



Impact of contrasting fertilizer technologies on N dynamics from sub-surface bands of ‘pure’ or blended fertilizer applications

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Abstract. Enhanced efficiency fertilizer (EEF) technologies that employ product coatings to delay nitrogen (N) release or are chemically stabilized to inhibit key steps of N transformations in soil, offer potential for improving N use efficiency (NUE) in agricultural systems. However, the dynamics of N release and transformation from single technologies may result in a spatial or temporal mismatch of N supply and demand during a growing season. This may be overcome by use of blends of different technologies, provided the reduction in the concentration of stabilizing products does not reduce effectiveness. Laboratory incubations quantified the N dynamics around fertilizer bands of polymer-coated urea (PCU) and nitrification inhibited (NI) urea and varying blends of these technologies, and referenced this against conventional urea and biodegradable, plant oil-coated urea (POCU) applied at the same rates in two contrasting soils over 60 days. Blends of NI-urea and PCU typically resulted in N concentrations and distribution that were intermediate to that of the constituent products in unblended applications. Changes in the proportions of each product were mirrored by urea-N concentrations around the bands in both soils, while the proportions of NI-urea in each blend were only related to the extent of nitrification inhibition in the Vertisol. A proportion of the POCU granules burst during initial water imbibition, resulting in initially higher mineral N concentrations *cf.* PCU. However, both CRFs delayed N release and generation of NO₃-N relative to granular urea, and mineral N distribution was similar within each soil. Soil type had a significant impact on banded N dynamics. Where there was little effect of N-fertilizer treatment on NO₃-N production in the Ferralsol, the greater impedance of solutes in the Vertisol contributed to a significant inhibitory effect of NI-urea on nitrification in both pure and blended DMPP-urea treatments. Using NO₃-N production as a benchmark for the risk of environmental loss, the efficacy of fertilizer treatments in this soil was of: DMPP-urea-PCU blends (higher ratio of PCU may offer small but insignificant benefit) > DMPP-urea = PCU > urea. These findings highlight the importance of soil properties in determining the N dynamics from different banded EEF products. Insights into the efficacy of biodegradable alternatives to polymer coatings and the efficacy of blended EEF products can improve the reliability of N supply while reducing environmental impacts, therefore offering greater opportunities to sustainably improve fertilizer NUE in cropping systems.



1 Introduction

Enhanced efficiency nitrogen (N) fertilizers (EEFs) are widely proposed as a key tool for meeting the competing demands of (i) improving agricultural productivity and (ii) reducing the impacts of agricultural activity on nearby environments. Broadly, these challenges involve improving N use efficiency (NUE) in agricultural systems. Current EEF technologies are
35 deployed in fertilizer products which utilize controlled-release or nutrient stabilizing mechanisms to regulate fertilizer N transformations in soil to enhance crop uptake of N and reduce potential losses to natural environments (AAPFCO, 2013).

Controlled-release mechanisms aim to regulate the duration, rate and pattern of N release (Shaviv, 2001), providing a predictable supply of labile N under a given set of environmental conditions. The coating material of controlled-release fertilizers (CRFs) has evolved over time, with polymer-coated products currently being the most widely utilized form of this
40 technology in agriculture. Recently, concerns of polymer “shell” persistence and the potential for these to carry unreleased N into natural ecosystems (Azeem et al., 2014; Trenkel, 2010) has led to investigation of a range of biodegradable coatings (Araujo et al., 2017; Li et al., 2016; Rotondo et al., 2018; Vadjung and Saengsuwan, 2018). Environmental concerns will continue to drive the development of biodegradable coatings in future CRF products, but the efficacy of these coatings for delivering a controlled release of N has only been explored in a very limited number of studies (e.g., Martinez et al., 2021).

Stabilized fertilizers function by reducing the rate of microbial activity at specific steps of N transformation, thereby
45 regulating the formation of N products with greater risks of loss (Trenkel, 2010). Commercial stabilizing products currently target either of two key processes, viz. the hydrolysis of urea to ammonium carbonate ($[\text{NH}_4]\text{CO}_3$) with urease inhibitors (UIs), and the oxidisation of ammonia (NH_3) to nitrite (NO_2^-) with nitrification inhibitors (NIs). A wide range of compounds inhibit urease (Byrnes and Freney, 1995; Medina and Radel, 1988) and nitrification (Juliette et al., 1993; McCarty, 1999;
50 McCarty and Bremmer, 1986; 1989) activities, with N-(n-butyl) thiophosphoric triamide (NBPT) and 3, 4 dimethylpyrazole phosphate (DMPP) being the most widely used compounds in urease and nitrification inhibitor formulations, respectively.

While the function and efficacy of EEFs is relatively well-defined under controlled-conditions (Azam et al., 2001; Benckiser et al., 2013; Creason et al., 1990; Du et al., 2006; Guardia et al., 2018; Shaviv et al., 2003), field performance of these technologies have not demonstrated consistent outcomes for yield improvement, N uptake and reductions in N losses
55 (Dang et al., 2021; LeMonte et al., 2016; Lester et al., 2016; Li et al., 2015; Martinez et al., 2021; Nauer et al., 2018; Rowlings et al., 2016; Schwenke and Haigh, 2019; Wang et al., 2016b). This inconsistency has eroded confidence in EEFs to deliver improved NUE. In particular, concerns have been raised that the delivery of N from CRFs may be too slow to effectively meet early and continued crop demand (Halvorson and Grosso, 2013; Nelson et al., 2009; Quemada et al., 2013; Thapa et al., 2016). Conversely, questions of inhibitor effectiveness have been raised where inhibition appears to be
60 relatively transient (Chen et al., 2015; Fu et al., 2018; Shi et al., 2016). These concerns, combined with the higher fertilizer costs (per kg N) associated with EEF technology, raise real concerns about the effectiveness and the economic returns that can be derived from the use of products reliant on individual EEF technologies. In response, attempts have been made to



blend the cheaper, readily soluble N products with variable proportions of EEF products to mitigate at least some of the potential N loss while minimising the financial risk.

65 Multiple field experiments have investigated the potential for blends and varying the ratios of urea and CRFs (typically, polymer coated urea [PCU]) to deliver improved NUE (Dhakai and Nelson, 2019; Guo et al., 2017; McKenzie et al., 2010; Nash et al., 2012; Noellsch et al., 2009; Patil et al., 2010; Payne et al., 2015; Wang et al., 2015). Whilst most of these experiments have found that blends of PCU and urea may reduce N losses (*cf.* urea alone), not all have reported positive outcomes for improved crop uptake and / or yield gains (McKenzie et al., 2010; Nash et al., 2012; Patil et al., 2010).
70 Where benefits have not been recorded, ‘failures’ have been attributed to a disparity in N supply relative to crop demand (i.e., not enough N available early in the season due to the delay in N supply from PCU) or conditions where N loss (of both urea and PCU-urea blends) has been negligible and the addition of PCU has therefore not provided any advantages relative to urea alone. In the instance of poor early-season N supply from PCU-urea blends, it is hypothesized that the urea component of the applied fertilizer may have been unable to supply adequate N to the crop as a result of rapid N loss from
75 this more soluble component of the blend (i.e., uncoated urea) under conditions conducive of N loss. Under such high loss conditions, the substitution of N-stabilized products for the more ‘vulnerable’ urea in blends with PCU may provide additional protection against immediate leaching and / or denitrification losses. The combination of NI-urea and PCU in fertilizer bands may therefore create a more ‘stable’ and sustained supply of N *cf.* PCU-urea blends or urea alone, but this approach has not been well tested. The concept of blending NI-urea and PCU products within fertilizer bands as a way of
80 improving crop NUE in the Australian sugar industry was tested in a pilot field program by (Bell et al., 2021), and in a much larger investigation by Connellan et al. (2021). Sugarcane crops using these blends were consistently able to achieve similar crop N uptake to urea applications at 20 - 25% higher application rates, although yield responses were variable in some situations and rarely exceeded those achieved from using either NI alone, or a blend of urea and PCU applied at rates chosen to minimise any price premiums paid for the PCU component (Connellan et al. 2021). Other evaluations of blended EEF
85 products are limited, and the extent to which these reports of improved fertilizer-blend efficacy in sugarcane can be extrapolated to other soils, climates and cropping systems is limited by a lack of mechanistic studies that seek to understand any potential benefits and pitfalls arising from the use of blended EEF technologies in concentrated fertilizer bands. This study seeks to address this knowledge gap by determining the release, transformation rates and distribution of N from bands of granular urea, CRFs, stabilized urea and blends of these products in contrasting soils under controlled-environment
90 conditions. A plant-oil coated CRF (Kingenta, 2018) is also included in this investigation for preliminary examination of N release dynamics from an example of a biodegradable fertilizer coating. This study aims to provide a mechanistic understanding of fertilizer-blend dynamics which may be used to predict the potential efficacy of different EEF blends in contrasting soils that underpin a variety of crop industries.



2 Methods

95 2.1 Soil collection, characterization, and preparation

Soil from the top 10 cm of the profile of a yellow Ferralsol and black Vertisol (IUSS Working Group, 2015) were collected from sugarcane farms in the Bundaberg district. The soils were chosen on the basis of their contrasting texture and physico-chemical properties (Table 1). Briefly, the Ferralsol was slightly acidic and had a low clay content, cation exchange capacity (CEC), and pH buffering capacity (pHBC). The Vertisol had a high clay content, CEC and pHBC, and a neutral pH. Prior to
100 incubation, soil was prepared by air-drying and sieving to < 2 mm. The moisture content at field capacity for both soils was determined using the column method described in Asher et al. (2002).

2.2 Incubation experiment

Prepared soil (air dry, sieved to < 2 mm) was added to a depth of 50 mm in round incubation pots (225 mm diameter PVC endcaps). Bulk density was adjusted to ca. 1.2 g cm³ by tapping the pot base ten times on the bench, which settled the soil to
105 a depth slightly below 5 cm. Deionized water was added to each pot to bring the soil mass to the pre-calculated field capacity moisture content (0.32 g g⁻¹ and 0.12 g g⁻¹ for the Vertisol and Ferralsol, respectively), taking into account the initial moisture content of air-dry soil. Pots were sealed with plastic cling wrap and the moisture was allowed to equilibrate overnight. Following moisture application, the Vertisol swelled approximately 0.5 cm in height, reducing the bulk density of this soil to ca. 1.1 g cm³. Treatments of N-fertilizer were then applied by removing a 10 mm diameter core of soil, inserting
110 the various fertilizer treatments to a depth of 50 mm, and then covering the fertilizer column with some of the previously removed soil to limit volatilization losses. All treatments were applied at a rate equivalent to the weight of fertilizer in a 50 mm section of a N-fertilizer band applied at 150 kg N ha⁻¹ at 1.8 m row spacing (1350 mg N pot⁻¹), with this rate and concentration chosen to reflect typical application rates and band spacings in the Australian sugarcane industry (i.e., Six Easy Steps approach; SRA, 2018). The N-fertilizer treatments described in this paper include: (i) granular urea; (ii) granular
115 urea coated with the NI DMPP and marketed as ENTEC[®] (DMPP-urea); (iii) polymer coated urea granules with a 90-day release duration and marketed as Agromaster Tropical[®] (PCU); (iv) urea granules with a biodegradable coating and an 80-day release duration supplied by Kingenta Australia Ag PTY LTD as Plant Oil Coated Urea (POCU); (v) a 1:2 fertilizer blend of DMPP-urea and PCU (1:2 DMPP-PCU); and (vi) 2:1 fertilizer blend of DMPP-urea and PCU (2:1 DMPP-PCU).

A moist glass fibre filter paper (125 mm diameter) was placed centrally on top of the soil to maintain humidity within the pot and allow for moisture adjustment without disturbing soil. Incubation pots were weighed to allow for on-going moisture
120 adjustment, with moisture maintained at field capacity ± 0.5 % (moisture adjustment every 7 days). Incubation pots were sealed with plastic wrap and three pinholes were inserted to prevent the development of anaerobic conditions. The incubation was maintained at 23±1°C and duplicate pots were placed in a completely randomised design across two benches.



2.3 Sample collection

125 Destructive sampling occurred at 10, 35 and 60 days after incubation initiation (DAI). Soil in each pot was collected
concentrically beginning with a 20 mm diameter central core (designated the '0-10 mm' position), and then in increments
moving outwards from that central core designated as the 10-30 mm, 30-50 mm, 50-70 mm, 70-90 mm and 90-110 mm
zones. These samples were collected using stainless steel sampling rings (diameters of 20, 40, 60, 80, and 100 mm) that were
pushed down through the soil to isolate and progressively remove each ring of soil, beginning from the central 20 mm core,
130 which contained the fertilizer band (the '0-10 mm' sample or 'fertosphere'). Soil in each increment was completely removed
from incubation pots and mixed to homogenize, with 2×10 g samples subsequently taken for analysis. One of the two soil
sub-samples was used for pH and EC measurement (1:5 soil:water) and subsequently extracted with 2M KCl for analysis of
mineral N species. The other sub-sample was extracted with deionized water for analysis of urea-N concentration. In
fertilizer treatments containing controlled release products, intact granules were recovered prior to removal of soil samples
135 from incubation pots. The granule samples were kept frozen until analysis for total N.

2.4 pH and electrical conductivity measurements

The pH and EC of soil samples were measured in a 1:5 (w/v) soil-water solution by adding 50 mL DI water to 10 g soil.
Measurements for pH and EC were recorded after 1 minute of stirring. Values for pH and EC were arithmetically averaged
from duplicate samples.

140 2.5 Soil extraction

Extraction of mineral N followed standard methods (Method: 7C2a, Rayment and Lyons, 2010), using a 1:5 (w/v) soil-
solution extraction with 2M KCl (i.e., 10 g soil with 50 mL 2M KCl). Samples were shaken end-over-end for 1 hour and
then removed and allowed to settle for approximately 15 minutes. A 10 mL aliquot of the extraction solutions was taken and
centrifuged at 3700 rpm (RCF: $1760 \times g$) for 9 minutes, followed by pouring off and re-centrifuging of the supernatant. All
145 solutions were frozen until analysis. The solutions were analysed for NH_4^+ -N (Searle, 1974), NO_2^- -N and NO_x -N (Best,
1976; Bremner, 1965) by segmented flow analysis (SFA). Urea-N was extracted in 1:5 soil-water (w/v) solutions with the
addition of 100 μL of 0.02 M Ag_2SO_4 (urease inhibitor, equiv. $62.4 \mu\text{g g}^{-1}$ soil) added immediately after water addition to
soil samples (Douglas and Bremner, 1970). Samples were shaken end-over-end for 1 hour and then removed and allowed to
settle for approximately 15 minutes. A 10 mL aliquot of the extraction solutions was taken and centrifuged at 3700 rpm
150 (RCF: $1760 \times g$) for 9 minutes, followed by pouring off and re-centrifuging of the supernatant. All solutions were frozen
until analysis. The urea extracts were analysed on a Thermo Fisher Ultimate 3000 U-HPLC system using a Thermo Fisher
DAD 3000 diode array detector following modified versions of the methods detailed for urea-N (MicroSolv, 2010). In each
batch, urea spikes were added to an aliquot of sample extract to monitor and quantify any urea loss during the analysis
process, which was minimal. Previous experiments indicated that the high concentrations of urea-N generated in this study



155 resulted in solubility issues such that recoveries were incomplete when extracted from soil with 2M KCl solution. Consequently, water extractions from soil were considered appropriate for determination of this N species.

2.6 Granule analysis

The organic N content of a sub-sample (*ca.* 0.1 g) of intact fertilizer granules was determined by digestion in concentrated sulphuric acid followed by the semi-micro Kjeldahl procedures outlined in Brech (1976), Searle (1974), and Youden and
160 Steiner (1975). A second sub-sample (*ca.* 1.7 g) was analysed for mineral N content using segmented flow analysis (Best, 1976; Bremner, 1965; Searle, 1974) of solution containing crushed and dissolved granules (1:5 w/v). A separate incubation experiment, in which the conditions of the diffusion experiment were replicated on smaller scale, was conducted to determine the moisture uptake in controlled release products over time. At 10, 35 and 60 DAI, PCU and POCU granules were removed and gravimetric moisture was determined using oven drying methods at 40°C (Topp et al., 2007). The
165 moisture content was incorporated into calculations to accurately determine the remaining organic and inorganic N (identified in Kjeldahl digestions of wet samples and SFA analysis of granule solution extractions), which is expressed as a percentage of CRF-N applied.

2.7 Data transformation

Urea-N values from HPLC analysis were expressed as the solution concentration (mg urea-N L⁻¹) of extractions. Mineral N
170 values are reported as mg N kg soil⁻¹ as these data were obtained from soil extracts (1:5 w/v 2M KCl) and the mg N kg soil⁻¹ unit is comparable to that used in other studies. Concentrations of NO₃⁻-N were determined by calculating the difference between NO_x-N and NO₂⁻-N (Best, 1976; Bremner, 1965). Quantities of net NO₃-N were calculated by dividing the reported concentration (mg/L) for each sampled zone by the extractant volume and multiplying by the volume of soil in that zone. All zones were summed to calculate the net NO₃-N production in each pot (mg). The proportion of total N recovered was
175 determined by summing the total N (mg) in soil solution with that remaining in granules in CRF and CRF-blend N-fertilizer treatments. It was assumed that any N immobilisation or solid phase sorption was minimal and / or consistent across treatments. This value was divided by either the soil solution or granule N content (mg) to calculate the respective proportion of total N (%). Ammonia (aqueous NH₃) concentrations were calculated by applying the pK_a value of the NH₄⁺ ↔ NH₃ reaction at 25°C (Emerson et al., 1975), then determining the NH₃ (molarity) from known K_a, H⁺ (molarity), and NH₄⁺
180 (molarity) values in soil solution, as per Equation 1 (Erickson, 1985). Quantities of ammonia were calculated by determining the moles in solution, multiplying by the molecular weight of NH₃, and expressing as mg NH₃ L⁻¹.

$$NH_3 = \frac{K_a \times NH_4^+}{H^+} \quad Eq. 1$$



2.8 Statistical analyses

185 Analysis of chemical data ($\text{pH}_{1:5\text{w}}$ and $\text{EC}_{1:5\text{w}}$) and N (urea-N, mineral N and aqueous NH_3) concentrations was conducted by
a three-way analysis of variance (ANOVA) for each soil with time, treatment, and distance from the fertosphere as factors.
Granule N retention data was analysed by two-way ANOVA for each soil with time and fertilizer treatment as factors.
Tukey's HSD method ($P < 0.05$) was used to test means for significant interaction effects identified in ANOVAs. For CRF
granule N, urea-N, mineral N and aqueous NH_3 datasets, statistical variation is represented in figures by standard error bars
190 fitted to mean values at each point. Statistical differences for pH and EC are not presented within the manuscript as the
number of interactions makes graphical presentation difficult to interpret. Where appropriate, significant effects are noted in
the text, with the full statistical results available in Supplementary Material Part B.

3 Results and Discussion

195 The dynamics of DMPP-urea, PCU, and blends of the two reported in this study do not consider the impact of crop N
demand, as the experiment was designed to determine N dynamics in banded NI-urea/CRF mixtures in a broad context (i.e.,
not crop specific), relative to granular urea and the individual EEF technologies. Therefore, while these studies explore the
principles underpinning the different N dynamics observed in blends of EEF technologies, the lack of plant N demand may
limit the extrapolation of these findings to the broader portfolio of field trials examining the impacts of EEF technologies
200 and blends on NUE (e.g., Bell et al. 2021; Connellan et al. 2021).

3.1 Soil chemical changes vary with fertilizer form and soil type

Rapid hydrolysis of concentrated urea bands results in significant increases in soil pH, EC and aqueous NH_3 concentrations
(Janke et al., 2021). This effect was observed in both soils treated with uncoated N-fertilizers (urea and DMPP-urea) and to a
lesser extent in blended DMPP-urea and PCU treatments (Figs. 1 – 3). The slow release of small amounts of urea-N to soil
205 solution from the single product CRF treatments (PCU, POCU) meant that these treatments demonstrated a much smaller
zone of impact for ureolytic-induced increases in pH, EC and aqueous NH_3 , although peak changes within the fertosphere
zone did not appear to be any less severe than that of granular urea (Figs. 1 – 3). Further, the longer duration of release of
urea-N from CRFs (Fig. 4), and thus the extended period of urea hydrolysis, meant that the pH remained high within the
fertosphere over the 60-day incubation in treatments with CRFs, either alone or in blends (Fig. 1). In the Ferralsol, the effects
210 of the chemically 'hostile' zone extended as far as *ca.* 8 - 9 cm from the fertosphere in the granular urea and DMPP-urea
treatment, with this zone reduced by 1 – 2 cm for the DMPP-urea/PCU blends and 3 – 4 cm for the two CRFs (Fig. 1, 2). The
higher pHBC and CEC of the Vertisol (Table 1) buffered these chemical changes more effectively, and combined with the
reduced permeability of this soil, resulted in a reduced zone of impact and less severe and more rapidly reversible pH and
EC increases. For the Vertisol, the effects of rapid urea hydrolysis extended as far as *ca.* 6 cm for urea, DMPP-urea, and the



215 2:1 DMPP-urea/PCU blend. The impacted zone decreased by *ca.* 1 and 2 cm for the 1:2 DMPP-urea/PCU and CRF treatments, respectively.

Acidification of soil solution occurred outside the boundaries of the hydrolysing zone for each combination of soil and N-fertilizer treatment (Fig. 1). There was a clear trend for higher $\text{NO}_3\text{-N}$ production in pots in which greater proportions of soil were at lower pH, suggesting nitrification was a casual factor for this acidification. However, correlation of specific zones of reduced pH and higher $\text{NO}_3\text{-N}$ concentrations was poor (data not shown), likely due to the diffusive mobility of $\text{NO}_3\text{-N}$ down high concentration gradients from the zones in which it was formed.

3.2 N dynamics in the Ferralsol

The high and widely distributed $\text{NH}_4\text{-N}$ concentrations arising from bands of uncoated fertilizers in the Ferralsol (i.e., urea and DMPP-urea; Fig. 5) were somewhat preserved by the broad zone and persistence of modified chemical conditions which were inhibitory to nitrification (Figs. 1 – 3). Since DMPP distribution from fertilizer bands is limited to approximately 1 – 3 cm (Janke et al., 2021) and significant concentrations of $\text{NH}_4\text{-N}$ were more widely distributed (up to 7 cm from fertosphere in both urea and DMPP-urea treatments), there was little effect of this inhibitor on nitrification *cf.* standard urea in this soil. Compared to the pure CRFs, the greater availability of substrate-N (i.e., $\text{NH}_4\text{-N}$ in solution) combined with wider zones of modified soil chemistry meant that peak $\text{NO}_3\text{-N}$ concentrations were higher (*ca.* 50 mg N kg^{-1} soil) but occurred further from the fertosphere in the 1:2 DMPP-urea/PCU blend (3 – 5 cm), 2:1 DMPP-urea/PCU blend (5 – 7 cm), and the urea and DMPP-urea treatments (7 – 9 cm). In contrast, narrower zones of soil impacted by urea hydrolysis meant that early and more rapid nitrification of released N was observed around bands of PCU (0-3 cm zone) and to a lesser extent POCU (0-5 cm zone) by 10 DAI in the Ferralsol (Fig. 5).

Despite the observed variation in $\text{NO}_3\text{-N}$ distribution between N-fertilizer treatments (Fig. 5), the differences in total $\text{NO}_3\text{-N}$ production were not large in the Ferralsol (*ca.* 30 – 70 mg N, Figs. 5, S2). Compared to standard urea, total $\text{NO}_3\text{-N}$ production was most reduced in the PCU treatment (21%) over 35 DAI, where the 1:2 and 2:1 DMPP-urea/PCU blends produced total quantities of $\text{NO}_3\text{-N}$ that were 9 and 14% lower, respectively (Fig. S2). While the controlled release mechanism appears to be more effective for regulating $\text{NO}_3\text{-N}$ production when compared to bands of uncoated product (i.e., urea, DMPP-urea) over the short-term (10 – 35 DAI) in the Ferralsol, this is likely an outcome of simply less N being available for nitrification. When PCU and DMPP-urea are blended, greater proportions of PCU do not appear to confer greater regulation of $\text{NO}_3\text{-N}$ formation. Rather, in the blended treatments, having a greater proportion of N as uncoated DMPP-urea reduced $\text{NO}_3\text{-N}$ production (*cf.* greater proportions of PCU in the blend) by 5% at 35 DAI, with this trend becoming stronger over time (9% difference at 60 DAI). It is unclear if the inhibitory effect of a greater proportion of DMPP-urea in blends was due to (i) a better synchrony of N and DMPP in and around bands (having a portion of the N derived from the coated component of this blend meant that N did not disperse as far as N from pure applications of uncoated products) or (ii) greater ureolytic-induced chemical changes (Figs. 1-3) in response to a greater proportion of N being rapidly hydrolysed (i.e., from the uncoated DMPP-urea, where N from the coated urea would usually be exposed to



‘benign’ chemical conditions under pure PCU applications). Over time, the inhibitory effect of DMPP weakened as greater amounts of N were released from fertilizer bands containing blends with higher proportions of PCU and where the inhibitor to N ratio declined with time (i.e., progressively less of the applied N is “protected” in blends with greater proportions of PCU). By 60 DAI, treatments containing a CRF (pure or blended) demonstrated higher net $\text{NO}_3\text{-N}$ production (Figs. 5, S2) as more slowly released N became available and was nitrified.

3.3 N dynamics in the Vertisol

In the Vertisol, very little nitrification was observed in any N-fertilizer treatment at 10 DAI (Figs. 6, S2), likely an outcome of (i) minimal $\text{NH}_4\text{-N}$ diffusion to zones outside the chemically hostile conditions that formed around bands of uncoated fertilizers (i.e., urea, DMPP-urea) in this high CEC soil; and (ii) limited release of N from coated products. By 35 DAI, the production of $\text{NO}_3\text{-N}$ from urea and CRF bands was significantly greater than that of the other fertilizer treatments (Figs. 6, 7). In pure CRF bands, the phenomenon was similar to that in the Ferralsol where the benign chemical conditions around the fertosphere did not inhibit nitrification of released N. In urea bands, the hostile conditions associated with urea hydrolysis had largely dissipated by 35 DAI in the Vertisol (Figs. 1 – 3), which meant, despite the somewhat reduced N distribution (*cf.* Ferralsol), higher concentrations of $\text{NH}_4\text{-N}$ were found outside the narrower zone of inhibition and were subject to rapid nitrification. In the Vertisol, this diffusion of N from the fertosphere likely occurred in response to steep concentrations gradients as aqueous NH_3 (Fig. 3), since the high CEC of this soil (Table 1) would have restricted movement of N as $\text{NH}_4\text{-N}$.

Unlike the Ferralsol, $\text{NH}_4\text{-N}$ was preserved and $\text{NO}_3\text{-N}$ formation inhibited in the DMPP-urea treatment when compared to urea (Fig. 6), with the variable efficacy of NI products between soil types previously reported by Janke et al. (2021). Effectively, the low permeability of the Vertisol *cf.* to the Ferralsol (lower CEC and coarser texture), meant that the distribution of DMPP and $\text{NH}_4\text{-N}$ around DMPP-urea bands were more closely aligned in the Vertisol, resulting in more complete and persistent inhibition of nitrification in this soil.

The efficacy of the DMPP-urea in the Vertisol meant that the DMPP-urea/PCU blend treatments were also highly effective at reducing $\text{NO}_3\text{-N}$ formation in this soil. Total $\text{NO}_3\text{-N}$ production, relative to urea, was reduced by *ca.* 56% for the 1:2 and 2:1 DMPP-urea/PCU treatments, at 35 DAI and by 61 (1:2) and 55% (2:1) at 60 DAI (Fig. S2). This reduction was considerably greater at 35 DAI and comparable at 60 DAI to either the PCU or DMPP-urea treatments (*ca.* 46 and 53 % reduction at 35 and 60 DAI, respectively). The efficacy of the DMPP-urea/PCU blends in reducing $\text{NO}_3\text{-N}$ formation can be attributed to the inhibitory effect of DMPP on nitrification and the persistence of DMPP in the vicinity of the band while N was slowly diffusing out of PCU granules. Slightly higher rates of nitrification were recorded in 2:1 DMPP-urea/PCU bands (*cf.* 1:2 DMPP-urea/PCU), and this was consistent with the greater proportion of non-coated fertilizer and initially higher $\text{NH}_4\text{-N}$ concentrations that contributed to more extensive diffusion of N (probably as aqueous NH_3 , Fig. 3) into zones beyond the distribution of DMPP. Higher $\text{NO}_3\text{-N}$ concentrations were observed in the 5 – 11 cm zone for this treatment (Fig. 6). Overall, although small differences were observed, the ratio of DMPP-urea to PCU in the blended treatments did not significantly affect the production of $\text{NO}_3\text{-N}$ (Fig. S2) in this soil.



3.4 Effective use of EEFs varies with soil type and loss conditions

These findings suggest that minimisation of fertilizer-N losses arising from $\text{NO}_3\text{-N}$ (i.e., leaching, denitrification) will require different fertilizer-N technology and management practices in different soil types. When fertilizers are banded at the rates used in this study (150 kg N ha^{-1} equivalent), there appears to be little advantage in using a CRF, DMPP-urea or blends of the two (*cf.* urea), in soils of high permeability and poor chemical buffering (i.e., low CEC and pHBC, coarse texture) for mitigating N losses over a season. Similar findings were demonstrated in studies comparing N in soil solution (urea *cf.* DMPP-urea; Janke et al., 2021) and leachate NO_3^- (urea *cf.* PCU; LeMonte et al., 2016) in sandy soils, with no significant differences found between urea and the EEFs. However, some small but insignificant benefits of EEFs have been observed in sugarcane crops grown on sandy-textured soils in field trials and these predominantly occurred in low yielding sites (Bell et al., 2021; Connellan et al., 2021) or where N loss conditions were high (Dowie et al., 2019). This suggests that in scenarios where NUE is poor due to large loss events (e.g., heavy rainfall and extensive leaching), EEFs may provide some protection against N loss, even in soils which may otherwise be poorly suited to banded EEF application. In this case, the timing of the loss event is likely to play a role in efficacy of the EEF product for mitigating N losses. For example, EEF protection against N leaching from a large rainfall event late in the season, when crop N demand is low, offers little benefit compared to when N loss is mitigated early in the season, where an application of standard granular urea would be vulnerable to loss. Indeed, observations from field trials indicate the greatest benefit of the EEFs appears to be in mitigating N losses that occur soon after fertilizer application (Bell et al., 2021; Connellan et al., 2021). The same field programs noted that total N loss from CRF products can sometimes be higher than urea in seasons/sites where prolonged loss conditions occur and the crop is unable to accumulate the slowly released N in crop biomass (e.g. due to waterlogging). This impact on seasonal or product ‘life-time’ losses has also been reported in several other studies of PCU (Janke et al., 2020; Parkin and Hatfield, 2014) and PCU-urea blends (Schwenke and Haigh, 2019), and is an important consideration for effective CRF utilization.

In contrast, in soils where solute movement is restricted, and possibly where chemical changes are better buffered (i.e., high clay soils), there may be a significant reduction in N losses with the use of banded EEFs and blends in the order of DMPP-urea/CRF blends (higher ratio of PCU may offer small but insignificant benefits) > DMPP-urea = PCU > urea. The soil type findings are supported by field research where a greater efficacy of banded DMPP-urea/PCU blends (Bell et al., 2021) or PCUs (DiBella et al., 2013) has been demonstrated in clay and / or organic matter dominant soils by reduced losses and greater crop N uptake.

3.5 Reduced N losses may not always translate into improved crop NUE

Preservation of N as $\text{NH}_4\text{-N}$ (Figs. 5, 6) and retention of N in granules (Fig. 7) may mitigate some $\text{NO}_3\text{-N}$ losses, although N present in the soil as NH_4^+ or retained inside CRF granules may not be accessible to plants. Studies by Lester et al. (2016)



and Schwenke and Haigh (2019) were unable to demonstrate increased N uptake in short-season crops (i.e., cereals and grains) despite reduced N losses and increased N availability in the soil profile after DMPP-urea and / or PCU applications. Similar findings have been reported with irrigated maize for both DMPP-urea or CRF products (Dang et al., 2021; Martinez et al., 2021). However, several trials in sugarcane (Dowie et al., 2019; Wang et al., 2016a), which has a longer N-uptake period, reported similar crop N uptake from lower rates of DMPP-urea than from urea applied at rates 20-25% higher (i.e., industry-standard rates), resulting in greater NUE with this EEF. Recent studies (Janke et al., 2020; Janke et al., 2022) suggest that banding NI and PCU fertilizers may prolong the inhibitory effects and N release from these products, further exacerbating differences in EEF efficacy between long and short-season crops. Additional constraints for improving N uptake from EEFs, despite reduced N losses, can arise due to unfavourable soil moisture dynamics. For example, in conditions where the fertilized layer is dry / drying, N largely remains in shallow fertilizer bands, but plant roots may follow moisture down the soil profile (Janke et al., 2020). Moisture dynamics were not addressed in this study but are an important factor in the multi-faceted challenge of synchronising N availability from fertilizer inputs with crop N demand.

3.6 Potential for biodegradable CRF coatings

The POCU treatment demonstrated earlier and more rapid release of urea-N compared to PCU in both soils, although this was particularly obvious in the Vertisol (Fig. 4). The early release (and subsequent hydrolysis) of greater amounts of urea-N was also evident in the greater increases in pH and EC (Figs. 1, 2) and to a lesser extent, aqueous NH_3 (Fig. 3) in soil treated with POCU (*cf.* PCU). Urea-N released from the POCU was rapidly hydrolysed, resulting in higher $\text{NH}_4\text{-N}$ concentrations up to 35 DAI in the Ferralsol (Fig. 5) and up to 10 DAI in the Vertisol (Fig. 6). By 35 DAI, $\text{NH}_4\text{-N}$ concentrations in the 0 – 5 cm zone of the POCU band were slightly lower than that of PCU in the Vertisol, suggesting rapid nitrification of the $\text{NH}_4\text{-N}$ released earlier in the POCU treatment. The nitrified N appeared to diffuse to zones further away from the band in response to established concentration gradients in both CRF treatments, with the distribution of $\text{NO}_3\text{-N}$ similar for both treatments by 60 DAI in the Vertisol (Fig. 6). There were slightly higher concentrations of $\text{NO}_3\text{-N}$ further away from the band in the POCU *cf.* the PCU treatment, which were likely a result of earlier nitrification of greater amounts of N and subsequent diffusion. A similar trend was observed in the Ferralsol, with peak $\text{NO}_3\text{-N}$ concentrations occurring closer to the fertilizer band for PCU (0 – 5 cm zone) than the POCU (7 – 11 cm) at 35 – 60 DAI.

These findings suggest that a proportion of these POCU granules may have experienced coating ‘failure’, whereby increases in osmotic pressure within granules in response to water penetration cause some of the POCU granules to rupture rather than swell. Micropores within the coating of many CRFs are stretched and enlarged during the imbibement and swelling process, allowing release of the coated nutrient (now dissolved in water) from the granule (Shaviv, 2001). The hypothesis of POCU granule ‘failure’ is supported by the detection of a greater proportion of ‘burst’ POCU granule coatings (*cf.* PCU) during granule recovery at sampling. The greater volumetric water content of the Vertisol at field capacity (*cf.* Ferralsol; Table 1) may have contributed to more rapid water uptake and hence more frequent rupturing of POCU granules, resulting in the initially higher urea-N and $\text{NH}_4\text{-N}$ concentrations in this soil (Figs. 4, 6). Observations of a high prevalence



of ‘burst’ granules have also been made in field experiments with banded POCU in a Vertisol (Martinez et al., 2021). The coating of this particular biodegradable CRF therefore does not appear to be as robust or elastic as commercial PCU coatings. However, despite the initial flush of N from apparently ‘failed’ POCU granules, mineral N distribution (Figs. 5, 6) and the proportion of N remaining in granules (only 2 – 7 % difference, Fig. 7) was relatively similar for bands of POCU and
350 PCU within each soil. This suggests that once moisture uptake by POCU granules has completed, the N release characteristics were very similar to that of PCU. Furthermore, a slightly higher availability of early N may have a positive impact (*cf.* PCU) when such products are deployed in short-season crops, with early-season N supply potentially better matched to plant demand. However, this hypothesis requires validation in studies in which realistic plant N demands are placed on similar POCU products.

355

4 Conclusions

Banding NI-urea and CRF technology in blends of differing proportions revealed that the NI concentration or proportion of CRF within the band did not have a significant impact on N dynamics. Both treatments slowed N release (with higher proportions of PCU) and higher NI concentrations (with higher proportions of NI-urea) limited the distribution of N from
360 these fertilizer bands (*cf.* urea bands). Consequently, $\text{NH}_4\text{-N}$ formed within blended bands was similarly protected against oxidisation by inhibitory conditions associated with band chemistry and the NI, irrespective of blend proportions. The efficacy of N preservation in blended vs ‘pure’ NI and CRF treatments varied depending on soil type, but in all instances the availability of N to meet potential plant demand was initially lower in treatments including the PCU. Timing of N availability with crop demand has significant implications for crop NUE, with the lower N supply in treatments including a
365 CRF likely to impact crops with greater early seasonal demand. Poor synchrony of N availability to crop demand also increases the risk of environmental losses, with the persistence of fertilizer N in soil once crop uptake has concluded vulnerable under wetter seasonal conditions. Importantly, soil characteristics appeared to play a large role in influencing the efficacy of EEFs and blends of NI-urea/CRF. When applied in concentrated fertilizer bands, these individual products and their blends are likely to be most effective in soils with lower permeability. This is due to better co-location of fertilizer N
370 with any applied NI and inhibitory conditions associated with band chemistry.

The plant-oil coated urea product that was evaluated initially released more N due to a higher prevalence of ‘burst’ granules, which was likely an outcome of somewhat poorer tolerance of increased osmotic pressure within granules (*cf.* PCU). However, overall dynamics and proportions of N in mineral forms (NH_4^+ and NO_3^-) were similar to that of PCU, suggesting this technology may be a suitable option for managing the competing requirements of (i) a predictable N supply
375 and (ii) avoiding the introduction of persistent polymers in the environment.

This study has provided fundamental understanding of the dynamics and efficacy of bands of blended NI-urea and PCU fertilizers, relative to the individual blend constituents and to granular urea. It also offers preliminary evidence to



persist with the development and testing of biodegradable CRFs as substitutes for the PCU products currently in use. Findings from this study may be utilized in agronomic decisions to reduce the impact of off-farm N losses and improve on-farm NUE.

Data availability. Original data are available upon request. Data necessary for this study's findings are presented in the manuscript.

Author contributions. Chelsea K Janke and Michael J Bell conceived and designed the experiments. Chelsea K Janke performed the experiments and analysed the data. Chelsea K Janke and Michael J Bell prepared the manuscript and both authors approve the submission of this paper.

Competing interests. The authors declare that they have no conflict of interest.

Acknowledgements. This research was supported by the Australian Government's National Environmental Science Programme (Tropical Water Quality Hub) projects 2.1.8 and 5.11 'Improved Water Quality Outcomes from On-Farm Nitrogen Management'.

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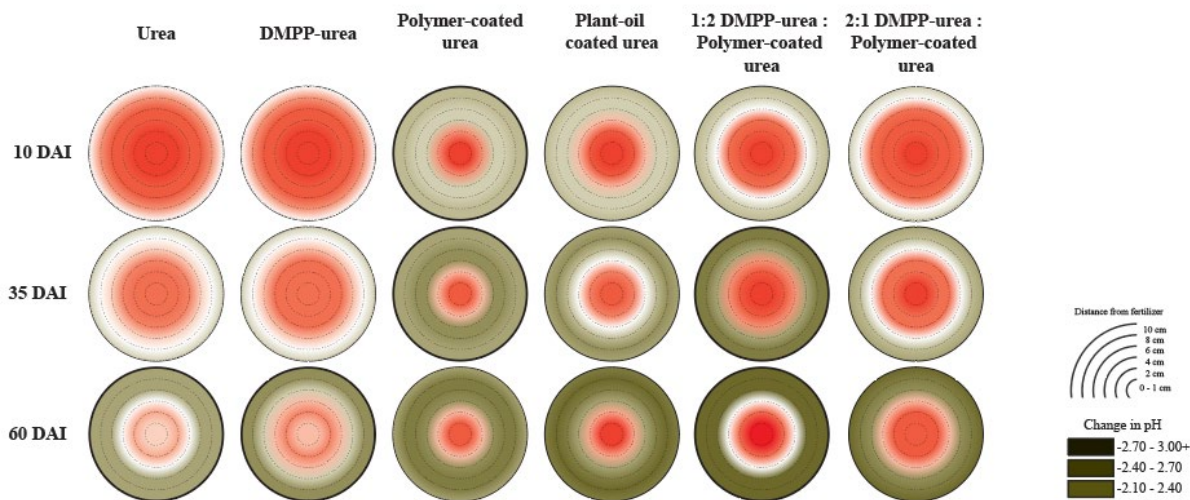
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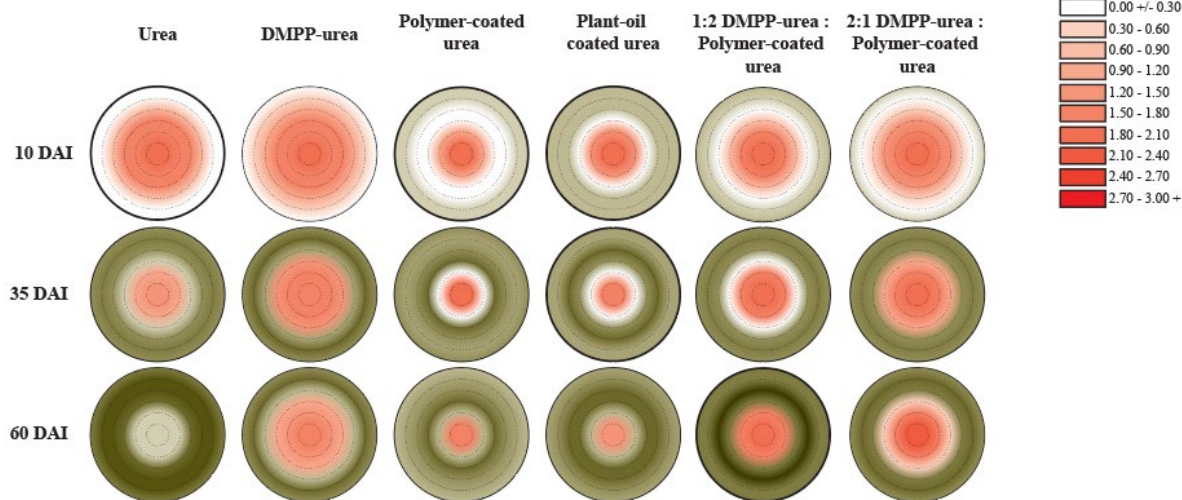
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Vertisol

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575 **Figure 1:** Heat maps displaying changes (with respect to untreated soil, Table 1) in pH distribution from bands of N-fertilizer treatments in the Ferralsol and Vertisol over time (days after incubation; DAI). Initial pH values were 6.30 (Ferralsol) and 7.15 (Vertisol). Changes in the pH of untreated soil are presented in the Supplementary Material (Part A, Table S1).



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Ferralsol

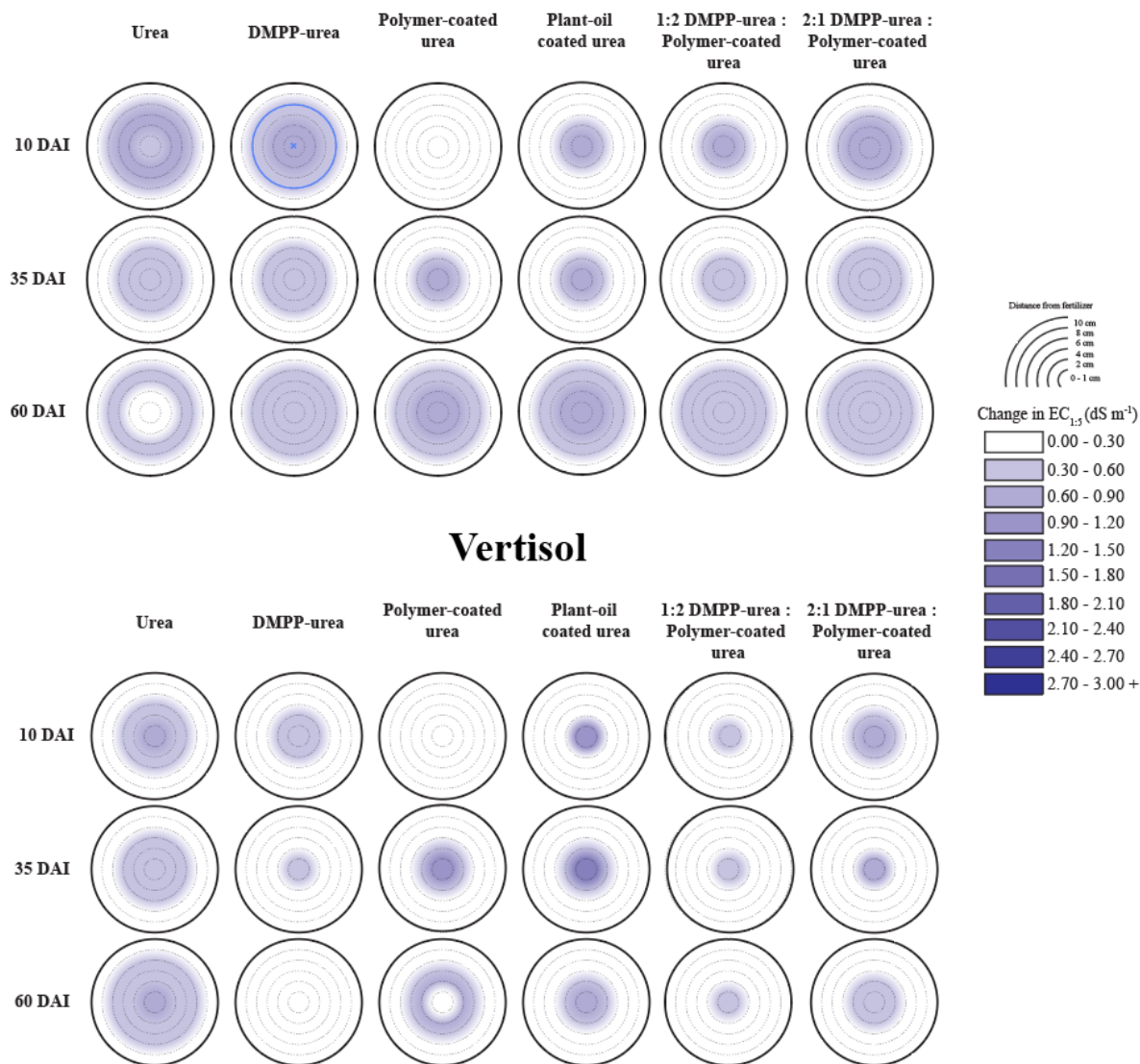
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610 **Figure 2: Heat maps displaying changes (with respect to untreated soil, Table 1) in electrical conductivity (EC, dS m⁻¹) distribution from bands of N-fertilizer treatments in the Ferralsol and Vertisol over time (days after incubation; DAI). Initial EC values were 0.07 dS m⁻¹ (Ferralsol) and 0.35 dS m⁻¹ (Vertisol). Changes in the EC of untreated soil are presented in the Supplementary Material (Part A, Table S1).**

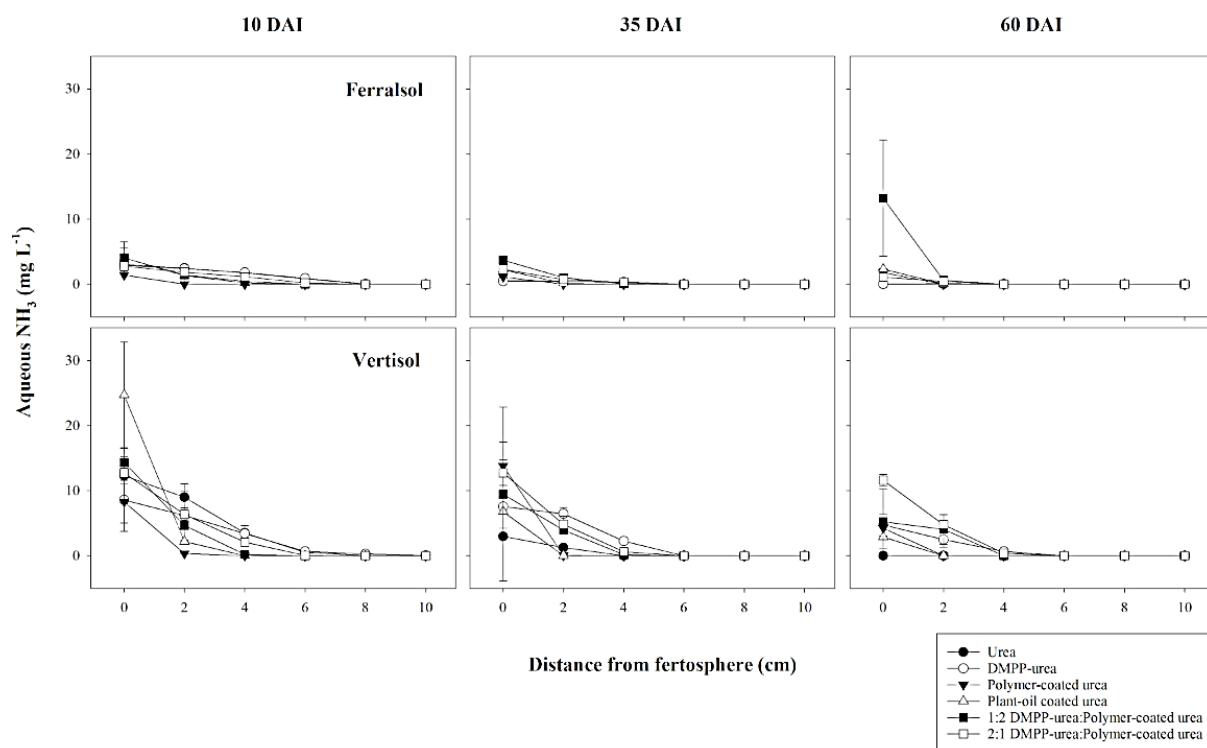


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635 **Figure 3: Aqueous NH₃ concentration (mg L⁻¹) over distance and time (days after incubation; DAI) from banded N-fertilizer treatments in a Ferralsol and Vertisol. Vertical bars represent the standard error of the mean.**

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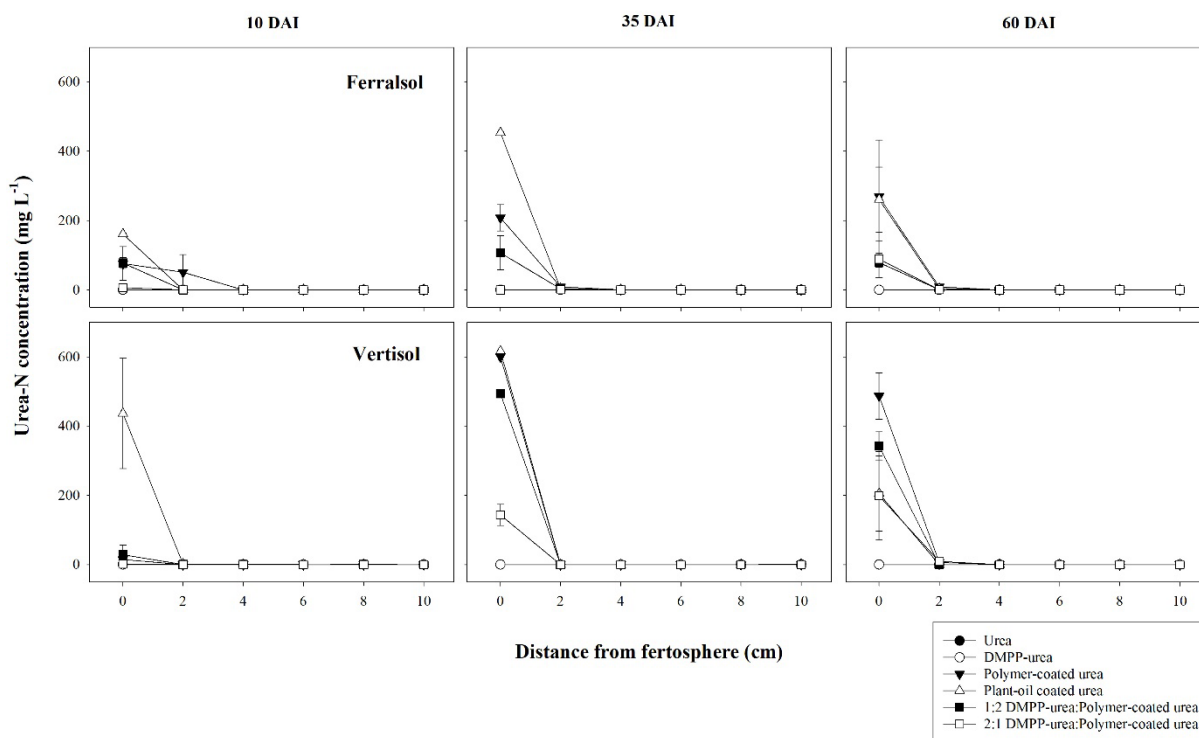


Figure 4: Urea-N concentration (mg L^{-1}) over distance and time (days after incubation; DAI) from banded N-fertilizer treatments in a Ferralsol and Vertisol. Vertical bars represent the standard error of the mean.

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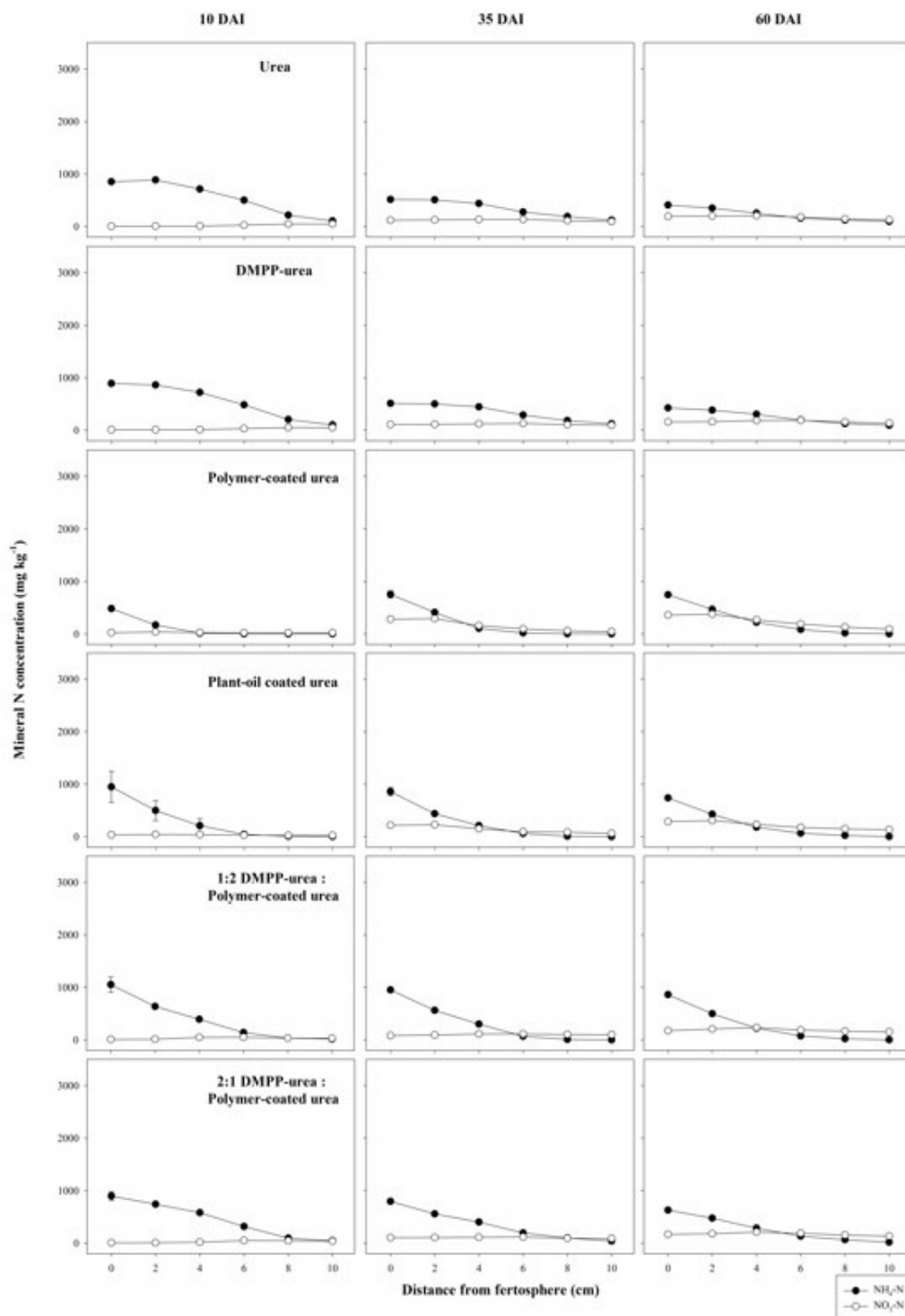
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715 **Figure 5: Concentration (mg N kg⁻¹ soil) of mineral N species from N-fertilizer treatments in the Ferralsol during a 60-day incubation (DAI). Vertical bars represent the standard error of the mean.**

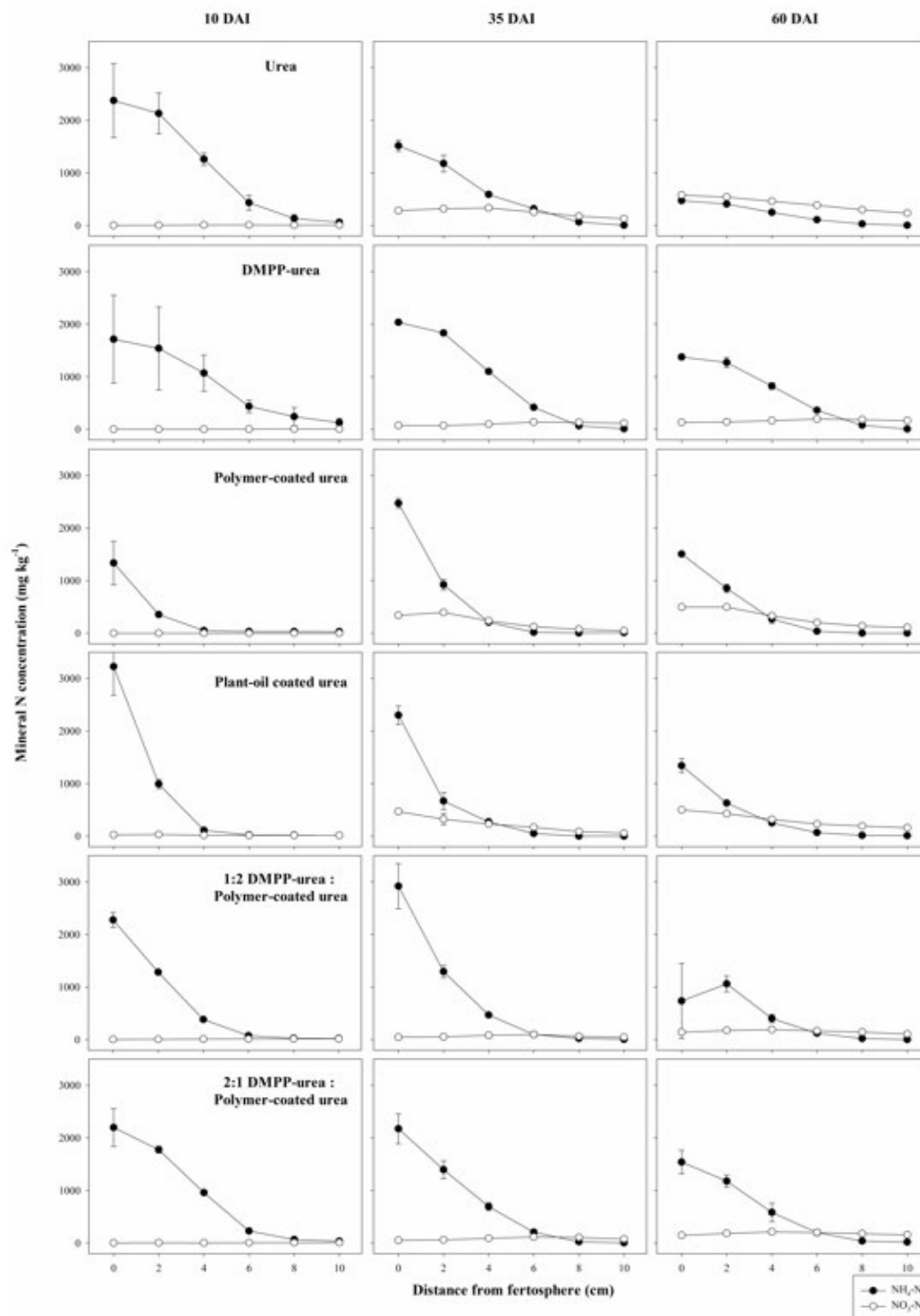


Figure 6: Concentration (mg N kg⁻¹ soil) of mineral N species from N-fertilizer treatments in the Vertisol during a 60-day incubation (DAI). Vertical bars represent the standard error of the mean.



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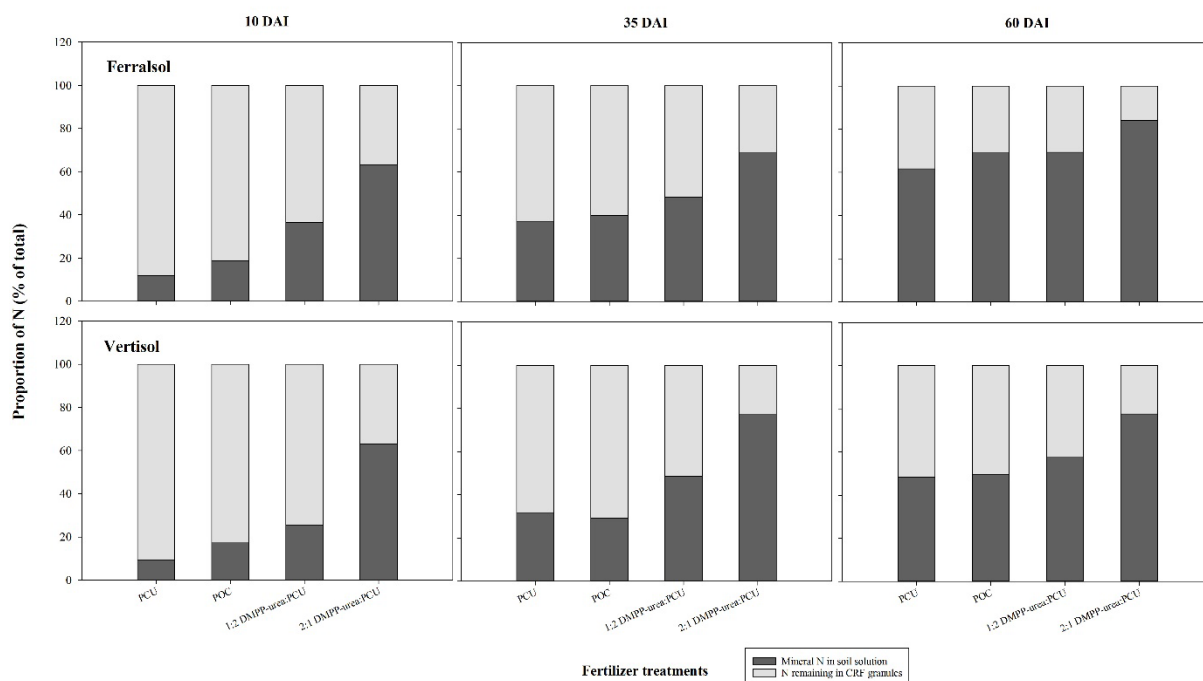


Figure 7: Proportion of total recovered N (%) in either soil solution (dark shading) or remaining within granules (light shading) over 60 days after incubation (DAI) in the Ferralsol and Vertisol.

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Table 1: Soil chemical properties and characteristics. Where, EC is the electrical conductivity; pHBC is the pH buffering capacity; C is carbon; TOC is total organic carbon; TN is total nitrogen; and GWC (FC) is the gravimetric water content at field capacity

Soil Order ¹	pH 1:5w	EC 1:5	(cmol OH ⁻ -kg ⁻¹ pH unit ⁻¹)	Particle size analysis (%)				Exchangeable cations cmol(+) kg ⁻¹					(%)			(mg kg ⁻¹)	(g g soil ⁻¹)	
				Coarse sand	Fine sand	Silt	Clay	Ca	Mg	K	Na	CEC	Total C	TOC	TN	NH ₄ ⁺ -N	NO ₃ ⁻ -N	GWC (FC)
Ferralsol	6.30	0.07	1.32	38	46	9	9	3.6	0.7	0.19	<0.08	4.5	1.21	1.17	0.06	3.2	6.8	0.12
Vertisol	7.15	0.35	3.51	5	26	21	52	12.8	13.2	0.79	1.82	28.5	2.36	2.32	0.18	4.8	8.5	0.32

¹IUSS Working Group, 2015