se in an aerosol water extract (unspiked; black line) and spiked with a dimethylselenonium propionate ( DMSeP ) standard (yellow line). These chromatograms were obtained, using an Ionosphere5C column (Chrompack, 100x3 mm, $5 \mu \mathrm{~m}$ ) and the procedure of Larsen et al. (2001), for the aerosol water extract S16_S3 from the 2015-2020 aerosol time series. DMSeP was spiked at $70 \mathrm{ng} \cdot \mathrm{L}^{-1}$ using a DMSeP standard synthesized in-house according to published procedure by W.-M. Fan et al. (1998).


Fig. S20. Recoveries of Se (blue boxplots) and S (yellow boxplots) species in water extracts of aerosols from the 2015-2020 aerosol time series, as well as in cloud water and precipitation samples collected during the campaign in 2019. Recoveries of Se and $S$ species are given in percentage of total Se and S concentrations in the samples, respectively. Recoveries of $S$ above $100 \%$ are likely related to high and variable blank values for sulfate $\left(\mathrm{SO}_{4}{ }^{2-}\right)$ leading to overestimation of $\mathrm{SO}_{4}{ }^{2-}$ in samples with low $S$ concentration. Sample size of each sample type set is indicated with $N$.

## S10: Additional information on factors driving the deposited chemical form of Se

Table S10. Correlation coefficients between the abundance of (sub-)groups of organic compounds identified by Py-GC-MS and the proportions of Se species (with respect to total Se concentrations) for the 2015-2020 aerosol time series. Positive and negative correlations (determined using Spearman correlation coefficient) are highlighted in blue and orange, respectively. Different significance levels are indicated by colour strength: $* * p<0.01$ (dark blue/orange), ${ }^{*} p<0.05$ (light blue/orange).

|  | Toluene | (Poly)aromatics | Alkanones | Carbohydrates | Carboxylic acids | N -comp. | Alkanes | Alkenes | Other aliphatics | Pheno | -comp | Steroids | n | Levoglucosan |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total Se | 0.157 | -0.121 | -0.152 | -0.230 | 0.141 | -0.053 | 0.211 | 0.184 | -0.149 | 0.099 | -0.314* | -0.015 | 0.130 | 0.255* |
| Total water soluble Se | 0.169 | -0.056 | -0.180 | -0.090 | 0.062 | -0.017 | 0.190 | 0.235 | -0.104 | 0.171 | -0.349" | 0.031 | 0.198 | 0.243 |
| OrgSe | -0.474** | -0.377 ${ }^{\text {² }}$ | -0.089 | -0.046 | 0.192 | 0.299* | -0.204 | -0.106 | 0.086 | -0.063 | -0.160 | 0.282* | 0.532" | 0.149 |
| Se ${ }^{\text {IV }}$ | 0.312* | 0.434" | -0.023 | -0.209 | -0.187 | 0.025 | -0.011 | -0.074 | -0.292* | 0.085 | 0.353" | -0.079 | -0.306 ${ }^{*}$ | -0.354" |
| $\mathrm{Se}^{\mathrm{VI}}$ | -0.098 | -0.196 | 0.296* | -0.178 | -0.208 | -0.068 | 0.184 | 0.265* | 0.246 | 0.334" | -0.017 | -0.258 | -0.037 | -0.108 |



Fig. S21. Ratio between selenate $\left(\mathrm{Se}^{\mathrm{VI}}\right)$ and sulfate $\left(\mathrm{SO}_{4}{ }^{2-}\right)$ for precipitation events collected during the campaign in 2019 with low and dominant Atlantic moisture sources. The average contributions of Atlantic moisture sources of the two shown classifications are $4 \pm 1 \%$ (low) and $41 \pm 19 \%$ (dominant). The letters x and y denote that the two sample set are significantly different (Mann-Whitney-U test; $p<0.01$ ).

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