# Assessing acetone for the GISS ModelE2.1 Earth system model

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8 Abstract. Acetone is an abundant volatile organic compound (VOC) in the atmosphere with important influence on ozone and

9 oxidation capacity. Direct sources include chemical production from other VOCs as well as anthropogenic, emissions, terrestrial 10 vegetation, biomass burning emissions, and ocean production. Sinks include chemical loss, deposition onto the land surface, and 11 ocean uptake. Acetone also has a lifetime, that is long enough to allow transport and reactions with other compounds remote from 12 its sources. The NASA Goddard Institute for Space Studies (GISS) Earth System Model, ModelE2.1, simulates a variety of Earth 13 system interactions. Previously, acetone had a very simplistic representation in the ModelE chemical scheme. This study assesses 14 a more sophisticated acetone scheme in which acetone is a full 3-dimensional tracer, with explicit sources, sinks, and atmospheric 15 transport. We first evaluate the new global acetone budget in the context of past literature. Estimated source and sink fluxes fall 16 within the range of previous models, although total atmospheric burden and lifetime fall on the lower end of published literature. 17 Acetone's new representation in the ModelE2.1 also results in more realistic spatial and vertical distributions, which we compare 18 against previous models and field observations. The seasonality of acetone-related processes was also studied in conjunction with 19 field measurements, and these comparisons show promising agreement but have shortcomings at high-emission urban locations, 20 where the model's resolution is too coarse to capture the behavior, Finally, we conduct a variety of sensitivity studies that explore 21 the influence of key parameters on the acetone budget and its global distribution. An impactful finding is that the production of 22 acetone from precursor hydrocarbon oxidation has strong leverage on the overall chemical source, indicating the importance of 23 accurate molar yields. Overall, our implementation is one that corroborates with previous studies and marks a significant

24 improvement to the development of the acetone tracer in the GISS ModelE2.1,

### 25 1 Introduction

Acetone (C3H6O) is an abundant oxygenated volatile organic compound (VOC) that has important connections to ozone and the atmosphere's self-cleansing oxidation capacity (Read et al., 2012). Acetone's dynamic presence in Earth's atmosphere can be described through sources, sinks, and mechanisms of transport. Extensive literature has discussed the nature of these sources and sinks, and some are more well-constrained than others.

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Primary sources of acetone in the atmosphere include anthropogenic, terrestrial vegetation, and biomass burning emissions. Past literature has found the fluxes of these sources to range between 1-2 Tg yr<sup>-1</sup>, 30-45 Tg yr<sup>-1</sup>, 2.5-4.5 Tg yr<sup>-1</sup>, respectively (Beale et al., 2013; Brewer et al., 2017; Elias et al., 2011; Fischer et al., 2012; Folberth et al., 2006; Jacob et al., 2002; Singh et al., 2000; Wang et al., 2020). Chemical production from other VOCs with 3 or more carbon atoms, each with their own molar yields, is another source of acetone in the atmosphere (Brewer et al., 2017; Fischbeck et al., 2017; Hu et al., 2013; Jacob et al., 2002; Singh et al., 2000; Weimer et al., 2017).

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79 Sinks of acetone include wet and dry deposition, as well as chemical loss. Wet deposition occurs within and below clouds due to 80 the solubility of acetone, and depends on its Henry's Law coefficient (Benkelberg et al., 1995). Dry deposition occurs on the land 81 surface. Chemical loss of acetone forms radicals through photolysis. Past literature has estimated the acetone sinks to be 10-30% 82 dry deposition and 40-85% chemical loss (Arnold et al., 2005; Elias et al., 2011; Fischer et al., 2012; Khan et al., 2015; Singh et 83 al., 1994). The estimated fluxes are 10-16 Tg yr<sup>1</sup> and 45-60 Tg yr<sup>1</sup> for total deposition and chemical loss, respectively (Arnold et 84 al., 2005; Brewer et al., 2017; Dufour et al., 2016; Elias et al., 2011; Fischer et al., 2012; Jacob et al., 2002; Khan et al., 2015; 85 Marandino et al., 2005; Singh et al., 2000; Wang et al., 2020). 86 87 The ocean surface is a bidirectional flux that provides both a source and a sink for acetone. Ocean surface conditions such as wind 88 speed, sea surface temperature, and seawater concentration of acetone can influence the direction and magnitude of ocean-acetone 89 exchange (Wang et al., 2020). Previous literature estimates an oceanic source flux of 25-50 Tg yr<sup>1</sup> and oceanic uptake flux of 90 35-60 Tg yr<sup>-1</sup>. However, there is little consensus in the literature on whether the ocean serves as a net source or sink of acetone, 91 with some studies indicating a net oceanic source (Beale et al., 2013; Jacob et al., 2002; Wang et al., 2020), and other studies 92 indicating a net oceanic sink (Brewer et al., 2017; Elias et al., 2011; Fischer et al., 2012; Wang et al., 2020). 93 94 In addition to a global annual mean atmospheric budget, previous studies have reported the seasonality of acetone-related processes. 95 Past studies have compared monthly estimates of acetone mixing ratios to field measurements of European sites from Solberg et 96 al. (1996) (Arnold et al., 2005; Elias et al., 2011; Jacob et al., 2002). Comparisons with these European sites have emphasized the 97 seasonal variability of acetone emissions, as nearly all sites portray a summer maximum and winter minimum of acetone 98 abundance. Vegetation emissions from June to September, along with chemical sources, have an especially strong contribution to 99 this seasonality. The winter minimum of acetone is aided by an ocean sink at coastal sites (Jacob et al., 2002). 00 .01 Other studies have described spatial distributions and seasonal dependence of ocean fluxes of acetone (Fischer et al., 2012; Wang .02 et al., 2020). A model by Fischer et al. (2012) proposed a net ocean sink of 2 Tg yr<sup>-1</sup> and characterized ocean uptake of acetone as 03 strongest in northern latitudes year-round and in the high southern latitudes during the winter. An oceanic acetone source was .04 dominant in the tropical regions, with an exception off the western coasts of Central America and Central Africa (Fischer et al., 05 2012). A model by Wang et al. (2020) that varied surface seawater acetone concentration through a machine learning approach .06 also proposed a net ocean sink year-round. This net sink was strongest in December-February, and weakest in March-May. 07 .08 The vertical distribution of acetone has been modeled between the seasons of May-October and November-April in the surface .09 and troposphere (Fischer et al., 2012). Acetone concentrations are generally higher in the lower altitudes due to proximity to surface 10 emissions. Surface-level acetone has been measured over a variety of terrestrial and oceanic sites around the world (de Gouw et al., 2004; Dolgorouky et al., 2012; Galbally et al., 2007; Guérette et al., 2019; Hu et al., 2013; Huang et al., 2020; Langford et al., 11 12 2010; Lewis et al., 2005; Li et al., 2019; Read et al., 2012; Schade and Goldstein, 2006; Singh et al., 2003; Solberg et al., 1996; 13 Warneke and de Gouw, 2001; Yoshino et al., 2012; Yuan et al., 2013), and in some cases, these measurements were taken over a variety of months to provide a sense of seasonality (Dolgorouky et al., 2012; Hu et al., 2013; Read et al., 2012; Schade and 14 15 Goldstein, 2006; Solberg et al., 1996). Additionally, vertical distributions of acetone have been measured through NASA's Atmospheric Tomography Mission (ATom) campaigns (Thompson et al., 2022). The ATom-1, ATom-2, ATom-3, and ATom-4 16 17 campaigns took place during July-August 2016, January-February 2017, September-October 2017, and April-May 2018,

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respectively. Each campaign provided mixing ratios for a variety of VOCs in profiles from the marine boundary layer up to the 22 .23 upper troposphere/lower stratosphere (Apel et al., 2021).

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The NASA Goddard Institute for Space Studies (GISS) ModelE2.1 Earth System Model (Kelley et al., 2020) has the capability of 25 26 simulating a variety of Earth system interactions, is used both to interpret and predict past and future climate, and routinely participates in the Climate Model Intercomparison Projects (CMIP) and Intergovernmental Panel for Climate Change (IPCC) 27 .28 reports. Here we used and enhanced this model by adding acetone as an independent chemical tracer (Kelley et al., 2020). .29 Previously, acetone had a very simplistic representation in the model's chemical scheme (Shindell et al., 2003), in which acetone's .30 spatial variation was parameterized based on the difference of the model's zonal mean distribution of isoprene and that tracer's 31 three-dimensional distribution. Acetone's lifetime is long enough to be transported remote from sources, but not long enough to become uniformly mixed, and therefore its simulated distribution should benefit from a more realistic implementation. We 32 .33 developed a greatly improved acetone tracer scheme by making prognostic calculations of the 3-dimensional distribution of acetone 34 as a function of time. We evaluated its atmospheric burden and lifetime as well as source/sink fluxes (anthropogenic emissions, .35 vegetation emissions, biomass burning, deposition, ocean, and chemistry) against other models and its concentration against field .36 measurements. This work aims to provide a holistic assessment of the abundance of acetone in the atmosphere.

### .37 2 Methodology

.38 Here we implement acetone in the GISS ModelE2.1 based on the literature rather than developing a new parameterization. Our 39 'Baseline' simulation is a climatological mean with year 2000 conditions, chosen to be relatively modern without precluding 40 comparison with models in older literature. The 1996-2004 mean of prescribed emissions from Hoesly et al. (2018) were used, .41 along with the 1996-2005 mean sea surface temperature and sea ice cover as described in Kelley et al. (2020). Acetone simulations 42 use full chemistry and not archived OH fields. An additional simulation, 'Nudged\_ATom', was conducted to compare more directly 43 with ATom field measurements. This simulation employed nudged winds from MERRA2 (Gelaro et al., 2017), ocean surface 44 conditions from PCMDI-AMIP 1.1.4 for 2016-2017 (Taylor et al., 2000) and from Hadley Center HadISST1.1 for 2018 (Met .45 Office, Hadley Centre, 2006), and trace gas and aerosol emissions changing with time during 2016-2018.

### 46 2.1 Sources

### 47 2.1.1 Anthropogenic emissions

48 Anthropogenic emissions were prescribed using the 1996-2004 averages of the Community Emissions Data System (CEDS) .49 emissions from Hoesly et al. (2018) as prepared for the GISS contributions to the Coupled Model Intercomparison Project, Phase 50 6 (CMIP6) (Kelley et al., 2020). These include sources from agriculture, the energy sector, the industrial sector, 51 residential/commercial/other, international shipping, solvents production and application, the transportation sector, and waste. In line with past studies, we base acetone emissions on that of ketones. VOC23-ketones emissions from Hoesly et al. (2018) were 52 53 scaled down by a ratio of acetone molecular weight to an average ketone molecular weight (58.08 g mol<sup>-1</sup>/75.3 g mol<sup>-1</sup>). 54 Maintaining the resulting spatial and temporal pattern of emissions, the magnitudes were then tuned to be close to that of Fischer 55 et al. (2012), resulting in a total of about 1 Tg yr<sup>1</sup>. This resulted in roughly 36.5% of CEDS VOC23-ketones used as acetone 56 emissions. Lacking an accurate way to obtain acetone aircraft emissions from the bulk VOCs available in the emission inventory, 57

we have neglected that sector in the simulations.

# 58 2.1.2 Terrestrial vegetation emissions

59 Emissions from land vegetation were derived from the Model Emissions of Gases and Aerosols from Nature (MEGAN), version 60 2.1 (Guenther et al., 2012), a new contribution to the ModelE. Emission response algorithms in the MEGAN2.1 model are derived

61 from input leaf area indices, solar radiation, temperature, moisture, CO<sub>2</sub> concentrations, and plant functional types and composition

62 of species (Guenther et al., 2012). The acetone vegetation emissions in the Baseline simulation in GISS ModelE2.1 <u>were calculated</u> 63 to equal  $36.1 \text{ Tg yr}^{-1}$ .

### .64 2.1.3 Biomass burning emissions

was 1.59 Tg yr<sup>-1</sup>.

Acetone emissions were prescribed from a 1996-2004 average of the NMVOC-C3H6O species from version 2.1 of the biomass burning dataset of van Marle et al. (2017), used by CMIP6. The acetone mass flux from biomass burning in the Baseline simulation

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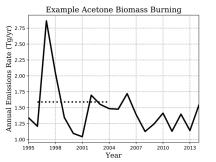
69 Figure 1 shows the biomass burning emission rate chosen for this study, and how it lies within the range of substantial interannual

variation. During the 20-year period shown, emissions averaged 1.46 Tg yr<sup>1</sup>, with a standard deviation of 0.402, and a spike in the

earlier years of emissions over 2.75 Tg yr<sup>-1</sup> is also observed (Figure 1). On top of any differences across emission inventories, the

years considered when reporting emissions may be the reason for disagreements between models (e.g. 2.40 - 2.80 Tg yr<sup>-1</sup> from the

2006 GFED-v2 emission inventory in Elias et al. (2011) and Fischer et al. (2012), compared to  $3.22 \text{ Tg yr}^{-1}$  from 1997-2001 in Folberth et al. (2006)).



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Figure 1. Illustration of interannual variation of NMVOC-C3H6O biomass burning emissions of van Marle et al. (2017) (solid
 line), used as acetone emissions in our simulation. Climatological-emissions simulations use the 1996-2004 mean (dotted line),

78 though emissions vary by month.

# 79 2.2 Sinks

### 80 2.2.1 Deposition

81 Both dry and wet deposition of acetone were included in the model, although dry deposition was, on average, 91% of total

82 deposition. The wet deposition scheme is given by Koch et al. (1999). Acetone and other species are transported within and below 83 clouds, and soluble gases are deposited depending on the conditions of the grid box they are in and a Henry's Law Coefficient

84 (Shindell et al., 2001). The Henry's Law Coefficient for acetone used in the GISS ModelE2.1 is 27 mol L<sup>-1</sup> atm<sup>-1</sup>, with a Henry

85 temperature dependence of acetone of 5300 J mol<sup>-1</sup> (Benkelberg et al., 1995; Zhou and Mopper, 1990). The dry deposition scheme

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91 uses resistance-in-series calculations, global seasonal vegetation data (Chin et al., 1996; Shindell et al., 2001; Wesely and Hicks,

.92 1977), and a reactivity factor of  $f_0=0.1$ . This resulted in an acetone deposition rate in the Baseline simulation of 22.2 Tg yr<sup>-1</sup>.

# .93 2.3 Chemistry

94 The GISS ModelE2.1 Baseline simulation estimates a net chemistry <u>flux of -20.6 Tg yr<sup>-1</sup></u>. The components can be broken up into 95 sources and sinks as follows.

### .96 2.3.1 Chemical sources

97 The Baseline simulation estimates chemical production to be 33.3 Tg yr<sup>-1</sup>. The acetone chemical scheme includes two production 98 reactions:

.99	$Paraffin + OH \rightarrow 0.35 Acetone$	(1)
200	Terpenes $+ \{OH, O_3\} \rightarrow 0.12$ Acetone	(2)

In the first reaction, acetone is produced by paraffin, a proxy tracer for paraffinic (saturated) carbon, and OH (Eq. 1). The molar yield of acetone from paraffin was found to be a strong leverage to the overall chemical source (see Section 3.5). A rate coefficient of  $8.1 \times 10^{-13}$  cm<sup>3</sup> molecule<sup>-1</sup> s<sup>-1</sup> was used (Shindell et al., 2003). Previous literature has suggested an acetone yield on a molecular scale of 0.72 (Fischbeck et al., 2017; Jacob et al., 2002; Weimer et al., 2017). Initial tests using a yield of 0.72 resulted in an

vorestimated chemistry source, leading us to re-evaluate this yield for the specific mixture of VOCs represented in the GISS
ModelE2.1.

207 Our model's anthropogenic emissions of paraffin is based on an aggregation of selected VOC groups. Based on year 2019 emissions 208 of the O'Rourke et al. (2021) dataset, we emit paraffin that is about 11% propane by mole, 22% butane and 21% pentane. 209 Multiplying these by each VOC's acetone molar yield (0.73, 0.95, 0.63, respectively), we estimate that 42% of paraffin from 210 anthropogenic sources becomes acetone in our model. Paraffin biomass burning emissions, estimated from year 2020 of SSP3 70 211 emissions (Riahi et al., 2017; Fujimori et al., 2017) contain mole fractions for propane of 9% and higher alkanes of 23%, and when !12 multiplied by acetone molar yields of 0.73 and 0.79, respectively, suggest that about 25% of paraffin from biomass burning sources 213 becomes acetone in our model. The molar yields used in these calculations were derived with suggestions from the literature 214 (Fischbeck et al., 2017; Jacob et al., 2002; Weimer et al., 2017). Refer to the "Chemical sources" section of the manuscript 215 supplement for a more detailed breakdown. Overall, an average of the 42% anthropogenic paraffin and 25% biomass burning 216 paraffin was used to conclude that approximately 35% of paraffin from emissions becomes acetone, leading to our refinement of !17 the molar yield in Eq. (1) to 0.35.

218 Additionally, reactions between terpenes and {OH, O<sub>3</sub>} were implemented with an acetone yield of 0.12 (Hu et al., 2013; Jacob et

219 al., 2002) (Eq. 2). The rates for these reactions are  $2.51 \times 10^{-11}$  e<sup>(4447)</sup> cm<sup>3</sup> molecule<sup>-1</sup> s<sup>-1</sup> for the OH reaction and  $1.40 \times 10^{-14}$  e<sup>(-7327)</sup>.

 $cm^3$  molecule<sup>-1</sup> s<sup>-1</sup> for the O<sub>3</sub> reaction, and these coefficients are enhanced from the standard  $\alpha$ -pinene one to consider the reactivity

221 variability across mono- and higher terpenes (Tsigaridis and Kanakidou, 2003).

# 222 2.3.2 Chemical sinks

<sup>123</sup> The chemical sink of acetone in the Baseline simulation is estimated to be  $53.8 \text{ Tg yr}^{-1}$ . The sinks of acetone include oxidation by <sup>124</sup> OH and Cl radicals, and photolysis:

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!25 Acetone + 0H →  $H_20$  + CH<sub>3</sub>C(0)CH<sub>2</sub> (assumed to decompose to HCH0)

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230	Acetone + $Cl \rightarrow$	$HCl + CH_{3}C($	0)CH2 (	(assumed to decom	pose to HCHO
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- 231 Acetone +  $hv \rightarrow CH_3CO + CH_3$
- !32 Acetone +  $hv \rightarrow CH_3 + CH_3 + CO$

:33 The first and second acetone destruction reactions above have rates of  $\frac{1.33 \times 10^{-13} + 3.82 \times 10^{-11}}{2.000}$  cm<sup>3</sup> molecule<sup>-1</sup> s<sup>-1</sup> and

234 7.70×10<sup>-11</sup> e<sup>(-1000T)</sup> cm<sup>3</sup> molecule<sup>-1</sup> s<sup>-1</sup>, respectively (Sander et al., 2011) (Eq. 3, 4). Previously, acetone photolysis (which only

:35 affected production of radicals and not acetone itself) did not utilize the model's photolysis scheme but was parameterized solely

!36 as a function of orbital geometry and atmospheric pressure. In the model updates, photolysis now consists of two separate reactions,

237 where acetone forms either CH<sub>3</sub>CO + CH<sub>3</sub> radicals or two CH<sub>3</sub> radicals and CO (Eq. 5, 6). The spectroscopic data used for acetone

238 photolysis is from JPL 2010 (Sander et al., 2011) and mapped onto Fast-J version 6.8d's wavelength intervals (Neu et al., 2007).

:39 The photolysis cross section for Eq. 5 is pressure-dependent while that of Eq. 6 is temperature-dependent, leading to variation in 240

yields with altitude and location. For example, in a standard atmosphere the ratio of the yield of CO to CH<sub>3</sub>CO decreases from

241 0.28 at the surface to 0.18 at 4 km altitude.

### 242 2.4 Ocean

243 Bidirectional fluxes of acetone are calculated over ocean based on the "two-phase" model of molecular gas exchange at the air-sea 244 interface of Liss & Slater (1974), as it is described in Johnson (2010). The fluxes are a function of simulated surface temperature 245 and near-surface wind speed but independent of salinity. Henry's Law constants and temperature dependence of solubility for 246 acetone are from Sander (1999). The source from ocean water and sink from the atmosphere are calculated assuming a constant 247 concentration of acetone in water (of 15 nM), the lower boundary layer atmospheric concentration, and the total transfer velocity 248 (a combination of water-side and air-side transfer velocities). The constant concentration of 15 nM follows the implementation by 149 Fischer et al. (2012) in the GEOS-CHEM model, who looked at observations and did not find a strong reasoning to make the concentration vary seasonally or spatially. The GISS ModelE2.1 Baseline simulation calculates the ocean to be a net source of 250 acetone, producing 3.94 Tg yr<sup>-1</sup>. 251

### 152 2.5 Sensitivity studies

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253 Sensitivity studies were conducted to determine the influence of key parameters on the acetone budget and its global distribution 254 (summarized in Table 1). Specifically, we were interested in seeing the sensitivity of simulated acetone to artificial perturbations 255 in given parameters, Sensitivity studies for chemistry modify the production of acetone. The Chem Cl0 and Chem Terp0 256 simulations provide no formation of acetone from chlorine or terpenes, respectively. The importance of paraffin is explored by 257 halving its yield of acetone to 17.5% in the Chem Par0.5 simulation, and by doubling its yield of acetone to 70% in the 258 Chem\_Par2.0 simulation. As vegetation was the most prominent source of acetone, the Veg\_0.7 simulation observes its reduction 259 by decreasing the MEGAN production of acetone by 30%. The Ocn 2.0 simulation aims to explore the impact of ocean acetone 260 concentration by doubling it from 15 nM to 30 nM globally. The Dep\_fo0 simulation tests dropping the reactivity factor for dry 261 deposition from 0.1 to Q. Finally, given the high interannual variability of biomass burning emissions, the BB\_2.0 simulation 262 explores the impact of doubling those emissions.

264 Table 1. Sensitivity studies conducted to observe the leverage that a specific parameter afforded the model. Simulation names, as 265 well as the parameter they target and a description, are included.

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GISS ModelE2.1 Sensitivity Simulation	Sensitivity Parameter	Description
Chem_Cl0	Chemistry Source	Acetone + Chlorine reaction rate = 0
Chem_Terp0	Chemistry Source	No reaction for production of acetone from terpenes
Chem_Par0.5	Chemistry Source	Half the yield of acetone from paraffin (17.5%)
Chem_Par2.0	Chemistry Source	Double the yield of acetone from paraffin (70%)
Veg_0.7	Vegetation	0.7 factor of acetone from MEGAN
Ocn_2.0	Ocean	Ocean acetone concentration from 15nM to 30nM
Dep_f <sub>0</sub> 0	Dry Deposition	f <sub>0</sub> changed from 0.1 to 0
BB_2.0	Biomass Burning	Double biomass burning emissions

# 281 3 Results and model evaluation

# 282 3.1 Global acetone budget and burden

. A global acetone budget table was compiled to place our estimates in context with past global modeling studies (Table 2) (Arnold

284 et al., 2005; Beale et al., 2013; Brewer et al., 2017; Dufour et al., 2016; Elias et al., 2011; Fischer et al., 2012; Folberth et al., 2006;

<sup>185</sup> Guenther et al., 2012; Jacob et al., 2002; Khan et al., 2015; Marandino et al., 2005; Singh et al., 2000, 2004; Wang et al., 2020).

186 The values of the individual fluxes in our model (global deposition, biomass burning, anthropogenic emissions, vegetation

!87 emissions, ocean net/source/sink, and chemistry net/source/sink) were mentioned previously.

288

189 Table 2. Global acetone budget table comparing burden, flux and lifetime estimates of acetone from the Baseline model to thirteen 190 previous studies.

	This Study – Baseline [2021]	<i>Wang et al.</i> [2020] <sup>a</sup>	<i>Wang et al.</i> [2020] <sup>b</sup>	Brewer et al. [2017]	Fischer et al. [2012]	Elias et al. [2011]	Jacob et al. [2002]	Other Estimates [2000-2016] <sup>e</sup>
Burden (Tg)	2.93	3.50	3.80	5.57	5.60	7.20	3.80	3.50 - 4.20
Global Deposition (Tg yr1)	22.2	25.2	12.4	12.4	12.0	<b>1</b> 9.0	<b>9</b> .00	6.00 - 26.0
Biomass Burning (Tg yr1)	1.59	4.00	2.40	2.60	2.80	2.40	4.50	3.22 – 9.0 <u>0</u>
Anthro Emissions (Tg yr-1)	1.00	0.50	3.40	3.60	0.73	1.60	1.10	1.02 – 2.0 <u>0</u>
Vegetation Emissions (Tg yr <sup>-1</sup> )	36.1	39.8	32.2	37.1	32.0	76.0	35.0	15 <u>.0</u> – 56 <u>.0</u>
Net Ocean (Tg yr <sup>-1</sup> )	3.94	-8.10	1.30	-7.50	-2.0 <u>0</u>	-8.0 <u>0</u>	13.0	4.00
Ocean Source (Tg yr <sup>-1</sup> )	15.2	33.4	45.7	51.8	80.0	20.0	27.0	20.0
Ocean Sink (Tg yr <sup>-1</sup> )	11.3	<b>4</b> 1.5	<b>4</b> 4.4	59.2	82.0	28.0	<b>1</b> 4.0	62.0
Net Chemistry (Tg yr <sup>-1</sup> )	-20.5	-11.1	-26.1	-22.5	-21.0	-53.0	-45.0	- <u>(5.50 –</u> 33.0)
Chem Source (Tg yr <sup>-1</sup> )	33.3	38.5	26.1	24.1	31.0	27.0	28.0	15.5 - 55.6
Chem Sink (Tg yr <sup>-1</sup> )	53.8	49.6	52.2	<b>4</b> 6.6	52.0	80.0	73.0	33.4 - 61.1

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Chemical Lifetime (days) <sup>c</sup>	19.9	25.8	26.6	43.6	39.3	32.9	19.0	20.9-35.6
Lifetime (days) <sup>d</sup>	12.3	11.0	12.7	17.2	14.0	21.0	14.5	12.8 – 35 <u>.0</u>

<sup>a</sup> CAM-Chem Model (Wang et al., 2020)

<sup>b</sup> GEOS-Chem Model (Wang et al., 2020)

<sup>c</sup> Chemical Lifetime = Burden/Chemical Sink

<sup>d</sup> Total Atmospheric Lifetime = Burden/Total Sink

<sup>e</sup> Singh et al. [2000, 2004], Arnold et al. [2005], Folberth et al. [2006], Marandino et al. [2006], Guenther et al. [2012], Beale et al. [2013], Khan et al. [2015], Dufour et al. [2016].

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Atmospheric burden describes the total amount of acetone that is in the atmosphere. The GISS ModelE2.1 Baseline simulation estimates the burden to be 2.93 Tg. Additionally, chemical lifetime and atmospheric lifetime can be derived from burden. The chemical lifetime of acetone is calculated as the burden divided by the chemical sink, whereas total lifetime is the burden divided by all sinks. The chemical and total atmospheric lifetimes for the Baseline simulation are calculated to be 19.9 and 12.3 days, respectively. These values are also placed in the context of previous literature in Table 2.

The GISS ModelE2.1 Baseline acetone budget is further compared to previous model studies in Figure 2. The calculated fluxes in the Baseline simulation that are less than one standard deviation away from the literature mean include anthropogenic and

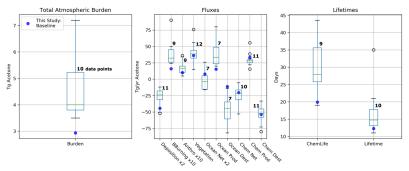
vegetation emissions, net ocean, net chemistry, chemical production, and chemical destruction (Figure S1). Biomass burning in
 GISS ModelE2.1 appears as an outlier when compared against 9 previous model studies but can be attributed to the high interannual
 *variation* with emissions (as discussed in Section 2.1.3). The value of acetone deposition is on the high (more negative) end in
 GISS ModelE2.1 relative to 11 previous studies. This might be partially attributed to differences in deposition parametrization

across models, as explored by our sensitivity study on dry deposition presented in <u>Section 3.5.2</u>. The values for oceanic acetone

sources and losses are smaller (in absolute values) than the mean from 7 previous model studies. Nevertheless, the net ocean flux

matches the literature well. Lastly, the total atmospheric burden and lifetime calculated by GISS ModelE2.1 are lower than the

revious papers, an expected consequence of the higher removal by deposition. The chemical lifetime is also calculated to be at the low end of published literature.



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Figure 2. Total atmospheric burden, fluxes, and lifetimes of acetone from the literature values in Table 2 (shown in boxes and

whiskers with outliers as open circles), and values from GISS ModelE2.1 (shown as solid <u>blue</u> circles). The number of models

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the biomass burning and anthropogenic emissions were multiplied by 10 for a better visualization of the distribution.

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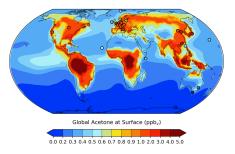
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used to create each box and whisker plot are labeled. Note that the deposition and ocean net fluxes were multiplied by 2 and that

# 351 3.2 Spatial distribution of acetone

\$52 The global distribution of acetone at the surface is given in Figure 3. It is evident that acetone mixing ratios are largest over the 153 continents, where anthropogenic, vegetation, and other terrestrial sources are located. Over the ocean, acetone mixing ratios are \$54 highest downwind of Central America and Central Africa. A comparison of the GISS ModelE2.1 results against twenty-six prior \$55 field measurements shows an overall great agreement, with a root mean squared error of 0.3494 and an R<sup>2</sup> value of 0.8306. To put \$56 these results into the context of model evaluation, a similar comparison to field measurements was done for the model's previous \$57 acetone scheme. The prior parameterization was designed as a rough representation of acetone oxidized from isoprene in the upper 158 troposphere, without regard for realism near the surface, and this is evident from the comparison with surface observations: a root ;59 mean squared error and R<sup>2</sup> value of 1.3620 and 0.0413, respectively. The improvement of the new acetone tracer model in the \$60 GISS ModelE2.1 is evident from these statistics.



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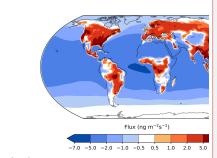
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Figure 3. GISS ModelE2.1 spatial distribution of annual mean acetone at surface for the Baseline simulation. Filled circles
represent data from twenty-six field measurements (de Gouw et al., 2004; Dolgorouky et al., 2012; Galbally et al., 2007; Guérette
et al., 2019; Hu et al., 2013; Huang et al., 2020; Langford et al., 2010; Lewis et al., 2005; Li et al., 2019; Read et al., 2012; Schade
& Goldstein, 2006; Singh et al., 2003; Solberg et al., 1996; Warneke & de Gouw, 2001; Yoshino et al., 2012; Yuan et al., 2013).
The root mean squared error and the R<sup>2</sup> value between the Baseline acetone estimations and the field measurements are 0.3494
and 0.8306, respectively. A nonlinear colorbar is used to better differentiate the details in the map.

69 A breakdown of the acetone bidirectional fluxes indicates that its chemical production is concentrated over the continents, while \$70 chemical destruction is primarily over the oceans (Figure 4), Hotspots of production over the continents include the Southeastern, \$71 United States and Central South America, East and Northern Asia, and Central Africa. Chemical sinks over the oceans are stronger ;72 in the tropics than in the high southern or northern latitudes. Annually, there is a net flux of about -20.46 Tg yr<sup>1</sup>, Observing the ;73 chemical flux across all four seasons, the net loss appears unaffected while the net source changes more significantly, following ;74 the seasonality of precursor compounds like isoprene and terpenes (Figure 4). Chemical production is strongest in the months of ;75 June/July/August, primarily in North America and Northern Asia. Production is weakest in the months of ;76 December/January/February, losing almost all production in North America and Northern Asia entirely. Still, a net negative flux ;77 is present for all four seasons (Figure 4).

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Figure 4. Annual average of acetone net chemistry fluxes (columnintegrated) in the Baseline simulation, with red indicating a net source and blue indicating a net sink. A nonlinear colorbar is used to better differentiate the details in the map. The weighted global mean of the net chemistry flux is shown in a box on the lower right.¶

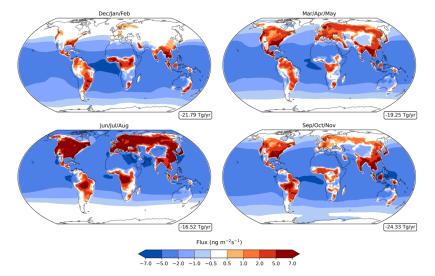
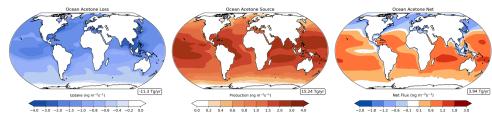


Figure 4. <u>Net acetone</u> chemistry fluxes (column-integrated) in the Baseline simulation for December-February (top left), March-May (top right), June-August (bottom left), and September-November (bottom right), with red indicating a net source and blue indicating a net sink. Nonlinear colorbars are used to better differentiate the details in the map. The weighted global means of the net chemistry fluxes are shown in boxes on the lower right <u>of each subplot</u>.

103 The ocean acetone sources and sinks are unevenly distributed across latitudes (Figure 5). Oceanic uptake of acetone is mostly 104 concentrated in the northern rather than the southern oceans, while the ocean acetone source is strongest in the tropics and decreases 105 at higher latitudes of both hemispheres, Combining these two unidirectional fluxes results in the ocean serving as a sink in the 106 northern high latitudes, a source in the tropical latitudes, and near neutral at the high southern latitudes (Figure 5). This finding 107 corroborates very well with findings from Fischer et al. (2012) and Wang et al. (2020). Additionally oceanic bidirectional fluxes 108 of acctone present trends over the four seasons (Figure S2). Overall, every season has a positive global mean net flux. However, 109 production becomes strongest in the months of December through May, and weakest in the months of June through November. 10 Off the coast of western South America, the ocean appears to be a net sink of acetone, even though this latitude band is generally 111 a source of acetone (Figure 5, Figure S2). This is especially evident in the months of June/July/August and 112 September/October/November. As the model simulates this location to have high levels of acetone at the surface (Figure 3), we

believe the acetone in the air is driving the ocean to be a sink there.



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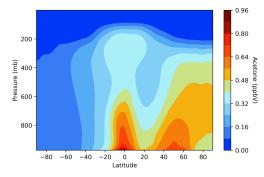
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21	Figure 5. Annual mean oceanic acetone uptake (left), oceanic acetone production (middle), and net bidirectional flux (right) in the
122	Baseline simulation, with red indicating a net source and blue indicating a net sink. Nonlinear colorbars are used to better
123	differentiate the details in the map. The corresponding weighted global means of the ocean fluxes are shown in boxes on the lower
124	right of each subplot,

# **3.3 Vertical distribution of acetone**

The vertical distribution of acetone varies by latitude, with near-surface air mixing ratios being higher in the tropics and in the northern midlatitudes (Figure 6). Acetone levels in the atmosphere decrease with height, a direct result of sinks dominating the sources, Prior to the implementation of an acetone tracer in the GISS ModelE2.1, when acetone was derived from the zonal mean of isoprene, the vertical distribution looked very different. Acetone was only concentrated around the tropics and did not extend

nearly as high into the <u>troposphere</u>. The complexity of Figure <u>6</u> supports the new acetone tracer scheme as a significant
 improvement to the GISS ModelE.



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Figure 6. GISS ModelE2.1 vertical distribution of acetone air mixing ratios across latitudes in the Baseline simulation.

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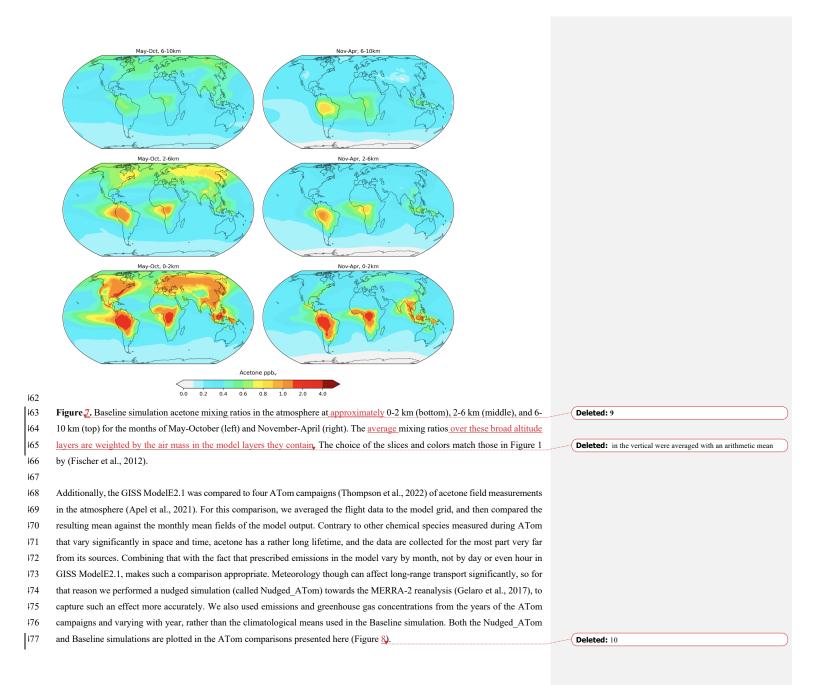
135 Another modeled vertical distribution of acetone, including a differentiation between two long seasons, is explored in Figure 7. In

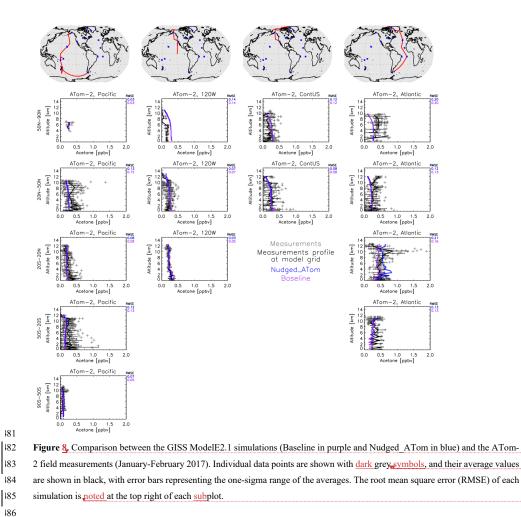
general, it was found that acetone mixing ratios are higher in the months of May-October than in November-April, and that this

relationship is stronger in the lower atmosphere (0-2 km) than the upper atmosphere (6-10 km). This finding corroborated well

138 with a similar analysis done by Fischer et al. (2012).

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$\langle \rangle \rangle$	Deleted: right	
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	<b>Deleted:</b> Figure 7. Annual average of acetone ocean bidirectional fluxes in the Baseline simulation, with red indicating a net source and blue indicating a net sink. A nonlinear colorbar is used to better differentiate the details in the map. The weighted global mean of the net chemistry flux is shown in a box on the lower right.	
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There are very few notable differences between the nudged and climatological simulations. An example is the tropical Atlantic

Ocean, where during ATom-2 (Figure 3), the nudged simulation calculates higher acetone concentrations, but without gain of skill.

Both model simulations miss the upper tropospheric peak that is found in the measurements, likely indicating a missing long-range

transported plume. There is a similar result for, ATom-3 (Figure S4) for the southern Atlantic Ocean mid-latitudes, where the

nudged simulation is higher. Contrary to the ATom-2 case, both simulations for the ATom-3 case calculate an upper tropospheric

maximum, which is not found in the measurements. The tropical and southern mid-latitude Atlantic Ocean regions are both

downwind of African biomass burning zones during ATom-2 and ATom-3, respectively, hinting to a primary and/or secondary

incorrect source of acetone related with biomass burning and subsequent long-range transport. Other than those few cases, for the

most part the two simulations are indistinguishable, indicating that our conclusions comparing climatological simulations to ATom

should be robust. (Figure, <u>8, Figures</u>, <u>83-S5</u>). This is important to remember in Section 3.5.3, where we perform sensitivity analyses
 using climatological simulations and comparing against all four ATom campaigns.

### 307 3.4 Seasonality of acetone

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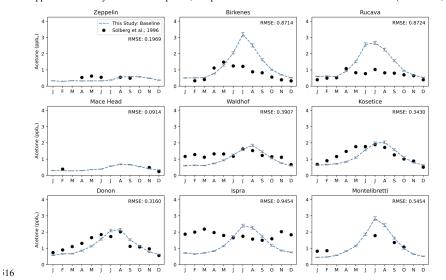
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right of each subplot,

Most European sites presented in Figure 3 have monthly-resolved measurements that can be used to analyze the seasonal behavior
 of acetone in the model (Figure 9, Figure S6) (Solberg et al., 1996). These sites differ with respect to their geographic locations

and their proximity to anthropogenic sources. Zeppelin, Birkenes, Rucava, and Mace Head are all coastal sites, while Waldhof,
Kosetice, Donon, Ispra, and Montelibretti are inland sites. Regarding anthropogenic sources, Zeppelin is the most remote location
and Birkenes and Rucava each have small sources. Mace Head is a site affected by the marine boundary layer, and Waldhof,
Kosetice and Donon are sites with small local anthropogenic sources that are generally located in higher emission regions.
Montelibretti and particularly Ispra are subject to the highest anthropogenic sources. The measurements taken at Ispra show an

opposite seasonality than what is expected, and previous studies have considered this anomalous (Jacob et al., 2002).



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The GISS ModelE2.1 matches the seasonality of the measurements especially well in Zeppelin, Mace Head, Waldhof, Kosetice,

Figure 2. Acetone over twelve months at nine European sites, similar to that of Jacob et al. (2002). The modeled estimates of

acetone at the surface from the Baseline simulation are shown as dashed blue lines and the grey error bars represent the one-sigma

range of the modeled concentrations in the climatological mean of 5 years. Field measurements from Solberg et al (1996) are

shown as solid black dots. Root mean squared error between the Baseline simulation and field measurements are shown at the top

i24 and Donon; the average root mean square error between the Baseline model and the measurements at these five sites are 0.27. The i25 Baseline model overestimates the measurements in Birkenes and Rucava (RMSE = 0.87 for both), even though these two sites

have low anthropogenic sources. This overestimation has been attributed to the vegetation source, which has a distinct seasonality

and is much stronger than any other source there. Interestingly, in Montelibretti, the model's overestimation of vegetation, yet underestimation of local emissions, results in a decent estimation of the sources there (RMSE = 0.55) (Figure 9).

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As mentioned previously, an analysis of the distribution of the regional sources and sinks at the nine European sites shows that,
except for Zeppelin and Mace Head, all studied European sites have vegetation as the dominant source that strongly contributes to
the simulated seasonality of concentrations (Figure <u>\$7</u>). Vegetation sources peak in the summer months and are lower in the winter.
Deposition is a major sink of acetone that is comparable in magnitude with the vegetation source. Ocean uptake of acetone follows
a weak seasonal cycle, being stronger in the summer months. <u>Relatively, the other fluxes (anthropogenic emissions, biomass</u>)

burning and ocean production) do not exhibit much seasonality at these locations (Figure 57).

We also compared the GISS ModelE2.1's surface acetone at observation sites with less temporal coverage (Figure <u>\$8</u>). In general,
the GISS ModelE2.1 matches the field measurements well. This is especially true for the non-summer seasons in Rosemount and
Berkeley, <u>USA</u>, and the summer peaks in Utrecht, <u>Netherlands</u> and Mainz, <u>Germany</u>. The model seems to be overestimating
acetone around Australia, as shown by comparisons with Cape Grim and Wollongong, while underestimating emissions in large
cities like Shenzhen and Beijing, <u>China</u>, London, <u>UK</u>, and Paris, <u>France (Figure S8)</u>

### 3.5 Sensitivity studies

The sensitivity simulations presented here have been described in Section 2.5 and in Table 1. We grouped them in two categories: those directly related with chemical sources and sinks, and those related with terrestrial and oceanic acetone fluxes. Overall, the sensitivity studies that presented the largest changes to total atmospheric burden included Chem\_Terp0, Chem\_Par0.5, Chem\_Par2.0, Veg\_0.7, Ocn\_2.0, and Dep\_f00 (all but Chem\_Cl0 and BB\_2.0) (Figures S92514).

# 558 3.5.1 Chemistry

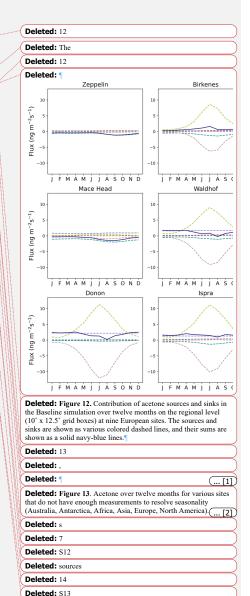
i59 Chemistry sensitivity tests that modified the production of acetone were analyzed with respect to the budget and global distribution

i60 of acetone. In the Chem\_Cl0 simulation, where no acetone oxidation by the chlorine radical occurs, the overall global acetone

i61 budget does not change. However, in some places like Rucava, Ispra, Montelibretti, and Shenzhen, the shape of the acetone

i62 concentration profile over the year changes slightly (Figure <u>10</u>, Figure <u>S15</u>).

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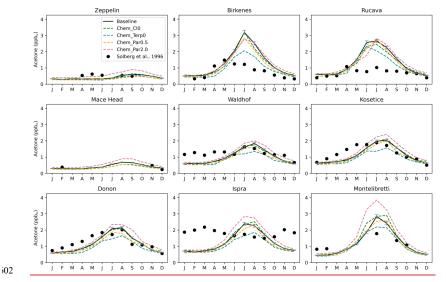
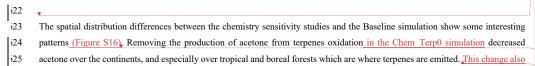


Figure 10. Similar to Figure 9, but with the chemistry sensitivity studies added. The modeled estimates of acetone at the surface from the Baseline simulation are shown as solid black lines, and the sensitivity studies are as follows: removing the acetone + chlorine reaction (dashed green lines), removing the production of acetone from terpenes (dashed blue lines), halving the yield of acetone from paraffin (dashed orange lines), and doubling the yield of acetone from paraffin (dashed pink lines). Field measurements from Solberg et al., (1996) are shown as solid black dots.

The Chem\_Terp0 simulation that removes the production of acetone from terpenes decreases the summer peak of acetone by as much as 35.5% in Birkenes, 25.5% in Mainz, and 25.3% in Berkeley (Figure <u>10</u>, Figure <u>\$15</u>). Other sites like Montelibretti, Ispra and Paris have their summer peak decreased by 22.6%, 22.2%, and 19.0%, respectively (Figure <u>10</u>, Figure <u>\$15</u>). Coastal and remote areas like Zeppelin, Mace Head and Dumont d'Urville, <u>Antarctica</u> are not impacted by the removal of terpenes (Figure <u>10</u>, Figure <u>\$15</u>).

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There seems to be some nonlinearities with the relationship between acetone abundance and its yield from paraffin, as the results from the Chem\_Par2.0 and Chem\_Par0.5 simulation reveal that doubling the yield has a stronger impact than halving it. For instance, in Montelibretti, doubling the yield from paraffin increases the summer peak by 35.7%, while halving the yield decreases the summer peak by only 8.3% (Figure 10). A similar relationship is observed at other sites: Ispra (19.1% increase with double paraffin, 2.5% decrease with half paraffin) and Berkeley (12.7% increase with double paraffin, 2.5% decrease with half paraffin) (Figure 10, Figure 515). Overall, we explored chemistry sensitivities that would tend to push acetone in both directions. The Baseline simulation falls between our tests, which we have identified as important uncertainties.



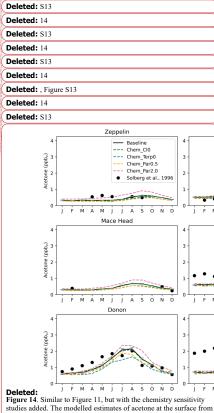


Figure 14. Similar to Figure 11, but with the chemistry sensitivity studies added. The modelled estimates of acctone at the surface from the Baseline simulation are shown as solid black lines, and the sensitivity studies are as follows: removing the acctone + chlorine reaction (dashed green lines), removing the production of acctone from terpenes (dashed blue lines), halving the yield of acetone from paraffin (dashed orange lines), and doubling the yield of acetone from paraffin (dashed pink lines). Field measurements from Solberg et al., (1996) are shown as solid black dots.

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48         tergenes oxidation products (Figure S16, top left), Jialving production of acctone from paraffin oxidation in the Chem Par0.5           49         simulation only decreased acctone concentrations over the continents, while doubling it in the Chem Par2.6 simulation increased           50         acctone concentrations over the continents and attengthened acetone destruction over the zropical occans (Figure S16, top right and           51         bittom middle, respectively.) Setting the acctone t chlorine reaction rate to 0 in the Chem C10 simulation resulted in negligible           52         changes across the globe (anomalies of <0.4 ng m <sup>2</sup> s <sup>-1</sup> ).           53         J.S.2 Terrestrial and occanic fluxes           54         Terrestrial and occanic fluxes sensitivities were analyzed at the same sites. The vegetation flux sensitivity, Veg. 0.7, reduced acetone production from MEGAN by 30%. This change decreased the summer peak of acetone down at nearly every location studied, but most notably by 32.6% in Birkenes, 22.9% in Rucava, and 22.2% in Rosemount (Figures S17, S18).           56         In the oceanic flux sensitivity simulation, Ocn 2.0, the concentration of acetone in the water was doubled from 15 mM to 30 nM.           57         In the oceanic flux sensitivity simulation varied with geographic location. For instance, in Birkenes is at 58°N, while Montelibretti, it was increased by 16.1% (Figure S17, S18).           56         Another broader finding from the occan sensitivity study is that doubling do cean acetone concentration did not change much atmospheric acetone throughout the year (Figure S13, S18).           66 <th>Formatted: Fo</th>	Formatted: Fo
<ul> <li>acetone concentrations over the continents and gtrengthened acetone destruction over the propical oceans (Figure S16, top right and bottom middle, respectively). Setting the acetone ± chlorine reaction rate to 0 in the Chem_Cl0 simulation resulted in negligible changes across the globe (anomalies of &lt;0.4 ng m<sup>2</sup> s<sup>-1</sup>).</li> <li><i>3,5.2 Terrestrial</i> and oceanic fluxes</li> <li>Terrestrial and oceanic fluxes sensitivities were analyzed at the same sites. The vegetation flux sensitivity, Veg_0.7, reduced acetone production from MEGAN by 30%. This change decreased the summer peak of acetone down at nearly every location studied, but most notably by 32.6% in Birkenes, 22.9% in Rucava, and 22.2% in Rosemount (Figures §17, S18).</li> <li>In the oceanic flux sensitivity simulation, Ocn_2.0, the concentration of acetone in the water was doubled from 15 nM to 30 nM. The results of this simulation varied with geographic location. For instance, in Birkenes, doubling ocean concentration reduced overall acetone by 13.9%, while in Montelibretti, it was increased by 16.1% (Figure §17). Even though Birkenes is more of a coastal city than Montelibretti, this result may simply be a temperature effect; Birkenes is at 58°N, while Montelibretti is at 42°N, and a warmer ocean may produce more acetone. Overall, in most places, the doubling ocean acetone concentration did not change much atmospheric acetone throughout the year (Figure \$13). Specifically, in this sensitivity study the emissions of doubled while the uptake of increased by 40%. This difference may be attributed to the fact that a higher ocean concentration will generally cause less resistance in the emission direction, but more resistance in the uptake differences in oceanic acetone ensitivity simulation, the reactivity factor, fa, was reduced from 0.1 to 0. As a result, the amount of acetone removed by deposition decreased, and the atmospheric acetone concentration increased chemical destruction, and an overall higher burden of acetone in the atmospheric (Figure</li></ul>	Formatted: Fo
<ul> <li>bottom middle, respectively). Setting the acetone + chlorine reaction rate to 0 in the Chem Cl0 simulation resulted in negligible</li> <li>changes across the globe (anomalies of &lt;0.4 ng m<sup>2</sup> s<sup>-1</sup>),</li> <li>3.5.2 Terrestrial and oceanic fluxes</li> <li>Terrestrial and oceanic fluxes sensitivities were analyzed at the same sites. The vegetation flux sensitivity, Veg_0.7, reduced acetone production from MEGAN by 30%. This change decreased the summer peak of acetone down at nearly every location studied, but most notably by 32.6% in Birkenes, 22.9% in Rucava, and 22.2% in Rosemount (Figures, 217, 518),</li> <li>In the oceanic flux sensitivity simulation, Ocn_2.0, the concentration of acetone in the water was doubled from 15 nM to 30 nM. The results of this simulation varied with geographic location. For instance, in Birkenes, doubling ocean concentration reduced overall acetone by 13.9%, while in Montelibretti, it was increased by 16.1% (Figure S17). Even though Birkenes is more of a coastal city than Montelibretti, this result may simply be a temperature effect; Birkenes is at 58°N, while Montelibretti is at 42°N, and a warmer ocean may produce more acetone. Overall, in most places, the doubling ocean acetone concentration did not change much atmospheric acetone throughout the year (Figure S13, Specifically, in this sensitivity study the emissions of doubled frinding from the oceanic uptake of acetone (Figure S13, Specifically, in this sensitivity study the emissions doubled while the uptake only increased by 40%. This difference may be attributed to the fact that a higher ocean concentration will generally cause less resistance in the emission direction, but more resistance in the uptake direction. The differences in oceanic acetone emissions and uptakes in this sensitivity study also resulted in increased chemical destruction, and an overall higher burden of acetone in the atmosphere (Figure S13).</li> <li>In the dry deposition decreased, and the atmospheric acetone concentration increases were found</li></ul>	Deleted: This concentration inc
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Baseline simulation are presented in Figure 1. Decreasing acetone production from MEGAN vegetation by 30% resulted in a	Deleted: incre

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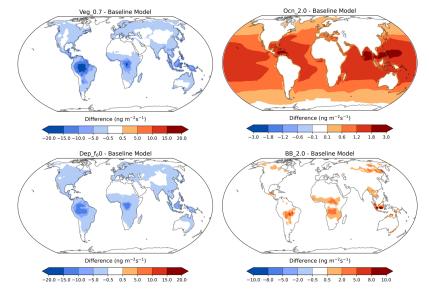
decrease of acetone mixing ratios over the tropical and boreal forests, where this source is most prominent (Figure 11, top left).

Doubling ocean acetone concentrations increased production of acetone from the oceans globally. This increase was stronger in

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'37 the tropics, due to the higher sea surface temperatures (Figure 11, top right). Reducing the reactivity factor for dry deposition decreased the amount of acetone removed by deposition over the continents (Figure 11, bottom left), in particular where acetone

- '38
- '39 concentration is elevated (Figure 3). Finally, doubling biomass burning emissions did not change acetone mixing ratios much,
- '40 other than over biomass burning hotspots like central South America, central Africa, Southeast Asia, and Siberia (Figure 11, bottom '41 right).



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Figure 11. Acetone anomalies from the Baseline simulation for the vegetation (top left), ocean (top right), dry deposition (bottom

744 left) and biomass burning (bottom right) sensitivities, with red indicating an increase and blue indicating a decrease of the specific

'45 flux. Nonlinear colorbars are used to better differentiate the details in the map.

### '46 3.5.3 ATom comparisons

'47	The ATom comparisons were replicated with the sensitivity simulations (Figure 12, Figures \$19,\$21). Doubling the paraffin
'48	yield of acetone seemed to have the most noticeable impacts on the vertical profiles. As seen during ATom-1 (July-August
'49	2016), doubling the paraffin yield decreases the root mean square error (RMSE) against measurements in the Northern
'50	hemisphere polar atmosphere and brings the model to closer agreement to observations, but decreases the agreement throughout
'51	the remote Pacific Ocean, which implies different chemical formation pathways over the more polluted northern hemisphere on
'52	the Atlantic Ocean side, compared to the Pacific Ocean. Nearly the exact opposite is calculated in the case of the halving of the
'53	paraffin yield of acetone, which adds confidence to the chemical pathway explanation (Figure 12). The doubling of the ocean
'54	acetone concentration shows a small improvement (decrease) in the RMSE over the tropical and north Atlantic Ocean during
'55	ATom-1 and an even smaller decrease over the north hemisphere Pacific Ocean, but an increase over the tropical and south
'56	Pacific Ocean, showing the potential role of different ocean concentrations of acetone across the globe (Figure 12). It needs to be
'57	noted though that the model performs fairly well in those regions already, so the small improvements mentioned do not largely
'58	affect the regional acetone concentrations, as also expected due to the rather weak acetone source from the ocean.

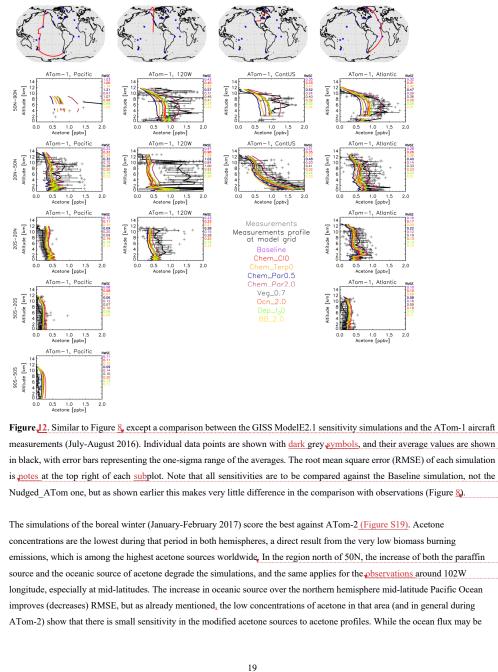
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# Deleted: (Figure 18)



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'91	During boreal fall (ATom-3), doubling the paraffin yield tends to overshoot most of the observations (Figure S20), contrary to
'92	what was calculated during boreal summer (ATom-1; Figure 12). This is the case for most ATom-3 Atlantic Ocean flights, while
'93	an improvement is calculated when comparing with the flights near the west coast of the US or the Pacific Ocean mid-latitudes.
'94	These results reveal that the model may be underestimating a paraffin source during boreal summer, which diminishes during
'95	boreal fall.
'96	
'97	The boreal spring season (April-May 2018; ATom-4; Figure S21) is the hardest for the model to simulate when it comes to
'98	northern hemisphere concentrations. All sensitivity studies greatly underestimate observations, in particular the long-range
'99	transport upper tropospheric amount near the polar latitudes but also the concentrations measured throughout the troposphere at
300	northern mid-latitudes. The model skillfully simulates tropical and southern hemisphere profiles, while it cannot reproduce the
301	higher concentrations at northern latitudes. The increased yield from paraffin or the increased oceanic concentration do reduce
302	RMSE, but still fall short on capturing the magnitude, or the shape, of the profiles of the spring hemisphere. We cannot infer
303	from our model simulations whether this is a missing source or an overestimated sink, but the latter appears to be more plausible,
304	given the large underestimation of all modeled profiles at northern mid-latitudes. In the southern hemisphere, the increase of
305	oceanic acetone clearly degrades model skill, as was frequently the case during the other campaigns presented above.
306	It is worth mentioning that for most cases the changes in the source of acetone do not alter the shape of the vertical profile. This
307	means that the transport or chemical sinks of acetone dictate its spatiotemporal distribution more than sources, while the sources
308	do affect the magnitude of that distribution, quite significantly under some of the sensitivity simulations described here.

small, these ATom comparisons reveal that they especially matter in the southern latitudes. These are the same latitudes where

the ocean appears to be in equilibrium (neither a strong source nor sink) (Figure 5).

# 309 4 Conclusion

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The development of acetone's representation in the NASA GISS ModelE2.1 from its previous simplistic parameterization of 310 311 instantaneous isoprene to a full tracer experiencing transport, chemistry, emissions, and deposition of its own, marks a significant 312 improvement to the model's chemical scheme. Calculations of the 3-dimensional distribution of acetone as a function of time, as 313 well as evaluations of its atmospheric burden and source/sink fluxes demonstrate the complexity of acetone's spatiotemporal 314 distribution in the atmosphere. An analysis was conducted to assess the simulated global acetone budget in the context of past 315 modeling studies. Further comparisons were made against field observations on a variety of spatial and temporal scales, which 316 indicated that the model agrees well with surface field measurements and vertical profiles in the remote atmosphere. The chemical 317 formation of acetone from precursor compounds such as paraffin was found to be an uncertain yet impactful factor. Vegetation 318 fluxes as calculated by MEGAN were identified as the dominant acetone source which dictates its seasonality. Additionally, the 319 acetone concentration in seawater was found to affect oceanic sources more than oceanic sinks. 320

Any feedback between acetone and the rest of the chemistry, and particularly ozone, have not been assessed here, and should be

the goal of a future study. Additionally, the current ocean-acetone interaction uses a constant concentration of acetone in the ocean.

123 It will be helpful to test a more realistic, non-uniform ocean acetone concentration, when this becomes available. Finally, other

324 atmospheric conditions such as surface wind speed may be considered further when modifying the ocean scheme.

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**Deleted:** The work presented here demonstrates the usefulness of the approach to evaluate a chemical species in the model and can be used for similar evaluations of other important gaseous and aerosol species.

### 340 Code Availability

The GISS ModelE code is publicly available at <u>https://simplex.giss.nasa.gov/snapshots/</u>. The most recent public version is E.2.1.2;
the version of the code used here is already committed in the non-public-facing repository and will be released in the future following the regular release cycle of ModelE, under version E3.1.

### 344 Data Availability

We have made available the simulated three-dimensional distributions of acetone from each simulation described in the paper (Baseline, sensitivity simulations in Table 1, and Nudged\_ATom). These are found in zip files, grouped by simulation, here: <u>https://doi.org/10.5281/zenodo.7567614</u>. Each zip file contains a series of netCDF format files with filenames (month]\_5yrAvg\_Acetone\_{simulation}.nc, where each file is a climatological average over 5 years of repeated forcing conditions.

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151 The exception is the transient-forcing simulation "Nudged\_ATom", which contains single-month averages of acetone from JUL

352 2016 through MAY 2018, to cover the ATom observational period. The file names for that simulation are of the form:

 $\label{eq:source} \ensuremath{\texttt{S3}} \qquad \{\ensuremath{\texttt{month}}\xspace_{\texttt{Acetone}}\ensuremath{\texttt{Nudged}}\xspace_{\texttt{ATom.nc.}}\xspace \ensuremath{\texttt{Acetone}}\xspace_{\texttt{v}$ 

These are hybrid sigma levels, but nominal pressure middles and edges are given in the plm and ple variables, respectively, and

the grid box surface areas are also provided.

### 356 Author Contribution

157 KT conceived the study and guided the model development which was done by GF. All simulations presented here were performed

by GF. DS advised during the whole development process. AR did the literature search and all comparisons against other modeling

studies. With the exception of the ATom analysis and plots which were done by KT, and comparisons against field measurements

and the rest of the plots were done by AR. AR drafted the first version of the manuscript, and all authors contributed to it. GF

361 prepared all model outputs for dissemination.

### 362 Competing Interests

363 The authors declare that they have no conflict of interest.

### 364 Acknowledgements

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- 368 Center.

### 369 References

Apel, E. C., Asher, E. C., Hills, A. J., and Hornbrook, R. S.: ATom: Volatile Organic Compounds (VOCs) from the TOGA
 instrument, Version 2, ORNL DAAC, https://doi.org/10.3334/ORNLDAAC/1936, 2021.

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The 3-dimensional model output of acetone concentrations will be made public at the GISS website at the time of publication in the discussion phase, as was done in other publications (e.g. <u>https://pubs.gis.nasa.gov/abs/bab/S000g.html</u>). This statement will be modified accordingly for final publication.

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- Arnold, S. R., Chipperfield, M. P., and Blitz, M. A.: A three-dimensional model study of the effect of new temperature-dependent 380 381 quantum yields for acetone photolysis, J. Geophys. Res. Atmospheres, 110, https://doi.org/10.1029/2005JD005998, 2005.
- 382 Beale, R., Dixon, J. L., Arnold, S. R., Liss, P. S., and Nightingale, P. D.: Methanol, acetaldehyde, and acetone in the surface waters 383 of the Atlantic Ocean, J. Geophys. Res. Oceans, 118, 5412-5425, https://doi.org/10.1002/jgrc.20322, 2013.
- 384 Benkelberg, H.-J., Hamm, S., and Warneck, P.: Henry's law coefficients for aqueous solutions of acetone, acetaldehyde and acetonitrile, and equilibrium constants for the addition compounds of acetone and acetaldehyde with bisulfite, J. Atmospheric 385 Chem., 20, 17-34, https://doi.org/10.1007/BF01099916, 1995. 386
- 387 Brewer, J. F., Bishop, M., Kelp, M., Keller, C. A., Ravishankara, A. R., and Fischer, E. V.: A sensitivity analysis of key natural 388 factors in the modeled global acetone budget, J. Geophys. Res. Atmospheres, 122, 2043-2058, https://doi.org/10.1002/2016JD025935, 2017. 389
- 390 Chin, M., Jacob, D. J., Gardner, G. M., Foreman-Fowler, M. S., Spiro, P. A., and Savoie, D. L.: A global three-dimensional model 391 of tropospheric sulfate, J. Geophys. Res. Atmospheres, 101, 18667-18690, https://doi.org/10.1029/96JD01221, 1996.
- Dolgorouky, C., Gros, V., Sarda-Esteve, R., Sinha, V., Williams, J., Marchand, N., Sauvage, S., Poulain, L., Sciare, J., and 392 Bonsang, B.: Total OH reactivity measurements in Paris during the 2010 MEGAPOLI winter campaign, Atmospheric Chem. Phys., 393 12, 9593–9612, https://doi.org/10.5194/acp-12-9593-2012, 2012. 394
- Dufour, G., Szopa, S., Harrison, J. J., Boone, C. D., and Bernath, P. F.: Seasonal variations of acetone in the upper troposphere-(95 396 lower stratosphere of the northern midlatitudes as observed by ACE-FTS, J. Mol. Spectrosc., 323, 67-77, 397 https://doi.org/10.1016/j.jms.2016.02.006, 2016.
- Elias, T., Szopa, S., Zahn, A., Schuck, T., Brenninkmeijer, C., Sprung, D., and Slemr, F.: Acetone variability in the upper 398 399 troposphere: analysis of CARIBIC observations and LMDz-INCA chemistry-climate model simulations, Atmospheric Chem. )00 Phys., 11, 8053-8074, https://doi.org/10.5194/acp-11-8053-2011, 2011.
- Fischbeck, G., Bönisch, H., Neumaier, M., Brenninkmeijer, C. A. M., Orphal, J., Brito, J., Becker, J., Sprung, D., van Velthoven, )01 P. F. J., and Zahn, A.: Acetone-CO enhancement ratios in the upper troposphere based on 7 years of CARIBIC data: new insights )02)03 and estimates of regional acetone fluxes, Atmospheric Chem. Phys., 17, 1985-2008, https://doi.org/10.5194/acp-17-1985-2017, )04 2017.
- Fischer, E. V., Jacob, D. J., Millet, D. B., Yantosca, R. M., and Mao, J.: The role of the ocean in the global atmospheric budget of )05 906 acetone, Geophys. Res. Lett., 39, https://doi.org/10.1029/2011GL050086, 2012.
- )07 Folberth, G. A., Hauglustaine, D. A., Lathière, J., and Brocheton, F.: Interactive chemistry in the Laboratoire de Météorologie )08 Dynamique general circulation model: model description and impact analysis of biogenic hydrocarbons on tropospheric chemistry, Atmospheric Chem. Phys., 6, 2273-2319, https://doi.org/10.5194/acp-6-2273-2006, 2006. )09
- Fujimori, S., Hasegawa, T., Masui, T., Takahashi, K., Herran, D. S., Dai, H., Hijioka, Y., and Kainuma, M.: SSP3: AIM implementation of Shared Socioeconomic Pathways, Glob. Environ. Change, 42, 268–283, )10 )11 https://doi.org/10.1016/j.gloenvcha.2016.06.009, 2017. 12
- Galbally, I., Lawson, S. J., Bentley, S., Gillett, R., Meyer, M., and Goldstein, A.: Volatile organic compounds in marine air at Cape )13 14 Grim, Australia, Environ. Chem. - Env. CHEM, 4, https://doi.org/10.1071/EN07024, 2007.
- )15
- Gelaro, R., McCarty, W., Suárez, M. J., Todling, R., Molod, A., Takacs, L., Randles, C. A., Darmenov, A., Bosilovich, M. G., Reichle, R., Wargan, K., Coy, L., Cullather, R., Draper, C., Akella, S., Buchard, V., Conaty, A., Silva, A. M. da, Gu, W., Kim, G.-116 K., Koster, R., Lucchesi, R., Merkova, D., Nielsen, J. E., Partyka, G., Pawson, S., Putman, W., Rienecker, M., Schubert, S. D., )17 )18 Sienkiewicz, M., and Zhao, B.: The Modern-Era Retrospective Analysis for Research and Applications, Version 2 (MERRA-2), )19 J. Clim., 30, 5419-5454, https://doi.org/10.1175/JCLI-D-16-0758.1, 2017.
- )20 de Gouw, J., Warneke, C., Holzinger, R., Klüpfel, T., and Williams, J.: Inter-comparison between airborne measurements of methanol, acetonitrile and acetone using two differently configured PTR-MS instruments, Int. J. Mass Spectrom., 239, 129-137, 121 122 https://doi.org/10.1016/j.ijms.2004.07.025, 2004.

- Guenther, A. B., Jiang, X., Heald, C. L., Sakulyanontvittaya, T., Duhl, T., Emmons, L. K., and Wang, X.: The Model of Emissions of Gases and Aerosols from Nature version 2.1 (MEGAN2.1): an extended and updated framework for modeling biogenic emissions, Geosci. Model Dev., 5, 1471–1492, https://doi.org/10.5194/gmd-5-1471-2012, 2012.
- Guérette, É.-A., Paton-Walsh, C., Galbally, I., Molloy, S., Lawson, S., Kubistin, D., Buchholz, R., Griffith, D. W. T., Langenfelds,
   R. L., Krummel, P. B., Loh, Z., Chambers, S., Griffiths, A., Keywood, M., Selleck, P., Dominick, D., Humphries, R., and Wilson,
   S. R.: Composition of Clean Marine Air and Biogenic Influences on VOCs during the MUMBA Campaign, Atmosphere, 10, 383,
   https://doi.org/10.3390/atmos10070383, 2019.
- Hoesly, R. M., Smith, S. J., Feng, L., Klimont, Z., Janssens-Maenhout, G., Pitkanen, T., Seibert, J. J., Vu, L., Andres, R. J., Bolt,
   R. M., Bond, T. C., Dawidowski, L., Kholod, N., Kurokawa, J., Li, M., Liu, L., Lu, Z., Moura, M. C. P., O'Rourke, P. R., and
   Zhang, Q.: Historical (1750–2014) anthropogenic emissions of reactive gases and aerosols from the Community Emissions Data
   System (CEDS), Geosci. Model Dev., 11, 369–408, https://doi.org/10.5194/gmd-11-369-2018, 2018.
- Hu, L., Millet, D. B., Kim, S. Y., Wells, K. C., Griffis, T. J., Fischer, E. V., Helmig, D., Hueber, J., and Curtis, A. J.: North
   American acetone sources determined from tall tower measurements and inverse modeling, Atmospheric Chem. Phys., 13, 3379–
   3392, https://doi.org/10.5194/acp-13-3379-2013, 2013.
- Huang, X.-F., Zhang, B., Xia, S.-Y., Han, Y., Wang, C., Yu, G.-H., and Feng, N.: Sources of oxygenated volatile organic
  compounds (OVOCs) in urban atmospheres in North and South China, Environ. Pollut., 261, 114152,
  https://doi.org/10.1016/j.envpol.2020.114152, 2020.
- Jacob, D. J., Field, B. D., Jin, E. M., Bey, I., Li, Q., Logan, J. A., Yantosca, R. M., and Singh, H. B.: Atmospheric budget of
   acetone, J. Geophys. Res. Atmospheres, 107, ACH 5-1-ACH 5-17, https://doi.org/10.1029/2001JD000694, 2002.
- Johnson, M. T.: A numerical scheme to calculate temperature and salinity dependent air-water transfer velocities for any gas,
   Ocean Sci., 6, 913–932, https://doi.org/10.5194/os-6-913-2010, 2010.

Kelley, M., Schmidt, G. A., Nazarenko, L. S., Bauer, S. E., Ruedy, R., Russell, G. L., Ackerman, A. S., Aleinov, I., Bauer, M.,
Bleck, R., Canuto, V., Cesana, G., Cheng, Y., Clune, T. L., Cook, B. I., Cruz, C. A., Genio, A. D. D., Elsaesser, G. S., Faluvegi,
G., Kiang, N. Y., Kim, D., Lacis, A. A., Leboissetier, A., LeGrande, A. N., Lo, K. K., Marshall, J., Matthews, E. E., McDermid,
S., Mezuman, K., Miller, R. L., Murray, L. T., Oinas, V., Orbe, C., García-Pando, C. P., Perlwitz, J. P., Puma, M. J., Rind, D.,
Romanou, A., Shindell, D. T., Sun, S., Tausnev, N., Tsigaridis, K., Tselioudis, G., Weng, E., Wu, J., and Yao, M.-S.: GISS-E2.1:
Configurations and Climatology, J. Adv. Model. Earth Syst., 12, e2019MS002025, https://doi.org/10.1029/2019MS002025, 2020.

Khan, M. A. H., Cooke, M. C., Utembe, S. R., Archibald, A. T., Maxwell, P., Morris, W. C., Xiao, P., Derwent, R. G., Jenkin, M.
 E., Percival, C. J., Walsh, R. C., Young, T. D. S., Simmonds, P. G., Nickless, G., O'Doherty, S., and Shallcross, D. E.: A study of global atmospheric budget and distribution of acetone using global atmospheric model STOCHEM-CRI, Atmos. Environ., 112, 269–277, https://doi.org/10.1016/j.atmosenv.2015.04.056, 2015.

- Koch, D., Jacob, D., Tegen, I., Rind, D., and Chin, M.: Tropospheric sulfur simulation and sulfate direct radiative forcing in the
   Goddard Institute for Space Studies general circulation model, J. Geophys. Res. Atmospheres, 104, 23799–23822,
   https://doi.org/10.1029/1999JD900248, 1999.
- Langford, B., Nemitz, E., House, E., Phillips, G. J., Famulari, D., Davison, B., Hopkins, J. R., Lewis, A. C., and Hewitt, C. N.:
   Fluxes and concentrations of volatile organic compounds above central London, UK, Atmospheric Chem. Phys., 10, 627–645,
   https://doi.org/10.5194/acp-10-627-2010, 2010.
- Lewis, A. C., Hopkins, J. R., Carpenter, L. J., Stanton, J., Read, K. A., and Pilling, M. J.: Sources and sinks of acetone, methanol,
   and acetaldehyde in North Atlantic marine air, Atmospheric Chem. Phys., 5, 1963–1974, https://doi.org/10.5194/acp-5-1963-2005,
   2005.
- Li, K., Li, J., Tong, S., Wang, W., Huang, R.-J., and Ge, M.: Characteristics of wintertime VOCs in suburban and urban Beijing:
   concentrations, emission ratios, and festival effects, Atmospheric Chem. Phys., 19, 8021–8036, https://doi.org/10.5194/acp-19-8021-2019, 2019.
- Liss, P. S. and Slater, P. G.: Flux of Gases across the Air-Sea Interface, Nature, 247, 181–184, https://doi.org/10.1038/247181a0,
   1974.

Marandino, C. A., Bruyn, W. J. D., Miller, S. D., Prather, M. J., and Saltzman, E. S.: Oceanic uptake and the global atmospheric acetone budget, Geophys. Res. Lett., 32, https://doi.org/10.1029/2005GL023285, 2005.

van Marle, M. J. E., Kloster, S., Magi, B. I., Marlon, J. R., Daniau, A.-L., Field, R. D., Arneth, A., Forrest, M., Hantson, S.,
 Kehrwald, N. M., Knorr, W., Lasslop, G., Li, F., Mangeon, S., Yue, C., Kaiser, J. W., and van der Werf, G. R.: Historic global
 biomass burning emissions for CMIP6 (BB4CMIP) based on merging satellite observations with proxies and fire models (1750–
 2015), Geosci. Model Dev., 10, 3329–3357, https://doi.org/10.5194/gmd-10-3329-2017, 2017.

 174
 Met Office, Hadley Centre: HadlSST 1.1 - Global sea-Ice coverage and SST (1870-Present), [Internet]. NCAS British

 175
 Atmospheric
 Data
 Centre
 2006, April
 3,
 2021.
 Available
 from

 176
 http://badc.nerc.ac.uk/view/badc.nerc.ac.uk
 ATOM
 dataent
 hadisst, 2006.

Neu, J. L., Prather, M. J., and Penner, J. E.: Global atmospheric chemistry: Integrating over fractional cloud cover, J. Geophys.
 Res. Atmospheres, 112, 2006JD008007, https://doi.org/10.1029/2006JD008007, 2007.

O'Rourke, P. R., Smith, S. J., Mott, A., Ahsan, H., McDuffie, E. E., Crippa, M., Klimont, Z., McDonald, B., Wang, S., Nicholson,
 M. B., Feng, L., and Hoesly, R. M.: CEDS v\_2021\_04\_21 Release Emission Data, https://doi.org/10.5281/zenodo.4741285, 2021.

Read, K. A., Carpenter, L. J., Arnold, S. R., Beale, R., Nightingale, P. D., Hopkins, J. R., Lewis, A. C., Lee, J. D., Mendes, L., and
 Pickering, S. J.: Multiannual Observations of Acetone, Methanol, and Acetaldehyde in Remote Tropical Atlantic Air: Implications
 for Atmospheric OVOC Budgets and Oxidative Capacity, Environ. Sci. Technol., 46, 11028–11039,
 https://doi.org/10.1021/es302082p, 2012.

Riahi, K., van Vuuren, D. P., Kriegler, E., Edmonds, J., O'Neill, B. C., Fujimori, S., Bauer, N., Calvin, K., Dellink, R., Fricko, O.,
Lutz, W., Popp, A., Cuaresma, J. C., Kc, S., Leimbach, M., Jiang, L., Kram, T., Rao, S., Emmerling, J., Ebi, K., Hasegawa, T.,
Havlik, P., Humpenöder, F., Da Silva, L. A., Smith, S., Stehfest, E., Bosetti, V., Eom, J., Gernaat, D., Masui, T., Rogelj, J., Strefler,
J., Drouet, L., Krey, V., Luderer, G., Harmsen, M., Takahashi, K., Baumstark, L., Doelman, J. C., Kainuma, M., Klimont, Z.,
Marangoni, G., Lotze-Campen, H., Obersteiner, M., Tabeau, A., and Tavoni, M.: The Shared Socioeconomic Pathways and their
energy, land use, and greenhouse gas emissions implications: An overview, Glob. Environ. Change, 42, 153–168,
https://doi.org/10.1016/j.gloenvcha.2016.05.009, 2017.

Sander, R.: Compilation of Henry's Law Constants for Inorganic and Organic Species of Potential Importance in Environmental
 Chemistry, 1999.

Sander, S. P., J Abbatt, J. R. Barker, J. B. Burkholder, R. R. Friedl, D. M. Golden, R. E. Huie, C. E. Kolb, M. J. Kurylo, G. K.
 Moortgat, V. L. Orkin, and P. H. Wine: Chemical Kinetics and Photochemical Data for Use in Atmospheric Studies Evaluation
 No. 17, JPL Publication 10-6, Jet Propulsion Laboratory, Pasadena, 2011.

197 Schade, G. W. and Goldstein, A. H.: Seasonal measurements of acetone and methanol: Abundances and implications for 198 atmospheric budgets, Glob. Biogeochem. Cycles, 20, https://doi.org/10.1029/2005GB002566, 2006.

Shindell, D. T., Grenfell, J. L., Rind, D., Grewe, V., and Price, C.: Chemistry-climate interactions in the Goddard Institute for
 Space Studies general circulation model: 1. Tropospheric chemistry model description and evaluation, J. Geophys. Res.
 Atmospheres, 106, 8047–8075, https://doi.org/10.1029/2000JD900704, 2001.

 Shindell, D. T., Faluvegi, G., and Bell, N.: Preindustrial-to-present-day radiative forcing by tropospheric ozone from improved simulations with the GISS chemistry-climate GCM, Atmospheric Chem. Phys., 3, 1675–1702, https://doi.org/10.5194/acp-3-1675-2003, 2003.

Singh, H., Chen, Y., Tabazadeh, A., Fukui, Y., Bey, I., Yantosca, R., Jacob, D., Arnold, F., Wohlfrom, K., Atlas, E., Flocke, F.,
 Blake, D., Blake, N., Heikes, B., Snow, J., Talbot, R., Gregory, G., Sachse, G., Vay, S., and Kondo, Y.: Distribution and fate of
 selected oxygenated organic species in the troposphere and lower stratosphere over the Atlantic, J. Geophys. Res. Atmospheres,
 105, 3795–3805, https://doi.org/10.1029/1999JD900779, 2000.

Singh, H. B., O'Hara, D., Herlth, D., Sachse, W., Blake, D. R., Bradshaw, J. D., Kanakidou, M., and Crutzen, P. J.: Acetone in the atmosphere: Distribution, sources, and sinks, J. Geophys. Res. Atmospheres, 99, 1805–1819, https://doi.org/10.1029/93JD00764, 1994.

- Singh, H. B., Tabazadeh, A., Evans, M. J., Field, B. D., Jacob, D. J., Sachse, G., Crawford, J. H., Shetter, R., and Brune, W. H.:
   Oxygenated volatile organic chemicals in the oceans: Inferences and implications based on atmospheric observations and air-sea
   exchange models, Geophys. Res. Lett., 30, https://doi.org/10.1029/2003GL017933, 2003.
- )15 Singh, H. B., Salas, L. J., Chatfield, R. B., Czech, E., Fried, A., Walega, J., Evans, M. J., Field, B. D., Jacob, D. J., Blake, D.,
- 116 Heikes, B., Talbot, R., Sachse, G., Crawford, J. H., Avery, M. A., Sandholm, S., and Fuelberg, H.: Analysis of the atmospheric
- distribution, sources, and sinks of oxygenated volatile organic chemicals based on measurements over the Pacific during TRACE P, J. Geophys. Res. Atmospheres, 109, https://doi.org/10.1029/2003JD003883, 2004.
- r, J. Geophys. Res. Aunospheres, 109, https://doi.org/10.1029/2005JD005865, 2004.
- Solberg, S., Dye, C., Schmidbauer, N., Herzog, A., and Gehrig, R.: Carbonyls and nonmethane hydrocarbons at rural European sites from the mediterranean to the arctic, J. Atmospheric Chem., 25, 33–66, https://doi.org/10.1007/BF00053285, 1996.
- 121 Taylor, K., Williamson, D., and Zwiers, F.: The sea surface temperature and sea ice concentration boundary conditions for AMIP
   122 II simulations, <u>PCMDI Report 60</u>, Program for Climate Model Diagnosis and Intercomparison, Lawrence Livermore National
   123 Laboratory, 2000.
- Thompson, C. R., Wofsy, S. C., Prather, M. J., Newman, P. A., Hanisco, T. F., Ryerson, T. B., Fahey, D. W., Apel, E. C., Brock, )24 )25 C. A., Brune, W. H., Froyd, K., Katich, J. M., Nicely, J. M., Peischl, J., Ray, E., Veres, P. R., Wang, S., Allen, H. M., Asher, E., )26 Bian, H., Blake, D., Bourgeois, I., Budney, J., Bui, T. P., Butler, A., Campuzano-Jost, P., Chang, C., Chin, M., Commane, R., Correa, G., Crounse, J. D., Daube, B., Dibb, J. E., DiGangi, J. P., Diskin, G. S., Dollner, M., Elkins, J. W., Fiore, A. M., Flynn, C. )27 )28 M., Guo, H., Hall, S. R., Hannun, R. A., Hills, A., Hintsa, E. J., Hodzic, A., Hornbrook, R. S., Huey, L. G., Jimenez, J. L., Keeling, 129 R. F., Kim, M. J., Kupe, A., Lacey, F., Lait, L. R., Lamarque, J.-F., Liu, J., McKain, K., Meinardi, S., Miller, D. O., Montzka, S. )30 A., Moore, F. L., Morgan, E. J., Murphy, D. M., Murray, L. T., Nault, B. A., Neuman, J. A., Nguyen, L., Gonzalez, Y., Rollins, )31 A., Rosenlof, K., Sargent, M., Schill, G., Schwarz, J. P., Clair, J. M. S., Steenrod, S. D., Stephens, B. B., Strahan, S. E., Strode, S. A., Sweeney, C., Thames, A. B., Ullmann, K., Wagner, N., Weber, R., Weinzierl, B., Wennberg, P. O., Williamson, C. J., Wolfe, )32 133 G. M., and Zeng, L.: The NASA Atmospheric Tomography (ATom) Mission: Imaging the Chemistry of the Global Atmosphere, )34 Bull. Am. Meteorol. Soc., 103, E761-E790, https://doi.org/10.1175/BAMS-D-20-0315.1, 2022.
- Tsigaridis, K. and Kanakidou, M.: Global modelling of secondary organic aerosol in the troposphere: a sensitivity analysis, Atmos
   Chem Phys, 2003.
- Wang, S., Apel, E. C., Schwantes, R. H., Bates, K. H., Jacob, D. J., Fischer, E. V., Hornbrook, R. S., Hills, A. J., Emmons, L. K.,
  Pan, L. L., Honomichl, S., Tilmes, S., Lamarque, J.-F., Yang, M., Marandino, C. A., Saltzman, E. S., Bruyn, W. de, Kameyama,
  S., Tanimoto, H., Omori, Y., Hall, S. R., Ullmann, K., Ryerson, T. B., Thompson, C. R., Peischl, J., Daube, B. C., Commane, R.,
  McKain, K., Sweeney, C., Thames, A. B., Miller, D. O., Brune, W. H., Diskin, G. S., DiGangi, J. P., and Wofsy, S. C.: Global
  Mtmospheric Budget of Acetone: Air-Sea Exchange and the Contribution to Hydroxyl Radicals, J. Geophys. Res. Atmospheres,
  125, e2020JD032553, https://doi.org/10.1029/2020JD032553, 2020.
- Warneke, C. and de Gouw, J. A.: Organic trace gas composition of the marine boundary layer over the northwest Indian Ocean in
   April 2000, Atmos. Environ., 35, 5923–5933, https://doi.org/10.1016/S1352-2310(01)00384-3, 2001.
- Weimer, M., Schröter, J., Eckstein, J., Deetz, K., Neumaier, M., Fischbeck, G., Hu, L., Millet, D. B., Rieger, D., Vogel, H., Vogel,
   B., Reddmann, T., Kirner, O., Ruhnke, R., and Braesicke, P.: An emission module for ICON-ART 2.0: implementation and
   simulations of acetone, Geosci. Model Dev., 10, 2471–2494, https://doi.org/10.5194/gmd-10-2471-2017, 2017.
- Wesely, M. L. and Hicks, B. B.: Some Factors that Affect the Deposition Rates of Sulfur Dioxide and Similar Gases on Vegetation,
   J. Air Pollut. Control Assoc., 27, 1110–1116, https://doi.org/10.1080/00022470.1977.10470534, 1977.
- Yoshino, A., Nakashima, Y., Miyazaki, K., Kato, S., Suthawaree, J., Shimo, N., Matsunaga, S., Chatani, S., Apel, E., Greenberg,
   J., Guenther, A., Ueno, H., Sasaki, H., Hoshi, J., Yokota, H., Ishii, K., and Kajii, Y.: Air quality diagnosis from comprehensive
   observations of total OH reactivity and reactive trace species in urban central Tokyo, Atmos. Environ., 49, 51–59,
   https://doi.org/10.1016/j.atmosenv.2011.12.029, 2012.
- Yuan, B., Hu, W. W., Shao, M., Wang, M., Chen, W. T., Lu, S. H., Zeng, L. M., and Hu, M.: VOC emissions, evolutions and contributions to SOA formation at a receptor site in eastern China, Atmospheric Chem. Phys., 13, 8815–8832, https://doi.org/10.5194/acp-13-8815-2013, 2013.
- Zhou, X. and Mopper, K.: Apparent partition coefficients of 15 carbonyl compounds between air and seawater and between air
   and freshwater; implications for air-sea exchange, Environ. Sci. Technol., 24, 1864–1869, https://doi.org/10.1021/es00082a013,
   1990.

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