# 1 Differentiating between crop and soil effects on soil moisture

## 2 dynamics

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- 9 Abstract. There is urgent need to for developing sustainable agricultural land use schemes. Intensive crop production has
- 10 induced increased greenhouse gas emissions and enhanced nutrient and pesticide leaching to groundwater and streams. On the
- 11 one side, cClimate change is also expected to increase drought risk as well as the frequency of extreme precipitation events in
- 12 many regions. On the other side, crop production has induced increased greenhouse gas emissions and enhanced nutrient and
- 13 pesticide leaching to groundwater and receiving streams. Consequently, sustainable management schemes require sound
- 14 knowledge of site-specific soil hydrological water processes, that accounting explicitly take into account for the interplay
- between soil heterogeneities and crops. In this study, we appliedy a principal component analysis to a set of 64 soil moisture
- 16 time series from a diversified cropping field featuring seven distinct crops and two weeding management strategies.
- 17 Results showed that aAbout 97% of the spatial and temporal variance of the data set was explained by the first five principal
- 18 components. Meteorological drivers accounted for 72.3% of the variance, 17.0% was attributed to different seasonal behaviour
- 19 of different crops. While the third (4.1%) and fourth (2.2%) principal component explained were interpreted as effects of soil
- 20 texture and cropping schemes on soil moisture variance, respectively, the effect of soil depth was represented by the fifth
- 21 component (1.7%). However, neither topography nor weed control had a significant effect on soil moisture variance. Contrary
- 22 to common expectations, soil and rooting pattern heterogeneity seemed not to play a major role. Findings of this study highly
- 23 depend on local conditions. However, we consider the presented approach generally applicable to a large range of site
- 24 conditions.

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#### 1 Introduction

- 26 Agriculture plays a major role to ensure the provision of food to a growing global population. At the same time, climate change
- 27 is putting yield stability at risk due to extreme weather events, rising the need for sustainable management of resources, such
- 28 as water and soil (Trnka et al., 2014). As part of the adaptation to more challenging conditions, the transformation from large
- 29 homogeneously cropped fields towards diversified agricultural landscape was identified not only to have positive effects on
- 30 multiple ecosystem services. The transformation from large homogeneously cropped fields towards diversified agricultural
- 31 landscapes has been identified as an opportunity that can contribute to climate adaptation due to the positive effects on multiple

32 ecosystem services (Tamburini et al., 2020), but also on the system's resilience to climatic extremes and cropping system 33 resilience to climatic extremes (Birthal and Hazrana, 2019). Additionally, crop diversification is highly beneficial by reducing 34 soil erosion through permanent soil cover (Paroda et al., 2015), and by improving resource use efficiency through wider crop 35 rotations (Rodriguez et al., 2021). 36 In terms of soil hydrological water dynamics, crop and management diversification can lead to improved water-stable macro-37 aggregation, reduced soil compaction and increased soil organic carbon, which can reduce soil water infiltration and improve 38 water retention (Alhameid et al., 2020; Fischer et al., 2014; Karlen et al., 2006; Koudahe et al., 2022; Nunes et al., 2018). 39 Korres et al. (2015) reported that spatial variability of soil moisture was mainly driven by soil characteristics, followed by crop cover and management. Soil moisture is also affected by soil texture and pore size distribution (Krauss et al., 2010; Rossini et 40 41 al., 2021; Pan and Peters-Lidard, 2008). The quantification of the impact of these effects on soil moisture variability is 42 important, for instance for hydrological applications and adopted management practices in agriculture (Hupet and Vanclooster, 43 2002). 44 However, aAs the diversity of independent variables in agricultural systems increases, demands for frequency and spacing of 45 soil moisture measurements and related data interpretation grow. Therefore, soil sensoring networks are receiving increased 46 attention, particularly in Precision Agriculture (PA; )-Bogena et al., 2022; Salam and Raza, 2020), where the main goal is to 47 increase efficiency and productivity at the farm level, while minimizing the negative impacts on the environment (Taylor and 48 Whelan, 2010). Soil sensor networks can meaningfully contribute to PA as they can be used for various purposes, including 49 the delineation of management zones (Khan et al., 2020; Salam and Raza, 2020). Still, one of the most important demands to 50 be fulfilled by soil sensoring networks is soil moisture monitoring, as accurate measurement of soil water content can play an 51 important role in improving water management and therefore, crop yields (Salam, 2020). 52 Wireless solutions, for instance based on LoRaWAN (Long Range Wide Area Network) technology, in combination with 53 electromagnetic soil moisture sensors avoid labour-intensive and destructive soil moisture measurements that disrupt field 54 traffic. The development of such wireless soil monitoring sensor networks (WSN) enables broad and affordable application 55 also in areas with low cellular coverage (Cardell-Oliver et al., 2019; Lloret et al., 2021; Placidi et al., 2021; Prakosa et al., 56 2021). 57 The evolvement of such systems WSN does not only have benefits for management but is also of high relevance for fostering 58 the understanding of hydrological dynamics in the vadose zone. High-resolution datasets measured under real farming 59 conditions can be used to characterize and analyse spatio-temporal dynamics of soil water. Due to the large size of data sets 60

conditions can be used to characterize and analyse spatio-temporal dynamics of soil water. Due to the large size of data sets that are recorded with wireless sensor networks WSN, sophisticated data analysis approaches are required to detect hidden patterns and determine influence factors on soil moisture variability (Vereecken et al., 2014). Methods include geostatistical analysis (Vereecken et al., 2014) or data driven approaches (Hong et al., 2016). With the introduction of multiple-points geostatistics, it became possible to not only analyse patterns but also connect them with factors affecting soil moisture, such as topography, texture, crop growth and water uptake, and land management (Brocca et al., 2010; Strebelle et al., 2003). Wavelet analysis can analyse both localized features as well as spatial trends through which non-stationary variation of soil

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properties can be considered (Si, 2008). Cross-correlation analysis allowed linking soil moisture variability to climatic 66 67 variables (Mahmood et al., 2012). Furthermore, temporal stability analyses detect spots in the investigated area which are 68 consistently wetter or drier than the mean soil moisture (Baroni et al., 2013; Vachaud et al., 1985, Vanderlinden et al., 2012). 69 This method was already successfully used to detect soil moisture patterns related to soil properties, vegetation, and topography 70 (Zhao et al., 2010). 71 Principal component analysis (PCA) is another method that was successfully applied for soil moisture variability analysis at 72 the field (Hohenbrink et al., 2016; Hohenbrink and Lischeid, 2015; Martini et al., 2017), catchment (Korres et al., 2010; 73 Lischeid et al., 2017; Nied et al., 2013; Graf et al., 2014), and regional (Joshi and Mohanty, 2010) scale. These studies build 74 on previous applications in climatology where the term "Empirical Orthogonal Functions" is used (Bretherton et al., 1992) and 75 are examples for how. Space and time dimensions can be disentangled and be assigned to influencing factors. Additionally, 76 the propagation of hydrological signals (e.g. precipitation events) over depth can be assessed (Hohenbrink et al., 2016). This 77 opens up great opportunities to for contributing to timprove the knowledge he knowledge of changing soil water hydrological 78 dynamics in complex diversified agricultural systems with increasing heterogeneity (e. g. soil texture) and site-specific 79 adjustment of crop soil types and field management which, to our knowledge, have hardly been studied so far. 80 We The main objective of this study was to identify the drivers of soil moisture variability in a diversified cropping field in 81 terms of soil texture, crop selection and field management by applying PCA. Special focus was put on the interpretation of 82 spatial and temporal effects of crop diversification and of soil heterogeneities on soil moisture dynamics. 83 For this, we analysed a high-resolution soil moisture data set measured by a novel underground LoRaWAN monitoring system 84 with soil moisture TDR sensors in different depths of the vadose zone at a spatial-temporally diversified agricultural field in 85 Northeast Germany, The novelty of this Internet of underground Things (IouT) soil moisture monitoring network WSN relies on its is characterized by its unique on-farm installation environment. The deployment of transmission units in 0.3 m soil depth 86 87 and and the deployment of 180 sensors in up to 0.9 m soil depth, allowsing high spatio-temporal resolution wireless data 88 transmission, and enablesing conventional farming practices like machinery traffic, tillage and mechanical weeding. The main

#### 92 2 Materials and methods

#### 93 2.1 Study site

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94 The study site (52°26'51.8"N 14°08'37.7"E, 66-83 m.a.s.l.) is located near the city of Müncheberg in the federal state of 95

Brandenburg in Northeastern Germany. The landscape is classified as a hummocky ground moraine that formed during the

objective of this study was to identify the drivers of soil moisture variability in a diversified cropping field in terms of soil

texture, crop selection, soil type and field management by applying PCA. Special focus was put on the interpretation of spatial

and temporal effects of crop diversification and of soil heterogeneities on soil moisture dynamics.

last glacial periods. Glacial and interglacial processes as well as subsequent erosion resulted in highly heterogeneous soils

97 (Deumlich et al., 2018), being classified as Dystric Podzoluvisols according to the FAO scheme (Fischer et al., 2008). In the top 0.3 m soil layer, total organic carbon was 0.94% and total nitrogen content was 0.07%, and pH was 6.12. Between January 1991 and December 2020, the mean annual temperature in Müncheberg was 9.6°C, and the mean annual sum of precipitation was 509 mm (DWD Climate Data Center (CDC), 2021).

#### 2.2 Experimental setup

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102 The data collection was carried out from December 2020 until mid of August 2021 in the patchCROP experiment (Grahmann 103 et al, 2021; Donat et al., 2022). This landscape experiment has been set up to study the multiple effects of cropping system 104 diversification on productivity, crop health, soil quality, and biodiversity. To that end, a cluster analysis was carried out based 105 on soil maps and multi-year (2010 to 2019) yield data to identify high and low yield potential zones in the 70-ha large field 106 (Donat et al., 2022). Afterwards, single experimental units comprising 30 patches with an individual size of 0.52 ha (72 m × 107 72 m) each, have been implemented in both, high and low yield potential zones where each of those zones is characterized by 108 varying soil conditions and a site-specific five-year, legume-based crop rotation (Grahmann et al., 2021). The remaining area 109 outside of the 30 patches was planted with winter rye. For the current study, twelve out of 30 patches were considered (Table 110 1, Figure 1). Specific patches were selected to capture the soil heterogeneities in terms of soil texture, but also the seasonal 111 patterns of the crop rotation that may have important effects on the soil water dynamics such as crop types, presence of cover 112 crops or fallow periods. In the cropping season 2020/2021, seven different main crops were grown. For subsequent data 113 interpretation, crops have been grouped into A) winter crops, B) fallow, followed by summer crops and C) cover crops, 114 followed by summer crops. In seven out of twelve considered patches, weed control was carried out with herbicide application, 115 referred as "conventional" pesticide application, while in the remaining five patches, "reduced" pesticide management was 116 carried out by mainly using mechanical weeding, by harrowing, blind harrowing, and hoeing. Only in the case of high weed 117 pressure herbicides were applied. Due to the potential impact of mechanical weeding, i.e., on rainwater infiltration, soil 118 evaporation and topsoil rooting intensity, we differentiate between these modes of weed control.

#### 2.3 Data collection

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#### 2.3.1 Soil moisture data

121 Soil moisture was recorded by a long-range-wide-area network (LoRaWAN) based monitoring systemWSN. In each patch, 122 one Dribox box equipped with a SDI-12 distributer (serial data interface at 1200 baud rate, TBS04, TekBox, Saigon, Vietnam) 123 connected to six TDR-sensors (TDR310H, Acclima, Meridian, USA) and attached to an outdoor remote terminal unit (RTU) 124 fully LoRaWAN compliant (TBS12B: 4+1 channel analogue to SDI-12 interface for 24 Bit A/D conversion of sensor signals, 125 TekBox, Saigon, Vietnam) was installed as LoRa node. The DriboxIt was deployed at least 0.3 m below ground to allow 126 normal-field traffic and soil tillage. The sensors and boxes were installed between August and November 2020. At two 127 georeferenced locations within each patch, At two georeferenced locations within each patch (Figure 2), TDR-soil moisture 128 sensors\_were installed in 0.3, 0.6 and 0.9 m depth, respectively. Sensors were, approximately 2 m apart from the Driboxes

- 129 <u>LoRa node</u> in angles between 45° and 60° (Figure 1). Soil <u>moisture</u> sensors at 0.3 m were placed horizontally, while sensors
- 130 at 0.6 and 0.9 m depth were placed vertically using auger-made tunnels boreholes and extension tubes for soil insertion.
- 131 Communication of Driboxes LoRa nodes was wireless and ere-autarkic in in terms of energy supply. and communication
- 132 was wireless throughout. Thus, no electric cabling except from connections between sensors and Driboxes LoRa nodes was
- 133 needed. Under optimum conditions, battery running time of the LoRa nodes can be up to 12 months but can be reduced to 8
- months when radio transmission is attenuated (e.g. due to near water-saturated soil) which then increases power consumption
- 135 (Bogena et al., 2009). The data were Data was recorded every 20 minutes by the LoRa nodes through a LoRa-WAN Gateway
- DLOS8 (UP GmbH, Ibbenbüren, Germany) which was equipped with the modem TL-WA7510N (TP Link, Hong Kong,
- 137 China) to transfer the data to a cloud from where collected data could be accessed directly after the measurement. The time
- series included in this study covered the period from December 01, 2020, until August 14, 2021 (Figure 2).

#### 139 **2.3.2 Weather data**

- 140 Precipitation and temperature data (Figure 3) with a 15 min temporal resolution were obtained from two weather stations
- 141 located in the Eastern and Western end of the main patchCROP field with a 15 min temporal resolution. Climatic water balance
- 142 was calculated from precipitation and potential evapotranspiration, both measured at the climate station by the German
- 143 Weather Service in Müncheberg (DWD Climate Data Center (CDC), 2021). This station was chosen due to its proximity to
- the study site.

#### 145 2.3.3 Remote senses data for vegetation dynamics

- 146 Furthermore, drone imagery from May 20, 2021, May 31, 2021, and July 06, 2021, was used for vegetation assessment. The
- drone fixed-wing UAV-based RS eBee platform (SenseFly Ltd., Cheseaux-Lausanne, Switzerland) was operated at noon time
- and recorded multispectral imagery with a Parrot Sequoia+ camera (green, red, NIR, and red edge bands, spatial resolution of
- 149 0.105 m) and thermal imagery of the surface (only on May 31, 2021) with a senseFly Duet T camera with a spatial resolution
- 150 of 0.091 m (Table 2). The multispectral imagery was processed with Pix4D to obtain the Normalized Difference Vegetation
- 151 Index (NDVI), following Eq. (1):

$$NDVI = \frac{NIR-Red}{NIR+Red} \tag{1}$$

- in which NIR is the intensity of reflected near-infrared light (reflected by vegetation) and Red the intensity of reflected red
- light (absorbed by vegetation). A digital elevation model with a spatial resolution of 1 m (GeoBasis-DE and LGB, 2021) was
- used to calculate the slope (ArcGIS 10.7.0; ESRI, 2011) (Table 2).

#### 156 **2.3.4 Soil information**

#### 157 Soil texture by layer

Manual classification of soil texture analysis by layer was carried out for part of the sensors by using collecting 140 samples in eight of twelve analysed patches. Samples were taken with taken with a 1 m-length Pürckhauer soil auger of 1 m length, in eight of twelve analysed patches. Manual sSoil textural class was manually determined estimated at the field by applying the protocol "Finger test to determine soil types-texture according to DIN 19682-2 and KA5" (Sponagel et al., 2005). Additionally, representative soil samples were collected and analysed at the laboratory to determine particle size distribution for sand, silt, and clay (soil texture based on the German particle classification) by. Soil texture was analysed following the DIN ISO 11277 (2002) reference method by wet sieving and sedimentation, using the SEDIMAT 4-12 (Umwelt-Geräte-Technik GmbH, Germany). The sand fraction in this method is defined between 2 and 0.063 mm, according to IUSS Working Group WRB (2015).To extrapolate the laboratory-based soil particle distribution from the laboratory to the manual soil textural classes manually determined at the field, the high and low yield potential laboratory samples were pooled separately. and Tthe average soil particle distribution was calculated by for each soil textural class was calculated and assigned to the respective soil layer with that specific<del>particular</del> soil textural class. The soil texture analysis showed that soil texture variability increased with depth. In the third layer (average bottom depth =  $\frac{92 \text{ cm}}{0.92}$  m), the sand and clay share-content across 133 sampling points varied between 53% to 94% and 2% to 22%, respectively. Soil samples ampling points were between located approximately 0.8 m and 2.5 m far-away from the soil moisture sensors to minimize damage risk. The transferability of texture information from the sampling point to the sensor location was not ensured due to high nugget effects. Furthermore, manual soil texture analysis data were not available for all analysed patches. Consequently, they were not included into further correlation analysis.

#### Topsoil proximally sensed data

In October 2019, the "Geophilus" soil scanner system (Lueck and Ruehlmann, 2013) was used in the entire field to map soil electrical resistivity (ERa) of the soil as a proxy for soil texture for the top soil, using reference soil samples to calibrate the readings. A total of four georeferenced reference soil samples were taken until 0.25 m soil depth, and locations were selected based on the proximal soil sensor data (sensor-guided sampling; Bönecke et al., 2021). The "Geophilus" system is based on sensor fusion of within which ERa sensors are coupled with a gamma-ray detector(y) sensor. Apparent electrical conductivity was measured by pulling one or more sensor pairs mounted on wheels across the field where each pair of sensors measured a different soil depth. Amplitude and phase were measured simultaneously using frequencies from 1 MHz to 1 kHz. Reference soil samples were taken in several points analysed via soil-particle size analysis according to DIN ISO 11277 (2002) and served as calibration information in order to estimate sand, silt and clay content in the top 0.25 m of soil for soil the entire field. A non-linear regression model was applied. The RMSE of sand content (5.7%) was considerably smaller than the standard deviation of the sand content in the first layer from the manual soil texture analysis (11.9%), indicating a satisfactory prediction performance. The gammay-sensor was used to minimize uncertainties, being less sensitive to soil moisture than the

ERa readings (Bönecke et al., 2021). The estimated sand content in the upper 0.25 m at the study site varied between 69.1% and 81.2% and averaged 79.0% (Table 1, Figure 1).

#### 2.4 Data processing

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192 Soil moisture data were available at 20-minute intervals. Transmission failures due to discharged batteries, signal disturbances 193 in sinks-after rainfall, in patches with a high density of biomass (e.g. maize), and theft of parts of the monitoring system WSN 194 led to data gaps that affected in some cases all sensors of the WSN and amounted to 81 out of 257 days of the measuring 195 period. The affected days, which were therefore skipped for the analysis. Whereas time series of eight sensors were excluded 196 due to a higher frequency of transmission failures, in total, 64 time series were used for the analysis, and additional data gaps 197 for single sensors were interpolated linearly. Of all 20,668 interpolated gaps, 96% were shorter than two hours, 3% between 198 two and six hours and 1% longer than six hours. In 26 cases, gaps exceeded the duration of one day. The interpolation was 199 justified as the differences between the values before and after the gaps were within the measuring accuracy of 1 vol-% of the 200 TDR-soil moisture sensors (Acclima Inc., 2019). As indicated by the retailer, sensors might suddenly jump to a soil moisture 201 value of 28.6% and go back to normal again after one or few time steps. Thus, a data deletion procedure of abrupt jumps to 202 28.6 was created. To ensure equal weighting for the subsequent analysis, all soil moisture time series were z-transformed to 203 unit variance and zero mean each (cf. Hohenbrink and Lischeid, 2015). As a consequence, differences of absolute values were 204 not considered by the further analysis.

#### 2.4 Statistical analysis

206 To identify common temporal patterns among single time series, the soil moisture data set was analysed by a principal 207 component analysis (PCA). In a first step, PCA decomposes the total variance of a multivariate data set into independent 208 fractions called principal components (PCs). The number of PCs is the same as the number of time series in the input data set. 209 Each PC consists of eigenvectors (loadings), scores, and eigenvalues. The scores reflect the temporal dynamics. The 210 importance of single principal components for single sites is represented by the loadings of each PC (Jolliffe, 2002; Lehr and 211 Lischeid, 2020). Loadings are the Pearson correlation coefficients of the single time series of the input data set with the scores 212 of each PC, respectively. The eigenvalues of the single PC are proportional to the variance that they explain. The PCs are 213 sorted in descending order of eigenvalues. Eigenvalues greater than one indicate that a PC explains more variance than -a 214 single input time series could contribute to the total variance of the entire input data set (Kaiser, 1960). More details on principal 215 component analysis for time series analysis are found in Joliffe (2002). The PCA was performed using the prcomp function in 216 R version 4.1.0 (R Development Core Team, 2021).

The scores of the principal components constitute time series. Every observed <u>soil moisture z-transformed</u> time series can be presented at arbitrary precision as a combination of various principal components. When the data set consists of time series of the same observable measured at different locations, the first principal component describes the mean behaviour inherent in the data set. Subsequent principal components reflect typical modifications of that mean behaviour at single locations due to

- different effects. Thus, generating synthetic time series as linear combinations of the first PC and another additional PC helps to assign this additional PC to a specific effect. To that end, scores of that component have either been added to or subtracted
- 223 from those of the first component using arbitrarily selected factors. The two resulting graphs show how the respective PC
- 224 causes deviations from the mean behaviour of the data set.
- 225 The relations to soil and vegetation parameters were tested by computing the Pearson correlation coefficients between the
- scores and arithmetic mean values of all input time series as well as the Pearson correlation coefficients between loadings and
- 227 sand content until 0.25 m depth, sensor depth, antecedent z-transformed water contents oil moisture, slope, and drone imagery
- 228 products- (NDVI and surface temperature). Eventually, the Wilcoxon-Mann-Whitney test was applied to check whether
- 229 loadings can be grouped by management parameters (crops, cover crops, weeding management). All statistical analyses were
- 230 conducted with R version 4.1.0 (R Development Core Team, 2021).
- 231 **3 Results**

- 232 3.1 Manual soil texture analysis
- 233 The transferability of texture information from the sampling point to the soil moisture sensor location was not ensured due to
- high nugget effects. Furthermore, manual soil texture analysis data were not available for all analysed patches. Consequently,
- 235 they were not included into further analysis.
- 236 **3.2 Principal component analysis**
- 237 The principal component analysis yielded five components with Eigenvalues exceeding one, which accounted for >97% of the
- 238 total variance of the data set (Table 3).
- 239 **3.2.1 First principal component**
- 240 The first principal component explained 72.3% spatiotemporal variance of the data set. All loadings on the first PC were
- 241 negative (Appendix A). The Pearson correlation coefficient of the scores of the first principal component with the mean values
- of all input time series was less than 0.999 (p < 0.01), the correlation between the scores and the cumulative climatic water
- balance  $(P ET_p)$  was -0.969 (p < 0.01). Thus, the time series of the negative scores of this component represented the mean
- 244 behaviour of soil moisture driven by external factors such as precipitation, temperature, and seasons in general which affected
- 245 time series in the same way, although to different degrees (cf., Hohenbrink et al., 2016; Lischeid et al., 2021).
  - 3.2.2 Second principal component
- 247 The second principal component explained 17.0% of the total variance. The loadings ranged from -0.801 to 0.760 with a
- 248 median of -0.030 (Figure 4). The loadings showed a crop type-group specific pattern. All winter crops (barley, oats, rye) had
- 249 positive loadings with only one exception in 0.9 m depth. The summer crops maize, soy, and sunflower exhibited negative

loadings. In contrast, the summer crop lupine exhibited mostly positive loadings, similar to the winter crops, although of slightly smaller magnitude. According to the Wilcoxon-Mann test, the group of barley, oats, rye, and lupine differed significantly from the group of maize, soy, and sunflower.

As described in the Methods section, synthetic time series were generated as a linear combination of PC1 and PC2 (Figure 5).
The graph resulting from applying a positive factor for PC2 represents a typical deviation from mean behaviour for sites that exhibit positive loadings, e.g., winter crops (blue line). The opposite holds for the summer crops which load negatively with PC2 (orange line). Both lines plot very close to each other in February and March. In contrast, the orange line shows lower values than the blue line in December and January, indicating lower soil moisture at the summer crop patches. The inverse holds for the subsequent summer period starting in early June, pointing to earlier and more rapid water uptake of the winter crops. In July and August, the approximately constant level of the blue curve indicates that only summer crops continue to

260 consume water while winter crops are in their ripening phase and eventually harvested.

Lupine and sunflower were the summer crops which were sown first (March 30, 2021, and April 2, 2021, respectively). Maize was sown on April 16, 2021, and soy on May 15, 2021. The loadings of lupine, which were rather performing like winter crops than summer crops, indicated that lupine showed an early onset of intensive evapotranspiration, compared to other summer crops, especially sunflower which was sown at the same time.

For further investigation of the vegetation effect on PCs, the loadings of PC2 were compared to drone imagery taken at the end of May, when sowing has been completed jen all patches, and imagerys taken at the beginning of July, when during winter crops  $\frac{1}{2}$  are in the ripening phase, was analysed. The second PC's loadings of the time series from different sensors were compared to the Normalized Difference Vegetation Index (NDVI; available for three dates) and surface temperature (only available for May 31, 2021) of the respective sensor location as a proxy for actual evapotranspiration. At the end of May, the NDVI, as a proxy for photosynthesis potential, was positively correlated with the loadings (Table 4). Surface temperature exhibited a negative correlation. The spatial pattern of surface temperature is assumed to be inversely related to that of actual evapotranspiration. Thus, both proxies, NDVI and surface temperature, support the inference that in this study positive loadings on this principal component represent sites with above-average plant activity and root water uptake at the end of May. This holds for sensors from all depths but was the closest for 0.9 m depth (Pearson correlation of r = -0.916 for surface temperature and of r = 0.946 for NDVI on May 31). The results in July compared to those in May support the observation. At the time when the winter crops are already in the ripening phase and the summer crops reach high levels of evapotranspiration, the correlations are being reversed and negative loadings indicate above-average plant activity for summer crops. On July 06, highest Pearson correlations for NDVI are found for 0.6 m depth (r = -0.917).

### 3.2.3 Third principal component

The third PC explained 4.1% of the total data set's variance. Loadings ranged between -0.787 and 0.244 with a median of 0.006. Extreme loadings (<-0.25) were found only for sensors in 0.9 m depth in patches 66, 89, 95 and 102 (Figure 6). The location of these patches shows a certain regional pattern, with the patches roughly followings an east-west direction rather

than showing a random location within the field. This may point to topography or soil structure causing deviations from mean soil moisture behaviour for patches located near this gradient., which, Hhowever, this pattern cannot be assigned to topography or structures apparent on the topsoil map (Figure 1). Loadings were closely related to the minima of the z-transformed soil moisture in the period from December to February (r = 0.70, p < 0.001, Figure 7). The most obvious difference between What distinguishes the orange line (negative loading on PC3) and from the blue line (positive loading on PC3) during the first half of the study period is that the latter reaches a maximum of soil moisture after rainfall much earlier compared to the former is the higher temporal variability and the delayed reaching of maxima in the first half of the study period (Figure 8).

#### 3.2.4 Fourth principal component

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- The fourth PC explained 2.2% of the total data set's variance. The loadings were clustered by crop groups. All fallow patches showed consistent positive loadings while the patches which were covered by winter crops, showed mainly negative loadings except in patch 95 where the loadings of the two sensors in 0.3 m depth were slightly above zero (Figure 9). According to the Wilcoxon-Mann test treatment group B (fallow, followed by summer crops) differed significantly from group A (winter crops) and C (cover crops, followed by summer crops) whereas there was no significant difference between group A and C. In contrast to crop groups A and B, patches that were covered by the cover crop phacelia during the winter months, did not show one-directional loadings.
- Figure 10 illustrates the effect of the fourth PC on time series. The blue line (positive loading) shows a hydrological behaviour which A positive factor would be typical for more sandy soils and for patches with fallow in autumn and winter (blue line). In contrastwhile the orange line (negative loading) depicts behaviour that one would expect in more loamy soils and for winter crops. The latter line exhibits slightly more due to its delayed responses to rainstorms and subsequent less steep recovery as it would be expected for more loamy soils. The patterns in the loadings thus show a differentiation between patches with winter crops and fallow patches in the winter months (Figure 9). However, it is not clear how winter crops on the one side and fallow on the other side could induce such a different soil water behaviour shown in Figure 10.

#### 3.2.5 Fifth principal component

- The fifth PC explained 1.7% of the data set's variance. The loadings showed a depth-related pattern. All time series from the 0.3 m depth exhibited negative loadings with two minor exceptions. Whereas all time series from 0.9 m depth showed positive loadings throughout, and time series from 0.6 m depth plot in between. Loadings in 0.6 m depth and 0.9 m depth were mostly more similar to each other than to the loadings of 0.3 m depth (Figure 11). The Pearson correlation coefficient between loadings and depth was r = 0.710 (p < 0.05). Thus it can be concluded that the fifth PC reflected the effect of soil depth on soil moisture variance. This effect differed between crops, with the three most negative loadings found in maize patches while the three most positive loadings were found in lupine patches.
- The hydrological signal after rainfall events exhibits damping over depth (blue line) while sensors in the upper layer react with a higher sensitivity (orange line) to weather conditions (Figure 10).

- 315 The soil water dynamics show a damping effect with increasing depth (Figure 12) from little damping for sensors in the upper
- depth (orange line) to higher damping for sensors in greater depth (blue line).
- 317 Neither patterns in topography nor in weeding management modes were reflected in the loadings of PC1-PC5. Due to the lack
- 318 of subsurface soil data, no additional findings could be derived from the Geophilus texture analysis.

#### 4 Discussion

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- 320 A PCA was conducted to identify the drivers of soil moisture variability in a diversified cropping field. Data consisted of
- 321 observed time series from 64 soil moisture probes. Results showed that tThe first five principal components described about
- 322 97% of the variance of the data set, which consisted of observed time series from 64 soil moisture probes and revealed various
- effects of weather, soil texture, soil depth, crops and management schemes (Table 3). The first principal component captured
- 324 72% of the total variance. Consequently, 72% of the observed dynamics could be described by a lumped model that would not
- 325 consider any within-field heterogeneity. These is figure results are is in the range of similar studies. In the study of Martini et al.
- 326 (2017) who found that the first PC explained 58% of the variance of a data set that comprised both agricultural fields as well
- 327 as grassland transects. Similarly, Lischeid et al. (2017) ascribed 70% of the variance of a forest soil hydrological moisture data
- 328 set to a single component. In the study by Hohenbrink et al. (2016), 85% of the variance of soil hydrological moisture data in
- 329 a set of anable field experiments with two different crop rotation schemes was attributed to the first principal component. The
- 330 strong influence of weather conditions as it is shown in our study is confirmed by Choi et al. (2007) who showed that rainfall,
- next to topography, explained most of the surface soil moisture variability.

#### 4.1 Crop effects

- 333 As Korres et al. (2015) stated that vegetation and management (e.g. planting and harvesting dates) are among the main causes
- 334 for spatial variability of soil moisture in agricultural fields besides soil parameters are vegetation and management (e.g.
- 335 planting and harvesting dates). The quantification of the impact of these effects on soil moisture variability is highly important.

for instance for hydrological applications and adopted management practices in agriculture (Hupet and Vanclooster, 2002).

- 337 Joshi and Mohanty (2010) investigated the spatial soil moisture variability at on the field to regional scale in the Southern
- 338 Great Plains regions in the US by means of PCA and assessed the effect of vegetation as limited since none of the first seven
- 339 PC showed strong correlations with vegetation parameters. In Western China, Wang et al. (2019) used a non-linear Granger
- 340 causality framework and quantified the vegetation effect on soil moisture variability with up to 8.2%. In this study, conducted
- 341 at the field scale, around 17% of the total variance at the field scale was attributed to the vegetation effect. When not considering
- the temporal component reflected by PC1 and thus only looking at the spatial variability, 61% of the remaining variance
- 343 (attributed to PC2 to PC64) is caused by the vegetation effect reflected by PC2. Korres et al. (2010) also used PCA to identify
- the drivers of spatial variability of soil moisture within a cropped area but did not find such a pronounced vegetation effect. In
- their study, more than two thirds of the spatial variability was related to soil parameters and topography. In contrast, the strong

influence of vegetation in our study may be due to the high level of crop diversification. Within single crop fields, vegetation effects are observable due to heterogeneous biomass or root development (Brown et al., 2021; Korres et al., 2010), but may be of a lower magnitude compared to fragmented field arrangements with different crops. The high impact of crop diversification on soil moisture variability is also visible when comparing our results to the results of a field under comparable conditions in the same region with only two crop rotations in which only 3.8% was explained by the different crop rotations (Hohenbrink et al., 2016). Joshi and Mohanty (2010) investigated spatial soil moisture variability at the field to regional scale in the Southern Great Plains regions in the US by means of PCA and assessed the effect of vegetation - in contrast to this study - as limited since none of the first seven PC showed strong correlations with vegetation parameters.

It needs to be considered that the proportion of the vegetation effect on soil moisture variability does not only vary spatially and over depth, but also over time. Under dry conditions, soil-plant interactions prevail while under moist conditions, percolation behaviour is predominant (Baroni et al., 2013). The scores are time series and reflect the effect size of a particular process represented by the respective PC. The more the scores of a certain PC deviate from zero during specific periods, the stronger the respective effect is. Consequently, the time series of PC2 scores indicates that the effect of vegetation on total variability varies by time. In accordance with literature, the absolute values of the scores of PC2, representing differences between the contrasting seasonality of crops, are highest in the dry months, May to August. This is mostly explained by the high water demand of summer crops, which are in their vegetative growth stage from May to August, whereas winter crops are already in their reproductive growth stage, including maturity, senescence and harvest where water uptake by crops is minimal or absent (Zhao et al, 2018). In the moist winter months January to March, as well as during the heavy rainfall event in July, the scores of PC2 are relatively small, showing that spatial variability at that time is caused by other factors.

The second principal component clearly differentiated between winter and summer crops, which was driven by the different seasonal patterns of root water uptake (Figure 4). In contrast, the fourth component differentiated between fallow followed by summer crops and winter crops, whereas phacelia followed by summer crop did not show a clear pattern-separated winter crops and fallow (Figure 9). Phacelia is grown as a cover crop and usually dies off in frost periods. Due to rather mild winter temperatures 2020/21, Phacelia was not terminated efficiently and kept growing until spring, until it was terminated mechanically. It was recently shown that the timing of removal of winter cover crops is key to provide soil water recharge for the subsequent crop, as the depletion of soil water in autumn is significant (Selzer and Schubert, 2023). Thus, some Phacelia patches exhibited negative loadings, similarly to the winter crop patches while other patches with most likely different termination dates exhibited positive loadings.

Hence, the fourth component obviously reflected the effect of the active root systemplant cover in the winter period, which can hardly be ascribed to different patterns of root water uptake. According to this component, soil moisture water dynamics at in the fallow patches mostly resembled more the typical behaviour one would expected for sandy soils, and that of winter crop patches showed a more damped behaviour that is usually observed in more loamy soils. Note that the term "fallow" refers to crop cover in autumn and winter only. Acharya et al. (2019) found that winter cover crops improved soil moisture from 3

381 for winter crops (orange line) in winter. However, it has also been observed that roots from winter crops can increase soil 382 porosity and therefore, water mobility in the soil (Lange et al., 2013; Scholl et al., 2014). 383 That feature pattern could point to a soil carbon effect on the soil's water holding capacity: Only in at the winter crop sites 384 patches, soil organic carbon (SOC) in soil increased continuously due to root growth and root exudation, whereas 385 mineralisation reduced SOC the organic carbon stock at in the fallow sitespatches. Effects of dense living root networks on 386 soil hydraulic conductivity have been reported, e.g., Scholl et al. (2014), Zhang et al. (2021) and Lange et al. (2013). Further 387 soil-vegetation interactions might play a role for the delayed seepage fluxes of winter crop and part of cover crop patches, such 388 as soil organic matter from cover crops and plant residues (Manns et al., 2014; Rossini et al., 2021). Usually, such effects are 389 assumed to occur only at larger time scales, which is closely related to problems of detecting changes soil organic carbon 390 (SOC) quantity or quality. So far, there is only anecdotal evidence for rather short-term soil organic carbon SOC quality 391 affecting soil hydraulic properties even at smaller time scales. Although this effect constituted only a minor share of soil 392 moisture variance (Table 3), it was clearly discernible as a separate principal component. This effect would be worth to be 393 tested in more detailed future studies. If it were to be confirmed, it would be a good example for how crop management shapes 394 soil properties.

to 5% in the top 0.3 m soil layer which is in line with the findings from Figure 10 that shows a higher water holding capacity

#### 4.2 Soil texture and soil depth effects

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- Texture is another highly important spatial variable that affects soil moisture. The pore size distribution, which is directly linked to texture has great influence on wetting processes as well as on the water retention capacity of soil (Krauss et al., 2010; Rossini et al., 2021). Furthermore, texture influences the evapotranspiration which is another main factor controlling soil moisture (Pan and Peters Lidard, 2008). For coarse grained soils as they are present in this case study, the water retention capacity is small, resulting in enhanced seepage fluxes (Scheffer and Schachtschabel, 2002; Krauss et al. 2010).
- 401 Loadings on the third principal component were not related to crop types. In contrast, a spatial pattern emerged: Only sensors 402 from 0.9 m depth from six adjacent patches exhibited strongly negative loadings (Figure 6), whereas all other sensors showed 403 minor positive or negative loadings. This points to an effect of subsoil substrates, that is, higher clay content and consequently 404 higher water holding capacity. That would be consistent with delayed response to seepage fluxes and reduced desiccation in 405 the vegetation period (Figure 8). The strong relation between z-transformed soil moisture minima at the beginning of the study 406 period (Figure 7) which might originate from a delayed response to a prior rainfall, at the beginning of the study period, and 407 the regional pattern of the location of the patches following a west-east direction within the experiment might be an indicator 408 of underlying soil structures causing this effect. Data on the texture at soil moisture the sensor locations in deeper layers would 409 be of high value to confirm the assumptions.
- Whereas the third principal component seems to reflect a local peculiarity, the fifth component obviously grasps a more generic feature. Loadings on this component are clearly related with depth (Figure 11). Strong positive loadings indicate a strongly damped behaviour of soil moisture time series: The blue line, representing sites with positive loadings on PC5 which is typical

for sensors at greater depth (Figure 12), exhibits clearly reduced amplitudes compared to the orange line, that is, sensors at shallow depth. Hohenbrink and Lischeid (2015) combined a hydrological model and principal component analysis to study the effect of soil depth and soil texture on damping of the input signal in more detail. A subsequent field study proved the relevance of that effect in a real-world setting (Hohenbrink et al., 2016). Moreover, Thomas et al. (2012) found that damping accounted for a large share of variance in a set of hydrographs from a region of 30,000 km<sup>2</sup>. Damping was also the most relevant driver of spatial variance in a set of time series of groundwater head at about the same scale (Lischeid et al., 2021).

#### 4.3 Limitations

Data gaps during the studied period occurred due to multiple technical and environmental factors. Data gaps in soil moisture time series were caused by repeated temporary failure of the WSN. There was a failure of one sensor that was replaced and one LoRa node was damaged by intruding water. More relevant, however, were failures of data transmission. Yildiz et al. (2015) point to the problem of optimizing transmission power for data and acknowledgement packets depending on energy dissipation under the given conditions. E.g., saturated soil conditions and dense biomass stands reduce the transmission signal from the node to the gateway (Bogena et al., 2009). The installation of a second gateway in September 2021 increased higher transmission coverage in the field. Another obstacle was snow cover on the gateways' solar panels. Finally, solar panels were subject to theft. However, higher level of maintenance and supervision helped to reduce the number and the length of data gaps.

428 <u>gaps</u>

PCA requires gapless time series. Gaps in single time series need to be either filled at the risk of introducing artefacts or the respective time period cannot be considered at all for analysis. This can be seen as a weakness of PCA. On the other hand, and in contrast to other time series analysis approaches, the time series need not to be equidistant. Assigning PCs to processes and effects is not straightforward and might be subject for debate. For example, in this study soil samples were taken at least at 0.8 m distance from the sensors to avoid disturbance of the measurements. Due to pronounced small-scale soil variability these samples are not fully representative for the measurement sites. In spite of these limitations, the PCA results clearly point to various effects worth to be studied in more detail in subsequent studies.

#### **5 Conclusion**

To disentangle and to quantify different effects of environmental processes in complex settings is a key challenge of agricultural and environmental research. It is an indispensable prerequisite for tailored field and crop management. Mechanistic models are a way to upscale findings from numerous single cause single effect studies. But there is urgent need to further validate model results and to study interactions between various effects in a systematic way. Principal component analysis is a step further to meet these challenges although not entirely without problems. In this study which focuses on the interplay between crops and soil heterogeneities in terms of soil moisture dynamics, the strength of the methodology in contributing to

444 disentangling different effects of complex spatially and temporally diversified cropping systems based on a comprehensive 445 real world data set is presented. The use of PCA has a high value for the application in environmental sciences, as it contributes 446 to process understanding of soil water dynamics by disentanglinge the different effects of complex spatially and temporally 447 diversified cropping systems. In this study, Momore than 97% of the observed spatial and temporal variance was assigned to 448 five different effects. Meteorological drivers explained 72.3% of the total variance (PC1). Different seasonal patterns of root 449 water uptake of winter crops compared to summer crops accounted for another 17.0% of variance (PC2). An additional share 450 of 2.2% of variance seemed to be related to the effects of a living rooting system different vegetation cover on and its interplay 451 with soil hydraulic properties (PC4). Heterogeneity of subsoil substrates explained 4.1 % of variance (PC3), and the damping 452 effect of input signals in the soilover depth another 1.7% (PC5). To summarize, plant-related direct and indirect effects 453 accounted for 19.2% of the variance (PC2 and PC4), and soil-related effects only for 5.8% (PC3 and PC5). In particular, the 454 plant-induced effects on soil hydraulic properties would be worthwhile to be studied in more detail. Findings of this study highly depend on local conditions. However, the methodology itself is generally applicable to other site

Findings of this study highly depend on local conditions. However, the methodology itself is generally applicable to other site conditions and can lead to improved management practices through improved knowledge about soil water dynamics.

Furthermore, information from this study can also help to develop both parsimonious and tailored mechanistic models for model upscaling. In this regard, principal component analysis of large soil moisture data sets from real-world monitoring setups performed a meaningful diagnostic tool for complex cropping systems.

Knowledge from data driven approaches can support adequate crop selection as a management option to encounter the increasing drought risk in the study region. It has been shown that principal component analysis has a high value for the application in environmental sciences, as it allows to draw conclusions about variabilities in large data sets from real world monitoring setups despite gaps in time series. Information from this study will contribute to elucidate management effects as well as to develop both parsimonious and tailored mechanistic models. Findings of this study highly depend on local conditions. However, we consider the presented approach generally applicable to a large range of site conditions. In this regard, principal component analysis of soil moisture time series performed as a powerful diagnostic tool and is highly recommended.

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#### 474 Competing interests

475 The authors declare that they have no conflict of interest.

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- 653 Figures and Tables

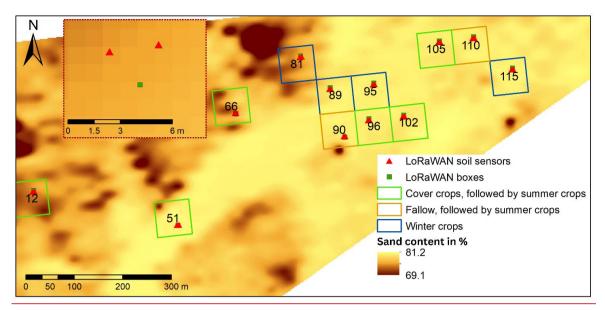
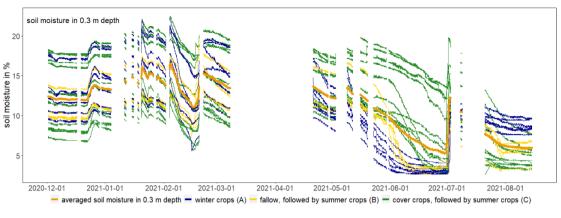
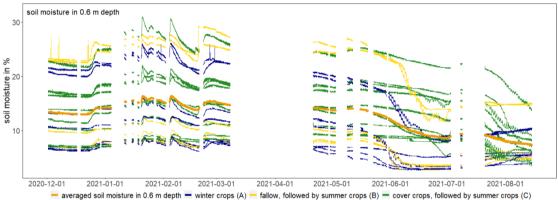


Figure 1: Sand content (in %) in the top 0.25 m soil depth, and location of the of the analysed patches, including soil sensors (triangle) and boxes (square) under different crop rotations atim the patchCROP landscape laboratory, patchCROP, Tempelberg, Brandenburg and boxes (square). The inset shows sensor and box location within one of the patches.





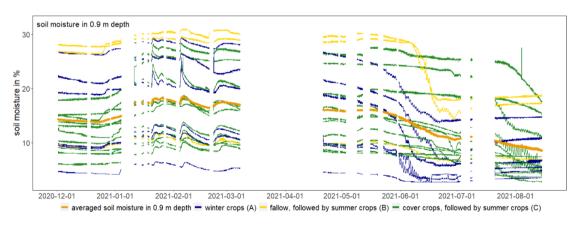


Figure 2: Input soil moisture time series per depth, differentiated between crop groups, and average soil moisture of all time series per depth from 2020-12-01 until 2021-08-15 at the patchCROP landscape laboratory, Tempelberg, Brandenburg, Germany.

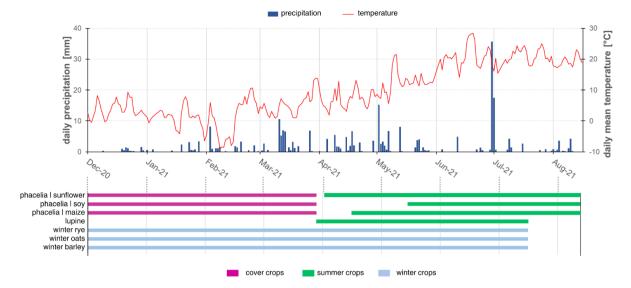
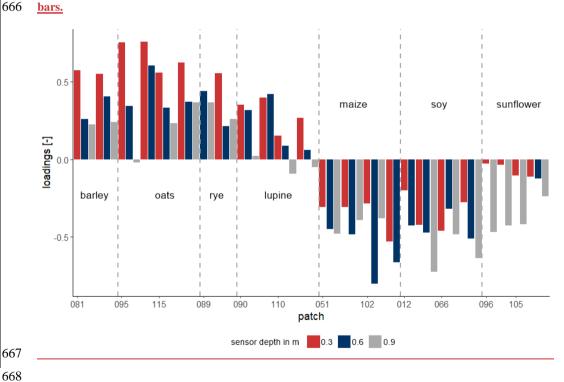


Figure 3: Measured daily precipitation, mean temperature and cultivated crops - differentiated between winter crops (light blue bars), summer crops (green bars) and cover crops (pink bars) - from 2020-12-01 until 2021-08-15 at the patchCROP landscape laboratory, Tempelberg, Brandenburg, Germany. Specific crops for the studied timeframe stated at the left side of the horizontal bars.



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Figure 4: Time series Loadings of time series on the second principal component at the patchCROP landscape laboratory, Tempelberg, Brandenburg, Germany, showing a crop group related pattern. Bars represent individual time series grouped by patch ID and sorted by crop.

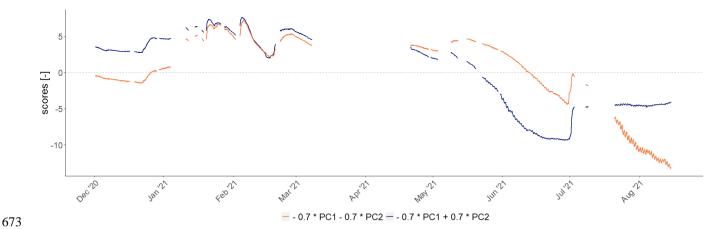


Figure 5: Effect of the second principal component on modification of the general mean behaviour presented by the first principal component at the patchCROP landscape laboratory, Tempelberg. The blue line represents deviations from mean soil moisture for time series with positive loadings on PC2 (winter crops) while the orange line represents deviations from mean soil moisture for time series with negative loadings on PC2 (summer crops).

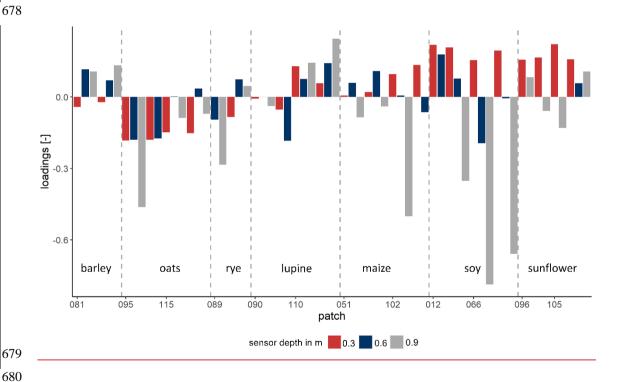


Figure 6: Loadings of time series on the third principal component at the patchCROP landscape laboratory, Tempelberg, Brandenburg, Germany with some of the sensors in deeper layers showing noticeably negative loadings. Bars represent individual time series grouped by patch ID and sorted by crop.

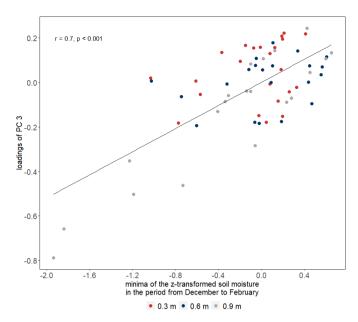


Figure 7: Relation between minima of the z-transformed soil moisture in the first months of the study period with loadings of third principal component showing that sensors with noticeably negative loadings showed distinctly negative z-transformed minima.

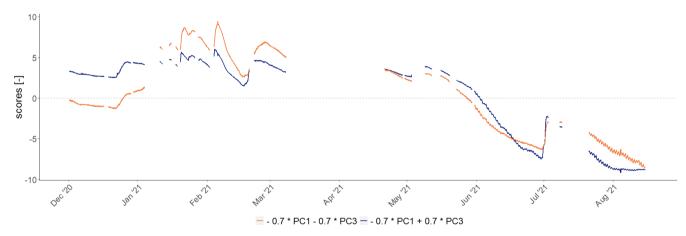


Figure 8: Effect of the third principal component on modification of the general mean behaviour presented by the first principal component at the patchCROP landscape laboratory, Tempelberg. The blue line represents deviations from mean soil moisture for time series with positive loadings on PC3 (majority of the time series) while the orange line represents deviations from mean soil moisture for time series with negative loadings on PC3 (part of the sensors in 0.9 m depth).



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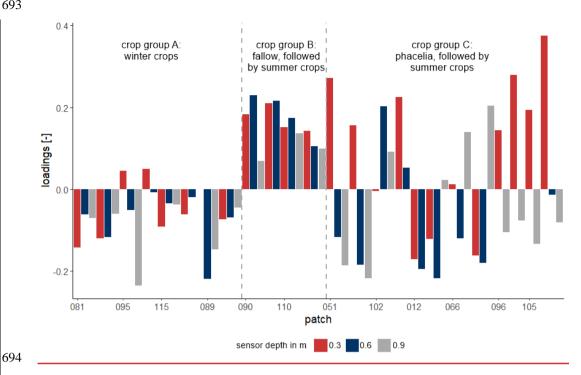


Figure 9: Loadings of time series on the fourth principal component at the patchCROP landscape laboratory, Tempelberg, Brandenburg, Germany showing mainly negative loadings for crop group A, positive loadings for crop group B and loadings with no clear pattern for crop group C. Bars represent individual time series grouped by patch ID, sorted by treatment group.

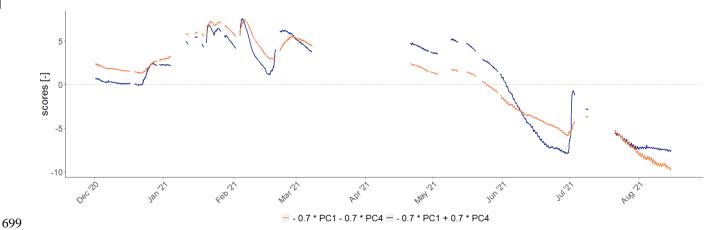


Figure 10: Effect of the fourth principal component on modification of the general mean behaviour presented by the first principal component at the patchCROP landscape laboratory, Tempelberg. The blue line represents deviations from mean soil moisture for time series with positive loadings on PC4 (single sensors of crop group A, all sensors of crop group B, and part of crop group C) while the orange line represents deviations from mean soil moisture for time series with negative loadings on PC4 (most sensors of crop group A and part of the sensors of crop group C).



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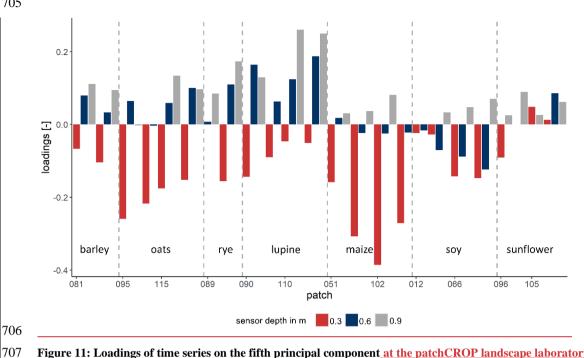


Figure 11: Loadings of time series on the fifth principal component at the patchCROP landscape laboratory showing a depth related pattern. Bars represent individual time series grouped by patch ID, sorted by crop.

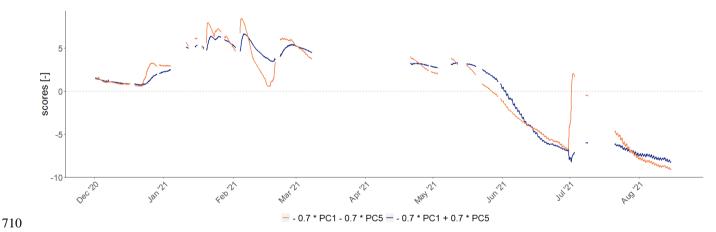


Figure 12: Effect of the fifth principal component on modification of the general mean behaviour presented by the first principal component at the patchCROP landscape laboratory, Tempelberg. The blue line represents deviations from mean soil moisture for time series with positive loadings on PC5 (sensors in greater depth) while the orange line represents deviations from mean soil moisture for time series with negative loadings on PC5 (sensors in shallow depth).

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Crop in winter season	Crop in summer season	Crop group	Sand content (in 1 m buffer zone around sensors) in [%]	Weed control	Patch ID
Winter barley		<u>A</u>	78.3	<u>conventional</u>	81
Win	ter oats	A	80.7	conventional	95
Win	ter oats	A	80.6	reduced	115
Win	iter rye	A	80.5	conventional	89
Fallow	Lupine	В	80.6	conventional	90
Fallow	Lupine	В	80.3	reduced	110
Phacelia	Maize	С	80.8	reduced	51
Phacelia	Maize	С	80.6	conventional	102
Phacelia	Soy	С	78.5	reduced	12
Phacelia	Soy	С	77.9	conventional	66
Phacelia	Sunflower	С	80.6	conventional	96
Phacelia	Sunflower	С	80.5	reduced	105

Table 2: Overview of normalized difference vegetation index (NDVI), surface temperature, and slope at the locations of analysed sensors -at the patchCROP experiment in Tempelberg, Brandenburg, Germany.

Crop	Patch ID	Sensor	NDVI	NDVI	NDVI	Surface	Slope_ in
		Position	2021-05-20	2021-05-31	2021-07-06	Temperature	₽° <del>}</del>
			[-]	[-]	[-]	2021-05-31 in [°C]	
Winter Bbarley	81	West	0.874	0.182	0.926	20.57	2.01
Winter bBarley	81	East	0.875	0.180	0.927	20.43	1.94
Winter o Oats	95	East	0.838	0.208	0.834	27.25	1.36
Winter Ooats	95	West	0.838	0.213	0.840	27.85	1.15
Winter o Oats	115	West	0.756	0.278	0.845	23.70	1.28
Winter Ooats	115	East	0.783	0.281	0.863	25.12	0.43
Winter Rrye	89	West	0.796	0.263	0.856	22.39	1.74
Winter rRye	89	East	0.787	0.206	0.822	24.95	1.67
Lupine	90	West	0.185	0.395	0.710	26.31	1.40
Lupine	90	East	0.203	0.391	0.733	24.96	1.27
Lupine	110	West	0.090	0.563	0.635	26.98	1.88
Lupine	110	East	0.090	0.567	0.639	26.76	2.50

Maize	51	West	-0.099	0.654	0.181	35.44	0.82
Maize	51	East	-0.096	0.638	0.217	35.29	0.93
Maize	102	West	-0.077	0.714	0.175	37.88	0.88
Maize	102	East	-0.058	0.728	0.178	38.03	0.90
Soy	12	West	-0.107	0.748	0.166	34.87	1.71
Soy	12	East	-0.108	0.723	0.162	34.44	1.11
Soy	66	West	-0.115	0.730	0.144	35.09	2.40
Soy	66	East	-0.114	0.661	0.147	34.39	2.13
Sunflower	96	West	-0.109	0.816	0.211	33.76	0.59
Sunflower	96	East	-0.101	0.827	0.229	34.70	0.69
Sunflower	105	West	0.178	0.610	0.564	29.79	1.04
Sunflower	105	East	0.030	0.696	0.399	34.53	1.00

Table 3: <u>Statistical characteristics and interpretations of principal components 1 to 5 for soil moisture dynamics of selected patches</u> at the patchCROP landscape laboratory, Tempelberg, Brandenburg, Germany.

	PC1	PC2	PC3	PC4	PC5
Eigenvalue	46.25	10.89	2.60	1.43	1.06
Proportion of variance in %	72.27	17.01	4.06	2.23	1.65
Proportion of variance (cumulative) in %	72.27	89.28	93.34	95.57	97.22
Interpretation	Mean behaviour	Winter vs. summer crops	Subsoil texture	Soil organic earbonwinter vegetation cover and influence of cover crops on soil	Damping of the input signal
Prevailing driver	weather	crop	soil	crop and soil	soil

Table 4: Pearson correlation coefficients between drone imagery products surface temperature and normalized difference vegetation index (NDVI)-at the patchCROP landscape laboratory, Tempelberg, Brandenburg, Germany taken on May  $31^{st}$ , 2021, and loadings of sensors in all depths or at single depths, respectively, on the second principal component. All correlations \_-arewere highly significant (p <0.01).

<u>Variable</u>	Sensors in all depths	0.3 m	0.6 m	0.9 m
Surface temperature	-0.853	-0.881	-0.909	-0.916

NDVI 2021-05-20	0.836	0.904	0.837	0.907
NDVI 2021-05-31	0.899	0.945	0.944	0.946
NDVI 2021-07-06	-0.860	-0.898	-0.917	-0.913

## 732 APPENDIX A

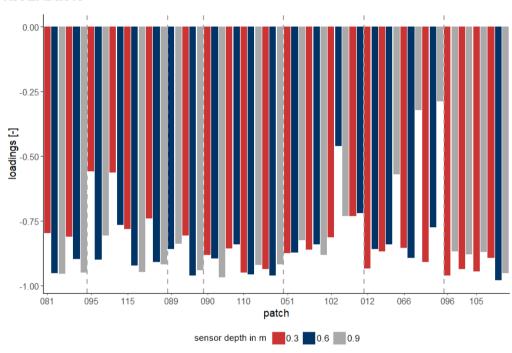


Figure 13: Loadings of time series on the first principal component at the patchCROP landscape laboratory, Tempelberg, Brandenburg, Germany. Bars represent individual time series grouped by patch ID and sorted by crop.