Author's response

Response to Anonymous Referee #1's comments on manuscript egusphere-2022-8951	
Response to Anonymous Referee #2's comments on manuscript egusphere-2022-89515	

Response to Anonymous Referee #1's comments on manuscript egusphere-2022-895

We thank the Reviewer for the immensely helpful comments. In response, we have revised the main text to improve clarity and grammar throughout. We respond to each specific comment in detail below. The reviewer comments are shown in *black italics*. Our replies are shown in indented black text, and the modified text is shown in corresponding screenshots. The annotated line numbers refer to the revised copy of the manuscript.

The manuscript attempts to explore the influence of adopted emission source profiles in CTMs on the simulated results of PM_{2.5} components by sensitivity analysis. The extent of the influence for different components were quantitatively analyzed, the impact laws and pathway were identified. The topic is interesting and their findings highlight the importance of effective utilization of emission source profiles in CTMs. Although the description of experiments is complete to allow their reproduction by fellow researchers, some explanations and discussions are not clear. I recommend its publication subject to the following amendments.

Major concern 1:

What is the design basis for the perturbation of emission source profile in the sensitivity experiments?

Response:

First, we analysis the source profile through the published literatures and existing source profile databases, we found that **the main components**

and their contents of different sources were significantly different, for example, 1) In industry process, the percentages of Ca, Fe, OC and SO₄²⁻ are relatively high, but the shares in different source profile database varied. In SPAPPC (database of Source Profiles of Air Pollution and published source profiles in China), these four components account for $16.4\pm14.9\%$, $10.4\pm14.4\%$, $6.9\pm6.1\%$, $6.2\pm6.4\%$, the proportions in SPECIATE (US EPA SPECIATE database) are 10.4±9.8%, 11.4±10.6%, 8.5±4.9%, 16.3±13.3%, respectively. 2) The transportation sector makes a dominant contribution of OC and EC, but still vary in wide range: In SPAPPC, the percentages of OC, EC are 40.8±15.0%, 23.1±13.8%, and in SPECIATE, the percentages are $40.6\pm16.4\%$, $36.1\pm21.5\%$, respectively. Besides that, the variations of main components in the same category of emission sources are also obvious, for example, the compositions of PM_{2.5} emitted by coal-fired power plants with different flue gas desulfurization facilities, e.g. wet/dry limestone, ammonia and double-alkali flue gas desulfurization, have been proved to be very different. So we take the variation range in the source profile as the range of the sensitivity experiment for each component. The detail of this step is shown in section 2.2 of the manuscript.

Second, we divided the components into several groups according to the pre-experiment due to the large number and complex chemical composition of $PM_{2.5}$. Through the pre-experiment, we found that the results for SNA (SO₄²⁻, NO₃⁻, and NH₄⁺) and Non-SNA were obviously different. Therefore, we divided the components in the source profile into four groups (Non-SNA, SO₄²⁻, NO₃⁻, and NH₄⁺). The second step could be found in section 4 of the manuscript.

Based on the two pieces of information mentioned above, 1) the perturbation on the percentage of each component in source profile must fell within the variation range of its measured value described in section 2.2. 2) The sum of the percentage of listed Non-SNA, SNA and Other components in PM_{2.5} source profile was 100%; Finally, the sensitivity experiment of **perturbation on Non-SNA, perturbation on SO**₄²⁻, **perturbation on NO**₃⁻, **and perturbation on NH**₄⁺ were determined. In the meantime, keeping the other modeling conditions unchanged except source profile.

In general, the perturbation on each component was fallen in the actual fluctuation percentage range of that component in source profile, and grouped based on pre-experiment results to design the sensitivity experiment. The design idea is shown in Figure RC1 as follows:



Figure RC1 The sketch of design idea

Major concern 2:

The discussion of the results should be extended. The authors mentioned that emission source profile adopted in CTMs has a significant impact on the simulation results of $PM_{2.5}$ components, so how to select the appropriate source profiles in the simulation? In the section of conclusion (Line 549-551), the author concluded that "the representativeness and

timeliness of the source profile should be considered". How to understand the "representativeness" and "timeliness" here?

Response:

Source profile, a physicochemical point of view of which reveals the signatures of source emission, play an important role in the application of CTMs for converting total emissions from source into the speciated emission and calculating source-specific emission of individual compounds (Reff et al., 2009; Hsu et al., 2019). In the past few years, source profile of PM_{2.5} from a variety of source types have been substantially developed all over the world, especially in US (Simon et al., 2010), Europe (Pernigotti et al., 2016) and East Asia (Liu et al., 2017; Bi et al., 2019). With the change of fuel and raw materials, the development of production technology and the innovation of pollution treatment technology in recent years, some components have changed significantly in the source profile. By comparing the source profile in exited databases and published literatures, we found that the components in PM_{2.5} source profiles have the following characteristics:

Firstly, the large variation of components content exists in source profiles. We take coal-fired power plants (PP) as an example here (Coalfired power plants remain the main coal consumers in China (NBS, 2021), source profile data were from SPECIATE¹ and SAPPC²). The dominant components generally are similar such as SO₄²⁻, Cl⁻, Ca, OC, Al and Si in PP source profiles, however, there are large and small differences in their contributions. In SAPPC, the average weight percentage of main components are sorted by SO_4^{2-} , Cl⁻, Ca, OC, Si, and their percentage range were 0.6%~47.4%, 0.1%~27.8%, 0.6%~24.1%, 0.3%~34%, 0.4%~28.3%, respectively. In SPECIATE, the main components in PP source profiles were SO_4^{2-} , Ca, OC, Al, Si, and their variation range were 0.4%~71.1%, 2.3%~24.8%, 0.7%~70.3%, 1.2%~19.7%, 1.9%~23.9%, separately. Our previous study also showed that the relatively large variation in the source profiles for industry emissions, vehicle emissions and residents coal combustion, it is called for the establishment of local profiles for these sources (due to their high uncertainties) through the uncertainty analysis (Bi et al., 2019).

Secondly, the main components (or the tracer components) of emission sources have changed because of the changing standards. On Jan. 1, 2012, China began to implement the new Emission Standards for Air **Pollutants** from Thermal Power Plants (GB13223-2011, https://www.mee.gov.cn/ywgz/fgbz/bz/bzwb/dqhjbh/dqgdwrywrwpfbz/20 1109/W020130125407916122018.pdf), which stipulates that SO_2 emissions from thermal power boilers in key areas shall be subject to the stricter standard. To meet new emission standards, the installation rate of desulphurization facilities in coal-fired power plants has greatly increased, which to some extent affects the composition of coal-fired sources in

Chinese cities. It has been reported that the percentages of Ca, Mg, SO_4^{2-} and Cl⁻ in PP profiles increased after the limestone-gypsum method was used in coal-fired power plants (Zhang et al., 2020; Bi et al., 2019), Ammonia desulphurization will increase NH_4^+ and SO_4^{2-} in particulate matter (Pan et al., 2016). Due to the changing standards of gasoline and diesel oil since the 1980s, Pb and Mn are no longer tracers of gasoline vehicle emissions (Bi et al., 2019). However, OC and EC still are the dominant species in vehicle emissions since the 1980s, despite the changing standards, this also could be seen from our manuscript. Especially China plans to achieve carbon neutral before 2060, more stringent standards will be introduced, the characteristics of source profiles' components will also change.

Thirdly, with the development of advanced sampling and chemical analysis techniques, more valuable information has been explored to further know about the source profiles. A number of recent studies found that, contrary to our previous belief, primary emission may be more important for some components, for example, sulfate (a major PM_{2.5} component) was largely from primary emissions rather than secondary formation in ambient air in certain circumstances (Dai et al., 2019; Ding et al., 2021; Yan et al., 2020).

Besides that, fuel, raw and auxiliary materials, process conditions, pollution removal facilities, source sampling methods and other

factors have a significant impact on the source profile of PM_{2.5}.

Therefore, The representativeness and timelessness of source profile, **from a macro perspective**, it needs to see whether it is a typical source profile, and whether it can represent the chemical composition of PM_{2.5} emitted by sources in the region in the study stage; **From the micro-view**, it is to evaluate whether the components' characteristics in the source profiles can represent the chemical compositions of the vast majority of such sources in the actual environment, which is based on the general chemical composition law of the source profile. We should **consider the regional emission character** and **the characteristics of regional emission period** when selected the source profiles.

This paper preliminarily explored the impact of emission source profiles on the simulation of $PM_{2.5}$ components, the detail about how to select the source profile will be further studied in our future work, to provide some new ideas for improving the uncertainty of model simulation.

Reference:

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¹ SPECIATE- U.S. Environmental Protection Agency's (EPA) SPECIATE database, https://www.epa.gov/air-emissions-modeling/speciate.

² SAPPC- SPAP database and published source profiles in China; SPAP-database of Source Profiles of Air Pollution, http://www.nkspap.com:9091/.

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- Reff, A., Bhave, P. V., Simon, H., Pace, T. G., Pouliot, G. A., Mobley, J. D., Houyoux, M.: Emissions Inventory of PM_{2.5} Trace Elements across the United States, Environ. Sci. Technol., 43, 5790-5796, <u>http://doi.org/10.1021/es802930x</u>, 2009.
- Simon, H., Beck, L., Bhave, P. V., Frank, D., Hsu, Y., Luecken, D., Mobley, J. D., Pouliot, G. A., Reff, A., Sarwar, G., Strum, M.: The development and uses of EPA's SPECIATE datebase, Atmos. Pollut. Res., 1, 196-206, <u>https://doi.org/10.5094/APR.2010.026</u>, 2010.
- Yan, Q., Kong, S., Yan, Y., Liu, H., Wang, W., Chen, K., Yin, Y., Zheng, H., Wu, J., Yao, L., Zeng, X., Chen, Y., Zheng, S., Wu, F., Niu, Z., Zhang, Y., Zheng, M., Zhao, D., Liu, D., Qi, S.: Emission and simulation of primary fine and submicron particles and water-soluble ions from domestic coal combustion in China, Atmos. Environ., 224, 117308, http://doi.org/10.1016/j.atmosenv.2020.117308, 2020.
- Zhang, J., Wu, J., Lv, R., Song, D., Huang, F., Zhang, Y., Feng, Y.: Influence of Typical Desulfurization Process on Flue Gas Particulate Matter of Coal-fired Boilers (In Chinese), Environ. Sci., 41, 4455-4461, https://doi.org/10.13227/j.hjkx.202003193, 2020.

Minor concern 1:

Line 21 and Line 27, there are two notes for CTM in one paragraph,

which appear to be repetitive.

Response:

We have deleted the duplicated notes for CTM in our manuscript.

20 The chemical transport model (CTM) is an essential tool for air quality prediction 21 and management, widely used in air pollution control and health risk assessment. However, the current models do not perform very well in simulating PM_{2.5} components. 22 Studies suggested that the uncertainties of model chemical mechanism, source emission 23 inventory and meteorological field can cause inaccurate simulation results. Still, the 24 emission source profile of PM2.5 has not been fully taken into account in current 25 numerical simulation. This study aims to answer (1) Whether the variation of source 26 27 profile adopted in chemical transport models (CTMs) has an impact on the simulation

Minor concern 2:

Line 57-59, *the references are verbose.*

Response:

We have removed redundant references in our manuscript.

- 57 adverse impact on human health (Shi et al., 2018) and ecosystem (Han et al., 2019;
- 58 Zhou et al., 2018), such as acid rain in southwest China (Han et al., 2019), food security
- 59 (Zhou et al., 2018), etc. ←

Minor concern 3:

Line 111-113, It is not clearly explained the role of source profiles in CTMs.

Response:

We have added extra explain in our manuscript and cited the source.

111 In particular, the emission source profile of $PM_{2.5}$ (Hereinafter referred to as

- 112 "source profile"), creating speciated emission inventories for CTMs (Hsu et al., 2019),
- has not been fully taken into account in the current numerical simulation by CTMs. In

Reference: Hsu, Y., Divita, F., Dorn, J.: SPECIATE 5.0 - Speciation Database Development Documentation, Final Report, M. MENETREZ, Abt Associates Inc./Office of Research and Development/U.S. Environmental Protection Agency Research Triangle Park, NC27711, https://www.epa.gov/sites/default/files/2019-07/documents/speciate 5.0.pdf, 2019.

Minor concern 4 and 5:

Line 257: "The detailed information on" should be "The information of ... "

Line 259: "Coefficient Divergence (CD)" would be appropriate.

Response:

We have replaced the sentence with the correct expression in our manuscript. New line is in 261-263.

261 CMAQ_SPA and CMAQ_SPE. The detailed information of on-source profiles is shown

262 in Figure S1. To determine the similarity between the two groups of source profiles,

263 Coefficient divergence-Divergence (CD) is calculated using the following formula

264 (Wongphatarakul et al., 1998):↩

Minor concern 6:

In the supplementary material, Fig. S1, the author selected code 91041, 900162.5, 91155, 91022 and 91162 as SPECIATE source profiles

for simulation. Detailed information of these source profiles need be provided by authors.

Response:

We have added a table (Table S26) in supplementary material to show the detail information of these source profiles.

Code	Profile Name	Controls	Profile Date	Profile Notes	Keywords
91041ª	Draft Sub- Bituminous Combustion - Composite	Mixture of Baghouse, None, Electrostatic Precipitator, Wet Scrubber, Mechanical Collectors, Dry Lime Scrubber, Ammonia Injection	2006-5-24	Replaced by Profile 91110. Median of Profiles 3191, 3192, 3690, 3694, and 3700.	Sub- Bituminous Coal Combustion; PM Composite
900162.5 ^b	Industrial Manufacturing - Average	Not Applicable	1989-1-5	Average profile developed from original profiles representing the source category group 3xxxxxx.	INDUSTRIAL
91155°	Residential Coal Combustion - Composite	Uncontrolled	2009-7-12	Median of Profiles 3761, 432012.5	Residential Coal Combustion; Inventory speciation
91022ª	Draft On-road Gasoline Exhaust - Composite	Mixture of Catalytic converter and Not available	2006-5-24	ReplacedbyProfile91122.MedianofProfiles311072.5,	On-road Gasoline Exhaust; PM Composite

Table S26 The selected information of source profile in SPECIATE and SPAPPC database

91162°	LDDV Exhaust - Composite	Mixture of Catalytic converter and Not available	2009-7-12	3517, 3884, 3892, 3904, 3947, 3951, 3955, 3959, and 4558. Median of Profiles 321042.5, 3912, 3963, 4675	LDDV Exhaust; Inventory speciation
Local ^d	РР	Mixture of Baghouse, None, Electrostatic Precipitator, Wet Scrubber, Mechanical Collectors, Dry Lime Scrubber,		Average of profiles power and heating power plant	
Local ^d	IN	Wet Scrubber, Dry Lime Scrubber,		Average of profiles steel, metallurgy, cement, glass, industrial boiler	
Local ^d	TR	Mixture of Catalytic converter		Average of profiles gasoline, diesel, gasoline- diesel exhaust	
Local ^d	RE			Average of profiles civil boiler	

a, Hsu, Ying, Randy Strait, Stephen Roe, David Holoman. 2006. 'SPECIATE 4.0 Speciation database development document - Final Report', Prepared for US EPA, RTP, NC, EPA Contract Nos. EP-D-06-001, Work Assignment Numbers 0-03 and 68-D-02-063, WA 4-04 and WA 5-05, by E.H. Pechan & Associates, Incorporation, Durham, NC. https://www.epa.gov/sites/production/files/2015-10/documents/speciatedoc_1206.pdf. b, Shareef, G. S. Engineering Judgement, Radian Corporation. August 1987.

c, Reff, Adam, Prakash V Bhave, Heather Simon, Thompson G Pace, George A Pouliot, J David Mobley, and Marc Houyoux. 2009. 'Emissions Inventory of $PM_{2.5}$ Trace Elements across the

United States', Environmental Science & Technology, 43, no. 15: 5790-96. DOI: 10.1021/es802930x.

d, Database of Source Profiles of Air Pollution (SPAP), measured by State Environmental Protection Key Laboratory of Urban Ambient Air Particulate Matter Pollution Prevention and Control & Tianjin Key Laboratory of Urban, Nankai University. Coal combustion by power plants (PP), industrial processes (IN), residential emission (RE) and transportation sector (TR).

Response to Anonymous Referee #2's comments on manuscript egusphere-2022-895

We thank the reviewer for the immensely helpful comments. In response, we have carefully addressed the referee's concerns with this work. Please see point-by-point response to the comments and the revised manuscript for details. The reviewer's comments are shown in black *italics*. Our replies are shown in indented black text.

The manuscript investigates the sensitivity of simulated $PM_{2.5}$ and its components' concentrations to the uncertainties in the componentspecified $PM_{2.5}$ source emission inventories using the CMAQ chemical transport model. The relatively-complete chemical components, including Al, Ca, Cl, EC, Fe, K, Mg, Mn, Na, OC Si, NH_4^+ , NO_3^- , SO_4^{2-} , and others, are taken into account in the emission inventory used. The authors showed that the influence of the relative contributions of different components to the total $PM_{2.5}$ emission (denoted as source profile changes in the manuscript) on simulated $PM_{2.5}$ concentration was insignificant, but its impact on $PM_{2.5}$ components could not be ignored. They also showed that these source profile changes caused the variations in simulated gaseous pollutants' concentrations. While such kind of model experiment should be a welcome addition to the literature on air quality model simulation, I do have concerns that the data and methodology used in this study would be sensible (or well introduced) and the conclusions applicable to the simulations done by other chemical transport models with different chemical and physical modules. Therefore, I cannot recommend publication the current version of this manuscript in GMD.

The major issues are follows:

1. What is the grid resolution of the MEIC emission inventory that was used for the model simulation in this study? Is the resolution sufficiently fine for the Dom3 (4 km× 4km) simulation? What does the area marked in green in Fig. 1 refer to? No information on the regional distributions of either $PM_{2.5}$ emission sources or their simulated concentrations is provided in the manuscript. Are all the 10 monitoring sites located in the cities of Dom3? Is there any site that is located near the desert area? Were the mineral dust emissions taken into account in the simulation?

Response:

Thank you for your reminder. More description of simulation area is placed in Fig. 1 and emission information in Fig. S2 of the revised manuscript. To address the reviewer's comment, additional interpretation has been made.

The grid resolution of the MEIC emission inventory was $0.25^{\circ} \times 0.25^{\circ}$; We extracted the emissions from the original national

emission inventory and reprocessed the emissions into 36km×36km, 12km×12km, and 4km×4km grids for Domains 1, 2, and 3, respectively. The Inventory Spatial Allocate Tool (ISAT) was used to provide grided PM_{2.5} emission inventory for the simulations. Considering the purpose of this paper is to explore how much the source profile changes will affect the simulation results, the resolution of the emission inventory is enough. For different scenarios, other modeling conditions remain the same except for the component-specified PM_{2.5} source emission inventories changed.

The area marked in green in Fig. 1 is Tianjin city in the third domain (Dom 3). The third domain with a horizontal resolution of $4 \text{ km} \times 4 \text{ km}$ mainly focuses on Tianjin region which is marked in Fig. RF1(b) as follow (In the revised manuscript, we have replaced Fig. 1 with Fig.RF1 below to make it more clearly).



Fig. RF1 Modeling domains of the CMAQ model. (a) The three nested domains in CMAQ model; (b) Land use and observation sites of Dom3.

Data source of Land use: GLOBELAND30, www.globeland30.org, National Geomatics Center of China.



Fig. RF2 The regional distribution of PM_{2.5} emission sources. (a) coal-fired power plant; (b) industry process; (c) transportation sector; (d) residential coal combustion.

The information of regional distribution of $PM_{2.5}$ emission sources are shown in Fig. RF2. In the revised manuscript, we have also provided the regional distributions of $PM_{2.5}$ emission sources (Fig. S2) in the supplementary material.

All the monitoring sites locate in the third domain which is shown in Fig. RF1(b). No sites are located in desert areas and the dust emissions are not taken into account in our simulation as the study region is far away from the deserts. The land use type of Dom3 is shown in Fig. RF1(b).

To fully address the reviewer's comment, additional interpretation has been made as follows:

In chemical transport models such as CMAQ, GEOS-Chem, CAMx, the $PM_{2.5}$ emission inventory is speciated in the chemical-composition dimension (Reff et al., 2009). Some commonly used emission inventories are listed in Table RF1. Different CTMs and their aerosol module have different regulations on $PM_{2.5}$ species types. Pollutants or species in emission inventory, especially for PM and VOCs, need to be speciated into chemical components for CTMs to match chemical mechanism. Taking CMAQ as an example, the aerosol module (AERO6) expands the definition of the PM Other species in earlier versions to include more detailed PM species (Chapel Hill, 2012); There are 18 $PM_{2.5}$ species in AERO6: OC, EC, SO_4^{2-} , NO_3^- , NH_4^+ , H_2O , Na, Cl, NCOM, Al, Ca, Fe, Si, Ti, Mg, K, Mn, and Other. Other CTMs also have similar regulation, the classification of $PM_{2.5}$ species in mainstream CTMs are shown in Table RF2.

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Scale	Name	Air pollutants
Global	EDGAR ¹	CO, NO _x , NMVOC, CH ₄ ; NH ₃ , NO _x , SO ₂ ; PM ₁₀ , PM _{2.5} , BC, OC
Global	EDGAR-	SO ₂ , NO _x , CO, NMVOC, PM ₁₀ , PM _{2.5} , BC, OC, NH ₃
	HTAP ²	
Global	GAINS ³	SO ₂ , NO _X , VOC, PM, NH ₃ , CO ₂ , CH ₄ , N ₂ O and the F-gases
Reginal	MIX,	SO ₂ , NO _x , CO, NMVOC, PM ₁₀ , PM _{2.5} , BC, OC, NH ₃ , and CO ₂
	$MEIC^4$	
Reginal	NEI ⁵	CO, NOx, PM ₁₀ , PM _{2.5} , SO ₂ , VOC, NH ₃
Reginal	REAS ⁶	SO ₂ , NO _x , CO, NMVOC, PM ₁₀ , PM _{2.5} , BC, OC, NH ₃ , and CO ₂

Table RF1 The air pollutants in emission inventory

Note:

1, Emissions Database for Global Atmospheric Research (EDGAR) (1970-). https://edgar. jrc.ec.europa.eu/dataset_ap61

2, The Task Force Hemispheric Transport of Air Pollution (HTAP) (2000-2010). https://je odpp.jrc.ec.europa.eu/ftp/jrc-opendata/EDGAR/datasets/htap_v2_2/ALL/

3, Greenhouse Gas and Air Pollution Interactions and Synergies (GAINS) (1990-).https://ga ins.iiasa.ac.at/gains/download/GAINS-tutorial.pdf.

4, A new Asian anthropogenic emission inventory (MIX) (2008, 2010); Multi-resolution Emission Inventory for China (MEIC) (2008-). http://meicmodel.org/

5, National emission inventory (NEI) (1970-), https://www.epa.gov/air-emissions-inventorie s/national-emissions-inventory-nei

6, Regional Emission inventory in Asia (REAS) (1950-2015). https://www.nies.go.jp/REAS/index.html#REASv3.2.1

7, European Monitoring and Evaluation Programme (EMEP) (1990-), https://www.eea.europ a.eu/data-and-maps/dashboards/national-air-pollutant-emissions-data

Aerosol module	PM _{2.5} species							
	Al, Ca, Cl, EC, Fe, K, Mg,							
	Mn, Na, OC, Si, Ti, NH4 ⁺ ,							
AEKU0	NO ₃ ⁻ , SO ₄ ²⁻ , NCOM, Other,							
	H_2O							
AERO5	OC, EC, NO ₃ ⁻ , SO ₄ ²⁻ , Other							
	Al, Ca, Cl, EC, Fe, K, Mg,							
aerosol.mod	Mn, Na, OC, Si, Ti, NH4 ⁺ ,							
	NO_3^- , SO_4^{2-} , Other							
MADE MOSAIC MAM	OC, EC, NO_3^- , SO_4^{2-} , Ca, Na,							
MADE, MOSAIC, MAM	Cl, H ₂ O, Other							
CE	OC, EC, NO ₃ ⁻ , SO ₄ ²⁻ , NH ₄ ⁺ ,							
Cr	Cl, Na, Other							
	Aerosol module AERO6 AERO5 aerosol.mod MADE, MOSAIC, MAM CF							

Table RF2 The speciated allocation for PM2.5 in mainstream CTMs

Note:

1, Particulate matter (aerosols): PM using three lognormal sub-distributions, or modes, two interacting modes (Aitken and accumulation) represent $PM_{2.5}$

https://www.airqualitymodeling.org/index.php/CMAQ_version_5.0_(February_2010_release) _OGD#Aerosol_Module.

2, Particulate matter in GEOS-Chem: $PM_{2.5} = (NH4 + NIT + SO4) * 1.10 + BCPI + B$ CPO + (OCPO + (OCPI * 1.05)) * (OM/OC ratio) + DST1 + DST2 * 0.30 + SAL A * 1.86 + SOA * 1.05. (NIT-NO3; BCPI and BCPO-EC; OCPO and OCPI-OC, NCO M; DST1-SO4, NH4, NO3, Cl, Na, K, Ca, Fe, Al, Si, Ti, Mn, Other, OC, NCOM; DST2-SO4, Cl, ASOL; SALA-SO4, Cl, Na, Mg, K, Ca.

http://wiki.seas.harvard.edu/geos-chem/index.php/GEOS-Chem_to_CMAQv5.0) http://wiki.seas.harvard.edu/geos-chem/index.php/Particulate_matter_in_GEOS-Chem.

3, Aerosols in WRF-Chem: PM using 3 or 7 log-normal modes, two interacting modes (Ai tken and accumulation) represent $PM_{2.5}$.

https://ruc.noaa.gov/wrf/wrf-chem/wrf_tutorial_2018/Aerosols.pdf.

^{4,} Aerosol Chemistry: $PM_{2.5} = PSO4 + PNO3 + PNH4 + PEC + NA + PCL + POA + SOA1 + SOA2 + SOA3 + SOA4 + SOPA + SOPB + FPRM + FCRS + (PFE + PMN + PK + PCA + PMG)$

⁺ PAL + PSI + PTI) (Fe, Mn, K, Ca, Mg, Al, Si and Ti are Optional Species).

https://camx-wp.azurewebsites.net/Files/CAMxUsersGuide_v7.20.pdf.

As total PM_{2.5} need to be speciated into its chemical components to match the chemical mechanism in CTMs, emission source profiles, which can provide "species" and "split factor" for PM2.5, are key inputs for creating chemically-resolved emission inventories for CTMs. However, the actual emission source profile of $PM_{2.5}$ and the sensitivity of simulated components' concentrations to the variation in PM_{2.5}source profiles are currently not well considered. In some studies, the $PM_{2.5}$ emission inventory is speciated using "None" or "simplified profiles" in the chemical-composition dimension (Reff et al., 2009). The corresponding literature-based data is presented in Table RF3 as bellow, we only selected the main components of $PM_{2.5}$ (SO₄²⁻, NO₃⁻, NH₄⁺, OC and EC) as example here. The species allocation coefficients of PM_{2.5} emission sources are commonly treated in the following ways: (1) allocated $PM_{2.5}$ components of source emissions by referring to source profile data in published literature or database like the US SPECIATE; (2) chemical profiles came from local measurement. With the development of production technology and the innovation of pollution treatment technology in recent years, some

source profiles have changed dramatically. The timeliness of $PM_{2.5}$ species allocation coefficients in current CTMs also need to be considered.

Although the number of $PM_{2.5}$ species and calculation method in different CTMs are different, no matter what kinds of CTMs, as long as it involves chemical components simulation for $PM_{2.5}$, the influence of source emission profiles should be considered. It remains unclear whether the variations of adopted emission source profiles of $PM_{2.5}$ had influence on the CTMs' performance and how much the influence would be and how it works. The purpose of this paper is to explore how much the $PM_{2.5}$ emission source profile changes will affect the simulation results. Taking CMAQ (one of the most widely used CTMs) and MEIC (a high-resolution inventory of anthropogenic air pollutants in China) as the carrier, we tested the sensitivity of the simulated chemical components to the variation of source profiles. The same kind of experiment is also applicable to other CTMs and emission inventories (e.g. NEI, EEI, REAS, HATP, etc.).

Thecomponentproportioninsource profile	PM _{2.5} components	Model	NMB	R	Study area	Period	Reference
9%	SO_4^{2-}	$CMAO_{2}4.7.1$	-45%	0.73	Eastern Chine	2010	(Chang at al. 2015)
1%	NO ₃ -	CMAQV4.7.1	29%	0.82	Eastern China	2010	(Cheng et al., 2013)
Not explicitly	SO_4^{2-}		-4.5%	0.87			
Specified	NO ₃ -	CMAQv4.7.1	10%	0.87	Qing Dao	Jan. 2016	(Zhang et al., 2017)
specified	$\mathrm{NH_4^+}$		-6%	0.9			
	SO_4^{2-}		-54%	0.6			
	NO ₃ -		-40%	0.8		2013	(Zheng et al., 2015)
	$\mathrm{NH_4^+}$	CMAQv5.0.1	-58%	0.7	Northern China		
	OC		-25%	0.8			
Not explicitly	EC		196%	0.6			
Specified	SO4 ²⁻		6%	0.7			
	NO ₃ -		6%	0.8			
	$\mathrm{NH_4}^+$	Revised CMAQ	-4%	0.8			
	OC		-28%	0.7			
	EC		183%	0.6			
	SO 2-		-84%	0.31		Jan. 2017	
	504-		-71%	0.26		Apr. 2017	(Sha et al., 2019)
Not explicitly	NO -	WDE Cham? 6.1	45%	0.51	Nanjing	Jan. 2017	
Specified	NO ₃	w KF-Chem5.0.1	67%	0.32		Apr. 2017	
	NILL.+		-34%	0.27]	Jan. 2017	
	$\mathrm{NH_{4}^{+}}$		-13%	0.31		Apr. 2017	

Table RF3 The adopted source profile and simulation result for different CTMs from published literatures

	SO4 ²⁻		-41%	0.82		D 2015 I	
Not explicitly	NO ₃ -	CMAQv5.0.2	41%	0.83	Qing Dao	Dec. $2015 \sim \text{Jan.}$	(Gao et al., 2020)
Specified	$\mathrm{NH_4^+}$		-5%	0.83		2016	
	SO4 ²⁻		-4%	0.83			
Not oveligitly	NO ₃ -		-4%	0.77			
Specified	$\mathrm{NH_{4}^{+}}$	RAQMS	4%	0.81	Beijing	Feb. to Mar. 2014	(Li et al., 2020)
specified	OC		-39%	0.92			
	EC		-9%	0.81			
Not avaliaitly	SO4 ²⁻		-56%~-29%				
Specified	NO ₃ -	CMAQv5.0.1	-47%~19%	-	China	2013	(Shi et al., 2017)
	$\mathrm{NH_4^+}$		-44%~1				
	SO. ²⁻	D. ² -	-16% and -6%		USA	Jan. 2006	(Foley et al., 2010)
	504		-19%~-0.2%			Aug. 2006	
	NO ₃ -		-5% and 1%	-		Jan. 2006	
Not explicitly	NH +		13% and 14%			Jan. 2006	
Specified	11114	CMAQv4.7	15% and -6%			Aug. 2006	
specified	OC		-20%			Jan. 2006	
			-49%			Aug. 2006	
	FC		-25%			Jan. 2006	
	LC		-32%			Aug. 2006	
0%	SO /2-		-34%~7%			Jan. 2002	
970	504		-18%~-37%	-		Jul. 2002	(Liu et al., 2010)
10/2	NO ₂ -	CMAQv4.5.1	16%~118%		USA	Jan. 2002	
1%	NO ₃		-69%~88%			Jul. 2002	
0%	$\mathrm{NH_4^+}$		-0.5%~61%			Jan. 2002	

			-43%~53%			Jul. 2002	
200/	00		-4%~13%			Jan. 2002	
30%	00		-71%~-64%			Jul. 2002	
240/	EC		-16%~18%			Jan. 2002	
24%0	EC		-39%~38%			Jul. 2002	
		CMAQv4.5.1	5%	0.7		Lan. 2002	
00/	SO ² -	CAMx-4.4.2	33%	0.6		Jan. 2002	
9%	5042	CMAQv4.5.1	-39%	0.5		Lul 2002	
		CAMx-4.4.2	-9%	0.6		Jul. 2002	
		CMAQv4.5.1	46%	0.8		Lan. 2002	
10/	NO -	CAMx-4.4.2	-21%	0.8	South Eastern USA	Jan. 2002	
1 %0	INO ₃	CMAQv4.5.1	-62%	0.2		Jul. 2002	
		CAMx-4.4.2	-80%	0.2			
	$\mathrm{NH_{4}^{+}}$	CMAQv4.5.1	-7%	0.8		Lan. 2002	— (Zhang et al., 2013)
00/		CAMx-4.4.2	-8%	0.7		Jan. 2002	
0%0		CMAQv4.5.1	-52%	0.7		Lul 2002	
		CAMx-4.4.2	-45%	0.7		Jul. 2002	
		CMAQv4.5.1	-15%	0.8		L., 2002	
200/	00	CAMx-4.4.2	-18%	0.8		Jan. 2002	
30%	00	CMAQv4.5.1	-73%	0.7		L-1 2002	
		CAMx-4.4.2	-47%	0.7		Jul. 2002	
		CMAQv4.5.1	-9%	0.7	-	Lev. 2002]
240/	EC	CAMx-4.4.2	5%	0.7		Jan. 2002	
24%	EC	CMAQv4.5.1	-47%	0.4		L.1 2002	
		CAMx-4.4.2	-33%	0.4		Jul. 2002	

00/			0.7% and -31%	0.85	USA		
9%	SO422		-2%	0.61	Europe		
10/	NO -	CMA O5 0	56%~59%	0.66	USA	1000 2010	$(\mathbf{X}_{1}^{i}) = -1 - 2015)$
1%0	NO ₃	CMAQV5.0	-6%	0.70	Europe	1990-2010	(Xing et al., 2015)
0%	NILI +		-13%	0.52	USA		
	1 N П4		34%	0.62	Europe		
	SO4 ²⁻		-16%	0.82			
Not overligitly	NO ₃ -		72%	0.64			
Specified	$\mathrm{NH_{4}^{+}}$	CMAQv4.5	13%	0.68	USA	2002~2008	(Friberg et al., 2016)
specified	OC		-30%	0.39			
	EC		-22%	0.5			
	SO4 ²⁻		-50%~29%	-	California	2013	(Chen et al., 2020)
Not overligitly	NO ₃ -		-27%~48%				
Not explicitly	$\mathrm{NH_4}^+$	CMAQv5.0.2	-32%~130%				
specified	OC		-35%~13%				
	EC		0~43%				
The emission inventories for	SO4 ²⁻						
SO4 ²⁻ , NO3 ⁻ and		CEOS Cham					
NH4 ⁺ emitted	NO ₃ -	GEOS-Chem	Quite different		China	2015	(Yan et al., 2020)
from residential		V11-01					
coal combustion	NH ⁺						
were established	1114						
Not explicitly	SO4 ²⁻	WPE Cham	$MB=5\mu g/m^3$	RMSE=12.5 μ g/m ³	BTH China	2014	(Listal 2018)
Specified	NO ₃ -	w Kr-Chein	$MB = -0.3 \mu g/m^3$	$RMSE=14.3\mu g/m^3$	ып, Ciiiila	2014	(L1 et al., 2018)

	$\mathrm{NH_4^+}$		MB=-0.4 μ g/m ³	RMSE= $8.2 \mu g/m^3$			
	SO4 ²⁻	$ \begin{array}{c} \frac{4^{2^{2}}}{D_{3}^{-}} \\ \frac{C}{C} \end{array} $ CAMx		0.32			
Local source	NO ₃ -			0.59	Tianjin	2017-2018	(Ma et al., 2022)
profile	OC			0.27			
	EC			0.47			

2. At the beginning of Sect. 2.2 it is stated that in addition to SPA and SPE, the PM_{2.5} emission source profile database from published literature was used. Where and what are the final, merged emission source profiles used in this study? The simulated PM_{2.5} and its components' concentrations using CMAQ_SPA are compared with those using CMAQ_SPE. However, no comparison with observed PM_{2.5} components' concentrations at the monitoring sites has been made to show the advantage of the SPA over the SPE.

Response:

More descriptions of source profiles are shown in Fig. S1 and Table S26 of our revised supplementary material. In addition, to address the reviewer's comment, we added an extra explanation as follows:

In this study, for SPE, the selected source profile of each source category group was the average/median profile developed from original profiles in SPECIATE database. The source profile codes for power plant (PP), industrial process (IN), residential coal combustion (RE), and transportation sector (TR) are 900162.5, 91155, 91022 and 91162, respectively. Please see Table RF4 for details. For SPA, the selected source profiles were from database of Source Profiles of Air Pollution, they are also available in our previous paper (Bi et al., 2019). The detailed information of source profiles as shown in the following Fig. RF3 and

Table RF4. They have also been updated in the revised supplementary material (Fig. S1 and Table S26).



Fig. RF3 The selected speciation profile of $PM_{2.5}$ for case CMAQ_SPE and CMAQ_SPA In SPE, the selected source profiles were average profile developed from original profiles of the source category group in SPECIATE database, the power plant (PP) source profile code was 91041, industrial process (IN) was 900162.5, Residential coal combustion (RE) was 91155, Transportation sector (TR) was 91022 and 91162. In SPA, the selected source profiles were from SPAPPC database which were measured from local emission sources.

Code	Profile Name	Controls	Profile Date	Profile Notes	Keywords
91041ª	Draft Sub- Bituminous Combustion - Composite	MixtureofBaghouse,None,ElectrostaticPrecipitator,Wet Scrubber,MechanicalCollectors,DryLimeScrubber,	2006-5-24	Replaced by Profile 91110. Median of Profiles 3191, 3192, 3690, 3694, and 3700.	Sub- Bituminous Coal Combustion; PM Composite

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		Ammonia Injection			
900162.5 ^b	Industrial Manufacturing - Average	Not Applicable	1989-1-5	Average profile developed from original profiles representing the source category group	INDUSTRIAL
91155°	Residential Coal Combustion - Composite	Uncontrolled	2009-7-12	Median of Profiles 3761, 432012.5	Residential Coal Combustion; Inventory speciation
91022ª	Draft On-road Gasoline Exhaust - Composite	Mixture of Catalytic converter and Not available	2006-5-24	ReplacedbyProfile91122.MedianofProfiles311072.5,3517,3884,3892,3904,3947,3951,3955,3959,	On-road Gasoline Exhaust; PM Composite
91162°	LDDV Exhaust - Composite	Mixture of Catalytic converter and Not available	2009-7-12	and 4558. Median of Profiles 321042.5, 3912, 3963, 4675	LDDV Exhaust; Inventory speciation
Local ^d	РР	Mixture of Baghouse, None, Electrostatic Precipitator, Wet Scrubber, Mechanical Collectors, Dry Lime Scrubber,		Average of profiles power and heating power plant	

Local ^d	IN	Wet Scrubber, Dry Lime Scrubber,	Average of profiles steel, metallurgy, cement, glass, industrial boiler
Local ^d	TR	Mixture of Catalytic converter	Average of profiles gasoline, diesel, gasoline- diesel exhaust
Local ^d	RE		Average of profiles civil boiler

a, Hsu, Ying, Randy Strait, Stephen Roe, David Holoman. 2006. 'SPECIATE 4.0 Speciation database development document - Final Report', Prepared for US EPA, RTP, NC, EPA Contract Nos. EP-D-06-001, Work Assignment Numbers 0-03 and 68-D-02-063, WA 4-04 and WA 5-05, by E.H. Pechan & Associates, Incorporation, Durham, NC. https://www.epa.gov/sites/production/files/2015-10/documents/speciatedoc_1206.pdf. b, Shareef, G. S. Engineering Judgement, Radian Corporation. August 1987.

c, Reff, Adam, Prakash V Bhave, Heather Simon, Thompson G Pace, George A Pouliot, J David Mobley, and Marc Houyoux. 2009. 'Emissions Inventory of PM_{2.5} Trace Elements across the United States', Environmental Science & Technology, 43, no. 15: 5790-96. DOI: 10.1021/es802930x.

d, Database of Source Profiles of Air Pollution (SPAP), measured by State Environmental Protection Key Laboratory of Urban Ambient Air Particulate Matter Pollution Prevention and Control & Tianjin Key Laboratory of Urban, Nankai University. Coal combustion by power plants (PP), industrial processes (IN), residential emission (RE) and transportation sector (TR).

By comparing SPA and SPE source profiles, our purpose is to show

that the source profile of same category of emission sources can vary greatly, and whether the variation of source profile adopted in CTMs has an impact on the simulation of chemical components in $PM_{2.5}$. Here we mainly tried to answer (1) Whether the variation of source profile adopted in CTMs has an impact on the simulation of $PM_{2.5}$ chemical components? (2) How much does it impact? (3) How does the impact work? Different

simulation scenarios were designed and the sensitivity of components simulation results to $PM_{2.5}$ sources profile was explored through with chemical components of source profiles perturbation. In fact, the emission inventory, and the selection of simulated area here only are the carrier to conduct this study. In the sensitive test, perturbations were added on the percentage of each component in source profile, hence the simulated result in each scenario could not match the observed data.

3. While the MEIC inventory includes four categories, i.e. power plants (PP), industrial processes (IN), residential emission (RE) and transport sector (TR), the SPA and SPE are shown to have different categories (perhaps more than the MEIC does). How were these chemical PM_{2.5} emission source profiles combined to match the MEIC categories? For instance, the residential emission should include not only coal burning but also straw burning, and the latter was seemly not considered in the simulations. Also, the chemical profiles for gasoline and diesel oil in the transport sector might be different.

Response:

Thank you for your valued advices. Just as you mentioned, in the database of Source Profiles of Air Pollution (SPAP) and U.S. Environmental Protection Agency's (EPA) SPECIATE database, these four source categories (coal-fired power plant, industry process,

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transportation sector and residential coal combustion) contain a series of sub-categories. But unfortunately, the MEIC inventory does not include the corresponding sub-categories. So we take the average values of all source profiles in each source category as representing source profile, the details could be seen in our previous work (Bi et al., 2019); Then multiply inventory emissions by profile fraction to get emissions of specific chemical compounds. The general step for speciation is shown in Fig. RF4.



Fig. RF4 Speciation in general step

Source: International Emissions Inventory Conference. SPECIATE and using the Speciation Tool to prepare VOC and PM chemical speciation profiles for air quality modeling, p31. https://www.epa.gov/sites/default/files/2017-10/documents/speciate_speciationtool_training.pdf.

In our study, we found that the simulated concentration of $PM_{2.5}$ components, not only primary components but also secondary components, indeed varied with the source profiles. The representativeness and timeliness of source profile should be considered due to the underappreciated impact of emission source profiles on the simulation of $PM_{2.5}$ components. Thank you for your valuable comments, we will deeply

discuss the influence of sub-source profiles on the simulation results in the follow-up study.

More descriptions have been added in Section 6 (Lines 558-562 in the revised manuscript).

Our study tentatively discussed the impact mechanism of emission source profiles on $PM_{2.5}$ components simulation results in CTMs. In the next work, we will use different source profile for simulation, compare the simulation results with local measured $PM_{2.5}$ components and discuss the influence of sub-source profiles variation on the simulation results. In addition, the size distribution, mixing state, aging and

4. How are the dynamic, microphysical and chemical processes of aerosols treated in the CMAQ model used for this study? Are the size distribution, mixing state, aging and solubility taken into account for different aerosol components? By which molecular form are the chemical components (Al, Ca, Cl, EC, Fe, K, Mg, Mn, Na, OC Si, NH₄⁺, NO₃⁻, and SO₄²⁻) emitted from the sources? Taking elemental Ca as an example, it should be emitted by CaO, CaCO₃, CaSO₄, or other compound, rather than merely by the cation Ca²⁺. The similar principle applies for anions (NO₃⁻ and SO₄²⁻). The difference in the exiting form of these emitted aerosol components might have large impacts on the thermodynamic equilibrium of ions in liquid aerosols and clouds.

Response:

Thank you for the reviewer's questions, please see the point-by-point response as follows:

For: How are the dynamic, microphysical and chemical processes of aerosols treated in the CMAQ model used for this study?

The key scientific algorithms simulating aerosol processes for the CCTM in CMAQ are: (1) aerosol removal by size-dependent dry deposition; (2) aerosol-cloud droplet interaction and removal by precipitation; (3) new particle formation by binary homogeneous nucleation in a sulfuric acid/water vapor system; (4) the production of an organic aerosol component from gas-phase precursors; and (5) particle coagulation and condensation growth (Byun and Young, 1999).

The particle dynamics of aerosol distribution using three interacting lognormal distributions, or modes. Two modes (Aitken and accumulation) are generally less than 2.5µm in diameter while the coarse mode contains significant amounts of mass above 2.5µm. The equation of lognormal distribution is as follow:

$$n(\ln D) = \frac{N}{\sqrt{2\pi} \ln \sigma_g} \exp\left[-0.5 \left(\frac{\ln \frac{D}{D_g}}{\ln \sigma_g}\right)^2\right]....(1)$$

Where *N* is the particle number concentration within the mode suspended in a unit volume of air, *D* is the particle diameter, D_g is the geometric mean diameter, σ_g is the geometric standard deviation of modal distribution. A brief summary is described by (Binkowski and Roselle, 2003) and fully described by (Whitby and McMurry, 1997). The aerosol species of PM_{2.5} in CMAQ are listed in Table RF5.

CMAQ species	Description
AECI, AECJ	Aitken (I) and accumulation (J) mode EC mass
APOCI, APOCJ	Aitken (I) and accumulation (J) mode OC mass
APNCOMI, APNCOMJ	Aitken (I) and accumulation (J) mode primary non-carbon organic matter mass
ASO4J	Accumulation (J) mode sulfate mass
ANO3J	Accumulation (J) mode nitrate mass
ACLJ	Accumulation (J) mode particulate chloride mass
ANH4J	Accumulation (J) mode particulate ammonium mass
ANAJ	Accumulation (J) mode sodium mass
AKJ	Accumulation (J) mode potassium mass
AMGJ	Accumulation (J) mode magnesium mass
ACAJ	Accumulation (J) mode calcium mass
AFEJ	Accumulation (J) mode iron mass
AMNJ	Accumulation (J) mode manganese mass
AALJ	Accumulation (J) mode aluminum mass
ASIJ	Accumulation (J) mode silicon mass
ATIJ	Accumulation (J) mode titanium mass
AH2OJ	Accumulation (J) mode particulate water mass
AOTHRJ	Accumulation (J) mode remaining unspeciated fine mode
	primary PM mass

Table RF5 Aerosol species of PM2.5 in CMAQ

The aerosol microphysics i.e. coagulation, condensation, new particle formation, deposition, etc.) are considered in CMAQ using aero_subs.F, aero_depv.F, coags.f, in CCTM module correspondingly. The microphysical process and the related numerical simulation in subroutines called by the CMAQ driver are covered in more detail in the literatures (Binkowski and Roselle, 2003; Byun and Young, 1999).

The aerosol chemical species are listed in Table RF4. ISORROPIA v2.2 in the reverse mode are used to calculate the condensation/evaporation of volatile inorganic gases to/from the gas-phase concentrations of coarse particle surfaces. ISORROPIA v2.2 is also used in the forward mode to calculate instantaneous thermodynamic equilibrium between the gas and fine-particle modes. The equilibria and the associated constants are shown in Table RF6.

Number	Reaction	<i>K</i> ⁰ (298.15K)
I1	$\operatorname{Ca(NO_3)}_{2(s)} \leftrightarrow \operatorname{Ca}_{(aq)}^{2+} + 2\operatorname{NO}_{3(aq)}^{-}$	6.067×10 ⁵
12	$\operatorname{Ca(Cl)}_{2(s)} \leftrightarrow \operatorname{Ca}_{(aq)}^{2+} + 2\operatorname{Cl}_{(aq)}^{-}$	7.974×10 ¹¹
13	$CaSO_4 \cdot 2H_2O_{(s)} \leftrightarrow Ca^{2+}_{(aq)} + SO^{2-}_{4(aq)} + 2H_2O$	4.319×10 ⁻⁵
I4	$K_2SO_{4(s)} \leftrightarrow 2K_{(aq)}^+ + SO_{4(aq)}^{2-}$	1.569×10 ⁻²
15	$\mathrm{KHSO}_{4(\mathrm{s})} \leftrightarrow \mathrm{K}^{+}_{(\mathrm{aq})} + \mathrm{HSO}^{-}_{4(\mathrm{aq})}$	24.016
16	$\mathrm{KNO}_{\mathfrak{Z}(\mathfrak{s})} \longleftrightarrow \mathrm{K}^+_{(\mathrm{aq})} + \mathrm{NO}^{\mathfrak{Z}(\mathrm{aq})}$	0.872
Ι7	$\mathrm{KCl}_{(s)} \leftrightarrow \mathrm{K}^{\scriptscriptstyle +}_{(\mathrm{aq})} + \mathrm{Cl}^{\scriptscriptstyle -}_{(\mathrm{aq})}$	8.680
18	$MgSO_{4(s)} \leftrightarrow Mg^{2+}_{(aq)} + SO^{2-}_{4(aq)}$	1.079×10 ⁵
19	$Mg(NO_3)_{2(s)} \leftrightarrow Mg_{(aq)}^{2+} + 2NO_{3(aq)}^{-}$	2.507×10^{15}
I10	$Mg(Cl)_{2(s)} \leftrightarrow Mg^{2+}_{(aq)} + 2Cl^{-}_{(aq)}$	9.557×10^{21}

Table RF6 Equilibrium relations and Constants

I11	$\mathrm{HSO}_{4(\mathrm{aq})}^{-} \longleftrightarrow \mathrm{H}_{(\mathrm{aq})}^{+} + \mathrm{SO}_{4(\mathrm{aq})}^{2-}$	1.015×10 ⁻²
I12	$\mathrm{NH}_{\mathfrak{Z}(g)} \leftrightarrow \mathrm{NH}_{\mathfrak{Z}(\mathrm{aq})}$	57.64
I13	$\mathrm{NH}_{\mathfrak{Z}(\mathrm{aq})} + \mathrm{H}_{2}\mathrm{O}_{(\mathrm{aq})} \leftrightarrow \mathrm{NH}_{\mathfrak{Z}(\mathrm{aq})}^{+} + \mathrm{OH}_{(\mathrm{aq})}^{-}$	1.805×10-5
I14	$HNO_{3(g)} \leftrightarrow H^{+}_{(aq)} + NO^{-}_{3(aq)}$	2.511×10^{6}
I15	$\operatorname{HNO}_{3(g)} \leftrightarrow \operatorname{HNO}_{3(aq)}$	2.1×10^{5}
I16	$\mathrm{HCl}_{(g)} \leftrightarrow \mathrm{H}^{\scriptscriptstyle +}_{(\mathrm{aq})} + \mathrm{Cl}^{\scriptscriptstyle -}_{(\mathrm{aq})}$	1.971×10^{6}
I17	$\mathrm{HCl}_{(\mathrm{g})} \leftrightarrow \mathrm{HCl}_{(\mathrm{aq})}$	2.5×10^{3}
I18	$H_2O_{(aq)} \leftrightarrow H^+_{(aq)} + OH^{(aq)}$	1.010×10 ⁻¹⁴
I19	$Na_2SO_{4(s)} \leftrightarrow 2Na^+_{(aq)} + SO^{2-}_{4(aq)}$	0.4799
I20	$(\mathrm{NH}_4)_2\mathrm{SO}_{4(s)} \leftrightarrow 2\mathrm{NH}^+_{4(\mathrm{aq})} + \mathrm{SO}^{2-}_{4(\mathrm{aq})}$	1.817
I21	$\mathrm{NH}_4\mathrm{Cl}_{(\mathrm{s})} \leftrightarrow \mathrm{NH}_{\mathrm{3(g)}} + \mathrm{HCl}_{(\mathrm{g})}$	1.086×10 ⁻¹⁶
I22	$NaNO_{3(s)} \leftrightarrow Na^{+}_{(aq)} + NO^{-}_{3(aq)}$	11.97
I23	$NaCl_{(s)} \leftrightarrow Na_{(aq)}^{+} + Cl_{(aq)}^{-}$	37.66
I24	$NaHSO_{4(s)} \leftrightarrow Na^{+}_{(aq)} + HSO^{-}_{4(aq)}$	2.413×10 ⁴
I25	$\mathrm{NH}_4\mathrm{NO}_{3(\mathrm{s})} \leftrightarrow \mathrm{NH}_{3(\mathrm{g})} + \mathrm{HNO}_{3(\mathrm{g})}$	4.199×10 ⁻¹⁷
I26	$\mathrm{NH}_{4}\mathrm{HSO}_{4(\mathrm{s})} \leftrightarrow \mathrm{NH}_{4(\mathrm{aq})}^{+} + \mathrm{HSO}_{4(\mathrm{aq})}^{-}$	1.383
I27	$(\mathrm{NH}_4)_3\mathrm{H}(\mathrm{SO}_4)_{2(\mathrm{s})} \leftrightarrow 3\mathrm{NH}_{4(\mathrm{aq})}^+ + \mathrm{HSO}_{4(\mathrm{aq})}^- + \mathrm{SO}_{4(\mathrm{aq})}^{2-}$	29.72

Source: (Fountoukis and Nenes, 2007)

Besides that, for a higher computational efficiency, a VBS-style approach (four surrogate species with specific vapor pressures) is widely used in models; For the nonvolatile POA configuration, mass is tracked separately in terms of its carbon (OC) and non-carbon (NCOM) content. With this approach in AERO6, mass can be added to the non-carbon species to simulate the aging of POA in response to atmospheric oxidants. Details are shown in CMAQ users guide (chapter 6, https://github.com/USEPA/CMAQ/blob/main/DOCS/Users_Guide/CMA Q_UG_ch06_model_configuration_options.md#6.11_Aerosol_Dynamics) and the literature (Binkowski and Roselle, 2003).

For: Are the size distribution, mixing state, aging and solubility taken into account for different aerosol components?

Yes, they are all taken into account for different aerosol components. For size distribution, taking $PM_{2.5}$ as an example, except for a very small fraction of OC, EC and non-carbon organic matter are allocated in the Aitken mode, the rest are allocated in the accumulation mode. The size distribution of different $PM_{2.5}$ components are shown in Table RF7.

Name	Aitken (I)	Accumulation (J)	Coarse (K)
EC	0.001	0.999	0
OC	0.001	0.999	0
NCOM	0.001	0.999	0
SO4 ²⁻	0	1	0
NO ₃ -	0	1	0
Cl ⁻	0	1	0
$\mathrm{NH_{4}^{+}}$	0	1	0
Na	0	1	0
K	0	1	0
Mg	0	1	0
Ca	0	1	0
Fe	0	1	0
Mn	0	1	0
Al	0	1	0

Table RF7 The size distribution of different PM_{2.5} components

Si	0	1	0
Ti	0	1	0
H_2O	0	1	0
Other	0	1	0

Source: AERO_EMIS.F in CCTM module

As regards as mixing state, internal mixing of the aerosol species is assumed within each mode (Aitken mode, accumulation mode, and coarse mode), the modes themselves were externally mixed (Han et al., 2011).

As to aerosol aging, the subroutine of poaage.F in AERO module calculates oxidative aging of POA using the following reaction (Table RF8):

Table RF8 Oxidative aging of POA

For solubility, the system modeled by ISORROPIA II consists of the following potential components: Gas phase: $NH_3(g)$, $HNO_3(g)$, HCl(g), $H_2O(g)$; Liquid phase: $NH_4^+(aq)$, $Na^+(aq)$, $H^+(aq)$, $Cl^-(aq)$, $NO_3^-(aq)$, SO_4^{2-} (aq), $HNO_3(aq)$, $NH_3(aq)$, HCl(aq), $HSO_4^-(aq)$, $OH^-(aq)$, $H_2O(aq)$, $Ca^{2+}(aq)$, $K^+(aq)$, $Mg^{2+}(aq)$; Solid phase: $(NH_4)_2SO_4(s)$,

 $NH_4HSO_4(s),(NH_4)_3H(SO_4)_2(s), NH_4NO_3(s), NH_4Cl(s), NaCl(s), NaNO_3(s), NaHSO_4(s), Na_2SO_4(s), CaSO_4(s), Ca(NO_3)_2(s), CaCl_2(s), K_2SO_4(s), KHSO_4(s), KNO_3(s), KCl(s), MgSO_4(s), Mg(NO_3)_2(s), MgCl_2(s); where the subscripts (g), (aq), (s) denote gas, aqueous and solid, respectively.$

For: By which molecular form are the chemical components (Al, Ca, Cl, EC, Fe, K, Mg, Mn, Na, OC Si, NH_4^+ , NO_3^- , and SO_4^{2-}) emitted from the sources? Taking elemental Ca as an example, it should be emitted by CaO, CaCO₃, CaSO₄, or other compound, rather than merely by the cation Ca^{2+} . The similar principle applies for anions (NO_3^- and SO_4^{2-}). The difference in the exiting form of these emitted aerosol components might have large impacts on the thermodynamic equilibrium of ions in liquid aerosols and clouds.

Generally, the PM samples emitted from the sources are collected on Teflon and quartz fiber filters and then sent for chemical component analysis. Elements analysis uses Teflon filters, common chemical analysis instruments are: inductively coupled plasma optical emission spectrometer (ICP-OES), inductively coupled plasma atomic emission spectrometer (ICP-AES), inductively coupled plasma mass spectrometer (ICP-MS) instruments and X-ray fluorescence. The total carbon (TC) mass in the samples are typically determined using thermal or thermal–optical methods. There are two widely utilized approaches to dividing OC and EC from TC, known as IMPROVE_A (from the Desert Research Institute– DRI) and NIOSH (method 5040; from the National Institute for Occupational Safety and Health – NIOSH), which are operationally defined by the time– temperature protocols, and the OC–EC split point is determined by optical reflectance/transmittance (Ho et al., 2003; Bi et al., 2019). PM samples collected on the quartz fiber filters are normally used for the determination of water-soluble inorganic ions via different types of ion chromatography (IC) with high-capacity cation-exchange and anion-exchange columns. In addition, the molecular form of particulate matter emitted by pollution sources is difficult to measure. Hence, data form in emission source profiles are chemical components NOT chemical compounds.

The emission input files for CTMs are generated from data provided by emission inventories, only the species that are specifically defined in the chemical mechanism will be included in model inputs and outputs. PM need to be speciated into chemical components for CTMs to match chemical mechanism, and the emission source profiles can provide "species" and "split factor" (Detail is shown in Fig. RF6). The species for PM_{2.5} in mainstream CTMs are listed in Table RF2. The process of modeling speciation requires components **rather than chemical compounds.** In addition, the "split factor" is species allocation, not about particle mode; In our previous measurement, we found the particle size distribution of different source profiles are quite different and some species

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are bimodal or multimodal (Fig. RF5) (Li et al., 2021; Du et al., 2021; Ma et al., 2016; Lyu et al., 2021), which is differ from the size distribution divided by CMAQ (In the CMAQ work, the assumption is that major part of $PM_{2.5}$ particulate mass emissions are in the accumulation mode with a small fraction in the Aitken mode; i.e. a fraction of 0.999 of $PM_{2.5}$ is assumed to be in the accumulation mode and the remaining fraction, 0.001, is assigned to the Aitken mode., detail are shown in Table RF7).





Fig. RF5 Mass concentration and percentage of species emitted from some sources.

Thank you for your valuable comments. Our study tentatively discussed the impact mechanism of emission source profiles on PM_{2.5} components simulation results in CTMs. We found the influences are connected to model chemical mechanisms since the variation of species allocations in emission sources directly affected the thermodynamic equilibrium system. We will continue exploring the influence of source profile changes on aerosol particle size distribution, microphysical and chemical processes in a follow-up study.

More descriptions have been added in Section 6 (Lines 562-565 in the revised manuscript).

562 on the simulation results. In addition, the size distribution, mixing state, aging and 563 solubility for different aerosol components might have something to do with source 564 profile, how much the influence of source profile changes on these physical and 565 chemical process, is deserved to do in the future.

Tuble KI 2 The spectrue anocation for TM _{2.5} in mainstream CTM5				
CTMs	Aerosol module	PM _{2.5} species		
		Al, Ca, Cl, EC, Fe, K, Mg,		
		Mn, Na, OC, Si, Ti, NH4 ⁺ ,		
$CMAQ^1$	AEKU0	NO ₃ ⁻ , SO ₄ ²⁻ , NCOM, Other,		
		H_2O		
	AERO5	OC, EC, NO ₃ ⁻ , SO ₄ ²⁻ , Other		
		Al, Ca, Cl, EC, Fe, K, Mg,		
GEOS-Chem ²	aerosol.mod	Mn, Na, OC, Si, Ti, NH4 ⁺ ,		
		NO_3^- , SO_4^{2-} , Other		
WDE $Chom^3$	MADE MOSAIC MAM	OC, EC, NO ₃ ⁻ , SO ₄ ²⁻ , Ca, Na,		
WKF-Clielli	MADE, MOSAIC, MAM	Cl, H ₂ O, Other		
CAM ₂ 4	CE	OC, EC, NO ₃ ⁻ , SO ₄ ²⁻ , NH ₄ ⁺ ,		
CAMIX	CF	Cl, Na, Other		

Table RF2 The speciated allocation for PM_{2.5} in mainstream CTMs

Note:

1, Particulate matter (aerosols): PM using three lognormal sub-distributions, or modes, two interacting modes (Aitken and accumulation) represent $PM_{2.5}$

https://www.airqualitymodeling.org/index.php/CMAQ_version_5.0_(February_2010_release) _OGD#Aerosol_Module.

2, Particulate matter in GEOS-Chem: $PM_{2.5} = (NH4 + NIT + SO4) * 1.10 + BCPI + B$ CPO + (OCPO + (OCPI * 1.05)) * (OM/OC ratio) + DST1 + DST2 * 0.30 + SAL A * 1.86 + SOA * 1.05. (NIT-NO3; BCPI and BCPO-EC; OCPO and OCPI-OC, NCO M; DST1-SO4, NH4, NO3, Cl, Na, K, Ca, Fe, Al, Si, Ti, Mn, Other, OC, NCOM; DST2-SO4, Cl, ASOL; SALA-SO4, Cl, Na, Mg, K, Ca.

http://wiki.seas.harvard.edu/geos-chem/index.php/GEOS-Chem_to_CMAQv5.0) http://wiki.seas.harvard.edu/geos-chem/index.php/Particulate_matter_in_GEOS-Chem.

3, Aerosols in WRF-Chem: PM using 3 or 7 log-normal modes, two interacting modes (Ai tken and accumulation) represent $PM_{2.5}$.

https://ruc.noaa.gov/wrf/wrf-chem/wrf_tutorial_2018/Aerosols.pdf.

4, Aerosol Chemistry: PM_{2.5} = PSO4 + PNO3 + PNH4 + PEC + NA + PCL + POA + SOA1 + SOA2 + SOA3 + SOA4 + SOPA + SOPB + FPRM + FCRS + (PFE + PMN + PK + PCA + PMG + PAL + PSI + PTI) (Fe, Mn, K, Ca, Mg, Al, Si and Ti are Optional Species).

https://camx-wp.azurewebsites.net/Files/CAMxUsersGuide_v7.20.pdf.

						•
species	species		Prescribed Bu	Prescribed Burning – Co		
name	description	AES	AED		pollutant	pollutant species
POC	organic carbon	Y	Y		DM2 5	
PEC	elemental carbon	Y	Y		PM2_5	PM2_5 POC
PSO4	sulfate	Y	Y		PM2_5	PM2_5 PEC
PNO3	nitrate	Y	Y		PM2_5	PM2_5 PSO4
PMFINE	unspeciated PM2.5	Y	N		PM2 5	PM2.5 PNO3
PNH4	ammonium	N	Y			
	non-carbon organic				PM2_5	PM2_5 PNH4
PNCOM	matter	N	Y		PM2_5	PM2_5 PAL
PFE	iron	N	Y		PM2 5	PM2 5 PCA
PAL	aluminum	N	Y		DM2 5	
PSI	silica	N	Y			
PTI	titanium	Ν	Y		PM2_5	PM2_5 PFE
PCA	calcium	N	Y		PM2_5	PM2_5 PK
PMG	magnesium	N	Y		PM2_5	PM2_5 PMN
РК	potassium	N	Y		PM2 5	PM2 5 PMOTHR
PMN	manganese	N	Y			
PNA	sodium	N	Y		PMZ_5	PMZ_5 PNA
PCL	chloride	N	Y		PM2_5	PM2_5 PNCOM
BUDO					PM2_5	PM2_5 PSI
PHZO	water	N	Y		PM2 5	PM2 5 PTI
PMOTHR	unspeciated PM2.5	N	Y			

Example modeling speciation

profile – AE6

Model species definitions

Fig. RF6 PM_{2.5} speciation- Modeling profile example

Source: International Emissions Inventory Conference. SPECIATE and using the Speciation Tool to prepare VOC and PM chemical speciation profiles for air quality modeling, p31. https://www.epa.gov/sites/default/files/2017-10/documents/speciate_speciationtool_training.pdf.

5. In Sect. 1 and Table S1, the deviations of $PM_{2.5}$ components simulated by CMAQ are presented. All these components (NH_4^+ , NO_3^- , SO_4^{2-} , and part of OC), except for EC and part of OC, are second aerosols, and their loadings in the atmosphere are controlled primarily by the emissions of gaseous precursors, instead of the emission of aerosols. The presentation here and associated arguments seems to be misleading as the effect of uncertainties in the gaseous emissions is not considered in this study.

Response:

Thank you for your advices. One of the important sources of these atmospheric components (NH_4^+ , NO_3^- , SO_4^{2-} , and OC) is formed by chemical conversion of gaseous precursors, which are second aerosols. But they still have some primary sources, a number of recent studies found that, primary emission may be also important. These components $(NH_4^+,$ NO_3^- , SO_4^{2-} , and OC) exist in primary emission sources such as coal-fired power plant, industry process, transportation sector and residential coal combustion, the detail is shown in Fig. RF7 (Fig. 2~ 5 in manuscript); For example, sulfate (a major PM2.5 component) is largely from primary emissions rather than secondary formation in ambient air in certain circumstances (Chen et al., 2017; Dai et al., 2019; Ding et al., 2021; Ding et al., 2019; Li et al., 2017; Yang et al., 2020; Yan et al., 2020), its weight percentage variation range is 0.7~71% in coal-fired power plant, 0.03%~40% in industry process, 0.02~40% in transportation sector, 1~40% in residential coal combustion, respectively.



Fig. RF7 Chemical profiles for PM2.5 emitted from coal-fired power plant (PP), industry processes (IN), transportation sector (TR), residential coal combustion (RE).

In our study, we found source profile variation could affect the simulation result of secondary components, **they could lever the whole aerosol equilibrium system.** The effects of source profile variation on the simulation results of different components were linked. When the percentages of Non-SNA, SO_4^{2-} , NO_3^{-} and NH_4^{+} in the source profile changed, they not only affected the simulated concentration of themselves, but also affected the simulation results of some other components through the thermodynamic equilibrium system (ISORROPIA II, SO_4^{2-} - NO_3^{-} - Cl^{-} - NH_4^{+} - Na^{+} - K^{+} - Mg^{2+} - Ca^{2+} - H_2O system). Section 5 in our manuscript

focused on these performances: in the sensitivity tests, when we only perturb the $PM_{2.5}$ source profile (primary emission) but not the emission inventory of gaseous precursors, the simulated result of secondary $PM_{2.5}$ components also changed, this side-fact indicates the crucial role of primary $PM_{2.5}$ components on the simulation of second components formation in CTMs.

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