Response of soil nutrients and erodibility to slope aspect in the northern agro-pastoral ecotone, China

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Abstract

Soil erosion, considered a major environmental and social problem, leads to the loss of soil nutrients and the degradation of soil structure, impacts plant growth. However, data on the effects of land use changes caused by vegetation restoration on soil nutrients and erodibility at different slope aspects is limited. This study was conducted to detect the response of soil nutrients and erodibility of different slope aspects in a typical watershed of the northern agro-pastoral ecotone in China. The following indexes were used to determine the improvement of soil nutrients and erodibility through a weighted summation method: comprehensive soil nutrient index and comprehensive soil erodibility index. The results showed that the vegetation types with the highest comprehensive soil quality index (CSQI) on the western, northern, southern, and eastern slopes were *Pinus sylvestris* and *Astragalus melilotoides* (1.45), *Caragana korshinskii* and *Capillipedium parviflorum* (2.35), *Astragalus melilotoides* (4.78), and *Caragana korshinskii* and *Lespedeza bicolor* (5.00), respectively. Slope aspect had a significant effect on understory vegetation characteristics, soil nutrients, and soil erodibility. Understory vegetation and soil characteristics explained 50.86–74.56% of the total variance in soil nutrients and erodibility of slope aspect.

Mean weight diameter and total phosphorus were the main factors affecting CSQI on different slope aspects. Our study suggested the combinations of species, such as *C. korshinskii* and *L. bicolor*, were the best species to include on any slope aspect in regard to improving soil nutrients and soil erodibility.

Keywords: Slope aspect; Soil nutrients; Soil erodibility; Soil erosion; Vegetation restoration; Land use

1. Introduction

Soil erosion, considered a major environmental and social problem, leads to the loss of soil nutrients and the degradation of soil structure, influences the functional capacity of soils on a global scale (Singh and Panda, 2017; Wen et al., 2021). Vegetation restoration is an important method of ecological restoration that aims to control soil erosion and prevent soil degradation (Schmiedel et al., 2017; Zhang et al., 2021). Vegetation restoration can improve the soil structure and nutrients, which in turn promotes the restoration of soil quality and function (Guo et al., 2021; Li et al., 2017). Changes in land use due to vegetation restoration play an important role in improving the environment and ecosystem function, as well as improving soil quality and soil nutrient cycling (Akiyama and Kawamura, 2007; Singh and Gupta, 2018).

Previous studies have shown that the plants selected for vegetation restoration projects drive land use change and alter soil properties, thus affecting soil erodibility (Wang et al., 2019b, a; Zhang et al., 2019). Many studies have also elucidated the influences of land use change on soil nutrients and have confirmed that revegetation is an effective way to enhance soil nutrients (Huang et al., 2020; Li et al., 2020; Yang et al., 2021; Zhu et al., 2020). Most studies have only focused on one aspect; thus, they lack comprehensive consideration and evaluation of the impact of land use changes caused by vegetation restoration on soil nutrients and erodibility. However, it is not clear which plants selected for restoration are the most effective in enhancing soil nutrients and reducing soil erodibility. The lack of a comprehensive understanding prevents us from gaining the best ecological benefits from vegetation restoration. Therefore, studies must be conducted on the response of soil nutrients and erodibility to different vegetation restoration types.

Soil erodibility is the sensitivity of the soil surface to erosion processes (Batista et al., 2023; Bryan et al., 1989). It is a necessary parameter for establishing soil loss equations and erosion models. There is currently no soil erosion model that can accurately predict soil erosion, although there are many related models (de Vente et al., 2013, 2008). At present, the soil erodibility K-factor, as defined in the general soil loss equation (USLE), is the most widely used measure (Wischmeier and Smith, 1978). In addition to K, other soil indexes have been adopted, including saturated hydraulic conductivity (SHC), soil disintegration rate (SDR), mean weight diameter (MWD), soil structural stability index (SSSI), clay ratio (CR), and soil organic carbon cementing agent index (SCAI), to quantify soil erodibility (Dong et al., 2022a; Guo et al., 2021; Wang et al., 2018; Zhang et al., 2019). Soil organic carbon, nitrogen, and phosphorus as well as their stoichiometry is also essential for assessing soil quality as well as ecosystem productivity and functionality (Borchard et al., 2017; Li et al., 2020; Masciandaro and Ceccanti, 1999; Schloter et al., 2003). A single index cannot fully reflect all soil properties; therefore, it is necessary to develop a comprehensive soil index using several related indicators.

In addition to soil properties, topographic factors also significantly affect soil nutrients and erodibility (Bangroo et al., 2017; Nabiollahi et al., 2018; Qin et al., 2016; Zhang et al., 2018). Slope aspect can affect the growth of plants due to a combination of factors, such as light, temperature, wind speed, and precipitation, which can cause significant changes in the ecological relationship between plants and the environment (Li et al., 2018; Tamene et al., 2020; Zhang et al., 2020). This is especially true for harsh climates such as cold, dry alpine regions in the north, in which plants are more sensitive to environmental changes. However, the optimal vegetation restoration type has primarily been studied by slope gradient and slope position (Dong et al., 2022a; Guo et al., 2021; Wen et al., 2021). There is a lack of systematic evaluation of the effects of land use changes caused by vegetation restoration on soil nutrients and erodibility on different slope aspects. Therefore, the classification of slope aspect needs to be further refined to elucidate the response of different slope aspects to changes in soil nutrients and erodibility caused by revegetation.

The ecologically fragile northern agro-pastoral zone in China is located in an erosion zone affected by both wind and water; soil erosion in this zone is considered very serious (Guo et al., 2019). Recently, the Chinese government has planned and carried out a series of ecological restoration projects in this region, including the Beijing-Tianjin Wind and Sand Source Control Project, the Beijing-Hebei Water Protection Forest Project, and the Sebei Forest Plantation Afforestation Project. These ecological restoration projects have effectively reduced land erosion and desertification, and have significantly delayed the onslaught of wind and sand (Wang et al., 2021b; Zeng et al., 2014; Zhang et al., 2017). However, the method used for afforestation, which mainly consists of plantations, is affected by differences in water, heat, wind, and sand in the different habitats, making it difficult to achieve vegetation restoration in some ecologically fragile areas, and the selection of suitable tree species is still equivocal.

Based on the abovementioned scientific gaps, we hypothesize that both slope aspect and land use types can significantly alter soil structure and properties to influence soil nutrients and erodibility under vegetation restoration. We further hypothesize that the western slope may have the lowest comprehensive soil quality index compared to other slope aspects. Therefore, we selected four slope aspects (west, north, south, and east) that have four different land use types (degraded land, grasslands, shrublands and woodlands) in a typical watershed of the northern agro-pastoral ecotone with three specific purposes: 1) to determine the impact of different vegetation types on different slope aspects on soil nutrient improvement and soil erodibility enhancement; 2) to determine the key influencing factors affecting soil nutrients and erodibility of the four slope aspects; and 3) to provide optimal revegetation models for improving soil nutrients and reducing soil erodibility on different slope aspects.

2. Materials and Methods

2.1. Study area

This study was conducted in the Yangcaogou Watershed (41°4′~41°8′ N, 114°58′~115°2′ E; Fig.1), Chongli District, Zhangjiakou City, Heibei Province, China. The watershed is located in a typical ecological transition zone of the agro-pastoral ecotone in northern China (Wu et al., 2023). The study site spans an area of 10.6 km^2 with an altitude ranging from 1084 to 1575 m. It belongs to a typical temperate continental monsoon semi-arid climate with an annual average temperature of 3.5 ℃. The average annual rainfall is 401.6 mm. The rainy season occurs from June to September (Chang et al., 2021; Guo et al., 2019). The main soil type is classified as chestnut soil in both the Chinese Soil Taxonomy and the World Reference Base for Soil Resources (Schad, 2017). Most of the study area consists of Proterozoic soil rock formations. Owing to irrational human reclamation and grazing, there is very serious soil and gully erosion. Over the past decade, due to the implementation of the Beijing–Tianjin Sandstorm Source Control Project, soil erosion and desertification has been effectively mitigated (Wang et al., 2020b). However, native plant populations have been diminished and instead the area is planted with trees, shrubs, and herbs.

2.2. Selection of sites and determination of slope aspect

The study was conducted during the 2021 growing season. A comprehensive field survey was conducted on the dominant plant species and soil properties of each of the following land use types: degraded land, grasslands, shrublands, and woodlands in the Yangcaogou watershed. Grasslands, shrublands, and woodlands were restored from degraded land over the past 12 years. The degraded land (loss of soil material from wind and water erosion, degradation of physical, chemical and biological properties of soil) was previously degraded cropland. All land use types were vegetated and restored in the form of engineering measures such as fish scale pits (Wang et al., 2014b) and parallel ditches (Barua and Alam, 2013).

In addition to the degraded land, the other three land use types were all sampled along complete slope aspects at the E, W, N, and S slopes. It includes 28 sample sites (20 m × 20 m) of a degraded land, two grasslands, two shrublands and two woodlands on each slope aspect. Three sampling quadrats (1 m × 1 m) were set up in each sample site to investigate and record the species, height, richness, coverage, aboveground biomass, belowground biomass, and litter biomass of herbs. Height was measured as the average height of herbs in the sample. Biomass coverage was determined following the visual method (Proulx and Mazumder, 1998). Richness was calculated by measuring the number of individuals of each herb in the quadrat and calculating the percentage of its occurrence (Dou et al., 2023). Belowground biomass and soil samples were collected with a 9 cm diameter soil drill. The measured land use types, major plant species, and understory vegetation characteristics at each selected field site are listed in Table S1.

Following the methods described by (Yimer et al., 2006), study sites were selected that included the four land use types on each of the four slope aspects: east, west, north, and south. Eastern, western, northern, and southern slopes are also known as semi-sunny, semi-shady, shady, and sunny slopes (Che et al., 2022; Chen et al., 2021b). In this region, four unrestored degraded land were selected as representatives from the western slope. The slope gradients and positions were similar for all selected sample sites (Fig. 1).

2.3. Soil sampling and analysis

Three quadrants were selected at each site to investigate vegetation and collect soil samples. For each sampling point, a steel cutting ring (100 cm^3) was used to obtain 75 soil samples (25 sites × three sampling points). The saturated hydraulic conductivity of the soil were evaluated using the constant head permeability test (Chandler and Chappell, 2008). The mean weight diameter was measured by screens with different pore sizes (0.25, 0.50, 1.00, 2.50 and 5.00 mm) (Campo et al., 2008). After air-drying via dry screening, 50 g of the soil samples were placed on the sieve of a soil aggregate analyzer (TTF-100 model, China), then completely immersed in water, and shaken up and down 30 times for 1 minute (Wang et al., 2014a). After shaking, samples were removed from the settling cylinder, and the remaining aggregates on each sieve were put into an aluminum box for drying. Finally, the samples were weighed and the dried aggregates were recorded.

Soil characteristics of different vegetation types at different slope aspects are listed in Table S2. Topsoil samples were collected from 0–10 cm using a cutting ring. Samples were brought back to the lab to oven-dried at 105℃ for 24 hours. Then, the soil bulk density (SBD) (Lardy et al., 2022; Moreira et al., 2020) and soil capillary porosity (SCP) (Singh and Pollard, 1958) were measured. In addition, 225 mixed soil samples (25 sites × three quadrats/site × three samples/quadrat) were collected as soil samples. Among them, the particle size distribution of clay content (Cl), silt content (Si), sand content (Sa) was determined by a Microtrac S3500 laser particle sizer (Malvern 3000, UK). Total nitrogen (TN) and total phosphorus (TP) were determined by the dichromate oxidation (Bremner, 1996) and $HClO_4$-H_2SO_4 methods (Kisand, 2005), respectively. Soil pH (Cornfield, 1954) was determined using a pH meter at a 2.5 soil:1 water ratio.

2.4 Calculation of soil indexes

Saturated hydraulic conductivity of the soil (K_S) (Campo et al., 2008), mean weight diameter (MWD) (Ortas and Lal, 2012), soil disintegration rate (SDR) (Guo et al., 2021), soil structure stability index (SSSI) (Nichols and Toro, 2011), soil organic carbon cementing agent index (SCAI) (Dong et al., 2022a) and K factor (Jiang et al., 2020; Li et al., 2012) were used to express the soil erodibility. These indexes were calculated using equations (1) - (5):

$$K_S = \frac{QL}{Aht} \tag{1}$$

where Q is the outflow volume (ml), A is the soil column section (mm^2), t is the time (min), h is the head difference (mm), and L is the height of the soil column (mm).

$$MWD = \sum_{i=1}^{n} (w_i/m_t)d_i \tag{2}$$

Where w_i is the mass of the i-th level of aggregates or other soil material (g), m_t is the sample mass, and d_i is the mean diameter of the i-th level of aggregates or other soil material (mm).

$$SDR = \frac{M_1 - M_2}{t_2 - t_1} \times 100\% \tag{3}$$

Where M_1 and M_2 are the weight of the soil before (t_1) and after (t_2) disintegration, respectively.

$$\mathrm{SSSI} = 100\% \times \frac{\mathrm{SOMC}}{\mathrm{Cl+Si}} \tag{4}$$

$$\mathrm{K} = \left\{0.2 + 0.3\exp\left[-0.0256\mathrm{Sa}\left(1 - \frac{\mathrm{Si}}{100}\right)\right]\right\}\left(\frac{\mathrm{Si}}{\mathrm{Cl+Si}}\right)^{0.3} \times \left(1 - \frac{0.25\mathrm{C}}{\mathrm{C} + \exp(3.72 - 2.95\mathrm{C})}\right)\left(1.0 - \frac{0.7\mathrm{SN1}}{\mathrm{SN1} + \exp(-5.51 + 22.9\mathrm{SN1})}\right) \tag{5}$$

Where SOMC is the content of soil organic matter (Kar et al., 2023), C = 0.583 × SOMC; Cl and Si represent the clay and silt content (%), respectively; SN1 = 1-Sa/100; K represents the soil loss rate per unit area under rainfall erosivity conditions for a specified soil on a standard plot (Jiang et al., 2020; Renard et al., 1997). A previous study indicates the rationality and validity of estimating K in the Zhangjiakou region using this model (Wang et al., 2020a).

In order to further evaluate soil nutrients and erodibility, comprehensive soil nutrient and erodibility index were calculated using equations 6 and 7, respectively:

$$\mathrm{CSNI} = \sum_{i}^{n} K_{ni} \cdot C_{ni} \tag{6}$$

$$\mathrm{CSEI} = \sum_{i}^{n} K_{ei} \cdot C_{ei} \tag{7}$$

Where K_{ni} and C_{ni} are the weight and score of soil nutrient index respectively, K_{ei} and C_{ei} are the weight and score of soil erodibility index respectively, and n is the number of indexes.

The weight of each soil nutrient index and soil erodibility index was determined using a principal component analysis (PCA) (Pandey et al., 2021; Wang et al., 2018). The scores of SHC, MWD, SSSI, SOC, TN, and TP scores were calculated using a "reverse S" function, which was calculated using equations 8.

$$f(x) = \begin{cases} 1 & ,x \geq b \\ \frac{x-a}{b-a} & ,a < x < b \\ 0 & ,x \leq a \end{cases} \quad (8)$$

The SDR and K factor scores were calculated by "S" function, as shown in equations 9.

$$f(x) = \begin{cases} 1 & ,x \leq b \\ \frac{x-a}{b-a} & ,a > x > b \\ 0 & ,x \geq a \end{cases} \quad (9)$$

Comprehensive soil quality index (CSQI) is used to express soil quality, which takes into account both soil nutrients and erodibility (De Laurentiis et al., 2019; Dong et al., 2022b). The CSQI was calculated as follows (Eq. 10):

$$CSQI = \frac{CSNI}{CSEI} \quad (10)$$

where CSQI (＞0), CSNI (0-1) and CSEI (0-1) are the comprehensive soil quality, nutrient, and erodibility indexes, respectively.

2.5. Statistical analysis

SPSS Ver. 20 software were used for data processing and statistical analysis, and ArcGIS 10.4.1 and Origin 2021 were used for graphing. A one-way analysis of variance (ANOVA) was used to compare soil nutrient and erodibility indexes of different slope aspects and different land use types. The effects of land use types, slope aspects and their interaction on soil nutrients and erodibility indexes were tested using a two-way ANOVA. Pearson's correlation coefficient was used to determine the correlation between soil nutrient, erodibility, and quality indexes and their influencing factors. The contributions of understory vegetation and soil characteristics to total variance in soil nutrients and erodibility indicators were determined using a redundancy analysis (RDA) (Capblancq et al., 2018; Peres-Neto et al., 2006). A random forest algorithm based on R software was used to analyze the importance of impact factors from different slope aspects (Schonlau and Zou, 2020; Vincenzi et al., 2011). The importance index was determined as the average accuracy reduction. When the importance index is higher, it means that the corresponding factor holds more weight (Chen et al., 2021a; Hao et al., 2015).

3. Results

3.1. Changes in the characteristics of understory vegetation on different slope

aspects

Slope aspect significantly influenced some of the characteristics of understory vegetation such as aboveground biomass (AGB) and belowground biomass (BGB). All measured characteristics of understory vegetation on the western slope were lower than that of other three slope aspects. AGB and BGB was significantly lower for the western slope than the eastern slope (Fig. 2). AGB and BGB on the eastern slope were significantly higher than those on the western slope by 63.40% and 78.40%, respectively (Fig. 2d, e). The measured plant characteristics from the eastern and western slopes were not significantly different from those on the northern and southern slopes. There were significant differences among the four land use types for all characteristics measured for the western slope (Table S1). BH, R, and AGB of understory vegetation were significantly higher for the woodland than for the other three land use types (Fig. 2). Overall, shrubland had the highest litter biomass on each slope aspect, while degraded land on the western slope had the lowest.

3.2. Changes in soil nutrients on different slope aspects

Slope aspect significantly affected soil nutrients. Soil organic carbon (SOC), total nitrogen (TN), and total phosphorus (TP) were significantly lower in soil collected from the western slope than the eastern slope (Fig. 2). SOC of the eastern slope was 0.96–1.38 times greater than that of other slopes, respectively (Fig. 2g). TN was highest on the eastern slope and was 0.39 g kg^{-1} and 0.28 g kg^{-1} greater than that on the western and northern slopes, respectively (Fig. 2h). Similarly, the TP of the eastern slope was significantly greater than that of the southern and eastern slopes by 59.60% and 17.37%, respectively (Fig. 2i). When all slope aspects were considered, comprehensive soil nutrient index (CSNI) was significantly lower on the western slope than on the other three slope aspects. The highest CSNI was found for both southern slope (0.81) and eastern slope (0.86) (Fig. 3). For a given slope aspect, land use types also significantly influenced soil nutrients (Fig. S1). For exemple, on the western slope, the SOC of forested land was significantly higher than other restored land uses by 11.81–150.84% depending on the comparison. SOC, TN, and TP of degraded land were significantly lower than that of other land use types. CSNI was influenced by land use type, slope aspect, and their interactions (Table 1). Compared to degraded land, CSNI was significantly higher for all three land uses, with the greatest increase in CSNI for shrubland (0.75), followed by woodland and grassland (Fig. 4).

3.3. Changes in soil erodibility under vegetation restoration

The effect of slope aspect on soil erodibility indicators was significant (Table 1 and 2). Among the four slope aspects, SHC of the soil collected from the eastern slope was the greatest, and was significantly greater than that of the western and northern slopes by 311.16% and 187.10%, respectively. MWD was highest on the eastern slope (3.65 mm), followed by the southern and northern slopes. MWD among the four slopes was significantly different. SSSI of the western slope was the lowest (0.41 g kg^{-1}), and it was significantly lower than the other three slope aspects. In contrast, the highest SCAI was found on the western slope, and it was significantly higher than the other slope aspects by 46.10%–59.70%, respectively. When all slope aspects were considered, the southern (0.26) and eastern (0.20) slopes had the highest comprehensive soil erodibility index (CSEI) reduction capacity (Fig. 3). For any given slope aspect, land use types also greatly influenced soil erodibility indicators (Table 2). On the western slope, MWD was significantly increased by 0.67 mm–1.59 mm. On the northern slope, the SHC of woodland was significantly higher than that of shrubland (by 117.67%) and grassland (by 94.24%), respectively. On the southern slope, the K in the grassland land use type was significantly lower than that in woodland and shrubland. On the eastern slope, soil disintegration rates of the three restored land uses were significantly different, with the highest SDR in the woodlands. CSEI was influenced by land use type, slope aspect, and their interactions (Table 1). The CSEI of all three restored land uses was significantly lower by (63.01–64.70%) compared to the degraded land (Fig. 4).

3.4. Changes in comprehensive soil quality index under vegetation restoration

When all slope aspects are considered, there were significant differences in comprehensive soil quality index (CSQI), with the eastern slope (2.46) having the greatest capacity to increase CSQI (Fig. 3). Compared to degraded land, the CSQI of grassland, shrubland and woodland increased significantly by 2.51, 2.65, and 2.44, respectively (Fig. 4). CSQI was influenced by land use type, slope aspect, and their interactions (Table 1).

The differences in CSQI of different vegetation types were compared to determine the optimal vegetation restoration type for different slope aspects. On the western slope, the WGCP grassland (*Capillipedium parviflorum*) and WWPS woodland (*Pinus sylvestris* and *Astragalus melilotoides*) had relatively high CSQIs. They were significantly higher than that of other vegetation types (Fig. 5a). Therefore, these two plant communities may be selected for restoration practices on the western slope. On the northern slope, the CSQI of the shrubland (NSCK) was significantly higher and second highest in grassland (NGBI). The combination of *Caragana korshinskii* and *Capillipedium parviflorum* (NSCK) could also be selected as taxa for restoration vegetation (Fig. 5b). On the southern slope, the CSQI of grassland (SGAM) was significantly higher than that of other vegetation types (Fig. 5c). The SGAM was dominated by the herb *Astragalus melilotoides*, which had the highest CSQI. *A. melilotoides* could be selected for improving soil quality on the southern slope. On the eastern slope, the CSQI of the shrubland (ESCK) was relatively higher than that of other sites (Fig. 5d). The ESCK was dominated by *Caragana korshinskii* and *Lespedeza bicolor*,which had the highest CSQI. Therefore, these species should be selected for improving soil quality on the eastern slope.

3.5. Key factors and their contributions on different slope aspects

The RDA followed by Monte Carlo permutation tests revealed that the variations in the nine measured soil quality indicators were significantly influenced by understory vegetation and soil characteristics on the four slope aspects ($P < 0.01$, Fig. 6). On the western slope, 62.7% of the total variance can be explained by understory vegetation and soil characteristics (Fig. 6a), with understory vegetation and soil characteristics explaining 43.11% and 19.59% of the total variance, respectively. For the northern slope, the understory vegetation and soil characteristics contributed 50.86% of the total variance of soil quality (Fig. 6b), of which understory vegetation and soil characteristics accounted 33.28% and 17.58% of the total variance, respectively. On the southern slope, the total variance in soil quality of 54.23% could be explained by understory vegetation and soil characteristics, of which the combination of soil and roots contributed 44.56% and 9.67% of total variance, respectively (Fig. 6c). However, on the eastern slope, the understory vegetation and soil characteristics contributed 74.56% of the total variance of soil quality (Fig. 6d), of which understory vegetation and soil characteristics accounted for 56.81% and 17.59% of the total variance, respectively.

The random forest analysis highlighted the importance of 21 modeling factors to determine the restoration characteristics of understory vegetation and the physical and chemical characteristics of topsoil on different slope aspects. MWD, TP, saturated hydraulic conductivity (SHC), and soil disintegration rate (SDR) were the main factors influencing understory vegetation and soil properties on different slope aspects.

The mean accuracy reduction was calculated using the random forest method. Using this calculation, we obtained an MWD of 13.40, TP of 13.30, SHC of 12.60, and SDR of 8.20 (Fig. S2).

4. Discussion

4.1. Effects of slope aspect on understory vegetation characteristics

Although slope aspect, one of the most important topographic factors, may impacts vegetation characteristics due to differences in sunlight, moisture, temperature, and soil, our results showed that most of the characteristics of understory vegetation had no significant differences based on the different slope aspects. This may be due to the fact that the understory plants were shaded by the taller trees and shrubs (Niinemets, 2010). Aboveground biomass was greater on the eastern and southern slopes than on the northern and western slopes. Vegetation density was lowest on the western slope. These findings indicated that aboveground biomass is closely related to sunshine hours. Sunshine hours affect the balance of heat and water (Chen et al., 2021b; Shi et al., 2021). This contributed to the low aboveground biomass of the western slope. Similarly, belowground biomass declined from the eastern, southern, northern, and western slopes. This may be due to the difference in the aboveground biomass of the four slope aspects. Aboveground biomass impacts belowground biomass (Sun et al., 2022), and the belowground biomass was significantly lower on the western slope than on the eastern slope.

4.2. Effects of slope aspect on soil nutrients

Our results show that the conditions related to slope aspect have significant effects on single soil nutrient indicators and the comprehensive soil nutrient index (Figs. 2, 5). In the same area, soil nutrients can vary depending on the slope aspect (Li et al., 2021; Sharma et al., 2010). On different slope aspects, TN, TP, and the comprehensive soil nutrient index of surface soil were highest on the eastern and southern slopes, while the soil organic carbon content was highest on the northern slope. Plants need to absorb a large amount of fast-acting nitrogen and phosphorus during vegetative growth, and the nutrients required for plant growth are converted from organic matter in the soil. The lowest SOC, TN, TP, and the comprehensive soil nutrient index on the western slope are due to the fact that it was located in the wind–water erosion zone of the northern agro-pastoral ecotone, and the topsoil has been lost due to long-term wind erosion.

The effect of different slope aspect conditions on soil pH was limited. This is because plant root systems and sediments were not abundant in the case of vegetation restoration of just 12a (Bai et al., 2020). The organic acid content was low when combined with organic matter during decomposition and vegetation restoration; therefore, it was insufficient to lower the pH of the surface soil (Seddaiu et al., 2013).

4.3. Effects of slope aspect on soil erodibility

Our results show that slope aspect has a significant effect on single soil erodibility indexes as well as comprehensive soil erodibility index. In general, soil erodibility decreases from the western slope to the eastern slope (Table 2), a pattern that may be related to the geographical location, altitude, temperature, and semi-arid climate of the region. Due to special location, the western and northern slopes are susceptible to year-round gales from the northwestern interior and Siberia, resulting in varying environmental conditions. However, the soil water content of the northern slope (shaded slope) is higher than that of the western slope, which may be more favorable for vegetation restoration on the northern slope (Liu et al., 2020); the western slope may be more vulnerable to erosion. Wind speed and soil moisture are key factors controlling the process of vegetation restoration (Hupet and Vanclooster, 2002; Meng et al., 2018), and these factors further influence soil erodibility (Sun et al., 2016).

4.4. Relationship between soil nutrients and soil erodibility

The comprehensive soil nutrient index was significantly positively correlated with saturated hydraulic conductivity, mean weight diameter, and soil structure stability index (Fig. 7), while the comprehensive soil nutrient index was highly significantly negatively correlated with the comprehensive soil erodibility index, with an explanation of 88% (Table S1). Many previous studies have reported similar results (Dong et al., 2022a; Zhu et al., 2018). In this study, higher saturated hydraulic conductivity, mean weight diameter, and soil structure stability index and lower soil disintegration rate, K, and SOC cementing agent index indicate better soil structure and lower soil erodibility. These characteristics can significantly reduce runoff and sediment loss, which can result in soil nutrient accumulation (Pan and Shangguan, 2006; Sun et al., 2015; Zheng et al., 2021). Therefore, revegetation increases soil nutrients and reduces soil erodibility, which further change vegetation and soil characteristics. In addition, these factors could reduce soil nutrient loss and further promote soil nutrient accumulation by reducing soil erodibility.

The comprehensive soil erodibility index was highly significantly negatively correlated with SOC, TN, and TP (Fig. 7). Previous studies have shown that soil organic matter and SOC are closely related to soil erodibility (Wang et al., 2019b). SOC acts as a cement for soil aggregation, which improves soil structural stability through the formation of aggregates, thus reducing soil erodibility. Soil nitrogen indirectly affects soil erodibility by promoting plant growth and development, increasing the accumulation of SOC in plants. In addition, nitrogen enrichment increased soil macroparticles and mean weight diameter, which directly affected soil erodibility. Similar to nitrogen, phosphorus is one of the essential elements for plant growth and development, and the phosphorus content of soil determines the development of soil microorganisms and root systems, which will further influence the input of soil organic carbon and the formation of soil aggregates.

4.5. Key factors impacting soil and vegetation related to slope aspect

The results derived from the random forest method showed that mean weight diameter and TP were the main influencing factors. The main adhesion agents for the formation of aggregates included clay content, SOC and cementation. The mean weight diameter was significantly and positively correlated with soil organic carbon and clay content. The magnitude of mean weight diameter affects soil structural stability and root establishment, which varies due to environmental factors on different slope aspects. Soil phosphorus is an important element necessary for plant growth and development, and rapid growth requires more soil phosphorus, so there were some differences between different land use types on different slope aspects. The difference in TP between slope aspect affected the amount of inorganic phosphorus available for uptake by plants, and the lower phosphorus content limited plant growth. By analyzing the main factors influencing surface soil quality in different slope aspects, timely application of phosphorus fertilizer in vegetation restoration projects could help accelerate the process of afforestation.

4.6. Optimal land use type and plant species based on slope aspect

Our study has shown that vegetation restoration can be an effective measure to improve soil nutrients and reduce soil erodibility. Moreover, the restored land use types and plant species to improve soil quality differed significantly depending on the slope aspect. Therefore, according to the differences in water, heat, wind, and sand on different slope aspects, the selection of land use and its corresponding vegetation types should be carefully considered. Our findings both agree with and differ from previous studies (Colgan et al., 2010; Dong et al., 2022a; Wang et al., 2021a). Studies that found contrasting results are likely due to the environmental conditions (e.g. climate, rainfall, topographic conditions, seed bank, soil texture) of the different slopes aspects. It is noteworthy that herbaceous vegetation on the western slope is prone to severe shallow nutrient loss and soil erosion because of strong wind conditions and sandy soil (Guo et al., 2020). Therefore, the use of herbaceous vegetation should be carefully considered as the primary restoration vegetation species. Fortunately, our proposal (*Caragana korshinskii* and *Lespedeza bicolor*) satisfied this requirement. In addition, wind also contributes to soil erosion in this region; however, limited research has been conducted on wind erosion and combined erosion by wind and water. Future studies should be conducted on combined erosion by wind and water study to better characterize soil erosion.

5. Conclusions

We found that some understory vegetation characteristics and soil properties varied significantly with slope aspect. Soil nutrients and erodibility reflected by soil organic carbon, total nitrogen, total phosphorus, saturated hydraulic conductivity, soil disintegration rate, mean weight diameter, soil structure stability index, soil erodibility factor, and soil organic carbon cementing agent index, respectively, were also influenced by slope aspect and land use. Furthermore, comprehensive soil nutrient, erodibility, and quality indexes also varied significantly with slope aspect, land use, and predominant plant species. Slope aspect strongly modified the relationship between comprehensive soil nutrient, erodibility, and quality indexes as well as understory vegetation characteristics and soil properties. Our study found that *Caragana korshinskii* and *Lespedeza bicolor* were the best taxa to include on any slope aspect to improve soil nutrients and prevent soil erosion. This study provides insight into the rational planning of vegetation restoration measures on all slope aspects in the northern agro-pastoral ecotone in semi-arid areas. Future work will focus on land degradation associated with soil erosion from water and storms in the region.

Date Availability

Data will be made available on request.

Author contributions.

Yuxin Wu: Writing-original draft. Guodong Jia: Project administration, Funding acquisition, Writing-review and editing. Xinxiao Yu: Project administration, Funding acquisition, Writing-review and editing. Honghong Rao: Methodology and Formal analysis. Xiuwen Peng: Investigation. Yusong Wang and Yushi Wang: Investigation.

Competing interests.

The author declares that the publication of this scientific paper has no conflict of interest.

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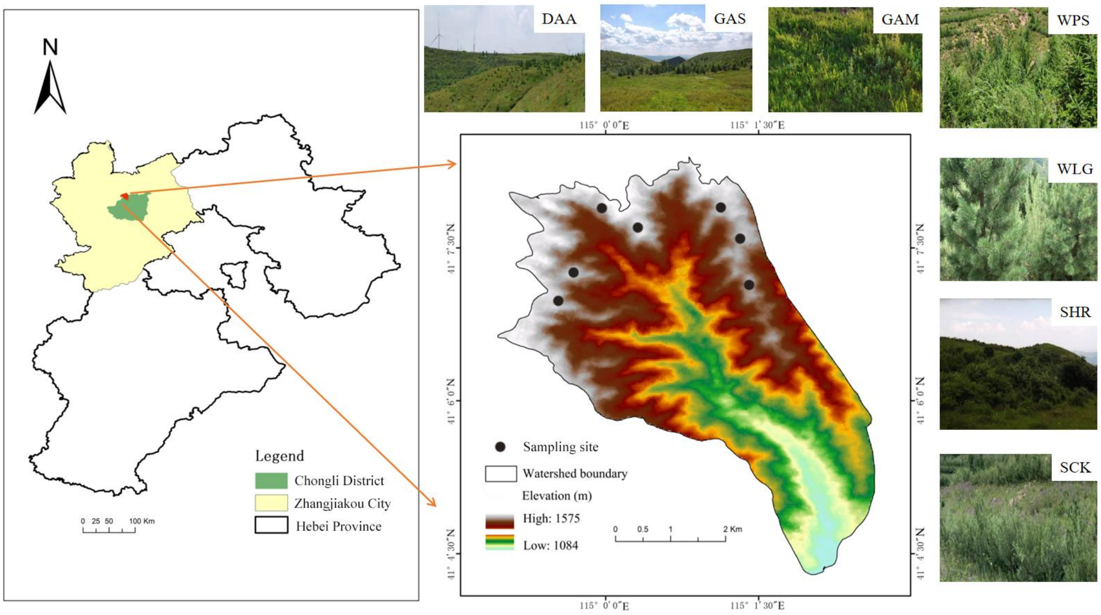

Fig. 1. Location map of the sampling points in the study area. The first letter: D, G, S and W represent degraded land, grassland, shrubland and woodland. The sampling sites from west to east were: DAA, degraded land; GAS, *Artemisia sacrorum*; GAM, *Astragalus melilotoides*; WPS, *Pinus sylvestris*; WLG, *Larix gmelinii*; SHR, *Hippophae rhamnoides*; SCK, *Caragana korshinskii*.

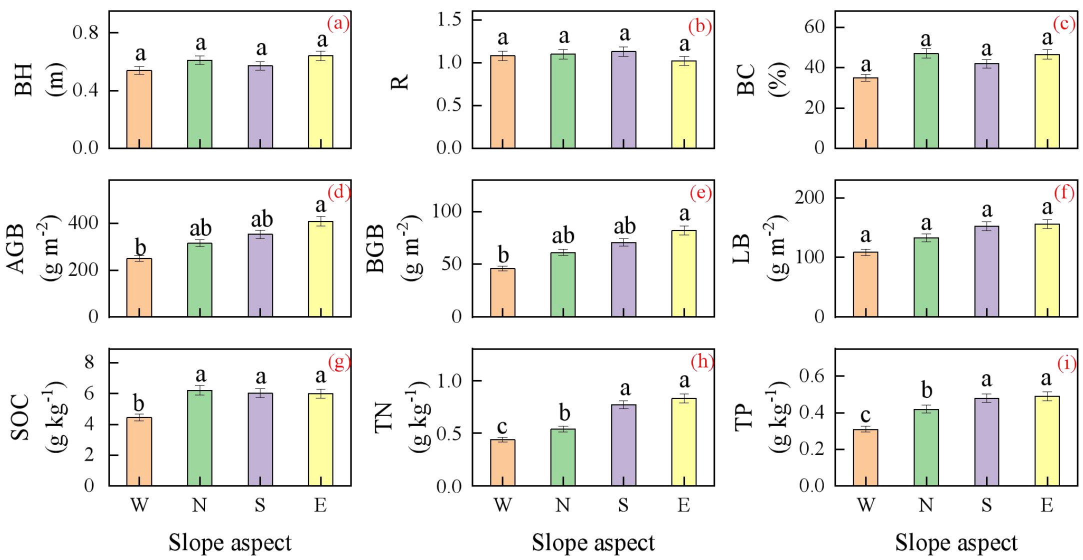

Fig. 2. Variation of understory vegetation characteristics and soil nutrients with slope aspects. BH, biomass height; R, richness; BC, biomass coverage; AGB, aboveground biomass; BGB, belowground biomass; LB, litter biomass; SOC, soil organic carbon; TN, total nitrogen; TP, total phosphorus; W, west; N, north; S, south; E, east. Different letters indicate significant differences among different seasons at P<0.05 level.

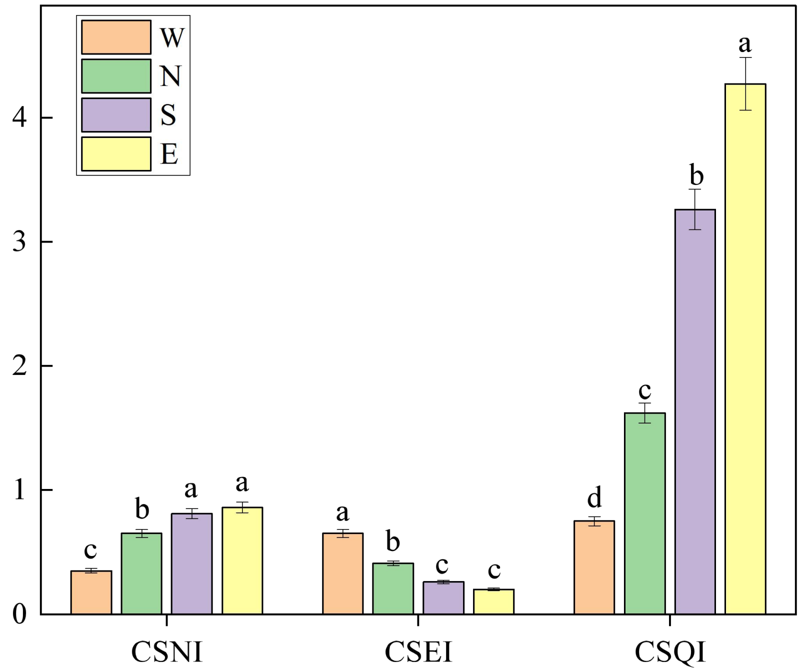

Fig.3. Variation of comprehensive soil nutrient, erodibility and quality index with slope aspects. CSNI, comprehensive soil nutrient index; CSEI, comprehensive soil erodibility index; CSQI, comprehensive soil quality index. Different letters indicate significant differences among different slope aspects at $P<0.05$ level.

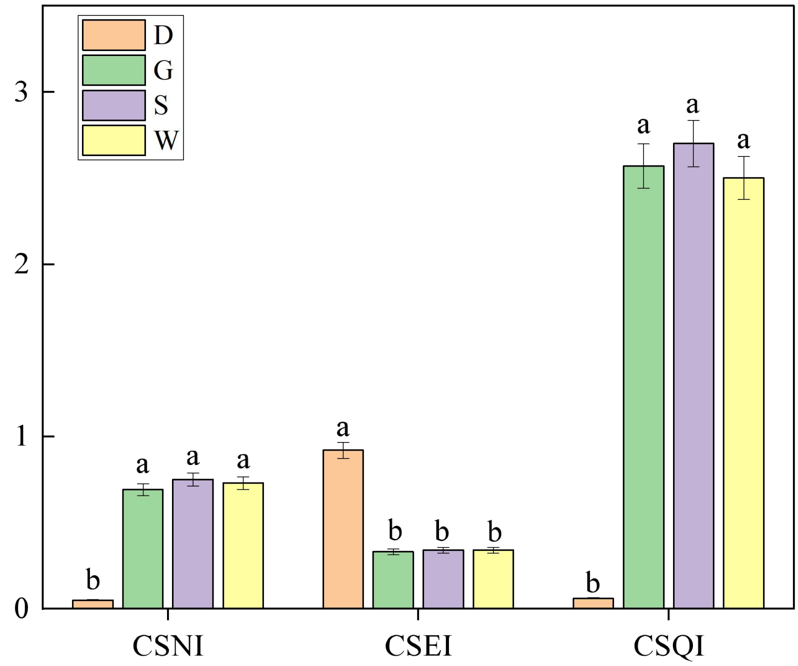

Fig. 4. Variation of comprehensive soil nutrient, erodibility and quality index with land use. Different letters indicate significant differences among different land use types at $P<0.05$ level.

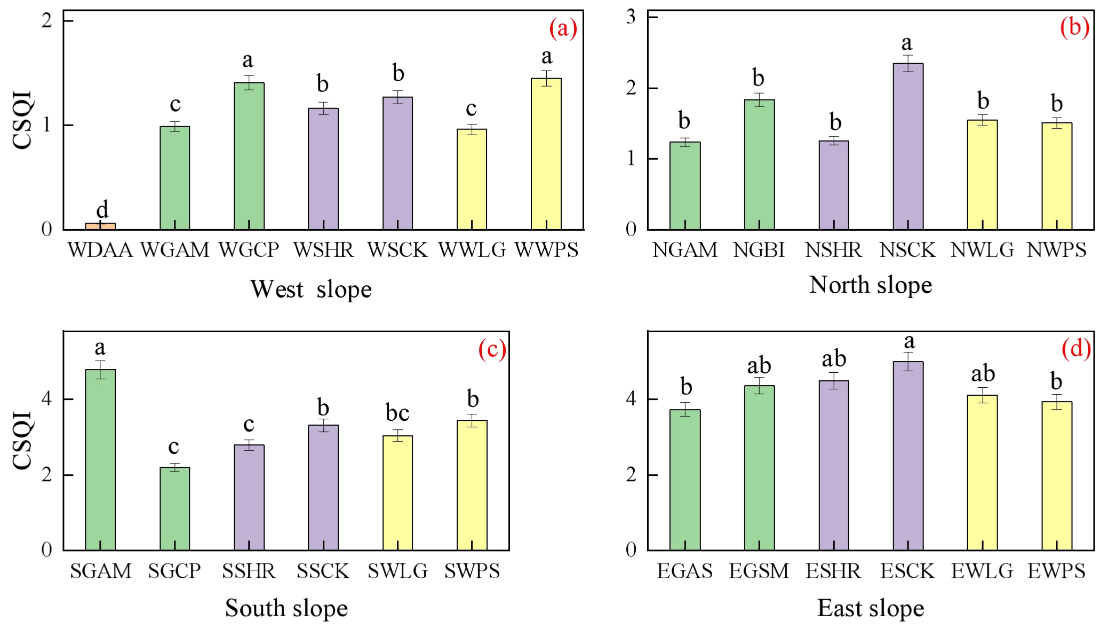

Fig. 5. Variation in comprehensive soil quality index with vegetation types along slope aspects. WDAA, *Artemisia annua*; WGAM, NGAM and SGAM, *Astragalus melilotoides*; NGBI, *Bothriochloa ischaemum*; EGSM, *Artemisia sacrorum*, *Astragalus melilotoides*; WGCP, NGCP and SGCP, *Capillipedium parviflorum*; WSHR, NSHR, SSHR and ESHR, *Hippophae rhamnoides*; WSCK, NSCK, SSCK and ESCK, *Caragana korshinskii*; WWLG, NSWG, SSWG and ESWG, *Larix gmelinii*; WWPS, NWPS, SWPS and EWPS, *Pinus sylvestris*. Different letters indicate significant differences among different seasons at P<0.05 level.

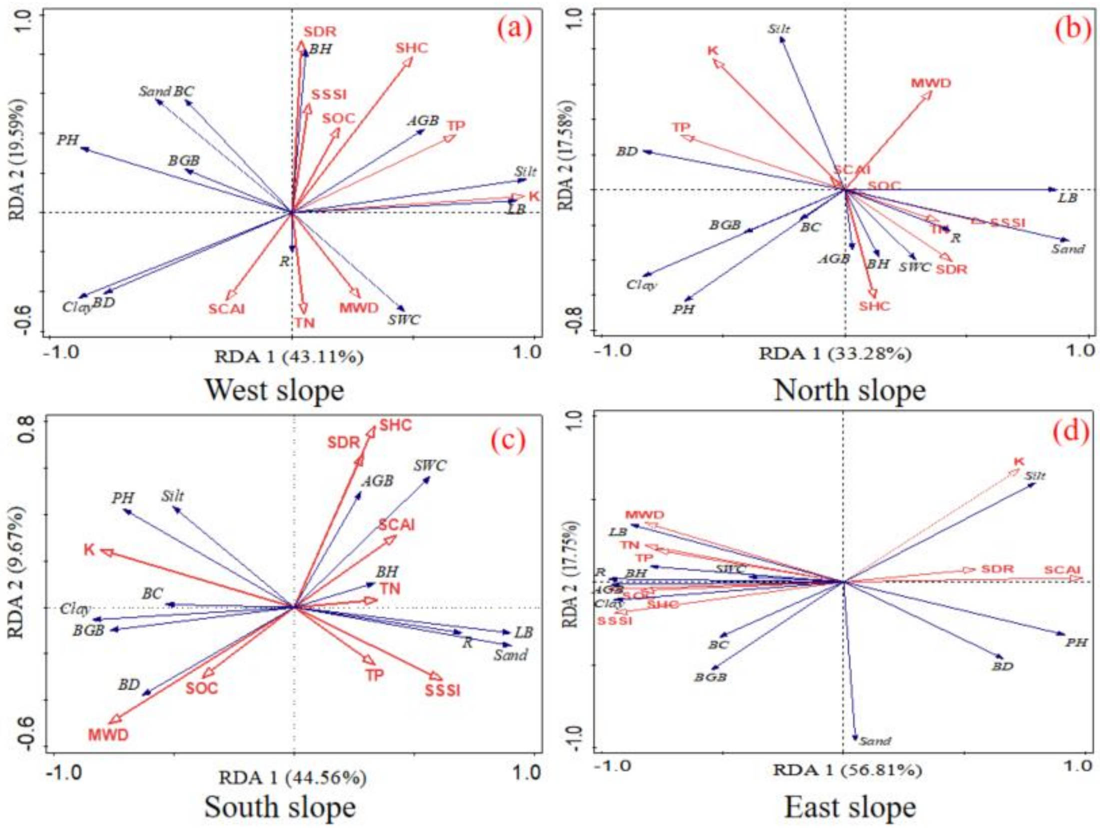

Fig. 6. Results of redundancy analysis (RDA) among soil quality parameters and characteristics of vegetation and soil on four slope aspects. BH: biome height; R: richness; BC: biome coverage; AGB: aboveground biomass; BGB: belowground biomass; LB: litter biomass; Sand: sand content; Silt: silt content; Clay: clay content; SWC: soil water content; SBD: soil bulk density; SOC: soil organic carbon; TN: total nitrogen; TP: total phosphorus; SHC, saturated hydraulic conductivity; SDR, soil disintegration rate; MWD, mean weight diameter; K, soil erodibility factor; SSSI, soil structure stability index; SCAI, SOC cementing agent index.

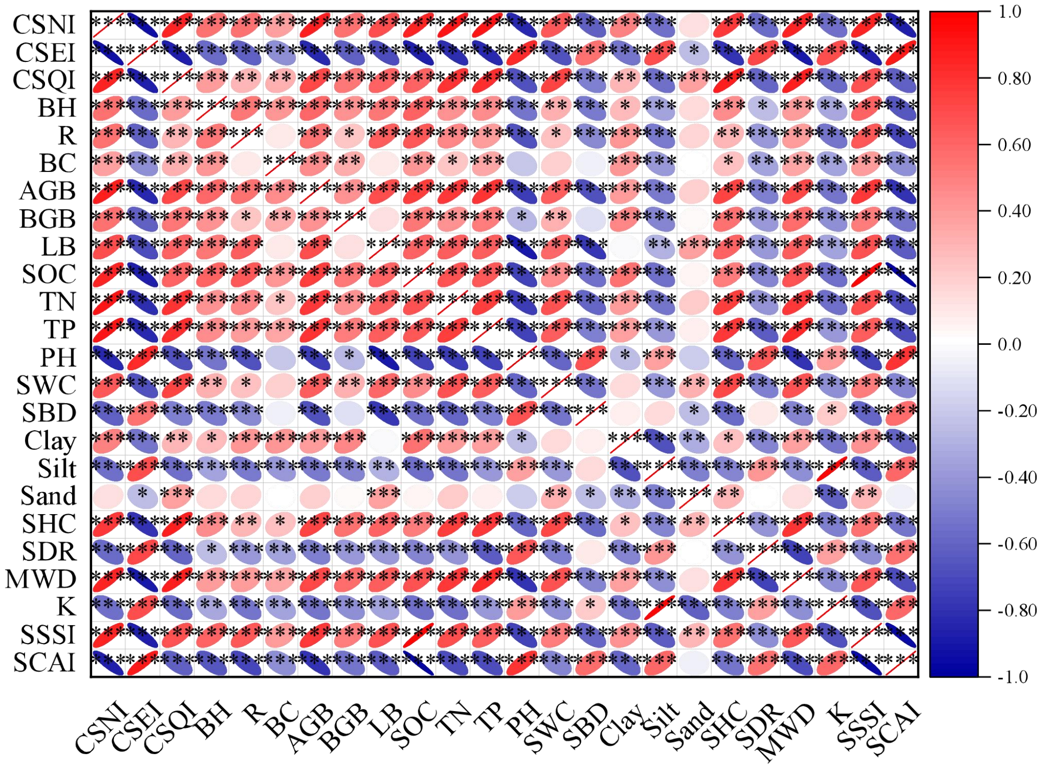

Fig. 7. Correlation analysis of CSNI, CSEI and CSQI with vegetation and soil characteristics. Red indicates a positive correlation, blue indicates a negative correlation, and the color depth indicates Pearson coefficients $*p < 0.05$, $**p < 0.01$ and $***p < 0.001$, n = 84. CSNI, comprehensive soil nutrient index; CSEI, comprehensive soil erodibility index; CSQI, comprehensive soil quality index.

Table 1

The two-way ANOVA result for soil nutrient and erodibility. SOC: soil organic carbon; TN: total nitrogen; TP: total phosphorus; CSNI: comprehensive soil nutrient index; SHC: saturated hydraulic conductivity; SDR: soil disintegration rate; MWD: mean weight diameter; K: soil erodibility factor; SSSI: soil structure stability index; SCAI: SOC cementing agent index; CSEI: comprehensive soil erodibility index; CSQI: comprehensive soil quality index.

soil variables	Land use type		Slope aspect		Land use ×Slope aspect	
Soil nutrient	F	P	F	P	F	P
SOC	1200.37	0.000	50.985	0.000	5.818	0.000

TN	520.016	0.000	79.681	0.000	24.354	0.000
TP	382.353	0.000	6.718	0.000	6.764	0.000
CSNI	832.059	0.000	46.447	0.000	6.851	0.000
Soil erodibility						
SHC	824.538	0.000	54.173	0.000	52.672	0.000
SDR	799.513	0.000	6.632	0.001	3.956	0.000
MWD	1667.15	0.000	180.654	0.000	10.673	0.001
K	859.009	0.000	14.423	0.000	23.822	0.000
SSSI	517.098	0.000	41.05	0.000	26.717	0.000
SCAI	693.653	0.000	15.553	0.000	6.623	0.000
CSEI	1120.468	0.000	38.983	0.000	6.369	0.000
Soil quality						
CSQI	642.05	0.000	103.399	0.000	35.679	0.000

Table 2

Soil erodibility indicators of different land use types at different slope aspect (mean ± SD). SHC, saturated hydraulic conductivity; SDR, soil disintegration rate; MWD, mean weight diameter; K, soil erodibility factor; SSSI, soil structure stability index; SCAI, SOC cementing agent index. Different capital letters indicate significant differences between slope aspects ($p<0.05$), different lowercase letters indicate significant differences between the land use types ($p<0.05$).

Slope aspect	Land use	SHC mm min^{-1}	SDR g min^{-1}	MWD mm	K $t\cdot hm^2\cdot h\cdot hm^{-2}\cdot MJ^{-1}\cdot mm^{-1}$	SSSI g kg^{-1}	SCAI mm kg^{-1} g^{-1}
W	Degraded land	0.13±0.02cC	1.64±0.19aA	0.79±0.02dD	0.33±0.01aA	0.25±0.01dB	20.23±0.81aA
	grassland	0.28±0.04bC	0.29±0.04cA	1.83±0.06bD	0.26±0.01dA	0.51±0.06bB	9.09±0.97bA
	shrubland	0.32±0.07bC	0.82±0.53bA	2.38±0.32aD	0.32±0.01bA	0.46±0.04cB	9.03±0.80bA
	Woodland	0.53±0.06aC	1.58±0.07aA	1.46±0.15cD	0.27±0.01cA	0.61±0.05aB	7.53±0.70cA
N	grassland	0.28±0.03bB	0.26±0.02cB	2.32±0.47bC	0.31±0.01aAB	0.50±0.06aA	8.30±0.94aB

	shrubland	0.31±0.04bB	0.73±0.44bB	2.84±0.12aC	0.29±0.04aAB	0.58±0.08aA	8.14±0.95aB
	Woodland	0.60±0.07aB	1.26±0.17aB	1.76±0.29cC	0.29±0.01aAB	0.57±0.03aA	7.90±0.39aB
	grassland	0.93±0.11bA	0.24±0.01cBC	3.28±0.04aB	0.25±0.01cB	0.51±0.10bA	9.16±1.74aB
S	shrubland	1.31±0.20aA	0.40±0.11bBC	3.32±0.06aB	0.31±0.01aB	0.53±0.03bA	8.27±0.40abB
	Woodland	1.45±0.14aA	1.17±0.06aBC	3.25±0.07aB	0.28±0.01bB	0.67±0.10aA	6.94±1.00bB
	grassland	1.55±0.18aA	0.24±0.01cC	4.06±0.14aA	0.29±0.01aB	0.59±0.02bA	7.28±0.29bB
E	shrubland	1.71±0.06aA	0.31±0.07bC	3.46±0.09bA	0.26±0.02bB	0.61±0.05bA	8.18±0.89aB
	Woodland	1.73±0.12aA	0.38±0.03aC	3.42±0.10bA	0.28±0.01bB	0.71±0.05aA	6.41±0.44cB