

# Satellite-derived Constraints on the Effect of Drought Stress on Biogenic Isoprene Emissions in the Southeast US

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**Abstract.** While substantial progress has been made to improve our understanding of biogenic isoprene emissions under unstressed conditions, there remain large uncertainties in isoprene emissions under stressed conditions. Here we use the US Drought Monitor (USDM) as a weekly drought severity index and tropospheric columns of formaldehyde (HCHO), the key product of isoprene oxidation, retrieved from the Ozone Monitoring Instrument (OMI) to derive top-down constraints on the response of summertime isoprene emissions to drought stress in the Southeast U.S. (SE US), a region of high isoprene emissions and prone to drought. OMI HCHO column density is found to be 6.7% (mild drought) - 23.3% (severe drought) higher than that in no-drought conditions. A global chemical transport model, GEOS-Chem, with the MEGAN2.1 emission algorithm can simulate this direction of change, but the simulated increases at the corresponding drought levels are 1.1-1.5 times of OMI HCHO, suggesting the need for a drought-stress algorithm in the model. By minimizing the model-to-OMI differences in HCHO to temperature sensitivity under different drought levels, we derived a top-down drought stress factor ( $\gamma_{d\_OMI}$ ) in GEOS-Chem that parameterizes using water stress and temperature. The algorithm led to an 8.6% (mild drought) - 20.7% (severe drought) reduction in isoprene emissions in the SE US relative to the simulation without it. With  $\gamma_{d\_OMI}$  the model predicts a non-linear increasing trend in isoprene emissions with drought severity that is consistent with OMI HCHO and a single site's isoprene flux measurements. Compared with a previous drought stress algorithm derived from the latter, the satellite-based drought stress factor performs better in capturing the regional scale drought-isoprene responses as indicated by the close-to-zero mean bias between OMI and simulated HCHO columns under different drought conditions. The drought stress algorithm also reduces the model's high bias in organic aerosols (OA) simulations by 6.60% (mild drought) to 11.71% (severe drought) over the SE US compared to the no-stress simulation. The simulated ozone response to the drought stress factor displays a spatial disparity due to the isoprene suppressing effect on oxidants, with an <1 ppb increase in O<sub>3</sub> in high-isoprene regions and a 1-3 ppbv decrease in O<sub>3</sub> in low-isoprene regions. This

36 study demonstrates the unique value of exploiting long-term satellite observations to develop empirical stress  
37 algorithms on biogenic emissions where in situ flux measurements are limited.

## 38 **1. Introduction**

39 Biogenic nonmethane volatile organic compounds (BVOCs) emitted by terrestrial ecosystems are of great importance  
40 to air quality, tropospheric chemistry, and climate due to their effects on atmospheric oxidants and aerosols (Atkinson,  
41 2000; Claeys et al., 2004; Pacifico et al., 2009). The dominant BVOC is isoprene ( $\text{CH}_2=\text{C}(\text{CH}_3)\text{CH}=\text{CH}_2$ ), comprising  
42 70% of the global total BVOC emitted from vegetation (Sindelarova et al., 2014). Isoprene emissions depend on  
43 vegetation/plant type, physiological status, leaf age, and meteorological conditions such as radiation, temperature, and  
44 soil moisture. These relationships provide the basic framework of isoprene emission models that are capable of  
45 coupling with meteorology and the land biosphere, the most widely used being the Model of Emissions of Gases and  
46 Aerosols from Nature (MEGAN) (Guenther et al., 1993, 2006, 2012, 2017). Recent work has shown stressed  
47 conditions - such as drought, heatwaves, and high winds - can induce large changes in isoprene emissions different  
48 from model predictions in the absence of those stress factors (Potosnak et al., 2014; Huang et al., 2015; Kravitz et al.,  
49 2016; Seco et al., 2015; Otu-Larbi et al., 2020; Seco et al., 2022). As stressed conditions are rarely sampled by field  
50 campaigns due to their infrequent and irregular nature and hence poorly constrained, stress impacts on isoprene  
51 emissions are among the least understood aspects in our predictivity of BVOC-chemistry-climate interactions.

52 A common stress for terrestrial vegetation worldwide is drought, characterized by low precipitation, high temperature,  
53 and low soil moisture (Trenberth et al., 2014). These conditions are primary abiotic stresses that will cause  
54 physiological impacts on plants affecting photosynthesis, stomatal conductance, transpiration, and leaf area. During  
55 short-term or mild droughts, the photosynthetic rate of plants quickly decreases due to limited stomatal conductance,  
56 while isoprene is not immediately impacted because of the availability of stored carbon and because the photosynthetic  
57 electron transport is not inhibited. Isoprene can even increase by several factors due to warm leaf temperatures which  
58 increases isoprene synthase activity (Potosnak et al., 2014; Ferracci et al., 2020). During prolonged or severe drought  
59 stress, after a lag related to photosynthesis reduction, isoprene emission eventually declines because of inadequate  
60 carbon availability. This conceptualized non-monotonic response of isoprene emission to drought has been  
61 demonstrated at the Missouri Ozarks AmeriFlux (MOFLUX) field site in Missouri (Potosnak et al., 2014; Seco et al.,  
62 2015), the only available drought-relevant whole canopy isoprene flux measurements to date, and qualitatively  
63 supported by ambient isoprene concentrations monitored by regional surface networks (Wang et al., 2017). It is  
64 noteworthy that the MOFLUX data covered only two drought events (summer 2011 and summer 2012), while the  
65 surface sites are sparsely distributed with an urban focus. More recently, the isoprene concentration measurements  
66 during the Wytham Isoprene iDirac Oak Tree Measurements (WiSDOM) campaign showed that isoprene was up to  
67 four times higher than normal in responses to a combined heatwave and drought episode (June-October 2018) over a  
68 mid-latitude temperate forest in the UK (Ferracci et al., 2020; Otu-Larbi et al., 2020), which supports the enhanced  
69 isoprene emissions at the MOFLUX site under mild droughts. However, these observations offer only limited  
70 constraints on drought stress impacts on isoprene emissions.

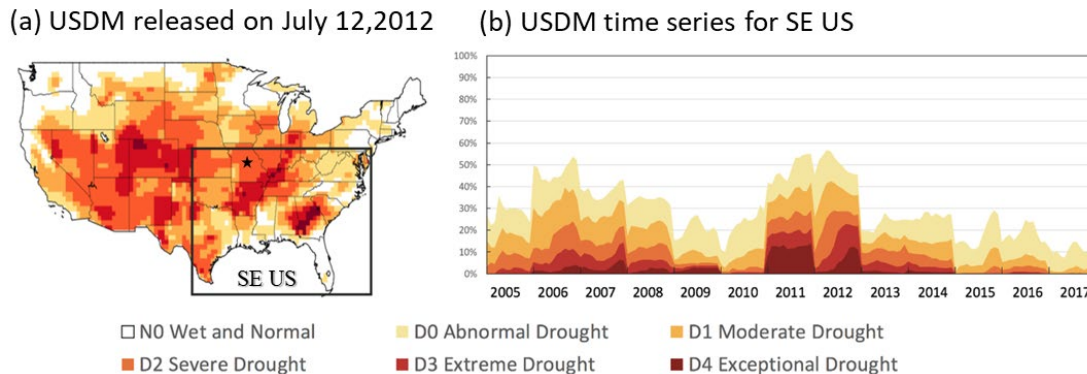
71 With wide spatiotemporal coverage, satellite provides arguably the best platform to capture drought development and  
72 impacts. Satellite observations of tropospheric formaldehyde (HCHO) columns have been used as a proxy of isoprene  
73 emissions for more than a decade (Abbot et al., 2003; Palmer et al., 2003), as HCHO is formed promptly and in high  
74 yield from isoprene oxidation (Sprengnether et al., 2002). Previous applications of satellite HCHO products provided  
75 “top-down” estimates on seasonality, magnitude, spatial distribution, and interannual variability of isoprene emissions  
76 globally and regionally (e.g., Marais et al., 2016; Kaiser et al., 2018; Stavrakou et al., 2018). While most of these  
77 studies focused on *unstressed* conditions, recent efforts have shown that satellite HCHO registered drought signals on  
78 a monthly scale (Zheng et al., 2017; Naimark et al., 2021; Li et al., 2022; Opacka et al., 2022). These signals are yet  
79 to be exploited to constrain isoprene response to drought.

80 The present study aims at improving the current quantification of satellite HCHO response to drought by accounting  
81 for sub-monthly variability of drought severity. We use a weekly time scale, the finest temporal scale of drought  
82 indices available, and separate five levels of drought severity defined by the US Drought Monitor. By comparison,  
83 previous investigations used binary classification (drought or not) on a monthly time scale. Our improvement in scale  
84 is expected to better capture the nonlinear response of isoprene emissions to drought severity as described above. The  
85 study region is the Southeast United States (SE US), which has large isoprene emissions due to substantial forest  
86 coverage and is also prone to drought due to large interannual variability in precipitation (Seager et al., 2009). In  
87 addition, the MOFLUX site is located in the SE US, which will allow us to evaluate if satellite-derived drought  
88 responses of HCHO are consistent with those from isoprene flux measurements at MOFLUX. Finally, we use these  
89 HCHO signals in conjunction with models to identify the model gaps in predicting isoprene responses to drought.

## 90 **2. Data and Method**

### 91 **2.1 Drought index**

92 There are many types of drought indices focusing on different factors, including precipitation, temperature,  
93 evaporation, runoff, and the impact of drought on ecosystems and vegetation (Palmer, 1965; McKee et al., 1993;  
94 Guttman, 1999; Vicente-Serrano et al., 2010; Chang et al., 2018). Drought indices also differ by time scale. As drought  
95 by definition is a prolonged period of water deficit, the shortest time scale of drought is weekly. Here we chose the  
96 United States Drought Monitor (USDM) drought index to identify drought periods. USDM’s weekly timescale and  
97 multiple drought severity levels (Svoboda et al., 2002) provide a better delineation of drought variability than the  
98 monthly or seasonal scale used in the previous analysis of drought signals in HCHO and isoprene (Wang et al., 2017;  
99 Naimark et al., 2021).



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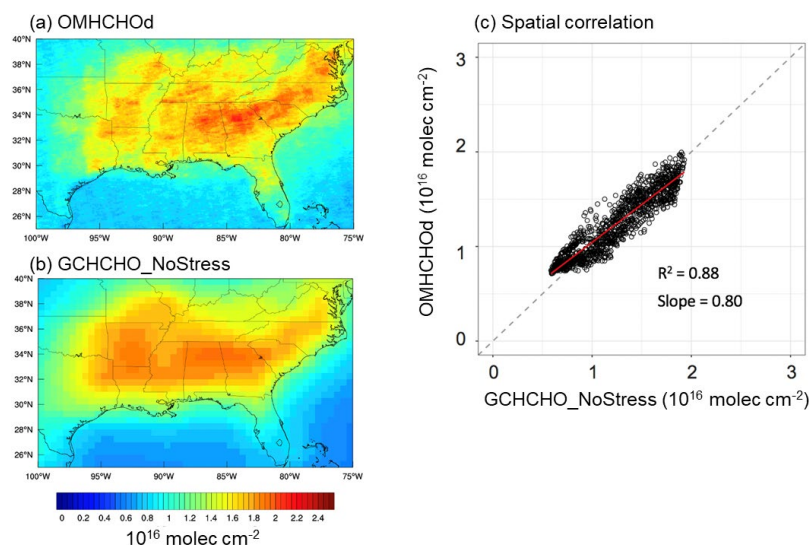
101 **Figure 1. (a) Drought distribution for the second week of July 2012 based on USDM. The black star indicates the location**  
 102 **of MOFLUX site. (b) Time series of drought frequency in the study area (black box in Figure 1a) for JJA from 2005 to**  
 103 **2017. N0 (white) for wet and normal, D0 (light yellow) for abnormal drought, D1 (yellow) for moderate drought, D2 (orange)**  
 104 **for severe drought, D3 (red) for extreme drought, and D4 (brown) for exceptional drought.**

105 The USDM is a composite drought index based on six key physical indicators including the Palmer Drought Severity  
 106 Index (PDSI, Palmer, 1965), CPC Soil Moisture Model Percentiles (Huang et al., 1996), U.S. Geological Survey  
 107 (USGS) Daily Streamflow Percentiles (<http://water.usgs.gov.waterwatch/>), Percent of Normal Precipitation (Willeke  
 108 et al., 1994), Standardized Precipitation Index (SPI, McKee et al., 1993), and remotely sensed Satellite Vegetation  
 109 Health Index (Kogan, 1995). Opinions of local experts are also considered (Svoboda et al., 2002). The USDM website  
 110 (<https://droughtmonitor.unl.edu/>) provides weekly ArcGIS shapefiles of the polygons covering the whole US under  
 111 five drought levels: D0 for abnormal drought, D1 for moderate drought, D2 for severe drought, D3 for extreme drought,  
 112 and D4 for exceptional drought. We used the method of Chen et al. (2019) to rasterize and convert USDM shapefiles  
 113 to  $0.5^\circ \times 0.5^\circ$  gridded indices with -1 indicating non-drought (N0) and 0-4 for D0-D4 drought, respectively. **Figure**  
 114 **1a** displays the spatial distribution of gridded USDM indices for the second week of July 2012, which clearly depicts  
 115 the extent and severity of the infamous 2012 Great Plains drought (Hoerling et al., 2014). **Figure 1b** shows the weekly  
 116 time series of USDM indices averaged over SE US ( $75\text{--}100^\circ\text{W}$ ,  $25\text{--}40^\circ\text{N}$ , black box in Figure 1a) for the summer  
 117 months (June, July, August; JJA) of 2005 -2017, our study period. During this period, abnormal drought (D0) appeared  
 118 every summer, while extreme and exceptional drought (D3-D4) were mainly concentrated in 2006-2008 and 2010-  
 119 2012. This pattern is consistent with the long-term drought statistics from other drought indices such as SPEI and  
 120 PDSI (Svoboda et al., 2015).

## 121 2.2 OMI HCHO and NO<sub>2</sub> product

122 We used the Ozone Monitoring Instrument (OMI) v003 level 3 tropospheric formaldehyde (HCHO) column density  
 123 (OMHCHOd) as described by Chance (2019). OMI was launched on NASA's Aura satellite in 2004 and has since  
 124 provided daily global measurements of ozone (O<sub>3</sub>) and its precursors with a nadir spatial resolution of  $24 \times 13 \text{ km}^2$ .  
 125 Since January 2009, OMI has been suffering from a major row anomaly. OMHCHOd data processing explored all  
 126 level 2 OMHCHO observations to filter out pixels with bad formaldehyde retrievals, high cloud fractions (>30%),  
 127 high SZA (>70°), and pixels affected by OMI's row anomaly (Chance, 2019). The spatial resolution is  $0.1^\circ \times 0.1^\circ$ .  
 128 Zhu et al. (2016) verified the OMHCHOd data using high-precision HCHO aircraft observations obtained during

129 NASA SEAC4RS activities in SE US from August to September 2013. They showed that OMI retrievals have accurate  
 130 spatial and temporal distribution but were biased low by 37% relative to the aircraft. We corrected this underestimation  
 131 by applying a uniform and constant factor of 1.5 to the OMHCHOd data, as did by Shen et al. (2019) in their long-  
 132 term analysis of OMI HCHO. **Figure 2a** presents the corrected OMHCHOd for the SE US averaged over JJA 2005-  
 133 2017, where higher levels of HCHO are clearly seen over forested regions in Missouri, Georgia, Arkansas, and Texas.  
 134 OMHCHOd values shown hereafter are those with the correction factor applied. Although it is not known if the  
 135 correction factor has temporal spatial variations during our study period, its application produced a good match  
 136 between OMI and simulated HCHO columns under non-drought (N0) conditions (Figure 2c). To examine the  
 137 concurrent changes of nitrogen oxides ( $\text{NO}_x = \text{NO}_2 + \text{NO}$ ) under droughts, we also used the level 3 tropospheric  
 138 column of  $\text{NO}_2$  from OMI during the same period (Nickolay et al., 2019).



139  
 140 **Figure 2. Mean 2005–2017 HCHO columns for June – August over the SE US of (a) OMI observation (OMHCHOd) and**  
 141 **(b) GEOS-Chem simulation (GCHCHO\_NoStress). (c) Scatterplot of spatial correlation between the two. The dashed line**  
 142 **indicates the 1:1 agreement.**

143 **2.3 GEOS-Chem chemical transport model**

144 We used the long-term simulation of the nested-grid GEOS-Chem global chemical transport model (version 12-02,  
 145 <http://www.geos-chem.org>) to obtain daily mean results of modeled formaldehyde columns and isoprene emissions  
 146 for North America during JJA 2005 – 2017. The simulation was driven by the Modern-Era Retrospective analysis for  
 147 Research and Applications, Version 2 (MERRA-2) meteorological data from NASA’s Global Modeling and  
 148 Assimilation Office (GMAO) with a horizontal resolution at  $0.5^\circ \times 0.625^\circ$ . Biogenic emissions were calculated using  
 149 MEGAN2.1, which is the prevailing version of MEGAN implemented in most chemical and climate models.  
 150 MEGAN2.1 has a soil dependence algorithm whose parameterization is based on plant wilting point threshold and  
 151 soil moisture (Guenther et al., 2017). However, this factor is disabled in GEOS-Chem as in many other CTMs due to  
 152 the unavailability of the required driving variables, such as wilting point and soil moisture, which cannot be simulated  
 153 well in most models (Trugman et al., 2018). Thus, outputs from the standard GEOS-Chem simulations do not have

154 drought effects on isoprene emissions and these outputs are referred to as NoStress\_GC. Anthropogenic emissions  
155 over North America were from the 2011 National Emissions Inventory (NEI2011, [http://www.epa.gov/air-emissions-](http://www.epa.gov/air-emissions-inventories)  
156 [inventories](http://www.epa.gov/air-emissions-inventories)) for the United States, with historical scale factors applied to each simulated year. Open fire emissions  
157 were from GFED4 (Giglio et al., 2013) for 2005–2017.

158 To better match with OMI overpassing time, model HCHO outputs at 13:30 local time were sampled  
159 (GCHCHO\_NoStress). **Figure 2b** shows GCHCHO\_NoStress averaged over the same domain and period as  
160 OMHCHOd in **Figure 2a**. The scatter plot (**Figure 2c**) shows a good spatial correlation between the two ( $R^2 = 0.88$ ).  
161 This correlation is consistent with other studies comparing GEOS-Chem and OMI HCHO columns in SE US during  
162 non-drought periods (Kaiser et al., 2018).

## 163 **2.4 Observations of ozone, organic aerosol, LAI, and isoprene flux**

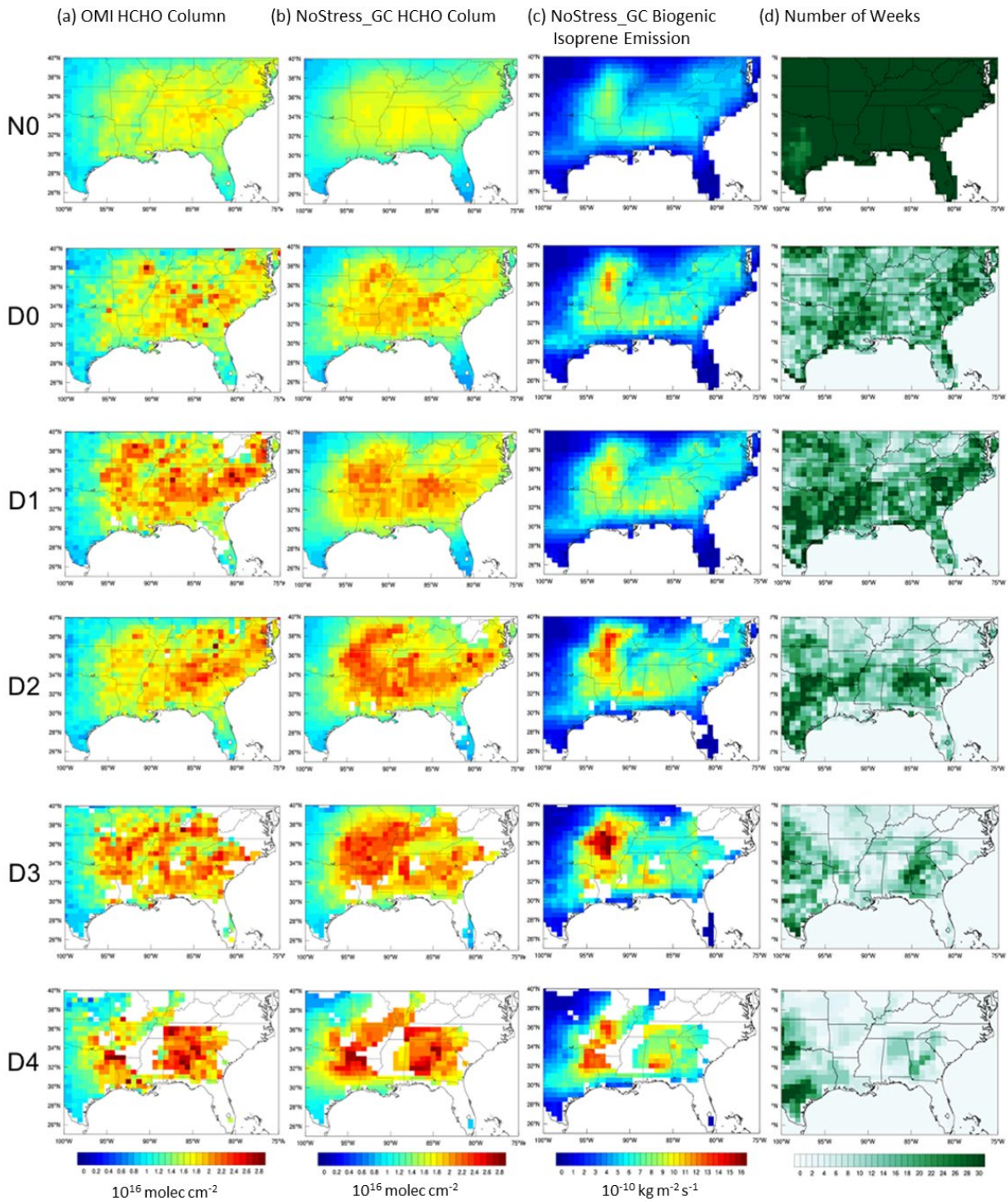
164 To evaluate how the drought stress factor changes the simulations of surface O<sub>3</sub> and organic aerosol (OA), we adopted  
165 the gridded (1° × 1°) hourly O<sub>3</sub> observations created by Schnell et al. (2014) using the modified inverse distance  
166 weighting method. The dataset aggregates several networks of O<sub>3</sub> measurements including the US Environmental  
167 Protection Agency’s (EPA) Air Quality System (AQS), Clean Air Status and Trends Network (CASTNET), and  
168 Environment Canada’s National Air Pollution Surveillance Program (NAPS). Following the same method, we created  
169 a gridded organic aerosol (OA) dataset using the organic carbon (OC) observations from the Interagency Monitoring  
170 of Protected Visual Environments (IMPROVE) network. A factor of 2.1 was used to convert OC to OA as suggested  
171 by other studies (Pye et al., 2017; Schroder et al., 2018). To examine the changes of leaf area index (LAI) under  
172 droughts, the MODerate resolution Imaging Spectroradiometer (MODIS) Collection 5 LAI products reprocessed by  
173 Yuan et al. (2011) with a resolution of 0.25° × 0.25° was used. These three datasets were further remapped through  
174 bilinear interpolation to match the spatial resolution of the USDM. The isoprene flux measurements at the MOFLUX  
175 site during 2012 May-September were used to derive a site-based drought stress algorithm. The site is located in the  
176 Ozarks region of central Missouri (38.74°N, 92.20°W, black star in Figure 1a). It is surrounded by a deciduous forest  
177 dominated by isoprene-emitting white and red oak species. The dataset is widely used to investigate isoprene emissions  
178 response to droughts (Potosnak et al., 2014; Seco et al., 2015; Jiang et al., 2018; Opacka et al., 2022).

## 179 **3. Observational Evidence of Drought Stress on Isoprene Emissions**

### 180 **3.1 Changes of HCHO column densities with drought**

181 To reveal drought responses of HCHO, we sampled weekly-mean HCHO columns onto the gridded spatial and  
182 temporal locations of each USDM category and generated average HCHO distributions at each drought level over the  
183 SE US. The outputs are shown in **Figure 3a** for OMI and **3b** for NoStress\_GC, respectively. The processing of weekly-  
184 mean HCHO data corresponds to the timing of USDM: a whole week includes Wednesday of the previous week to  
185 Tuesday of the present week. There are 12 consecutive weeks from June to August in each year of 2005-2017, giving  
186 a total of 156 weeks’ gridded HCHO data to be assigned to individual USDM categories by week and location. **Figure**  
187 **3d** shows the number of weeks underlying the gridded averages of HCHO for each USDM category. As severe

188 droughts are less frequent than mild droughts, some locations in SE US did not experience D2-D4 droughts during the  
 189 study period and hence are shown as white in **Figure 3**.

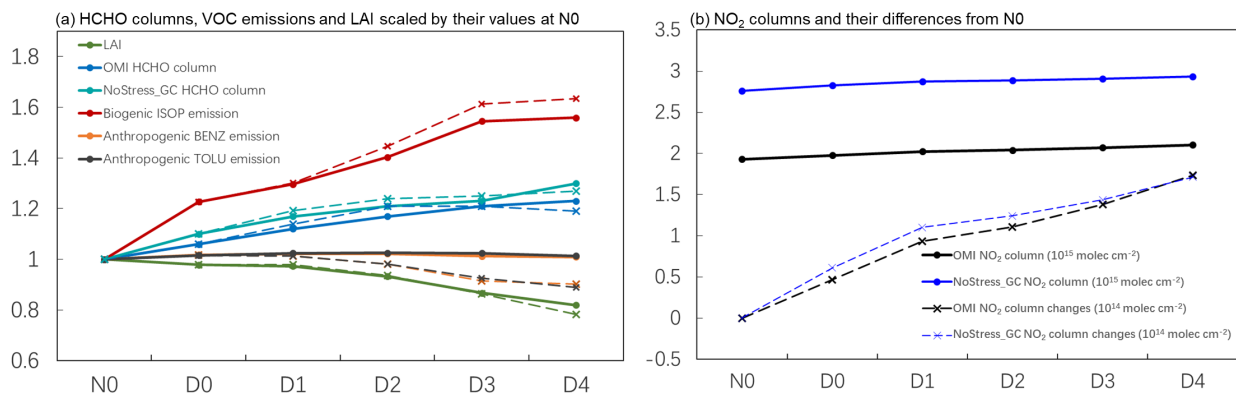


190

191 **Figure 3.** The mean spatial distributions of (a) OMI HCHO column density; (b) NoStress\_GC HCHO column density, (c)  
 192 NoStress\_GC isoprene emissions, and (d) the number of weeks during JJA 2005 to 2017 in the southeast US under different  
 193 USDM drought levels (N0, D0-D4).

214 OMI HCHO column density increases with increasing drought severity in almost all locations in the SE US (**Figure**  
 215 **3a**). Relative to no-drought condition (N0), the mean HCHO column from OMI is 6.7%, 12.6%, 16.5%, 21.2%, and  
 216 23.2% higher under D0 - D4 drought in the entire SE US, respectively. These HCHO changes are statistically  
 217 significant at a 95% confident interval, indicating that the OMI HCHO products contain significant drought signals.  
 218 The increasing rate of OMI HCHO with USDM is not linear, faster under mild droughts (D0-D2) and flattening under  
 219 more severe droughts (D2-D4). This is qualitatively consistent with the conceptualized model of the nonlinear  
 220 response of isoprene emissions to drought described before (Potosnak et al., 2014).

221 Model HCHO column density also increases with increasing drought severity (**Figure 3b**). GCHCHO\_NoStress is  
 222 9.90%, 15.1%, 19.5%, 21.8%, and 29.1% higher under D0-D4 drought than that of N0, respectively. These increases  
 223 are 1.1-1.5 times those of OMI under all drought levels. The model comparison against OMI HCHO also changes  
 224 with drought severity. GCHCHO\_NoStress has a minimal bias ( $0.05 \times 10^{16}$  molec  $\text{cm}^{-2}$ ) under N0. As drought severity  
 225 increases, the mean bias over the entire SE US increases to  $0.10 \times 10^{16}$  molec  $\text{cm}^{-2}$ ,  $0.09 \times 10^{16}$  molec  $\text{cm}^{-2}$ ,  $0.11 \times$   
 226  $10^{16}$  molec  $\text{cm}^{-2}$ ,  $0.08 \times 10^{16}$  molec  $\text{cm}^{-2}$ , and  $0.15 \times 10^{16}$  molec  $\text{cm}^{-2}$  under D0 - D4 levels, respectively. The spatial  
 227 correlation between OMI and NoStress\_GC degrades with USDM, with  $R^2$  being smaller than 0.65 under D0 - D4  
 228 levels compared to  $R^2$  of 0.70 under N0. Worsening model performance with increasing drought severity suggests the  
 229 model lacks a process that changes with drought. As isoprene accounts for more than 80% of the contribution of non-  
 230 methane VOCs to the HCHO column in the southeast US (Palmer et al., 2003; Millet et al., 2006), the missing process  
 231 is most likely drought-induced changes in isoprene emissions.



212  
 213 **Figure 4. (a) Relative changes of regional-mean OMI HCHO column, NoStress\_GC simulated HCHO column, isoprene**  
 214 **emissions, anthropogenic benzene emission, anthropogenic toluene emission, and MODIS leaf area index (LAI) under**  
 215 **different drought levels in the southeast US. All data are scaled to their respective values at N0. The dotted lines are the**  
 216 **arithmetic mean of all grids, and the solid lines are the corrected mean excluding the missing area. (b) Regional-mean**  
 217 **tropospheric NO<sub>2</sub> columns from OMI and NoStress\_GC (solid lines), and their respective changes from non-drought (N0)**  
 218 **conditions (dashed lines). Note the different scales between the solid and dashed lines. The calculation is based on the grids**  
 219 **with the presence of all USDM levels.**

220 **Figure 4a** displays the relative changes in the regional mean HCHO column from OMI and NoStress\_GC, emissions  
 221 of isoprene and select anthropogenic VOCs from NoStress\_GC, and MODIS LAI as a function of USDM indices,  
 222 each scaled by its respective value at N0. The dotted line is the arithmetic mean of all available grids under each  
 223 dryness category, and the solid line is the mean for those grids with valid data in all dryness categories (i.e., removing



224 white areas shown in Figure 3). In either calculation, NoStress\_GC overestimates the relative increase of HCHO under  
225 D0-D4 by 10-50% compared to OMI. After correcting for no data areas at D2-D4, isoprene emissions in NoStress\_GC  
226 are 22.7%, 29.6%, 40.3%, 54.5%, and 56.0% higher in D0-D4 than N0. Note that LAI is observed to decrease by 5-  
227 10% per USDM level (**Figure 4a**), which makes the predicted increase of isoprene emissions with drought severity  
228 even more remarkable. This is likely caused by the increasingly higher temperature under droughts, given the  
229 exponential relationship of isoprene emissions with temperatures in MEGAN (Guenther et al., 2006).

230 By comparison, the modeled increase of the HCHO column with drought is 12-25%, more buffered than that of  
231 isoprene emissions. This is mainly caused by the loss of HCHO to photolysis, which is expected to increase under  
232 droughts with clearer skies (Wang et al., 2017; Naimark et al., 2021). In addition, HCHO formation also depends on  
233 the abundance of oxidants, such as hydroxyl radicals (OH) and NO<sub>x</sub>, that oxidize isoprene. High isoprene emissions  
234 can suppress OH under the low-NO<sub>x</sub> conditions that prevail in part of the SE US (Wells et al., 2020), leading to the  
235 buffered response in HCHO. Previous studies (Travis et al., 2016; Kaiser et al., 2018) showed that the NEI2011  
236 anthropogenic inventory in the model were biased high in the SE US and a reduction of 60% of NO<sub>x</sub> emission was  
237 suggested. By comparing to OMI NO<sub>2</sub> column, we found NoStress\_GC indeed overestimates NO<sub>2</sub> columns by ~42%  
238 in the SE US (**Figure 4b**), but the absolute bias in NO<sub>2</sub> is nearly constant from N0 to D4 (solid lines in Figure 4b).  
239 NO<sub>2</sub> column also shows an increasing trend from N0 to D4, yet with a much smaller rate (less than 9%) than HCHO.  
240 The model captures the relative change in NO<sub>2</sub> column with USDM (dashed lines in Figure 4b), despite the high bias  
241 due to the NEI2011 inventory, which indicates that the changes in natural sources of NO<sub>x</sub> (e.g., biomass burning and  
242 soil NO<sub>x</sub>) with droughts are well represented by NoStress\_GC. To further examine the effect of high bias of NO<sub>x</sub> on  
243 simulated HCHO, we conducted a sensitivity simulation of reducing the NEI2011 NO<sub>x</sub> emissions by 50% over the SE  
244 US during JJA 2011-2013. Most of the SE US was under droughts during the summertime of 2011-2012, while 2013  
245 was a less drought-stricken year (**Figure 1**). The sensitivity simulation resulted in a small reduction of the simulated  
246 HCHO column and the change was nearly constant among the USDM levels (**Figure S1a-b**), ranging from  $-0.04 \times 10^{16}$   
247 molec cm<sup>-2</sup> (2.6%) to  $-0.05 \times 10^{16}$  molec cm<sup>-2</sup> (3.5%). This rules out the possibility that the high NO<sub>x</sub> bias in the model  
248 is the reason for the overestimation of HCHO under droughts. Given the suppression effect of isoprene on OH and the  
249 well-captured NO<sub>2</sub> relative changes under droughts, the overestimation of HCHO columns by the model is unlikely  
250 to be caused by model chemistry, and more likely by the overestimation of isoprene emissions under drought  
251 conditions.

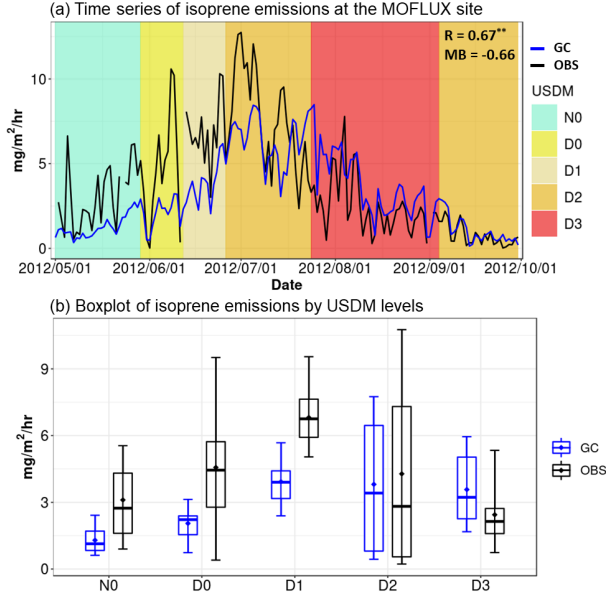
252 While oxidation of anthropogenic VOCs also produces HCHO, using benzene and toluene as indicator species, we  
253 found no change in anthropogenic VOC emissions with drought in the model (**Figure 4a**). This insensitivity rules out  
254 anthropogenic VOCs as a key driver of model overestimation of HCHO under drought conditions. If anything, we  
255 expect anthropogenic VOC emissions to increase during drought due to higher evaporative emissions driven by higher  
256 temperature and more fossil fuel consumption driven by more demand for space cooling. Wildfires are another  
257 important source that can lead to high HCHO levels, but their contributions to HCHO are more likely to be  
258 underpredicted in GEOS-Chem partly due to insufficient hydrocarbon emissions and the underrepresented fire plume  
259 chemistry (Alvarado et al., 2020; Liao et al., 2021; Zhao et al., 2022). A deeper planetary boundary layer (PBL) is

260 expected under droughts primarily caused by a larger sensible height flux released from dry soil (Miralles et al., 2014).  
261 Indeed, the MERRA-2 PBL height used in our simulation increases by 12.42%, 17.79%, 20.99%, 26.21%, and 29.52%  
262 from D0 to D4 relative to the value of 1589 m at N0 in the SE US during the midday (13:30 LT). Considering the  
263 PBL heights in MERRA-2 agree well with observations with only an overall 200 m low bias (Guo et al., 2021), we  
264 do not expect mixing heights to be the main cause of the high bias of HCHO column under drought conditions. To  
265 further quantify the effects of wildfires and PBL on the changes of HCHO column with drought, we conducted two  
266 additional sensitivity tests: (1) turning off the GFED4 wildfire emission inventory during 2011-2013 JJA, and (2)  
267 keeping PBL constant as in 2013 (normal year) during 2011-2012 (drought years) JJA. The results in **Figure S1c-d**  
268 show overall negligible changes in HCHO column in the SE US, which verifies our assumptions above.

269 In summary, the model overestimates HCHO increases during drought as compared to OMI. This overestimation is  
270 attributed to the model overestimation of isoprene emissions during drought. Drought stress effect on isoprene  
271 emissions is thus required in GEOS-Chem to resolve the discrepancy in HCHO responses to drought between OMI  
272 and the model.

### 273 **3.2 Isoprene flux measurement**

274 To further evaluate isoprene emissions in NoStress\_GC, we compared the isoprene flux measurements at the  
275 MOFLUX site (Potosnak et al., 2014; Seco et al., 2015) with predicted isoprene emissions at the model grid that  
276 contains the site. At the time of writing, the MOFLUX site is the only long-term, canopy-level, biogenic isoprene flux  
277 measurement site in the Northern midlatitude that sampled droughts. The site experienced multiple drought levels in  
278 the summer of 2012, which allows for the model-observation comparison across different drought severities as shown  
279 in **Figure 5**. The abnormal dry conditions (D0) started in early June, which developed to moderate drought (D1) in  
280 late June, worsened to severe drought (D2) and extreme drought (D3) in July-August, and bounced back to D2 in  
281 September (**Figure 5a**). The model generally captures the daily variability of isoprene emissions with a statistically  
282 significant correlation coefficient (R) of 0.67, but its biases differ by USDM levels. The model underestimates isoprene  
283 flux from N0 (bias of -1.81 mg/m<sup>2</sup>/hr) to D1 (bias of -2.89 mg/m<sup>2</sup>/hr), has a minimal bias (-0.47 mg/m<sup>2</sup>/hr) at D2, and  
284 changes to an overestimate at D3 (bias of 1.2 mg/m<sup>2</sup>/hr) (**Figure 5b**). While differences are expected when comparing  
285 a single-point flux measurement with the grid-mean model prediction, such differences most likely result in a  
286 systematic bias that should not relate to the temporal variability of drought. The fact that the model bias changes from  
287 being underpredicting to overpredicting as drought severity increases further confirms the importance of the model  
288 lack of a drought suppression effect on isoprene emissions during severe to exceptional droughts (D3 and D4). This  
289 is qualitatively consistent with that of the HCHO biases described above.



290

291 **Figure 5. (a) Comparison of daily time series of isoprene emissions observed at the MOFLUX site (OBS) and simulated by**  
 292 **MEGAN2.1 in GEOS-Chem (GC). The background is color-coded according to the USDM drought severity. R and MB at**  
 293 **the upright corner show the correlation coefficient and mean bias, respectively. (b) Boxplot of isoprene emissions separated**  
 294 **by USDM drought levels. The upper and lower whiskers represent the 90% and 10% quantiles, respectively.**

#### 295 4. Drought Stress Algorithm

296 The MEGAN2.1 isoprene emission routines in GEOS-Chem use a simplified mechanistic representation of the major  
 297 environmental factors controlling biogenic emissions (Guenther et al., 2012), in which the isoprene emission factor  
 298  $\gamma_{2.1}$  is the product of a canopy-related normalization factor ( $C_{FAC}$ ) multiplied by other factors representing light ( $\gamma_{PAR}$ ),  
 299 temperature ( $\gamma_T$ ), leaf age ( $\gamma_{AGE}$ ), LAI ( $\gamma_{LAI}$ ), carbon dioxide ( $CO_2$ ) inhibition ( $\gamma_{CO_2}$ ), and soil moisture ( $\gamma_{SM}$ ):

$$300 \gamma_{2.1} = C_{FAC} \gamma_{PAR} \gamma_T \gamma_{AGE} \gamma_{LAI} \gamma_{CO_2} \gamma_{SM} = \gamma_0 \gamma_{SM} \quad (1)$$

301 where  $\gamma_0$  is the product of the non-drought factors. Because of the lack of reliable soil moisture databases,  $\gamma_{SM}$  is  
 302 always set to be one in GEOS-Chem as in many other chemical transport models, which means no water stress term  
 303 in the standard model configuration (i.e., NoStress\_GC). We show above that NoStress\_GC overestimates isoprene  
 304 emissions and consequently HCHO column densities under drought conditions in the SE US. In this section, we  
 305 describe the approach whereby observational constraints from the MOFLUX isoprene flux measurement and OMI  
 306 HCHO were separately used to derive a drought stress factor  $\gamma_d$  which replaces  $\gamma_{SM}$  in Equation (1) to simulate the  
 307 response of isoprene emissions to drought in MEGAN2.1 implementation of GEOS-Chem (hereafter referring to as  
 308 GC/MEGAN2.1). The drought stress factor  $\gamma_d$  derived from the MOFLUX isoprene flux measurement is denoted as  
 309  $\gamma_{d\_MOFLUX}$  and that from OMI HCHO as  $\gamma_{d\_OMI}$ . Their corresponding simulations are referred to as  
 310 MOFLUX\_Stress\_GC and OMI\_Stress\_GC, respectively. In either algorithm, the underlying assumption is that the  
 311 GEOS-Chem model has no significant bias in predicting isoprene fluxes or HCHO columns due to factors other than  
 312 isoprene emissions under drought conditions. The assumption is reasonable because the GEOS-Chem model uses

313 reanalysis meteorology, state-of-the-science isoprene oxidation schemes, time-specific anthropogenic emissions and  
314 fire emissions, and natural emissions calculated online using model meteorology as described in Section 2.3. The  
315 discussion in Section 3.1 validated some aspects of the assumption such as NO<sub>x</sub> emissions, wildfire emissions, and  
316 PBL.

#### 317 **4.1 MOFLUX-based Drought Stress Algorithm**

318 The  $\gamma_{d\_MOFLUX}$  was derived following Jiang et al. (2018) by implementing photosynthesis and water stress parameters  
319 with a formula of:

$$320 \quad \gamma_{d\_MOFLUX} = \gamma_0 \gamma_{d\_isoprene} \begin{cases} \gamma_{d\_isoprene} = 1 & (\beta_t \geq 0.3) \\ \gamma_{d\_isoprene} = V_{cmax}/\alpha & (\beta_t < 0.3, \alpha = 77) \end{cases} \quad (2)$$

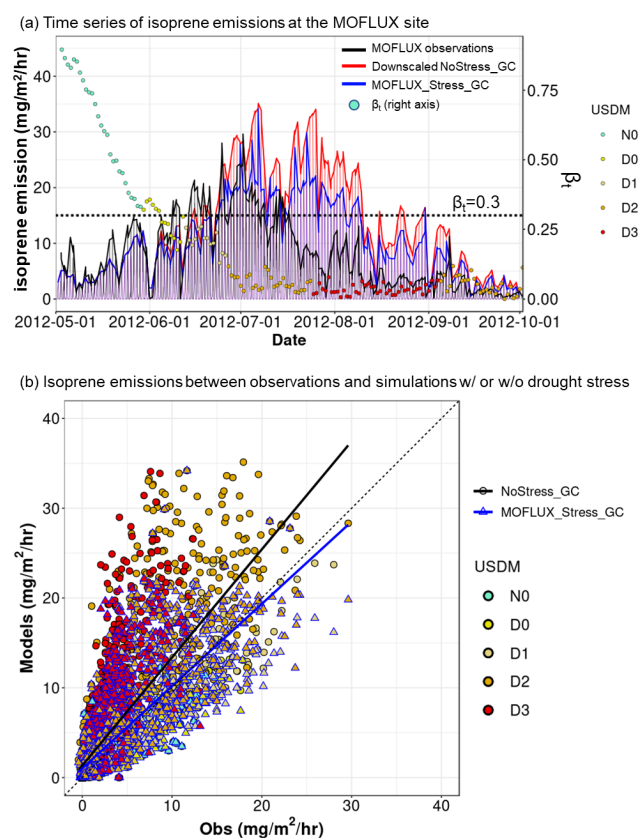
321 where  $V_{cmax}$  is the maximum carboxylation rate by photosynthetic Rubisco enzyme and  $\beta_t$  represents the water stress  
322 ranging from zero (fully stressed) to one (no stress). This simplified method intends to use the decreased  
323 photosynthetic enzyme activity to physiologically represent the variation in isoprene emissions under drought stress  
324 via dividing  $V_{cmax}$  by an empirical parameter  $\alpha$  when the water stress is below a threshold.

325 Since the default GEOS-Chem does not have these photosynthetic parameters, we adopted the ecophysiology module  
326 created by Lam et al. (2022) that is based on the photosynthesis calculation in the Joint UK Land Environmental  
327 Simulator (JULES; Best et al., 2011; Clark et al., 2011) as an online component in GEOS-Chem so that it simulates  
328 photosynthesis rate and bulk stomatal conductance dynamically and consistently with the underlying meteorology that  
329 drives GEOS-Chem. The outputs of  $V_{cmax}$  and  $\beta_t$  from the ecophysiology module were passed to MEGAN2.1 in GEOS-  
330 Chem to parameterize the drought stress according to Equation 2. In addition to GEOS-Chem meteorology, the  
331 ecophysiology module uses soil parameters from the Hadley Centre Global Environment Model version 2 – Earth  
332 System Model (HadGEM2-ES). In general, the implementation of the ecophysiology module much improved the  
333 simulated stomatal conductance and dry deposition velocity relative to site observations on average for seasonal  
334 timescales, but the  $\beta_t$  computed has not been calibrated to intermittent drought conditions. Instead of adopting the  
335 values of  $V_{cmax}$  and  $\beta_t$  from Jiang et al. (2018) which were based on the Community Land Model, we need to determine  
336 the  $\beta_t$  threshold and the  $\alpha$  value specific to GEOS-Chem with the ecophysiology module of Lam et al. (2022). To  
337 calibrate  $\beta_t$ , we first examined the statistical distribution of  $\beta_t$  at the MOFLUX grid (**Figure S2**) during May-September  
338 2011 and 2012 when multiple USDM drought categories occurred. Then we decided on a value of 0.3 as the threshold  
339  $\beta_t$  below which the drought stress will be triggered in the model because this value is greater than 75% quantile of all  
340 the  $\beta_t$  values from D0 to D3, thus capturing most of the drought conditions.

341 We note the observed isoprene flux at MOFLUX is consistently higher than predicted values during the non-drought  
342 period (e.g., N0 in Figure 5a). This systematic bias is expected because we compare the single-point observations with  
343 grid-mean isoprene emission fluxes. To correct the systemic bias, we scaled down the model isoprene emissions at  
344 the MOFLUX grid by a factor of 1.42, which is the ratio of the average hourly isoprene fluxes between observations  
345 and simulations at the MOFLUX grid during non-drought conditions ( $\beta_t > 0.3$ ). The factor of 1.42 was applied to

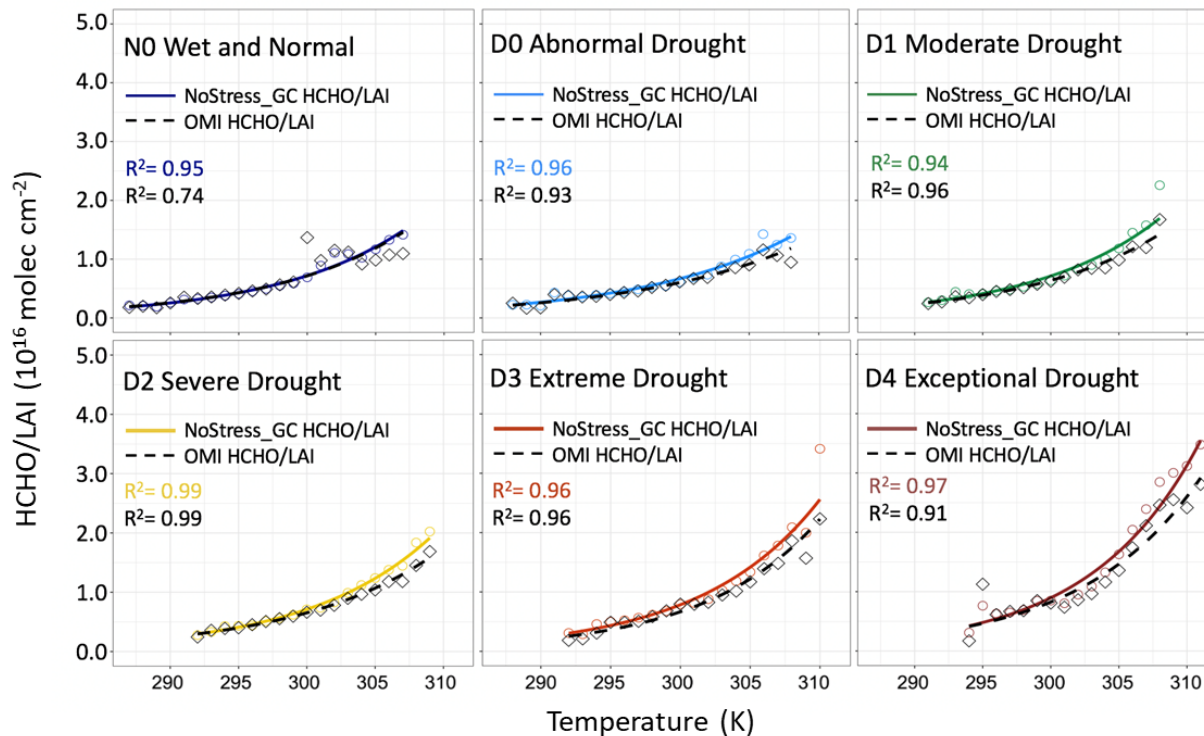
346 downscale modeled isoprene fluxes at the MOFLUX grid during the entire time series, including drought conditions.  
 347 The resulted time series are shown in **Figure 6a**. Based on the downscaled model prediction, we derived that  $\alpha=77$   
 348 under drought conditions ( $\beta_t < 0.3$ ), which minimized the mean bias under drought conditions between the modeled  
 349 and observed isoprene fluxes at the MOFLUX grid.

350 **Figure 6b** shows the comparison of the hourly NoStress\_GC and MOFLUX\_Stress\_GC isoprene emissions with  
 351 observations in May-September 2012. The overall mean bias is reduced from 2.05 mg/m<sup>2</sup>/hr to 0.01 mg/m<sup>2</sup>/hr despite  
 352 the fact that the stress factor is only applied to drought conditions. The correlation coefficient (R) and index of  
 353 agreement (IOA) also increase from 0.77 to 0.85 and from 0.80 to 0.93, respectively. All the changes in the comparison  
 354 metrics indicate the model simulations are improved considerably based on the single-point measurement.



355  
 356 **Figure 6. (a)** Hourly time series of isoprene emissions at the MOFLUX site from observations (black line) and simulations  
 357 with (MOFLUX\_Stress\_GC; blue line) and without drought stress (NoStress\_GC; red line; after downscaling). The dots  
 358 color-coded by USDM levels represent the daily values of  $\beta_t$  (right axis). The dashed line indicates the threshold of 0.3. (b)  
 359 Comparison of isoprene emissions between observations (Obs) and simulations with (MOFLUX\_Stress\_GC; blue-bordered  
 360 triangle) and without (NoStress\_GC; black-bordered circle) drought stress. Data are color-coded by USDM levels.

361 **4.2 OMI-based Drought Stress Algorithm**



362  
 363 **Figure 7. Response of HCHO/LAI ratio ( $10^{16}$  molec  $\text{cm}^{-2}$ ) to temperature (K) in different drought levels averaged over JJA**  
 364 **2005-2017. The colored solid line is the modelled NoStress\_GC HCHO/LAI ratio, and the black dashed line is the observed**  
 365 **HCHO/LAI ratio from OMI. The exponentially fitted formulas and the resulted coefficient of determination ( $R^2$ ) are**  
 366 **labelled in each subplot.**

367 Isoprene emission increases exponentially with temperatures below  $\sim 310$  K (Guenther et al., 2006) in the absence of  
 368 other stress factors such as drought. Indeed, an exponential relationship between biogenic isoprene emission per unit  
 369 LAI and temperature is predicted by MEGAN2.1 at all USDM levels (**Figure S3**). However, the predicted temperature  
 370 sensitivity is found to increase substantially with drought severity with no sign of plateauing or slow-down even under  
 371 the most severe drought conditions when MOFLUX measurements measured a decrease in isoprene emissions (c.f.  
 372 Figure 5). Similarly, we found NoStress\_GC overestimates HCHO sensitivities to high temperatures ( $> 300$  K) under  
 373 drought conditions (D0-D4) (**Figure 7**), but no such overestimation is seen under non-drought (N0) or low temperature  
 374 conditions during drought ( $< 300$  K). This indicates the role of drought stress on isoprene emissions is likely through  
 375 suppressing the dependence of emissions on temperatures during drought. Leaf level measurements conducted during  
 376 the 2012 drought at the MOFLUX site provide independent evidence of drought suppression of the isoprene response  
 377 to increasing temperature for less drought-resilient tree species (Geron et al., 2016). Taking advantage of these  
 378 empirical observations, we derived the OMI-based drought stress algorithm by minimizing the differences in HCHO  
 379 column sensitivities to temperatures between OMI and GEOS-Chem under drought conditions as shown in **Figure 7**.  
 380 When calculating the relationships between HCHO column densities and temperatures, we first scaled HCHO column  
 381 by LAI on a grid-by-grid basis to account for the regional differences in isoprene emissions due to different vegetation  
 382 coverage as well as the effect of LAI changes with drought (c.f. **Figure 4**). Each point in **Figure 7** represents the mean

383 HCHO/LAI ratio, denoted as  $\Omega$ , within each 1K temperature interval. We used exponential functions ( $\ln\Omega = kT + b$ ) to  
 384 separately fit the temperature (T) dependence of HCHO/LAI ratio ( $\Omega$ ) under different drought levels (**Figure 7**) for  
 385 both the model and OMI. The resulting formulas were listed in **Table 1** and the  $R^2$  of most fitting lines is greater than  
 386 0.9.

387 **Table 1. Fitted exponential formulas of NoStress\_GC and OMI HCHO/LAI ratio ( $\Omega$ ,  $10^{16}$  molec  $\text{cm}^{-2}$ ) to surface air**  
 388 **temperature (T, K), and fitted value of HCHO/LAI ratio at 290K, 300K, and 310K.**

USDM	NoStress_GC HCHO/LAI ( $\Omega$ , $10^{16}$ molec $\text{cm}^{-2}$ )				OMI HCHO/LAI ( $\Omega$ , $10^{16}$ molec $\text{cm}^{-2}$ )			
	Fitting Formula	290K	300K	310K	Fitting Formula	290K	300K	310K
N0	$\ln\Omega = 0.104T - 31.42$	0.25	0.72	2.03*	$\ln\Omega = 0.101T - 30.78$	0.26	0.72	1.97*
D0	$\ln\Omega = 0.091T - 27.83$	0.27	0.67	1.66	$\ln\Omega = 0.085T - 25.92$	0.26	0.60	1.40
D1	$\ln\Omega = 0.108T - 32.83$	0.24	0.71	2.10	$\ln\Omega = 0.100T - 30.56$	0.23	0.64	1.74
D2	$\ln\Omega = 0.110T - 33.33$	0.24	0.71	2.14	$\ln\Omega = 0.098T - 29.97$	0.24	0.65	1.75
D3	$\ln\Omega = 0.118T - 35.72$	0.24	0.78	2.56	$\ln\Omega = 0.121T - 36.62$	0.20	0.67	2.23
D4	$\ln\Omega = 0.125T - 37.59$	0.26	0.90	3.13	$\ln\Omega = 0.115T - 34.62$	0.26	0.83	2.60

389 \* Asterisk indicates that the temperature does not reach this value in actual data and is an extrapolated value.

390 As the fitting equations suggest, both NoStress\_GC and OMI HCHO/LAI ratios increase with temperature under all  
 391 conditions, but the former shows a higher sensitivity to temperature under drought conditions. This can be clearly seen  
 392 from the higher HCHO/LAI ratios of NoStress\_GC ( $\Omega_{GC}$ ; solid lines) than those of OMI ( $\Omega_{OMI}$ ; dashed lines)  
 393 especially when the temperature is greater than 300 K under D0-D4. To better explain this, we also calculated the  
 394 fitted value of HCHO/LAI at three temperatures of 290K, 300K, and 310K in **Table 1**. Since it is difficult for the N0  
 395 condition to reach a temperature of 310K, the values were extrapolated and marked with an asterisk in the table. The  
 396 results show that the model overestimates the temperature dependence at all drought levels. At 290K, all biases  
 397 between  $\Omega_{OMI}$  and  $\Omega_{GC}$  are less than  $0.05 \times 10^{16}$  molec  $\text{cm}^{-2}$ . At 310K, the bias between the two is  $0.06 \times 10^{16}$  molec  
 398  $\text{cm}^{-2}$  (3.0%) at N0 but increases by more than a factor of 4 to  $0.26 \times 10^{16}$  molec  $\text{cm}^{-2}$  (18.6%),  $0.36 \times 10^{16}$  molec  
 399  $\text{cm}^{-2}$  (20.7%),  $0.39 \times 10^{16}$  molec  $\text{cm}^{-2}$  (22.3%),  $0.33 \times 10^{16}$  molec  $\text{cm}^{-2}$  (14.8%), and  $0.53 \times 10^{16}$  molec  $\text{cm}^{-2}$  (20.4%) at D0-  
 400 D4 drought, respectively. As isoprene emission is a fixed function of temperature in MEGAN2.1, the overdependence  
 401 of HCHO column on temperature is caused by the previous two weeks' temperatures being higher under drought,  
 402 which leads to a higher value of  $\gamma_T$  reflecting the temperature “memory” effects on isoprene emissions (**Figure S4**).  
 403 Based on the fitted formulas in **Table 1**, the ratio between  $\frac{\Omega_{OMI}}{\Omega_{GC}}$  under each level from D0 to D4 can be derived by:

$$404 \quad \frac{\Omega_{OMI}}{\Omega_{GC}} = \frac{e^{k_{OMI}T + b_{OMI}}}{e^{k_{GC}T + b_{GC}}} = e^{(k_{OMI} - k_{GC})T} e^{(b_{OMI} - b_{GC})} \quad (3)$$

405 where  $k_{OMI}$  ( $k_{GC}$ ) and  $b_{OMI}$  ( $b_{GC}$ ) represent the slopes and intercepts of the formulas in **Table 1** for OMI (GC)  
 406 HCHO column;  $T$  is surface temperature, and  $e$  is the exponential constant. By averaging the values of  $k_{OMI}-k_{GC}$  and  
 407  $b_{OMI}-b_{GC}$  from D0 to D4, we can obtain:

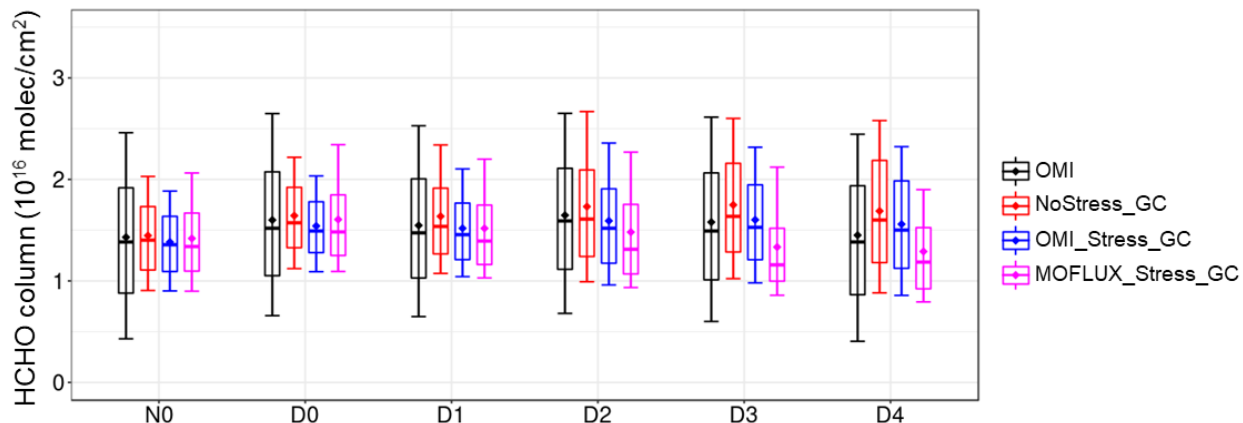
$$408 \quad \frac{\Omega_{OMI}}{\Omega_{GC}} = 380.10e^{-0.02T} (\beta_t < 0.6, T > 300 K) \quad (4)$$

409 where  $\beta_t < 0.6$  represents the 75% quantile of the  $\beta_t$  values from D0 to D4 for the whole SE US study region in JJA  
 410 2005-2017 (**Figure S2**).

411 The formula of  $\gamma_{d\_OMI}$  is thus:

$$412 \quad \gamma_{d\_OMI} = \gamma_0 \gamma_{d\_isoprene} \begin{cases} \gamma_{d\_isoprene} = 1 (\beta_t \geq 0.6 \text{ or } T \leq 300K) \\ \gamma_{d\_isoprene} = \frac{\Omega_{OMI}}{\Omega_{GC}} = 380.10e^{-0.02T} (\beta_t < 0.6, T > 300K) \end{cases} \quad (5)$$

413 Note the threshold of  $\beta_t$  in equation 5 is different from the value used by  $\gamma_{d\_MOFLUX}$  because all the SE US grids were  
 414 considered in deriving  $\beta_t$  for  $\gamma_{d\_OMI}$ . Another difference is that the factor is activated only if the temperature is higher  
 415 than 300K when significant biases between  $\Omega_{OMI}$  and  $\Omega_{GC}$  are found (**Figure 7**).



416  
 417 **Figure 8.** Boxplot of HCHO column statistical distributions for OMI observations (black) and different GEOS-Chem  
 418 simulations: without drought stress (NoStress\_GC; red) and with drought stress factors derived from MOFLUX  
 419 observations (MOFLUX\_Stress\_GC; blue) and from OMI HCHO constraints (OMI\_Stress\_GC; pink).

420 **Figure 8** compares the statistical distributions of HCHO column densities from OMI, NoStress\_GC,  
 421 MOFLUX\_Stress\_GC, and OMI\_Stress\_GC during May-September 2012 over the SE US. Compared to OMI,  
 422 NoStress\_GC simulation has a mean high bias of  $0.02 \times 10^{16}$  molec cm<sup>-2</sup> -  $0.24 \times 10^{16}$  molec cm<sup>-2</sup> during D0-D4. The  
 423  $\gamma_{d\_OMI}$  algorithm reduces the high bias to  $-0.05 \times 10^{16}$  molec cm<sup>-2</sup> -  $0.11 \times 10^{16}$  molec cm<sup>-2</sup>. By contrast, the  $\gamma_{d\_MOFLUX}$   
 424 algorithm reduces the HCHO simulations too much over the SE US and causes an overall underestimation of  $0.02 \times 10^{16}$



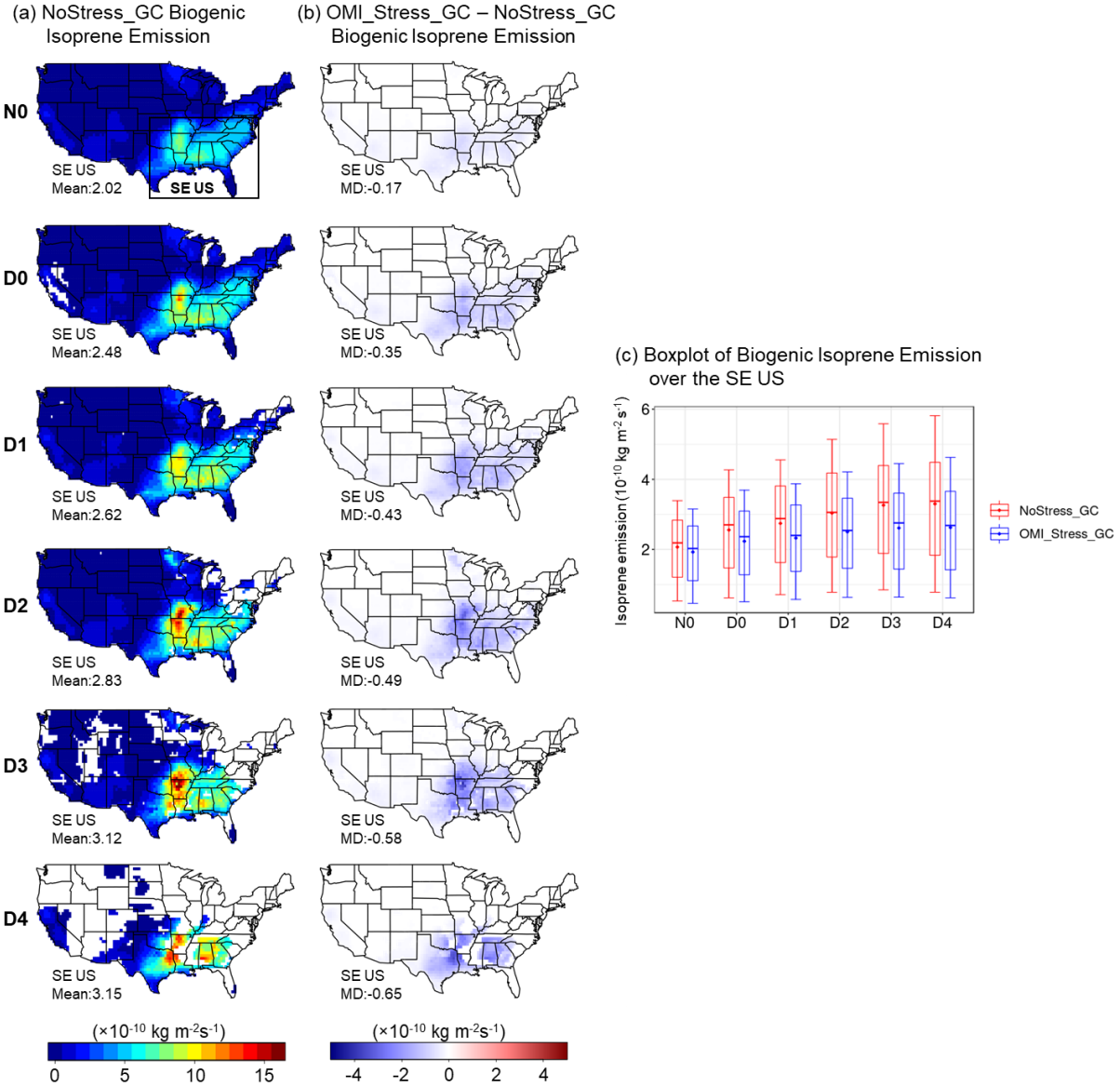
425 molec cm<sup>-2</sup> -  $0.25 \times 10^{16}$  molec cm<sup>-2</sup>. The  $\gamma_{d\_MOFLUX}$  algorithm also narrows the statistical distribution of HCHO as  
426 indicated by the smaller boxes and shorter whiskers compared to OMI. This suggests that the  $\gamma_{d\_MOFLUX}$  algorithm  
427 based on the single-site observations is incapable of representing the drought stress over the SE US, possibly because  
428 the MOFLUX site has thin soil layers and thus is vulnerable to water stress (Opacka et al., 2022). Isoprene emissions  
429 measured here are therefore more sensitive to droughts and the same extent of drought stress is likely too strong to be  
430 applied to other regions in the SE US. As a result, the  $\gamma_{d\_OMI}$  algorithm is used in the next section to further evaluate  
431 how this algorithm would change the responses of atmospheric compositions to droughts.

## 432 5. Changes in Simulated Biogenic Isoprene Emissions, HCHO, O<sub>3</sub>, and OA

433 In this section, we evaluated the changes in biogenic isoprene emissions and HCHO column densities by running a  
434 long-term (2005-2017, JJA) simulation, after adding the OMI-based drought stress factor for isoprene emissions  $\gamma_{d\_OMI}$   
435 in GEOS-Chem. Since isoprene is an important precursor for the formation of tropospheric O<sub>3</sub> and OA, maximum  
436 daily 8-hour average (MDA8) O<sub>3</sub>, and OA changes were also examined. We used the ComplexSOA mechanism in  
437 GEOS-Chem (Pye et al., 2010; Marais et al., 2016) which includes more detailed pathways of isoprene to secondary  
438 organic aerosols such as aqueous-phase reactive uptake and the formation of organo-nitrates.

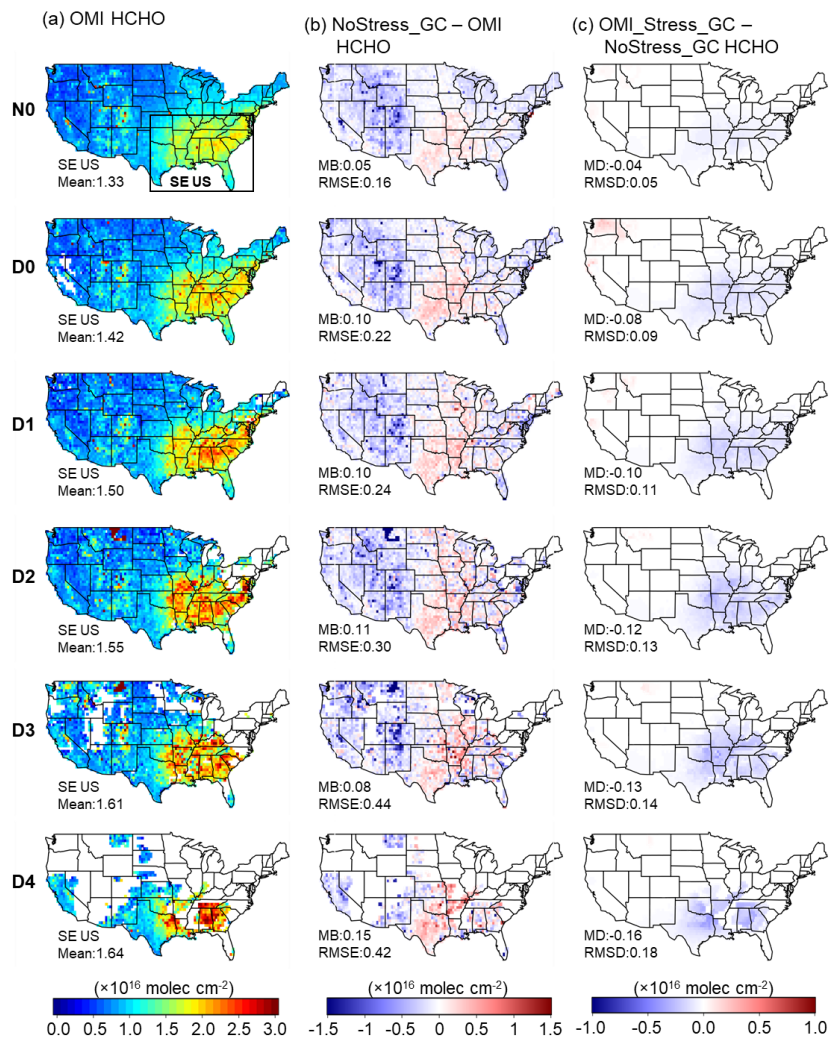
439 **Figure 9** shows the changes in biogenic isoprene emissions resulting from adding  $\gamma_{d\_OMI}$  drought stress in GEOS-  
440 Chem. Here we expanded the maps to the entire contiguous US to examine whether the drought stress algorithm can  
441 impose large changes on other US regions although such changes need to be interpreted with caution. The numbers at  
442 each panel indicate the means of isoprene emissions of NoStress\_GC and the mean differences (MD) relative to the  
443 OMI\_Stress\_GC over the SE US. As expected, the biggest decrease in isoprene emissions is found in the SE US with  
444 the regional-mean emissions reduced by  $0.17 \times 10^{-10}$  kg m<sup>-2</sup> s<sup>-1</sup> (8.60%),  $0.35 \times 10^{-10}$  kg m<sup>-2</sup> s<sup>-1</sup> (14.24%),  $0.43 \times 10^{-10}$  kg  
445 m<sup>-2</sup> s<sup>-1</sup> (16.57%),  $0.49 \times 10^{-10}$  kg m<sup>-2</sup> s<sup>-1</sup> (17.49%),  $0.58 \times 10^{-10}$  kg m<sup>-2</sup> s<sup>-1</sup> (18.66%), and  $0.65 \times 10^{-10}$  kg m<sup>-2</sup> s<sup>-1</sup> (20.74%)  
446 from N0 to D4, respectively (**Figure 9c**). Despite lowering emissions relative to NoStress\_GC, OMI\_Stress\_GC  
447 simulates an increase of isoprene emissions under drought conditions compared to non-drought in the SE US; the  
448 respective increases are  $0.28 \times 10^{-10}$  kg m<sup>-2</sup> s<sup>-1</sup> (15.20%),  $0.34 \times 10^{-10}$  kg m<sup>-2</sup> s<sup>-1</sup> (18.40%),  $0.49 \times 10^{-10}$  kg m<sup>-2</sup> s<sup>-1</sup> (26.47%),  
449  $0.69 \times 10^{-10}$  kg m<sup>-2</sup> s<sup>-1</sup> (37.46%), and  $0.65 \times 10^{-10}$  kg m<sup>-2</sup> s<sup>-1</sup> (35.23%) from D0 to D4 relative to N0 (**Figure 9c**). This  
450 increase results from the top-down constraints by the corresponding changes in OMI HCHO column densities with  
451 USDM and consequently exhibits the behavior of non-uniform increases with drought severity (e.g., peak increase of  
452 37.5% at D3, followed by a ~2% reduction at D4), which is consistent with the MOFLUX flux measurements.

453 For other regions, such as California and Minnesota, biogenic isoprene emissions decreased slightly by less than  
454  $0.5 \times 10^{10}$  kg m<sup>-2</sup> s<sup>-1</sup>. The smaller effect of the drought stress factor imposed on regions other than the SE US is  
455 understandable because of the lower isoprene emissions.



456

457 **Figure 9. Simulated biogenic isoprene emissions during JJA 2005-2017 by USDM dryness category by NoStress\_GC (a),**  
 458 **OMI\_Stress\_GC minus NoStress\_GC (b), and statistical distributions of SE US isoprene emissions between the two**  
 459 **simulations (c). Numbers at the bottom-left corner of each panel indicate the SE US (black box) regional mean of biogenic**  
 460 **isoprene emissions for NoStress\_GC (left column), and mean differences (MD) between OMI\_Stress\_GC and NoStress\_GC**  
 461 **(middle column).**

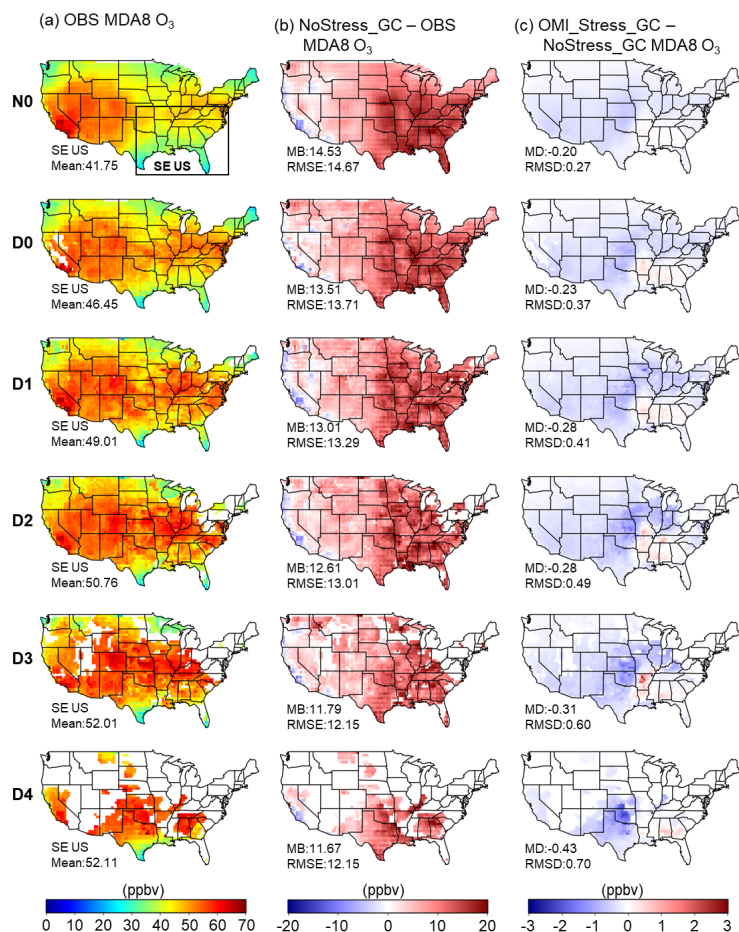


462

463 **Figure 10.** Mean HCHO column densities during JJA 2005-2017 by USDM dryness category for OMI (a), NoStress\_GC  
 464 minus OMI (b), and OMI\_Stress\_GC minus NoStress\_GC (c). Numbers at the bottom-left corner of each panel indicate the  
 465 SE US (black box) regional mean of OMI HCHO column (left column), mean bias (MB), and root mean square error  
 466 (RMSE) in HCHO column densities between NoStress\_GC and OMI (middle column), and mean differences (MD) and root  
 467 mean square deviation (RMSD) between OMI\_Stress\_GC and NoStress\_GC (right column). MD and RMSD are calculated  
 468 in the same way as MB and RMSE; the different names are used to distinguish between model-to-model comparison and  
 469 model-to-observation comparison, respectively.

470 The changes in the HCHO column are shown in **Figure 10**. Different from the overestimation in the SE US,  
 471 NoStress\_GC underestimates HCHO column densities in the western US compared to OMI (**Figure 10b**). This  
 472 negative bias should be interpreted with care because the scaling factor of 1.5 (c.f. section 2.2) is derived over the SE  
 473 US and may not hold in other regions. For the SE US overall, the drought stress factor reduces modeled HCHO  
 474 columns by  $0.08 \times 10^{16}$  molec  $\text{cm}^{-2}$  (5.43%),  $0.10 \times 10^{16}$  molec  $\text{cm}^{-2}$  (6.46%),  $0.12 \times 10^{16}$  molec  $\text{cm}^{-2}$  (7.22%) and  
 475  $0.13 \times 10^{16}$  molec  $\text{cm}^{-2}$  (7.62%),  $0.16 \times 10^{16}$  molec  $\text{cm}^{-2}$  (8.91%) under D0-D4, respectively, relative to NoStress\_GC  
 476 (**Figure 10c**). This leads to a better agreement with OMI as OMI\_Stress\_GC has nearly zero MB under D0-D4 (**Figure**  
 477 **S5**; MB =  $-0.05 \times 10^{16}$  molec  $\text{cm}^{-2}$   $\sim$   $-0.02 \times 10^{16}$  molec  $\text{cm}^{-2}$ ). The RMSE is also reduced by 3%-13% relative to the  
 478 NoStress\_GC simulation compared to observations. The changes in both metrics indicate that the drought algorithm

479 considerably improves the model performance in capturing the biogenic isoprene response to drought as evidenced by  
 480 HCHO column. Similar to the changes in biogenic isoprene emissions, the OMI\_Stress\_GC only slightly decreases  
 481 HCHO column densities (<5%) compared to the NoStress\_GC simulation in other US regions.

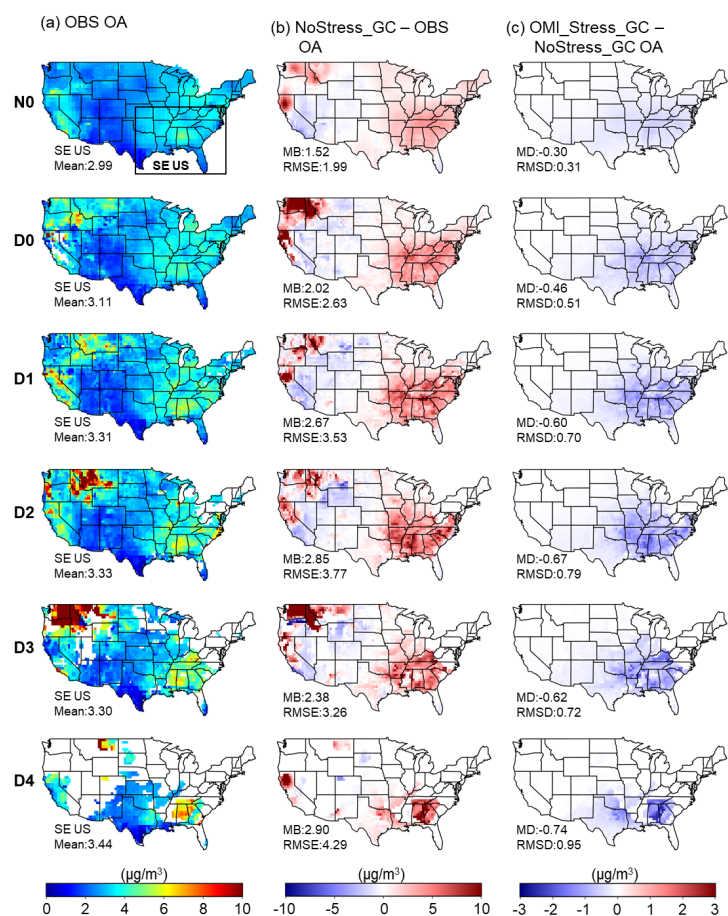


482

483 **Figure 11. Same as Figure 10 but for surface maximum daily 8-hour average (MDA8) O<sub>3</sub>.**

484 **Figure 11a** displays the observed MDA8 O<sub>3</sub> changes with USDM. Similar to the changes of the HCHO column with  
 485 USDM levels, O<sub>3</sub> in the SE US exhibits a gradual increase, relative to the mean of 41.74 ppbv at N0, of 4.70 ppbv,  
 486 7.26 ppbv, 9.01 ppbv, 10.26 ppb, and 10.36 ppbv under D0-D4, respectively. This is consistent with our previous  
 487 study (Li et al., 2022; Lei et al., 2022) which investigated O<sub>3</sub> changes with drought severity in more detail. The  
 488 NoStress\_GC simulation has a high bias in MDA8 O<sub>3</sub> across all USDM categories (**Figure 11b**). High positive bias  
 489 is a common issue of surface O<sub>3</sub> simulations in chemical transport models, which is a research question and can be  
 490 attributed to the uncertainties in various processes, such as NO<sub>x</sub> emissions, isoprene oxidation pathways, O<sub>3</sub> dry  
 491 deposition velocity, boundary layer dynamics (Fiore et al., 2005; Lin et al., 2008; Squire et al., 2015; Travis et al.,  
 492 2016; Travis and Jacob, 2019). Despite the systematic high bias, NoStress\_GC captures the increasing trend of MDA8  
 493 O<sub>3</sub> with increasing dryness but with a respectively smaller increment (relative to N0) of 3.62 ppbv, 5.67 ppbv, 7.01  
 494 ppbv, 7.41 ppbv, and 7.41 ppbv under D0 to D4. This discrepancy between NoStress\_GC and observations can also

495 be inferred from the fact that the MB between model and observations decreases from 14.53 ppbv at N0 to 11.67 ppbv  
 496 at D4 (**Figure 11b**). **Figure 11c** shows the difference in MDA8 O<sub>3</sub> between OMI\_Stress\_GC and NoStress\_GC. In  
 497 the SE US where isoprene emissions are the highest and reduced the most by the drought stress algorithm,  
 498 OMI\_Stress\_GC shows a small increase in MDA8 O<sub>3</sub> of less than 1 ppbv. This increase in O<sub>3</sub> can be explained by an  
 499 increase of OH resulting from reducing isoprene emissions under low-NO<sub>x</sub> conditions in the SE US (Wells et al.,  
 500 2020). For the SE US study domain as a whole, the change in MDA8 ozone was negligible but negative (regional  
 501 mean of -0.5 ppbv). Although the drought factor does not reduce the overall high bias, it makes the model more  
 502 consistent with the observed increment in MDA8 O<sub>3</sub> for the subregion with increased O<sub>3</sub> (e.g., 90–94°W, 32–35°N) as  
 503 drought severity increases. Since NO<sub>x</sub> has a high positive bias from the NEI2011 inventory (**Figure 4**), the  
 504 improvement of MDA8 in these regions is likely to be underestimated. Over northeastern Texas, Oklahoma, and  
 505 Kansas where isoprene emission is also reduced by the drought algorithm yet from a much lower emission base  
 506 compared to other SE US areas, OMI\_Stress\_GC simulates 1-3 ppbv lower MDA8 O<sub>3</sub> under drought conditions (D0-  
 507 D4), leading to a better agreement with observations. For regions with lower isoprene and higher NO<sub>x</sub> concentrations,  
 508 O<sub>3</sub> formation is more sensitive to the changes in isoprene, which explains the reduction in MDA8 O<sub>3</sub> caused by the  
 509 drought stress factor.



510

511 **Figure 12. Same as Figure 10 but for organic aerosol (OA).**

512 The changes in OA with USDM are shown in **Figure 12**. Observed OA in the SE US shows an average increase  
513 (relative to N0) of 0.12  $\mu\text{g}/\text{m}^3$ , 0.32  $\mu\text{g}/\text{m}^3$ , 0.34  $\mu\text{g}/\text{m}^3$ , 0.31  $\mu\text{g}/\text{m}^3$ , and 0.45  $\mu\text{g}/\text{m}^3$  under D0 to D4, respectively. The  
514 extremely high values over the northwest states (e.g., Washington and Montana) are likely associated with higher  
515 wildfire emissions under droughts (Wang et al., 2017). The NoStress\_GC simulation considerably overestimates OA  
516 in the SE US with an MB of 1.52  $\mu\text{g}/\text{m}^3$  (50.83%) at N0 and the overestimation becomes even higher to 2.02-2.90  
517  $\mu\text{g}/\text{m}^3$  (64.95%-85.58%) at D0-D4 (**Figure 12b**), thus causing an overprediction of the drought-OA relationship.  
518 Zheng et al (2020) reported a similar level of overestimation and attributed this to the overdependence of isoprene-  
519 derived secondary organic aerosol (SOA) on sulfate. As isoprene is one of the dominant sources of OA in the SE US  
520 (Xu et al., 2015; Budisulistiorini et al., 2016), our analysis suggests that the model overestimation of isoprene  
521 emissions under drought conditions is another reason for this high OA bias in the SE US. Indeed, the drought stress  
522 factor greatly improves the OA simulation by reducing the MB by 0.30  $\mu\text{g}/\text{m}^3$  (6.60%), 0.46  $\mu\text{g}/\text{m}^3$  (8.98%) 0.60  
523  $\mu\text{g}/\text{m}^3$  (10.07%), 0.67  $\mu\text{g}/\text{m}^3$  (10.85%), 0.62  $\mu\text{g}/\text{m}^3$  (10.88%), 0.74  $\mu\text{g}/\text{m}^3$  (11.71%) under N0 to D4 over the SE US  
524 relative to NoStress\_GC, thus lowering the MB to be within 1.22-2.18  $\mu\text{g}/\text{m}^3$  (40.82% - 65.52%; **Figure S5**) compared  
525 with observations. We also examined the change of three major SOA components in **Figure S6**. Anthropogenic SOA  
526 (ASOA) barely changes; isoprene SOA (ISOA) decreases the most as expected since the drought stress factor is  
527 applied to isoprene emissions only. Interestingly, terpene SOA (TSOA) also shows a slight decrease, suggesting  
528 positive feedback between ISOA and TSOA.

529 In summary, the OMI-based drought stress factor shows good performance in correcting the overestimation of  
530 biogenic isoprene in default GEOS-Chem simulations under drought conditions. The drought stress factor was  
531 constrained by the observed exponential fitting between the HCHO to LAI ratio and temperature, not by observed  
532 HCHO columns directly. It nearly eliminates the high HCHO bias compared with OMI observations in the SE US  
533 under drought conditions, which consequently improves the simulation of OA. MDA8 O<sub>3</sub> slightly increases in the  
534 areas with high isoprene emissions, leading to no improvement in model bias but a better agreement with the observed  
535 O<sub>3</sub> increment with drought severity. Places with lower isoprene emissions show an MDA8 O<sub>3</sub> reduction of 1-3 ppbv,  
536 indicating the region-specific O<sub>3</sub> responses to the changes of isoprene due to the nonlinearity of O<sub>3</sub> chemistry.

## 537 **6. Conclusions**

538 Using long-term (JJA 2005-2017) weekly USDM drought index and OMI HCHO column data over the SE US, we  
539 revealed a step-increase pattern of HCHO by 6.7%, 12.6%, 16.5%, 21.2%, and 23.2% from D0 to D4 relative to non-  
540 drought conditions (N0), respectively, which indicates the increasingly higher isoprene emissions with drought on a  
541 regional scale although the rate of increase decreases under severe droughts. Compared with OMI observations, the  
542 GEOS-Chem simulated HCHO column density exhibits a similar pattern, but the changes are 1.1-1.5 times higher  
543 with a respective increase of 9.90%, 15.1%, 19.5%, 21.8%, and 29.1% from D0 to D4. Since there are no big changes  
544 in anthropogenic VOCs under droughts, biogenic isoprene emissions are the key drivers for the increase of HCHO,  
545 and a drought stress factor is missing in the MEGAN2.1 biogenic inventory in the default GEOS-Chem simulations  
546 causing the overestimation of the HCHO changes in response to droughts.

547 The MOFLUX site provides the only long-term ground-based isoprene flux observations covering multiple drought  
548 severities. We developed a drought stress algorithm based on the MOFLUX site following Jiang et al. (2018), and the  
549 algorithm improves the HCHO simulation at the MOFLUX grid while underestimating HCHO after all the SE US  
550 grids are included. By comparison, the OMI-based drought stress algorithm derived from the different HCHO-  
551 temperature sensitivities between OMI and GEOS-Chem can reflect better spatial coverage and nearly removes the  
552 positive bias between OMI and the default simulations seen from a test simulation in May-September 2012 over the  
553 SE US.

554 The long-term simulation with the OMI-based drought stress factor can significantly reduce the biogenic isoprene  
555 emissions by  $0.35 \times 10^{-10} \text{ kg m}^{-2} \text{ s}^{-1}$  (14.24%),  $0.43 \times 10^{-10} \text{ kg m}^{-2} \text{ s}^{-1}$  (16.57%),  $0.49 \times 10^{-10} \text{ kg m}^{-2} \text{ s}^{-1}$  (17.49%),  $0.58 \times 10^{-10}$   
556  $\text{ kg m}^{-2} \text{ s}^{-1}$  (18.66%) and  $0.65 \times 10^{-10} \text{ kg m}^{-2} \text{ s}^{-1}$  (20.74%) from D0 to D4, respectively, which consequently leads to a  
557 better agreement between OMI and simulated HCHO column. Despite lowering emissions relative to the no-stress  
558 simulation, OMI\_Stress\_GC simulates a non-uniform trend of increasing isoprene emissions with drought severity  
559 that is consistent with OMI HCHO and MOFLUX. Relative to N0, the simulated increase in isoprene emissions is 15-  
560 18% under D0-D1, increasing to 26% at D2 and peaking at 37% at D3, followed by a slight decrease to 35% at D4.

561 The observed MDA8 O<sub>3</sub> and OA over the SE US show a similar increase pattern with HCHO. The OMI-based drought  
562 stress algorithm also helps reduce the mean bias of OA by  $0.30 \text{ }\mu\text{g/m}^3$  (6.60%),  $0.46 \text{ }\mu\text{g/m}^3$  (8.98%),  $0.60 \text{ }\mu\text{g/m}^3$   
563 ( $10.07\%$ ),  $0.67 \text{ }\mu\text{g/m}^3$  (10.85%),  $0.62 \text{ }\mu\text{g/m}^3$  (10.88%),  $0.74 \text{ }\mu\text{g/m}^3$  (11.71%) from N0 to D4 over the SE US compared  
564 with the high positive bias of more than  $2.02 \text{ }\mu\text{g/m}^3$  (50.83%) without the drought stress. By contrast, the MDA8 O<sub>3</sub>  
565 response to the reduced biogenic isoprene caused by the drought stress factor presents a spatial disparity due to the  
566 nonlinear O<sub>3</sub> chemistry. Places with high isoprene emissions show an increase of MDA8 O<sub>3</sub> by less than 1 ppbv, which  
567 slightly improves the simulated drought-O<sub>3</sub> relationship. For the regions with low isoprene emissions in the SE US,  
568 the drought stress factor reduces MDA8 O<sub>3</sub> by 1-3 ppbv.

569 This study reveals an increasingly higher level of biogenic isoprene under drought conditions over the regions with  
570 high vegetation coverage. As drought is predicted to become more frequent in a warming climate (Cook et al., 2018),  
571 it is essential to update current biogenic emission inventories by adding a drought stress factor and to improve the  
572 constraints of isoprene chemistry in the climate chemistry models in order to have a better projection of air quality in  
573 the future. We demonstrate the feasibility of applying satellite data to the development of drought stress algorithms  
574 when ground-based measurements are limited. Our attempt here is a top-down approach and used temperature as the  
575 only parameter to adjust isoprene emissions under drought conditions. The water stress threshold in our algorithm is  
576 used only as a triggering parameter; that is, it is used to determine whether a grid is in drought or not and thus can be  
577 replaced with other drought-identifying approaches. One issue with our approach is the type of temperature data to be  
578 used in the algorithm. Ideally, it should be leaf temperature because this is what regulates stomata at the process level.  
579 However, leaf temperature is not readily available from meteorological fields that drive CTMs. MEGAN uses 2 m air  
580 temperature to parameterize isoprene emissions, and thus our algorithm uses the same temperature. More biogenic

581 emission flux observations covering different vegetation types and drought severities will be helpful to better depict  
582 the relationships between biogenic VOCs and drought stress.

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### 590 **Data Availability**

591 GEOS-Chem model is publicly available at <http://www.geos-chem.org>. LAI is obtained from  
592 [http://geoschemdata.wustl.edu/ExtData/HEMCO/Yuan\\_XLAI/v2021-06/](http://geoschemdata.wustl.edu/ExtData/HEMCO/Yuan_XLAI/v2021-06/). O<sub>3</sub> and organic carbon observational data  
593 can be downloaded via [https://aq5.epa.gov/aq5web/documents/data\\_mart\\_welcome.html](https://aq5.epa.gov/aq5web/documents/data_mart_welcome.html). Observational isoprene  
594 measurements at MOFLUX are from Potosnak et al. 2014 and Seco et al. 2015 and are available upon request from  
595 co-author Alex Guenther. Satellite ΩHCHO is available publicly at  
596 [https://cmr.earthdata.nasa.gov/search/concepts/C1626121562-GES\\_DISC.html](https://cmr.earthdata.nasa.gov/search/concepts/C1626121562-GES_DISC.html).

### 597 **Competing interests**

598 The authors declare that they have no conflict of interest.

### 599 **Author contributions**

600 YW conceived the research idea. NL and WL conducted the model simulation and data analysis. JCYL and APKT  
601 created the ecophysiology module. AG, MJP and RS provided the field observations. All authors contributed to the  
602 interpretation of the results and the preparation of the manuscript

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